# A hierarchical Bayesian model to investigate trade-offs <sup>2</sup> between growth and reproduction in a long-lived plant

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## 21 **Abstract**

A trade-off between growth and fecundity, reflecting the inability of simultaneously investing in 22 both functions when resources are limited, is a fundamental feature of life history theory. This 23 particular trade-off is the result of evolutionary and environmental constrains shaping reproduc-24 tive and growth traits, but it remains difficult to pinpoint in natural populations of long-lived 25 plants. We developed a hierarchical Bayesian model to estimate the inter-individual correlation 26 among growth and reproduction, using observations at individual level over several years com-27 bined with resource simulations from an ecophysiological-based model (CASTANEA). In the 28 Bayesian model, the resource, simulated by CASTANEA and incorporated as a latent variable, 29 is allocated to tree growth, reproductive buds initiation and fruit maturation. Then, we used 30 individual random effects correlated among energetic sinks to investigate potential trade-offs. 31 We applied this original approach to a Mediterranean coniferous tree, Atlas Cedar (Cedrus 32 atlantica), at two contrasted levels of competition, high versus low density population. We 33 found that trees initializing many reproductive buds had a higher growth. Moreover, a negative 34 correlation was detected between growth and fruit survival during maturation. Finally, trees 35 investing more resource to maturate fruits initiated less reproductive buds. The level of com-36 petition did not impact the sign of these three correlations, but changed the level of resource 37 allocation: low density population favored growth whereas high density favored reproduction. 38 The level of resource have an impact on individual strategies. This new modeling framework 39 allowed us to detect various individual strategies of resource allocation to growth versus late-40 stage reproduction on the one hand, and to early-versus late-stage reproduction on the other 41 hand. Moreover, the sign of the correlation between growth and reproductive traits depends 42 on the stage of reproduction considered. Hence, we suggest that the investigation of potential 43 trade-offs between growth and reproduction requires to integrate the dynamics of resource and 44 sink's phenology, from initiation to maturation of reproductive organs. 45

Keywords: *Cedrus atlantica*, conifer, masting, mechanistic model, resource allocation, trade off, tree growth

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## 48 2 Introduction

Trade-offs between life-history traits are expected to be ubiquitous throughout the living world. 49 Although the concept of trade-off is used in many disciplines, we will follow here the definition 50 in evolutionary ecology proposed by Stearns (1989): "trade-offs represent the costs paid in the 51 currency of fitness when a beneficial change in one trait is linked to a detrimental change in 52 another". In this sense, a trade-off corresponds to a negative correlation between two traits 53 related to fitness, which are observed on a set of individuals, or at different life-stages for a given 54 individual. In particular, trade-offs between traits may occur when the resources, i.e. energy and 55 nutrients, are limited. In this case, trade-offs are the results of resource allocation to different 56 sinks sharing a common pool of resources. Thus, some of the trade-offs observed within a pop-57 ulation may result from the variation in allocation strategies among individuals and/or among 58 life-stages. The inter-individual variability of allocation strategies, as any other phenotypic 59 trait, is expected to be driven by a combination of environmental and genetic factors (Garland 60 and Carter, 1994). The theoretical model of van Noordwijk and de Jong (1986) demonstrated 61 that the patterns of resource acquisition versus resource allocation can change the sign of the 62 correlation between two life history traits: when resource acquisition is highly variable among 63 individuals while the fraction allocated to each life history traits is similar, the between-trait 64 correlation is positive. In the reverse case, when resource allocation is highly variable among 65 individuals, but not resource acquisition, the correlation becomes negative. To better track 66 trade-offs, one solution is to account for the variability of resources acquisition and allocation 67 between individuals together with the measurement of life history traits: this approach requires 68 combining ecophysiology with population ecology (Olijnyk and Nelson, 2013). 69

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Trade-offs between reproduction and growth-related life-history traits have received much attention (Bell, 1980; Lovett Doust, 1989), in particular in annual and perennial plants (Obeso, 2002; Thomas, 2011; Lauder *et al.*, 2019). Most studies investigated trade-offs through the phenotypic correlations between growth and reproduction at individual or population level, and reported either positive, negative or no correlations (Sánchez-Humanes *et al.*, 2011; Thomas, 2011; Wu *et al.*, 2020). Four main hypotheses related to resource allocation have been advanced

to explain such idiosyncratic patterns (Pulido et al., 2014): resource allocation to growth and 77 reproduction can be either (i) based on a hierarchy, where the resource is first allocated to repro-78 duction and then to growth (Wardlaw, 1990; Suzuki, 2001) or (ii) linked to different resource 79 pools, so that the resource allocated to growth is independent from that allocated to reproduction 80 (Cremer, 1992; Yasumura et al., 2006; Knops et al., 2007; Żywiec and Zielonka, 2013) or (iii) 81 linked to a single resource pool, with a constant fraction of resource allocated to each sink 82 (Despland and Houle, 1997; Pérez-Ramos et al., 2010; Berdanier and Clark, 2016; Lebourgeois 83 et al., 2018) or, finally, (iv) linked to a single resource pool with competition for resources 84 allocated to different sinks (Koenig and Knops, 1998; Martín et al., 2015; Lebourgeois et al., 85 2018). Only in cases (i) and (iv) the expected trade-off can be observed, while cases (ii) and (iii) 86 can lead to positive or non significant correlation. Besides resource allocation schemes, climatic 87 conditions can have contrasted direct effects on reproduction and growth and generate negative 88 environmental correlation without functional trade-off, for instance when favorable conditions 89 for growth are unfavorable for reproduction (Knops et al., 2007; Mund et al., 2020). 90

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Trade-off can also occur between early and late stages of reproduction. This is well illus-92 trated by the Quercus sp study by Knops et al. (2007), where the meaningful trade-off occur 93 between current and future reproduction. Indeed, due to the high reproductive costs in trees, 94 an important seed production in a given year can decrease the level of resource to invest in 95 future reproduction, explaining why less seeds are produced the year after (Sala et al., 2012). 96 This trade-off between current and future reproduction was confirmed by removal experiments, 97 where removing fruits increase the number reproductive buds and increase fruit production in 98 the following year (Elmqvist et al., 1991; Fox and Stevens, 1991; Fox, 1995; Santos-del Blanco 99 and Climent, 2014). 100

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Another important trade-off in monoecious plants may occur between male and female reproduction. Studies investigating the cost of reproduction usually neglect the male function and in particular the abortion of reproductive flowers. The major assumption supporting this simplification is that male function needs fewer resources than female function (Elmqvist *et al.*, 1991;

Obeso, 2002) and is not limiting for reproduction. However, this hypothesis may no longer hold 106 under changing climate (Schermer et al., 2019). In trees, studies of the trade-off between male 107 and female reproduction are rare due to the difficulty to measure male reproductive biomass. 108 But studying reproduction only from the female point of view may underestimate the initial 109 reproductive effort and lead to a wrong estimate of correlation between growth and reproductive 110 traits (Knops et al., 2007; Knops and Koenig, 2012). Hence, two hypotheses are commonly 111 used: male biomass is relatively constant across years whereas reproductive female biomass 112 follows available resource (Knops and Koenig, 2012), or both male and female biomass co-vary 113 through time (Houle, 2001). 114

