Population dynamics of the primary malaria vector *Nyssorhynchus darlingi* in a high transmission setting dominated by fish farming in western Amazonian Brazil

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Abstract

Vale do Rio Juruá in western Acre, Brazil, has reported highest malaria numbers since 2005, and is considered persistent transmission hotspot. Fish farming development was encouraged to improve standard of living, resulting in productive breeding sites for Amazonian malaria vector species, including *Nyssorhynchus darlingi* that, combined with the high human density and mobility, adds to the local malaria burden. This study reports entomological profile of immature and adult *Ny. darlingi* at three sites in Mâncio Lima, Acre, during the rainy and dry season (February to September, 2017). From 63 fishponds, 10,859 larvae were collected, including 5,512 first-instar Anophelinae larvae and 4,927

second, third and fourth-instars, of which 8.5% (n = 420) were *Ny. darlingi*. This species was most abundant in not-abandoned fishponds and in the presence of emerging aquatic vegetation. Seasonal analysis of immatures in urban landscapes found no significant difference in the numbers of *Ny. darlingi*, corresponding to equivalent population density during the rainy to dry transition period. However, in the rural landscape, significantly higher numbers of *Ny. darlingi* larvae were collected in August (IRR = 5.80, $p = 0.037$) and September (IRR = 6.62, $p = 0.023$) (dry season), compared to February (rainy season), suggesting important role of fishponds for vector population maintenance during the seasonal transition in this landscape type. Adult sampling detected mainly *Ny. darlingi* (~93%), with similar outdoor feeding behavior, but different abundance according to landscape profile: urban site 1 showed higher peaks of human biting rate in May (46 bites/person/hour), than February (4) and September (15), while rural site 3 shows similar HBR during the same sampling period (22, 24 and 21, respectively). This study contributes to a better understanding of the larvae biology of the main malaria vector in the *Vale do Rio Juruá* region and, ultimately will support vector control efforts.

Introduction

The link between anthropogenic environmental change and the emergence of malaria is well-documented in the Amazon basin [1-3]. Increased human population and land use/land cover change (LULC) influence the biological community, including Anophelinae mosquitoes, particularly those with some degree of synanthropy and competence to transmit *Plasmodium* sp. that circulate in the Amazon region [4]. This vast

region is responsible for 99.5% of human malaria in Brazil, mainly *Plasmodium vivax* (>90% in 2019) [5]. Disease indicators vary according to the types of LULC and the socio-environmental aspects of occupied environments, influencing spatiotemporal malaria distribution trends [6]. Although from 2008-2016 Brazil reported annual reductions of the disease, with 2016 having the lowest incidence in the past 35 years, in 2017 the incidence increased by 50% compared with the previous year, decreasing only in 2019 [5]. This resurgence emphasizes the need for routine and integrated surveillance, even when disease rates are low, a characteristic feature of seasonal infectious diseases [7]. A key factor involved in the successful eradication policy of mosquito-borne diseases with a broad distribution and different focal transmission, such as malaria in the Amazon, is the identification and characterization of vector sources, following evaluation of potential tools for an integrated intervention framework [8].

Fish farming has been associated with malaria risk in the Amazon in Brazil [9], Peru [10], Colombia [11,12], and in sub-Saharan Africa in Nigeria [13] and Cote d'Ivoire [14]. The Vale do Juruá, Mâncio Lima municipality, is a classic example of the potential hazards of extensive fish farming in a periurban/urban setting. A local government program provided resources to residents to construct fish farms, frequently located in their backyards. The unwanted effect of this development program was the increased number of suitable larval habitats of *Nyssorhynchus darlingi* and other local malaria vectors which affect density and spatial distribution and threaten control strategies in the area [15-17]. Nowadays, the Vale do Juruá in western Acre is the region with the highest malaria numbers in Brazil, for both *P. vivax* and *P. falciparum.* In a scenario where anthropogenic fish farms have been demonstrated to be major contributors to vector abundance and

Plasmodium transmission, larval source management (LSM) can be a practical component of integrated vector management (IVM) to reduce or eliminate immature stages of mosquito vectors [18-20]. Further, the recognition that variation in larval habitats, particularly in nutrient availability, strongly influences mosquito fitness, longevity, and malaria transmission dynamics, has renewed interest in larval environments [21,22]. On the other hand, LSM as part of a vector-borne disease control management plan has limitations when dealing with natural aquatic habitats in rural and forest areas, especially when breeding sites are extensive, inaccessible, and require frequent intervention such as clearing aquatic vegetation [23, 24]. To address the application and effectiveness of any control strategies on mosquito borne-disease transmission, local vector biology information is essential, considering the diversity of *Ny. darlingi* in different environmental profiles of the Amazon Basin, reflected in malaria epidemiology. Although entomological surveys addressing Anopheline larvae and the main vector *Ny. darlingi* presence in fishponds have been conducted in the Vale do Jurua [15-17], these studies did not focus on follow-up with short intervals between observations (one/two months per collection), nor characterize environmentally the fishponds associated with larvae sampled.

In the present study, an entomological survey of larvae and adult malaria vectors was conducted to evaluate the presence of the main vector *Ny. darlingi* in fishponds and neighboring households in Mâncio Lima, Acre. To address this, our study examined: (i) aquatic habitat parameters associated with Anophelinae larval abundance; (ii) differences in the abundance of *Ny. darlingi* during the rainy to dry seasonal transition; (iii) the microgeographic effect of urban and rural landscapes on the population dynamics of *Ny.*

112 *darlingi*; and (iv) a comparison of human biting rates (HBR) and patterns of *Ny*. *darlingi*
113 biting times influenced by different landscape scenarios.

114

115 Methods

116 Ethics Statement

117 This study was approved by the World Health Organization Ethics Review
118 Committee (0002669). Verbal consent was obtained from residents for collections on their
119 properties, with the collaboration of the Mâncio Lima Endemics Diseases Coordination. A
120 monthly report of fishpond physiochemical conditions was provided to each resident. Adult
121 captures were conducted only by the authors, who used antimalarial prophylaxis as
122 recommended by the Brazilian Ministry of Health.

123

124 Study Area

125 The municipality of Mâncio Lima is located in western Acre state, Brazil (7° 36' 50"
126 S 72° 53' 45" W) along Highway BR 364 (Fig 1). An Anophelinae larval survey in artificial
127 and natural breeding-sites reported four times more immatures in fishponds [15] compared
128 with natural habitats. A time-series analysis (2003 to 2013), strongly suggested a
129 spatiotemporal association between fish farming and malaria incidence [16]. The estimated
130 population of Mâncio Lima is 17,545 [25], with the municipality registering for *P. vivax*:
131 6,632 infections in 2016 (API = 378 per 1000 habitants) and 7,049 infections in 2017

(API= 400); for *P. falciparum*: 1,172 in 2016 (API= 70) and 1,752 in 2017 (API = 99.8) (http://www2.datasus.gov.br/DATASUS, 2018). Notifications for monthly malaria shows significant linear correlation (>0.5) with rainfall: for *P. vivax* in 2016: $r = 0.75$, in 2017: $r = 0.43$; for *P. falciparum*: in 2016: $r = 0.51$, in 2017: $r = 0.47$ (S1 Fig). The most recent livestock census (2016) registered a total of 5,392 cattle in Mâncio Lima, mainly in rural areas (unpublished document, Institute of Agriculture and Forestry Defense of Acre, 2016).

Fig 1. Satellite image of Mâncio Lima municipality, showing the three study sites. Site 1, urban, yellow, near Federal Highway BR 364 and Mâncio Lima town; Site 2, urban, red, near Federal Highway BR 364 and more distant from Mâncio Lima; Site 3, rural, green, distant from both BR 354 and Mâncio Lima. Each site shows the residence and two perimeters: 0.5 and 1.0 km (source: ©2017 Google-Images ©2017 TerraMetrics, Nasa, Cartographical data©2017 Maplink). The insert is a map of Brazil indicating the location of Mancio Lima in Acre state.

Study Design

This research entailed an observational study of malaria vector ecology. For the Anophelinae survey, independent geographical areas were delimited based on two sampling criteria: the presence of a human residence occupied for at least the past 12 months for adult mosquito collection, and nearby fishponds for larval collection, whether economically active (used for pisciculture at the moment of the survey), or abandoned. Two perimeters

153 (500 m and 1000 m) were virtually attributed for each residence to delimit each study site,
154 and to support the localization of fishponds (Fig 1). These distances were chosen based on
155 the flight range of *Ny. darlingi* in a rural settlement in Rondonia state, between 500 and
156 1000 m [26]. To test the influence of an urban area on local transmission, two sites (Sites 1
157 and 2) were selected near Federal Highway BR 364; and one site (Site 3) that was more
158 distant from the highway (Fig 1). Highway BR 364 is important for socio-economic
159 landscape concepts in Acre state: usually, urban landscape profiles include paved streets
160 and have several residences and other human dwellings (schools, hospital, commercial
161 facilities), and this may be reflected in a higher number of families and houses, leading to
162 the establishment of more fishponds. On the other hand, rural landscape profiles consist of
163 a lower human presence, fewer dwellings, and primary or secondary forest cover, if the
164 landscape has not been exploited for logging, agriculture, or livestock [27]. The presence of
165 at least one fishpond near the house (within at least 500 m), positive previous larvae and
166 adult captures (in December 2016), ease of access to the property, and co-operation of the
167 residents were other considerations for the three residence selections and the respective
168 representative sites.

169

170 Larval and Adult Capture

171 Monthly larval collections were performed for six months in 2017 spanning rainy
172 and dry season (February, March, April, May, August, and September). Each fishpond was
173 sampled by 1) determining fixed sampling-points along fishpond margins (n=4, A-D); and
174 2) sampling by dipper at 10 sampling-points along each margin. The 10 dips were evenly

distributed according to the length of each margin. Dippers were standard: 10 cm in diameter, with a volume of 350 ml and a 1.5 m long handle, and white in color for better visibility of immatures [28]. Larval specimens were placed in 50 ml microtubes labeled according to sampling date, site, fishpond number and sampling-point margin letter (A-D) and number (1-10). All material was fixed in the field in 80% ethanol. Presence of aquatic fauna collected in the dippers were also recorded (i.e., *Culex* sp., amphibians, fish).

Adult collections were performed at each of the three sites in February, May, and September 2017. We used human landing catch (HLC), performed only by the professionally trained authors (two people indoors and two peridomestic simultaneously, rotating every two hours at each spot), using manual aspirators to capture mosquitoes, for 12 h /night (18:00-06:00). Collected mosquitoes were separated by date, location, and hour of capture. In months and sites with low mosquito density, we sampled one additional night (12h) and adjusted later for analysis. In February, there were two night collections at Site 1, and one in Sites 2 and 3, respectively. In May and September, two collections were done at Site 2, and one in Sites 1 and 3. Mosquitoes were stored in silica gel in microtubes (50 ml) identified with a code that included: month, site, date, and hour of collection. On rainy nights, adult captures were suspended and conducted on the following non-rainy night.

Field-collected specimens were identified at the Laboratory of Infectious Diseases of the Federal University of Acre (UFAC - campus Cruzeiro do Sul, Acre state) and at the School of Public Health of the University of São Paulo (USP - campus São Paulo, São Paulo state). Adults and the larval stages L2–L4 were identified using a stereomicroscope and entomological keys (Forattini, 2002). Because of the challenge to identify L1 morphologically [29], three larval groups were defined: Anophelinae L1 stage;

Anophelinae L2–L4 stages and *Ny. darlingi* L2–L4 stages: in this approach, Anophelinae L2–L4 group included no *Ny. darlingi* species. After morphological identification, adults and larvae were sent to the Biotechnological Institute of University of State of São Paulo (UNESP - Campus Botucatu, São Paulo State) for further molecular analysis.

