

# 1 **High regional and intra-generic variation in susceptibility to** 2 **mass bleaching in Indo-Pacific coral species**

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4 Paul R. Muir<sup>1,2</sup>, Terence Done<sup>1,3</sup>, J. David Aguirre<sup>4</sup>

5 <sup>1</sup> Biodiversity and Geosciences, Queensland Museum, 78-102 Flinders St. Townsville Qld 4810

6 Australia. <sup>2</sup> Global Change Institute, University of Queensland, Brisbane, Queensland 4072, Australia.

7 <sup>3</sup> Australian Institute of Marine Science, Townsville Qld 4810 Australia. <sup>4</sup> School of Natural and

8 Computational Sciences, Massey University, Auckland, New Zealand.

9

10 **Corresponding author:** [paul.muir@qm.qld.gov.au](mailto:paul.muir@qm.qld.gov.au)

11

## 12 **Abstract**

13 **Aim:** Mass bleaching is a major threat to reef-building corals and the ecosystems they  
14 underpin. Here, we identified regional variation in the nature of this threat in terms of  
15 the bleaching-susceptibility of individual coral species on some Indian Ocean and  
16 Pacific Ocean reefs.

17 **Location:** 22 sites in the central Great Barrier Reef, Australia (GBR) and 30 sites in  
18 the central Maldives Archipelago (MA).

19 **Time period:** 2002 for the GBR and 2016 for the MA.

20 **Major taxa studied:** Corals (Order Scleractinia).

21 **Methods:** Following marine heat-wave conditions, timed in-situ surveys were used to  
22 record bleaching responses (tissue colour) of large samples of individual coral  
23 colonies. Responses of 106 shared species were analysed for sites with similar levels  
24 of temperature stress, depth of occurrence and mortality. In each region, phylogenetic  
25 mixed models were used to partition the effects on responses of species of deep-time  
26 phylogeny, contemporary history and local-scale, among-site variability.

27 **Results:** Relative susceptibility to bleaching varied widely between regions: only 27  
28 of the 106 shared species were in the same quartile for relative susceptibility in both

29 regions. Few species were highly susceptible in both regions. Closely related species  
30 varied widely in their individual susceptibilities. Phylogenetic effects were moderate  
31 in both regions, but contemporary phenotypic effects indicative of recent evolution  
32 and acclimatization were greater in the MA, consistent with a stronger history of  
33 recent bleaching.

34 **Main conclusions:** The high regional and intra-generic variation in coral bleaching-  
35 susceptibility described here suggests there may be important differences in the extent  
36 to which these Indian and Pacific Ocean coral populations are exhibiting responses to  
37 deep-time evolutionary changes on the one hand, versus recent adaptation, on the  
38 other. There is a concerning scarcity of this type of data, by which coral species most  
39 at risk from bleaching in particular regions may be more accurately identified.

40

## 41 **1. INTRODUCTION**

42

43 Coral reefs are one of the most threatened ecosystems on the planet, with mass coral  
44 bleaching widely regarded as their greatest current threat (Heron et al., 2016).

45 Repeated, severe bleaching events have significantly damaged reefs across the tropics  
46 and subtropics, and mass bleaching events are predicted to increase in frequency and  
47 severity in the near future (Hoegh-Guldberg et al., 2007; Hughes et al., 2017; Hughes,  
48 et al., 2018a). Indeed, in 2020, the Great Barrier Reef (GBR) underwent its third  
49 severe event within three years (Hughes & Pratchett, 2020). Local and possible  
50 regional species extinctions related to repeated bleaching events have been reported  
51 (Glynn, 2011; Muir et al., 2017; Sheppard et al., 2020) and there is now a risk of  
52 global extinctions of reef-building coral species (Carpenter et al., 2008; IUCN, 2017;  
53 Richards & Day, 2018).

54           Rising sea temperatures, the ultimate cause of most mass bleaching events  
55 (Brown, 1997), are a global phenomenon, but the search for pragmatic management  
56 interventions are focused at local and regional scales (Anthony et al., 2015;  
57 Wooldridge & Done, 2009). While conventional management measures continue in  
58 the form of reducing impacts from agricultural runoff, overfishing and coastal  
59 development (GBRMPA, 2018; Steneck et al., 2019), current biological research  
60 seeks to support active interventions such as reef restoration, assisted evolution,  
61 assisted migration, hybridisation, seed banking, artificial refugia and assisted  
62 recruitment (van Oppen et al., 2017). However, given the scale of the problem, with  
63 around 750 species of scleractinian reef-corals (Hoeksema & Cairns, 2019) and over  
64 280,000 km<sup>2</sup> of reefs threatened around the globe (Spalding et al., 2001), these  
65 management and research efforts face a daunting task.

