

Agricultural fertilization with poultry manure results in persistent environmental contamination with the pathogen *Clostridioides difficile*

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Summary

During a field experiment applying broiler manure for fertilization of agricultural land, we detected viable *Clostridioides* (formerly, *Clostridium*) *difficile* in broiler feces, manure, dust, and fertilized soil. A large diversity of toxigenic *C. difficile* isolates was recovered, including PCR ribotypes common from human disease. Genomic relatedness of *C. difficile* isolates from dust and from soil, recovered more than two years after fertilization, traced their origins to the specific chicken farm that had delivered the manure. We present evidence of long-term contamination of agricultural soil with manure-derived *C. difficile* and demonstrate the potential for airborne dispersal of *C. difficile* through dust emissions during manure application. *Clostridioides* genome sequences virtually identical to those from manure had been recovered from chicken meat and from human infections in previous studies, suggesting broiler-associated *C. difficile* are capable of zoonotic transmission.

Introduction

The anaerobic gut bacterium *Clostridioides difficile* (formerly, *Clostridium difficile* (Lawson *et al.*, 2016)) is the most frequent infectious cause of antibiotic-associated diarrhea and among the leading culprits of healthcare-associated infections (Martin *et al.*, 2016). However, modelling studies have suggested that transmission in the community and in the healthcare system were equally relevant for sustaining *C. difficile* in the human population (Durham *et al.*, 2016; McLure *et al.*, 2019). Patients asymptomatically colonized with *C. difficile* upon hospital admission have a six-fold increased risk of suffering a *C. difficile* infection (CDI) (Zacharioudakis *et al.*, 2015), and even without developing CDI themselves they may increase the overall burden of nosocomial CDI significantly by spreading the pathogen to other patients (Longtin *et al.*, 2016; Blixt *et al.*, 2017; Donskey *et al.*, 2018). In addition, CDI occurs independent from healthcare at increasing incidence (Ofori *et al.*, 2018), but reservoirs and pathways of transmission outside of the hospital environment are incompletely understood (Warriner *et al.*, 2017; Rodriguez Diaz *et al.*, 2018).

Toxigenic *C. difficile* seems widespread in various environments, since it was recovered from domestic wastewater (Moradigaravand *et al.*, 2018; Numberger *et al.*, 2019) and river sediments (Zidaric *et al.*, 2010), from retail compost (Lim *et al.*, 2020), soil (Janezic *et al.*, 2016) and root vegetables (Lim *et al.*, 2018; Tkalec *et al.*, 2019). It was also found to colonize various mammals and birds, including wildlife, pets, and livestock (Weese, 2020). Notably, fattening pigs have been proposed as a potential source for transmission of *C. difficile* to humans, since strains with highly related genomes were isolated from both, pigs and farm workers (Knetsch *et al.*, 2018). *Clostridioides difficile* was also detected in chicken feces and

chicken meat repeatedly (Zidaric *et al.*, 2008; Weese *et al.*, 2010; Harvey *et al.*, 2011; Abdel-Glil *et al.*, 2018; Heise *et al.*, 2021), even though there is no evidence for significant CDI in birds (Weese, 2020). Livestock manure often contains *C. difficile* even after being treated by composting or fermentation in biogas plants (Usui *et al.*, 2017; Dharmasena and Jiang, 2018; Le Maréchal *et al.*, 2020). As a consequence, the disposal of manure or manure-derived products as fertilizer on agricultural land may lead to environmental contamination with *C. difficile* spores. The survival of *C. difficile* in fertilized agricultural soil and its release with surface water runoff or dust has as yet not been investigated, in contrast to other manure-derived pathogens (Blaustein *et al.*, 2015; Thiel *et al.*, 2020).

The spread of pathogenic bacteria can be tracked by comparing their genome sequences (Croucher *et al.*, 2015; Besser *et al.*, 2019; Thiel *et al.*, 2020). Within the EnteroBase platform, we have recently established a publicly accessible database for *Clostridioides* genomic data that currently (January 2021) contains 20,972 draft genomes and their associated metadata (Frentrup *et al.*, 2020). Standardized sequence data assembly and quality control in conjunction with core-genome multilocus sequence typing (cgMLST) and hierarchical clustering of cgMLST allelic profiles - as implemented in EnteroBase - facilitates the detection of *C. difficile* spread, since isolates from transmission chains frequently can be identified by being related at the HC2 level (i.e. constituting chains of genomes with pairwise differences of maximally two cgMLST alleles) (Frentrup *et al.*, 2020). Moreover, widespread epidemic strains commonly are related at the HC10 level, and PCR ribotypes correlate well with clusters at the HC150 level (which we dubbed 'core-genome sequence typing complexes'; CC) (Frentrup *et al.*, 2020).

In the present study, we detected the persistence of viable *C. difficile* in agricultural soil for several years following its fertilization with manure from broiler chickens. Genomic relatedness of *C. difficile* isolates from soil and from dust released during the fertilization process traced their origins to the specific chicken farm that had delivered the manure.

Results

Diversity of *C. difficile* isolates in chicken manure. Chicken manure was sampled at three different locations, including two farms and a manure trading cooperative. Altogether 146 *C. difficile* isolates were obtained from manure samples by applying an anaerobic enrichment protocol (Janezic *et al.*, 2018) and their genomes were sequenced. Genomic data indicated that 98% of the isolates carried both toxin genes, *tcdA* and *tcdB* (Suppl. Figure S1A), and only three isolates were non-toxigenic. Analysis of genome sequences with EnteroBase showed that manure isolates were related to 13 CCs (i.e. hierarchical clusters at the level HC150), which we had previously shown to correlate well with PCR ribotypes (RT, Figure 1A) (Frentrup *et al.*, 2020). The majority of isolates (94%) from Farm 1 were related to CC3 (Table 1, Figure 1B), which corresponds to PCR ribotype 001 (Frentrup *et al.*, 2020), and repeated samplings showed that this predominance of CC3 at Farm 1 was evident over a period of at least one year (Suppl. Figure S1B). In contrast, only one isolate (4%) from Farm 2 was CC3, and none from the manure trader (Figure 1B). Instead, isolates from the latter two suppliers were distributed among a number of different CCs, the most predominant of which were CC71 (RT014/020), CC88 (RT014), CC2 (RT002), CC86 (RT005), and CC391 (RT081) (Figure 1).

115 ***Close genomic relationships identify source of environmental C. difficile.*** A 2.1-hectare
 116 agricultural field was fertilized with 12 tons of poultry manure from Farm 1 (Thiel *et al.*, 2020).
 117 Prior to fertilization, our enrichment approach failed to detect any *C. difficile* in soil from this
 118 field. After fertilization, however, *C. difficile* got enriched and cultivated consistently from soil
 119 samples collected at multiple points in time for up to 143 weeks (Figure 2). Moreover, one
 120 dust sample collected during manure spread by using an aerosol collection device (Thiel *et*
 121 *al.*, 2020) at the edge of the field tested positive for *C. difficile* by enrichment (Figure 2). Of
 122 note, *C. difficile* was detected in soil and dust by enrichment culture only, whereas cultivation
 123 and quantification by direct plating on selective agar medium was not successful. Altogether,
 124 we collected 144 *C. difficile* isolates from fertilized soil and from dust, and from poultry feces
 125 and manure from Farm 1. Bacterial genome sequencing and cgMLST-based hierarchical
 126 clustering analysis with EnteroBase (Frentrup *et al.*, 2020) resulted in three HC2 clusters
 127 (HC2_1232, HC2_5435, HC2_5465; Figure 3) and four singletons. Generally, hierarchical
 128 clustering at the level HC2 indicates close genomic relationships of *C. difficile* isolates; it was
 129 previously shown to correlate with events of transmission between hospital patients (Frentrup
 130 *et al.*, 2020).
 131
 132 Two HC2 clusters (HC2_1232, HC2_5435) included genomes from two or more different
 133 sources, including chicken feces collected at Farm 1, manure from Farm 1, dust collected
 134 during the application of manure to the field, and fertilized soil from multiple points in time
 135 (Table 3 and Figure 3). This result confirmed that the *C. difficile* strains that were recovered
 136 during and after fertilization indeed originated from Farm 1, i.e. they had been disseminated
 137 onto the agricultural field through the fertilization process. These close genomic relationships
 138 were found among *C. difficile* isolates from all soil samples, indicating the persistence of

139 viable, manure-derived *C. difficile* in the soil for up to 143 weeks after fertilization (Figure 3).
 140 Likewise, the detection of closely related *C. difficile* in mineral dust showed that viable cells of
 141 the pathogen got aerosolized during the fertilization process and transported in an ascending
 142 dust plume at a distance of at least 20 meters from the applying tractor (Figure 3).

