Close-kin methods to estimate census size and effective population size

Robin S. Waples[1] and Pierre Feutry[2]

[1]Northwest Fisheries Science Center, National Marine Fisheries Service, National Oceanic and Atmospheric Administration, Seattle, WA, 98112 USA
[2]Commonwealth Scientific and Industrial Research Organisation, Oceans & Atmosphere, GPO Box 1538, Hobart 7001, TAS, Australia

Running head: Close-kin, N, and N_e

Corresponding author:
Robin Waples (robin.waples@noaa.gov)
Northwest Fisheries Science Center
2725 Montlake Blvd. East
Seattle, WA, 98112 USA
(206) 860-3254

August 2021

Abstract

The last two decades have witnessed rapid developments and increasing interest in use of (1) genetic methods to estimate effective population size (N_e) and (2) close-kin mark-recapture (CKMR) methods to estimate abundance based on the incidence of close relatives. Whereas N_e estimation methods have been applied to a wide range of taxa, all CKMR applications to date have been for aquatic species. These two fields of inquiry have developed largely independently, and this is unfortunate because deeper insights can be gained by joint evaluation of eco-evolutionary processes. In this synthesis, we use simple analytical models and simulated pedigree data to illustrate how various factors (life-history traits; patterns of variation in individual reproductive success; experimental design; stochasticity; marker type) can affect performance of the estimators. We show that the N_e/N ratio and the probability of a close-kin match both depend on a vector of parental weights that specify relative probabilities that different individuals will produce offspring. Although age-specific vital rates are central to both methodologies, for CKMR they can potentially bias abundance estimates unless properly accounted for, whereas they represent the signals of genetic drift that N_e estimation methods depend upon. Coordinating N_e and CKMR estimation methods using the same or overlapping datasets would facilitate joint evaluation of both the ecological and evolutionary consequences of abundance.

Key words: siblings, parentage analysis, reproductive skew, mark-recapture, genetics, abundance

Table of Contents

1 | INTRODUCTION

Quantifying abundance is central to many aspects of organismal biology, with two metrics being particularly important. The number of individuals (census size, N) informs a wide range of studies in basic and applied ecology, including population growth rates, competition and predation, immigration and emigration, behavior, extinction risk, harvest management, and ecosystem function (Krebs 2009). The evolutionary analogue of census size is effective population size (N_e), which determines rates of genetic drift, inbreeding, and loss of genetic variability and mediates the effectiveness of natural selection (Charlesworth 2009). N_e is usually smaller than N and sometimes much smaller (Frankham 1995; Hauser and Carvalho 2008; Palstra and Fraser 2012), which means that some abundant populations could nevertheless experience rapid genetic change—a scenario that has been proposed for many marine species with huge populations and high fecundity (Hedgecock and Pudovkin 2011).

Both key population-size metrics are challenging to estimate in nature. Many animal species are vagile, cryptic, or rare, making detection and enumeration difficult (Seber 1986; Taylor and Wade 2000; Holmes 2001; Doak et al. 2005), and this is particularly true in the marine realm (**"Managing fisheries is hard: it's like managing a forest, in which the trees are invisible and keep moving around;" J. Shepard; http://jgshepherd.com/thoughts/)**. Effective population size is defined in terms of key demographic parameters (mean and variance in offspring number; Crow and Denniston 1988, Caballero 1994), but collecting these data often entails formidable logistical challenges, especially for long-lived species. As a consequence, indirect genetic methods have been used to estimate N_e since the early 1970s (Krimbas and Tsakas 1971; Luikart et al. 2010). Similarly, mark-recapture (MR) methods to estimate abundance have been a cornerstone of wildlife ecology for the past half century (Jolly 1965; Seber 1982; Williams et al. 2002). For species that are difficult to capture and/or physically tag, naturally occurring genetic markers can be used to identify "recaptures" of the same individual from passively-collected samples such as feces or hair. For example, Kendall et al. (2008) used genotypes derived from hair samples to obtain a robust estimate of abundance of a threatened population of grizzly bears (*Ursus arctos*) in Glacier National Park, USA, and Buckworth et al. (2012) used 'genetagging' to estimate harvest rate and catchability in Spanish mackerel (*Scomberomorus commerson*) from northern Australia.

More recently, researchers have begun to capitalize on the realization that, if naturally-

occurring genetic marks are used, it is not necessary to capture individuals more than once, and

the single "capture" can simply involve collecting DNA from a dead individual (or non-lethally

from a living one). Close kin can also be used in a MR framework because individuals "mark"

their relatives with shared genes (Skaug 2001). The two kinship categories that have been used

to date with close-kin mark-recapture (CKMR) are parent-offspring pairs (POPs) and siblings.

Successful application of CKMR methods to a number of fish species (Rawding et al. 2014;

Bravington et al. 2016a, 2019; Hillary et al. 2018; Bradford et al. 2018; Ruzzante et al. 2019;

Thomson et al. 2020), together with refinements in statistical methodology (Bravington et al.

2016b; Skaug 2017; Conn et al. 2020), have generated a great deal of interest in using this

approach more broadly to study other taxa (e.g., Stewart et al. 2018; Oleksiak and Rajora 2020).

Basic principles of CKMR have also been used to place an upper bound on the pre-Columbian

human population size in the Caribbean (Fernandes et al. 2020).

Considerable advances in the estimation of effective size have also been made over the

past ~15 years. Prior to the mid-2000s, most genetic estimates of N_e used the temporal method

(KIrimbas and Tsakas 1971; Nei and Tajima 1981), which requires at least two samples spaced

in time. Most recent applications, however, use one of two single-sample methods (Palstra and

Fraser 2012): a bias-adjusted method based on linkage disequilibrium (LD; Waples and Do

2008) or a method based on frequencies of siblings (Wang 2009), and these methods have been

widely applied to marine, freshwater, and terrestrial species. Precision and bias of both single-

sample methods has been extensively evaluated using simulated data (Waples and Do 2010;

Tallmon et al. 2010; England et al. 2011; Gilbert and Whitlock 2015; Wang 2016), as have

effects of age structure on LD estimates (Robinson and Moyer 2013; Waples et al. 2014).

Although these recent advances in methods for estimating N and N_e have occurred

simultaneously, they have also occurred largely independently. That is unfortunate, as CKMR

estimates of abundance and related quantities depend on the same key demographic parameters

that are needed to calculate N_e. Furthermore, genotypes for individuals sampled for CKMR also

can be used to generate indirect genetic estimates of effective size, as was recently demonstrated

for southern bluefin tuna, *Thunnus maccoyii* (Waples et al. 2018a). Recent work by Akita

(2020a,b) has made a good start in this direction for some specialized scenarios, and here we

provide a more general treatment. Although POPs-based CKMR has been successfully applied

to semelparous species (Rawding et al. 2014), most species are iteroparous with overlapping

generations, and that is our focus here.

Bravington et al. (2016b) provide a brief yet authoritative introduction to the statistical

underpinnings of CKMR, but not in a format that is very accessible to biologists interested in

practical applications. Our synthesis here is motivated by the recognition that (1) a full

understanding of the evolutionary ecology of natural populations requires being able to quantify

both abundance and effective size; and (2) as a consequence, it is important to maximize the

synergistic effects of the two types of analyses. We begin by reviewing basic principles of

population demography that are central to both CKMR and N_e. We then consider how the same

genetic data can be used both to estimate abundance and related quantities via CKMR and to

estimate effective size using the popular single-sample estimators. Next, we introduce some

simple analytical tools that can be used to allow unbiased, closed-form CKMR estimation of

abundance when the number of factors affecting the probability of a close-kin match is limited.

Simulated data are then used to model these relatively simple scenarios, and results illustrate

how some of the most common covariates (annual mortality; changes in fecundity with age;

variance in reproductive success; non-random sampling; intermittent breeding) can affect bias

and precision of the estimates and how these factors can be accounted for. We show that the

N_e/N ratio and the probability of a close-kin match both depend on a vector describing the

relative probabilities different individuals have of producing offspring. We also show that,

although age-specific vital rates are central to both types of analyses, they have different

consequences for performance of CKMR estimates than they do for estimates of effective size.

2 | EVOLUTIONARY DEMOGRAPHY

2.1 | Vital rates

General features of population demography that underpin models used here are described

below (see Table 1 for notation and definitions). The focal population is assumed to be isolated

and iteroparous, with separate sexes. Time periods, assumed to be years, are discrete and age is

indexed by x. Reproduction follows the seasonal birth-pulse model (Caswell 2001). At age x,

individuals produce on average b_x offspring and then survive to age $x+1$ with probability s_x.

Both b_x and s_x can differ with age and sex. First reproduction occurs at age α (age at maturity)

and maximum age is ω. Often, offspring are not enumerated until they reach a certain age, at

which point they are considered "recruits." We designate z as age at recruitment ($1 \leq z \leq \alpha$) and

scale b_x to production of offspring that survive to age z. Cumulative survival from age z through

age x is $l_x = \prod_{i=z}^{x-1} s_i$, with $l_z = 1$. Letting N_x be the number of individuals of age x alive at any

given time, common definitions of census size include total abundance, $N_T = \sum_{x=1}^{\omega} N_x$ and number

of adults, $N_A = \sum_{x=\alpha}^{\omega} N_x$. Except as noted we focus on N_A.

2.2 | Reproduction

Both effective population size and CKMR depend on the population pedigree, which

describes genealogical relationships among individuals. Each year, the actual pedigree of a

population is one realization of the stochastic process of reproduction and recombination. We

use a generalized Wright-Fisher (WF) model of reproduction in which all adults have the

potential to breed and produce offspring. Let k_i be the reproductive output (RO = number of

offspring) produced by the ith parent, and let μ_k and σ_k^2 be the mean and variance of k across N_A

potential parents. In the classical WF model of random reproductive success, each adult has an

equal and independent probability of being the parent of each offspring; this produces a

multinomial distribution of offspring number, with $E(\sigma_k^2) = \mu_k(N_A-1)/N_A$ (Crow and Denniston

1988), which is closely approximated by the Poisson variance $\sigma_k^2 = \mu_k$. The generalized model

(Waples 2020) differs in allowing adults to have unequal probabilities of parenting a given

offspring. Define a vector of relative parental weights, $\mathbf{w}$, where w_i is the relative probability that

adult i will be the parent of a given offspring. These relative weights can be converted to a

vector of absolute (standardized) parental weights, $\mathbf{W}$, where $W_i = w_i/\Sigma w_i$ and $\Sigma W_i = 1$. If all

weights are equal, we recover the original WF model, with $E(\sigma_k^2) \approx \mu_k$; in real populations,

however, different individuals generally will have different parentage probabilities, which leads

to overdispersed variance in reproductive success ($\sigma_k^2 > \mu_k$) relative to the random expectation.

This in turn has important consequences for both N_e and CKMR.

Within a single year of reproduction, two components contribute to variation in

individual weights: (1) an among-age effect, which arises from systematic changes in expected

fecundity with age; and (2) a within-age effect, which reflects difference in expected

reproductive success of individuals of the same age and sex. The among-age effect is

determined by the b_x vector from a life table. [As note above, vital rates and other demographic

parameters can vary by sex, but to reduce complexity of the notation we omit the subscripts for

167 sex, with the understanding that all analyses involve individuals of a single sex, nominally

168 female.] The within-age effect depends on the relationship between an individual's expected

169 reproductive success and the b_x value that represents mean fecundity for its age. Thus, the

170 relative parental weight for individual i of age x can be written as $w_i = b_x + \delta_i$, where δ_i is the

171 deviation from the age-specific mean. A positive δ means that the individual is expected to be

172 above average for its age; a negative δ indicates the reverse. If all $\delta_i=0$, the within-age effect

173 disappears, expected age-specific variance in reproductive success equals the mean

174 ($E[\sigma^2_{k(x)}] \approx b_x$), and individuals of the same age effectively behaves like a mini Wright-Fisher

175 population.

   Empirical data for $\sigma^2_{k(x)}$ are rarely published, but there are many reasons why individuals

177 of the same age and sex might not have identical expectations of reproductive success. Some

178 might have established breeding territories or acquired harems and others not; some might have

179 phenotypes that favor them in sexual selection. In many species, fecundity depends more on size

180 than age, in which case variation in w_i for same-age individuals will be correlated with variance

181 in size-at-age, leading to ($E[\sigma^2_{k(x)}] > b_x$. (see Waples et al. 2018a for an example involving

182 southern bluefin tuna). The magnitude of the within-age effect is quantified by the age-specific

183 index $\phi_x = E(\sigma^2_{k(x)})/b_x$, where $E(\sigma^2_{k(x)})$ and b_x have been scaled to their expected values in a

184 stable population (Waples et al. 2011; Waples 2016).