66

67           Despite the threat to reef-building corals globally, there is surprisingly little  
68 data on the bleaching susceptibility of individual species. Susceptibility has been  
69 documented mostly for coral genera (Baird et al., 2018; Guest et al., 2012; Hughes et  
70 al., 2018b; McClanahan et al., 2007; Marshall & Baird, 2000; Chou et al., 2016; Loya  
71 et al., 2001) with only a small number of species examined. While wide geographical  
72 variation in relative susceptibility has been established for entire coral communities  
73 (McClanahan et al. 2020, Thompson & van Woesik, 2009), there are few comparative  
74 studies of individual taxa and these are mostly restricted to comparisons at a genus-  
75 level (Guest et al., 2012, McClanahan et al., 2004; Pratchett et al., 2013). While much  
76 of the current research is aimed at active interventions to prevent loss of coral species  
77 following repeated mass bleaching events, the species most at risk from bleaching  
78 have not been well documented. For instance, the IUCN “Red List” which is widely

79 used to categorise the risk and conservation status of coral taxa, is compiled from  
80 mainly genus-level bleaching data from a small number of regions (Carpenter et al.,  
81 2008) and thus might be misrepresenting the status of many coral species and regional  
82 populations. The lack of data limits the extent to which management and interventions  
83 could be targeted to the most susceptible species.

84

85 Here, we investigate the susceptibility of a broad range of Indo-Pacific scleractinian  
86 species to moderate bleaching conditions in shallow reef waters of the central  
87 Maldives Archipelago (MA) during 2016 and the central Great Barrier Reef (GBR)  
88 during 2002. To the best of the authors' knowledge, our datasets are unique in  
89 capturing in-situ bleaching responses for diverse assemblages of coral species. For  
90 each region, phylogenetic mixed model analyses were used to estimate longer-term  
91 phylogenetically heritable effects, shorter-term effects commensurate with adaptation  
92 and/or acclimatisation (hereafter, contemporary phenotypic effects) and local-scale  
93 variation. Comparison between regions were restricted to corals exposed to similar  
94 levels of temperature stress at a similar depth of occurrence, but since mass bleaching  
95 is caused by a complex array of stressors that can occur over many weeks (Skirving et  
96 al. 2019) we avoided absolute comparisons. Instead, we compared the responses of  
97 the 106 species present in both regions in terms of their relative susceptibilities. We  
98 then assessed how the available data might be used to estimate species risk across  
99 wide geographic regions and closely related species. The findings have important  
100 implications for how coral bleaching, the greatest ongoing threat to these critically  
101 important, yet vulnerable species, is assessed and how research and management  
102 efforts might be prioritized.

103

## 104 2. MATERIALS AND METHODS

105

106 For the GBR in 2002 and the MA in 2016, sea temperatures exceeded bleaching  
107 thresholds over wide areas for several weeks, resulting in moderate mortalities over  
108 the ensuing months. To control for variation due to differing levels of temperate stress  
109 we restricted our analyses to sites with a Degree Heating Week (DHW) index of  
110 between 4 and 7 derived from satellite data (NOAA, 2017). This index is widely used  
111 to estimate the duration and severity of the temperature anomaly, with 4-7 DHW  
112 currently considered a moderate event (NOAA, 2017; Skirving, et al., 2019). Colony  
113 mortalities at selected sites reached 68% in the MA (Cowburn et al. 2019) and 86% in  
114 the GBR (Done et al. 2003). Since there can be a marked depth gradient in bleaching  
115 response (Muir et al., 2017; Baird et al., 2018), we focussed on corals assayed  
116 between 3 to 11 m depth in both regions. We further confined our analyses to the  
117 proportion of colonies with severe bleaching or recently partial or complete mortality  
118 to account for the possibility that less affected individuals recovered before our  
119 surveys (see below).

