

1 **Limiting the loss of terrestrial ecosystems to safeguard nature for biodiversity** 2 **and humanity**

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- 32 **Keywords:** Biodiversity; Convention on Biological Diversity; Framework Convention on Climate
33 Change; Convention to Combat Desertification; ecosystems; post-2020; retention; Sustainable
34 Development Goals; targets

35 **ABSTRACT**

36 Humanity is on a pathway of unsustainable loss of the natural systems upon which we, and all life,
37 rely. To date, global efforts to achieve internationally-agreed goals to reduce carbon emissions, halt
38 biodiversity loss, and retain essential ecosystem services, have been poorly integrated. However,
39 these different goals all rely on preserving natural ecosystems. Here, we show how to unify these
40 goals by empirically deriving spatially-explicit, quantitative area-based targets for the retention of
41 natural terrestrial ecosystems. We found that at least 67 million km² of Earth's natural terrestrial
42 ecosystems (~79% of the area remaining) require retention – via a combination of strict protection
43 but more prominently through sustainably managed land use regimes complemented by restoration
44 actions – to contribute to biodiversity, climate, soil and freshwater objectives under four United
45 Nations' Resolutions. This equates to retaining natural ecosystems across ~50% of the total
46 terrestrial (excluding Antarctica) surface of Earth. Our results show where retention efforts could be
47 focussed to contribute to multiple goals simultaneously. The retention targets concept that we
48 present explicitly recognises that such management can and should co-occur alongside and be
49 driven by the people who live in and rely on places where natural and semi-natural ecosystems
50 remain on Earth.

51

52 **INTRODUCTION**

53 Despite the dependence of humanity's wellbeing on the natural world, we continue to erode nature,
54 often irreversibly (IPBES 2019). While commitments to 'sustainability' abound, and have been at the
55 core of international agreements since the seminal 'Rio Declaration on Environment and
56 Development' almost 30 years ago (United Nations 1992), humans continue to degrade and destroy
57 ecosystems at unsustainable rates in many parts of the world (Mackey et al. 2015; Watson et al.
58 2016; Thomas et al. 2017). This loss is accompanied by the finality of species extinction (Baisero et
59 al. 2020), the erosion of ecosystem function and evolutionary processes (Watson et al. 2018), the
60 loss of connection between people and nature (Ives et al. 2018), poorer-quality fresh water
61 (Mapulanga & Naito 2019), loss of soil resources (Banwart 2011), depressed yields from living
62 natural resources (Spera et al. 2020), and increased harm to people and nature from climate change
63 (Maxwell et al. 2019). We lack a clear understanding of how much further biodiversity loss can occur
64 before permanent, irrevocable damage is wrought on our life-support system (Maron et al. 2018a).
65 However, evidence suggests we are at imminent risk of breaching (or have already exceeded) certain
66 critical planetary thresholds (Steffen et al. 2015; Lenton et al. 2019). Given that the pressures on

67 natural systems are accelerating as the human population grows and consumption intensifies
68 (Leclère et al. 2020; Wiedmann et al. 2020), it is of urgent importance that we set a limit to the loss
69 of natural ecosystems, to prevent further, irreparable damage.

70 The goals enshrined in multiple global agreements quite clearly depend upon retaining existing (and
71 restored) natural and semi-natural ecosystems. For example, the global community's aspirations
72 regarding biodiversity conservation are comprehensively reflected in the stated objectives of the
73 Convention on Biological Diversity (CBD), as well as various other 'biodiversity-related conventions'
74 (e.g. CITES, Convention on the Conservation of Migratory Species of Wild Animals, World Heritage
75 Convention, Convention on Wetlands (Ramsar)). Our reliance on, and need to conserve nature is
76 further enshrined in the UN Sustainable Development Goals (e.g. Goal 15 "protect, restore and
77 promote sustainable use of terrestrial ecosystems, sustainably manage forests, combat
78 desertification, and halt and reverse land degradation and halt biodiversity loss"; also, Goals 6 (clean
79 water), 14 (marine environment), and others). However, the targets underpinning the achievement
80 of the goals captured in these agreements rarely articulate a desired, measurable outcome state –
81 the amount of nature, where, and managed in what way, that is required for us to achieve particular
82 biodiversity and sustainability objectives (Butchart et al. 2016; Elder & Olsen 2019).

83 Notably, dialogue relating to the proposed Post-2020 Global Biodiversity Framework under the CBD
84 hints at an important shift, with an emphasis on *net* outcomes for biodiversity; namely, net
85 improvements in ecosystems to 2050, delivered via an interim goal to increase the area, connectivity
86 and integrity of natural ecosystems by at least 5% by 2030 (Secretariat of the Convention on
87 Biological Diversity 2020). In large part, this goal's achievement will necessitate limiting further
88 losses of ecosystems – in other words, retaining the vast majority of what we have now. Some
89 ongoing natural ecosystem depletion is inevitable to meet economic and social development
90 imperatives (Zeng et al. 2020), underscoring the vital role of restoration activities to counterbalance
91 losses using protocols such as offsets/ecological compensation (Simmonds et al. 2020). However,
92 noting that restoration entails substantial time lags (Crouzeilles et al. 2016), is unfeasible or has
93 uncertain outcomes in many instances (Gann et al. 2019), and that losses caused by pervasive
94 threats (degradation by invasive species, illegal activities, etc.) rarely trigger compensation
95 requirements (Maron et al. 2018b), a foundation of ecosystem retention is critical for achieving this
96 ambitious outcome of a net increase in ecosystems extent (and net improvement in condition) by
97 2030.

98 International agendas that embed goals/targets which explicitly or implicitly require the retention of
99 ecosystems have been largely implemented in isolation from one another, even with the emergence
100 of the umbrella of the UN Sustainable Development Goals (CBD Subsidiary Body on Scientific
101 Technical and Technological Advice 2019). This separation has predominated despite the fact that
102 the requisite policies and on-ground actions needed to contribute to the achievement of the
103 goals/targets captured among these conventions are likely to align in many instances (Secretariat of
104 the Convention on Biological Diversity 2018; Jung et al. 2020; Soto-Navarro et al. 2020). The
105 inefficiencies of this status quo – in terms of both missed opportunities for synergies, and the
106 unification of agendas underpinned by a common response to managing nature including the
107 retention of ecosystems – are now being recognised (CBD Subsidiary Body on Scientific Technical
108 and Technological Advice 2019). Together with the opportunities presented by a move towards
109 ambitious net outcomes-based goals in a post-2020 world (Bull et al. 2020; Secretariat of the
110 Convention on Biological Diversity 2020), integrated delivery of various international agreements
111 underscore the key role that the retention of ecosystems, framed by spatially-explicit ecosystem
112 retention targets, could play in safeguarding our natural assets globally. Identification of retention
113 targets – limits to loss – that set out what nature we need, and where, to achieve the full suite of
114 goals associated with Earth’s remaining natural terrestrial ecosystems, has the potential to unite
115 nature conservation and sustainable development goals across the international environmental
116 agenda (Maron et al. 2018a).

117 Here, we present a method for quantifying ecosystem retention targets that are needed to achieve
118 the nature-reliant ambitions of global policy agreements relating to biodiversity conservation,
119 climate stabilisation, soil maintenance and water quality regulation. We set out an analytical
120 approach for translating each of these goals to spatially-explicit guidance on how much and where
121 existing natural terrestrial ecosystems should be retained to give us the best chance at meeting
122 these various global objectives. We examine appropriate mechanisms for the management of these
123 ecosystems, noting that a combination of strict protection and sustainable use will be required to
124 align environmental, social and economic imperatives, rights and responsibilities. Further, we
125 suggest that international cooperation will be crucial for determining country-level contributions to
126 global ecosystem retention efforts.