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The aim of this study is to detect the trade-off between growth and reproduction at different 116 stages of the reproduction cycle and considering different functions (male and female) within a 117 tree population. To tackle this objective, we developed a new and original methodological ap-118 proach by combining two different models for resource acquisition and allocation, respectively. 119 On the one hand, we used an ecophysiological model to simulate the acquisition of resource 120 at the individual tree level (Net Primary Production, NPP). This ecophysiological model simu-121 lates climate effects through several processes such as tree photosynthesis, respiration and soil 122 evaporation. On the other hand, we developed a hierarchical Bayesian model based on resource 123 allocation processes, which simultaneously accounts for (i) the timing of growth and reproduc-124 tive processes, and (ii) the variation among individuals and years in resource allocation and in 125 the phenotypic gender, both considered as latent variables. This approach allows to estimate 126 the inter-individual correlation between growth and reproduction while accounting for resource 127 heterogeneity between trees. 128

## **3** Material and Methods

#### **3.1** Hierarchical Bayesian model

<sup>131</sup> We developed a Bayesian hierarchical model linking resource to three energetic sinks: growth, <sup>132</sup> reproductive buds initiation and female cones survival. Potential correlations between these

sinks are accounted for by correlated random effects. The model is composed of two layers (see 133 Fig. 1 for a general overview): the process model which describes how resource is allocated 134 to each sink, and the data model which describes the links between the process model and 135 the observations. The observations consist in repeated counts of male and female organs to 136 characterize the reproductive sink and repeated measure of size increment to characterize the 137 growth sink. Description of model implementation and procedure is present in Appendix S1: 138 Section S2. We presented prior parameters in Table S1, accuracy of the model in Table S2 and 139 posterior distribution in Table S3. 140

#### 141 **3.1.1 Process model**

**Resource allocation to energetic sinks:** The resource  $(NPP_{i,t})$  for individual *i* at year *t* deter-142 mines its growth  $(BAI_{i,t})$ , the number of initiated reproductive buds  $(IB_{i,t})$  and the probability of 143 female cones survival  $(p_{i,t}^{FCS})$ . The use of correlated random effects (referred as  $\epsilon_{x,i}$ ) allowed us 144 to investigate potential trade-offs among these three energetic sinks. We tested two alternative 145 models of inter-individual variation for each growth and reproduction trait to identify two po-146 tential types of phenotypic trade-off, either driven by NPP resource or not, respectively named 147 "model 1" and "model 2". In model 1, inter-individual variation acts on the capacity to valorize 148 the amount of available resource (NPP), i.e. we introduced a random individual effect on the 149 growth or reproduction trait response to NPP. In model 2, inter-individual variation is directly on 150 the trait, i.e. we introduced a random effect on the intercept of the model. With this approach, 151 there is no constraints on the triplet  $\{\gamma; \beta_1; \beta_2\}$ . Instead, the correlated individual random effect 152 used here  $(\epsilon_{x,i})$  allow more flexibility, and possible synergies or antagonisms between sinks to 153 emerge from the model. To determine which one of the two model is the best, we included 154 the variable  $Y \sim \text{Bernoulli}(p_Y)$  indicating whether NPP was the only resource responsible of 155 inter-individual correlation (Y = 0) or not (Y = 1): 156

$$BAI_{i,t} = (\gamma_d + \underbrace{Y * \epsilon_{1,i}}_{\text{model 1}}) * NPP_{i,t} + \underbrace{(1 - Y) * \epsilon_{1,i}}_{\text{model 2}}$$

$$IB_{i,t} = X_{i,t} * ((\beta_{1,d} + \underbrace{Y * \epsilon_{2,i}}_{\text{model 1}}) * NPP_{i,t} + \underbrace{(1 - Y) * \epsilon_{2,i}}_{\text{model 2}})$$

$$(1)$$

$$\log_{i,t}(p_{i,t}^{FCS}) = \beta_0 + (\beta_{2,d} + \underbrace{Y * \epsilon_{3,i}}_{\text{model 1}}) * NPP_{i,t} + \underbrace{(1 - Y) * \epsilon_{3,i}}_{\text{model 2}}$$

$$\epsilon \sim \mathcal{N}_3 (0, \Sigma)$$

where  $\gamma_d$  is the slope parameter that depends on the density d. In the model 1,  $\epsilon_{1,i}$  is an individual 157 random effect associated to the slope parameter whereas in the model 2, the individual random 158 effect does not constrains  $NPP_{i,t}$  on growth increment  $(BAI_{i,t})$ . Then,  $\beta_{1,d}$  is the slope parameter 159 than depends on density d and  $\epsilon_{2,i}$  constrains the slope parameter (model 1) or is individual 160 random effect for the effect of  $NPP_{i,t}$  on bud initiation ( $IB_{i,t}$ ) (model 2). Parameter  $\beta_{2,d}$  is the 161 slope depending of density d and  $\epsilon_{3,i}$  is associated to the slope parameter (model 1) or is the 162 individual random effect for the effect of  $NPP_{i,t}$  on female cones survival  $(p_{i,t}^{FCS})$  (model 2).  $X_{i,t}$ 163 is a Bernoulli variable  $(X_{i,t} \sim \mathcal{B}(p_X))$ , indicating if individual *i* produced  $(X_{i,t} = 1)$  or not  $(X_{i,t})$ 164 = 0) reproductive buds at year t, allowing to consider null values in the observations, because 165 some trees never produced cones (see Fig 3). The intercept  $\beta_0$  is fixed to constrain  $p_{it}^{FCS} \approx 0$ 166 when  $NPP_{i,t} = 0$ . Note that the model does not account for other possible drivers of female 167 cone survival such as pollen limitation. 168

This model explicitly considers that individual random effect on the three sinks,  $\{\epsilon_{1,i}; \epsilon_{2,i}; \epsilon_{3,i}\}$ , are related to each other through the variance-covariance matrix  $\Sigma$ . Their pairwise correlations can thus be used to investigate constrains in resource allocation to the three sinks. Indeed, for a given amount of resource available for an individual *i*,  $NPP_{i,t}$ , the sign of the correlation between  $\epsilon_{l,i}$ , and  $\epsilon_{k,i}$  indicates how much resource is respectively allocated to sinks *l* and *k*. Correlations were computed as  $\rho_{l,k} = \sum_{l,k} / (\sum_{l,l} \sum_{k,k})$ .

<sup>175</sup> **Male and female reproduction:** The model considers that initialized buds  $(IB_{i,t})$  develop <sup>176</sup> into a number of initiated male cones  $(IMC_{i,t})$  and a number of initiated female cones  $(IFC_{i,t})$ <sup>177</sup> according to the phenotypic gender  $(PG_{i,t})$  of individual *i* at year *t*, using the following model:

$$logit (PG_{i,t}) \sim \mathcal{N} (\bar{PG}, \sigma_{PG})$$

$$IMC_{i,t} = PG_{i,t} * IB_{i,t}$$

$$IFC_{i,t} = (1 - IMC_{i,t})$$
(2)

Here, phenotypic gender thus correspond to maleness, i.e. the ratio of male vs male and female
initiated cones.