120

### 121 2.1 Field Surveys

122 A total of 14 locations (22 sites) in the central GBR and 10 locations (30 sites) in the  
123 central MA met the DHW criteria for inclusion in the analyses (Figure 1). At each  
124 site, timed surveys using SCUBA or snorkel dives of 45-80 minutes duration were  
125 conducted, recording bleaching responses and recent mortality of individuals  
126 according to standard methods (Supporting Information, Table S1). Species  
127 identifications were mostly made *in situ*, with high-resolution macro photographs or  
128 small samples taken from problematic colonies (under permit) for later analysis by

129 comparison with material available in the Queensland Museum, consultation with  
130 relevant experts and reference to standard texts (see Supporting Information for  
131 details). Both datasets were converted to current valid species (Hoeksema & Cairns,  
132 2019). For certain genera (e.g. *Montipora*, *Porites*, *Goniopora* and *Cycloseris*) it was  
133 not possible to accurately identify many of the species *in situ*, so these were assessed  
134 as grouped species and only considered in a supplementary analysis.

135

136 The MA locations were surveyed two months after the peak of the temperature  
137 anomaly, whereas the GBR locations were surveyed 4-6 months after. To assess the  
138 effect of survey timing, we analysed a supplementary dataset of 142 individual corals  
139 (five species) monitored *in situ* after a moderate bleaching event in the GBR (Baird &  
140 Marshall, 2002, data kindly provided by A. Baird as: DOI to be supplied). These data  
141 showed that colonies with mild to moderate bleaching had reasonable recovery,  
142 whereas severely bleached colonies showed little recovery or >50% tissue death  
143 (Supporting Information, Table S2). Thus, the proportion of severely affected or  
144 recently dead/partially dead individuals remained consistent from 2.5 to 5.5 months  
145 after peak temperatures. We were therefore confident that our analyses were robust to  
146 differences in the timing of our surveys.

147

## 148 2.2 Statistical Analyses

149 The proportion of severely affected or recently dead/partially dead individuals  
150 were analysed using phylogenetic generalised linear mixed models implemented in  
151 the MCMCglmm package (Hadfield, 2019) in R v3.6.1. Separate models were applied  
152 to each region to accommodate different fixed effects. For the MA analysis, fixed  
153 effects included depth (3-5 m and 9-11 m), shading (shaded microhabitats and

154 unshaded) and the depth  $\times$  shading interaction. For the GBR analysis, the fixed effects  
155 included two levels of depth (3-5 m and 6-9 m), seven levels of habitat/zone  
156 (nearshore-fringing reef, coastal-fringing reef, offshore-northern lagoon, offshore-  
157 northern reef slope, offshore- southern back reef, offshore-southern lagoon and  
158 offshore-southern reef slope) and habitat/zone  $\times$  depth interaction. The random effects  
159 used in both analyses were: phylogenetically heritable effects, contemporary  
160 phenotypic effects, and location. Only taxa with more than two records in each region  
161 were included in the analyses. See S2, Supporting Information for further detail.

162

163         The super tree of Huang and Roy (2016) was used in the phylogenetic mixed  
164 model analyses, it currently being the most comprehensive for the group. We also  
165 explored the effect of phylogenetic uncertainty on our results and undertook analyses  
166 using a restricted phylogeny based only on molecular data (Huang & Roy, 2016).  
167 These analyses are not considered further here as they recovered similar results to the  
168 analyses conducted using the super tree phylogeny, indicating our methods were  
169 robust to uncertainty in the construction of the super tree (Supporting Information,  
170 Tables S4 and S5). A supplementary analysis using the Bleaching Mortality Index  
171 (BMI, McClanahan et al., 2004) with species data pooled into genera was conducted  
172 to allow comparison with previous reports.

173

### 174 **3. RESULTS**

175

176 Analysis of 2888 individuals from 180 species, 55 genera and 14 families from the  
177 GBR showed moderate phylogenetically heritable effects, weak contemporary effects  
178 and weak local-scale effects (Table 1). For the MA, analysis of data from 4480

179 individuals representing 144 species, 47 genera and 14 families also showed moderate  
180 phylogenetic effects and minor local-scale effects. However, contemporary effects  
181 were 2.7 times greater than those for the GBR. In both regions, the species responses  
182 varied widely within genera (Figure 2). Genus *Acropora* had the largest number of  
183 species present in each region (35 species in both regions) and species were present in  
184 both the lowest and highest quartiles of bleaching response in each region (Figure 3).  
185 Species of the genera *Pocillopora*, *Echinopora*, *Leptastrea*, *Platygyra* and  
186 *Dipsastraea* showed a greater range of responses in the MA relative to the GBR  
187 (Figure 2).