Environmental Variables

Fishponds were classified and measured according to environmental and physical-chemical conditions. For the environment, categorical variables included periodicity (permanent or temporary during the 6-month study period); abandoned fishpond- no maintenance by the owner (yes or no); associated vegetation on the margins of fishpond (if present: emerging, submerged, floating); the presence of *Culex* sp., amphibians, and fish. For periodicity and abandoned by the owner, classification was at the fishpond level; for vegetation and presence of other animals, classification was at the sampling-point level. Physical-chemical variables included pH, temperature, and conductivity, measured using an ExTECH multiparameter (extech.com/) probe that presented continuous values. However, due to functionality limitations, data from this device were collected only in the first three months (February, March, April). For the remaining three months (May, August, September), pH, nitrates (mg/L), nitrites (mg/L), carbonate hardness (KH) and dissolved chlorine (mg/L) were collected using a JBL ProScan kit (jbl.de/en), by immersion of a test strip in the water and reading by smartphone app downloaded at Google Play Store (play.google.com/store/apps/details?id=de.jbl.proscan). The data collected using the JBL Proscan kit had a more limited range, i.e., categorical variables. Collections using the

ExTECH multiparameter probe were obtained at the sampling-point level; for the JBL Proscan test kit, data were obtained at the fishpond level. Turbidity and shading were also obtained only during the last three months of the survey (May, August, September). Water turbidity was determined at the fishpond level using a LaMotte (lamotte.com) water column test kit, with discrete values ranging between 0 - 200 JTU (where 0 represents translucent water), at the fishpond level. Shading by canopy was collected at the sampling-point level with a TerraGes spherical densitometer according to the manufacturer's specifications (terrages.pt), with continuous values ranging between 0 – 24.96 1⁄4"-squares (where 0 represents shaded and 24.96 represents completely exposed), at the sampling-point level. This information is summarized in S1 Table.

Monthly precipitation data were obtained from the CPTEC/INPE website (clima1.cptec.inpe.br/). Adverse weather/air conditions (rain, mist, wind, smoke from burning) were noted when they occurred during the adult night collections. Field information was digitally stored through Open Data Kit (ODK). Data were compiled in EXCEL (Microsoft). Visual resources (photographs) were also obtained from each sampling-point, by ODK function. Georeferencing of the residences and fishponds was conducted using GPS Garmin device and Google Earth Pro TM software.

Data Setting and Statistical Analysis

Statistical analyses were conducted to establish the association between larval groups, environment, and physical-chemical variables, using multilevel regression models.

The seasonal pattern, according to rainfall trends of western Amazon Basin (Rainy Season: Oct-Feb; Dry Season: Apr-Sept, see S1 Fig), was analyzed, considering the repeated measures framework used for larval sampling: the month of the collection was assumed to be a variable factor, with February being the chosen reference baseline according to rainfall seasonality effect on Culicidae biology abundance [29]. Therefore we chose February to represent a rainy month; September to represent a dry month; and the interval between February and September as the rainy-dry transition (S1 Fig). Larval counts of three groups (Anophelinae L1; Anophelinae L2 – L4; *Ny. darlingi* L2 - L4) were considered the outcome variables.

Overdispersion was observed in data distribution resulting from large numbers of zero values, thus a binomial negative regression analysis was used [30]. According to assumptions of a negative binomial distribution, and the respective nature of dependent variables, regression coefficients are presented as incidence rate ratios (IRRs), defined by the number of events (Anophelinae counts) by fishpond (analysis unit) [31]. For all tests, the statistical significance level assumed was 0.05. An initial univariate regression was performed to verify any associations between single independent variables. Considering the non-randomized approach, multivariate regression was performed to verify adjustments in the coefficients. A cut-off value for p of less than 0.2 of univariate analysis was chosen, and the order of insertion of the independent variable in the multivariate regression was from the lowest to the highest p-value considering the univariate analysis [32]. Multicollinearity was assessed for the following independent variables used in multivariate analysis, since they were measured at the same sampling level: linear correlation for numerical variables

(continuous physical-chemical) and Spearman rank correlation coefficient for ordinal variables (categorical physical-chemical).

Considering the hierarchical data structure (samplings-points nested within fishponds), a mixed-effects model was conducted, mainly due to its flexibility in repeated measures modeling of unbalanced data [33]. The dataset was structured in a long format, with the *i*th row functioning as a time-point per specific sampling-point, and respective fishpond (the subject of the analysis) [34]. Considering the biology of Anophelinae the three study sites were not considered independent, the usual procedure for mixed models that simulates repeated measures ANOVA, due to geographic proximity between sites (mainly Sites 1 and 2) [35]. In addition to the overall regression, to distinguish effects among sites, regressions were performed for each site. An unconditional model was built first, followed by a model with a random component to indicate the subject of the repeated term. A two-level model was chosen, combining sampling-points at the first level and fishponds at the second level as the random component, according to the data structure (sampling-points nested in fishponds). Due to some gaps in variables measured during the monthly survey, a full dataset was the primary design (six-months), using the respective independent environmental variables: periodicity, abandoned, associated vegetation, presence of animals, and collection month. For physical-chemical variables, the turbidity of water and shading, which were not possible to measure during the whole six-month survey, three-month datasets were designed according to independent variables: pH, temperature and conductibility (continuous data) in February, March, and April; shading, turbidity and pH, nitrates, nitrites, carbonated hardness and dissolved chlorine (ordinal data), measured in May, August, and September. For three-month physical-chemical datasets, individual

regressions were not performed for each site, due to reduced sampling effort. This information is summarized in S1 Table.

For comparison of categorical variables between sites, as well as the hypothesized adult abundance difference between indoor and outdoor, a Chi-square test was used. Outliers and systematic errors were verified through box-plot graphs. All statistical analyses were performed using Stata 14.2 (data analysis and statistical software - StataCorp LP, College Station, TX, USA). A robust option for the variance component estimators (VCE) was chosen according to the Stata configuration.

RESULTS

Sampling sites

Sixty-three fishponds in the three sites were identified and followed during the 2017 study period. Total numbers of fishponds monitored throughout the field survey was variable because of seasonal precipitation or occasionally being emptied by owners, and some fishponds could not be reached across flooded fields at some point during the sampling period. Fig 2 shows satellite images for the three sites with each nearby residence and the fishponds surveyed; number of fishponds sampled by period, along with dry conditions and other characteristics is presented in S2 Table. S3 Table shows environment variables by site: both urban Sites 1 (~66%, 92/139) and 2 (~88%, 126/143) show a higher number of not abandoned fishponds compared with rural Site 3 (~22%, 12/64) ($p < 0.001$). No significant difference in fishpond periodicity was identified between three sites

307 (p=0.625); all sites had high numbers of permanent fishponds: (Site 1: ~86%, 120/139; Site
308 2: 85% 122/143; Site 3: ~84%, 46/55).

309

310 **Fig 2: Satellite image of three study sites, with respective residence marked by**
311 **pushpin and fishponds by flags.** Note, using the virtual perimeters of each residence-
312 index, the different scales and fishpond distributions at each site. (Source: ©2017 Google-
313 Images ©2017 TerraMetrics, Nasa, Cartographical data©2017 Maplink).

314

315 Larval Collection

316 During the six-month sampling period of 2017, 10,859 larvae were collected:
317 n=5,512 corresponded to the group of Anophelinae L1 stage species; n=4,927 to the group
318 of Anophelinae L2–L4 stage species; and n=420 to the group of *Ny. darlingi* L2–L4 stages.
319 Urban Site 1 shows the highest number of larvae (n=6,065), followed by rural Site 3
320 (n=3,017) and urban Site 2 (n=1,777). Rural Site 3 had the highest density of larvae per
321 fishpond (54.85), followed by urban Sites 1 (43.63) and 2 (12.46). For the *Ny. darlingi* L2–
322 L4 group, urban Site 1 had the highest total number (249) and density per breeding site
323 (1.80); urban Site 2 had 97 and 0.68, respectively; rural Site 3 had 74 and 1.32,
324 respectively.

325 Anophelinae species and *Ny. darlingi* (both in L2–L4 stages) distributed by
326 fishpond, site, and period are shown in Fig 3. The higher proportion of non-*Ny. darlingi*
327 Anophelinae species compared with *Ny. darlingi* in practically all fishponds during the

rainy season from February through May (except for Fishpond number 07, Site 1, in February) is noteworthy. The exclusive presence of *Ny. darlingi* was observed in some fishponds at urban Site 2, however, these count values were minimal (1 or 2 specimens). For the dry season (August and September), urban Site 1 had more fishponds that were positive exclusively for *Ny. darlingi*: 01 in August; 01, 04, 08, 11 18, 25, and 26 in September. Urban Site 2 also had some fishponds with *Ny. darlingi* exclusively, but also with low counts. Rural Site 3 had *Ny. darlingi* in August and September in fishponds 54, 55 and 58 where this species was not observed during the rainy season months.

Fig 3. Summary of larvae collected in the study. Anophelinae spp. *and Ny. darlingi* (both L2 to L4 stages), distributed by fishpond, site and period.

Statistical Analysis

Table 1 shows the IRR coefficient results for the six-month dataset using univariate analysis, by identified Anophelinae larval group. Table 2 shows the multivariate analysis results, according to the selection criteria for independent variables. The Spearman rank coefficient detected no correlation between independent variables of the six-month dataset, with values lower than 0.1, except for the presence of *Culex* sp. and amphibians (0.34), and submerged aquatic vegetation with floating vegetation (0.23).

Table 1: Incidence Rate Ratio (IRR), 95% Confidence Interval and *p* values for Anophelinae identity-group, for six month survey univariate two level negative binomial regression:

Anophelinae identity-group	Independent Variable		Univariate Two-level Negative Binomial							
			Overall		Site 1		Site 2		Site 3	
			IRR (95% C.I.)	*p*	IRR (95% C.I.)	*p*	IRR (95% C.I.)	*p*	IRR (95% C.I.)	*p*
Anophelinae species (L1)	Month Collection	February	1		1		1		1	
		March	1.23 (.78-1.92)	0.373	1.19(.64-2.23)	0.572	.87 (.23-3.26)	0.839	1.19(.61-2.30)	0.613
		April	1.29 (.73-2.25)	0.374	1.16(.58-2.33)	0.658	3.75(.90-15.60)	0.069	.39 (.18-.82)	0.013
		May	.94 (.55-1.61)	0.842	1.07(.56-2.06)	0.822	1.05(.28-3.89)	0.940	.35 (.13-.81)	0.016
		August	.56 (.24-1.28)	0.175	.13 (.04-.35)	0.000	4.21(1.27-13.99)	0.019	.64 (.19-2.20)	0.487
		September	.63 (.29-1.34)	0.256	.25 (.09-.64)	0.004	1.99 (.63 - 6.29)	0.239	1.21(.38-3.85)	0.742
	Periodicity	Temporary	1		1		1		1	
		Permanent	2.19 (.55-8.70)	0.262	1.32 (.45-3.87)	0.608	5.01 (.73-34.07)	0.100	2.12(1.20-3.75)	0.009
	Abandoned	Yes	1		1		1		1	
		No	.37 (.13-1.02)	0.056	1.63 (.70-3.77)	0.249	2.49 (.26-23.44)	0.423	1.08 (.32-3.59)	0.900
	Associated Vegetation	Emerging	1.86(1.07-3.25)	0.021	1.05 (.43-2.59)	0.906	3.98(1.25-12.68)	0.020	1.45 (.93-2.28)	0.097
		Submerged	.74 (.39-1.38)	0.343	.48 (.20-1.14)	0.101	.83 (.39-1.75)	0.629	.64 (.16-2.47)	0.521