127

128 **METHODS**

129 **Global goals dependent upon natural terrestrial ecosystem retention**

130 There are four current global agreements made under UN Resolutions for which natural terrestrial
131 ecosystem retention is directly relevant: the Convention on Biological Diversity (CBD), the
132 Framework Convention on Climate Change (UNFCCC), the Convention to Combat Desertification
133 (UNCCD), and the Sustainable Development Goals (SDGs). These international agreements, each
134 adopted and ratified by the vast majority of nations, contain clearly-articulated statements about
135 what we expect from natural ecosystems, and are the logical starting point to mapping out how
136 much terrestrial nature we need to retain, if we are to achieve each of these agreements' goals.

137 We examined the goals, objectives, targets, and/or indicators under each of the four agreements to
138 identify statements about desired outcomes that specifically depend upon (in whole or part) the
139 retention of natural terrestrial ecosystems. This formed the basis of translating explicit quantifiable
140 milestones and/or aspirational goals noted in the respective agreements into terrestrial ecosystem
141 retention targets (Figure 1). Hereafter, we refer to the retention targets examined here that
142 correspond to the four respective agreements as the 'biodiversity conservation target', 'carbon
143 storage target', 'soil maintenance target' and 'freshwater quality target'.

144 In this analysis, we only considered areas for retention that are characterised by natural vegetation.
145 While we refer here and throughout to ecosystems, we note that our maps select areas of natural
146 land cover (albeit, along a spectrum of condition/human influence) categorised into broad classes,
147 which are not of sufficiently high resolution to represent the full gamut of terrestrial ecosystems. We
148 use the term 'natural ecosystems' throughout to underscore that this analysis, and the retention
149 targets concept, is focussed on identifying and securing areas that are vegetated, and by proxy,
150 support (or have potential to support) the suite of interacting and interconnected biotic and abiotic
151 attributes (species assemblages and processes) that comprise ecosystems (notwithstanding issues of
152 degradation – see Discussion). Importantly, the retention targets we propose here should be
153 interpreted and considered in the context of ecosystem-specific targets, such as those being
154 proposed for the Post-2020 Global Biodiversity Framework (Watson et al. 2020) and other
155 complementary approaches to target-setting and planning (Dinerstein et al. 2020).

156 We did not include barren areas in our definition of natural ecosystems (e.g. ecosystems
157 characterised by bare sand, exposed rock, ice and snow – including all of Antarctica). We
158 acknowledge the value of these natural, largely non-vegetated environments, especially for
159 biodiversity, but have excluded them here as this analysis is centred on one of the key actions
160 needed to achieve multiple global goals – the retention of existing natural vegetation. Further, our
161 analysis has a terrestrial focus, and does not extend to marine systems. Nonetheless, the framework

162 we present for translating internationally-agreed goals to spatially-explicit quantitative ecosystem
 163 retention targets could be transferable to the marine realm. Amenable datasets may include marine
 164 ecosystems (Spalding et al. 2007; Spalding et al. 2012), human pressures (Halpern et al. 2015),
 165 carbon export in oceans (Henson et al. 2012; Roshan & DeVries 2017), and species ranges and
 166 protected area (various datasets produced by IUCN, BirdLife International, UNEP-WCMC; taxa-
 167 specific data (Kaschner et al. 2011)).

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169

170

171 **Figure 1. Objectives and associated targets/indicators that require retention of natural terrestrial**
 172 **vegetated ecosystems under international agreements. The basis for each agreement’s translation**

173 **to the quantitative, spatially-explicit targets for the retention of natural terrestrial vegetated**
174 **ecosystems that we examine, linked to the objectives/targets of each respective agreement, is**
175 **outlined in the central part of the figure. A representation of the ‘apex retention target’ is also**
176 **presented – this is made up of all the areas of natural terrestrial vegetated ecosystems that are**
177 **captured by one or more of the four individual retention targets, and describes the amount of**
178 **retention as a percentage of the total remaining terrestrial vegetated ecosystems on Earth**
179 **(adapted from Maron et al. (2018a)).**

180 **Quantifying terrestrial ecosystem retention requirements**

181 For each of the four retention targets – biodiversity conservation, carbon storage, soil maintenance,
182 and freshwater quality – we produced maps that captured the amount and location of existing
183 natural terrestrial vegetated ecosystems that need to be retained to contribute to the achievement
184 of internationally-agreed goals. To represent natural ecosystems in this analysis (hereafter, we refer
185 to this as the ‘natural ecosystems layer’), we used the MODIS Land Cover Type product MCD12Q1 at
186 250 m (native data resolution = 500 m) that was processed and modified by Borrelli et al. (2017).
187 This layer, covering approximately 84% of Earth’s surface, is based on the International Geosphere
188 Biosphere Programme (IGBP) system and reports seventeen land cover classes, including ten natural
189 terrestrial vegetation classes (Supporting Information Table S1). We excluded non-vegetated and/or
190 aquatic land cover types from our analysis (permanent wetlands, barren, snow/ice, water). Further,
191 the three developed and mosaicked land classes (croplands, urban and built up, cropland/natural
192 vegetation mosaics) were excluded from the analysis. For the ‘grasslands’ ecosystem category, we
193 masked out grazing lands for the year 2000 using a spatial dataset that combines agricultural census
194 data with satellite-derived land cover to map pasture extent (Ramankutty et al. 2008). The map we
195 produced indicated that approximately 83.8 million km² of natural terrestrial vegetated ecosystems
196 remain on Earth (approximately 62% of the non-Antarctic land surface; noting that this does not
197 account for the condition of this vegetation).

198 To determine how much and where natural terrestrial vegetated ecosystems need to be retained,
199 we intersected existing mapping products linked to each of the four targets (described below) with
200 the natural ecosystems layer. This allowed us to calculate the amount and location of natural
201 terrestrial vegetated ecosystems that are spatially congruent with places recognised as having
202 importance for each of the targets considered in this analysis. All spatial statistics were calculated
203 using a Mollweide projection, with all mapping analyses being undertaken at a raster pixel resolution

204 of approximately 250 m (i.e. at the resolution of natural ecosystems layer). An overview of the
205 workflow for deducing spatially-explicit retention targets is presented in Figure 2.

206 *Biodiversity conservation target*

207 Language of the draft Post-2020 Global Biodiversity Framework under the CBD (August 2020), and
208 the so-called Aichi Targets that it will supersede, is focussed on preventing extinctions, recovering
209 threatened species, expanding the protection and management of important sites, and
210 maintaining/enhancing ecosystem extent and resilience (including connectivity and intactness)
211 (Secretariat of the Convention on Biological Diversity 2020). To translate these various aims to a
212 biodiversity conservation retention target, we used the map produced by Allan et al. (2019), to
213 identify natural terrestrial vegetated ecosystems that should be retained for multiple biodiversity
214 conservation objectives (Figure 1). These maps assessed the spatial distribution and areal
215 requirements needed to conserve 28,594 species, whilst also accounting for existing ecoregional
216 representation targets and protected areas, Key Biodiversity Areas, and the maintenance of all large,
217 contiguous areas with low human pressure ('wilderness'). Detailed methods including the input data
218 used to produce this map, and limitations on its interpretation, are provided Allan et al. (2019). We
219 identified all natural terrestrial vegetated ecosystems from our natural ecosystems layer that are
220 overlapped by this map, to identify the extent and distribution of natural ecosystems that require
221 retention to contribute to the achievement of the biodiversity conservation objectives captured by
222 the Allan et al. (2019) map. We assume that natural terrestrial vegetation retention in these places is
223 integral to the persistence of biodiversity as represented by the Allan et al. (2019) analysis.

224 *Carbon storage target*

225 Forests play a key role in storing carbon. To determine how much and where natural ecosystems
226 should be retained to stabilise levels of atmospheric carbon, we used spatial outputs from a
227 harmonized land-use transition model for Shared Socioeconomic Pathway (SSP) 1 – a globally-
228 accepted sustainable development pathway that respects environmental boundaries (van Vuuren et
229 al. 2017). Spatial outputs for SSP1 are based on the IMAGE 3.0 integrated assessment model (IAM)
230 (Stehfest et al. 2014) and describe annual transitions in different land-use categories (e.g. primary
231 forest, secondary forest, pasture and cropland) between the years 1500 and 2100 at a spatial
232 resolution of 0.5 x 0.5 degrees (approximately 25 km² at the equator). The land use transitions
233 capture subsequent effects on energy, water, and carbon exchanges between the land surface and
234 the atmosphere, and are thus able to track the consequences of land use transitions on the global
235 climate and carbon cycles. We chose SSP1 (of the five SSPs available (Riahi et al. 2017)) because the

236 land-use transitions captured in this scenario have the greatest probability of limiting global
237 warming to less than 2°C above pre-industrial levels (Hurtt et al. 2020).