188

189       Of the 218 species with sufficient data for analysis, only 106 species (64  
190 genera, 14 families) were present in both regions. Categorizing responses into  
191 quartiles based on their posterior median estimates from the phylogenetic mixed  
192 model analyses showed only 27 species (25.4%) were in the same quartile in both  
193 regions, 45 species differed by one quartile and 34 by two or more quartiles (Figure  
194 3). Only six species were in the most susceptible quartile in both regions. The  
195 phylogenetically heritable effects for each species showed a weak correlation between  
196 regions, markedly different from a slope of  $\pm 1.0$  expected if bleaching responses  
197 were phylogenetically constrained between regions (Figure 4). The contemporary  
198 effects also showed a low correlation between regions (Figure 4).

199

200       As contemporary effects were 2.7 times greater in the MA than the GBR  
201 (Table 1) and the region had a stronger history of recent bleaching, we examined  
202 species with the strongest contemporary effects consistent with recent adaptation or  
203 acclimatisation (“C” in Supporting Information, Table S3). A wide range of genera



204 (21 of 47 analysed) were represented in the upper quartile (38 species) for  
205 contemporary effects in the MA, including the ecologically significant genera:  
206 *Acropora*, *Dipsastraea*, *Cycloseris*, *Favites*, *Goniastrea*, *Hydnophora*, *Lobophyllia*,  
207 *Pavona* and *Pocillopora*.

208

209 For many genera, re-analysis of our data using the bleaching mortality index  
210 (BMI) at the level of genus (Supporting Information, Figure S2) showed results  
211 somewhat similar to those of the study that defined it (McClanahan et al., 2004).  
212 However, the relative responses were markedly different to those derived from the  
213 phylogenetic mixed models of our main analyses.

214

#### 215 **4. DISCUSSION**

216

217 We report the first regional comparison of bleaching susceptibility for a broad range  
218 of common Indo-Pacific reef corals at the level of species. Comparing shallow-reef  
219 populations exposed to similar levels of moderate thermal stress we found relative  
220 responses varied only slightly at local scales, but showed wide differences between  
221 the regions. The findings have important implications for how mass bleaching  
222 research is conducted and how this severe threat is addressed.

223

224 Prior to this study, the small amount of data available indicated low to  
225 moderate levels of regional variation in the relative susceptibility to mass bleaching  
226 for Indo-Pacific coral taxa. While entire assemblages varied widely in their overall  
227 susceptibility (McClanahan et al. 2020), the major coral genera often showed a  
228 similar hierarchy of susceptibility for different regions and bleaching events (e.g.

229 Loya et al. 2001, Marshall & Baird 2000, Hughes et al. 2017). The few direct  
230 comparative studies also showed low to moderate regional variation. For example,  
231 relative bleaching responses between East Africa and the north/central GBR were  
232 similar for 19 coral genera (McClanahan et al., 2004), while a comparison of sites  
233 within SE Asia and Australia found only two of 18 genera varied significantly in their  
234 relative susceptibility for two sites (Guest et al., 2012). Our study showed markedly  
235 higher levels of regional variation when susceptibility was considered at the level of  
236 species. Only 25% of the 106 species with sufficient data for analysis were present in  
237 the same quartile of susceptibility in the MA and GBR (Figure 3). This comparison  
238 was for an identical assemblage of currently accepted species for similar levels of  
239 temperature stress and depth of occurrence and for events that produced similar levels  
240 of mortality.