		Floating	1.06 (.60-1.89)	0.874	.73 (.31-1.70)	0.473	1.83 (.84-4.01)	0.129	.82 (.46-1.47)	0.511
	Presence	*Culex* sp.	1.73(1.16-2.57)	0.007	1.69(1.04-2.74)	0.032	1.79 (.64-5.05)	0.264	1.84 (.9-3.78)	0.094
		Amphibian	1.53 (.96-2.44)	0.073	1.32 (.86-2.03)	0.193	5.78 (.56-59.4)	0.140	1.34 (.59-3.01)	0.474
		Fish	3.34(1.58-7.04)	0.002	5.55(2.19-14.09)	0.000	.45 (.072-2.84)	0.398	2.09 (.72-6.06)	0.175
Anophelinae species (L2, L3, L4)	Month Collection	February	1		1		1		1	
		March	1.12 (.69-1.79)	0.649	1.30(.677-2.51)	0.426	1.12 (.44-2.86)	0.807	.77 (.38-1.54)	0.457
		April	1.17 (.69-2.01)	0.550	1.20 (.66-2.18)	0.543	4.65(1.11-19.46)	0.035	.30 (.17-.54)	0.000
		May	1.07 (.63-1.82)	0.796	1.25 (.64-2.46)	0.512	1.93 (.72-5.21)	0.193	.33(.15-.73)	0.007
		August	.85 (.38-1.92)	0.709	.28 (.08-.87)	0.028	3.51(1.22-10.18)	0.020	1.09(.33-3.67)	0.878
		September	.87 (.41-1.84)	0.717	.49 (.16-1.56)	0.232	2.28 (.72-7.25)	0.161	1.11 (.45-2.95)	0.835
	Periodicity	Temporary	1		1		1		1	
		Permanent	2.88 (.54-15.31)	0.214	1.24 (.21-7.33)	0.805	13.9(1.54-125.35)	0.019	1.69 (.82-3.5)	0.150
	Abandoned	Yes	1		1		1		1	
		No	.44 (.15-1.26)	0.129	2.34 (.83-6.64)	0.107	1.82 (.18-18.42)	0.610	1.15 (.35-3.77)	0.811
	Associated Vegetation	Emerging	1.82(1.01-3.31)	0.050	.95 (.43-2.07)	0.902	3.71(1.06-13.04)	0.040	1.89 (.68-5.19)	0.217
		Submerged	.55 (.27-1.12)	0.098	.29 (.13-.68)	0.004	.56 (.22-1.45)	0.235	.98 (.21-4.60)	0.985

		Floating	1.41 (.78-2.53)	0.252	1.37 (.61-3.10)	0.445	1.77 (.65-4.86)	0.264	.73 (.64-.83)	0.000
	Presence	*Culex* sp.	1.78(1.16-2.74)	0.009	1.86(1.09-3.19)	0.023	1.06 (.32-3.57)	0.917	2.2 (1.1-4.39)	0.025
		Amphibian	1.37 (.74-2.52)	0.313	1.11 (.55-2.23)	0.762	5.44 (.87-33.83)	0.069	1.29 (.46-3.65)	0.620
		Fish	2.92(1.23-6.93)	0.015	2.84(1.08-7.51)	0.035	3.92 (.33-46.71)	0.279	4.18 (.54-32.11)	0.169
Ny. darling (L2, L3, L4)	Month Collection	February	1		1		1		1	
		March	.74 (.34-1.60)	0.449	.55 (.18-1.65)	0.294	.94 (.22-3.93)	0.928	1.31 (.35-4.93)	0.683
		April	1.01(.48-2.11)	0.984	.89 (.32-2.45)	0.831	1.82(.27-12.35)	0.538	.21 (.02-2.26)	0.194
		May	.81 (.41-1.57)	0.532	.74 (.30-1.81)	0.512	.56 (.11-2.79)	0.480	1.03 (.14-7.58)	0.971
		August	1.39 (.57-3.41)	0.465	.39 (.09-1.58)	0.190	3.01 (.75-11.94)	0.118	5.80(1.08-30.94)	0.039
		September	1.55(.68-3.54)	0.291	.91 (.28-2.88)	0.879	1.61(.42-6.09)	0.481	6.25(1.19-32.76)	0.030
	Periodicity	Temporary	1		1		1		1	
		Permanent	1.21 (.38-3.77)	0.742	.86 (.22-3.36)	0.835	5.89 (.46-75.12)	0.172	.654 (.18-2.27)	0.504
	Abandoned	Yes	1		1		1		1	
		No	1.81 (.66-4.95)	0.243	10.27(2.7-39.07)	0.001	1.00 (.21-4.87)	0.996	.91 (.19-4.27)	0.906
	Associated Vegetation	Emerging	2.08(1.25-3.46)	0.005	1.61 (.94-2.76)	0.079	3.07 (.96-9.78)	0.058	2.08 (.49-8.72)	0.318
		Submerged	1.45 (.68-3.07)	0.334	1.27(.50-3.22)	0.611	.79 (.26-2.42)	0.692	2.77 (.43-17.63)	0.279

		Floating	1.37 (.73-2.57)	0.320	1.62 (.73-3.60)	0.232	1.48 (.43-5.0)	0.527	*	0.000
	Presence	*Culex* sp.	1.23 (.73-2.07)	0.433	1.27 (.72-2.25)	0.407	2.2 (.55-8.78)	0.261	.57 (.21-1.55)	0.274
		Amphibian	1.32 (.63-2.76)	0.455	1.05 (.53-2.08)	0.888	6.53(2.34-18.21)	0.000	.88 (.065-11.78)	0.924
		Fish	1.24 (.6-2.57)	0.558	1.55 (.64-3.73)	0.328	.71 (.11-4.85)	0.723	1.48 (.22-9.81)	0.683

Statistically significant values at the 0.05 level are highlighted

*IRR value omitted due low decimal number (10^{-10}).

Table 2: Incidence Rate Ratio (IRR), 95% Confidence Interval and *p* values for Anophelinae species identity-group, for six-month survey multivariate two level negative binomial regression:

Anophelinae identity-group	**Independent Variable**		**Multivariate Two-level Negative Binomial**					
			Site 1		**Site 2**		**Site 3**	
			IRR (95%C.I.)	***p***	**IRR (95%C.I.)**	***p***	**IRR (95%C.I.)**	***p***
Anophelinae species (L1)	Month Collection	February	1		1		1	
		March	1.14 (.56-2.31)	0.721	.89 (.23-3.50)	0.870	1.21(.62-2.37)	0.579
		April	1.30 (.52-3.24)	0.568	2.30(.57-9.28)	0.242	.33 (.16-.69)	0.004
		May	1.31 (.56-3.04)	0.534	1.07(.29-3.92)	0.916	.25 (.13-.50)	0.000
		August	.21 (.07-.56)	0.003	3.70(1.05-13.03)	0.041	.54 (.16-1.79)	0.317
		September	.31 (.10-.91)	0.033	1.19 (.33-4.22)	0.789	1.19(.31-4.55)	0.797

	Periodicity	Temporary			1		1	
		Permanent			2.98 (.53-16.77)	0.214	1.98 (.88-4.45	0.096
	Associated Vegetation	Emerging			5.27 (.91-30.46)	0.063	2.79(1.51-5.17)	0.001
		Submerged	.67 (.38-1.20)	0.182				
		Floating			.62 (.14-2.73)	0.531		
	Presence	*Culex* sp.	1.50 (.99-2.27)	0.056			1.07 (.66-1.74)	0.773
		Amphibian	.97 (.55-1.73)	0.932	6.10(1.13-32.94)	0.036		
		Fish	1.94 (.72-5.23)	0.192			2.03 (.76-5.39)	0.157
Anophelinae species (L2, L3, L4)	Month Collection	February	1		1		1	
		March	1.43 (.76-2.67)	0.268	1.09 (.42-2.80)	0.864	.81 (.37-1.76)	0.593
		April	2.00 (1.11-3.59)	0.020	2.56 (.73-8.99)	0.141	.46 (.17-1.25)	0.129
		May	2.48 (1.29-4.76)	0.006	1.64 (.58-4.62)	0.351	.46 (.15-1.41)	0.175
		August	.55 (.18-1.68)	0.293	2.74 (.95-7.88)	0.062	1.71 (.55-5.33)	0.353
		September	.81 (.26-2.50)	0.721	1.46 (.31-6.77)	0.628	2.40 (.60-9.58)	0.214
	Periodicity	Temporary			1		1	
		Permanent			9.01(1.36-59.77)	0.023	1.12 (.61-	0.721

							2.06)	
	Abandoned	Yes	1					
		No	1.65 (.76-3.56)	0.201				
	Associated Vegetation	Emerging			3.04(1.14-8.15)	0.027	66 (.22-1.96)	0.454
		Submerged	.34 (.19-.58)	0.000				
	Presence	*Culex* sp.	2.24 (1.36-3.69)	0.001			1.96 (.82-4.72)	0.131
		Amphibian			6.47(1.89-22.14)	0.003		
		Fish	1.12 (.43-2.86)	0.819			6.14 (.86-43.76)	0.070
Ny. darlingi (L2, L3, L4)	Month Collection	February	1		1		1	
		March	.65 (.21-1.98)	0.452	1.11 (.22-5.54)	0.897	1.31 (.35-4.86)	0.681
		April	1.01 (.36-2.81)	0.989	1.36 (.20-9.14)	0.753	.21 (.02-2.31)	0.204
		May	.66 (.26-1.64)	0.371	.83 (.16-4.36)	0.825	1.04 (.14-7.55)	0.967
		August	.48 (.11-2.00)	0.314	2.71 (.57-12.88)	0.209	5.80(1.11-30.41)	0.037
		September	1.11 (.34-3.65)	0.864	1.32 (.31-5.67)	0.705	6.62(1.29-33.89)	0.023
	Periodicity	Temporary			1			
		Permanent			3.84 (.55-26.57)	0.173		
	Abandoned	Yes	1					

		No	11.40 (3.06-42.52)	0.000				
	Associated Vegetation	Emerging	2.12 (.83-5.45)	0.117	2.23 (.79-6.23)	0.126		
		Submerged					*	0.000
	Presence	Amphibian			5.86(2.42-14.156)	0.000		

Statistically significant values at the 0.05 level are highlighted

*IRR value omitted due low decimal number (10^{-10}).