143

144 **PCR ribotypes and antibiotic susceptibilities.** *Clostridioides difficile* isolates (n=19)
 145 selected to represent sources (i.e. manure from the different suppliers, fertilized soil, and
 146 dust) and genomic diversity (at the level of CCs) proved to be phenotypically susceptible to
 147 the antibiotics vancomycin, metronidazole, moxifloxacin, clindamycin, and tetracycline (Table
 148 2). None of the genome sequences (n=278) carried resistance-causing mutations in the
 149 gyrase gene *gyrA* (Zaiß *et al.*, 2010), confirming the lack of fluoroquinolone resistance in our
 150 strain collective (not shown). PCR ribotypes determined in the laboratory were fully
 151 concordant with ribotype predictions based on hierarchical clustering in EnteroBase (Table 2).

152

153 **Closely related clinical and poultry meat isolates.** Hierarchical clustering of cgMLST allelic
 154 profiles in EnteroBase routinely determines genomic relationships at multiple phylogenetic
 155 levels among all >20,000 entries in the *Clostridioides* database (Frentrup *et al.*, 2020).
 156 Remarkably, a limited number of genome sequences from several previous studies were
 157 closely related (at HC2 level) to those from Farm 1 (Figure 3). Most notably, virtually identical
 158 *C. difficile* genome sequences had been recovered from retail chicken meat (n=6; Figure 3),
 159 which had been purchased in one region in Germany (Berlin and Brandenburg), but had been
 160 produced in a number of different cutting plants in Germany and the Netherlands (Heise *et al.*,
 161 2021). Additional closely related genomes originated from isolates from human patients
 162 suffering from *C. difficile* infection in Germany (n=2), the Netherlands (n=5) and Hungary

(n=1; Figure 3). Of note, these genomic similarities were not due to impaired quality of the sequence data, since >99% of cgMLST alleles were successfully called for all genome sequences. Moreover, no genes of the whole-genome MLST set (Frentrup *et al.*, 2020) were differentially present (not shown), indicating that accessory genomes were virtually identical among all these isolates, too,

Discussion

Chicken manure carried diverse *C. difficile*, including clinically relevant strains. Almost all *C. difficile* isolates from manure in our study carried the *tcdA* and *tcdB* genes in their genomes, and hence must be considered fully virulent and able to cause gastrointestinal disease in humans. This result is in concordance with most previous studies on poultry-associated *C. difficile* (e.g. (Dharmasena and Jiang, 2018; Berger *et al.*, 2020; Le Maréchal *et al.*, 2020; Heise *et al.*, 2021)) even though there is little evidence that *C. difficile* may cause disease in birds (Weese, 2020).

In manure samples from three suppliers, we found a total of 13 CCs (core-genome sequence-type complexes) of *C. difficile*. CCs correlate well with PCR ribotypes (Frentrup *et al.*, 2020) (Table 2), and ribotypes 001, 014/020 and 005 have been reported from poultry feces (Indra *et al.*, 2009; Hussain *et al.*, 2016; Abdel-Glil *et al.*, 2018; Le Maréchal *et al.*, 2020) and from broiler meat (De Boer *et al.*, 2011; Tkalec *et al.*, 2020) in the past. In our manure samples, the most predominant strains were CC3 (RT001), CC71 (RT014/020) and CC88 (RT014). Remarkably, these are also among the most prevalent strains causing human *C. difficile*

infections in Europe (Davies *et al.*, 2016). However, our isolates from broiler chickens were not resistant to fluoroquinolones or clindamycin, in contrast to the vast majority of clinical RT001 isolates from human CDI (Zaiß *et al.*, 2010; Eyre *et al.*, 2018). This striking difference in antibiotic resistances suggests that *C. difficile* RT001 in chickens constitutes a population separate from the epidemic RT001 strain causing healthcare-associated CDI in humans, with limited exchange. This notion was confirmed by hierarchical clustering, which indicated that all our CC3 *C. difficile* from broiler manure (n=199) were related to a single HC10 cluster (HC10_783; Suppl. Table 1) that currently includes only 15 (7%) human-associated *C. difficile* isolates in EnteroBase. This separation was not observed for RT014/020, which is antibiotic resistant more rarely (Zaiß *et al.*, 2010; Eyre *et al.*, 2018), and where 13 isolates from broilers were affiliated to nine different HC10 clusters (Suppl. Table 1), the larger of which included numerous isolates from diverse host species and geographic origins. Fluoroquinolone and clindamycin resistance in poultry-associated *C. difficile* has occasionally been reported (from the USA and Zimbabwe (Harvey *et al.*, 2011; Dharmasena and Jiang, 2018; Berger *et al.*, 2020)). Since macrolides and fluoroquinolones are the two antibiotics most heavily used in the poultry industry in Europe, and resistance against these drugs is widespread among other gastrointestinal pathogens from chickens (Roth *et al.*, 2019), lowered susceptibilities might also have been expected from broiler-associated *C. difficile*, but yet this was not detected in our samples.

Long-term persistence of manure-derived C. difficile in fertilized agricultural soil.

Clostridioides difficile has been reported from a wide range of different environmental samples, including soil (Rodriguez Diaz *et al.*, 2018). To our best knowledge, however, our study is the first to use genome sequence analysis to trace environmental *C. difficile* back to

its source. As one result, we show that *C. difficile* in fertilized soil indeed originated from chickens in Farm 1. Hence, our field experiment demonstrated that manure-derived *C. difficile* remained viable in fertilized soil over the entire study period, i.e. for at least 143 weeks, or almost three years. The continued bacteriological detection of *C. difficile* in all samples investigated suggested that its survival may be much longer than the sampling period, even though precise extrapolation was not possible due to the failure of quantitative cultivation. The observed long-term contamination of the soil certainly was enabled by the ability of *C. difficile* to produce endospores, which can stay viable for many years (Yang and Ponce, 2011). In contrast, these bacteria are unlikely to perform much metabolic activity or even proliferate under ambient conditions in the soil, since their physiology is adapted to life in the intestines of warm-blooded animals.

We previously reported that chicken manure carried additional pathogens, including *Enterococcus faecium* and extended-spectrum beta-lactamase (ESBL) producing *Escherichia coli* (Thiel *et al.*, 2020). However, ESBL *E. coli* died off within a few days during manure storage (Siller *et al.*, 2020) and enterococci rapidly declined in soil within weeks after fertilization (Thiel *et al.*, 2020). In the present study, in contrast, we demonstrate that viable *C. difficile* remained detectable in fertilized soil for several years and hence represented a long-lasting pollution.

Potential for long-distance dispersal of *C. difficile*. Hierarchical clustering indicated that altogether 13 entries in the *Clostridioides* database shared identical HC2 clusters (HC2_1232, HC2_5465) with isolates from Farm 1, i.e. they had highly similar cgMLST profiles with at most two allelic differences, despite their origins from unrelated, previous studies. Seven of

these isolates had been recovered from retail chicken meat from various cutting plants in Germany and the Netherlands (Heise *et al.*, 2021), indicating widespread dissemination of *C. difficile* HC2_1232 by the poultry industry. Furthermore, the occurrence of the same HC2 clone in human CDI in Germany, Hungary and the Netherlands indicates that this strain is able to cause human disease. Consequently, this *C. difficile* HC2 clone poses a risk of zoonotic transmission.