241

242 Phylogenetic analyses allow the dependence resulting from a shared evolutionary  
243 history along a phylogeny to be considered (Housworth et al., 2004) and here provide  
244 an indication of the origins of variation in bleaching susceptibility between the two  
245 assemblages. A strong inter-regional correlation in the phylogenetic contribution to  
246 bleaching response would have suggested that bleaching susceptibility was strongly  
247 phylogenetically constrained. By contrast, we found only a weak correlation (Figure  
248 4) suggesting divergence in deep-time that has affected susceptibility to bleaching in  
249 each region. This result, and indeed our general findings, are consistent with recent  
250 genetics studies that indicate wide divergence within many species of Indian and  
251 Pacific Ocean reef corals, fishes and other reef fauna (summarized Richards et al.,  
252 2016). Several cryptic regional endemic species or subspecies have also been resolved  
253 within species previously considered pan Indo-Pacific (Richards et al., 2016; Arrigoni

254 et al., 2020), suggesting regional divergence of reef faunas may be greater than  
255 currently recognised. Such cryptic regional endemics would account for some the  
256 regional variation documented here. The existence of more regionally restricted taxa  
257 would have implications for management and conservation: taxa with smaller range  
258 sizes are likely to have a greater risk of extinction (Richards et al., 2016), further  
259 strengthening the case for species-level and regional assessments of susceptibility.

260

261 Our analyses indicate that phylogeny imposes only limited constraints on the  
262 bleaching responses in each region. This suggests that regional divergence of entire  
263 lineages, as well as recent acclimation and adaptation, play a role in mediating  
264 bleaching responses. The greater proportion of variation partitioned to contemporary  
265 effects in the MA assemblage in 2016, relative to the 2002 GBR assemblage is  
266 consistent with this view (Table 1). The MA study area experienced several bleaching  
267 events in the two decades before our surveys (summarized Muir et al., 2017),  
268 potentially providing a strong pressure for contemporary phenotypic responses. By  
269 contrast, although several events had been reported on the GBR in the two decades  
270 before our 2002 surveys (Berkelmans & Oliver, 1999; Berkelmans et al., 2004), our  
271 study sites had not yet been exposed to the globally widespread and severe events of  
272 the present century, and thus any associated selective tendency towards reduced  
273 susceptibility to subsequent bleaching.

274

275 Reduced susceptibility following a series of recent bleaching events has been  
276 reported previously, but mainly for communities (Thompson & van Woesik, 2009;  
277 McClanahan et al. 2020) or genera. Relatively low bleaching responses for the genera  
278 *Pocillopora* and *Acropora* in Singapore and Malaysia (Guest et al., 2012, Chou et al.,

279 2016) and for *Acropora* and *Montipora* in French Polynesia (Pratchett et al., 2013)  
280 have been attributed to adaptation/acclimation following several bleaching events.  
281 Here, we partitioned individual species responses between deep-time and  
282 contemporary effects to better understand the potential for adaption and/or  
283 acclimation. This method could provide a means of selecting species for different  
284 streams of management and intervention. For example, in the MA the species with the  
285 strongest contemporary phenotypic resistance following repeated bleaching events  
286 (“C” in Supporting Information, Table S3) are potential candidates for interventions  
287 such as reef restoration (Chamberland et al., 2017; dela Cruz & Harrison, 2017),  
288 assisted migration (van Oppen et al., 2017) and isolation of genes that confer  
289 resistance (van Oppen et al., 2015). Conversely, species with weak contemporary  
290 effects and higher bleaching responses (“P” in Supporting Information, Table S3) are  
291 potentially at increased risk and require heightened monitoring and first consideration  
292 for interventions such as artificial refugia (Coelho et al., 2017), assisted evolution  
293 (van Oppen, et al., 2015) and ‘seed-banking’ of cryopreserved gametes or larvae  
294 (Daly et al., 2018). Clearly, more data from other events and regions are required to  
295 fully realize the potential offered by this approach: our categories are only applicable  
296 for the MA and will likely change as the populations are challenged by further  
297 bleaching events.

298

299 High regional and intra-generic variation in bleaching susceptibility have  
300 important implications for the prioritization of reef management and conservation.  
301 The IUCN Red List (Carpenter et al., 2008) is widely used to set these priorities, but  
302 is currently based on mostly genus-level bleaching data from few locations. Obtaining  
303 the data to address this issue may be challenging as many species require expert

304 taxonomic input for accurate identification, the dominant genera *Acropora*, *Porites*  
305 *and Montipora* being particularly difficult. In addition, high local-scale variation can  
306 be problematic for *in situ* bleaching assessments (Chou et al., 2016; Cantin &  
307 Spalding, 2008), but our methods produced low local-scale variation and consistent  
308 species responses within each region (Table 1). Thus, despite the challenges there are  
309 the means to better inform management priorities and mitigate species loss.