185   Finally, in iteroparous species individual weights can vary across years, for two major

186 reasons: (1) if b_x varies with age; (2) if an individual's condition factor changes, which in many

187 species depends on whether it has reproduced in a recent year (see Section 3.2.2).

189 **2.3 | Effective population size**

190 **2.3.1 | Demographic factors that affect N_e**

191   Close-kin methods provide information about abundance in the recent past, so

192 contemporary (rather than long-term) effective size is our focus. Mean and variance in offspring

193 number are the major factors that determine effective size. For seasonal reproduction in an age-

194 structured species, the most relevant effective-size metric is the inbreeding effective number of

195 breeders per year (N_b) because it relates directly to the annual number of adults (N_A), which also

196 is generally the focus of CKMR analyses. N_b can be calculated using the standard discrete-

197 generation formula for inbreeding effective size (Crow and Denniston 1988; Caballero 1994):

198
$$N_b = \frac{\mu_k N_A - 1}{\mu_k - 1 + \sigma_k^2 / \mu_k} \quad .$$
(1)

With separate sexes, N_b can be calculated separately for males and females and an overall

effective size obtained using Wright's (1938) sex-ratio adjustment. For simplicity we focus on

census size and effective size for females (N_f, $N_{b(f)}$), but exactly analogous equations apply to

males.

Equation 1 is exact when based on realized population parameters; it can be reformulated

as an expectation using the parental weights described above. Let $CV_W^2 = \sigma_W^2 / \overline{W}^2$ be the squared

coefficient of variation (CV) of the individual weights (which is the same whether calculated

using raw or standardized weights). Then, the expectation for the effective number of female

breeders per year is just (Waples 2020):

208
$$\mathrm{E}(N_{b(f)}) \approx \frac{N_f}{1 + CV_W^2} , \text{ so } \frac{N_b}{N_f} \approx \frac{1}{1 + CV_W^2} \quad .$$
(2)

Although N_b relates directly to annual abundance estimated by CKMR, virtually all

evolutionary theory depends on effective size per generation (N_e) rather than N_b per year.

Calculating N_e in iteroparous species is complicated by the necessity of integrating information

across multiple episodes of reproduction. All else being equal, N_e per generation is positively

correlated with generation length and negatively correlated with lifetime variance in number of

offspring, $\sigma_{k\bullet}^2$ (Hill 1972). The program *AgeNe* (Waples et al. 2011) calculates $\sigma_{k\bullet}^2$ using the

species' age-specific vital rates (s_x, b_x, ϕ_x), under the assumptions (adopted from Hill 1972 and

Felsenstein 1971) that vital rates and population size are constant and survival and reproduction

are independent over time. Demographic N_e also can be converted from an estimate of N_b based

on the N_b/N_e ratio, which largely depends on 2-3 life history traits (Waples et al. 2013, 2014).

2.3.2 | Genetic methods for estimating N_e

The sibship method developed by Wang (2009) considers three types of relationship in

the one-generation pedigree: full sibling pairs (FSPs), half sibling pairs (HSPs), and unrelated

(U), which nevertheless can be related through previous generations. In a random sample of *n*

progeny, $Y = n(n-1)/2 \approx n^2/2$ pairwise comparisons are possible, and expected proportions of

siblings are simple functions of N_e (from Wang 2009, ignoring the term for selfing):

226
$$E[Q_{HS} + 2Q_{FS}] = \frac{4}{N_e},\tag{3}$$

227 where Q_{HS} is the fraction of pairs that are half siblings (maternal and paternal HSPs combined)

228 and Q_{FS} is the fraction that are full siblings. Substituting for $Q = R/Y$, where R is the number of

229 sibling matches, and rearrangement leads to the estimator of effective size:

230
$$\widehat{N}_e = \frac{4Y}{R_{HSP} + 2R_{FSP}},\tag{4}$$

which has the number of pairwise comparisons in the numerator and the total number of sibling

doses (one for each HSP and two for each FSP) in the denominator. When applied to offspring

from a single cohort, Equation 4 provides an estimate of N_b. Wang et al. (2010) described an

extension of this sibship method to estimate N_e when generations overlap, but their model

assumes that $\phi_x = 1$ for each age and sex so is of somewhat limited general applicability.

Linkage disequilibrium (LD) quantifies associations of alleles at different gene loci.

Random LD is generated every generation by mating among a finite number of parents, and that

provides the basis for estimating effective size. Loci on the same chromosome ("linked")

provide information about historical N_e, so unlinked markers are most suitable for estimating

contemporary N_e. For unlinked loci, N_e can be estimated from r^2 (a measure of LD at pairs of

loci) and sample size (n = number of individuals) (Hill 1981):

242
$$\widehat{N}_e \approx \frac{1}{3(r^2 - 1/n)}.\tag{5}$$

The program LDNe (Waples and Do 2008) adjusts for bias in Equation 5 caused by ignoring

second-order terms in $1/n$ and $1/N_e$.

When applied to offspring from a single cohort, the LD method primarily reflects the

signal from parental N_b, with some influence from N_e in previous generations (Waples 2005;

Waples et al. 2014). When applied to mixed-age samples, LDNe generally underestimates

generational N_e by 10-40% (Waples et al. 2014). Blower et al. (2019) developed simulation

software for age-structured species that can help researchers interested in using the LD method to

estimate effective size to determine sample sizes of individuals and loci that should produce

desired levels of precision.

3 | MARK-RECAPTURE METHODS

Traditional mark-recapture (MR) involves three elements:

• Capture and mark a number (n_1) of individuals;

• Release them to mix with the general population;

• Collect a second sample of n_2 individuals; those marked are the "recaptures" (R).

The fraction of original marks that are recaptured provides information about population

abundance, but this metric also can be affected by a variety of other factors, including behavioral

effects on sampling probabilities, tag loss, mortality, immigration/emigration, and incomplete

mixing between samples. Accordingly, a complex statistical framework for conducting MR

analyses has been developed to account for these factors (Lebreton et al. 1992; Pollock 2000;

Lindberg 2012). Commonly, MR analyses are conducted using Integrated Population Models

(IPMs) that combine information from multiple datasets within a single pseudo-likelihood

framework to estimate population abundance, trends, and vital rates. IPMs have been widely

used in fisheries stock assessments for several decades (Fournier and Archibald 1982; Punt et al.

2020) and more recently have been applied to wildlife management (Arnold et al. 2018).

In CKMR, close-kin matches are analogous to MR recoveries. Full siblings and parent-

offspring pairs on average share half their genes, and half siblings share one quarter. As with

MR, close-kin matches provide information about abundance but also are affected by a variety of

other factors, which include (but are not limited to) age-specific survival and fecundity, spatial

structure, sampling selectivity, variance in offspring number, and correlations of individual

reproductive success over time. Most CKMR applications are complicated enough that they are

best conducted by building a log- pseudo-likelihood framework, analogous to the IPMs used

with MR, that incorporates outcomes of each pairwise comparison (Bravington et al. 2016a,b;

Hillary et al. 2018; Conn et al. 2020).

Estimation frameworks for MR and CKMR can be quite complicated and are largely

opaque to all but the most dedicated aficionados. In this synthesis, for heuristic purposes we take

a different approach. First, by focusing on the simplest versions of the models, we illustrate

fundamental similarities between MR and CKMR methods to estimate abundance and sibling-

based methods to estimate effective size, and we show how to account for some common factors

that can affect the estimates. Second, we suggest a new way to formulate the probability of a

sibling match that can be useful in illustrating the influence of various covariates.   Finally, we

use simulated data to illustrate how factors like age-specific changes in vital rates, adult

mortality, intermittent breeding, within-age variance in reproductive success, and experimental

design can affect performance of abundance and effective size estimators.

3.1 | Traditional mark-recapture estimation

In the simplest "cartoon" version of MR (if the probability of a recovery is only

influenced by population size), the estimate of abundance is simply $\hat{N} = n_1/P$, where $P = R/n_2$ is

the fraction of the second sample that is marked. This leads to the abundance estimator

$$\hat{N} = Y/R, \tag{6}$$

where $Y = n_1 n_2$ is the number of potential comparisons of individuals in the two samples, and the

denominator is the number of matches.

3.2. | Close-kin mark-recapture estimation

In the cartoon version of CKMR, all parents are equivalent in terms of producing

offspring and all Y pairwise comparisons to search for close kin are independent. Under these

conditions, the probability that a random pair of individuals will produce a close-kin match is

$P_{match} = 1/N_f$, and the expected number of matches is just the product of the number of

comparisons and the probability of success: $(E(R) = P_{match}*Y = Y/N_f$. This leads to the naive

estimator of adult abundance as $\widehat{N_A} = Y/R$, which is identical in form to Equation 6 for MR and

Equation 4 for sibship estimation of N_e.

More generally, estimation of abundance using the incidence of close-kin matches is

based on two basic principles:

A. It is possible to specify the probability that a random pair of individuals will produce

308   a close-kin match, and this probability is inversely related to adult abundance, such

309   that $P_{match} = f_1(1/N,Z) = f_2(Z)/N$, where f_1 and f_2 are functions and Z represents other

310   covariates than also can affect close-kin probabilities.

311   B. Pairwise comparisons to search for a match are independent, so the expected number

312   of matches (R) is the product of the number of comparisons (Y) and the probability

313   that each produces a match: $E(R) = Y*P_{match} = Y*f_2(Z)/N$.

Principle B is not generally true in the strict sense, at least for POPs (if A is the mother of X, B

cannot be); however, except in very small populations, sampling generally will be sparse enough

that this assumption does not lead to appreciable errors (Bravington et al. 2016b; see also Section

S2.2 in Supporting Information). Principle A is reasonable if one has a flexible enough concept

of 'function.' If A and B hold to an acceptable degree, then rearrangement produces

$\hat{N} = Y^* f_2(Z)/R.$ (7)

Because Y and R are empirically determined, estimation of N is possible provided one can define

the function f_2 so that all relevant covariates are accounted for. In certain cases, Equation 7 can

be expressed as $\hat{N} = Y^* C/R$;, i.e., the function $f_2(Z)$ takes the simple form of a scalar conversion

factor C, which can be calculated as a function of individual parental weights W (see section 2.2)

and other life-history parameters. Supporting Information S2 illustrates how to calculate C in a

number of cases.

326   The core problem for CKMR, therefore, is to account for all other factors besides N that

can affect the probability of close-kin matches. Among the most important covariates to consider

in calculating probabilities of a close-kin match are mortality, fecundity, sampling selectivity,

temporal correlations, population structure, age, and sex. Unless the sex ratio is even and vital

rates are the same in both sexes, the distribution of genetic "marks" passed on to their offspring

by male and female parents will be different. In general, therefore, it is best to make sex-specific

CKMR estimates, for which mtDNA can be useful (see Section 5.3).

3.2.1 | Parent-offspring pairs

335   Notwithstanding the difficulty in developing a general closed formula for CKMR, one

key metric is central to all CKMR analyses because it describes the probability that any given

pair of individuals will produce a close-kin match: the relative reproductive output (RRO) of

different potential parents. For POPs, consider the probability $[P(i \rightarrow j)]$ that female i is the

mother of offspring j. A general expression for this probability is (Bravington et al. 2016b):

$P(i \rightarrow j) = k_{i,tj}/TRO_{tj} = RRO_i,$ (8)

where $k_{i,tj} = RO_{i,tj}$ is the number of offspring produced by female i in the year of j's birth and

$TRO_{tj} = \Sigma k_{i,tj}$ is the total number of offspring by all females in that year. To estimate abundance,

it is necessary to combine these probabilities across all sampled offspring and potential parents.

POPs provide information about N_f in the year the offspring were born, so if samples are taken

from multiple cohorts, a time-series of abundance estimates can be obtained via age-structured

population dynamics, as in many fish stock assessments (Quinn and Deriso 1999).

Because actual reproductive output is rarely known, CKMR generally deals with

expected relative reproductive output, $E(RRO)$, based on expectations for different classes of

parents (Bravington et al. 2016b). This concept provides a direct link to the parental weighting

scheme described previously: parental weights for a given individual i are calculated as

$W_i = E(RO_i)/E(TRO)$, so $E(RRO_i) = W_i$.