238 We used spatial outputs for SSP1 to render a map of carbon retention through *primary forests*
239 (defined as forested land that was not converted to an alternative land use type – urban, cropland,
240 grazing land or secondary forest – between 1500 and 2050 according to the IMAGE 3.0 IAM (Hurtt et
241 al. 2020); see below for explanation of focus on primary forest). The state of primary forest land in
242 2050 was subtracted from the state of primary forest land in 2015. The resulting index shows, for
243 each 25 km² grid cell, the proportion of primary forest required to be retained if forests are to
244 effectively help stabilise levels of atmospheric carbon. We note that the index ascribes each 25 km²
245 grid cell a proportional retention value – that is, grid cells classed as 100% are not necessarily 100%
246 covered by primary forest, but instead contain some primary forest in that cell, 100% of which must
247 be retained to achieve the carbon storage target.

248 Once we identified the proportion of forested areas to be retained in each 25 km² cell, we
249 overlapped this with treed natural areas (forests, categories 1-5) from our natural ecosystems layer
250 (Supporting Information Table S1). To allocate the spatial distribution of forest to be retained from
251 this layer, we preferentially selected pixels starting with those with the highest coverage of trees
252 based on the MOD44B Continuous Fields layer (an input into the MODIS-derived ecosystem layer
253 used in this analysis). This is a 250 m spatial resolution biophysical parameter derived from the
254 MODIS satellite. It reports annual estimates of the percentages of 1) surface vegetation cover, 2)
255 bare soil, and 3) tree cover. The rationale for selecting forest pixels in this way was that pixels with a
256 higher coverage of trees are more likely to be representative of primary forest; sparser areas on the
257 other hand may be representative of more open (treed) ecosystem types or degraded forests. We
258 continued this process until the area of forested pixels selected corresponded with the proportional
259 retention value identified by the index of forest retention (to 2050) produced from SSP1 output at a
260 25 km² grid cell resolution. To explore the implications of this for calculation of our ‘headline’
261 retention target, we repeated the sampling of forest pixels for the carbon goal preferentially
262 selecting forest pixels that were also captured by at least one other of the biodiversity conservation,
263 soil maintenance or freshwater quality target in this analysis. We found that the difference between
264 the methods only considering the pixels with higher coverage of tree cover versus preferentially
265 selecting forest pixels was less than 200,000 km², which corresponds to less than 0.15% of the global
266 area analysed.

267 Unlike the biodiversity, freshwater and soil goals in our analysis that span various natural terrestrial
268 ecosystem types, our goal for carbon retention focuses only on primary forest ecosystems. This
269 means that our representation of carbon retention will be an underestimate, particularly in parts of
270 the world where other (non-forest ecosystems) play a key role in carbon storage. We note that the
271 management (and retention) of other terrestrial ecosystem types (e.g. peatlands, mangroves,
272 grasslands) can greatly benefit global efforts to mitigate climate change (Goldstein et al. 2020).
273 However, we chose to focus on forested ecosystems as they contain the majority of terrestrial
274 carbon stocks on Earth (Goldstein et al. 2020) and the spatial resolution of the study (approx. 25km²)
275 precluded assessment of some carbon-dense ecosystems that have narrow or patchy geographic
276 distributions (e.g. some mangrove or peatland ecosystems). Furthermore, we focused on primary
277 forests as their effective management confers a multitude of co-benefits for other social and
278 environmental values (Watson et al. 2018), and effectively managing carbon stocks in degraded
279 forests (e.g. forests that are logged or used for livestock grazing) requires a complex set of
280 management interventions that are beyond the scope of this study to capture.

281 The way we have identified forests for retention for the carbon goal should also be considered when
282 interpreting the maps produced in this study – we reiterate that this study is not intended to be a
283 high-resolution spatial prioritisation exercise to identify exactly where on the Earth’s surface native
284 ecosystems need to be retained – rather, these maps are illustrative of the general location of where
285 these ecosystems occur across the globe, noting that resolution mismatches of input layers
286 prevented us from mapping forest retention for the carbon goal with the same degree of spatial
287 precision as was the case for the other targets.

288 *Soil maintenance target*

289 The retention of natural ecosystems in areas where their removal would result in unsustainable
290 rates of soil loss is crucial. Although several threats to land-based natural capital and soil quality are
291 globally significant (salinization, acidification, nutrient depletion, contamination, waterlogging, etc.),
292 for this analysis we focussed on soil erosion because it is a clear sign of land degradation that
293 constitutes an irreplaceable loss and that cannot reasonably be restored. Indeed, soil erosion is the
294 most immediate risk to land-based natural capital if natural vegetation is cleared.

295 Current global rates of soil loss by water erosion have been estimated by Borrelli et al. (2017), using
296 the RUSLE-based (Revised Universal Soil Loss Equation) (Renard et al. 1997) modelling platform
297 Global Soil Erosion Modelling (GloSEM). RUSLE incorporates rainfall erosivity and climate, erodibility
298 of the soil, topography, and local farming systems and practices to predict the amount of soil lost

299 (due to inter-rill and rill erosion processes) per unit area and time ($\text{t ha}^{-1}\text{yr}^{-1}$) (Renard et al. 1997).
300 Using data made available by Borrelli et al. (2017), we modelled the likely rate of soil loss in tons per
301 ha per year if mapped natural terrestrial vegetated ecosystems were cleared. The possible new land
302 use for 3,252 sub-national administrative units of 202 countries from the Global Administrative Unit
303 Layers (United Nations Food and Agriculture Organization 2015), after clearing for agriculture, was
304 either grazing or cropping depending on the main land use in the region according to Ramankutty et
305 al. (2008). We assumed that administrative units with less than 50% coverage of grazing would be
306 converted to the dominant current crop for that unit, whereas areas with more than 50% of
307 coverage of grazing would be converted to grazed areas. The dominant crop per country was
308 identified using data from the UN Food and Agriculture Organization (United Nations Food and
309 Agriculture Organization 2016). Based on this land cover change, we then assigned values that
310 measure the effect of cropping on the soil erosion process (C-Factors), based on thresholds
311 identified by Borrelli et al. (2017). All grazing areas had the highest C-Factor values (0.5).

312 Sustainable or tolerable (Verheijen et al. 2009) soil erosion rates can be defined as rates of soil
313 erosion that are equivalent to rates of soil formation consisting of mineral weathering as well as dust
314 deposition. Soil formation rates vary widely across the globe depending on climate, geology,
315 topography, and other factors (García-Ruiz et al. 2015). A global average soil formation rate of 2 t ha^{-1}
316 yr^{-1} is an appropriate upper estimate of formation rates, based on calculation of rock weathering
317 rates from exported water chemistry globally (Wakatsuki & Rasyidin 1992; Renard et al. 1997;
318 Panagos et al. 2015). Using our map of areas where erosion rates were predicted to exceed 2 t ha^{-1}
319 yr^{-1} if natural terrestrial vegetated ecosystems were replaced by agricultural land use, we identified
320 areas from our natural ecosystems layer that should be retained to avoid unsustainable erosion, and
321 thus contribute to land degradation neutrality goals.