#### 180 **3.1.2 Data model**

The model uses repeated observations of growth, male reproduction and female reproduction, such as presented in our case in the section 3.3. We assume that observed growth  $(BAI_{i,t}^{ODS})$  is related to the latent growth variable of the process model  $(BAI_{i,t})$  through :

$$BAI_{it}^{ODS} \sim \mathcal{N}\left(BAI_{i,t}, \sigma_{BAI}\right)$$
 (3)

The number of initiated male cones  $(IMC_{i,t})$  is a continuous variable while the observed abundance of male cones  $(IMC_{i,t}^{obs})$  is a categorical ordered variable as described in section 3.3. To link  $IMC_{i,t}$  and  $IMC_{i,t}^{obs}$  we used the following observational model:

$$\begin{cases} \pi_{i,t} = [F(s_0), F(s_1) - F(s_0), F(s_2) - F(s_1), F(s_3) - F(s_2), 1 - F(s_3)] \\ IMC_{i,t}^{\text{obs}} \sim \text{Multinomial}(\pi_{i,t}, 1) \end{cases}$$
(4)

where F(.) denotes the cumulative distribution function of a normal distribution with mean *IMC<sub>i,t</sub>* and variance  $\sigma_{IMC}$ . { $s_0, s_1, s_2, s_3$ } is a set of fixed thresholds determining the boundaries between each value of the notation, and are derived from F(.). Note that this approach is equivalent to consider a probit link in the case of binary data.

<sup>191</sup> Finally, the observed count of female cones  $(FC_{i,t}^{obs})$  is linked to the latent variable describing <sup>192</sup> number of current bud initiated  $(p_{i,t}^{FCS})$  and previous year reproduction (bud initiation,  $IB_{i,t-1}$ , <sup>193</sup> female cones initiation,  $IFC_{i,t-1}$ ) through the following Poisson observational model:

$$FC_{i,t}^{\text{obs}} \sim \mathcal{P}(p_{i,t}^{FCS} * IFC_{i,t-1} * IB_{i,t-1})$$

$$\tag{5}$$

#### **3.2** Study species and site

We applied this model to Atlas cedar, Cedrus atlantica (Manetti ex Endl.) Carrière, a Mediter-195 ranean coniferous tree species. This monoecious species reaches it sexual maturity between 196 15 and 30 years old (Toth, 1978), and carries male and female organs irregularly dispersed 197 on the crown of the tree. Male reproduction, from reproductive bud initiation to pollen mat-198 uration, is achieved within one year, whereas female reproduction, from initiation to cone 199 maturation, spreads over two years (Fig. 2). Reproductive buds are initiated during summer 200 (June-July for male buds and late August for female buds) at year t-1, followed by pollination 201 in September, when female cones open and receive pollen. Female cones close their scales in 202 October-November, when male cones fall down. A time lag between pollination and fecunda-203 tion characterizes coniferous species (Williams, 2009), with a duration of nine month for C. 204 atlantica. Pollen germinates in spring at year t, then ovules are fertilized and seeds begin to 205 maturate until autumn of the same year, and mature seeds are dispersed in year t+1 (Toth, 1978). 206 Hence, two generations of female cones may co-exist on a tree and it is possible to distinguish 207 green one-year cones from brownish two-years cones. Cedrus atlantica can be referred as a 208 masting species, characterized by a seed production highly variable among years and synchro-209 nized among individuals (Kelly and Sork, 2002; Krouchi et al., 2004). 210

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The study site is a 35 years-old experimental plantation located in Mont-Ventoux in France, a Mediterranean mountain, at 1170 meters of elevation (44°07' 05" N, 5°20' 38" E). All the trees were planted in similar pedo-climatic conditions. The initial tree density was 2700 stems ha<sup>-1</sup>. In this experiment, two thinning strategies had been applied leading to contrasted competitor densities at the time of observations: high (1200 stems ha<sup>-1</sup>) versus low density (250 stems ha<sup>-1</sup>). More details on the silvicuture experiment and tree growth response are available in Guillemot *et al.* (2015).

#### **3.3** Observations of growth and reproduction

We monitored 40 individual trees in the high-density stand and 31 individual trees in the lowdensity stand. These individuals were randomly sampled within each stand and measured each

year from 2002 to 2005 (except for growth, with a longer dataset from 1989 to 2015). We first 222 measured the diameter at 1.3 meters  $(DBH_{i,t})$  for each individual tree *i* and each year *t*. Annual 223 basal area increment  $(BAI_{i,t}^{\text{obs}})$  was computed as:  $BAI_{i,t}^{\text{obs}} = (\pi \times DBH_t^2/4) - (\pi \times DBH_{t-1}^2/4)$ 224 Male cones abundance  $(M_{i,t})$  was recorded as a qualitative ordered variable, consisting in a 225 score ranging from 0 to 4: "0" means no male cone is observed; "1" means few male cones are 226 dispersed in the canopy; "2" means male cones are abundant in one branch; "3" means male 227 cones are abundant on two branches and "4" means male cones are abundant over the whole tree 228 canopy. These scores were converted into multinomial observations  $IMC_{it}^{obs}$  as follows: 229

$$\begin{cases}
M_{i,t} = 0 \implies IMC_{i,t}^{\text{obs}} = [1, 0, 0, 0, 0] \\
M_{i,t} = 1 \implies IMC_{i,t}^{\text{obs}} = [0, 1, 0, 0, 0] \\
M_{i,t} = 2 \implies IMC_{i,t}^{\text{obs}} = [0, 0, 1, 0, 0] \\
M_{i,t} = 3 \implies IMC_{i,t}^{\text{obs}} = [0, 0, 0, 1, 0] \\
M_{i,t} = 4 \implies IMC_{i,t}^{\text{obs}} = [0, 0, 0, 0, 1]
\end{cases}$$
(6)

Female cones  $(FC_{i,t}^{obs})$  were individually counted over the whole canopy from the ground with binoculars and at two seasons each year (in spring then in late summer) to discriminate one-year versus two-year female cones based on their color. The identification and distinct count of one-year versus two-years cones allowed us to determine the number of cones produced each year. The raw relationship between  $FC_{i,t}^{obs}$  and  $BAI_{i,t}^{obs}$  in Fig. 3 shows a positive correlation between growth and female reproduction and individual variation.

#### 236 3.4 Resource simulation using CASTANEA

We used the ecophysiological model CASTANEA to estimate the carbon resource available each year for each tree. CASTANEA aims to simulate carbon and water fluxes of monospecific forest ecosystems (Dufrêne *et al.*, 2005). The model simulates radiation transfer, photosynthesis, autotrophic respiration, carbon allocation to different tree compartments, evapotranspiration and water balance. A complete description of the model is presented in Dufrêne *et al.* (2005) with subsequent modifications described in Davi *et al.* (2009) and in Davi and Cailleret (2017). The model input data are daily climate data, soil characteristics (texture, depth and stone content), and <sup>244</sup> initial tree characteristics (height, diameter and age). We calibrated and validated the model for <sup>245</sup> *C. atlantica* on our study site for both stand densities. The values of CASTANEA parameters, <sup>246</sup> simulations design, and validation procedure are described in Appendix S1: Section S1. In <sup>247</sup> the following hierarchical Bayesian model, resource was modeled for each individual through <sup>248</sup> the Net Primary Productivity (*NPP<sub>i,t</sub>*) simulated from 1999 to 2006, and corresponding to the <sup>249</sup> difference between gross primary productivity and autotrophic respiration.