CKMR estimates based on POPs are greatly simplified by the fact that each offspring has

exactly one mother. This means that variation in offspring number does not, by itself, affect

expected number of matches (but it can if there is unmodeled correlation between reproductive

output and sampling probability; see Section 5.2.1).

In the simplest POP-based examples, $P_{POP} = 1/N_f$, which would lead to an estimator of

adult female census size, directly analogous to a standard Lincoln-Petersen MR estimate (Seber

1982), as

$$\widehat{N}_f = Y/R_{POP} \qquad ,\qquad\qquad\qquad\qquad (9)$$

which has the same general form as Equations 4 and 6. Because $E(R_{POP})$ is not sensitive to

variation in offspring number, CKMR estimates based on POPs are not confounded by a signal

related to N_e or N_b. However, as illustrated in Section 5.2.2, precision of POPs-based estimates

is inversely related to variation in reproductive success.   If one or a few parents dominate

reproduction, the number of matches will vary widely depending on whether the prolific

parent(s) are sampled.

3.2.2 | Siblings

CKMR for siblings is considerably more complicated because (1) parents are not

observed directly, and (2) the probability of a sibling match depends heavily on adult mortality

and the extent to which reproductive output varies both within and across years. As

demonstrated below, only cross-cohort sibling comparisons are suitable for estimating

abundance. Because full siblings from different cohorts will be rare in large, randomly mating

populations, CKMR analyses generally focus on finding HSPs, and the probability that two

offspring share a mother can be taken as a close approximation to the probability that they are a

maternal HSP (MHSP). Our examples focus on MHSPs, but exactly equivalent equations apply

to paternal HSPs.

Realistic CKMR applications based on siblings must consider two ordered time periods,

0 and t. For two samples of n_1 and n_2 offspring, $Y = n_1 n_2$ across-sample comparisons are

possible. A general formula for the probability that two randomly selected offspring, one from

time 0 and one from time t, have the same female parent is (from Equation 3.9 in Bravington et

al. 2016b):

382
$$P_{MHSP} = \sum_{i=1}^{N_{f0}} \left[\frac{k_{i,0}}{TRO_0} \times \frac{k_{i,t}}{TRO_t} \right] = \sum_{i=1}^{N_{f0}} \left[RRO_{i,0} \times RRO_{i,t} \right] \qquad . \tag{10}$$

It is necessary to sum across all possible females that could have been alive when potential

siblings were born. Females that mature after time 0 cannot produce a sibling match for

offspring born in time 0, so the summation is Equation 10 is taken over the N_{f0} females that were

sexually mature at time 0.

Equation 10 does not generally lead to a closed-form estimator of adult abundance, unless

the species' biology allows a number of simplifying assumptions (see Hillary et al. 2018 for an

example involving the white shark, *Carcharodon carcharias,* which also required estimating

adult mortality from the data). However, some heuristic insights into factors that influence

P_{MHSP} can be achieved if the equation is reformulated as follows (see Supporting Information

Section S1 for details):

393
$$P_{MHSP} = \overline{RRO_{t*}} + N_{f0} [cov(RRO_{i,0}, RRO_{i,t})], \tag{11}$$

where the "*" in RRO_{t*} indicates that this vector only includes data for the subset of time-t

females that were mature at time 0. It might appear from Equation 11 that the term N_{f0}

represents a signal of female population size at time 0, but that is misleading: the denominator of

the population covariance is N_{f0}, so that signal cancels out. Instead, the term $\overline{RRO_{t*}}$ provides

information about adult abundance at time t. Two general cases can be identified.

3.2.2.1 | Case 1. Offspring are from the same cohort ($t = 0$).

If only a single time period is involved, $RRO_{i,0} = RRO_{i,t} = RRO_i$, with $\overline{RRO_i} = 1/N_f$.

Noting that the covariance of a random variable with itself is its variance, Equation 11 becomes

403
$$P_{MHSP} = 1/N_f + N_f [var(RRO_i)]. \tag{12}$$

Equation 12 shows that, when offspring from the same cohort are compared, the probability of a

sibling match increases as a function of the realized variance in offspring number. This variance

is the primary signal for N_b, so siblings from the same cohort provide information related to

effective size and not directly about adult abundance. For this reason, CKMR estimates using

siblings generally attempt to exclude within-cohort comparisons (Bravington et al. 2016b;

Hillary et al. 2018; Thomson et al. 2020).

3.2.2.2 | Case 2. Offspring are from different cohorts ($t \neq 0$).

In comparing offspring from different cohorts, it is always important to account for adult

mortality: if an adult female dies between times 0 and t, it can't be a parent in time t, so $RRO_{i,t} =$

0. These null parents are included in the summation in Equation 10 and will reduce $\overline{RRO_{t*}}$ and

therefore P_{MHSP} unless this mortality is accounted for. Other common factors that affect sibling

probabilities are considered below.

Changes in fecundity with age. All adults that survived to reproduce again at time t will

be at least one year older. If b_x changes with age (fecundity increases with age in many species,

especially fish and other poikilotherms with indeterminate growth), this will affect $\overline{RRO_{t*}}$ (and

perhaps the covariance term as well), unless this effect is accounted for.

Persistent individual differences. If within-age individual deviations from mean

fecundity (the δ_i defined in Section 2.2) are non-zero and positively correlated over time, such

that some individuals are consistently above or below average at producing offspring, this will

(all else being equal) lead to positive $cov(RRO_{i,0}, RRO_{i,t})$ and a higher proportion of siblings, as

well as reduced N_e (Lee et al. 2011, 2020). Body size is one factor that can easily lead to

positive temporal correlations in reproductive success. Fecundity often depends on size more

directly than on age, and an individual that is large (or small) for its age at time 1 is also likely to

differ in a similar way in subsequent years. Whether an individual is 'sexy' or has secured a

territory are other factors that could produce persistent individual differences.

Skip breeding. Skip breeding leads to a complex pattern of $cov(RRO_{i,0}, RRO_{i,t})$ that

can vary both in magnitude and sign, depending on the species' life history and the time interval

for cross-cohort comparisons (Figure 1). For example, consider a species for which females

generally skip t years after reproducing before doing so again. Comparisons of offspring born $<t$

years apart will produce few if any sibling matches, while comparisons exactly t years apart will

produce sibling matches at a rate determined not by female population size as a whole, but by the

$\approx N_f/t$ females that actually reproduce each year. It should be possible to develop an adjustment

for skip breeding using the vector (θ_t) of the probability that an individual will reproduce in the

current year, given that it last reproduced *t* years before (Shaw and Levin 2013). For female

loggerhead turtles (*Caretta caretta*), θ_t = [0.025, 0.443, 0.634, 0.743, 1] for *t*=1,5 (Waples and

Antao 2014), indicating a substantial effect of prior reproduction lasting four years. With

adequate data, CKMR can estimate parameters like θ_t based on age-gap patterns in the data (see

Section 6.1).

4 | SPATIAL STRUCTURE and SAMPLING

Spatial structure, where reproduction and/or sampling is concentrated in certain

geographic locations, will *not* bias CKMR estimates if either of the following conditions are met

(Conn et al. 2020): (1) offspring (siblings analyses) or parents and offspring (POPs analyses) mix

thoroughly before sampling occurs; or (2) sampling is equiprobable, meaning that every

individual has an equal chance of being sampled. To illustrate the first criterion, if reproduction

is panmictic in area A, where adults are sampled, and offspring randomly migrate to areas B and

C, sampling in just one of those areas still produces a random sample of all offspring (see

Bravington et al. 2016a for an empirical example involving southern bluefin tuna). If the second

criterion (and other CKMR assumptions) are met, abundance estimates will reflect the total

number of adults. If spatially-restricted reproduction is persistent and there are independent or

semi-independent populations or stocks for which separate abundance estimates are desired, the

naive CKMR estimate of a single overall abundance might be unsatisfactory. If estimates are

available for key parameters describing the degree of reproductive connectivity and

distributional overlap, it might be possible to develop area-specific abundance estimates (Davies

et al. 2017). Fortunately, if population stratification is strong enough to substantially affect

CKMR estimates, this should be apparent from even a fairly coarse preliminary sampling design

(Conn et al. 2020). Furthermore, spatial patterns in distribution of close-kin can provide insights

into population subdivision and connectivity (Økland et al. 2009; Palsbøll et al. 2010; Wang

2014; Feutry et al. 2017).

In general, the most important proviso regarding sampling is that whether an individual is

sampled should have no effect on the probability that one of its close kin is also sampled. If this

criterion is satisfied, conditioned on covariates associated with the sampled individuals, then

expected values of *RRO* are multiplicative, which makes the CKMR analyses tractable. Joint

sampling of close kin in the same spatio-temporal stratum at higher frequencies than would occur

by chance will increase the probability of close-kin matches; this potential source of bias can be

reduced by appropriate filtering of the data (as Thomson et al. 2020 did with the school shark,

Galeorhinus galeus).

Sampling considerations for single-sample genetic estimators are similar to those for sib-

based CKMR, except that non-random collection of siblings can be difficult to avoid when

sampling within cohorts. Without independent information, it is not possible to reliably

compensate for a non-random sample by excluding some or all putative siblings, although some

approaches can reduce bias (Waples and Anderson 2017). Biases to genetic estimates of N_b from

undetected population structure should be similar to those for CKMR-based estimates of

abundance. If the focal population is part of a metapopulation but sampling is local, the LD

method primarily estimates local N_b unless migration is very high in genetic terms ($>5\text{-}10\%$;

Waples and England 2011; Gilbert and Whitlock 2015).

5 | WORKED EXAMPLES

5.1 | Simulations

We simulated demographic and genetic data (using the software CKMRPop, Anderson

2021) to illustrate effects of some common factors on CKMR and N_e estimators. Scenarios

modeled here were simple enough that the function f_2 from Equation 7 could be specified and

CKMR estimates could be made using that closed-form equation. Below we briefly describe the

model and summarize the main results and conclusions. More details about the simulations and

explanations for all the calculations can be found in Supporting Information.

The core scenario involved a single closed population with the following features

(notation follows that introduced in Section 2 and defined in Table 1):

• age at maturity (α) = 3

• maximum age (ω) = 10

• equal primary sex ratio and same vital rates in males and females

• annual probability of survival was constant at $s_x = 0.7$

• relative fecundity was either constant with age ($b_x = 1$ for all adults) or proportional to

age ($b_x = x$ for all adults)

• The N_A adults produced a constant number of yearling offspring per year (N_1 = cohort
size = 2000)
• number of offspring produced by individuals of the same age and sex was drawn from a
negative binomial distribution with either Poisson (~random) variance ($\phi = 1$) or
overdispersed variance ($\phi = 10$).
The life table for this core population is shown in Table 2. With constant cohort size and random
survival, total abundance each year averages 3239 individuals of each sex ($N_T = 6478$), of which
3400 are juveniles and $N_A = 3078$ are adults (hence $N_f = N_m = 1539$). Table 3 defines six
Scenarios (A-F) that represent variations on this life table for which we simulated data, in
addition to some other scenarios described below.
CKMRPop does not do kin-finding based on genetic data. Because we wanted to focus
on bias and precision related to experimental design and the species' life history, we used the
true pedigree recorded by CKMRPop to identify POPs and siblings, which is equivalent to
assuming that all close-kin matches were made without error.
512
5.2 | Results
Bias can arise in CKMR estimates when the model does not fully account for all factors
that affect the probability of a close-kin match (P_{match}). In the examples considered here, we
accounted for potential biases by calculating a conversion factor C that is the ratio of the true
P_{match} to the naïve value of $1/N_f$. This conversion factor thus plays the role of the function f_2 in
Equation 7. The value of C for any given application is determined by the focal species' vital
rates and the experimental design, but it is independent of population size (see Supporting
Information for details about calculation of C). C can then be used to accurately predict the
expected number of close-kin matches and to obtain an unbiased estimate of abundance.
522
5.2.1 | Bias
5.2.1.1 | Parent-offspring pairs
In our simulations, POP-based estimates of adult N were unbiased using the naive model
that assumes $P_{POP} = 1/N_f$, (Table 3 and Figure 2) except when:
1) Fecundity and probability of being sampled both increased with age; or
2) The sample of potential adults also included immature individuals.