322 *Freshwater quality target*

323 We produced a map of natural terrestrial vegetated ecosystems that should be retained to
324 contribute to freshwater quality maintenance (e.g. via in-situ and catchment-level processes
325 associated with filtration of impurities, reduction of sedimentation and pollutant run-off, etc.), to
326 contribute to key global goals including SDG 6. To do this, we used three different datasets that map
327 semi-aquatic (vegetated) ecosystems, areas of natural freshwater importance globally, and river
328 basins that contribute disproportionately to global freshwater discharge. First, we used the Global
329 Lakes and Wetlands Database 3 (GLWD) (30 Arc second resolution) (Lehner & Döll 2004). This
330 dataset identifies areas that are predominantly water (e.g., lakes, rivers, marshes, flooded forests)

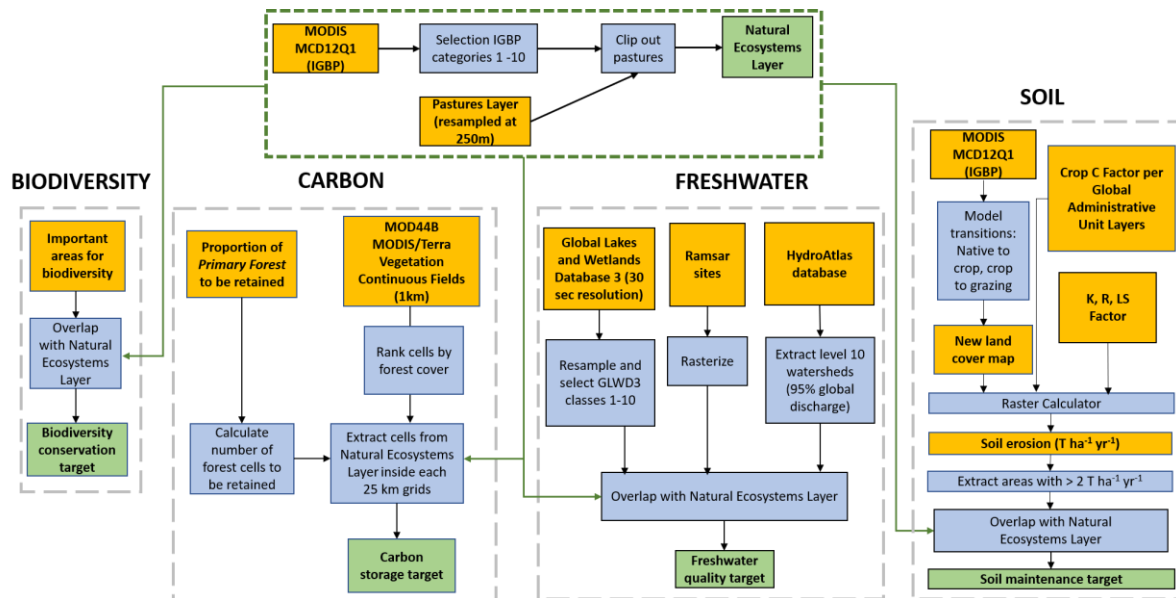
331 and those that are partially water (e.g., intermittent lakes and wetland complexes). Here, we
332 included classes 1-10 from the GLWD 3, but excluded classes 11 and 12 – these are attributed as
333 covering less than 50% of the 30 Arc second raster grid upon which the GLWD 3 is based (Lehner &
334 Döll 2004). To refine our analysis to natural terrestrial vegetated ecosystems associated with
335 (overlapping) large semi-aquatic systems, we did not capture these small (<0.5 km²) ephemeral
336 wetland complexes, noting that some such vegetation may have been captured in the
337 complementary river analysis described below. We overlapped the layer produced using the GLWD 3
338 (classes 1-10) with our natural ecosystems map to identify terrestrial vegetated areas congruent
339 with mapped wetlands. We undertook the same process for mapped (non-estuarine or marine)
340 Ramsar sites, whereby natural ecosystems overlapped by Ramsar sites were selected.

341 To complement this ‘wetlands’ component of our freshwater quality target, we also used the
342 HydroATLAS database developed by Linke et al. (2019) to identify river basins responsible for the
343 majority of the planet’s freshwater discharge. HydroATLAS captures twelve nested levels of sub-
344 basins at the global scale, each depicting consistently sized sub-basin polygons at scales ranging from
345 millions (level 1) to tens of square kilometres (level 12). Using level 10 basins – a reasonable
346 approximation of regional patterns of water discharge – we identified those basins collectively
347 responsible for 95% of global freshwater discharge (approximately 75,000 individual basins). We
348 then selected natural terrestrial vegetated ecosystems within these basins, under the rationale that
349 these natural ecosystems are important for contributing to the regulation of water quality being
350 discharged from these basins. To represent the amount and distribution of natural terrestrial
351 vegetated ecosystems to be retained to contribute to global freshwater quality, we combined the
352 areas of natural ecosystems selected for the GLWD overlay, Ramsar sites and river basin analyses
353 (noting some overlap among the three).

354 **Mapping outputs and analysis**

355 From these analyses, we produced four maps, displaying natural terrestrial vegetated ecosystems
356 that should be retained to contribute to respective biodiversity conservation, carbon storage, soil
357 maintenance and freshwater quality imperatives. On each map, we also showed existing natural
358 ecosystems that were not captured by the respective retention targets. In addition to showing the
359 distribution of natural ecosystem retention, these maps also allowed us to calculate the extent of
360 natural ecosystems required to be retained to contribute to each of the four goals, as well as the
361 percentage of natural ecosystems requiring retention compared to 1) the total extent of natural
362 ecosystems remaining in 2012, and 2) the terrestrial surface of the planet (excluding Antarctica). We

363 combined the four maps into a single global map to calculate an ‘apex retention target’ – in other
 364 words, the total amount (and percentage remaining) of natural terrestrial vegetated ecosystems
 365 that need to be retained to support the achievement of all four goals. To translate these findings to
 366 discrete units, we broke down natural ecosystem retention values by country.



367
 368 **Figure 2. Overview of workflow for deriving maps for each of the four retention targets considered**
 369 **in this analysis. Yellow boxes represent input/derived datasets; Blue boxes represent GIS**
 370 **operations; Green boxes represent outputs. All raster inputs were resampled to match the**
 371 **resolution of the natural ecosystems layer.**

372

373 RESULTS

374 At least 67 million km² (79%) of Earth’s remaining natural terrestrial vegetated ecosystems, covering
 375 approximately half the terrestrial (excluding Antarctica) surface of the planet, should be retained if
 376 internationally-agreed goals associated with biodiversity conservation, carbon storage, soil
 377 maintenance and freshwater quality are to be achieved (Table 1; Figure 3). This value represents the
 378 ‘apex retention target’ for natural terrestrial ecosystems – the amount needed for co-achievement
 379 of various environmentally-based goals. This target sets a limit to loss, and when combined with the
 380 spatial outputs from which it was derived, establishes (broadly) where we should aim to retain

381 natural ecosystems, lest we risk compromising one or more of the environmental goals that nations
 382 of the world are committed to achieving.

383 **Table 1. Area of natural terrestrial vegetated ecosystems to be retained under each target (a); the**
 384 **area required to meet for all four targets (b). The percentage values do not add to 100%, as there**
 385 **are considerable areas of overlap of natural terrestrial vegetated ecosystems requiring retention.**