Considering the February baseline value and a statistical significance at 95% C.I., seasonality differences were not detected in the overall regression for the three identified Anophelinae larval groups. For each site, Anophelinae L1 shows a particular pattern in the univariate regression: a decrease in counts for August and September in urban Site 1, and April and May in rural Site 3; and an increase in counts for urban Site 2 in August. The IRR values observed in the multivariate analysis were maintained relative to the univariate IRR values, indicating that these results were not influenced by possible confounding factors. The Anophelinae L2 - L4 group also shows a unique pattern for each site, however, in this case the IRR values were substantially different between the univariate and multivariate analysis. For the *Ny. darlingi* L2 - L4 group, there was a similar pattern in both urban sites (1 and 2), with no significant statistical difference in monthly larval numbers for the baseline value (February). In rural Site 3, an increase was observed for August [5.8 (95% C.I.:1.11-30.41)] and for September [6.62 (95% C.I.:1.29-33.89)]. Multivariate and univariate regression showed comparable IRR values in the three sites. For periodicity and

abandoned characteristics, permanent condition was significant for Anophelinae L1 group in rural Site 3, for univariate analysis only for the Anophelinae L2 - L4 group, whereas urban Site 2 shows an increase in larval number for both univariate and multivariate regression. The non-abandoned condition for *Ny. darlingi* shows an increase in larval number in urban Site 1, for both univariate and multivariate regression. Emerging associated vegetation shows an increase of larval number in the overall regression for all three Anophelinae groups. Presence of *Culex* sp. and fish was significant for Anophelinae L1 and L2, L3 and L4, in the overall regression. Urban Site 1 showed a similar association for both groups, however only the Anophelinae L2 - L4 group maintains this value in multivariate analysis. The presence of amphibians was positively associated with Anophelinae L1 and L2- L4 groups in urban Site 2 only in the multivariate regression. The *Ny. darlingi* group showed a positive association with amphibian presence only in urban Site 2, for both univariate and multivariate regressions.

Table 3 shows IRR coefficient results for a three-month dataset using univariate analysis, by identified Anophelinae larval group. Table 4 shows multivariate analysis, according to the selection criteria for independent variables. The Spearman rank test shows a high correlation between the categorical physical-chemical variables nitrates and nitrites (r=0.89), and a low correlation between carbonate hardness and pH (r=0.38), and carbonate hardness and dissolved chlorine (r=0.4). Physical-chemistry variables for continuous values (pH, temperature, and conductibility) were not statistically associated at 95% C.I. with the abundance of any of the three larval groups. Turbidity shows a significant negative association for the Anophelinae L1 only in the univariate regression [0.98 (IC95%:.97-.99)]. *Ny. darlingi* L2 - L4 shows a significant positive association with turbidity in

multivariate analysis only [1.01 (IC95%:1.00-1.01), p=0.045]. Shading reduction shows a significant negative association with the abundance of both Anophelinae L1, and Anophelinae L2 - L4 in both univariate and multivariate regressions, but for the *Ny. darlingi* group, the univariate was not significant at the 0.05 level, however, it was near the limit [0.96 (IC95%:.93-1.00) with p=0.052], whereas in the multivariate analysis shading was significant [0.95 (IC95%:.92-.99), p=0.02].

Table 3: Incidence Rate Ratio (IRR), 95% Confidence Interval and p values for Anophelinae species identity-group, for three-month survey univariate two level negative binomial regression:

Independent Variable		**Univariate Two-level Negative Binomial**					
		Anophelinae species (L1)		**Anophelinae species (L2, L3, L4)**		***Ny. darlingi* (L2, L3, L4)**	
		IRR (95% C.I.)	***p***	**IRR (95% C.I.)**	***p***	**IRR(95% C.I.)**	***p***
Physical-Chemistry (continual values)	pH	1.02 (.78-1.33)	0.901	.91 (.69-1.21)	0.521	1.05(.72-1.54)	0.779
	Temperature	.92 (.81-1.06)	0.259	.95 (.82-1.1)	0.505	.87 (.71-1.06)	0.165
	Conductibility	1.00 (.99-1.01)	0.432	1.01 (.99-1.07)	0.297	1.02 (.99-1.009)	0.488
Turbidity (discrete value)		.98 (.97-.99)	0.050	.98 (.97-1.00)	0.120	1.01 (.99 - 1.02)	0.100
Shading (continual value)		.95 (.91-.99)	0.008	.96 (.94-	0.035	.96 (.93-1.00)	0.052

					.99)			
Physical-Chemistry (categorical values)	pH	>6	1		1		1	
		6.4	.14 (.05-.39)	0.000	.06 (.03-.12)	0.000	.87 (.43-1.76)	0.695
		6.6	.26 (.03-2.36)	0.231	.19 (.01-3.02)	0.237	1.97 (.61-6.5)	0.261
		6.8	.79 (.20-3.15)	0.747	.14 (.01-3.06)	0.210	.87 (.13-5.79)	0.884
		7	.31 (.04-2.20)	0.245	.10 (.02-.46)	0.003	*	0.000
		7.2	*	0.000	*	0.000	*	0.000
		7.6	*	0.000	*	0.000	*	0.000
	nitrates (mg/L)	0	1		1		1	
		10	1.53 (.51-4.64)	0.449	1.39 (.51-3.83)	0.519	.91 (.38-2.19)	0.839
		25	1.39 (.22-7.84)	0.705	1.36 (.23-7.95)	0.734	.72 (.15-3.59)	0.688
		40	.22 (.16-.32)	0.000	.06 (.04-.09)	0.000	*	0.000
	nitrites (mg/L)	0	1		1		1	
		0.25	1.05 (.44-2.49)	0.917	.84 (.40-1.77)	0.651	.96 (.42-2.21)	0.934
		0.5	1.54 (.16-14.64)	0.707	1.27 (.09-16.31)	0.854	.47 (.04-5.57)	0.547
	carbonated hardness (KH)	0	1		1		1	
		1.5	1.90 (.98-3.68)	0.056	2.05 (1.01-4.19)	0.049	.89 (.60-1.34)	0.604

		3	.71 (.25-1.99)	0.515	93 (.22-3.96)	0.921	.24 (.04-1.28)	0.094
		4.5	.17 (.06-.47)	0.001	*	0.000	.30 (.08-1.2)	0.089
		6	.10 (.04-.27)	0.000	*	0.000	.56 (.14-2.26)	0.415
		8	.13 (.06-.26)	0.000	.33 (.14-.79)	0.013	.82 (.19-3.49)	0.791
		15	3.49(1.69-7.17)	0.001	.23 (.12-.43)	0.000	2.08(.93-4.62)	0.073
	dissolved chlorine (mg/L)	0	1		1		1	
		0.8	.94 (.26-3.39)	0.922	.66 (.15-2.99)	0.588	.82 (.27-2.51)	0.730
		1.5	2.04 (.94-4.42)	0.070	.72 (.28-1.88)	0.506	2.21(1.17-4.19)	0.015

Statistically significant values at the 0.05 level are highlighted

*IRR value omitted due low decimal number (10^{-10}).

Table 4. Incidence Rate Ratio (IRR), 95% Confidence Interval and *p* values for Anophelinae species identify-group, for three-month survey multivariate two level negative binomial regression:

Independent Variable	**Multivariate Two-level Negative Binomial**					
	Anophelinae species (L1)		**Anophelinae species (L2, L3, L4)**		***Ny. darlingi* (L2, L3, L4)**	
	IRR (95% C.I.)	***p***	**IRR (95% C.I.)**	***p***	**IRR(95% C.I.)**	***p***

Turbidity (discrete value)			.99 (.98-1.00)	0.087	.99 (.98-1.01)	0.216	1.01(1.00-1.01)	0.045
Shading (continual value)			.95 (.92-.98)	0.003	.97 (.94-.99)	0.041	.95 (.92-.99)	0.020
Physical-Chemistry (categorical values)	pH	>6	1		1		1	
		6.4	.18 (.03-1.06)	0.059	.08 (.025-.24)	0.000	.28 (.06-1.37)	0.118
		6.6	.71 (.06-8.19)	0.782	.68 (.04-10.58)	0.781	.84 (.12-5.62)	0.856
		6.8	6.95 (.35-136.97)	0.202	4.73 (.49-45.68)	0.179	.53 (.04-7.78)	0.645
		7	.34 (.02-6.38)	0.473	.27 (.03-2.73)	0.268	*	0.000
		7.2	*	0.000	*	0.000	*	0.000
		7.6	*	0.000	*	0.000	*	0.000
	nitrates (mg/L)	0	1		1		1	
		10	1.13 (.43-2.93)	0.804	1.12 (.46-2.72)	0.796	.74 (.27-2.07)	0.574
		25	.79 (.06-9.86)	0.861	.91 (.15-5.43)	0.915	.74 (.08-7.16)	0.794
		40	.68 (.18-2.62)	0.577	.30 (.09-1.04)	0.057	*	0.000
	carbonated hardness (KH)	0	1		1		1	
		1.5	1.39 (.75-2.58)	0.287	1.38 (.78-2.44)	0.272	.85 (.57-1.26)	0.423
		3	.33 (.14-.75)	0.008	.39 (.18-1.21)	0.103	.24 (.09-.63)	0.003
		4.5	.01 (.01-.52)	0.020	*	0.000	.13 (.01-1.88)	0.135

		6	.03 (.02-.38)	0.006	*	0.000	.37 (.08-1.71)	0.203
		8	.15 (.05-.46)	0.001	.25 (.07-.86)	0.027	3.87 (.99-15.17)	0.052
		15	.04 (.02-.51)	0.013	*	0.000	.34 (.03-4.23)	0.401
	dissolved chlorine (mg/L)	0	1				1	
		0.8	1.67 (.63-4.37)	0.299			1.34 (.64-2.78)	0.435
		1.5	4.23 (1.58-11.36)	0.004			3.41(1.51-7.68)	0.003

Statistically significant values at the 0.05 level are highlighted

*IRR value omitted due low decimal number (10^{-10}).

For ordinal physical-chemistry variables, increased pH values were associated with decreased in larval counts in all three groups, for both univariate and multivariate regression. Similarly, the highest nitrate level (40 mg/L) was associated with decreased larval counts for all three larval groups, and this was maintained in multivariate analysis for the *Ny. darlingi* group. Nitrites were not significantly associated with larval counts, and excluded from the multivariate regression analyses. For carbonated hardness, whereas both Anophelinae L1 and L2, L3 and L4 groups show a highly significant negative association (except Anophelinae L1 at 15 KH in univariate analysis, not kept in the multivariate regression), the *Ny. darlingi* group did not show statistical significance for any range, except a decrease in larval numbers observed at a range of 3 KH in the multivariate analysis [0.24 (IC95%:.09-.63)]. Dissolved chlorine showed a significant positive association at a range of 1.5 mg/L for the Anophelinae L1 in multivariate analysis [4.23 (IC95%:1.58-

11.36)], and the *Ny. darlingi* for both univariate [2.21 (IC95%:1.17-4.19)] and multivariate analysis [3.41 (IC95%:1.51-7.68)].

Adult Collection

A total of 692 Anophelinae specimens was collected and identified as *Ny. darlingi*. Fig 4 shows HBR for each site, adjusted for two night captures depending on site (Site 1 in February; Site 2 in May and September). There was a significant difference in the proportion of indoor vs. outdoor *Ny. darlingi* among the 3 sites (X^2=19.833, p<0.001), with a higher abundance in the peridomestic area. The proportion indoors was higher in Site 3 (~25%) than in Site 1 (~12%) or Site 2 (~11%). In Site 1, May showed a higher number of *Ny. darlingi* in all night captures (21/173 indoor/outdoor) than February (3/13 indoor/outdoor) or September (5/29 indoor/outdoor). Site 2 showed the lowest adult collections: February (3/10 indoor/outdoor), May (0/8 indoor/outdoor), September (0/7 indoor/outdoor). In Site 3, mosquito numbers were consistently high for outdoor collections, and increased for indoor captures in the last two months: February (11/104, indoor/outdoor), May (51/101 indoor/outdoor), September (42/97, indoor/outdoor).