In situation 1), joint effects of age-specific changes in selectivity and fecundity increased the

number of POP matches beyond the number expected under the naive model, leading to $C>1$

(calculations shown in Table S1). In situation 2), dilution of the 'adult' sample by inclusion of

juveniles that could not produce offspring reduced the fraction of comparisons that produced a

POP match, leading to $C<1$. With equiprobable sampling of the entire population, this produced

an unbiased estimate of overall abundance (N_T) but would represent a bias if the goal were to

estimate N_A. After adjusting P_{POP} by the factor C, $\widehat{N}_{A(adj)}$ for these two scenarios was unbiased

as well (Table 3).

5.2.1.2 | Siblings

Because sibling-based estimates of abundance depend on cross-cohort comparisons of

offspring to search for a shared parent, in general one of the following conditions must be met

for CKMR to be feasible: a) mixed-age samples can be aged precisely and sorted into cohorts, or

b) a series of samples from individual cohorts is collected over time. Our simulations followed

the second approach; in each replicate for sibling analyses we collected samples of yearlings in 5

consecutive years, which allowed us to compare cohorts born 1-4 years apart (the 'age gap').

Even when all adults are equally likely to produce offspring, it cannot be the case that the

probability of a sibling match is $1/N_f$, because some parents will die and cannot be the parent of

an offspring born in a later year. If all adults are otherwise equally likely to produce offspring,

the conversion factor C can be calculated from cumulative survival rates over time. If fecundity

changes with age, the bias adjustment is more complicated but can be done using Equation 10 to

calculate P_{MHSP} based on the population's vital rates (as illustrated in Table S2). By making

these adjustments, unbiased abundance estimates can be obtained for a wide range of scenarios,

for very small to very large populations (Table 3 and Figure S1).

Figure 3 illustrates the effects of intermittent breeding on the number of sibling matches

as a function of the age gap. Allowing a random subset of adults to reproduce every year does

not affect the total number of cross-cohort comparisons (columns 'Total 1-4'), but it does reduce

N_b per year and hence increases the number of within-cohort siblings. In scenarios where the

probability that the same individual reproduces in consecutive years is low or 0 (as in the

'Alternate' scenario in Figure 3), the expected number of sibling matches can vary dramatically

depending on the age gap—a factor that must be considered to obtain an unbiased estimate of

abundance. Fortunately, the empirical distribution of age gaps between sibling matches (or

between POP matches in a long-term POP study; see Bravington et al. 2016a) can be used to

estimate the nature and magnitude of skip breeding.

5.2.2—Precision

CKMR methods are categorical (each pairwise comparison is either a POP/HSP or not),

so only integer numbers of close-kin matches are possible (Figure 4). [In theory, probabilistic

kin assignments could be used, as they commonly are for calling genotypes in genomics-scale

datasets (Korneliussen et al. 2014), but this would cause hideously daunting consequences for

CKMR likelihood calculation, for little gain in demographic information—M. Bravington, pers

com. January 2021.] Precision of CKMR estimates is primarily determined by two key

parameters: mean and variance in number of close-kin matches (R).

A crude approximation to a minimum CV for $\hat{N}_A$ from MR or CKMR is $1/\sqrt{R}$

(Bravington et al. 2016b), in which case R=50 would translate into a CV of about 15%. This

rule-of-thumb assumes a Poisson variance in R, implying that var(R) $\approx$ 50. In our simulations we

monitored var(R) and compared it to the Poisson expectation. Also, because of the inverse

relationship between $\hat{N}_A$ and R, $\hat{N}_A$ is skewed high and is infinitely large when no recoveries are

found. This same issue applies to Wang's sibship method (see Equation 4), and for this reason it

is common for evaluations of precision of effective size estimators to focus on the distribution of

$1/\hat{N}_e$ rather than $\hat{N}_e$, and to use the harmonic mean rather than arithmetic mean as a measure of

central tendency (e.g., Wang 2009; Waples and Do 2010). Accordingly, here we calculate

empirical CVs of $1/\hat{N}_A$ and $1/\hat{N}_b$ based on variances across replicates, as well as medians or

harmonic means.

5.2.2.1 | Parent-offspring pairs

We found that var(R) among POP replicates was generally close to or slightly smaller

than the Poisson variance (Figure S2A), which implies that in the simplest forms of POP-based

CKMR the above rule-of-thumb is fairly accurate. An exception occurred for Scenario B, which

had substantially overdispersed variance in reproductive success (ϕ=10); in this case, the large

589 within-age effect on σ_k^2 caused var(R) to be over twice as large as the mean (Figure S2A), which

590 also considerably widened the distribution of $\widehat{N}_A$ (Figure 2).

592 **5.2.2.1 | Siblings**

593   For the same scenarios from Table 3, the ratio var(R)/mean(R) was generally slightly

594 higher for sibling analyses than for POPs (Figure S2A). For scenarios B (ϕ=10) and C (fecundity

595 increasing with age), var(R)/mean(R) for siblings was > 1 (2.22 and 1.23, respectively), which

596 presumably reflected the fact that both of these scenarios included unequal parental weights.

597   For a given life-history scenario, the ratio var(R)/mean(R) (and hence precision) is

598 sensitive to adult abundance. As illustrated in Figure 5 (Panels A and B), although larger

599 samples of individuals are needed from larger populations to produce a fixed number of close-

600 kin matches, required sample sizes increase more slowly than adult abundance, so in very large

601 populations only a small fraction of the population has to be sampled each year. This occurs

602 because increasing the sample size by a factor X increases the number of possible comparisons to

603 search for siblings by the factor X^2. In theory, the effects of overdispersed variance and other

604 factors that cause non-independence of the data should attenuate as sampling becomes sparse

605 compared to overall abundance (Bravington et al. 2016b).

606   Our simulations of populations with large N_A produced results consistent with this theory.

607 For the smallest populations we modeled ($N_A = 770$), var(R)/mean(R) was somewhat elevated

608 (1.26) from the Poisson expectation of 1.0 even for ϕ=1, but the variance rapidly dropped to at or

609 below the Poisson variance as population size increased (Figure S2B). With moderately high

610 within-age variance in reproductive success (ϕ=10), departures from the Poisson expectation are

611 greater and the rate of attenuation lower, but even so var(R)/mean(R) dropped to ~1.0 for adult

612 abundances > 10^5 (Figure S2B). All of these scenarios were designed to produce an average of

613 50 total half-sibling matches, and those that achieved a var(R)/mean(R) ratio close to 1.0

614 produced CVs of $1/\widehat{N}_A$ that were close to the expected 0.15 (Figure 5C).

615   For two scenarios (A and B from Table 3), we compared CKMR precision for estimating

616 N_A with precision to estimate N_b using Wang's sibship method, as a function of the number of

617 consecutive years of samples that were used (2-5). In Scenario A, $\phi = 1$ and fecundity was

618 constant with age, so $N_A = N_b = 3078$; precision was higher for $\widehat{N}_b$ with only 2 years of data, but

619 for longer studies precision of $\hat{N}_A$ was higher (Figure 6). This latter result occurred because the

total number of pairwise comparisons increases faster for cross-cohort comparisons used in

CKMR than it does for within-cohort comparisons used to estimate N_b. Results for Scenario B

were quite different: with $\phi = 10$, true N_b (388) is greatly reduced compared to adult N (3078),

and this produces a stronger drift signal that is relatively easier to estimate precisely.

Conversely, the high variance in reproductive success substantially increased variability in the

number of cross-cohort sibling matches. As a consequence, $CV(1/\hat{N}_A)$ was 4 times as large as

$CV(1/\hat{N}_b)$ for 2 years of samples and still more than twice as large with 5 years of samples

(Figure 6). These results illustrate an important difference between CKMR and estimators of

effective size: overdispersed variance in reproductive success reduces precision of $\hat{N}_A$ but

increases precision of $\hat{N}_b$ and $\hat{N}_e$.

CKMR and Wang's sibship method share a limitation in that once enough genetic

markers are used to reliably identify all close kin, additional markers cannot further increase

precision, unless doing so allows more distant kin relationships to be integrated into the analysis.

The LD method does not suffer from this limitation; in fact, the number of pairs of loci that can

be used to compute mean r^2 increases with the square of the number of loci, so in theory

precision to estimate N_e and N_b could become arbitrarily large with genomics-scale datasets. In

practice, limited recombination within chromosomes and lack of independence of overlapping

pairs of the same loci limit the information content in large genetic datasets (Thompson 2013;

Waples et al. 2021). Despite these limitations, if more than roughly 500-2000 SNPs are used,

precision of the LD method for estimating N_b will generally exceed the maximum possible

precision of Wang's sibship method, which occurs when all sibling relationships are correctly

identified (Waples 2021).

5.3 Costs

Data shown in Figure 5 provide an opportunity to illustrate the relationship between

experimental design and genotyping costs (see also Table S5 for more details). The following

parameters apply to prices discussed below:

•   Costs are in $USD and apply to SNP discovery and genotyping, but not DNA extraction,

using DArTseq/DArTcap technology (Jaccoud et al. 2001) as implemented in Feutry et

al. (2020);

- Costs include discovering and genotyping 1,500+ high quality SNPs (low error rate, high
call rate, and MAF>0.1);

- Costs assume 94 samples per plate, with 1 positive and 1 negative control;

- Costs include a fixed component for SNP discovery (1 plate of DArTseq @ $3800) + a
fixed component for DArTcap SNP panel synthesis + $11.4/individual);

- Costs assume a species with at least moderate levels of genetic diversity; species with
low diversity might require more sequencing at extra cost for SNP discovery.

Estimates are based on several CKMR projects conducted by CSIRO Hobart, using methods

described in Feutry et al. (2020), which in our experience is sufficient for accurate inference of

POPs and siblings in species with N_A up to ~2×10^6. These costs represent only one possible

commercial option, but a proven one that is globally available. Other methodologies exist, some

laboratories might be able to save money by performing analyses in-house, and costs are likely to

keep dropping in the future. Nevertheless, estimates shown in Figure 5D can serve as a useful

reference point (and upper limit) for planning and experimental design.

Data shown in Figure 5D are total costs for an experimental design that involves 5

consecutive years of samples at a level designed to produce about 50 total HSPs and $CV(1/\widehat{N}_A)$

of about 0.15. Costs rise more slowly than population size for two reasons: a) the fraction of the

population that has to be sampled to produce a fixed level of precision declines as N_A increases

(Figure 5B); and b) in large studies, fixed costs of SNP discovery become relatively less

important. Figures above do not include sampling, costs for which can be substantial but will

vary widely across applications. On the other hand, resulting genotypes should be suitable for

performing all of the analyses discussed in this paper, including identification of POPs and

siblings and calculation of mean r^2 for the LD analyses.

Also not included are costs for mtDNA analysis to estimate the fractions of HSPs that are

of maternal and paternal origin, which can be modelled using haplotype frequencies (Bravington

et al. 2016b; Thomson et al. 2020). Costs are modest (~5K$ per 50 HSPs) because only the

close-kin matches have to be tested but might be relatively important for small studies. For

POPs, parental sex is generally determined morphologically or using sex-specific markers.

6 | DISCUSSION

CKMR estimates of abundance and genetic methods to estimate effective population size

share a dependence on core features of population demography: means and variances of age-
specific vital rates and covariances in realized reproductive success over time. Effective size can
be predicted based on a vector of parental weights, **W**, that is mathematically equivalent to the
expected-relative-reproductive-success (*ERRO*) concept that underpins CKMR. A CKMR study
will generate some or all of the demographic data needed to calculate effective size directly, and
genotypes for the sampled individuals provide an opportunity to apply indirect genetic methods
for estimating N_e or N_b. Nevertheless, demographic parameters and other covariates influence
CKMR and effective size estimators in different ways.
689

6.1 | CKMR
As illustrated in Equations 8 (for POPs) and 10 (for siblings), the probability of a close-
kin match is inversely proportional to total reproductive output (*TRO*) of the population. *TRO* is
closely related to spawning stock biomass, which typically is an important metric in fisheries
stock assessments. In many other applications, however, adult abundance is the parameter of
primary interest, in which case it is necessary to disentangle the relationship between *TRO* and
the N_A adults responsible for that reproductive output. For some simple scenarios, this can be
done using Equation 7 and a scalar conversion factor *C*, calculated as described in detail in
Supporting Information S2.2. This approach was adopted for analyzing the simulated data,
which illustrated predictable effects of some common factors on performance of CKMR and
effective size estimators. Key results that were demonstrated include the following:
•   Large population size does not by itself cause bias (Figure S1; Table 3), but large
populations require larger sample sizes (*n*) to detect the same number of close kin (*R*),
where $E(R) = Y*P_{match}$. However, the number of pairwise comparisons (*Y*) increases with
n^2 while the probability of a match declines only according to $1/N_A$. Therefore, the
fraction of the population that must be sampled to achieve a fixed level of precision is
smaller in large populations, which helps reduce costs for large projects (Figure 5).
•   Precision of CKMR estimates primarily depends on two factors: mean and variance in
number of close-kin matches. The relationship between $\hat{N}_A$ and *R* is inverse and non-
linear, so as *R* increases the curve becomes flatter (Figure 4). With respect to precision,
therefore, the experimental-design goal should be to make E(*R*) large enough that random
sampling error and limited errors in pedigree reconstruction have relatively little

influence on $\hat{N}_A$. Our simulations showed that the assumption of a Poisson variance to R

is generally reasonable, at least for large populations. But the empirical var(R) can be

substantially higher than the Poisson expectation in populations of small to moderate

size, and this effect is exacerbated when adults have very unequal probabilities of

producing offspring. To compensate for this increased variance and maintain a target

level of precision, it might be necessary to boost the sampling effort to increase E(R).