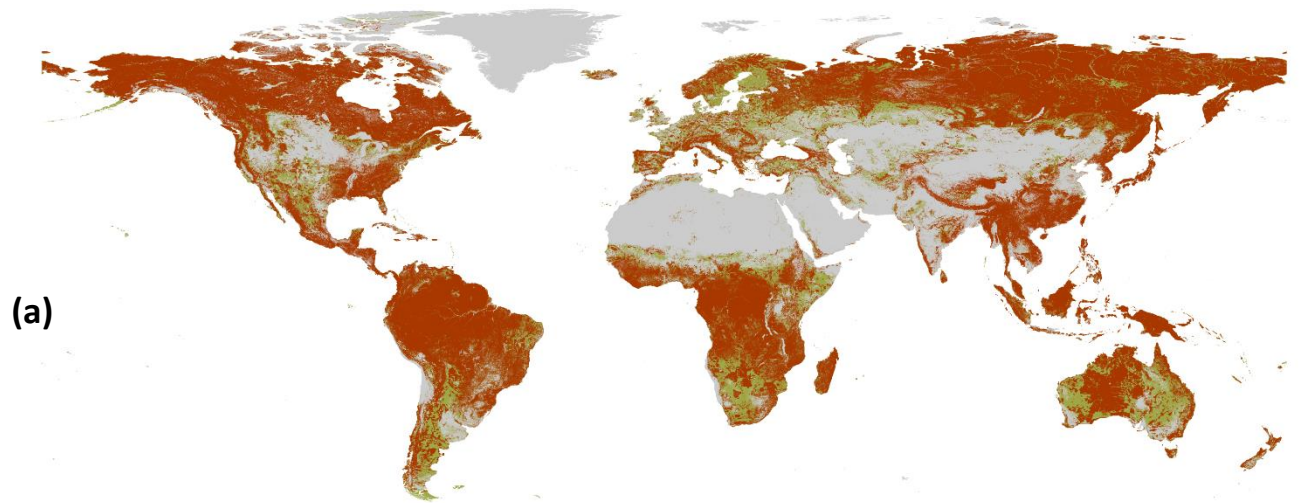
a) Target	Area to be retained (km ²)	Area as a % of remaining natural terrestrial vegetated ecosystems	Area as % of terrestrial Earth surface (excl. Antarctica)
Biodiversity conservation	43,551,955	51%	32%
Carbon storage	20,984,782	25%	15%
Soil maintenance	36,493,119	43%	27%
Freshwater quality	12,133, 826	14%	9%
b) Overlap among targets			
Number of targets contributed to			
1	34,430,437	41%	26%
2	22,314,968	26%	17%
3	9,263,759	11%	7%
4	1,499,633	2%	1%
Total	67,508,798	79%	50%

386

387 Biodiversity conservation required the largest extent of natural ecosystem retention (Table 1; Figure
 388 4), with over 43 million km² (51% of remaining ecosystems) required to contribute to this goal with
 389 its various objectives (species persistence, ecosystem representation, securing important sites such
 390 as existing KBAs, retention of contiguous areas of low human pressure ('wilderness')). There was
 391 relatively low overlap among the spatial distribution of retained ecosystems satisfying multiple goals
 392 – approximately 27% of retained ecosystems contributed to two targets, while only 2% contributed
 393 to all four targets (Table 1).

394

395



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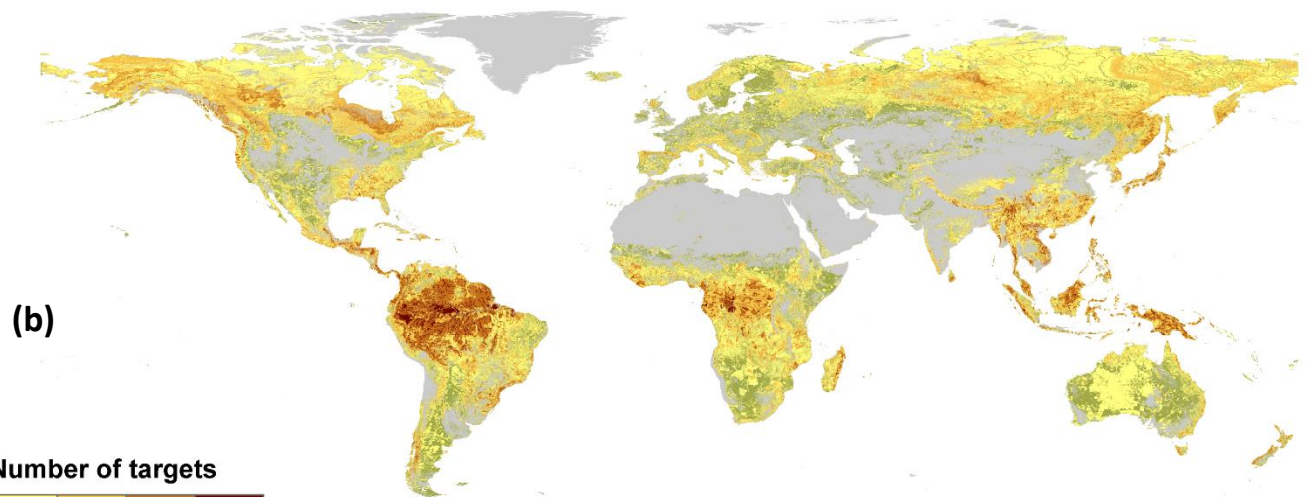
397

■ Natural vegetation critical for at least one target

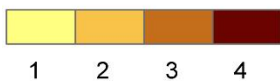
■ Other native vegetation

396

397



Number of targets



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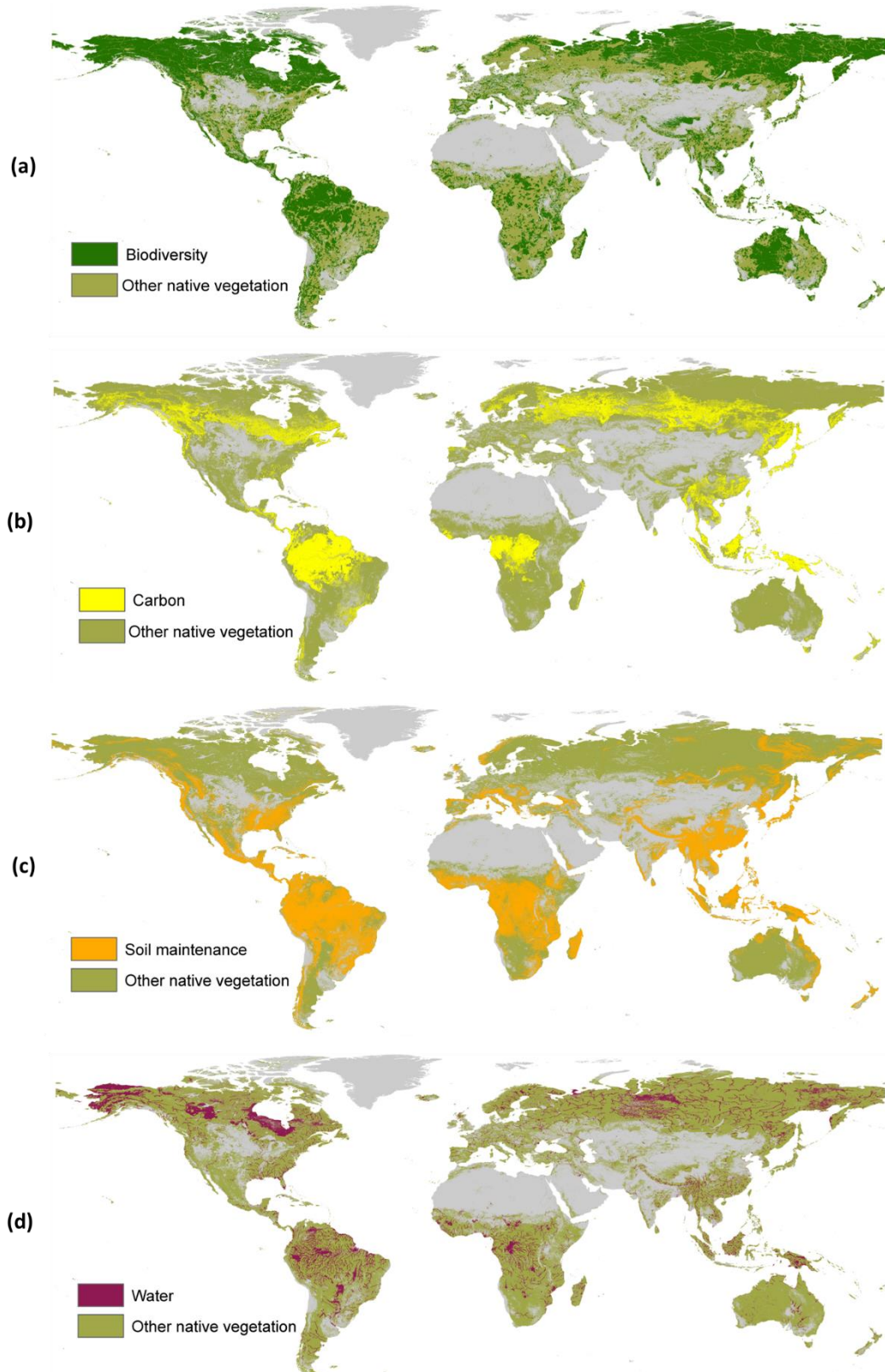
403

404

405

■ Other native vegetation

Figure 3. Natural terrestrial vegetated ecosystems requiring retention to contribute to the achievement of at least one target (a); and overlap of targets (b). Natural terrestrial vegetated ecosystems ('native vegetation') that is less critical for retention under any of the four targets are shown in light green. Grey shading represents areas of the land surface excluded from the analysis - either non-vegetated natural land cover types (e.g. rock, ice, barren), or where natural ecosystems have been replaced by other (anthropogenic) land cover types.