It should be noted that pathogen genomic similarity alone does not prove direct transmission between remote places, but should be interpreted with particular care in the absence of additional, epidemiological evidence (Besser *et al.*, 2019). However, several plausible scenarios for long-distance transport of poultry-associated *C. difficile* exist. Chicken meat contaminated with *C. difficile* (De Boer *et al.*, 2011; Harvey *et al.*, 2011; Candel-Pérez *et al.*, 2020; Heise *et al.*, 2021) gets distributed to customers through widely ramified retail chains. Similarly, pork products (i.e., meat or manure) were suspected to promote the long-distance spread of *C. difficile*, after closely related *C. difficile* genomes had been detected in fattening pigs and humans across large geographic distances, without any documented epidemiological connections (Knetsch *et al.*, 2018; Knight *et al.*, 2019). Another potential path for the long-range dissemination of livestock-associated *C. difficile* may be the transport of colonized, live animals, e.g. from farms to slaughterhouses (Heise *et al.*, 2021). Potentially even more important is the globalized structure of the poultry industry, which ships industrially produced broiler chicks by airfreight for stocking fattening farms globally (Lowder *et al.*, 2009). It would be interesting to investigate the colonization status of chickens upon their arrival at fattening farms.

In addition, here we show that mineral dust from agricultural operations may carry aerosolized, manure-derived *C. difficile*. This dust may stay airborne for several days and during this time may get transported over several hundred kilometers, depending on atmospheric conditions (Faust *et al.*, 2020; Thiel *et al.*, 2020). Poultry manure is particularly prone to aerosolization due to its high dry-matter content (Kabelitz *et al.*, 2020; Thiel *et al.*, 2020) and therefore, its application for fertilization of agricultural fields likely contributes to the airborne dispersal of chicken-associated *C. difficile* over long distances. Aerosolized *C. difficile* is considered a potential source of human infection when inhaled (Best *et al.*, 2010), similar to other enteric pathogens (Jahne *et al.*, 2015). Hence, *C. difficile* in agricultural dust may represent a risk of airborne zoonotic transmission. Taken together, our results corroborate the relevance of a 'One Health' approach for curbing the spread of *C. difficile* between human, livestock, and environmental reservoirs.

Experimental procedures

Manure samples. To capture the diversity of *C. difficile* isolates in manure samples, three samplings were performed on three different sites. Manure samples from two broiler fattening farms and one manure trading cooperative were investigated. In addition, chicken feces were sampled by collecting 30 chicken droppings from each of 11 stables in Farm 1. Farm 1 is an intensive poultry-fattening farm in Brandenburg, Germany, housing about 19,000 animals per stable on wood pellets. Manure from this farm was sampled three times (May 30th 2017, November 8th 2017 and May 19th 2018). In Farm 2, which is located in Saxony-Anhalt,

Germany, manure was collected in four different stables on August 14th 2017. Manure from the trader was sampled on March 27th 2017.

Field experiment. In a field experiment, 12 tons of chicken manure from Farm 1 (see above) were applied to a 2.1-hectare agricultural field, which had not been fertilized with animal manure for 15 years. Details of this experiment have been published previously (Thiel *et al.*, 2020). Briefly, dust particles that were released during the fertilization process were collected by impingement into 5 mL phosphate-buffered saline (PBS) at a height above ground of 1.50 m and at a distance from the tractor of 20, 50, and 100 m, respectively. Soil samples were taken on three representative sites on the field site prior to fertilization, directly after, and two, four, seven, ten, 14, 19 and 143 weeks later.

Isolation of *C. difficile* isolates. Ten g of poultry feces, manure and soil samples were mixed with 90 g Luria-Bertani broth (Roth) each and subsequently homogenized for 30 s with a bag mixer (Interscience). After sedimentation of coarse particles (30 min, room temperature), supernatants and impingement suspensions from the aerosol collector were diluted to extinction with PBS and subsequently streaked on ChromID *C. difficile*-agar (Biomérieux). After incubation at 37°C for 24 h, *C. difficile* colonies were identified by species-specific PCR (locus TR10) (Zaiß *et al.*, 2009). In addition, for enrichment cultures, 0.5 mL of suspensions were added to 10 mL brain heart infusion (BHI) broth (Roth) supplemented with 0.1% taurocholic acid (Sigma), 0.1% cysteine (Sigma) and *C. difficile* selective Supplement (Oxoid) in Hungate tubes (Janezic *et al.*, 2018). After seven days of incubation at 37°C, an ethanol shock was performed by adding an equal amount of absolute ethanol to 0.5 mL culture and incubation for 1 h at room temperature. The culture was centrifuged at 2,500 x g for 5

minutes, the resulting cell pellet was resuspended in 200 µL PBS, and 100 µL were plated on ChromID *C. difficile*-agar and incubated at 37°C for 24 h. Again, bacterial colonies were tested by *C. difficile*-specific PCR (Zaiß *et al.*, 2009).

Antibiotic susceptibility testing. Isolates from agar-plates were transferred to anaerobic BHI broth (Roth) in Hungate tubes and grown for two days at 37°C. Subsequently, the culture was diluted 1:5 with PBS and 100 µl was spread on Columbia blood-agar (Oxoid). For each antimicrobial agent, an E-test strip was applied to the agar surface, followed by 24 hours of incubation at 37°C. The tests were interpreted visually by reading the minimum inhibitory concentration (MIC). MICs were determined for vancomycin, metronidazole, moxifloxacin (Biomérieux), clindamycin and tetracycline (Liofilchem). For interpretation, MIC breakpoints for antibiotic resistance were applied according to Pirš *et al.* (Pirš *et al.*, 2013): metronidazole, ≥2 µg/mL; vancomycin, ≥2 µg/mL; moxifloxacin, ≥4 µg/mL; clindamycin, ≥8 µg/mL; tetracycline, ≥16 µg/mL.

PCR ribotyping. PCR ribotyping of *C. difficile* isolates was performed as reported previously (Indra *et al.*, 2008), applying capillary electrophoresis and the Webribo database (<https://webribo.ages.at/>).

Whole genome sequencing. Genomic DNA was extracted by using the DNeasy Blood & Tissue kit (Qiagen), libraries were prepared as described previously (Steglich *et al.*, 2018) and sequenced on an Illumina NextSeq 500 machine using a Mid-Output kit (Illumina) with 300 cycles. Illumina sequencing reads were uploaded to EnteroBase (<http://enterobase.warwick.ac.uk/>) and assembled with the embedded standardized pipeline

330 (Frentrup *et al.*, 2020). Thirty-two sequences did not pass the quality check in EnteroBase
 331 (Frentrup *et al.*, 2020) and were excluded from further analyses. For 278 genomes, cgMLST
 332 allelic profiles (>99% complete) were determined and cgMLST-based hierarchical clustering
 333 performed using EnteroBase tools. To visualize genomic relatedness, rapid-neighbor-joining
 334 and minimum-spanning trees were calculated applying GrapeTree (Zhou *et al.*, 2018;
 335 Frentrup *et al.*, 2020). PCR ribotypes were predicted based on genomic relatedness at the
 336 level HC150 (i.e. hierarchical clusters of genome sequences with pairwise differences of
 337 maximally 150 cgMLST alleles; for details see (Frentrup *et al.*, 2020)).
 338
 339 Sequences of the *gyrA* gene (cgMLST locus CD630_00060) were scanned for the mutations
 340 Thr-82-Ile and Asp-71-Glu, which are associated with fluoroquinolone resistance in *C. difficile*
 341 (Zaiß *et al.*, 2010).
 342
 343 All genome sequencing data were submitted to the European Nucleotide Archive
 344 (www.ebi.ac.uk/ena) under the study accession number PRJEB42049. A list of all analyzed
 345 genomes can be found in Supplementary Table1.
 346
 347 **Detection of toxin genes.** DNA from selected isolates (n=19) was tested for the presence of
 348 toxin genes *tcdA*, *tcdB*, *cdtA* and *cdtB* by PCR (Persson *et al.*, 2008). The presence or
 349 absence of toxin genes *tcdA* and *tcdB* was determined for all genomes in this study (n=275)
 350 based on allelic numbers for toxin gene loci in EnteroBase (i.e., allele number 0 was
 351 interpreted as absence of gene).
 352

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Conflict of interest.