Regarding HBR per hour, Site 1 shows more activity in May between 19:00-20:00 (HBR= 46), while February presents low numbers between 19:00-20:00 (HBR=4), as does September (18:00-19:00=13). Site 2 also presents low numbers (with peaks reaching at maximum of four mosquitoes/hr). In contrast, *Ny. darlingi* from Site 3 showed higher outdoors peaks during the first part of the night (February, 19:00-20:00=22; May, 18:00-19:00=24; September, 20:00-21:00=21), also demonstrating, besides low values, some

451 indoor peaks that exceeded outdoors ones, in May (00:00-01:00, indoor=12, outdoor=7;
452 02:00-03:00, indoor=7, outdoor=0) and September (for 00:00-01:00, indoor=10,
453 outdoor=7).

454

455 **Fig 4. Blood-feeding pattern by Human Biting Rate (HBR: *Ny. darlingi* per**
456 **human captures per hour), by night-capture, site and period.**

457

458 # Discussion

459 For effective control of Amazon malaria transmission it is essential to recognize the
460 diverse eco-epidemiologic profiles of the disease in local areas: municipalities, cities,
461 districts, subdistricts, along with “off the grid” areas: mining, rubber extraction (*seringal*),
462 rural settlements and indigenous populations. For such heterogeneity, the design and the
463 application of specific control methodologies according to each eco-epidemiologic profile
464 is needed [36]. The Brazilian Amazonian Basin has a total area of five million km²
465 (corresponding to an estimated 60% of the federal territory), but only 15% of the human
466 population, most in big cities and state capitals [25]. This heterogeneous distribution is
467 reflected in local characteristics of vector biology, thus malaria epidemiology, following
468 human dynamics that drive Amazonian occupation [37]. Interdisciplinary methods for
469 disease intervention are common but rarely tailored to specific local conditions [38, 39].
470 For effective eradication at a global scale, many aspects of public health need to be
471 included, such epidemiological and syndromic surveillance, early diagnosis, clinical

472 treatment, environmental sanitation, and improved methods for economical land use to
473 reduce inequity and poverty [40].

474 Ours is the first study to conduct a detailed microgeographic spatiotemporal analysis of
475 larvae and adult Anophelinae, with a focus on the major vector, *Ny. darlingi*, in Vale do
476 Jurua, western Acre, characterized by high malaria transmission associated with urban and
477 periurban fishponds. In this area, we determined that *Ny. darlingi* larval dynamics was not
478 affected by seasonality in urban landscapes, similar to findings in previous studies in the
479 area [15,16]. This feature may help to maintain the population density of *Ny. darlingi*
480 during the transition of rainy to dry seasons. We detected other fishpond characteristics
481 associated with *Ny. darlingi* abundance: active fishponds, emergent vegetation (normally
482 secondary growth that has emerged from deforested areas), and shade. A particular
483 fishpond characteristic verified by the present study was the presence of *Ny. darlingi* larvae
484 in water with dissolved chlorine, suggesting possible resilience for chemical pollution [41],
485 although the increase of pH and nitrates was observed as a limiting factor. Adult collections
486 were conducted for a single night per study site per month and therefore our conclusions
487 are preliminary. Most *Ny. darlingi* were collected outdoors, during the first part of the night
488 (18:00-00:00), a pattern reported for this species in other Amazonian occupied areas [42-
489 46], although we observed a greater abundance (not significant) and peaks of indoor
490 activity in the rural landscape.

491 In Amazonian malaria transmission, the most common type of breeding site, whether
492 natural or artificial (or both), contributes substantially to the dynamics and seasonality of
493 malaria [26,47,48]. Two main variables of natural aquatic habitats that affect larval survival
494 are water flow intensity during the rainy season (larval mortality rate); and low water

capacity during the dry season (loss of available aquatic niche) [24,48]. These conditions are generally neutralized in artificial aquatic habitats such as dams, micro dams, cisterns, fishponds, and other types of flow-limited water bodies [23], increasing the vectorial capacity of primary vectors, such as *Ny. darlingi* [49]. A successful breeding site in a malaria-endemic region should provide geographic and temporal coexistence for the epidemiologic triad: vector, etiological agent, and human reservoir, according to ecological strategies of Anopheline species [50,51] as well *Plasmodium* sp. [52], facilitating adaptation to host behavior [53]. The presence of the primary malaria vector in human residences and adjacent fishponds in Mancio Lima suggests that transmission may occur both in and around houses, although our HBR data demonstrate that most biting occurs outdoors.

A lower proportion of *Ny. darlingi* larval specimens was identified in the present study (8.5%) compared to that found in the same municipality in earlier studies (16.1% [15]; 22.5% [54]), and in other distinctive local Amazonian environments [24,48]. However, *Ny. darlingi* L1 larvae were not morphologically identified herein and this stage represented more than 50% of the total numbers of larvae surveyed. Furthermore, similar to other entomological studies in malaria-endemic areas, our adult survey detected only *Ny. darling* [43-46], although we recognize that HLC can generate a bias due to the mainly anthropophilic behavior of this species, as well for *Nyssorhynchus* sp. in general. Biodiversity of Anophelinae can be an indicator of environmental disruption, a putative signal of future outbreaks [55]. There is both a notably increased abundance and/or the emergence of *Nyssorhynchus* species in human-colonized Amazonian areas [44,56], and low natural abundance of this genus in primary Amazon forest [57,58]. *Nyssorhynchus*

518 *darlingi* is not always the dominant species in the *Nyssorhynchus* larval community that
519 emerges with anthropogenic change: for example, in Mâncio Lima, Acre state, it is *Ny.*
520 *albitarsis s.l.* [15]; in Labrea, Amazon state, *Ny. triannulatus* [24]; and in Pôrto Velho,
521 Rondônia state, *Ny. braziliensis* [59]. However, *Ny. darlingi* may be the species that best
522 adapts to human behavior in the Amazon region relative to vectorial capacity [49,60-62].

523 Our study was noteworthy for the micro-geographical analysis of larvae sampled,
524 measuring different characteristics of vector ecology at sampling-point and fishpond levels.
525 Anophelinae species L1 showed different behavior among the three sites: urban Site 1 had a
526 decrease in August and September; however there was an increase in August in urban Site
527 2; whereas for rural Site 3, there was a decrease in larval counts in April and May. These
528 results did not change in the multivariate analysis, in contrast with Anophelinae L2 - L4,
529 which present an inverse association in urban Site 1 after adjustment, indicating some
530 influential cofactor that was not measured by this study.

531 Nevertheless, the primary vector *Ny. darlingi* L2 - L4 group - identified to species level-
532 shows a singular pattern: no difference of larval numbers in fishponds detected in urban
533 Sites 1 and 2 during the rainy to dry season transition, in both univariate and multivariate
534 analysis. There was also no seasonal difference for *Ny. darlingi* in early study [15],
535 however, they incorporated a larger time frame (2 years) with larger intervals between
536 larval sampling efforts (5-6 months). Interestingly, in our Site 3 (rural), there was a
537 significant increase in larval numbers between the February baseline and both August and
538 September, months that correspond to the dry season. Possibly, fishponds play a more
539 important role in the maintenance of *Ny. darlingi* during the transition from rainy to dry

season in rural landscapes than our study demonstrates. Similar results were found in rural settlements with the presence of artificial breeding-sites [10,47].

Seasonal malaria is common in the Amazon region, associated with *Ny. darlingi* population density and rainfall patterns. In urban and suburban areas in Rondônia state [45], malaria increased at the end of the dry season and the beginning of the next rainy season in landscapes with mainly natural breeding-sites (riverside malaria); in contrast, in landscape dominated by artificial breeding-sites (so-called dryland malaria), both malaria and *Ny. darlingi* remain high throughout the year. Here, a simple linear correlation between monthly precipitation and *P. vivax* notifications showed a positive association for rainfall seasonality and malaria cases, mainly in 2016 (2016 r = 0.75; 2017 = 0.43), indicating some seasonal effect on malaria numbers (S1 Fig). However, these monthly notifications could have been more informative had they been adjusted for the appropriate landscape profile (urban/rural). For *Ny. darlingi* larvae, a major sampling effort with more sites in each landscape type in a multi-year survey is needed to confirm this seasonal pattern.

In urban Site 1, the increase of *Ny. darlingi* larvae in active fishponds, not detected for Anophelinae L1 and L2 - L4 groups, supports the earlier study [15], demonstrating that economically active fishponds are important larval habitats for primary vectors. Emerging aquatic vegetation was strongly associated with all three Anophelinae groups in the overall regression, reinforcing the recommendation by WHO [63], that cleaning the margins can be an effective environmental control for Amazon *Nyssorhynchus* sp. The presence of *Culex* sp. species and egg rafts was constant in the survey, suggesting they share the same ecological niche as the Anophelinae L2 - L4 group. Most *Culex* sp. were identified as subgenus *Melanoconion*, a group that contains species that are regional arbovirus vectors

563 [64]. Thus, fish farming may open larval habitats for other Culicidae species of
564 epidemiological importance. The presence of fish was common in the fishponds surveyed
565 (even abandoned ones), showing that Anophelinae larval species readily coexist with the
566 local fish community, or amphibians according to a microecological food web of aquatic
567 habitat [65]. Prospects for putative biological control seem unclear in this case unless exotic
568 larvivorous fish species were to be utilized, but they represent other risks for the local
569 environment and are not a feasible option [66].

570 Water turbidity was slightly associated with Anophelinae numbers, with *Ny. darlingi*
571 being found previously in turbid water [24]. We report a significant association with shaded
572 or low light environments for the three Anophelinae groups, a feature associated previously
573 with *Ny. darlingi* ecology [48]. High values of pH (>7) and nitrates (40 mg/L) appear to be
574 limiting factors for the Anophelinae aquatic habitat. Although carbonated hardness (an
575 alkalinity indicator), shows a similar pattern in the decrease of Anophelinae L1 and L2 - L4
576 groups, for *Ny. darlingi* there was no significant association. More surprisingly, the
577 increase in *Ny. darlingi* larvae in waters with dissolved chlorine suggested possible
578 tolerance of immatures to polluted aquatic habitats. This was also detected for the
579 Anophelinae L1, representing an important feature of opportunistic species that invade new
580 aquatic niches in human occupation without environmental sanitation, and may be linked to
581 phenotypic plasticity of ion regulation of Amazon mosquito Culicidae larvae under
582 different physical-chemical conditions [67].

583 Aside from the non-identification of L1 larvae, mentioned above, a second limitation of
584 this study was that we planned to measure the perimeter of each fishpond to test for an
585 association with larval abundance [10]. We initially measured each fishpond but, due the

high number of ponds (n=63), it was not realistic to accurately measure change in water level in each one for each of the six months. Thirdly, there were some technical problems with measuring instruments, resulting in gaps in some of the independent variables of the survey, reducing sampling effort. Fourthly, we intensively sampled two urban sites but only one rural one, mainly due to complex logistical issues. Finally, there is an important relationship between households with malaria incidence and distance to breeding sites for *Ny. darlingi* [48] and *P. vivax* infection [68] measurement of which was beyond the scope of our study.

Nevertheless, our study does provide important information about temporal variation and environment features of larvae of the primary vector *Ny. darlingi* at micro-spatial levels (sampling points of fishponds), as well as *Ny. darlingi* adult profiles in nearby households. Tailored LSM strategies accounting for this heterogeneity, such the use of biological larvicides [69], need to be routinely incorporated in malaria integrated control to reduce transmission in Mâncio Lima, and in other cities of Vale do Jurua region.