## **4 Results**

#### **4.1** Resource and phenotypic gender differences between densities

The model fitted accurately male fecundity (Brier score closed to 0). We obtained a lower 252 accuracy for both female cones production (Bayesian p-value = 0.84) and tree growth (Bayesian 253 p-value = 0.18). The posterior predictive checks of the hierarchical Bayesian model are pre-254 sented in Appendix S1: Table S2. All trees had a high probability to be reproductive in both 255 plots ( $p_X$ , posterior median [CI95%] = 0.84 [0.79; 0.88]). Remaining posterior parameters are 256 described in the following part and in Appendix S1: Table S3. The resource NPP<sub>i,t</sub> simulated 257 with CASTANEA model varied between years (due to climate) and between individuals (due to 258 variations in tree size and stand density; Appendix S1: section S1).  $NPP_{i,t}$  had a mean value 259 of 11990 gC tree<sup>-1</sup> year<sup>-1</sup> (sd = 3030 gC tree<sup>-1</sup> year<sup>-1</sup>) and 8751 gC tree<sup>-1</sup> year<sup>-1</sup> (sd = 2335 260 gC tree<sup>-1</sup> year<sup>-1</sup>), respectively for the low-density and high-density plot (Fig. 4 a). 261

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The phenotypic gender did not differ between the two density stands, with a mean maleness equal to 0.29 (sd = 0.04) for the high-density stand and to 0.28 (sd = 0.06) for the low-density stand (Fig. 4 b).

#### **4.2** Correlation between individual random effects

<sup>267</sup> We found that Y was equal to 0 for all iterations, meaning that the best model to infer trade-offs <sup>268</sup> in resource allocation is the model 2, where the  $\epsilon_{x,i}$  are not linked to NPP. We reported the <sup>269</sup> posterior distribution of correlations in Fig. 5 (a), (b) and (c) and pairwise covariations between

individual random effects  $\epsilon_{1,i}$ ;  $\epsilon_{2,i}$ ;  $\epsilon_{3,i}$  in Fig. 5 (d), (e) and (f).

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The estimated correlation between inter-individual variations in growth  $(\epsilon_{1,i})$  and reproduc-272 tive buds initiation ( $\epsilon_{2,i}$ ) was positive ( $\rho_{1,2} = 0.54[0.32, 0.70]$  with  $pr[\rho_{1,2} > 0] = 1$ , Fig. 5 273 a,d). Secondly, the estimated correlation between inter-individual variations in growth  $(\epsilon_{1,i})$ 274 and female cones survival ( $\epsilon_{3,i}$ ) was significantly negative ( $\rho_{1,3} = -0.36[-0.65, 0.24]$  with 275  $pr[\rho_{1,3} < 0] = 0.91$ , Fig. 5 b,e) indicating a trade-off. Lastly, the estimated correlation be-276 tween inter-individual variations in reproductive buds initiation ( $\epsilon_{2,i}$ ) and female cones survival 277  $(\epsilon_{3,i})$  was negative  $(\rho_{2,3} = -0.85[-0.96, -0.19])$ , with  $pr[\rho_{2,3} < 0] = 0.98)$ , indicating a clear 278 phenotypic trade-off between bud initiation and female cone survival (Fig. 5 c,f). These three 279 correlations mainly consist in between-density correlations and within-density correlation in the 280 high density stand (low resources), but only one within-density correlation is found in the low 281 density stand (high level of resources) between female cone survival and bud initiation. 282

## **4.3** Effects of stand density and phenotypic gender on individual allocation

#### 284 strategies

We found a higher growth response to available resource in the low density stand than in the high 285 density stand  $(pr[\gamma_{1200} > \gamma_{250}] = 0.25)$ . However, we found a higher reproductive response, 286 for initiation and maturation, to available resource in the high density than in the low density 287  $(pr[\beta_{1,1200} > \beta_{1,250}] = 0.86 \text{ and } pr[\beta_{2,1200} > \beta_{2,250}] = 0.94)$ . The ratio of resources allocated 288 to growth was higher in the low density stand, whereas the ratio of resources allocated to 289 reproduction was higher in the high density stand. Lastly, individual allocation strategies ( $\epsilon_{x,i}$ ) 290 were not affected by the phenotypic gender. Probabilities to obtain a significant correlation 291 P < 0.05 between  $\epsilon_{1,i}$  and phenotypic gender (PG) was equal to 0.25, to 0.24 with  $\epsilon_{2,i}$  and to 292 0.16 with  $\epsilon_{3,i}$ . 293

## <sup>294</sup> 5 Discussion

This study revealed significant phenotypic correlations between individual tree growth and re-295 productive traits using an original methodology combining an ecophysiological model which 296 simulates the yearly acquisition of resource at tree level, with a hierarchical Bayesian model 297 which specifies how the resource is allocated to growth and reproduction throughout the pheno-298 logical cycle. Although similar Bayesian models have been previously developed (Buoro *et al.*, 299 2010), this is the first time to our best knowledge that such a combined approach is used in a 300 tree species. Considering different reproductive stage (i.e. initiation or maturation) and allowing 301 alternatively for a driving role of NPP on the allocation scheme or not, our results shed light on 302 the contradictory findings reported in the literature about the sign and biological meaning of the 303 correlations between growth and reproduction (Knops et al., 2007). 304

# 5.1 Origins of the phenotypic correlations between growth and reproduc tion components

We detected a net increase between cone production and NPP, because trees in lower density, 307 having more NPP tends to growth and reproduce more (Appendix S1: Figure S2). We also found 308 a positive correlation between growth and initiation of reproductive organs: trees which invested 309 more in reproduction had also a higher growth. This result was obtained while our model does 310 not specify any competition between sinks in case of limiting resources. Resources appeared 311 to be invested proportionally in bud initiation and growth with a constant fraction devoted for 312 reproduction, which is referred as the "resource matching" hypothesis. Under this hypothesis, 313 trees both grow and reproduce more when more resources are available (Bogdziewicz et al., 314 2020). 315

Then, we found a negative correlation between growth and cone survival, suggesting that these functions of growth and cone maturation compete with each other during resource allocation. Similarly, a recent study in *Fagus sylvatica* demonstrated that reproduction drives the inter-annual variability of growth and that years with high reproductive effort were also years with reduced growth (Hacket-Pain *et al.*, 2018). However we cannot rule out that other processes <sup>321</sup> such as pollen limitation contribute to the negative correlation observed here between growth <sup>322</sup> and reproduction, as showed by Knops *et al.* (2007). Indeed, in wind pollinated species, springs <sup>323</sup> with high level of precipitation usually favor growth, but can reduce the efficiency of pollen <sup>324</sup> release and fertilization. In these cases, negative correlation between growth and reproduction <sup>325</sup> can emerge from contrasted effects of meteorological conditions, without any direct link between <sup>326</sup> growth and reproduction. But we demonstrated here that resources could also contribute to this <sup>327</sup> trade-off between growth and fruit maturation and it is not only dependent to pollination.

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Finally, we also found a trade-off between buds initiation and female cones survival: trees investing more resource to maturate cones initiated less reproductive buds. This result is consistent with previously published experimental data, in particular removal experiments which demonstrated a trade-off between current and future reproduction (for instance in *Pinus halepensis*, Santos-del Blanco and Climent 2014).