310

311 By assessing coral bleaching responses at greater taxonomic resolution over a  
312 wide range of taxa, we found much greater variation within genera and between  
313 regions than had previously been reported. By analysing these data using  
314 phylogenetic comparisons we found evidence for this variation having origins in both  
315 deep-time and recent adaptation and/or acclimation in response to repeated events.  
316 Overall, we conclude that there is an alarming lack of reliable data on species'  
317 susceptibility with which to address the coral bleaching phenomenon, the greatest  
318 current threat to these already endangered species. Somewhat encouragingly, we  
319 found that only a small proportion of species were highly susceptible in both regions  
320 and that bleaching responses were weakly phylogenetically constrained, with some  
321 potential for adaptation and/or acclimation to moderate levels of bleaching stress.

322

323

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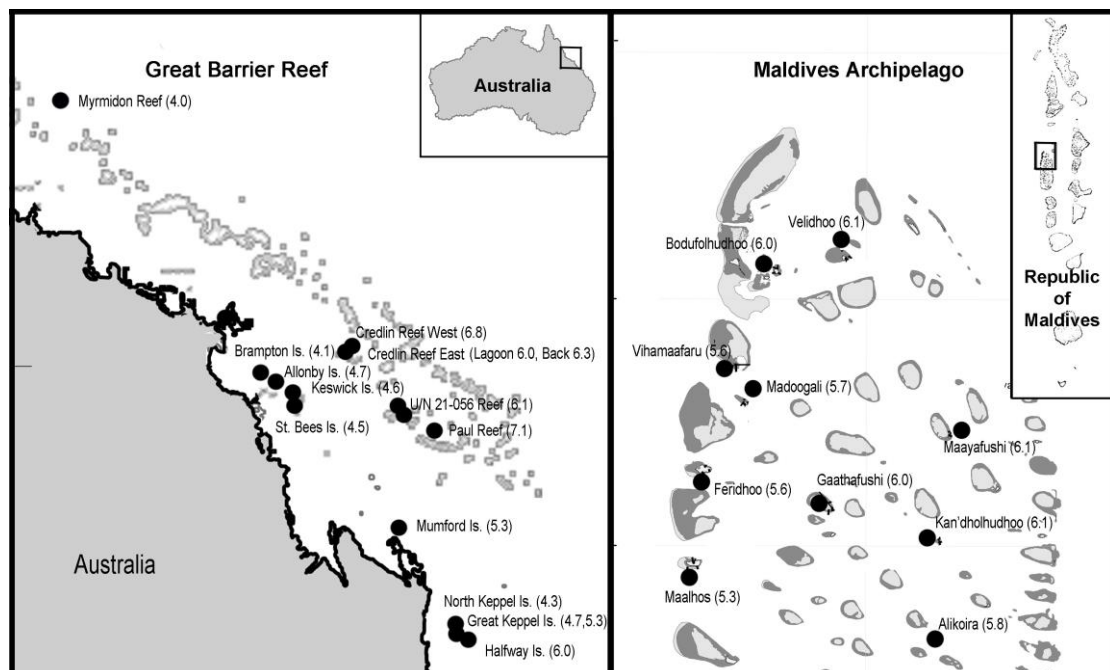
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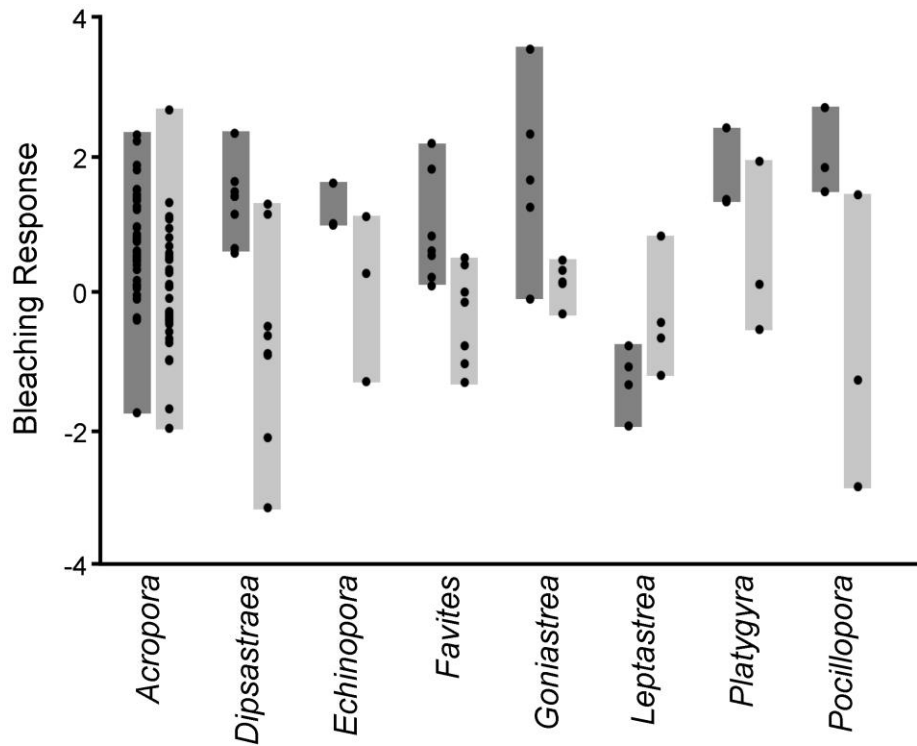
## FIGURES AND TABLES