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Support Information

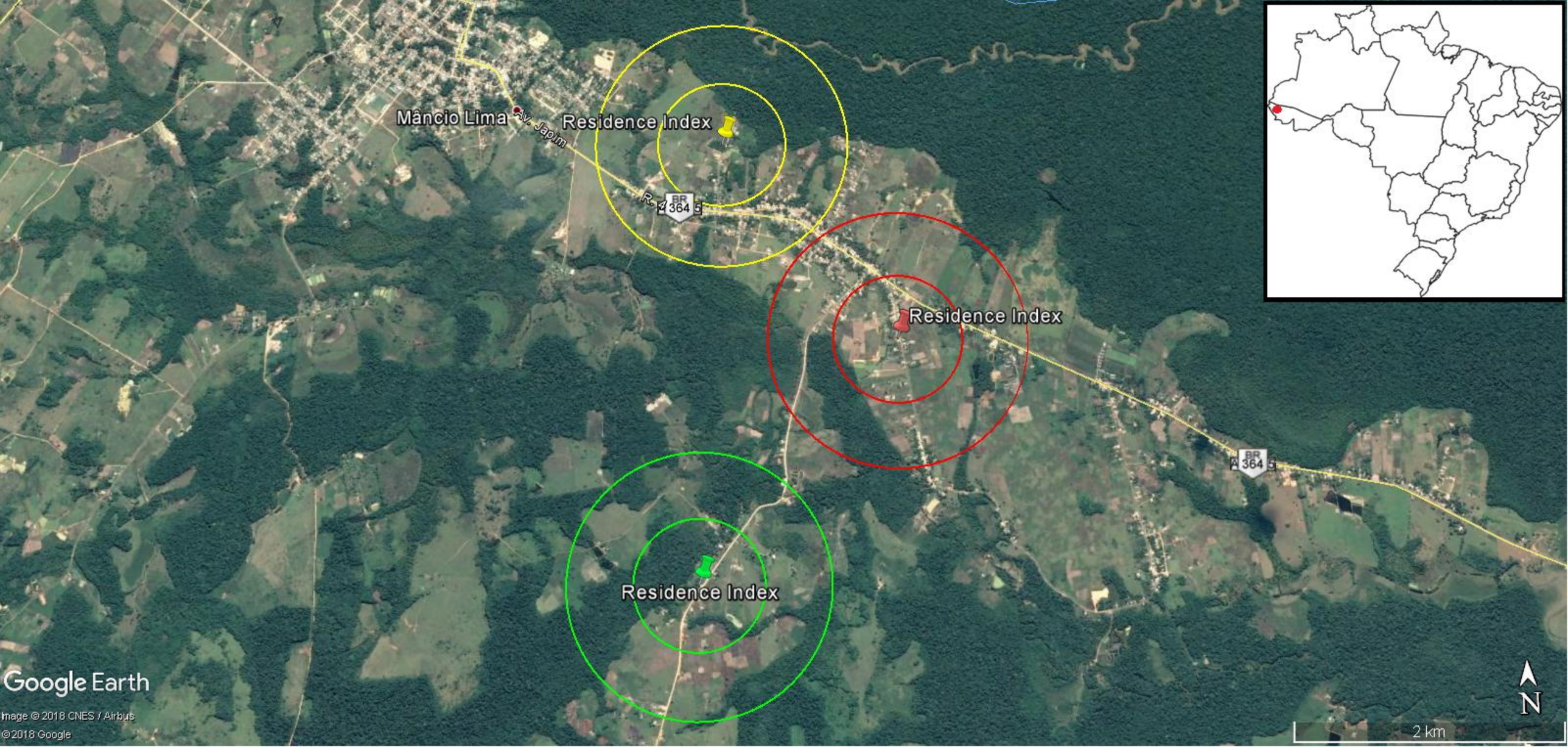
S1 Fig. Monthly distribution of reported malaria cases (*P. vivax* and *P. falciparum*) and accumulated monthly precipitation. Mancio Lima municipality 2016 and 2017. Pearson correlations: in 2016 for *P. vivax* and precipitation r = 0.75; for *P. falciparum* and precipitation r = 0.51; in 2017 for *P. vivax* and precipitation r = 0.43; for *P. falciparum* and precipitation r = 0.47. (Source: Malaria: http://www2.datasus.gov.br/DATASUS; Precipitation: clima1.cptec.inpe.br).

S1 Table. Independent variable and response options (by classification or instrument measure), level of subject analysis (sampling-point or fishpond), and sampling effort during 2017 according to the sampling schedule (6-month for Feb, Mar, Apr, May, Aug, Sept; 3-month for Feb, Mar, Apr; 3-month for May, Aug, Sept).

S2 Table. Fishpond numbers by site and collection month in Mancio Lima, Acre, Brazil 2017. Fishpond column: numbers of fishponds identified per respective month; values in brackets are total fishponds surveyed for the respective month. Dry column: fishponds with absence of water (by seasonality or owner management). Not surveyed column: fishponds unsampled (due to flooded fields or no possible access to property).

867 **Supplementary Table 3: Environment independent variables by site in Mancio Lima,**

868 **Acre, Brazil 2017.** *fishpond level. **sampling-point level.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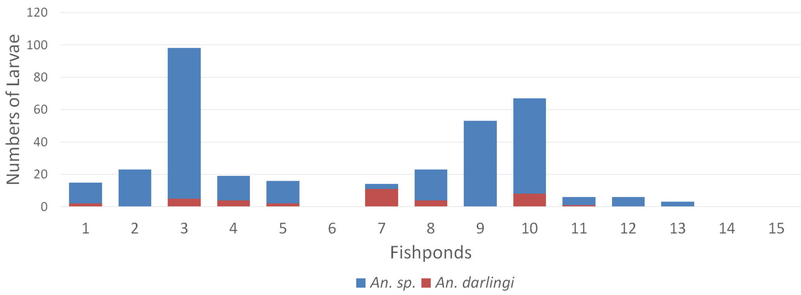
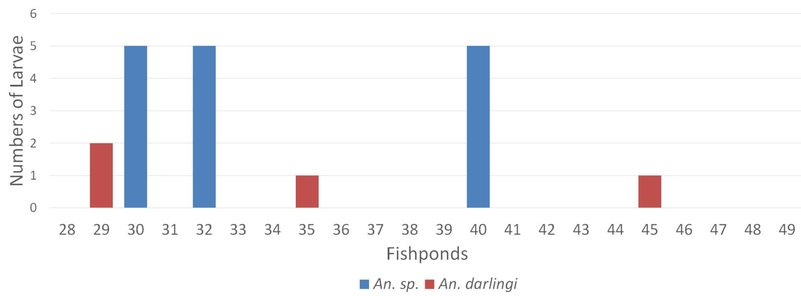
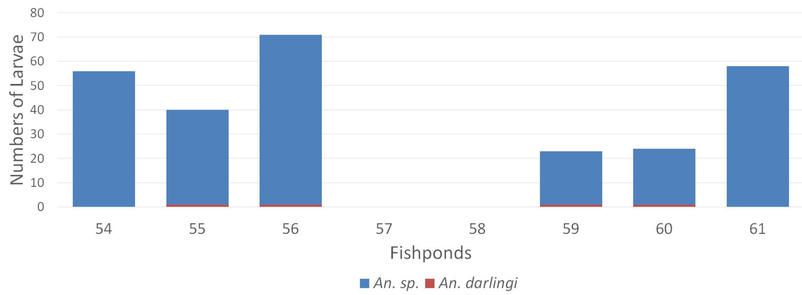
2 km

February







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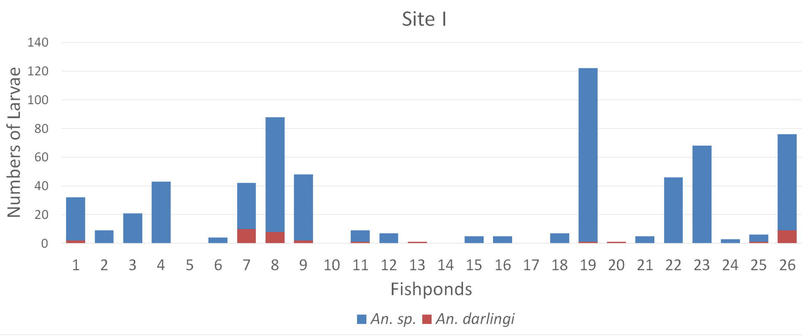
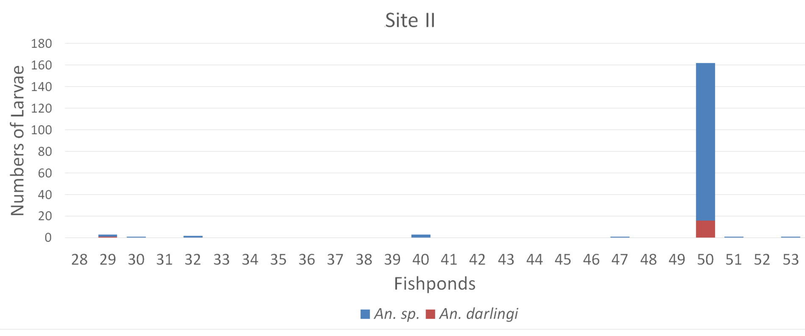
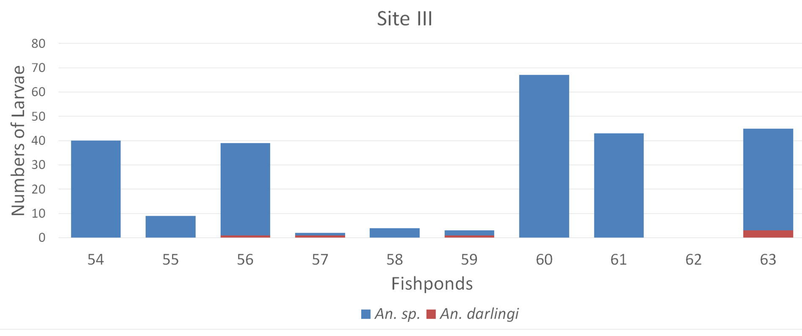
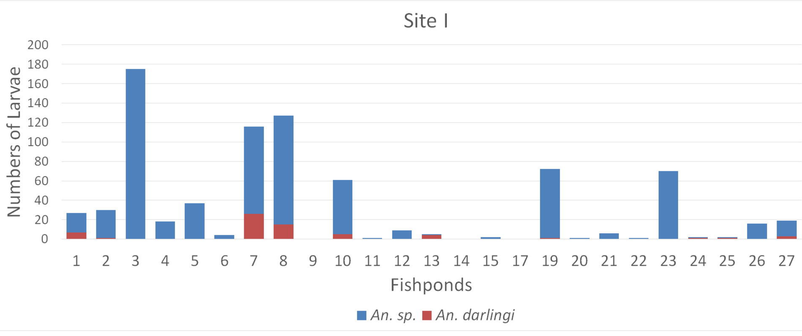
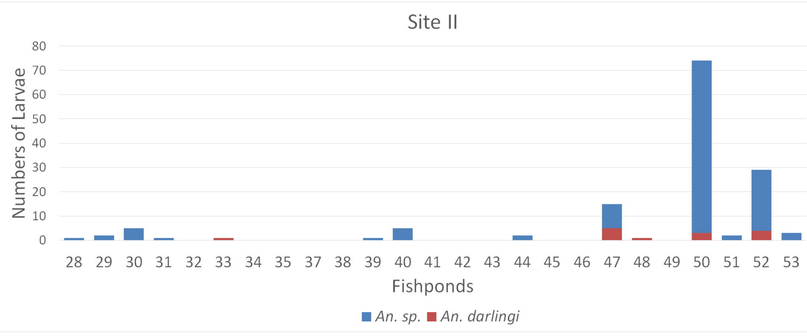