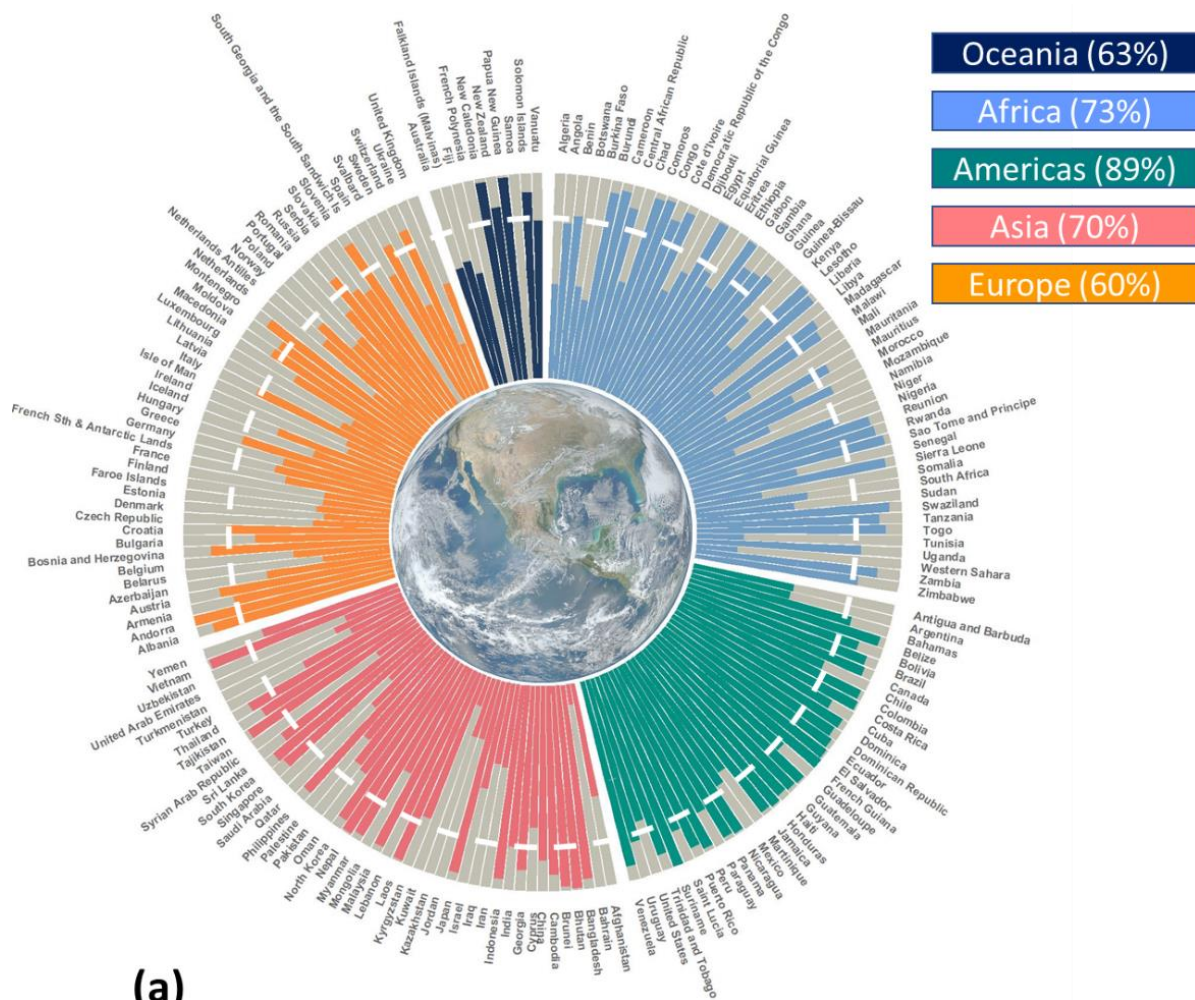


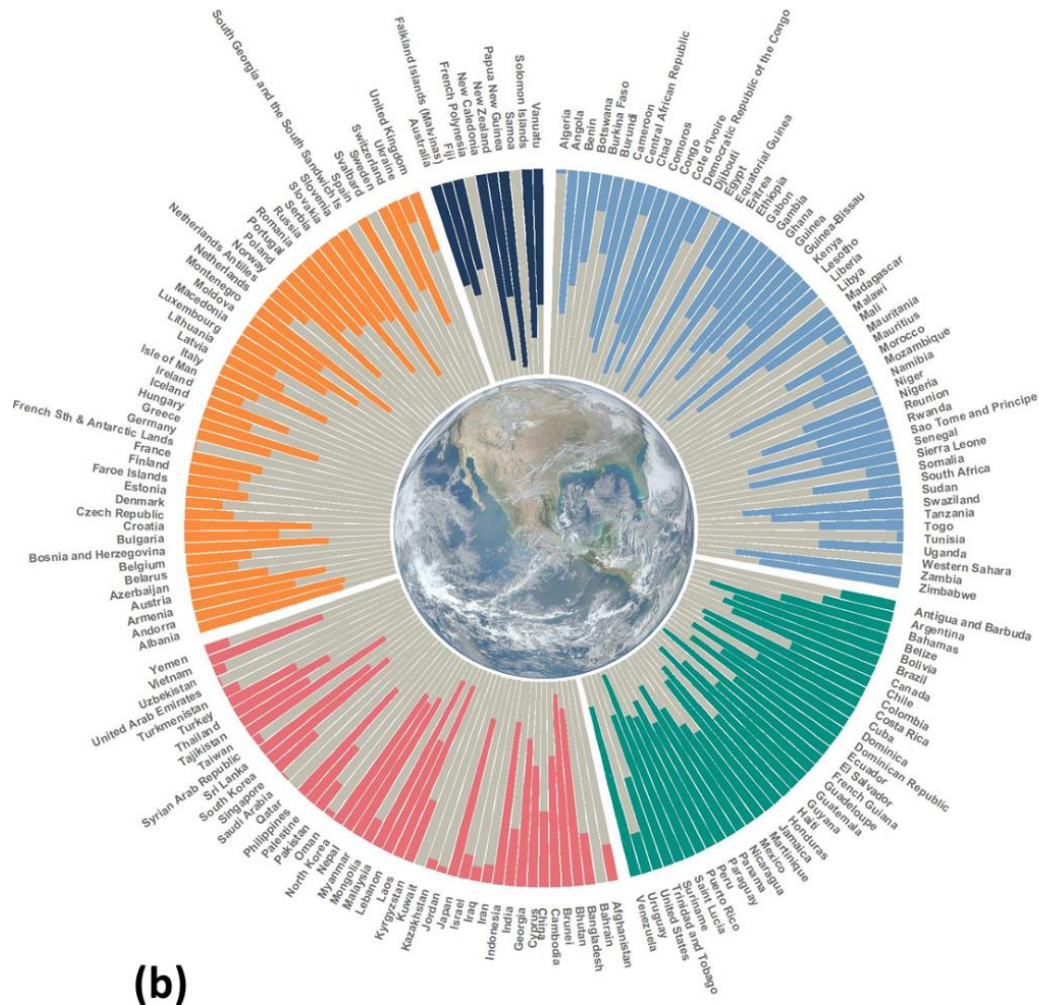
406

407

408 **Figure 4. Natural terrestrial vegetated ecosystems required for retention to meet each of the four**
409 **targets: biodiversity conservation (a); carbon storage (primary forest only) (b); soil maintenance**
410 **(c); and freshwater quality (d). Natural terrestrial vegetated ecosystems ('other native vegetation')**
411 **is shown in light green.**

412 Individual countries differed markedly in their ecosystem retention requirements (Figure 4a-d;
 413 Figure 5). The area for retention required to meet the apex target of 79% is disproportionately
 414 shared among several large countries, with Russia, Canada, Brazil, United States and Australia
 415 accounting for more than half (52%) the extent alone. Of the 66 countries with ecosystem retention
 416 targets of at least 90% of existing natural terrestrial ecosystem extent, the vast majority (n=63) are in
 417 Africa, Asia and the Americas. Imperatives to meet national socio-economic goals are likely to be
 418 particularly acute in many such countries, presenting a conflict between the achievement of global
 419 environmental goals, and development.