Region	Proportion of Variation Explained		
	Phylogenetic	Contemporary	Local-Scale
Maldives	0.367 (0.065,0.638)	0.115 (0.047, 0.258)	0.032 (0.011,0.105)
GBR	0.454 (0.264, 0.658)	0.042 (0.094,0.103)	0.025 (0.006,0.095)

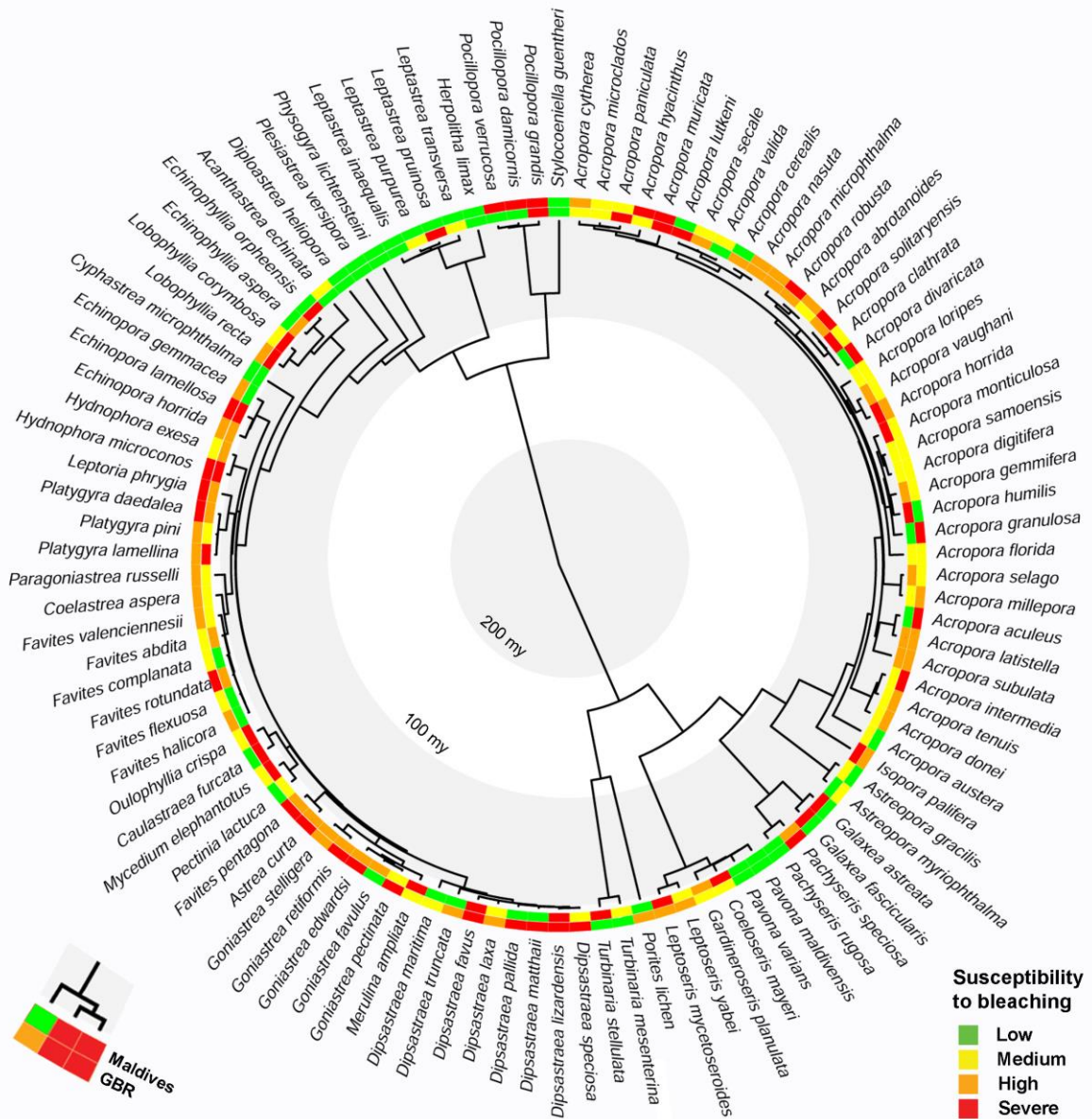
**Table 1.** Summary of phylogenetic mixed model analyses for the bleaching responses of scleractinian species in the Maldives Archipelago and Great Barrier Reef (GBR). The proportion of variation in bleaching explained by phylogenetically heritable (Phylogenetic), recent adaptation/acclimatization (Contemporary) and local-scale effects. Proportions given as posterior medians with 0.025 and 0.975 quantiles in brackets.