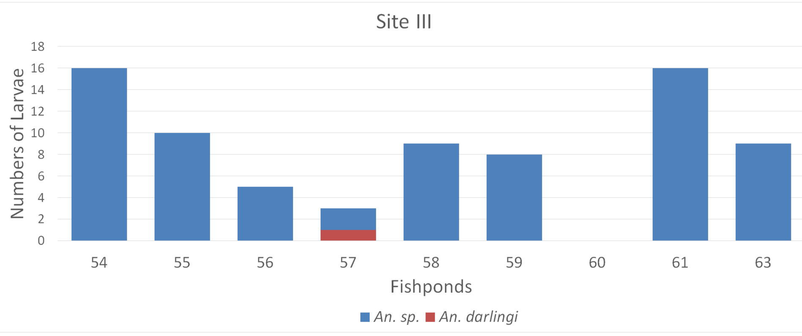




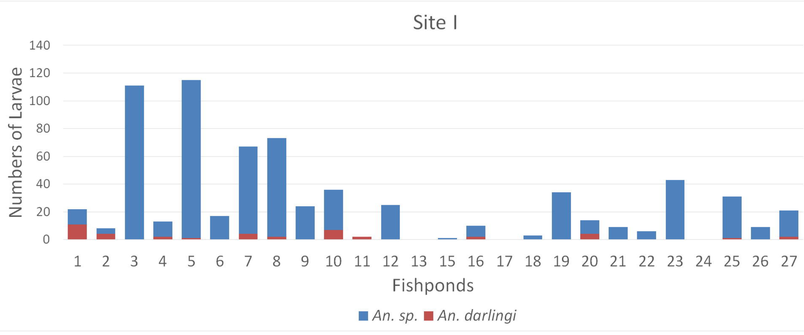
April

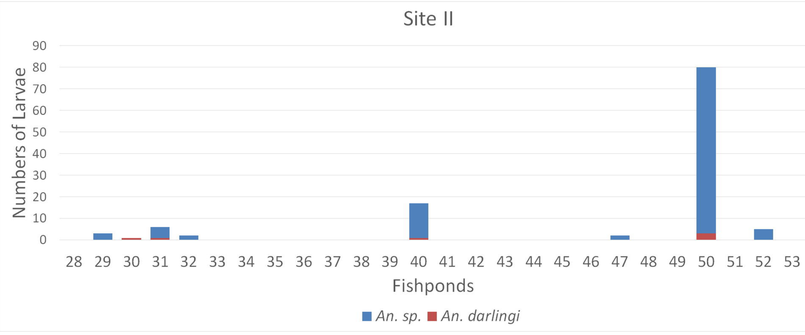


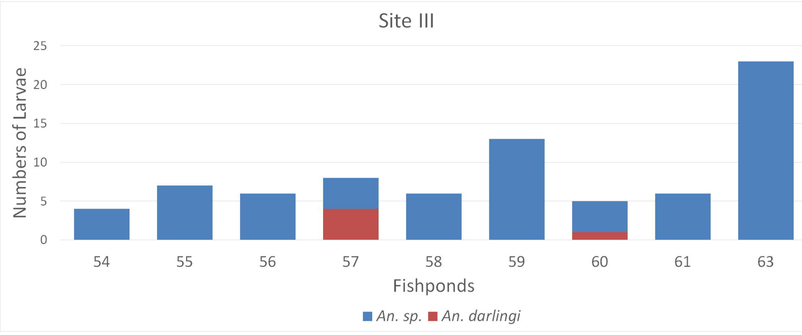




May







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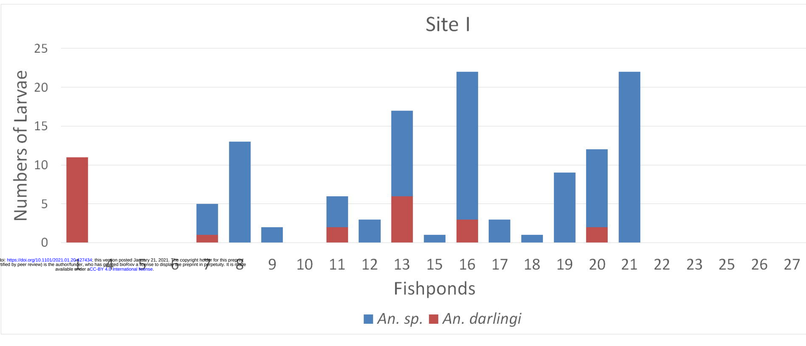

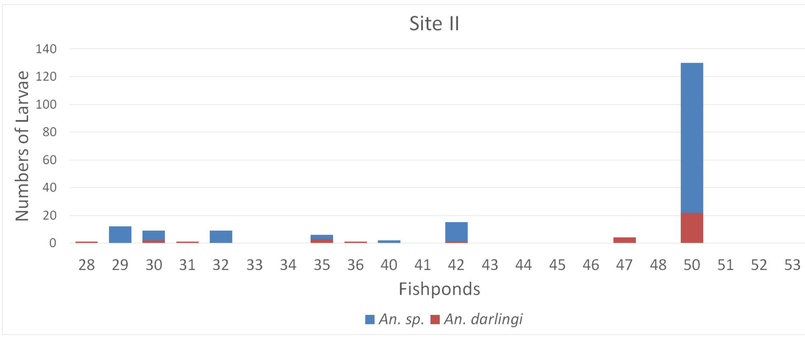

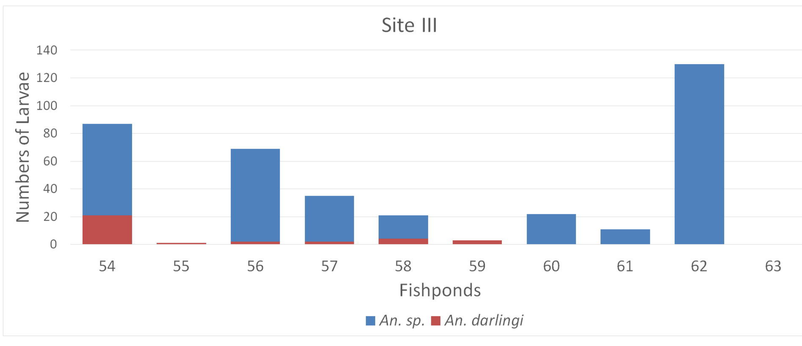

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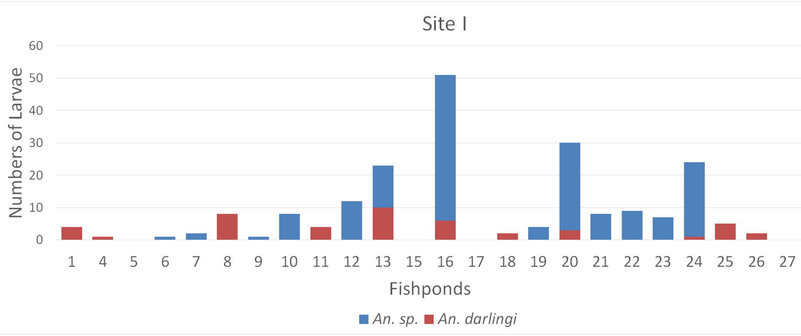
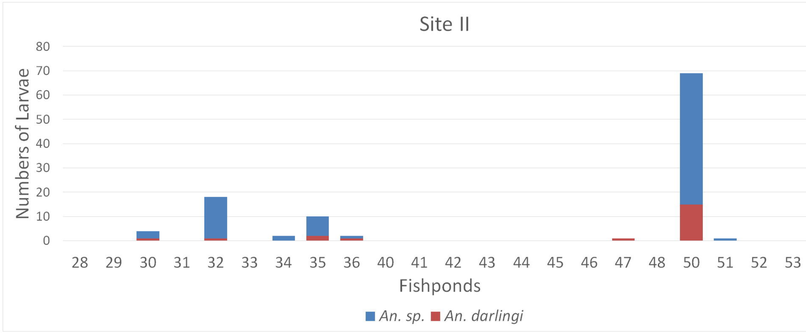
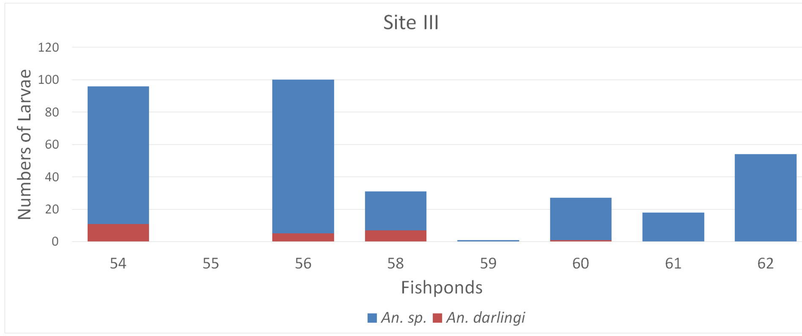
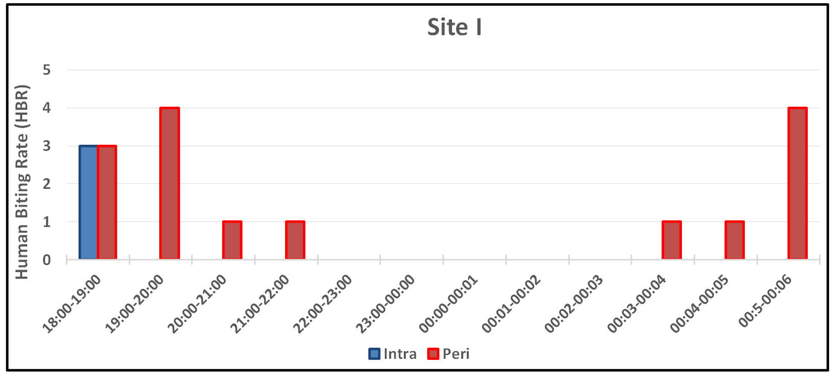
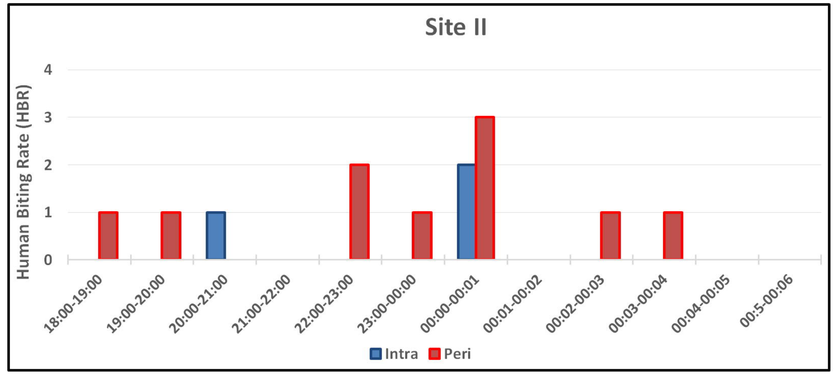