The authors declare no conflict of interest.

References

- Abdel-Glil, M.Y., Thomas, P., Schmoock, G., Abou-El-Azm, K., Wieler, L.H., Neubauer, H., and Seyboldt, C. (2018) Presence of *Clostridium difficile* in poultry and poultry meat in Egypt. *Anaerobe* **51**: 21–25.
- Berger, F.K., Mellmann, A., Bischoff, M., von Müller, L., Becker, S.L., Simango, C., and Gärtner, B. (2020) Molecular epidemiology and antimicrobial resistance of *Clostridioides difficile* detected in chicken, soil and human samples from Zimbabwe. *Int J Infect Dis* **96**: 82–87.
- Besser, J.M., Carleton, H.A., Trees, E., Stroika, S.G., Hise, K., Wise, M., and Gerner-Smidt, P. (2019) Interpretation of whole-genome sequencing for enteric disease surveillance and outbreak investigation. *Foodborne Pathog Dis* **16**: 504–512.
- Best, E.L., Fawley, W.N., Parnell, P., and Wilcox, M.H. (2010) The potential for airborne dispersal of *Clostridium difficile* from symptomatic patients. *Clin Infect Dis* **50**: 1450–1457.
- Blaustein, R.A., Pachepsky, Y.A., Shelton, D.R., and Hill, R.L. (2015) Release and removal of microorganisms from land-deposited animal waste and animal manures: a review of data and models. *J Environ Qual* **44**: 1338–1354.
- Blixt, T., Gradel, K.O., Homann, C., Seidelin, J.B., Schønning, K., Lester, A., et al. (2017) Asymptomatic carriers contribute to nosocomial *Clostridium difficile* infection: A cohort study of 4508 patients. *Gastroenterology* **152**: 1031-1041.e2.
- De Boer, E., Zwartkruis-Nahuis, A., Heuvelink, A.E., Harmanus, C., and Kuijper, E.J. (2011) Prevalence of *Clostridium difficile* in retailed meat in the Netherlands. *Int J Food Microbiol* **144**: 561–564.
- Candel-Pérez, C., Santaella-Pascual, J., Ros-Berruezo, G., and Martínez-Graciá, C. (2020) Occurrence of *Clostridioides* [*Clostridium*] *difficile* in poultry giblets at slaughter and retail pork and poultry meat in southeastern Spain. *J Food Prot* Epub ahead of print.
- Croucher, N.J., Page, A.J., Connor, T.R., Delaney, A.J., Keane, J.A., Bentley, S.D., et al. (2015) Rapid phylogenetic analysis of large samples of recombinant bacterial whole genome sequences using Gubbins. *Nucleic Acids Res* **43**: e15.
- Davies, K.A., Ashwin, H., Longshaw, C.M., Burns, D.A., Davis, G.L., and Wilcox, M.H. (2016) Diversity of *Clostridium difficile* PCR ribotypes in Europe: Results from the European, multicentre, prospective, biannual, point-prevalence study of *Clostridium difficile* infection

397 in hospitalised patients with diarrhoea (EUCLID), 2012 and 2. *Euro Surveill* **21**:
398 pii=30294.

399 Dharmasena, M. and Jiang, X. (2018) Isolation of toxigenic *Clostridium difficile* from animal
400 manure and composts being used as biological soil amendments. *Appl Environ Microbiol*
401 **84**: e00738-18.

402 Donskey, C.J., Sunkesula, V.C.K., Stone, N.D., Gould, C. V., McDonald, L.C., Samore, M., et
403 al. (2018) Transmission of *Clostridium difficile* from asymptotically colonized or
404 infected long-term care facility residents. *Infect Control Hosp Epidemiol* **39**: 909–916.

405 Durham, D.P., Olsen, M.A., Dubberke, E.R., Galvani, A.P., and Townsend, J.P. (2016)
406 Quantifying transmission of *Clostridium difficile* within and outside healthcare settings.
407 *Emerg Infect Dis* **22**: 608–616.

408 Eyre, D.W., Davies, K.A., Davis, G., Fawley, W.N., Dingle, K.E., De Maio, N., et al. (2018) Two
409 distinct patterns of *Clostridium difficile* diversity across Europe indicating contrasting
410 routes of spread. *Clin Infect Dis* **67**: 1035–1044.

411 Faust, M., Wolke, R., Münch, S., Funk, R., and Schepanski, K. (2020) A new Lagrangian in-
412 time particle simulation module (Itpas v1) for atmospheric particle dispersion. *Geosci*
413 *Model Dev Discuss*: [preprint].

414 Frentrup, M., Zhou, Z., Steglich, M., Meier-Kolthoff, J.P., Göker, M., Riedel, T., et al. (2020) A
415 publicly accessible database for *Clostridioides difficile* genome sequences supports
416 tracing of transmission chains and epidemics. *Microb Genomics* **6**: e000410.

417 Harvey, R.B., Norman, K.N., Andrews, K., Hume, M.E., Scanlan, C.M., Callaway, T.R., et al.
418 (2011) *Clostridium difficile* in poultry and poultry meat. *Foodborne Pathog Dis* **8**: 1321–
419 1323.

420 Heise, J., Witt, P., Maneck, C., Wichmann-Schauer, H., and Maurischat, S. (2021) Prevalence
421 and phylogenetic relationship of *Clostridioides difficile* strains in fresh poultry meat
422 samples processed in different cutting plants. *Int J Food Microbiol* **339**: 109032.

423 Hussain, Isfaqu, Borah, P., Sharma, R.K., Rajkhowa, S., Rupnik, M., Saikia, D.P., et al.
424 (2016) Molecular characteristics of *Clostridium difficile* isolates from human and animals
425 in the North Eastern region of India. *Mol Cell Probes* **30**: 306–311.

426 Indra, A., Huhulescu, S., Schneeweis, M., Hasenberger, P., Kernbichler, S., Fiedler, A., et al.
427 (2008) Characterization of *Clostridium difficile* isolates using capillary gel electrophoresis-
428 based PCR ribotyping. *J Med Microbiol* **57**: 1377–82.

429 Indra, A., Lassnig, H., Baliko, N., Much, P., Fiedler, A., Huhulescu, S., and Allerberger, F.
430 (2009) *Clostridium difficile*: A new zoonotic agent? *Wien Klin Wochenschr* **121**: 91–95.

431 Jahne, M.A., Rogers, S.W., Holsen, T.M., and Grimberg, S.J. (2015) Quantitative microbial
432 risk assessment of bioaerosols from a manure application site. *Aerobiologia (Bologna)*
433 **31**: 73–87.

434 Janezic, S., Mlakar, S., and Rupnik, M. (2018) Dissemination of *Clostridium difficile* spores
435 between environment and households: Dog paws and shoes. *Zoonoses Public Health*
436 **65**: 669–674.

437 Janezic, S., Potocnik, M., Zidaric, V., and Rupnik, M. (2016) Highly divergent *Clostridium*
438 *difficile* strains isolated from the environment. *PLoS One* **11**: e0167101.

439 Kabelitz, T., Ammon, C., Funk, R., Münch, S., Biniash, O., Nübel, U., *et al.* (2020) Functional
440 relationship of particulate matter (PM) emissions, animal species, and moisture content
441 during manure application. *Environ Int* **143**: 105577.

442 Knetsch, C.W., Kumar, N., Forster, S.C., Connor, T.R., Browne, H.P., Harmanus, C., *et al.*
443 (2018) Zoonotic transfer of *Clostridium difficile* harboring antimicrobial resistance
444 between farm animals and humans. *J Clin Microbiol* **56**: e01384-17.

445 Knight, D.R., Kullin, B., Androga, G.O., Barbut, F., Eckert, C., Johnson, S., *et al.* (2019)
446 Evolutionary and genomic insights into *Clostridioides difficile* sequence type 11: a diverse
447 zoonotic and antimicrobial-resistant lineage of global One Health importance. *MBio* **10**:
448 e00446-19.