(b)

421

422 **Figure 5. Country-level ecosystem retention values (coloured section of bars) as proportion of total**
423 **amount of natural terrestrial ecosystems remaining. The white dashed line represents the apex**
424 **global retention target of 79%. Percentage values in the key are the average natural terrestrial**
425 **ecosystem retention value per continent (a). Amount of natural terrestrial ecosystems requiring**
426 **retention as a percentage of total land area of each country (b). Countries with land area greater**
427 **than 500 km² shown here for display purposes, with all national retention values presented in**
428 **Supporting Information Table S2. Overseas dependencies are classified according to their country**
429 **(and continent) of association.**

430

431 DISCUSSION

432 Our analysis indicates that at least 79% of the remaining extent of natural ecosystems should be
433 retained, with any loss from the areas we identify for retention potentially compromising our ability

434 to achieve globally-agreed environmental goals of humanity. This equates to keeping at least half of
435 the planet under natural vegetation coverage – a striking finding that aligns with other recent
436 analyses that call for greatly increased ambition to conserve at least half of the natural world (e.g.
437 Dinerstein et al. 2020; Wilson 2016). Importantly, achieving such an ambitious outcome (a global
438 retention target of 79%) is not predicated on the exclusion of people from natural and semi-natural
439 landscapes, nor should it compromise development imperatives – indeed, changes in the way we
440 manage and draw from the biosphere can see this ambition become a reality (Tallis et al. 2018). This
441 retention-centred approach to framing and establishing global environmental targets that we
442 present sets a limit to the loss of nature. This is vital in a world where the depletion of natural
443 systems is rampant – for example, over 4 million km² of land was converted to human land uses
444 between 1993 and 2009 (Watson et al. 2016), while up to a million species are threatened with
445 extinction (IPBES 2019).

446 **Unifying multiple goals**

447 The retention targets approach we demonstrate can help achieve multiple goals under a unified
448 approach – so far, an elusive prospect (Scharlemann et al. 2020). It addresses key points of criticism
449 of previous global targets, such as the Aichi Targets (and especially Target 11) under the CBD with
450 their lack of focus on outcomes (Barnes et al. 2018). Moreover, it maps a path towards integrating
451 conservation with the various expectations we have of nature (e.g. service provision), which has
452 been lacking from global endeavours such as those encapsulated in the UN Sustainable
453 Development Goals (Zeng et al. 2020). This is because it points to a scientifically-formulated,
454 spatially-explicit and measurable outcome state, which is aligned with the goals of these various
455 agreements.

456 Our method, founded on the notion of setting a limit to loss via retention targets, offers a
457 complementary approach to various other proposals for setting global environmental targets, in that
458 it establishes what we need (the desired outcome) which can then lay a foundation for how to get
459 there (mechanisms). One such example is the ‘Global Safety Net’, which provides a spatially-explicit
460 representation of where vastly increased conservation efforts need to be concentrated, to achieve
461 complementary biodiversity and carbon objectives (Dinerstein et al. 2020). Based upon a detailed
462 suite of biodiversity and carbon data, the outputs of this analysis provide a scientifically-robust
463 roadmap for future land use planning and decision-making. We concur with Dinerstein et al. (2020)
464 that protected areas will always be the key cornerstone of conservation efforts, but that they alone
465 cannot preserve species, ecosystems and ecological functions (unless the ambition for protected

466 area coverage is dramatically and perhaps implausibly increased). Indeed, our results demonstrate
467 that we need vastly more nature retained than is contained within current protected areas, many of
468 which are only ‘paper parks’ (Di Minin & Toivonen 2015). Even ambitions to increase their coverage
469 will, in the absence of a more holistic framing, potentially fall short of securing all of the ecosystems
470 we must keep for the full gamut of services that nature provides us, as well as providing sufficient
471 space for wild species to thrive, and for evolutionary processes to continue. This underscores the
472 necessity of framing environmental targets in the retention (and restoration of) natural systems
473 across tenures and land use regimes.

474 Interestingly, our results revealed limited overlap between multiple targets in some parts of the
475 world – a likely function of our input datasets being constrained to certain elements of the biota
476 (e.g. ecoregional representation targets for biodiversity, primary forest only for carbon, focus of
477 vegetation retention in subset of river basins). Given that our approach is a conservative protocol for
478 identifying sites for retention, we stress that opportunities for synergies must be carefully
479 considered. Our analysis is not an optimisation, and differs from spatial prioritisations. Jung et al.
480 (2020) used an optimisation approach to identify high priority sites for species conservation, carbon
481 retention and water provisioning. This approach, which identified priority sites but did not establish
482 the amount and spatial distribution of land needed to meet specific outcomes-based targets, is
483 another valuable complement to our retention target maps – especially when it comes to identifying
484 what and where particular actions should be most urgently undertaken. For example, natural
485 ecosystems identified by our study as warranting retention, and by Jung et al. (2020) to be of high
486 priority, might warrant more targeted interventions (strict protection, carefully managed use to
487 allow for provisioning of and access to ecosystem services). Conversely, lower priority sites may be
488 more amenable to sustainably-managed mixed use (e.g. intensification of existing agricultural
489 practice) landscapes, albeit, where the focus is on long-term ecosystem retention (complemented by
490 restoration, which will be especially critical for depleted habitats for many threatened species).

491 **Implementation – opportunities and considerations**

492 As noted elsewhere (Maron et al. 2018a; Dinerstein et al. 2020), achieving retention targets will
493 require a carefully developed mixture of protection and sustainable management of natural and
494 semi-natural vegetation (in parallel with restoration of degraded sites) beyond strictly protected
495 areas, alongside transformative changes in consumption patterns (supply and demand of food and
496 resources) (Leclère et al. 2020). Many pragmatic decisions and some trade-offs will be needed. Well-
497 managed protected areas are long-standing mechanisms for securing ecosystems (Maxwell et al.

498 2020), which can be complemented by newer approaches such as other effective area-based
499 conservation measures (OECMs) (Dudley et al. 2018), including privately held land (Clements et al.
500 2018). The stewardship of Indigenous people and other local communities with connection to the
501 land and knowledge of its management, will be necessary for the preservation of large swathes of
502 natural and semi-natural areas, where such retention is aligned with the aspirations of Indigenous
503 custodians (Garnett et al. 2018; O'Bryan et al. 2020). Security of tenure and access is an essential
504 enabling condition for this stewardship to be equitable and undertaken in a manner that is
505 consistent with local peoples' aspirations. Strategies favouring commercial land uses that retain
506 important natural/semi-natural vegetation offer a further pathway to retention; for example,
507 sustainable forest management (e.g. via certification schemes (Kleinschroth et al. 2019)) or wildlife-
508 sympathetic and nomadic livestock grazing (Fynn et al. 2016; Liao et al. 2020), as compared with
509 conversion to intensive single-crop agriculture.