• Sibling-based CKMR analyses should exclude same-cohort comparisons, whereas

Wang's method is restricted to single cohorts. Like Jack Sprat and his wife[1], the two

methods neatly carve up the data into non-overlapping layers. Relative precision to

estimate N_A and N_b from the same experimental design depends on three factors: the

number of years of samples, annual mortality, and variance in reproductive success

(Figure 6). In longer studies, the number of cross-cohort comparisons increases faster

than those within cohorts, which increases relative precision of $\widehat{N}_A$; however, this

advantage is partially offset by adult mortality, which reduces the effective number of

cross-cohort comparisons. Overdispersed variance in reproductive success ($\phi>1$)

increases var(R) and CV($1/\widehat{N}_A$) but has the opposite effect on CV($1/\widehat{N}_b$) (Figure 6),

because large ϕ reduces N_b and increases the drift signal.

• Although $\phi>1$ increases var(R) and reduces precision for CKMR (except perhaps in very

large populations), it does not bias abundance estimates based on either POPs or cross-

cohort siblings (Figures 2 and S1).

• *P_POP* is a simple inverse function of $1/N_A$, leading to unbiased estimates of adult

abundance, provided that there are no unmodelled covariates that affect both an adult's

sampling probability AND its *ERRO* among the sampled offspring. If *ERRO* and

sampling probability of adults both increase over time (as in Scenario D, Table 3), this

potential source of bias can be accounted for as illustrated in Table S1.

• POPs-based abundance estimates depend heavily on which individuals are included in the

sample of potential parents. If non-breeding adults are not sampled, the estimate will

apply to the number of breeders in a given year (Figure 1A); conversely, if immature

---

[1] Jack Sprat could eat no fat. His wife could eat no lean. But, together both, they licked the platter clean (Opie and Opie 1997, p. 238)

740   individuals are sampled as potential adults, the naive estimate will be skewed away from

741   N_A toward overall N (Figure 2).

- Cross-cohort sibling comparisons provide information about adult abundance in the year the younger sibling was born (Equation 11). POP comparisons provide information about adult abundance in the year(s) offspring were born. Sampling from multiple cohorts thus can provide information about population trend.

- It is always necessary to consider adult mortality for sibling-based CKMR. This source of random noise increases var(R) and reduces precision, even when average mortality rates can be accurately estimated (Figure S2).

- Random and uncorrelated individual variation in *ERRO* does not bias sibling-based estimates of abundance (Figure 3). However, factors that cause positive or negative correlations in *ERRO* over time do affect P_{MHSP} (Equation 11) and must be accounted for to avoid bias. Effects of some of these factors, such as intermittent breeding and persistent individual differences, are relatively straightforward (Figure 1). Less intuitive is the fact that any systematic change in *ERRO* with age also establishes predictable correlations in individual reproductive success over time, and these effects can be accounted for as illustrated in Table S2.

758   Despite its heuristic value, the simplified approach outlined in Equation 7 and illustrated

759 by the simulated examples has some important limitations. For simplicity and tractability, we

760 assumed that:

- All close kin were correctly identified;
- The single, closed population was constant in size;
- Sex ratio was equal and vital rates were constant over time and known without error;
- All sampled individuals could be aged accurately;
- Sampling was random unless adults were sampled in proportion to relative fecundity, which also was known precisely.

In the sibling simulations, these assumptions allowed us to combine all sibling matches across

five years of samples to estimate a single, constant parameter (N_A for CKMR, N_b for Wang's

sibship method).

None of the CKMR applications published to date meet all these assumptions. More

generally, CKMR experimental designs involve sampling and testing for POPs and/or siblings

from multiple time periods, across which abundance and vital rates can change, perhaps sharply.

Incorrectly assuming parameter stationarity can easily lead to bias (Skaug 2001). When values

of key covariates are uncertain, the estimation process can integrate over plausible ranges, albeit

at a cost with respect to precision. For all these reasons, the most robust implementations of

CKMR incorporate information from parent-offspring pairs and siblings into a single, log-

pseudo-likelihood framework, within which all demographic parameters can be jointly estimated

in a way that allows coherent statements about uncertainty.

Another major benefit of using an integrated modeling framework is that many, if not all,

key covariates that affect P_{match} can also be estimated from the data. For example, with POPs

plus ages one can estimate age-specific fecundity. Adult survival can be inferred from the

distribution of age gaps between siblings (Hillary et al. 2018; see also Table S3). Intermittent

breeding can be evaluated using both POPs and HSPs by examining lag times between repeated

offspring assignments to the same parent (Bravington et al. 2016a; Thomson et al. 2020). To the

extent that independent information about covariates is available from other sources, this can be

included within the overall estimation process.

6.2 | Effective size

Many of the factors discussed above that complicate CKMR abundance estimates are part

of the genetic drift signal and hence create no problems for genetic methods that estimate annual

N_b. This applies to age-specific changes in vital rates, skewed sex ratios, skewed offspring

distributions, and skip breeding. As with CKMR, however, estimates of effective size can be

very sensitive to non-random sampling that leads to positive correlations of sampling

probabilities among relatives (Goldberg and Waits 2010; Whiteley et al. 2012). Although

detailed life-history and age-structure information for the population as a whole is necessary for

demographic estimates of N_e or N_b, ageing of individuals is primarily important for sorting

offspring into single cohorts. Single-sample genetic methods estimate the effective number of

breeders that produced the sampled offspring (Waples 2005; Wang 2009). Researchers should

keep in mind that, depending on the experimental design and the biology of the target species,

this might not represent the entire adult population (e.g., under skip breeding in Figure 1).

6.3 | Practical applications

Meaningful CKMR estimates require (1) genetic data robust enough to reliably identify close-kin pairs; and (2) sufficient sampling effort to produce the desired level of precision. We have assumed that all close-kin matches are correctly identified, which implies that criterion (1) has been met (see Bravington et al. 2016b for a summary of basic principles of kin identification). In practice, some level of uncertainty is associated with both genotyping and kin identification, so pedigree-reconstruction models should account for potential errors and evaluate their consequences.

To date, CKMR has only been applied to a handful of species, but these cover a wide range of population sizes and life histories, from relatively small populations of salmonids (Rawding et al. 2014; Ruzzante et al. 2019) and fecundity-limited sharks (Hillary et al. 2018) to a large, mobile, highly-fecund tuna with 2 million estimated adults (Bravington et al. 2016a). Most of these applications fall into one of two categories: at-risk species with high conservation value, or high-value fisheries. All published applications to date are for aquatic species, but CKMR methods can also potentially be applied to terrestrial species, as is commonly done with genetic methods to estimate effective size.

Even when various factors constrain the types of analyses that are feasible, it can be possible to extract valuable information using CKMR. Despite the inability to reliably sample adults, juvenile samples of the cryptic and critically endangered speartooth shark, *Glyphis glyphis*, have provided important insights into connectivity (Feutry et al. 2017). Inability to precisely age the school shark, *Galeorhinus galeus*, added considerable uncertainty to CKMR estimates of abundance, which nevertheless were more precise than those from fishery-dependent methods (Thomson et al. 2020; Simpfendorfer et al. 2021). Among the most important factors that affect the feasibility of CKMR methods are the following (see also Tables 4 and 5):

6.3.1 | Biology

- Long lifespan and/or precise ageing facilitates sampling of multiple cohorts
- Low adult mortality increases precision of HSPs
- For HSPs, it is necessary to be able to either a) assume that fecundity is constant with age, or b) use independent sources of information, or c) estimate b_x from CKMR data, generally using POPs + ages.

832   • In small populations, the lack of independence of close-kin comparisons can become

833   an issue.

6.3.2 | Ageing

835   • Minimum requirements: must be able to sort into potential parents and offspring

836   (POPs) or into consecutive cohorts (siblings; so age gap is known)

837   • Need actual ages to estimate age-specific fecundity with POPs

838   • Uncertainty in ageing can be modeled but can substantially reduce precision

6.3.3 | Sampling

840   • Must be able to sample individuals independent of their kinship, or model lack of

841   independence

842   • For POPs, if adults cannot be sampled independent of their *ERRO*, must be able to

843   estimate selectivity

6.4 | Future considerations

We expect to see some or all of the following developments in the future:

• As whole-genome-sequencing of non-model species becomes more common, it will

become feasible to extend the array of close-kin categories available to CKMR beyond

the one-generation pedigree to include more distant relationships, including cousins and

niblings (nieces and nephews).

• Conversely, as long-term studies continue, complications posed by other relationships

that have the same kinship probabilities as half siblings (e.g., grandparent – grandchild)

will increase and require effective treatments. Current applications often can ignore these

complications because of the species' biology (esp delayed age at maturity) and

experimental design.

• Single-sample genetic estimators can be used to provide a continuous time-series of

annual estimates of N_b, which are generally done independently each year (Ruzzante et

al. 2016; Whitely et al. 2015; Waples et al. 2018a,b). One could develop an analytical

framework analogous to that used by CKMR that jointly considers multiple years of data

to estimate temporal trajectories of N_b and other population parameters, but we are not

aware of any attempts to do that. Furthermore, because CKMR and the two single-

sample estimators of effective size use the same data, it should be possible to incorporate

all of the estimation of annual abundance, N_b, and related parameters into a single

pseudo-likelihood framework that would help shed light on the important ratio of

effective size to census size. An initial attempt to combine CKMR and effective size

information, based only on POP data for microsatellites used in the CKMR study of

southern bluefin tuna, showed that the range of N_e/N ratios that were most consistent with

the combined data was about 0.1 to 0.5 (Figure 7). This ruled out any substantial effect

of sweepstakes reproductive success for southern bluefin tuna, even though the species'

longevity and high fecundity are similar to several marine species for which "tiny"

estimates of N_e/N have been reported (Hedgecock and Pudovkin 2011).

• Tightly linked markers on the same chromosome provide insights into N_e in the more

distant past. If detailed information about genomic structure is available, it can be

leveraged to provide a temporal series of $\hat{N}_e$ estimates back in time. Applications of this

approach include humans (Tenesa et al. 2007), red drum (*Sciaenops ocellatus*;

Hollenbeck et al. 2016), and Atlantic salmon (*Salmo salar*; Lehnert et al. 2019), and a

recently-developed modification (Santiago et al. 2020) appears to considerably improve

performance. Apart from using more distant kin to extend pedigree inference back one or

a few more generations, there does not appear to be any comparable potential for CKMR

to provide insights into historical abundance.

Finally, an aspirational goal would be to integrate estimation of abundance with CKMR

and estimation of generational effective population size, N_e. This should uncover some

interesting parallels: unlike annual N_b but similar to sib-based CKMR, N_e per generation is

sensitive to correlations in individual reproductive success over time. Hill's (1972) model for N_e

when generations overlap is robust to random demographic variation if N is not too small (Engen

et al. 2005a; Waples et al. 2011). However, allowing for environmental stochasticity requires a

different formulation (Engen et al. 2005), and even this becomes problematical under strong

density dependence, when vital rates vary over time and generation length is not well defined

(Myhre et al. 2016). Successfully integrating generational N_e and CKMR into a single

estimation framework thus remains a challenging prospect.