449 Lawson, P.A., Citron, D.M., Tyrrell, K.L., and Finegold, S.M. (2016) Reclassification of
450 *Clostridium difficile* as *Clostridioides difficile* (Hall and O'Toole 1935) Prévot 1938.
451 *Anaerobe* **40**: 95–99.

452 Lim, S.-C., Foster, N.F., Elliott, B., and Riley, T. V. (2018) High prevalence of *Clostridium*
453 *difficile* on retail root vegetables, Western Australia. *J Appl Microbiol* **124**: 585–590.

454 Lim, S.-C., Knight, D.R., Moono, P., Foster, N.F., and Riley, T. V. (2020) *Clostridium difficile* in
455 soil conditioners, mulches and garden mixes with evidence of a clonal relationship with
456 historical food and clinical isolates. *Environ Microbiol Rep* **12**: 672–680.

457 Longtin, Y., Paquet-Bolduc, B., Gilca, R., Garenc, C., Fortin, E., Longtin, J., *et al.* (2016) Effect
458 of detecting and isolating *Clostridium difficile* carriers at hospital admission on the
459 incidence of C. difficile infections; A quasi-experimental controlled study. *JAMA Intern*
460 *Med* **176**: 796–804.

461 Lowder, B. V., Guinane, C.M., Zakour, N.L.B., Weinert, L.A., Conway-Morris, A., Cartwright,
462 R.A., *et al.* (2009) Recent human-to-poultry host jump, adaptation, and pandemic spread
463 of *Staphylococcus aureus*. *Proc Natl Acad Sci U S A* **106**: 19545–19550.

464 Le Maréchal, C., Gateau, C., Poezevara, T., Couturier, J., Rouxel, S., Syed Zaidi, R., *et al.*
 465 (2020) Characterization of *Clostridioides difficile* strains isolated from manure and
 466 digestate in five agricultural biogas plants. *Anaerobe* **62**: 102180.

467 Martin, J.S.H., Monaghan, T.M., and Wilcox, M.H. (2016) *Clostridium difficile* infection:
 468 Epidemiology, diagnosis and understanding transmission. *Nat Rev Gastroenterol Hepatol*
 469 **13**: 206–216.

470 McLure, A., Clements, A.C.A., Kirk, M., and Glass, K. (2019) Modelling diverse sources of
 471 *Clostridium difficile* in the community: Importance of animals, infants and asymptomatic
 472 carriers. *Epidemiol Infect* **147**: 1–9.

473 Moradigaravand, D., Gouliouris, T., Ludden, C., Reuter, S., Jamrozy, D., Blane, B., *et al.*
 474 (2018) Genomic survey of *Clostridium difficile* reservoirs in the East of England
 475 implicates environmental contamination of wastewater treatment plants by clinical
 476 lineages. *Microb Genomics* **4**: e000162.

477 Nummerger, D., Riedel, T., McEwen, G., Nübel, U., Frentrup, M., Schober, I., *et al.* (2019)
 478 Genomic analysis of three *Clostridioides difficile* isolates from urban water sources.
 479 *Anaerobe* **56**: 22–26.

480 Ofori, E., Ramai, D., Dhawan, M., Mustafa, F., Gasperino, J., and Reddy, M. (2018)
 481 Community-acquired *Clostridium difficile*: epidemiology, ribotype, risk factors, hospital
 482 and intensive care unit outcomes, and current and emerging therapies. *J Hosp Infect* **99**:
 483 436–442.

484 Persson, S., Torpdahl, M., and Olsen, K.E.P. (2008) New multiplex PCR method for the
 485 detection of *Clostridium difficile* toxin A (*tcdA*) and toxin B (*tcdB*) and the binary toxin
 486 (*cdtA/cdtB*) genes applied to a Danish strain collection. *Clin Microbiol Infect* **14**: 1057–
 487 1064.

488 Pirš, T., Avberšek, J., Zdovc, I., Krt, B., Andlovic, A., Lejko-Zupanc, T., *et al.* (2013)
 489 Antimicrobial susceptibility of animal and human isolates of *Clostridium difficile* by broth
 490 microdilution. *J Med Microbiol* **62**: 1478–1485.

491 Rodriguez Diaz, C., Seyboldt, C., and Rupnik, M. (2018) Non-human *C. difficile* reservoirs
 492 and sources: Animals, food, environment. *Adv Exp Med Biol* **1050**: 227–243.

493 Roth, N., Käsbohrer, A., Mayrhofer, S., Zitz, U., Hofacre, C., and Domig, K.J. (2019) The
 494 application of antibiotics in broiler production and the resulting antibiotic resistance in
 495 *Escherichia coli*: A global overview. *Poult Sci* **98**: 1791–1804.

496 Siller, P., Daehre, K., Thiel, N., Nübel, U., and Roesler, U. (2020) Impact of short-term storage
 497 on the quantity of extended-spectrum beta-lactamase–producing *Escherichia coli* in
 498 broiler litter under practical conditions. *Poult Sci* **99**: 2125–2135.

499 Steglich, M., Hofmann, J.D., Helmecke, J., Sikorski, J., Spröer, C., Riedel, T., *et al.* (2018)
500 Convergent loss of ABC transporter genes from *Clostridioides difficile* genomes is
501 associated with impaired tyrosine uptake and p-cresol production. *Front Microbiol* **9**: 901.

502 Thiel, N., Münch, S., Behrens, W., Junker, V., Faust, M., Biniash, O., *et al.* (2020) Airborne
503 bacterial emission fluxes from manure-fertilized agricultural soil. *Microb Biotechnol* **13**:
504 1631–1647.

505 Tkalec, V., Jamnikar-Ciglenecki, U., Rupnik, M., Vahnjal, S., Zelenik, K., and Biasizzo, M.
506 (2020) *Clostridioides difficile* in national food surveillance, Slovenia, 2015 to 2017. *Euro*
507 *Surveill* **25**: pii=1900479.

508 Tkalec, V., Janezic, S., Skok, B., Simoncic, T., Mesaric, S., Vrabic, T., and Rupnik, M. (2019)
509 High *Clostridium difficile* contamination rates of domestic and imported potatoes
510 compared to some other vegetables in Slovenia. *Food Microbiol* **78**: 194–200.

511 Usui, M., Kawakura, M., Yoshizawa, N., San, L.L., Nakajima, C., Suzuki, Y., and Tamura, Y.
512 (2017) Survival and prevalence of *Clostridium difficile* in manure compost derived from
513 pigs. *Anaerobe* **43**: 15–20.

514 Warriner, K., Xu, C., Habash, M., Sultan, S., and Weese, S.J. (2017) Dissemination of
515 *Clostridium difficile* in food and the environment: Significant sources of *C. difficile*
516 community-acquired infection? *J Appl Microbiol* **122**: 542–553.

517 Weese, J.S. (2020) *Clostridium (Clostridioides) difficile* in animals. *J Vet Diagnostic Investig*
518 **32**: 213–221.

519 Weese, J.S., Reid-Smith, R.J., Avery, B.P., and Rousseau, J. (2010) Detection and
520 characterization of *Clostridium difficile* in retail chicken. *Lett Appl Microbiol* **50**: 362–365.

521 Yang, W.W. and Ponce, A. (2011) Validation of a *Clostridium* endospore viability assay and
522 analysis of greenland ices and atacama desert soils. *Appl Environ Microbiol* **77**: 2352–
523 2358.

524 Zacharioudakis, I.M., Zervou, F.N., Pliakos, E.E., Ziakas, P.D., and Mylonakis, E. (2015)
525 Colonization with toxinogenic *C. difficile* upon hospital admission, and risk of infection: A
526 systematic review and meta-analysis. *Am J Gastroenterol* **110**: 381–390.

527 Zaiß, N.H., Rupnik, M., Kuijper, E.J., Harmanus, C., Michielsen, D., Janssens, K., and Nübel,
528 U. (2009) Typing *Clostridium difficile* strains based on tandem repeat sequences. *BMC*
529 *Microbiol* **9**: 1–11.

530 Zaiß, N.H., Witte, W., and Nübel, U. (2010) Fluoroquinolone resistance and *Clostridium*
531 *difficile*, Germany. *Emerg Infect Dis* **16**: 675–677.