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The three observed correlations differ according the stand densities (Fig. 5). Trees in low 335 density tend to allocate, proportionally, more resources to growth than in high density stand, 336 while trees in high density tend to allocate in general more resources to bud initiation and fruit 337 maturation than in low density stand. When resources are scarce (i.e. high density stand), trees 338 are expected to exhibit a stronger trade-off and to invest most of their resources in reproduction 339 whereas when resources are abundant, trees are expected to invest it preferentially for growth 340 (Lauder et al., 2019). Furthermore, correlations between growth and reproductive traits were 341 only present in the high density stand, except for the last correlation between initiation and cone 342 maturation, found in both densities. Our study also revealed that NPP is not the only driver to 343 the different correlations. Additional drivers might be important too (such as reserves, nitrogen, 344 or other sources of energy). 345

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Finally, we found that *C. atlantica* trees were overall more females both at high and low competitor density, and that the phenotypic gender was relatively constant across years (expect in 2004 with a lower median value for low density plot, Fig 4).

14

#### **5.2** Evolutionary consequences of the observed trade-offs

In the studied system, C. atlantica trees displayed different strategies of resource allocation: 351 some individuals allocated resource preferentially to growth and others to cone survival; some 352 individuals allocated resource preferentially to cone survival and other to reproductive buds 353 initiation. These strategies were observed in an even-aged plantation where the studied indi-354 viduals had the same age, experiment similar climatic conditions and also similar density of 355 competitor within stand. As trade-offs were identified through correlations between individual 356 random effect, they integrate the genetic and plastic inter-individual variation of traits involved 357 in growth and reproduction. The observed trade-off cannot therefore be entirely attributed to 358 genetic factors although we expect the plastic component of trait variation to be minimized in 359 this plantation as compared to a natural population. 360

361

This approach could be used in natural populations, using a similar model if any proxy 362 of spatial resource heterogeneity were available as a covariate, or a simplified model if not. 363 Therefore, we expect that the heterogeneity of density in a natural population would also result 364 in the coexistence of multiple allocation strategies, but along a wider range of trait values. 365 Moreover, the heterogeneity in tree density should not only affect the amount of available 366 resource for individuals, but also change the conditions of evolution. Finally, our study also 367 suggests that different individual strategies are expected at the core versus the margins of natural 368 population range (Sexton et al., 2009). Indeed, we showed that trees at low-density grew 369 more than individuals at high-density, which affected the slope (although not the sign) of the 370 correlation between growth and reproduction. But slopes of higher values where found for 371 low density stand. Such differences in resources acquisition could also occur across time and 372 ontogenic stages (Thomas, 2011; Barringer et al., 2013) and may affect the capacity of migration 373 and adaptation (Aitken et al., 2008). 374

#### **5.3** A new approach for modelling resource acquisition and allocation

The main novelty of our approach is to combine the ecophysiological model CASTANEA to simulate resource acquisition with a hierarchical Bayesian model to account explicitly for re-

source allocation through latent variables (e.g.  $IB_{i,t}$ ,  $PG_{i,t}$ ). CASTANEA allowed us to obtain 378 and validate proxys of the available resource, through the carbon Net Primary Production (NPP). 379 The model correctly reproduced the variation in the level of resource between plots at differ-380 ent competitor densities. Higher NPP per individual was simulated for the low density stand, 381 consistent with reduced competition. This positive effect of reduced competition on resource 382 acquisition at tree level was also observed with tree growth data (Guillemot et al., 2015) and 383 simulated with another version of the CASTANEA model (Guillemot et al., 2014). Based on 384 tree ecophysiology, climate and soil conditions, variations in the levels of resource were also 385 simulated among years. In particular, lower resources were simulated during the year 2003 386 due to higher stress. Drought stress is known to directly impact ecosystem productivity (Ciais 387 et al., 2005) and its duration determines the growth of Mediterranean trees (Linares et al., 2013; 388 Lempereur et al., 2015). 389

390

A second advantage of our approach is the use of latent variables in the Bayesian model, 391 which allowed us to explicitly model resource allocation (including the phenotypic gender) over 392 the two-years reproductive cycle. This Bayesian model revealed the existence of trade-offs that 393 could not observed based on raw data (Fig. 3), if we look at cone production and tree growth. 394 Many perennial species have a similar reproductive cycle over two years, with bud initiation 395 in the first year, followed by maturation in the second year. The hierarchical Bayesian model 396 developed in this study could thus be applied to other species, using direct measurements of 397 resource proxys (e.g. photosynthesis), or resources estimates obtained from an ecophysiological 398 model (for example the CASTANEA is currently calibrated for twelve European trees species). 399 Our model can also be used in species with a shorter reproductive cycle, where female fruit 400 mortality would occur the same year of it initiation. 401

402

Another advantage is the use of two different resource allocation scheme in the same equation (i.e. use of *Y*) to identify correlation. It was possible to test if individual random effects were strictly based on NPP resource or other kinds of resources or processes. We found that for all iterations that model 2, with individual random effects outside the slope, was the best, meaning

that allocation to growth and reproduction are the result of not only NPP. Deeper exploration is
 needed to investigate what are the main limiting energies responsible of these trade-offs.

## **5.4** Towards a synthesis between models of resource allocation to repro-

410 duction

With the Bayesian model, it was possible to test two potential models with a constrain on the 41 resource allocation and no constrain and allowed us to identify inter-individual correlation. We 412 found that trade-offs emerged from the estimation of the correlations between individual random 413 effect without the inclusion of any constraints on resource allocation. This is not the case of 414 other models that impose detailed on mechanisms driving resource acquisition and allocation, 415 such as Dynamic Energy Budget model (DEB) or Resource Budget model (RBM). In DEB 416 models, reserves used for growth cannot be used to increase reproduction, as direct competition 417 occurs only among reserves used for building structure and not with those used for paying 418 maintenance costs (Kooijman, 2009). Hence, the fraction of reserves used for growth and repro-419 duction is fixed. In RBM models, the resource is allocated to reproduction only when resource 420 level exceeds a threshold (Isagi et al., 1997). Indeed, RBM models consider that plants cannot 421 have high fruit production during several years due to resource depletion (Crone and Rapp, 422 2014). Our results support the modeling choice used in DEB and RBM models where they 423 include allocation to reproduction. But modeling explicitly the variation of resource partition-424 ing between growth and reproduction across the phenological cycle could improve these models. 425

426

We also found here that accounting for the reproductive phenology, which determines the timing of resource allocation, can improve the estimation of trade-off between growth and reproductive sinks, and can highlight the idiosyncratic correlation patterns observed in trees so far. Indeed, in the study system, the sign of the correlation between growth and reproduction differs depending of the reproductive stage (initiation or maturation). But with shorter reproductive cycle, we could expect different results for the relation between growth and fruit initiation, because plants may invest resources for initiation and maturation during the same year.

434

## 435 6 Conclusion

Combining a simulation, process-based model with a Bayesian statistical estimation model can 436 be a fruitful approach to investigate the potential trade-off between reproduction and growth. 437 Process-based models integrate species ecology through their ecophysiological characteristics, 438 and allow to explicitly model the effects of environmental stresses on reproduction, and in 439 particular on reproductive phenology. However, the processes involved in tree reproduction are 440 still not well understood and most models assume that a constant fraction of resource is allocated 441 to reproduction (Vacchiano et al., 2018). Further improvements of the resource allocation 442 component in these models are needed to understand how tree growth and reproduction jointly 443 respond to climate change. 444

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## **453** 8 Author Contribution

V.J., F.L. & H.D. initiated the idea. F.L & F.C. provided experimental data. H.D. and V.J.
developed the ecophysiological model for *C. atlantica*. J.P. & E.W. designed the Bayesian
model with advice from all authors. V.J., E.W. and J.P. ran simulation. V.J. started the redaction
in consultation with F.L., H.D. and S.O-M. All authors contributed critically to the drafts and
gave final approval for publication.