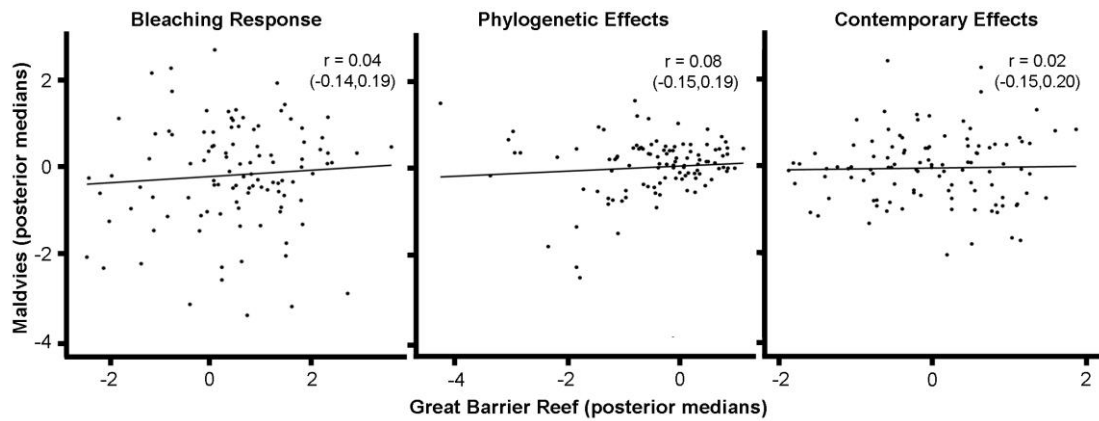
**Figure 1.** Locations where scleractinian coral bleaching responses were surveyed in the Great Barrier Reef and Maldives Archipelago. The severity of the bleaching at each location is given as peak degree heating weeks (DHW) in brackets.



**Figure 2.** Species' bleaching responses (●) varied widely within many of the main scleractinian coral genera in both the Great Barrier Reef (■) and Maldives Archipelago (■). Responses shown as posterior medians derived from phylogenetic mixed model analyses. The genera shown had more than two shared species between regions with sufficient data for analysis.



**Figure 3.** Susceptibility to moderate coral bleaching for scleractinian species in the Maldives Archipelago (inner ring) and Great Barrier Reef (outer ring) plotted onto the currently accepted phylogeny of the group (Huang and Roy, 2016). Susceptibility was categorized using quartiles of the species' posterior medians derived from separate phylogenetic mixed models that analysed the proportion of the population that was severely affected.



**Figure 4.** The bleaching responses of scleractinian species varied widely between regions, with phylogenetic mixed model analyses indicating that the variation had origins in both deep-time, phylogenetically heritable effects and short-term contemporary effects. These effects were poorly correlated between regions. Responses are shown as posterior medians (see Supporting Information, Figure S3 for details).