532 Zhou, Z., Alikhan, N.F., Sergeant, M.J., Luhmann, N., Vaz, C., Francisco, A.P., *et al.* (2018)
533 Grapetree: Visualization of core genomic relationships among 100,000 bacterial
534 pathogens. *Genome Res* **28**: 1395–1404.

535 Zidaric, V., Beigot, S., Lapajne, S., and Rupnik, M. (2010) The occurrence and high diversity
536 of *Clostridium difficile* genotypes in rivers. *Anaerobe* **16**: 371–375.

537 Zidaric, V., Zemljic, M., Janezic, S., Kocuvan, A., and Rupnik, M. (2008) High diversity of
538 *Clostridium difficile* genotypes isolated from a single poultry farm producing replacement
539 laying hens. *Anaerobe* **14**: 325–327.

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552 **Figure legends**

553

554 **Figure 1.** Rapid-neighbour-joining phylogenetic trees based on cgMLST allelic differences
555 between *C. difficile* isolates (n=146) from manure samples. The scale bar indicates the
556 branch length corresponding to sequence differences at 200 cgMLST loci. **A** Colors indicate
557 CCs (core-genome sequence-typing complexes); RT, PCR ribotype. **B** Colors indicate origins
558 of manure.

559

560 **Figure 2.** Distribution of isolates recovered from chicken feces and manure from Farm 1 and
561 from samples collected during the field experiment (n=144). Colors represent CCs, which
562 were determined based on cgMLST allelic profiles in EnteroBase.

563

564 **Figure 3. A** Rapid-neighbor-joining phylogenetic tree based on cgMLST allelic profiles from all
565 isolates (n=144) sampled during the field experiment. Colors indicate HC2 clusters. The scale
566 bar indicates the branch length corresponding to sequence differences at 200 cgMLST loci. **B**
567 Minimum-spanning trees for two HC2 clusters. Numbers on branches indicate the number of
568 cgMLST allelic differences and the colors represent the source from which the isolates were
569 extracted from.

570

571 **Supplementay Figure S1.** Rapid-neighbour-joining phylogenetic trees based on cgMLST
572 allelic differences between *C. difficile* isolates (n=146) from manure samples (compare Figure
573 1). The scale bar indicates the branch length corresponding to sequence differences at 200
574 cgMLST loci. **A** Colors indicate the complement of toxin genes *tcdA* and *tcdB*. **B** Colors
575 indicate sampling dates.

Table 1. Core genome sequence type complexes (CC)

CC	PCR Ribotype	Farm1	Farm2	trader
2	002	4 (4.3 %)	1 (4.2 %)	7 (25 %)
3	001	88 (93.6 %)	1 (4.2 %)	0
71	014/020	0	10 (41.7 %)	3 (10.7 %)
86	005	1 (1.1 %)	2 (8.3 %)	4 (14.3 %)
88	014	0	2 (8.3 %)	10 (35.7 %)
391	081	0	5 (20.8 %)	0
5408	029	0	1 (4.2 %)	0
5410	novel	0	2 (8.3 %)	0
207	003	1 (1.1 %)	0	0
34	014	0	0	1 (3.6 %)
596	011/049	0	0	1 (3.6 %)
645	029	0	0	1 (3.6 %)
1643	011/049	0	0	1 (3.6 %)

Table2. Genotypes and antibiotic susceptibilities of 19 selected *C. difficile* isolates

				MIC** [µg/ml]					Gene content (predicted*** PCR)			
Isolate	Source	Origin	CC (RT predicted* PCR)	VAN	MTZ	MXF	CLI	TET	<i>tcdA</i>	<i>tcdB</i>	<i>cdtA</i>	<i>cdtB</i>
CD-17-00892	manure	trader	88 (RT014 RT014)	0.5	0.094	0.75	1.5	0.016	++	++	-	-
CD-17-01035	manure	trader	1643 (RT011/049 RT049)	0.5	0.125	0.75	3	0.047	++	++	-	-
CD-17-01037	manure	trader	34 (RT014 RT014)	0.5	0.125	0.75	2	0.032	++	++	-	-
CD-17-01039	manure	trader	645 (RT029 RT029)	0.75	0.125	0.75	0.5	0.047	++	++	-	-
CD-17-01040	manure	trader	596 (RT011/049 RT049)	0.5	0.125	0.75	3	0.047	++	++	-	-
CD-17-01068	manure	Farm 1	207 (RT003 RT003)	0.38	0.094	1	0.5	0.023	++	++	-	-
CD-17-01070	dust	field experiment	3 (RT001 RT001)	0.5	0.094	0.5	1.5	0.032	++	++	-	-
CD-17-01381	manure	Farm 2	71 (RT014/020 RT014)	0.75	0.125	0.75	1.5	0.75	++	++	-	-
CD-17-01390	manure	Farm 2	391 (RT081 RT081)	0.5	0.125	0.75	2	0.032	++	++	-	-
CD-17-01395	manure	Farm 2	5408 (n.a. RT029)	0.75	0.125	1	3	0.032	-	-	-	-
CD-17-01424	manure	Farm 2	5410 (n.a. novel)	0.38	0.032	1	2	0.023	-	-	-	-
CD-17-01524	manure	Farm 1	86 (RT005 RT005)	1	0.19	0.75	6	0.032	++	++	-	-
CD-18-00685	manure	Farm 1	2 (RT002 RT002)	0.5	0.064	1	2	0.023	++	++	-	-
CD-19-00355	fert. soil wk. 7	field experiment	2 (RT002 RT002)	0.5	0.125	0.75	2	0.064	++	++	-	-
CD-19-00409	fert. soil wk. 14	field experiment	71 (RT014/020 RT014)	0.75	0.125	1	4	0.032	++	++	-	-
CD-19-00417	fert. soil wk. 14	field experiment	3 (RT001 RT001)	0.5	0.19	0.5	1.5	0.047	++	++	-	-
CD-19-00426	fert. soil wk. 19	field experiment	2 (RT002 RT002)	0.38	0.25	1	4	0.047	++	++	-	-
CD-19-00513	manure	Farm 1	3 (RT001 RT001)	0.75	0.75	0.5	4	0.047	++	++	-	-
CD-20-00542	fert. soil wk. 143	field experiment	3 (RT001 RT001)	0.75	0.75	0.75	6	0.047	++	++	-	-