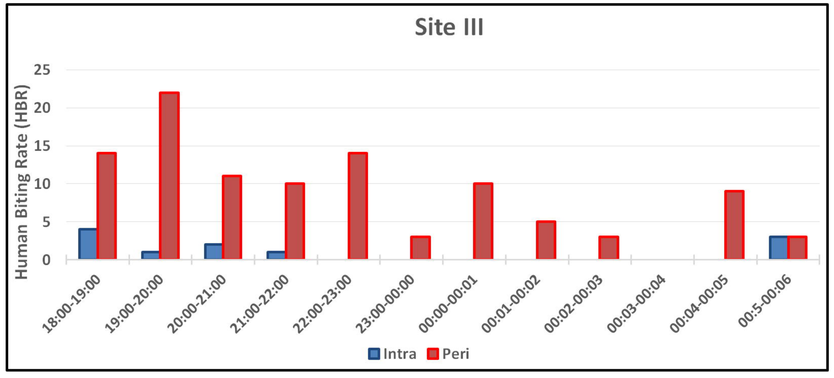




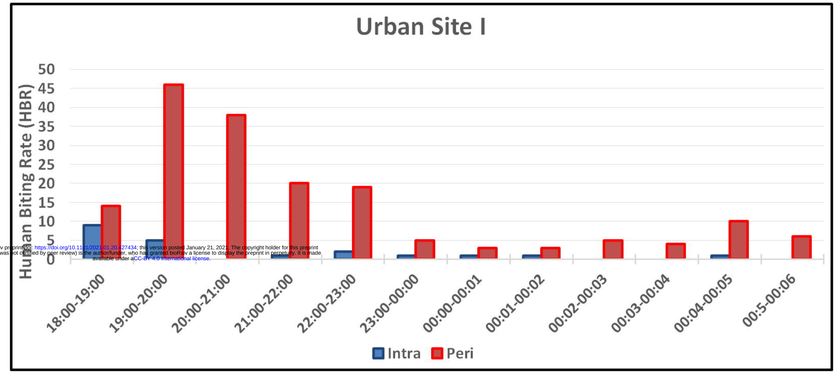
February

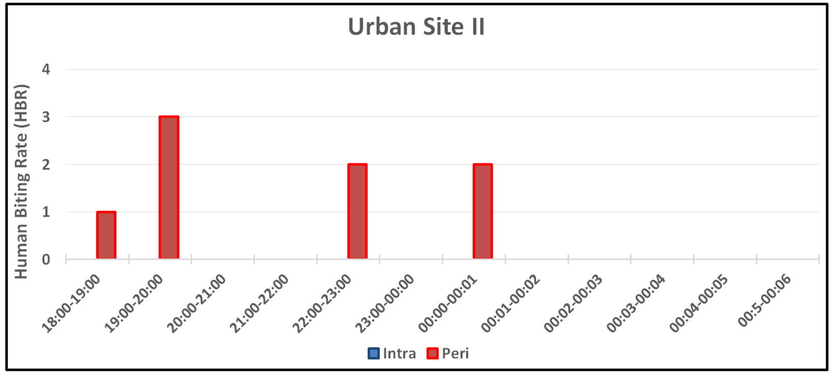


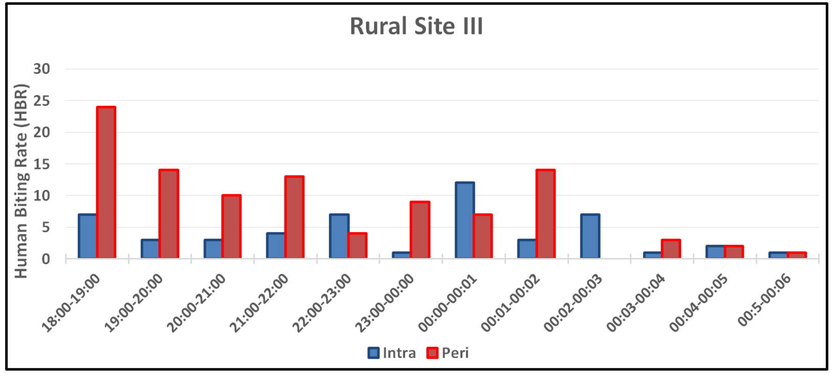




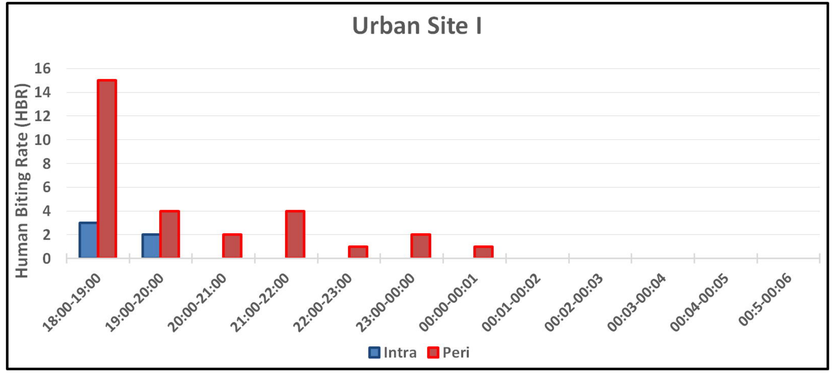
May

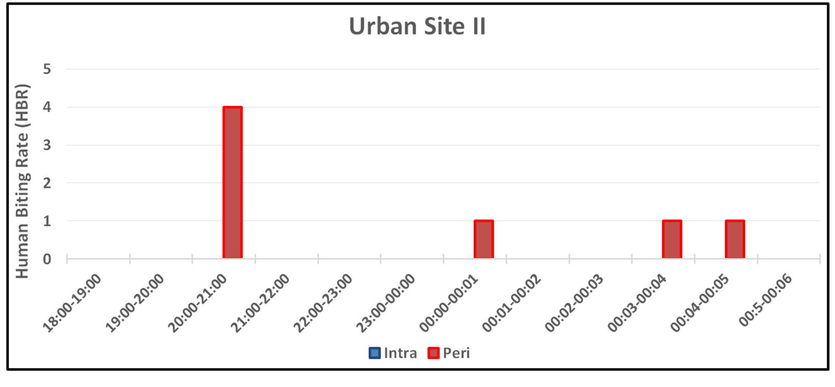


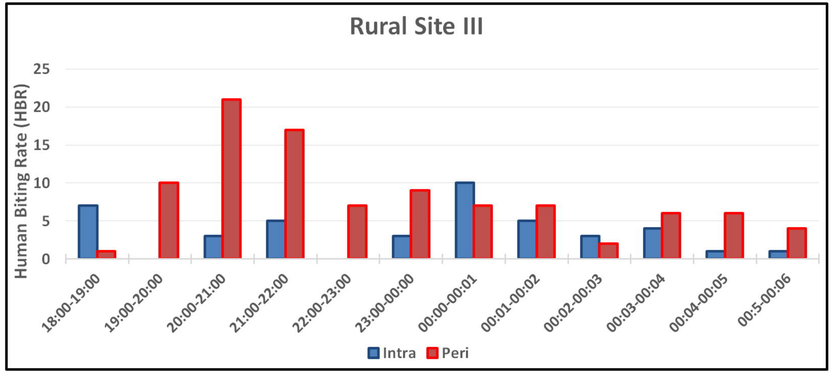


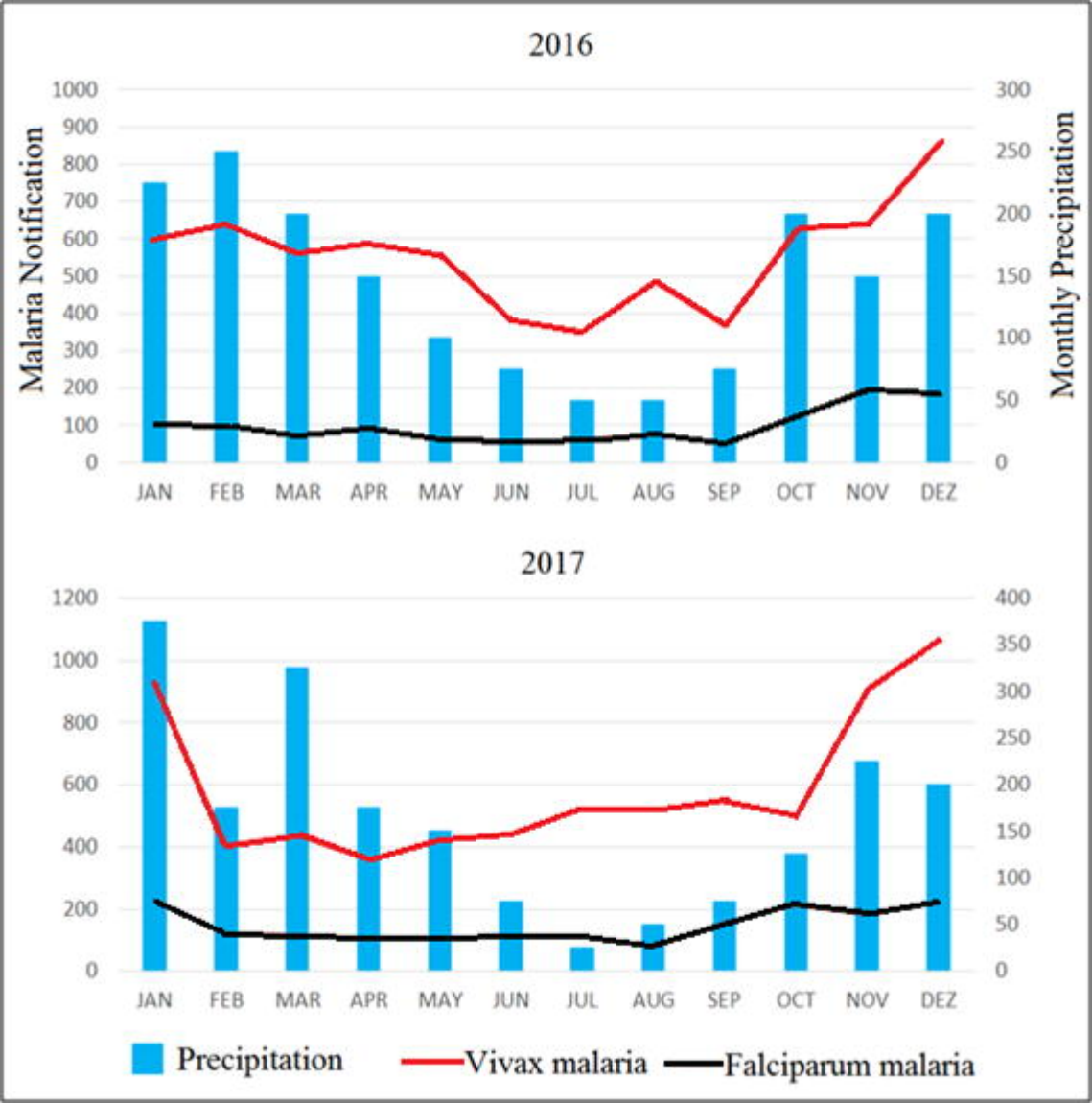


September








2016
Malaria Notification
Monthly Precipitation
1000
900
800
700
600
500
400
300
200
100
0
300
250
200
150
100
50
0
JAN
FEB
MAR
APR
MAY
JUN
JUL
AUG
SEP
OCT
NOV
DEZ
2017
1200
1000
800
600
400
200
0
400
350
300
250
200
150
100
50
0
JAN
FEB
MAR
APR
MAY
JUN
JUL
AUG
SEP
OCT
NOV
DEZ
Precipitation
Vivax malaria
Falciparum malaria