Acknowledgments

Mark Bravington contributed a great deal to our understanding of the material presented

here. We thank Eric Anderson, Peter Grewe, Daniel Ruzzante, and Hans Skaug for useful

discussions. Marco Andrello, Mark Bravington, Ryan Kovach, Jann Martinsohn, Todd

Seamons, Antonella Zanzi, and two anonymous reviewers provided insightful comments on

earlier versions of the manuscript. Eric Anderson kindly modified his CKMRpop simulation

package to run under Windows and performed some simulations for population sizes (adult $N >$

10^5) that were too large to run locally. RW's work on this project was supported by an

International Science Fellowship from the National Marine Fisheries Service and by a Froelich

Fellowship from CSIRO. The authors declare no conflict of interest. This manuscript did not

generate new empirical data.

References

Akita, T., 2020a. Nearly unbiased estimator of contemporary Ne/N based on kinship relationships. *Ecology and evolution*, *10:*10343-10352.

Akita, T., 2020b. Nearly unbiased estimator of contemporary effective mother size using within-cohort maternal sibling pairs incorporating parental and nonparental reproductive variations. *Heredity* 124:299-312.

Anderson EC. 2021. CKMRpop: Forward-in-time Simulation and Tabulation of Pairwise Relationships in Age-structured Populations v. 0.1.0 (github.com/eriqande/CKMRpop).

Arnold, T.W., Clark, R.G., Koons, D.N. and Schaub, M., 2018. Integrated population models facilitate ecological understanding and improved management decisions. *The Journal of Wildlife Management*, *82*(2), pp.266-274.

Blower, D.C., Riginos, C. and Ovenden, J.R., 2019. NeOGen: A tool to predict genetic effective population size (Ne) for species with generational overlap and to assist empirical Ne study design. *Molecular ecology resources*, *19*:260-271.

Bradford, R, Thomson, R, Bravington, M, Foote, D, Gunasekera, RM, Bruce, B, Harasti, D, Otway, N, Feutry, P. 2018. A close-kin mark-recapture estimate of the population size and trend of east coast grey nurse shark. Report to the National Environmental Science Program, Marine Biodiversity Hub. CSIRO Oceans & Atmosphere, Hobart, Tasmania. available at: https://www.nespmarine.edu.au/document/close-kin-mark-recapture-estimate-population-size-and-trend-east-coast-grey-nurse-shark.

Bravington, M.V., P. G. Grewe, C. R. Davies. 2014. Fishery-independent estimate of spawning biomass of Southern Bluefin Tuna through identification of close-kin using genetic markers. FRDC Report 2007/034, Australia, 2014.

Bravington, M.V., Grewe, P.M. and Davies, C.R., 2016a. Absolute abundance of southern bluefin tuna estimated by close-kin mark-recapture. *Nature communications*, *7*:13162.

Bravington, M.V., Skaug, H.J. and Anderson, E.C., 2016b. Close-kin mark-recapture. *Statistical Science*, *31*:259-274.

Bravington, M., Feutry, P., Pillans, R.D., Hillary, R., Johnson, G., Saunders, T., Gunesekera, R., Bax, N.J. and Kyne, P.M. (2019). Close-Kin Mark- Recapture population size estimate of *Glyphis garricki* in the Northern Territory. Report to the National Environmental Science Program, Marine Biodiversity Hub. CSIRO Oceans and Atmosphere, Hobart.

Buckworth, R., Ovenden, J., Broderick, D., Macbeth, G., McPherson, G. and Phelan, M., 2012. GENETAG: Genetic mark-recapture for real-time harvest rate monitoring. Pilot studies in northern Australia Spanish mackerel fisheries. Darwin Australia: Northern Territory Government, Department of Resources; 2012. Available at: http://hdl.handle.net/102.100.100/100224?index=1

Caballero, A., 1994 Developments in the prediction of effective population size. Heredity 73: 657–679.

Caswell, H. 2001. Matrix Population Models, Ed. 2. Sinauer, Sunderland, MA.

Charlesworth, B., 2009. Effective population size and patterns of molecular evolution and variation. *Nature Reviews Genetics*, *10*(3), p.195.

Conn, P.B., Bravington, M.V., Baylis, S. and Ver Hoef, J.M., 2020. Robustness of close-kin mark–recapture estimators to dispersal limitation and spatially varying sampling probabilities. *Ecology and Evolution* 10:5558-5569.

Crow, J.F., and C. Denniston, 1988 Inbreeding and variance effective population numbers. *Evolution* 42: 482-495.

Davies C, Bravington M, and Thomson R. 2017. Advice on Close-Kin Mark-Recapture for estimating abundance of eastern Atlantic bluefin tuna: a scoping study. CSIRO Marine Laboratories, Hobart, Tasmania, Australia.

Doak, D.F., Gross, K. and Morris, W.F., 2005. Understanding and predicting the effects of sparse data on demographic analyses. *Ecology*, *86*(5):1154-1163.

Engen, S., R. Lande, and B.-E. Sæther. 2005. Effective size of a fluctuating age-structured population. Genetics 170:941–954.

England, P.E., G. Luikart, and R.S. Waples. 2010. Early detection of population fragmentation using linkage disequilibrium estimation of effective population size. Conservation Genetics 11:2425–2430.

Felsenstein, J., 1971 Inbreeding and variance effective numbers in populations with overlapping generations. Genetics 68:581–597.

Fernandes, D.M., Sirak, K.A., Ringbauer, H. *et al.* 2020. A genetic history of the pre-contact Caribbean. *Nature* (https://doi.org/10.1038/s41586-020-03053-2; published 23 December 2020)

Feutry P, Berry O, Kyne PM, Pillans RD, Hillary RM, Grewe PM, Marthick JR, Johnson G, Gunasekera RM, Bax NJ (2017) Inferring contemporary and historical genetic connectivity from juveniles. *Molecular Ecology* 26:444-456.

Feutry, P, Devloo-Delva, F, Tran Lu Y, A, Mona, S, Gunasekera, RM, Johnson, G, Pillans, RD, Jaccoud, D, Kilian, A, Morgan, DL, Saunders, T, Bax, NJ, and PM Kyne. 2020. One panel to rule them all: DArTcap genotyping for population structure, historical demography, and kinship analyses, and its application to a threatened shark. Molecular Ecology Resources 20(6): 1470-1485.

Fournier, D., Archibald, C.P., 1982. A general theory for analysing catch at age data. Can. J. Fish. Aquat. Sci. 39, 1195‑1207.

Frankham, R., 1995. Effective population size/adult population size ratios in wildlife: a review. *Genetics Research*, *66*(2):95-107.

Gilbert, K.J. and Whitlock, M.C., 2015. Evaluating methods for estimating local effective population size with and without migration. *Evolution*, *69*(8): 2154-2166.

Goldberg CS, Waits LP (2010) Quantification and reduction of bias from sampling larvae to infer population and landscape genetic structure. Molecular Ecology Resources, 10, 304–313.

Hauser, L., and G. R. Carvalho. 2008. Paradigm shifts in marine fisheries genetics: Ugly hypotheses slain by beautiful facts. Fish Fish. 9:333–362.

Hedgecock, D, A. I. Pudovkin. 2011. Sweepstakes reproductive success in highly fecund marine fish and shellfish: A review and commentary. Bull. Mar. Sci. 87:971–1002.

Hedrick, P., 2005. Large variance in reproductive success and the Ne/N ratio. *Evolution 59*:1596-1599.

Hill, W.G., 1972 Effective size of population with overlapping generations. *Theoretical Population Biology* 3:278–289.

Hill, W. G., 1981 Estimation of effective population size from data on linkage disequilibrium. Genet. Res. 38:209–216.

Hillary, R.M., Bravington, M.V., Patterson, T.A., Grewe, P., Bradford, R., Feutry, P., Gunasekera, R., Peddemors, V., Werry, J., Francis, M.P. and Duffy, C.A.J., 2018. Genetic relatedness reveals total population size of white sharks in eastern Australia and New Zealand. *Scientific reports*, *8*(1), p.2661.

Hollenbeck CM, Portnoy DS, Gold JR (2016) A method for detecting recent changes in contemporary effective population size from linkage disequilibrium at linked and unlinked loci. *Heredity* 117:207-16.

Holmes, E.E., 2001. Estimating risks in declining populations with poor data. *Proceedings of the National Academy of Sciences*, *98*(9):5072-5077.

Jaccoud, D., Peng, K., Feinstein, D., & Kilian, A. (2001). Diversity arrays: A solid state technology for sequence information independent genotyping. *Nucleic Acids Research*, *29*(4), e25.

Jolly, G. M. (1965). Explicit estimates from capture–recapture data with both death and immigration-stochastic model. *Biometrika* **52:**225–247.

Kendall KC, Stetz JB, Roon DA, Waits LP, Boulanger JB, Paetkau D. 2008. Grizzly bear density in Glacier National Park, Montana. *Journal of Wildlife Management* 72:1693-1705.

Korneliussen, T.S., Albrechtsen, A. and Nielsen, R., 2014. ANGSD: analysis of next generation sequencing data. *BMC bioinformatics*, *15*(1), p.356.

Krebs, CJ. 2009. Ecology: The Experimental Analysis of Distribution and Abundance, 6th Edition. *Pearson International Edition. San Francisco, USA.*

Krimbas, C.B. and Tsakas, S., 1971. The genetics of Dacus oleae. V. Changes of esterase polymorphism in a natural population following insecticide control—selection or drift?. *Evolution*, *25*(3):454-460.

Lebreton, J.D., Burnham, K.P., Clobert, J. and Anderson, D.R., 1992. Modeling survival and testing biological hypotheses using marked animals: a unified approach with case studies. *Ecological monographs*, *62*(1), pp.67-118.

Lee, A.M., S. Engen, and B.-E. Sæther, 2011 The influence of persistent individual differences and age at maturity on effective population size. *Proceedings of the Royal Society of London, Series B: Biological Sciences*, 278: 3303-3312.

Lee, A.M., Myhre, A.M., Markussen, S.S., Engen, S., Solberg, E.J., Haanes, H., Røed, K., Herfindal, I., Heim, M. and Sæther, B.E., 2020. Decomposing demographic contributions to the effective population size with moose as a case study. *Molecular ecology*, *29*(1):56-70.

Lehnert, SJ, T Kess, P Bentzen, MP Kent, S Lien, J Gilbey, M Clément, NW Jeffrey, RS Waples, and IR Bradbury. 2019. Genomic signatures and correlates of widespread population declines in salmon. Nature Communications 10:2996.

Lindberg, M.S., 2012. A review of designs for capture–mark–recapture studies in discrete time. *Journal of Ornithology*, *152*(2):355-370.

Luikart, G., N. Ryman, D. A. Tallmon, M. K. Schwartz, and F. W. Allendorf. 2010. Estimation of census and effective population sizes: the increasing usefulness of DNA-based approaches. *Conservation Genetics* 11:355–373.

Myhre, A. M., S. Engen, B.-E. Sæther. 2016. Efective size of density-dependent populations in fluctuating environments. Evolution 70:2431-2446

Nei M, Tajima F (1981) Genetic drift and estimation of effective population size. *Genetics* **98:**625–640.

Økland, J.M., Haaland, Ø.A. and Skaug, H.J., 2009. A method for defining management units based on genetically determined close relatives. *Ices Journal of Marine Science*, *67*(3):551-558.

Oleksiak, Marjorie F., and Om P. Rajora. 2020. Marine Population Genomics: Challenges and Opportunities. pp 3-35 in Population Genomics: Marine Organisms, Springer Nature, Cham Switzerland.

Opie, I, and Opie, P. 1997. The Oxford Dictionary of Nursery Rhymes (2nd ed.). Oxford University Press, Oxford, UK.

Palsbøll PJ, Zachariah Peery M, Berube M (2010) Detecting populations in the 'ambiguous' zone: kinship-based estimation of population structure at low genetic divergence. Molecular Ecology Resources, 10, 797–805.

Palstra, F.P., and D.J. Fraser, 2012 Effective/census population size ratio estimation: a compendium and appraisal. *Ecology and Evolution* 2:2357–2365.

Pollock, K.H., 2000. Capture-recapture models. *Journal of the American Statistical Association*, *95*(449):293-296.