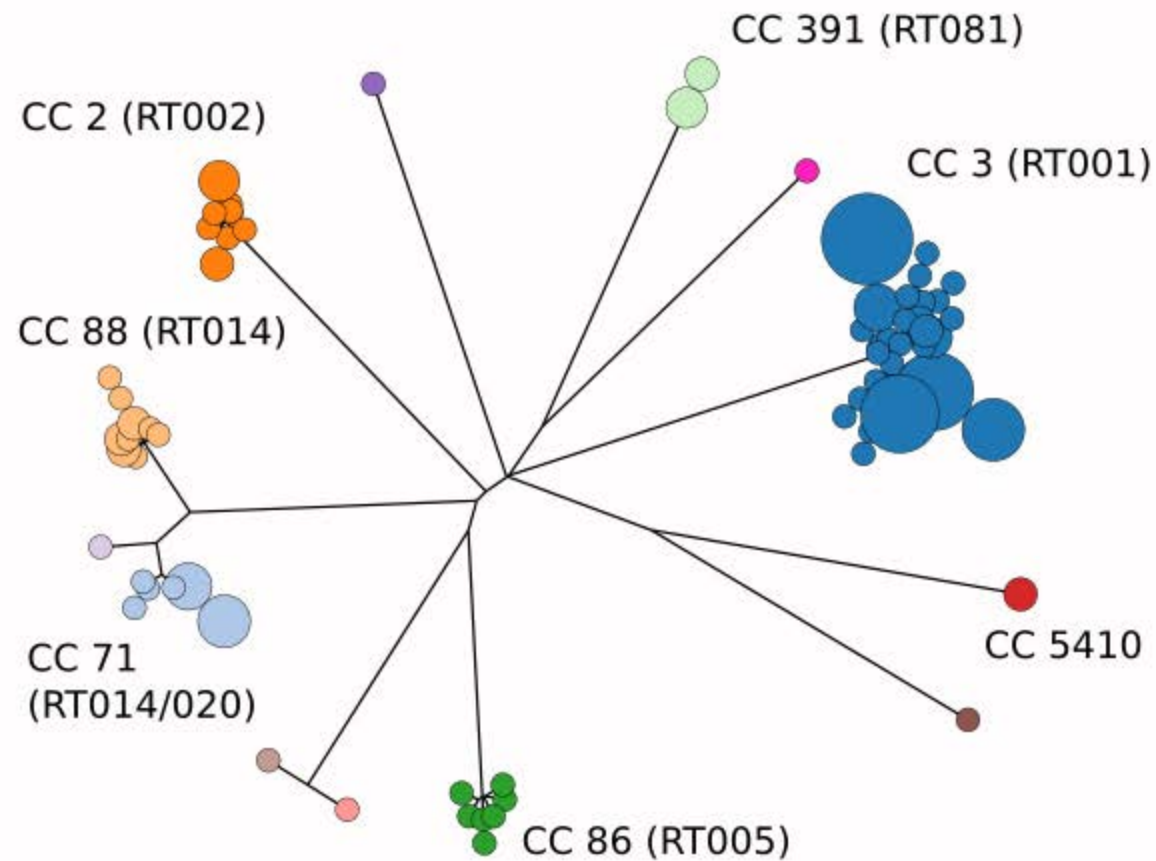
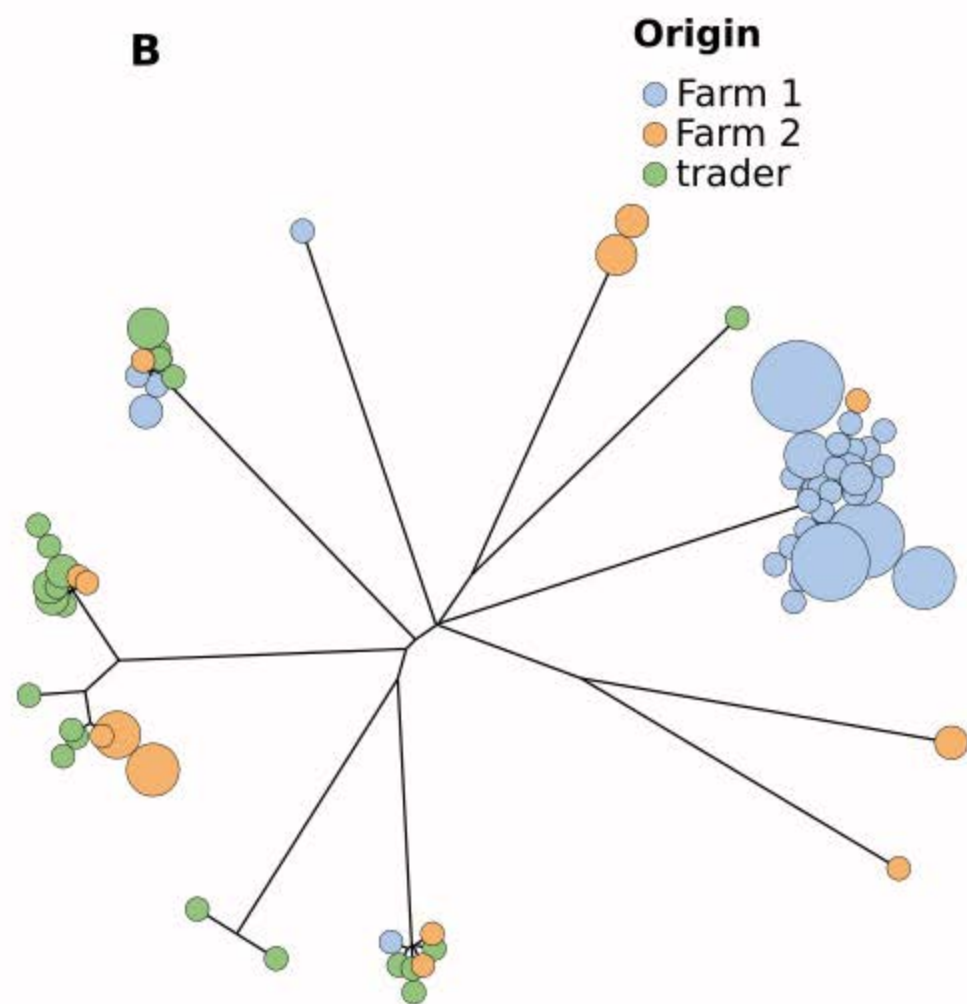
* Ribotypes were predicted based on hierarchical clustering in EnteroBase

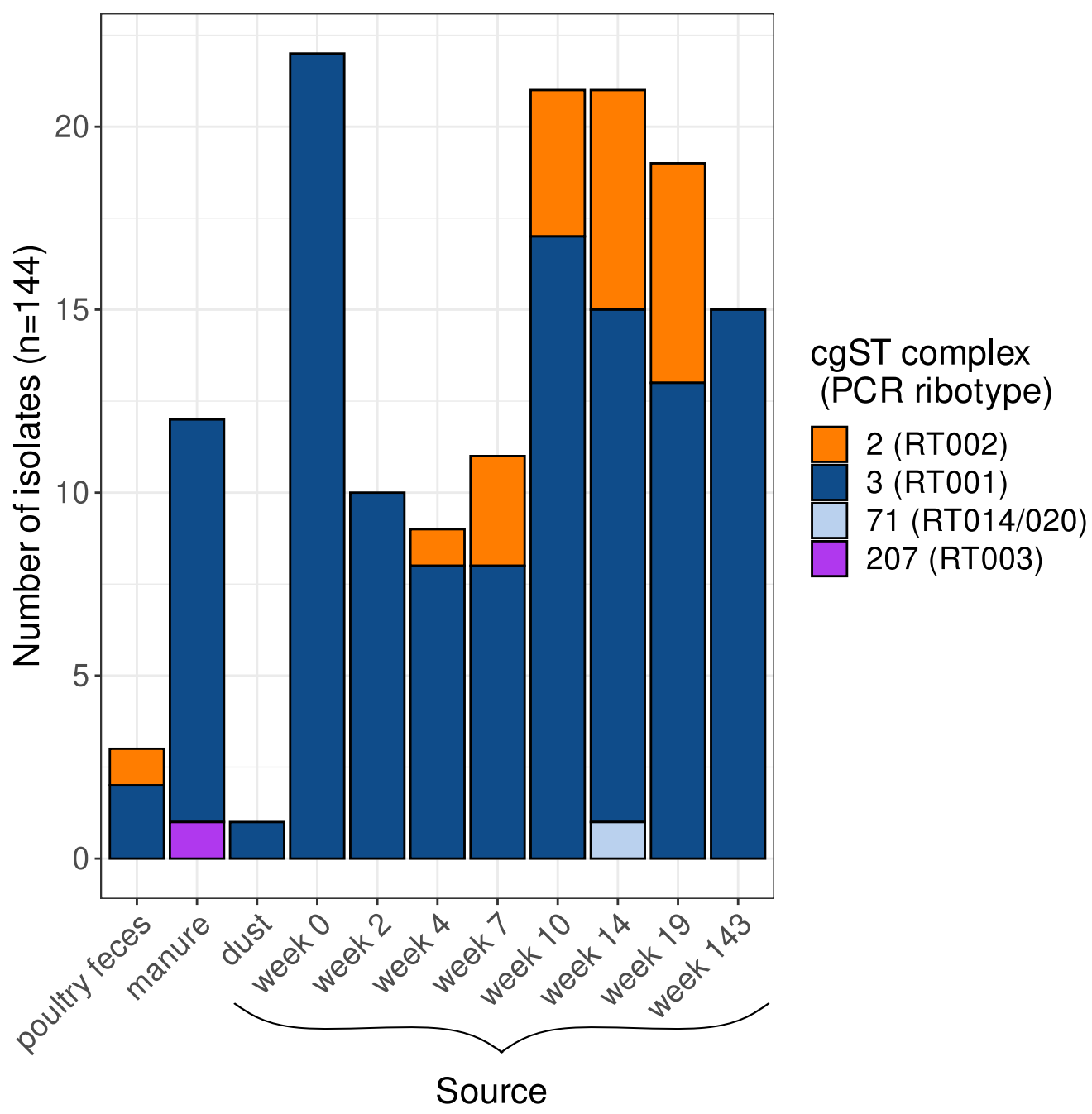
** VAN: vancomycin; MTZ: metronidazole; MXF: moxifloxacin; CLI: clindamycin; TET: tetracycline; *tcdA*: gene encoding for toxin A; *tcdB*: gene encoding for toxin B; *cdtA* and *cdtB*: genes encoding for the binary toxin