## **9** Data accessibility statement

- 460 Data used, CASTANEA model input simulation and hierarchical Bayesian model code are
- <sup>461</sup> available from the Zenodo repository <https://doi.org/10.5281/zenodo.4433892>. The
- 462 CASTANEA model is an open-source software available on the CAPSIS plateform: http:
- 463 //capsis.cirad.fr/capsis/models

## **References**

465	Aitken SN, Yeaman S, Holliday JA, Wang T, Curtis-McLane S. 2008. Adaptation, migration
466	or extirpation: climate change outcomes for tree populations. Evolutionary Applications,
467	<b>1</b> (1): 95–111. doi:10.1111/j.1752-4571.2007.00013.x.

Barringer BC, Koenig WD, Knops JM. 2013. Interrelationships among life-history traits in
three California oaks. *Oecologia*, 171(1): 129–139. doi:10.1007/s00442-012-2386-9.

Bell G. 1980. The Costs of Reproduction and Their Consequences. *The American Naturalist*,
 116(1): 45–76. doi:10.1086/283611.

Berdanier AB, Clark JS. 2016. Divergent reproductive allocation trade-offs with canopy
exposure across tree species in temperate forests. *Ecosphere*, 7(6): 1–10. doi:10.1002/ecs2.
1313.

Bogdziewicz M, *et al.* 2020. From theory to experiments for testing the proximate mechanisms
of mast seeding: an agenda for an experimental ecology. *Ecology Letters*, 23(2): 210–220.
doi:10.1111/ele.13442.

Buoro M, Prévost E, Gimenez O. 2010. investigating evolutionary trade-offs in wild populations of atlantic salmon (salmo salar): incorporating detection probabilities and individual heterogeneity. *Evolution*, 64(9): 2629–2642. doi:10.1111/j.1558-5646.2010.01029.x.

Ciais P, *et al.* 2005. Europe-wide reduction in primary productivity caused by the heat and
drought in 2003. *Nature*, 437(7058): 529–533. doi:10.1038/nature03972.

Cremer K. 1992. Relation between reproductive growth and vegetative growth of Pinus radiata.
 *Forest Ecology and Management*, 52: 179–199. doi:10.1016/0378-1127(92)90501-Y.

Crone EE, Rapp JM. 2014. Resource depletion, pollen coupling, and the ecology of mast seeding: Mechanisms of mast seeding. *Annals of the New York Academy of Sciences*, 1322(1):
 21–34. doi:10.1111/nyas.12465.

Davi H, Barbaroux C, Francois C, Dufrêne E. 2009. The fundamental role of reserves and
 hydraulic constraints in predicting LAI and carbon allocation in forests. *Agricultural and Forest Meteorology*, 149(2): 349–361. doi:10.1016/j.agrformet.2008.08.014.

Davi H, Cailleret M. 2017. Assessing Drought-Driven Mortality Trees with Physiological
 Process-Based Models. *Agricultural and Forest Meteorology*, 232: 279–290. doi:10.1016/j.
 agrformet.2016.08.019.

Despland E, Houle G. 1997. Climate influences on growth and reproduction of Pinus banksiana
 (Pinaceae) at the limit of the species distribution in Eastern North America. *American Journal* Of Botany, 84(8): 928–937. doi:10.2307/2446283.

<sup>497</sup> Dufrêne E, Davi H, François C, Le Maire G, Le Dantec V, Granier A. 2005. Modelling
<sup>498</sup> carbon and water cycles in a beech forest. Part I: Model description and uncertainty analysis
<sup>499</sup> on modelled NEE. *Ecological Modelling*, 185(2-4): 407–436. doi:10.1016/j.ecolmodel.
<sup>500</sup> 2005.01.004.

Elmqvist T, Cates RG, Harper JK, Gardfjell H. 1991. Flowering in Males and Females of a
 Utah Willow, Salix rigida and Effects on Growth, Tannins, Phenolic Glycosides and Sugars.
 *Oikos*, 61(1): 65–72. doi:10.2307/3545407.

Fox JF. 1995. Shoot Demographic Responses to Manipulation of Reproductive Effort by Bud
 Removal in a Willow. *Oikos*, 72(2): 283–287. doi:10.2307/3546230.

Fox JF, Stevens GC. 1991. Costs of Reproduction in a Willow: Experimental Responses Vs.
 Natural Variation. *Ecology*, 72(3): 1013–1023. doi:10.2307/1940601.

Garland T, Carter P. 1994. Evolutionary Physiology. *Annual Review of Ecology and Systematics*, 56: 193–228. doi:10.1146/annurev.es.22.110191.001205.

<sup>510</sup> Guillemot J, Delpierre N, Vallet P, Francois C, Martin-StPaul NK, Soudani K, Nicolas M,

<sup>511</sup> Badeau V, Dufrene E. 2014. Assessing the effects of management on forest growth across

France: Insights from a new functional-structural model. *Annals of Botany*, **114**(4): 779–793.

<sup>513</sup> doi:10.1093/aob/mcu059.

514	Guillemot J, Klein EK, Davi H, Courbet F. 2015. The effects of thinning intensity and tree
515	size on the growth response to annual climate in Cedrus atlantica: a linear mixed modeling
516	approach. Annals of Forest Science, 72(5): 651–663. doi:10.1007/s13595-015-0464-y.

Hacket-Pain AJ, *et al.* 2018. Climatically controlled reproduction drives interannual growth
variability in a temperate tree species. *Ecology Letters*, 21(12): 1833–1844. doi:10.1111/ele.
13158.

Houle G. 2001. Reproductive costs are associated with both the male and female functions in Alnus viridis ssp. crispa. *Écoscience*, 8(2): 220–229. doi:10.1080/11956860.2001.11682648.

Isagi Y, Sugimura K, Sumida A, Ito H. 1997. How Does Masting Happen and Synchronize?
 *Journal of Theoretical Biology*, 187: 231–239. doi:10.1006/jtbi.1997.0442.

Kelly D, Sork VL. 2002. Mast Seeding in Perennial Plants: Why, How, Where? Annual
 *Review of Ecology and Systematics*, 33(1): 427–447. doi:10.1146/annurev.ecolsys.33.020602.
 095433.

Knops JMH, Koenig WD. 2012. Sex Allocation in California Oaks: Trade-Offs or Resource
 Tracking? *PLOS ONE*, 7(8): 1–8. doi:10.1371/journal.pone.0043492.

Knops JMH, Koenig WD, Carmen WJ. 2007. Negative correlation does not imply a tradeoff
 between growth and reproduction in California oaks. *Proceedings of the National Academy* of Sciences, 104(43): 16982–16985. doi:10.1073/pnas.0704251104.

Koenig WD, Knops J. 1998. Scale of mast-seeding and tree-ring growth. *Nature*, 396: 225–226.
 doi:10.1038/24293.