510 Other effective levers, such as moratoria on conversion of sensitive ecosystems, are increasingly
511 applicable to commercial products and their supply chains (Newton et al. 2013). Jurisdictional law
512 and policies regarding the use and management of natural resources also have a central role to play
513 in framing where and to what extent certain elements of the biota can be utilised (e.g. harvest limits
514 (Di Minin et al. 2019)), and in imposing environmental conditions on permits for certain activities,
515 either through sectorial policies and regulations, or following an environmental impact assessment
516 (EIA) process, which is well-established globally. One of EIAs central tenets, the mitigation hierarchy,
517 is a globally-ubiquitous instrument legislated by governments, and required by many financiers and
518 industries, to minimise the impacts of development and balance environmental harm with
519 equivalent gains elsewhere (zu Ermgassen et al. 2020). Harnessing the corporate sector generally
520 (e.g. through commitments to 'science-based targets' (Andersen et al. 2020)), and the mitigation
521 hierarchy specifically (Simmonds et al. 2020) holds substantial promise for resolving tensions
522 between development and ecosystem conservation.

523 The retention target framing explicitly recognises that much of the nature we need is where we live,
524 work, produce our food and extract resources, and that keeping this nature *in situ* necessitates a
525 multi-faceted approach. However, we note that for any given site for which retention of natural
526 ecosystems is required, the management and governance standards needed will differ depending on
527 the objective to which that place contributes. Where ecosystem retention contributes to multiple
528 goals, the strictest standard required to ensure adequate management should apply. For example, a
529 site for which retention contributes to both biodiversity and carbon goals might be amenable to
530 sustainable production (e.g., grazing of livestock) from a carbon perspective, but this production may

531 unsustainably affect its biodiversity values. Here, targeted management to maintain the biodiversity
532 values would be necessary, extending to strict protection if other land uses were incompatible with
533 conserving the site's biodiversity values.

534 A pitfall of global environmental targets is that they tend to be taken *prima-facie* without a clear
535 understanding of the assumptions and limitations inherent to the target-setting, especially as this
536 applies to their implementation on the ground (Garcia et al. 2020). Ultimately, achieving targets such
537 as the retention concept we present, requires changes in behaviours and decision-making at
538 multiple scales by many different people. In fact, the way we make decisions on the targets, and the
539 role of human agency in this process, is likely more important than the targets themselves (Garcia et
540 al. 2020). In the context of retention targets, this particularly matters as countries (and communities
541 therein, especially Indigenous communities) that support a high proportional extent of ecosystems
542 requiring retention will carry a disproportionate burden when it comes to achieving a global apex
543 target such as the one we present in this analysis. Maron et al. (2020) suggested that the loss of
544 nature in many such countries will profoundly affect local people where there may be a more direct
545 reliance on ecosystem services (clean water access, for example), in the absence of extensive and
546 maintained infrastructure. Despite the local (and broader regional and global benefits) of keeping
547 nature *in situ*, an expectation that such peoples and nations retain the vast majority of their natural
548 ecosystems presents complex challenges, due to issues around power imbalances, governance,
549 capacity and competing (development) aspirations (Maron et al. 2020). Recognising that broad-
550 scale, intensive development is but one of several key pathways to improved socio-economic
551 conditions, the support (finance, capacity and empowerment; distribution of the cost of retention)
552 of other (external) actors may be needed in some instances (but only if requested) to ensure that
553 development can occur sustainably in these places, while not compromising local ecosystem service
554 provision nor global environmental goals (Maron et al. 2020). Meaningful engagement with rights-
555 holders and other stakeholders at multiple scales, from the outset of any decision-making on
556 development and retention objectives, will be essential to these endeavours.

557 Establishing an apex environmental target for global efforts to be directed towards does have
558 limitations. For example, Purvis (2020) contended that a single apex target focussed on biodiversity
559 is unwise, given that the different motivations for conserving biodiversity (e.g., preserving intrinsic vs
560 utilitarian values) need fundamentally different strategies. We acknowledge this, but also recognize
561 that the global agreements on which our analysis is based provide the most holistic and pluralistic
562 set of goals for scientifically-formulated targets for the retention of natural ecosystems. In this
563 respect, our apex target is relevant to the policy processes that multilateral environmental

564 agreements trigger among its signatories. For example, such a target would act as a useful empirical
565 marker for the achievement of overarching aspirations of these agreements, such as the CBD's 2050
566 Vision of humanity "living in harmony with nature".

567 **Limits to loss**

568 Our analysis acts as a starting point for establishing where and what nature should be retained to
569 help us meet global environmental commitments. We make a range of assumptions – most notably,
570 that all selected areas of natural vegetation are of equal value (and condition) with respect to their
571 contribution to the goals they are selected under. Variability in condition (and threats) has
572 substantial consequences for how sites are managed and retained – in the case of biodiversity for
573 example, extensive pressure-free lands will have different management imperatives compared with
574 habitat remnants in human mosaics. Improving the inputs for selecting sites for retention, and the
575 resolution of the ecosystem data (including information on condition (e.g., Hansen et al. 2020), may
576 allow for a more refined retention map and target value to be produced. This, however, may be
577 more suited to regional- or local-level analyses, where higher resolution data inputs are available,
578 and where relevant local actors can be involved in setting targets and deciding how to achieve them.
579 We reiterate that this analysis sets a minimum threshold for retention, given that we did not
580 comprehensively account for all elements of the biota linked to each goal (e.g., our carbon target
581 only deals with primary forest; we did not consider biodiversity in arid (barren) environments), as
582 well as the relatively narrow scope of goals that we considered; additional goals can only increase
583 this value.

584 With ongoing losses a certainty, and numerous questions around the capacity of many ecosystems
585 to recover (at least in timeframes congruent with human lifetimes/several generations, let alone
586 time-bound international environmental agreements), balancing losses with gains to achieve net
587 outcomes cannot be relied upon to maintain ecosystems at the required retention amount.
588 Moreover, the extensive degradation within remaining natural ecosystems – almost 60% of forests
589 globally have compromised ecological integrity and are in a degraded state (Grantham et al. 2020) –
590 further underscores that the retention values we present should be considered as conservative
591 targets. Retaining vegetation well above the apex target will act as a buffer for the (inevitable) net
592 losses incurred as a result of continued land use change, resource extraction and degradation. Of
593 course, a small amount of net loss could be absorbed in areas that our maps indicate are less critical
594 for retention for any of the four targets we consider (~21% of remaining). However, we urge great
595 caution in this interpretation, as the value of these sites for other environmental and/or sustainable

596 development imperatives, as well as the ecosystem services they may supply to proximal
597 communities, has not been assessed here. Thus, we strongly advise that these ecosystems should
598 not be considered ‘worthless’ regarding their contribution to nature and people, with any actions in
599 them (i.e. development) subject to rigorous scrutiny. This especially applies where the loss of such
600 ecosystems will negatively affect the supply of services to local communities, albeit without
601 necessarily detracting from the achievement of goals that are far broader in their scope.

602 Increasingly, nations and other influential actors are coalescing around the idea that transformative
603 change to the way in which we live, and manage our biosphere, is needed. Noting that time is
604 running out to halt and reverse the degradation of nature, we propose that a good starting point for
605 initiating this change is asking the simple question: what do we need and want from nature? This
606 question manifests here in the form of a retention target for a least four of nature’s contributions to
607 people, which can allow us to map the places and inform the actions needed to maintain natural
608 ecosystems. For the benefit of biodiversity, and for people too, we need to keep a great deal of the
609 world’s remaining natural ecosystems in place. This analysis provides a starting point to the
610 questions of ‘where’ and ‘how much’, this should be.

611

612 **ACKNOWLEDGEMENTS**

613 Pasquale Borrelli is funded by the EcoSSoil Project, Korea Environmental Industry & Technology
614 Institute (KEITI), Korea (Grant No. 2019002820004).

615

616 **AUTHOR CONTRIBUTIONS**

617 The data and methods for the spatial quantification of retention targets presented in this paper
618 were prepared and developed by MM, JSS, JEMW, AFSC, AER, JAA, SLM, PB, PPa, SL, TS, LJS and RKR.
619 Spatial analyses were conducted by AFSC and AER. JSS and MM wrote the manuscript. All authors
620 contributed to the conceptualisation of this paper, to subsequent drafting and editing of the paper,
621 and approved its final version.

622

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