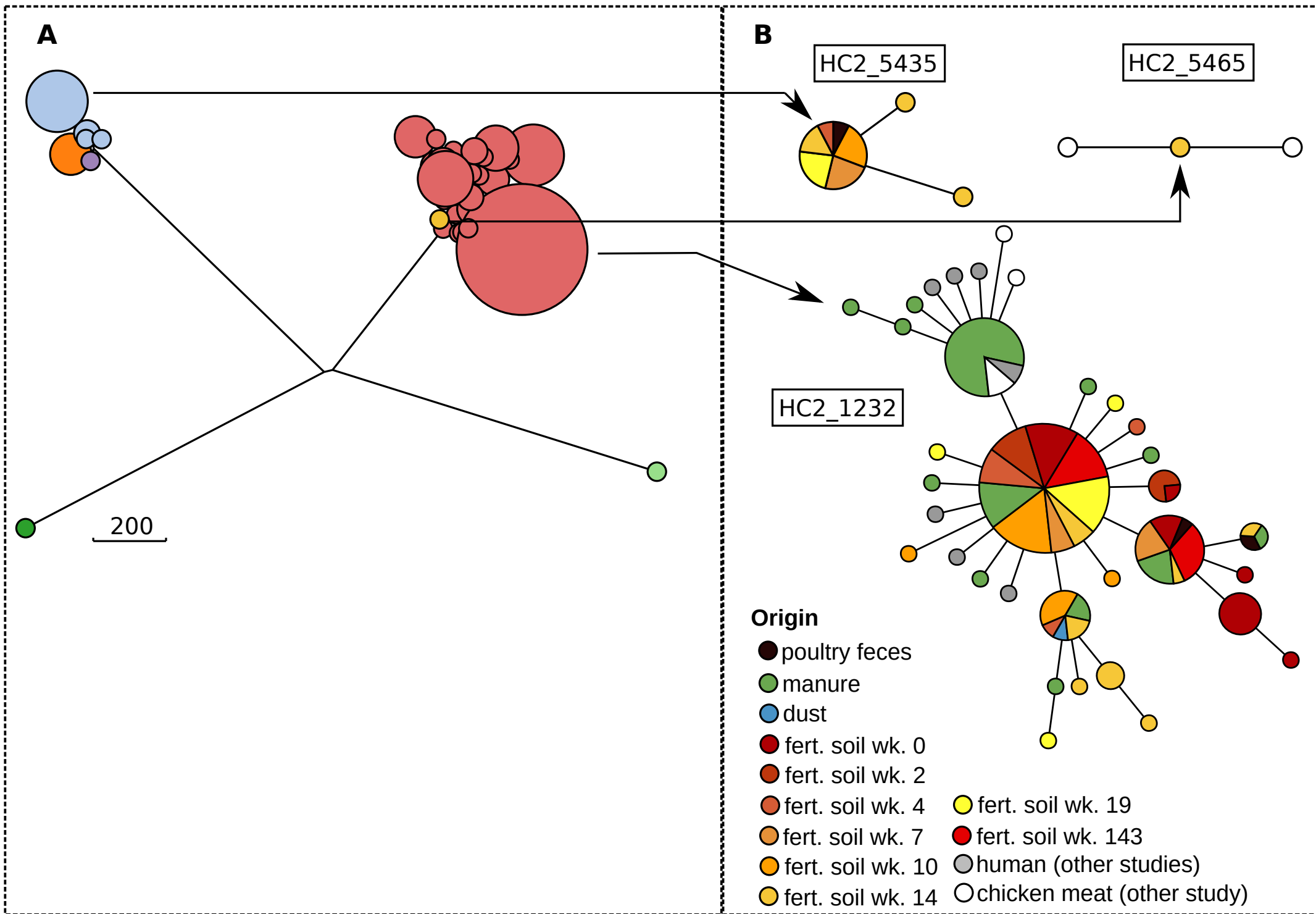
*** Toxin gene prediction was based on corresponding wgMLST loci (+=present; -=absent)

Table 3. HC2 clusters

Source	HC2_1232	HC2_4410	HC2_5435	HC2_5465	HC2_12193	HC2_12207	HC2_12213
poultry feces	2	0	1	0	0	0	0
manure	11	1	0	0	0	0	0
dust	1	0	0	0	0	0	0
fert. soil wk. 0	22	0	0	0	0	0	0
fert. soil wk. 2	10	0	0	0	0	0	0
fert. soil wk. 4	8	0	1	0	0	0	0
fert. soil wk. 7	8	0	3	0	0	0	0
fert. soil wk. 10	17	0	3	0	1	0	0
fert. soil wk. 14	13	0	4	1	2	0	1
fert. soil wk. 19	13	0	3	0	2	1	0
fert. soil wk. 143	15	0	0	0	0	0	0

A**B**



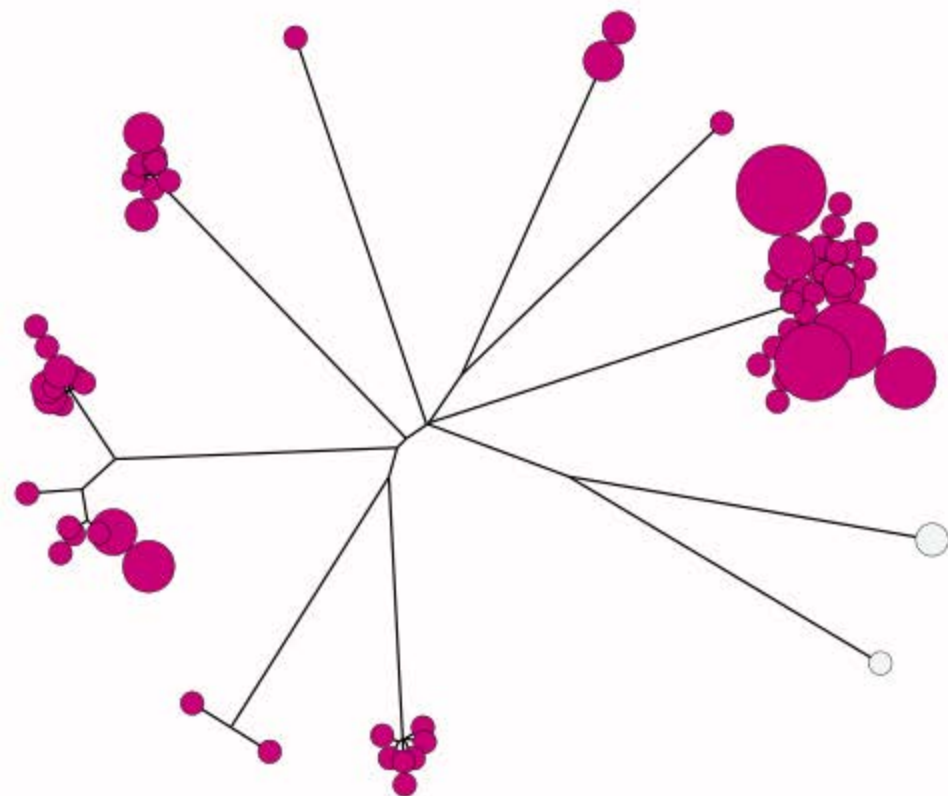


tcdA/tcdB genes

● present

○ absent

A



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Sampling timepoints

● November2017

● March2017

● August2017

● May2017

● May2018

B