Kooijman B. 2009. Dynamic Energy Budget Theory for Metabolic Organisation. Cambridge:
 Cambridge University Press, 3rd edition. doi:10.1017/CBO9780511805400.

Krouchi F, Derridj A, Lefèvre F. 2004. Year and tree effect on reproductive organisation of
 Cedrus atlantica in a natural forest. *Forest Ecology and Management*, 197(1-3): 181–189.
 doi:10.1016/j.foreco.2004.05.013.

Lauder JD, Moran EV, Hart SC. 2019. Fight or flight? Potential tradeoffs between drought
 defense and reproduction in conifers. *Tree Physiology*, 39(7): 1071–1085. doi:10.1093/
 treephys/tpz031.

Lebourgeois F, Delpierre N, Dufrêne E, Cecchini S, Macé S, Croisé L, Nicolas M. 2018.
Assessing the roles of temperature, carbon inputs and airborne pollen as drivers of fructification in European temperate deciduous forests. *European Journal of Forest Research*, 137(3):
349–365. doi:10.1007/s10342-018-1108-1.

546 Lempereur M, Martin-StPaul NK, Damesin C, Joffre R, Ourcival JM, Rocheteau A, Ram-

bal S. 2015. Growth duration is a better predictor of stem increment than carbon supply
 in a Mediterranean oak forest: implications for assessing forest productivity under climate
 change. *New Phytologist*, 207(3): 579–590. doi:10.1111/nph.13400.

Linares JC, Taïqui L, Sangüesa-Barreda G, Seco JI, Camarero JJ. 2013. Age-related
 drought sensitivity of Atlas cedar (Cedrus atlantica) in the Moroccan Middle Atlas forests.
 *Dendrochronologia*, 31(2): 88–96. doi:10.1016/j.dendro.2012.08.003.

Lovett Doust J. 1989. Plant reproductive strategies and resource allocation. *Trends in Ecology*& *Evolution*, 4(8): 230–234. doi:10.1016/0169-5347(89)90166-3.

Martín D, Vázquez-Piqué J, Carevic FS, Fernández M, Alejano R. 2015. Trade-off between
 stem growth and acorn production in holm oak. *Trees*, 29(3): 825–834. doi:10.1007/
 s00468-015-1162-y.

Mund M, Herbst M, Knohl A, Matthäus B, Schumacher J, Schall P, Siebicke L, Tamrakar
R, Ammer C. 2020. It is not just a 'trade-off': indications for sink- and source-limitation to
vegetative and regenerative growth in an old-growth beech forest. *New Phytologist*, 226(1):
111–125. doi:10.1111/nph.16408.

Obeso JR. 2002. The costs of reproduction in plants. *New Phytologist*, 155: 321–348. doi:
 10.1046/j.1469-8137.2002.00477.x.

<sup>564</sup> Olijnyk AM, Nelson WA. 2013. Positive phenotypic correlations among life-history traits

remain in the absence of differential resource ingestion. *Functional Ecology*, 27: 165–172.
 doi:10.1111/1365-2435.12015.

Pulido F, Moreno G, Garcia E, Obrador JJ, Bonal R, Diaz M. 2014. Resource manipulation
 reveals flexible allocation rules to growth and reproduction in a Mediterranean evergreen oak.
 *Journal of Plant Ecology*, 7(1): 77–85. doi:10.1093/jpe/rtt017.

Pérez-Ramos IM, Ourcival JM, Limousin JM, Rambal S. 2010. Mast seeding under in creasing drought: results from a long-term data set and from a rainfall exclusion experiment.
 *Ecology*, 91(10): 3057–3068. doi:10.1890/09-2313.1.

Sala A, Hopping K, McIntire EJB, Delzon S, Crone EE. 2012. Masting in whitebark
pine (Pinus albicaulis) depletes stored nutrients. *New Phytologist*, 196(1): 189–199. doi:
10.1111/j.1469-8137.2012.04257.x.

Santos-del Blanco L, Climent J. 2014. Costs of female reproduction in a conifer tree: a wholetree level assessment. *Journal of Ecology*, 102(5): 1310–1317. doi:10.1111/1365-2745.
12283.

Schermer É, *et al.* 2019. Pollen limitation as a main driver of fruiting dynamics in oak
populations. *Ecology Letters*, 22(1): 98–107. doi:10.1111/ele.13171.

Sexton JP, McIntyre PJ, Angert AL, Rice KJ. 2009. Evolution and ecology of species
 range limits. Annual Review of Ecology, Evolution, and Systematics, 40(1): 415–436. doi:
 10.1146/annurev.ecolsys.l.

Stearns SC. 1989. Trade-Offs in Life-History Evolution. *Functional Ecology*, 3(3): 259–268.
 doi:10.2307/2389364.

Suzuki A. 2001. Resource allocation to vegetative growth and reproduction at shoot level in
 Eurya japonica (Theaceae): a hierarchical investment? *New Phytologist*, 152: 307–312.
 doi:10.1046/j.0028-646X.2001.00251.x.

589 Sánchez-Humanes B, Sork VL, Espelta JM. 2011. Trade-offs between vegetative growth and

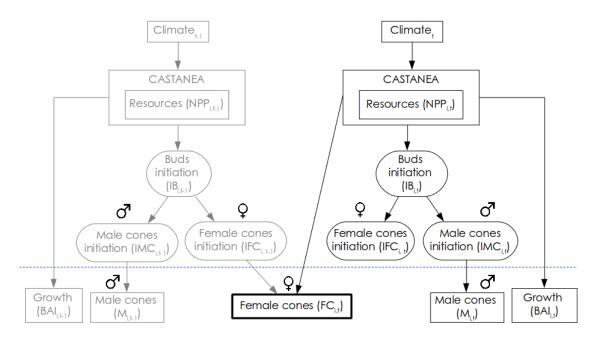
24

590	acorn production in Quercus lobata during a mast year: the relevance of crop size and hierarchi-
591	cal level within the canopy. <i>Oecologia</i> , <b>166</b> (1): 101–110. doi:10.1007/s00442-010-1819-6.
592	Thomas SC. 2011. Age-Related Changes in Tree Growth and Functional Biology: The Role
593	of Reproduction. FC Meinzer, B Lachenbruch, TE Dawson, editors, Size- and Age-Related
594	Changes in Tree Structure and Function, Dordrecht: Springer Netherlands, volume 4, 33-64.
595	doi:10.1007/978-94-007-1242-32.
596	Toth J. 1978. Contribution à l'étude de La Fructification et de La Régénération Du Cèdre
597	de l'Atlas Dans Le Sud de La France. Ph.D. thesis, Faculté des Sciences et Techniques de
598	Marseille Saint-Jérome.
599	Vacchiano G, et al. 2018. Reproducing reproduction: How to simulate mast seeding in forest
600	models. <i>Ecological Modelling</i> , <b>376</b> : 40–53. doi:10.1016/j.ecolmodel.2018.03.004.
601	van Noordwijk A, de Jong G. 1986. Acquisition and Allocation of Resources: Their Influence
602	on Variation in Life History Tactics. The American Naturalist, 128(1): 137-142. doi:
603	10.1086/284547.
604	Wardlaw IF. 1990. The control of carbon partitioning in plants. New Phytologist, 116: 341–381.
605	doi:10.1111/j.1469-8137.1990.tb00524.x.
606	Williams CG. 2009. Conifer reproductive biology. Dordrecht ; New York: Springer. OCLC:
607	ocn297148407.
608	Wu HX, Ker R, Chen Z, Ivkovic M. 2020. Balancing Breeding for Growth and Fecundity in
609	Radiata Pine Breeding Program. Evolutionary Applications. doi:10.1111/eva.13164.
610	Yasumura Y, Hikosaka K, Hirose T. 2006. Resource allocation to vegetative and reproductive
611	growth in relation to mast seeding in Fagus crenata. Forest Ecology and Management, 228-
612	233. doi:10.1016/j.foreco.2006.04.003.