Punt, A.E., Dunn, A., Elvarsson, B.Þ., Hampton, J., Hoyle, S.D., Maunder, M.N., Methot, R.D. and Nielsen, A., 2020. Essential features of the next-generation integrated fisheries stock assessment package: A perspective. *Fisheries Research*, *229*, p.105617.

Quinn, T.J. and Deriso, R.B., 1999. *Quantitative fish dynamics*. Oxford University Press.

Rawding DJ, Sharpe CS, Blankenship SM. 2014. Genetic-based estimates of adult Chinook salmon spawner abundance from carcass surveys and juvenile out-migrant traps. *Transactions of the American Fisheries Society* 143(1):55-67.

Robinson, J. D., and G. R. Moyer, 2013 Linkage disequilibrium and effective population size when generations overlap. Evol. Appl. 6:290–302.

Ruzzante, D.E., McCracken, G.R., Parmelee, S., Hill, K., Corrigan, A., MacMillan, J. and Walde, S.J., 2016. Effective number of breeders, effective population size and their relationship with census size in an iteroparous species, *Salvelinus fontinalis. Proceedings of the Royal Society B: Biological Sciences 283:*20152601.

Ruzzante, D.E., McCracken, G.R., Førland, B., MacMillan, J., Notte, D., Buhariwalla, C., Flemming, J.M. and Skaug, H. 2019. Validation of close-kin mark-recapture (CKMR) methods for estimating population abundance. *Methods Ecol Evol.* 10:1445–1453.

Santiago, E., Novo, I., Pardiñas, A.F., Saura, M., Wang, J. and Caballero, A., 2020. Recent demographic history inferred by high-resolution analysis of linkage disequilibrium. *Molecular Biology and Evolution*, *37*(12), pp.3642-3653.

Seber, G. A. F. 1982. The estimation of animal abundance and related parameters. Macmillan, New York.

Seber, G.A., 1986. A review of estimating animal abundance. *Biometrics 42*:267-292.

Shaw, A. K., and S. A. Levin. 2013. The evolution of intermittent breeding. *Journal of Mathematical Biology*, 66:685–703.

Simpfendorfer C, Cox S, Stokes K and Waples R (2021) Independent Expert Peer Review of the Close Kin Mark Recapture Assessment for School Shark. Report to AFMA. AFMA Research Project Number 190844.

Skaug, H. J. (2001). Allele-sharing methods for estimation of population size. *Biometrics* **57** 750–756.

Skaug, H.J., 2017. The parent–offspring probability when sampling age-structured populations. *Theoretical population biology*, *118*, pp.20-26.

Stewart, J.D., Jaine, F.R., Armstrong, A.J., Armstrong, A.O., Bennett, M.B., Burgess, K.B., Couturier, L.I., Croll, D.A., Cronin, M.R., Deakos, M.H. and Dudgeon, C.L., 2018.

Research priorities to support effective manta and devil ray conservation. *Frontiers in Marine Science*, *5*:314.

Tallmon, D.A, D. Gregovich, R.S. Waples, C.S. Baker, J. Jackson, B. Taylor, E. Archer, K.K. Martien, and M.K. Schwartz. 2010. When are genetic methods useful for estimating contemporary abundance and detecting population trends? Molecular Ecology Resources 10:684–692.

Taylor, B.L. and Wade, P.R., 2000. "Best" Abundance Estimates and Best Management: Why They Are Not the Same. In Ferson S, Burgman M, eds, *Quantitative methods for conservation biology* (pp. 96-108). Springer, New York, NY.

Tenesa A, Navarro P, Hayes BJ et al. (2007) Recent human effective population size estimated from linkage disequilibrium. *Genome Research* 17:520–526.

Thompson, E.A., 2013. Identity by descent: variation in meiosis, across genomes, and in populations. *Genetics*, *194*(2):301-326.

Thomson, R.B., Bravington M.V., Feutry, P., Gunasekera, R. and Grewe, P. (2020). Estimating the abundance of School Shark in Australia using close kin genetic methods. Final Report to FRDC Number 2014/024, August 2020. 108 pp. Available at: https://www.frdc.com.au/project/2014-024.

Wang, J., 2009 A new method for estimating effective population sizes from a single sample of multilocus genotypes. *Molecular Ecology* 18:2148–2164.

Wang, J., 2014. Estimation of migration rates from marker-based parentage analysis. *Molecular Ecology*, *23*(13):3191-3213.

Wang, J., 2016. A comparison of single-sample estimators of effective population sizes from genetic marker data. Molecular Ecology, 25, 4692–4711.

Wang, J., Brekke, P., Huchard, E., Knapp, L.A. and Cowlishaw, G., 2010. Estimation of parameters of inbreeding and genetic drift in populations with overlapping generations. *Evolution: International Journal of Organic Evolution*, *64*(6):1704-1718.

Waples, R.S. 2005. Genetic estimates of contemporary effective population size: To what time periods do the estimates apply? *Molecular Ecology* 14:3335-3352.

Waples, R.S. 2016. Life history traits and effective population size in species with overlapping generations revisited: the importance of adult mortality. Heredity 117:241-250.

Waples, R.S. 2020. An estimator of the Opportunity for Selection that is independent of mean fitness. Evolution 74:1942-1953.

Waples, RS. 2021. Relative precision of the sibship and LD methods for estimating effective population size with genomics-scale datasets. Journal of Heredity (published online 20 July 2021; DOI: 10.1093/jhered/esab042).

Waples, R.S., and C. Do. 2008. *LDNE*: A program for estimating effective population size from data on linkage disequilibrium. *Molecular Ecology Resources* 8:753-756.

Waples, R.S., and C. Do. 2010. Linkage disequilibrium estimates of contemporary N_e using highly variable genetic markers: A largely untapped resource for applied conservation and evolution. Evolutionary Applications 3:244-262.

Waples, R.S., C. Do, and J. Chopelet. 2011. Calculating N_e and N_e/N in age-structured populations: a hybrid Felsenstein-Hill approach. *Ecology* 92:1513-1522.

Waples, R.S., and P.R. England. 2011. Estimating contemporary effective population size based on linkage disequilibrium in the face of migration. *Genetics* 189:633–644.

Waples, R.S., G. Luikart, J.R. Faulkner, D.A. Tallmon. 2013. Simple life history traits explain key effective population size ratios across diverse taxa. Proc. Royal Society London, Ser. B. 280: 20131339.

Waples, R.S., and T. Antao. 2014. Intermittent breeding and constraints on litter size: consequences for effective population size per generation (N_e) and per reproductive cycle (N_b). *Evolution* 68:1722-1734.

Waples, R.S., T. Antao, and G. Luikart. 2014. Effects of overlapping generations on linkage disequilibrium estimates of effective population size. *Genetics* 197:769-780.

Waples, R.S., and E.C. Anderson. 2017. Purging putative siblings from population genetic datasets: A cautionary view. *Molecular Ecology* 26:1211-1224.

Waples RS, Grewe PG, Bravington MV, Hillary R, Feutry P. 2018a. Robust estimates of a high N_e/N ratio in a top marine predator, southern bluefin tuna. Science Advances 4:eaar7759.

Waples RS, Scribner K, Moore J, Draheim H, Etter D, Boersen, M. 2018b. Accounting for age structure and spatial structure in eco-evolutionary analyses of a large, mobile vertebrate. Journal of Heredity 109:709-723.

Waples, R.S., Waples, R.K., and Ward, E.J., 2021. Pseudoreplication in genomics-scale datasets. Molecular Ecology Resources (published online 4 August 2021; DOI: 10.1111/1755-0998.13482).

Whiteley AR, Coombs JA, Hudy M, Robinson Z, Nislow KH, Letcher BH (2012) Sampling strategies for estimating brook trout effective population size. Conservation Genetics, 13, 625–637.

Whiteley AR, Coombs JA, Cembrola M, O'Donnell MJ, Hudy M, et al. 2015. Effective number of breeders provides a link between interannual variation in stream flow and individual reproductive contribution in a stream salmonid. *Molecular Ecology* 24:3585-602.

Williams BK, J. D. Nichols, and M. J. Conroy. 2002. Analysis and Management of Animal Populations. Academic Press, San Diego, CA, USA, 2002.

Wright, S., 1938. Size of population and breeding structure in relation to evolution. *Science*, *87*:430-431.

Table 1. Definitions and notation used in this study. Unless specified, variables refer to one time period or year.

x	Age (in years)
z	Age at recruitment, when individuals are first inventoried
α	Age at maturity
ω	Maximum age
s_x	Probability of surviving from age x to age $x+1$
l_x	Cumulative probability of surviving to age x
N_1	Cohort size = number of offspring produced per time period that survive to age 1
N_x	Number of individuals of age x alive at any given time; $N_x = N_1 l_x$
N_T	Total number of individuals of all ages alive at any given time; $N_T = \Sigma N_x$
N_A (N_f, N_m)	Number of adults (females, males) with age$\geq\alpha$ alive at any given time
N_e	Effective population size per generation
N_b	Effective number of breeders in one year
$\hat{N}_A$, $\hat{N}_e$, $\hat{N}_b$	Estimates of adult census size and effective size
n	Number of individuals sampled
Y	Number of pairwise comparisons to search for close-kin matches
POP	Parent-offspring pair
HSP	Half-sibling pair
R	Number of close-kin matches identified (R_{POP} or R_{HSP})
k_i	Number of offspring produced by parent i
μ_k	Mean k_i for all parents
σ_k^2	Variance of k_i for all parents
b_x	Expected number of offspring for parents of age x
δ_i	Difference between expected number of offspring for individual i and the mean for that age and sex (b_x)
B_x	Total expected number of offspring produced by parents of age x: $B_x = N_x b_x$
$\sigma_{k(x)}^2$	Variance of k for adults of age x
ϕ_x	Ratio of variance to mean reproductive success for adults of age x; $\phi_x = \sigma_{k(x)}^2/b_x$
$\sigma_{k\bullet}^2$	Variance in lifetime number of offspring produced by the N_1 individuals in a cohort
RO	Reproductive output = number of offspring produced. $RO_i = k_i$ for individual i
TRO	Total reproductive output of all parents; $E(TRO) = \sum_{x=\alpha}^{\omega} B_x$
RRO	Relative reproductive output of an individual, compared to total reproductive output of all parents; $RRO_i = RO_i/TRO$.
$ERRO$	Expected relative reproductive output, based on a population's vital rates. For an individual of age x, $ERRO = E(RO_x)/E(TRO) = b_x / \sum_{x=\alpha}^{\omega} B_x$.
w_i	Relative probability that adult i will be the parent of a given offspring
W_i	$W_i = w_i/\Sigma w_i = ERRO_i$ = absolute probability that adult i will be the parent of a given offspring; $\Sigma W_i = 1$
CV_W^2	Squared coefficient of variation of the parental weights; $CV_W^2 = \sigma_W^2/\overline{W}^2$, where σ_W^2 = var(W). CV_W^2 is the same whether computed using w or W.

Table 2. Life table for a hypothetical species used in the simulations. s_x is the probability of surviving from age x to $x+1$, l_x is cumulative survival through age x, and b_x is relative fecundity of an individual of age x. This example assumes an equal sex ratio and the same vital rates in males and females, but data are shown for only one sex (nominally females). The hypothetical population produces 2000 offspring in each cohort, of which $N_1 = 1000$ are females. Expected numbers in successive age classes are defined by $N_x = l_x N_1$. Two fecundity scenarios are considered: constant fecundity and fecundity increasing with age. $TRO = \Sigma N_x b_x$ is total reproductive output of the population in one time period, in the same relative units as b_x, and $ERRO$ is the expected relative reproductive output of an individual of the specified age (all individuals of the same age are assumed to have the same $ERRO$). $ERRO$ is identical to the standardized parental weights described in Section 2.2.