<sup>613</sup> Żywiec M, Zielonka T. 2013. Does a heavy fruit crop reduce the tree ring increment? <sup>614</sup> Results from a 12-year study in a subalpine zone. *Trees*, 27(5): 1365–1373. doi: <sup>615</sup> 10.1007/s00468-013-0884-y.

## **10** List of Figures

#### **Process-model**



#### Data-model

Figure 1: Graphical representation of the hierarchical Bayesian model used with the process and data models. Right angles boxes represent observed variables and elliptic boxes represent unobserved variable (i.e. latent variables). We represent the previous year (t - 1) in gray color and current year (t) in black color. Resource (referred as  $NPP_{i,t}$ ) is simulated from the ecophysiological model (CASTANEA) with climate data. Resource is then allocated to radial growth  $(BAI_{i,t})$  and reproduction, which initiates buds  $(IB_{i,t})$ . Buds differ then between male  $(IMC_{i,t})$  and females  $(IFC_{i,t})$  according to phenotypic gender (no presented in the figure,  $PG_{i,t}$ ). During the year t we have the maturation of males cones  $(M_{i,t})$  of the same year t and the the female cones survival based on female cones initiated the previous year t - 1 (black bold box,  $FC_{i,t}$ ).

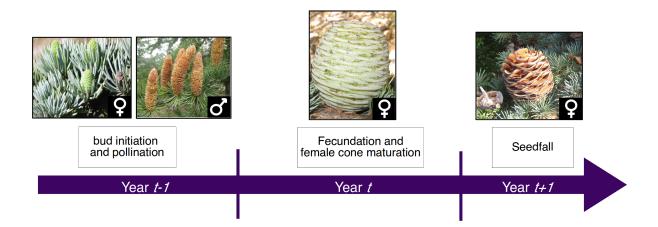


Figure 2: Overview of the general reproductive and growth cycle of *Cedrus atlantica*. Male reproduction and growth is carry on in one year. The duration of female reproduction takes two years from bud initiation during  $year_{t-1}$  followed by the maturation of female cone during  $year_t$ . Seedfall occurs during  $year_{t+1}$ .

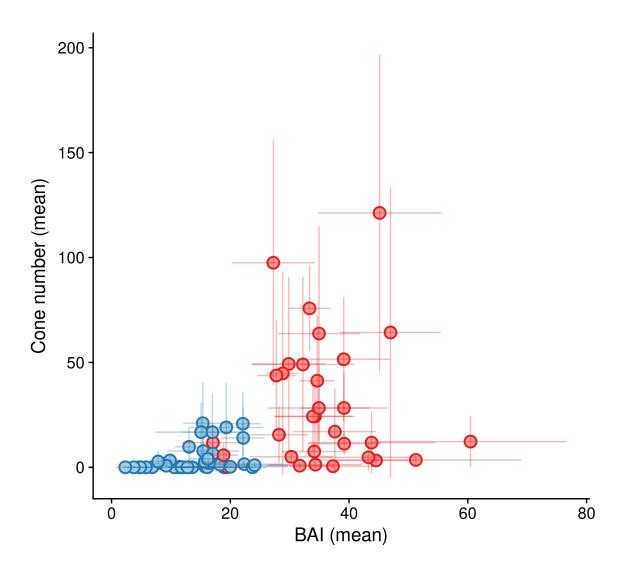


Figure 3: Raw data of female cone and growth for both density stand. Mean values with inter-annual standard deviations of female cone number  $(FC_{i,t}^{obs})$  and mean growth increment  $(BAI_{i,t}^{obs})$ , refereed here as BAI or Basal Area Increment in cm<sup>2</sup> year<sup>-1</sup>) for each individuals. Blue is for high density (1200 stems ha<sup>-1</sup>) and red color for low density stand (250 stems ha<sup>-1</sup>).

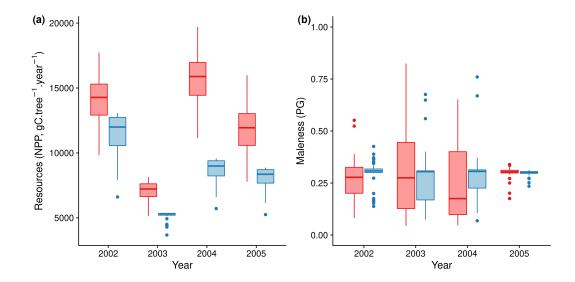


Figure 4: Output simulation of resource obtained from the ecophysiological model CASTANEA and estimation of the phenotypic gender from the hierarchical Bayesian model. (a) Boxplot of resource (Net Primary Productivity, in gC m<sup>-2</sup> year<sup>-1</sup>) simulated with the CASTANEA model for both stand density and years (b) Boxplot of phenotypic gender estimated for both stand density and years, ranging in y-axis from femaleness (0) to maleness (1). For both graphics, blue color is for the high density (1200 stems ha<sup>-1</sup>) and red for the low density stand (250 stems ha<sup>-1</sup>).

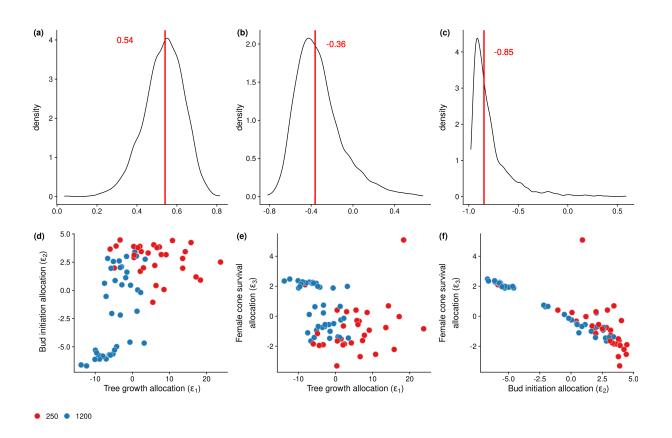


Figure 5: Correlation among sinks and their probabilities. (a) Density of the correlation  $\rho_{1,2}$  is positive (b) Density of the correlation  $\rho_{1,3}$  is negative (c) Density of the correlation  $\rho_{1,2}$  is negative. The red line represent the median posterior value for these three graphics. (d) Positive correlation between growth allocation ( $\epsilon_1$ ) and allocation to buds initiation ( $\epsilon_2$ ) (e) Negative correlation between growth allocation ( $\epsilon_1$ ) and female cone survival ( $\epsilon_3$ ) (f) Negative correlation between allocation to buds initiation ( $\epsilon_2$ ) and female cone survival ( $\epsilon_3$ ). Dots represent mean individual values, with blue and red color, respectively for high and low density of the stand.