Age (x)	s_x	l_x	N_x	Constant Fecundity b_x	$N_x b_x$	ERRO	Increasing Fecundity b_x	$N_x b_x$	ERRO
1	0.7	1.000	1000	0	0	0	0	0	0
2	0.7	0.700	700	0	0	0	0	0	0
3	0.7	0.490	490	1	490	0.00065	3	1470	0.00040
4	0.7	0.343	343	1	343	0.00065	4	1372	0.00053
5	0.7	0.240	240	1	240	0.00065	5	1200	0.00067
6	0.7	0.168	168	1	168	0.00065	6	1008	0.00081
7	0.7	0.118	118	1	118	0.00065	7	826	0.00094
8	0.7	0.082	82	1	82	0.00065	8	656	0.00107
9	0.7	0.058	58	1	58	0.00065	9	522	0.00121
10	0	0.040	40	1	40	0.00065	10	400	0.00134

Total 1+ (N_T) 3239   TRO 1539   TRO 7454
Adults 3+ (N_A) 1539

Table 3. Comparison of expected and estimated parameters for six scenarios evaluated in the simulations. Except for scenario F, for which cohort size and adult N are 10 times as large, all scenarios have a constant cohort size of N_1 = 2000 newborns, which produces adult N_A = 3078 (1539 of each sex; Table 2). n_{off} and $n_{parents}$ are annual sample sizes of offspring and potential parents, respectively. In selective sampling (Scenario D), the relative probability that an adult was sampled was equal to its expected fecundity (b_x); in other scenarios, adult sampling was random. Samples of potential parents were drawn only from mature (age 3+) females except in Scenario E, where the sample was collected from the entire population. Sampling of adults has no effect on siblings, so results for Scenarios D and E are only shown for POPs. Exp(R) and Obs(R) are expected and observed numbers of close-kin matches, respectively, and the conversion factor C is the constant that relates the true probability of a match to the naive value $1/N_f$ (see Eq. S3). For sibship analyses there is a separate C value for each age gap between offspring, and values shown here are weighted mean $\bar{C}$ values applicable to the total number of comparisons. Raw (naïve) $\widehat{N}_A$ is the estimate from Equation 7, assuming all $P_{match} = 1/N_f$; the adjusted estimator $\widehat{N}_{A(adj)}$ accounts for effects of mortality, changes in fecundity with age, and sampling using the conversion factor C.

	Scenario	A	B	C	D	E	F
	N_1	2000	2000	2000	2000	2000	**20000**
	fecundity	constant	constant	**increasing**	**increasing**	constant	constant
	adult sampling	random	random	random	**selective**	**random**[a]	random
	ϕ	1	**10**	1	1	1	1
	N_A	3078	3078	3078	3078	3078	**30783**
POPs	n_{off}	300	300	300	300	300	600
	$n_{parents}$	300	300	300	300	300	600
	Exp(R)	58.5	58.5	58.5	67.5	27.9	23.4
	Obs(R)	59.2	59.2	58.9	66.9	27.8	23.1
	C	1.00	1.00	1.00	1.15	0.48	1.00
	$\widehat{N}_A$ raw	3072	3182	3135	2802	6722	32556
	$\widehat{N}_{A(adj)}$	3072	3182	3135	3094	3201	32556
Siblings	n_{off}	100	100	100	-	-	300
	Exp(R)	63.7	63.7	84.3	-	-	57.4
	Obs(R)	63.8	67.7	82.4	-	-	58.8
	$\bar{C}$	0.4905	0.4905	0.648			0.4905
	$\widehat{N}_A$ raw	6504	6577	4854	-	-	61348
	$\widehat{N}_{A(adj)}$	3122	3157	3145	-	-	29447
	True N_b	3078	388	2665	-	-	30783
	$\widehat{N}_b$	3074	390	2658	-	-	28805

[a]In Scenario E, sampling of potential parents was random but with respect to the entire population rather than just mature adults

Table 4. Summary of major factors that, unless explicitly accounted for, can influence performance of CKMR methods to estimate abundance. 'RS' = reproductive success.

Factor	POPs	HSPs
Ageing individuals	Necessary if required to separate parents/offspring, account for changes in fecundity with age, or calculate ERRO in the year of offspring's birth	Required to sort individuals into cohorts and calculate age gaps
Sampling of cohorts	Samples from single cohorts can be used for semelparous species. Sampling multiple cohorts is essential for most iteroparous species	Sampling multiple cohorts is essential because only cross-cohort comparisons provide information about census size
Age-specific fecundity	Affects precision; needs to be estimated if selectivity varies with age	Affects precision and bias; can be estimated independently or from POPs + ages
Adult mortality	Not required unless sampling of adults is non-lethal	Required to adjust probability a parent produces sibs in different years, but can be estimated from HSP data
Overdispersed variation in RS ($\phi>1$)	Does not affect P_{POP} but reduces precision	Reduces precision. For bias, see below
HSPs from the same cohort		Siblings from same cohort estimate N_b, not adult N
HSPs from different cohorts		
Residual RS uncorrelated across years		No bias in $\widehat{N}_A$
Persistent individual differences in RS		Increases P_{HSP} and will bias $\widehat{N}_A$ unless accounted for
Skip breeding	Downward bias in $\widehat{N}_A$ if skip breeders not sampled. With adequate samples, pattern of skip breeding can be estimated from the data	Complex effects on P_{HSP}. With adequate samples, pattern of skip breeding can be estimated from the data
Non-random sampling concerns	POPs collected together more often than by chance; Probability of sampling an adult correlated with its RS	Siblings born in different years collected together more often than would occur by chance
Precision	In simple scenarios with $\phi\approx1$, CV($1/\widehat{N}_A$) = 0.15 requires ~50 POPs, which implies sample sizes on the order of $\sqrt{N_A}$ individuals	Somewhat more sensitive than POPs to factors that inflate variance, but var(R_{HSP}) generally converges on $\bar{R}_{HSP}$, for large N
Benefits of genomics-scale datasets	Scant: performance asymptotes at low numbers of genetic markers after POPS reliably identified	Once all siblings are correctly identified (which generally requires a few thousand SNPs), no benefits from adding more markers

Table 5. Summary of major factors that, unless explicitly accounted for, can influence performance of some genetic methods to estimate effective size. 'RS' = reproductive success.

Factor	Sibship method	LD method
Ageing individuals	Required to sort individuals into cohorts to estimate N_b or to estimate N_e from mixed-age samples	Required to sort individuals into cohorts to estimate N_b; not required to estimate N_e from mixed-age samples
Sampling of cohorts	A single-cohort sample provides a point estimate of N_b. Multiple such samples provide a temporal series of estimates	A single-cohort sample provides a point estimate of N_b. Multiple such samples provide a temporal series of estimates. Mixed-age samples can estimate N_e
Changes in fecundity or mortality with age; skewed sex ratio; overdispersed variance in RS ($\phi>1$)	No bias; these are part of the signal of reduced N_e and N_b	No bias; these are part of the signal of reduced N_e and N_b
Skip breeding and persistent individual differences in RS	No bias to single-cohort samples estimating N_b	No bias to estimates of N_b or N_e, but this affects the ratio N_b/N_e and influences the bias adjustments to account for age structure
Non-random sampling issues	Estimate is downwardly biased if siblings are collected together more often than would occur by chance	Estimate is downwardly biased if relatives are collected together more often than would occur by chance
Precision	For large populations, sample sizes on the order of $\sqrt{N_e}$ individuals	Precision jointly determined by samples of individuals and loci
Benefits of genomics-scale datasets	Once all siblings are correctly identified (which generally requires a few thousand SNPs), no benefits from adding more markers	With more than about 500-2000 SNPs, precision of LD method will meet or exceed maximum precision possible with sibship method
Changes in N and/or N_e	No effect on estimates based on single cohorts	Some bias toward recent N_e for 3-5 generations

Figure 1. Schematic diagram showing possible effects of skip breeding on CKMR estimates. Top: If non-breeders are not sampled, POPs will provide an abundance estimate only for the breeding population. Bottom: A multi-year study for a population in which most adults alternate between being breeders and non-breeders; the thin dashed arrows indicate that a fraction of adults breed in consecutive years. Two consecutive years of POPs collections will sample the entire adult population. Few if any HSPs will be found for pairs of individuals separated in age by odd numbers of years (sampling scheme A), while the number of HSPs for comparisons of individuals born an even number of years apart (sampling scheme B) will reflect the number of breeders per year more than the adult census size, N_A. Genetic methods that estimate N_b based on single-cohort samples also would produce an estimate related to the number of actual breeders each year, rather than the total number of adults.

Figure 2. POPs results for simulated scenarios A-E from Table 2. Symbols show medians of $\widehat{N}_A$ across 500 replicate simulated samples; vertical lines show empirical confidence intervals (CIs) defined by the upper 95% and lower 5% quantiles. All scenarios have the same true adult $N_A = 3078$ (dotted line). Estimated N_A ($\widehat{N}_A$; Y axis) is the raw estimate based on the observed number of POPs and assuming that $P_{POP} = 1/N_f$; all estimates adjusted using the conversion factor C to account for deviations from this assumption were close to true N_A (see Table 3). Note the log scale on the Y axis.

Figure 3. Mean numbers of siblings recorded in the simulations as a function of the age gap between offspring, for scenarios that involved intermittent breeding. The last column shows the total number of cross-cohort siblings found (age gaps 1-4). The scenario labeled 'Breeders: All' (black bars) is scenario A from Table 3. In the '50%' and '10%' scenarios (red and green bars), only a random half or 10% of all adults were allowed to reproduce each year; in the 'Alternate' scenario (yellow bars) only individuals aged 3, 5, 7, or 9 were allowed to reproduce, so every surviving parent skipped one year after reproduction. Note the log scale on the Y axis.

Figure 4. Precision in estimating adult census size (N_A) using CKMR analyses based on the number of half sibling pairs (HSPs). Data are for Scenario A from Table 3, with true $N_A = 3078$ (black dotted line) and 89 offspring sampled from 5 consecutive cohorts. Black triangles are possible estimates of N_A using R = integer total numbers of HSPs (maternal and paternal combined) in Equation 7, with the mean conversion factor ($\bar{C} = 0.4905$) playing the role of the function f_2. This level of sampling produces an average of $\bar{R} = 50$ sibling matches which, if var(R) were Poisson, would produce a standard error of $\bar{R}$ of about 7 and a 95% confidence interval (CI) around $\bar{R}$ of [36-64] (dashed red line); this in turn would (following red arrows) lead to a 95% CI of $\hat{N}_A = [2400\text{-}4240]$. In the simulations, actual var(R) among replicates was slightly higher than the Poisson expectation (1.13 times the mean; Figure S2B), which would lead to a slightly wider CI, but the empirical CV($1/\hat{N}_A$) was still close to 0.15 (Figure 5C).

Figure 5. Relationships between annual sample size (*n*), adult abundance (*N_A*), precision, and cost.   In panels A and B, the red line represents sample sizes that are expected to produce a total of 50 cross-cohort half siblings, assuming samples are collected for 5 consecutive years. Panel C shows results of simulations based on two reproductive-skew variations described in Table 3: Scenario A (ϕ=1, solid red line and symbols) and Scenario B (ϕ=10, dotted blue line and symbols). Cohort sizes were varied to produce adult abundances that ranged more than three orders of magnitude, and corresponding sample sizes were adjusted as shown in Panel A to produce a constant expected number of 50 total HSPs. In panel D, the green dashed line shows approximate total cost (across all 5 years) for SNP discovery and genotyping for sample sizes expected to produce 50 HSPs. Vertical axes: A) Annual sample size, *n*; B) *n* as a fraction of *N_A*; C) CV of $1/\widehat{N}_A$; D) left, *n*; right, total cost in thousands of USD. Note the log scale on the *X* axis, and on the *Y* axes in Panels A and D.

Figure 6. Precision for using close-kin sibling data to estimate adult N_A (black lines) or N_b (gray lines) as a function of the number of consecutive years during which samples are collected. CV (*Y* axis) is the coefficient of variation of $1/\widehat{N}_A$ or $1/\widehat{N}_b$, using data for all years. Fecundity was constant with age and true N_A was 3078. Results apply to Scenario A ($\phi = 1$, solid lines; true N_b = 3078) or Scenario B ($\phi = 10$, dashed and dotted lines; true N_b = 388).

Figure 7. Using complementary properties of genetic and demographic estimators to bracket the plausible range for the ratio N_e/N_A in southern bluefin tuna. The solid black line is the CKMR estimate of adult abundance (2×10^6) from Bravington et al. (2014); Waples et al. (2018a) used the same genetic and demographic data, respectively, to delimit a lower bound (dashed line) and an upper bound (dotted line) to N_e. Genetic methods can provide robust lower bounds for N_e but, because the drift signal is so small, have difficulty distinguishing large from very large or infinite effective sizes. Conversely, life history constrains how large N_e can be but, without very detailed demographic information, cannot rule out very small N_e/N_A ratios. Collectively, and in conjunction with a robust estimate of N_A from CKMR, the two effective-size methods constrain the plausible range of N_e/N_A to about 0.1-0.5.