Influence of warming temperatures on coregonine embryogenesis within and among species

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ABSTRACT:

The greatest known global response of lakes to climate change has been an increase in water

temperatures. The responses of many lake fishes to warming water temperatures are projected to

be inadequate to counter the speed and magnitude of climate change. We experimentally

evaluated the responses of embryos from a group of cold, stenothermic fishes (Salmonidae

Coregoninae) to increased incubation temperatures. Study groups included cisco (*Coregonus*

artedi) from lakes Superior and Ontario (USA), and vendace (*C. albula*) and European whitefish

(*C. lavaretus*) from Lake Southern Konnevesi (Finland). Embryos from artificial crossings were

incubated at water temperatures of 2.0, 4.5, 7.0, and 9.0°C, and their responses were quantified

for developmental and morphological traits. Embryo survival, incubation period, and length-at-

hatch were inversely related to incubation temperature whereas yolk-sac volume increased with

incubation temperature within study groups. However, varying magnitudes of responses among

study groups suggested differential levels of developmental plasticity to climate change.

Differential levels of parental effects indicate genetic diversity may enable all study groups to

adapt to cope with some degree of changing environmental conditions. Our results suggest that

the coregonines sampled within and among systems may have a wide range of embryo responses

to warming incubation conditions.

KEYWORDS: climate change, embryo incubation, thermal habitat, reaction norm, parental

effect, Coregonus

INTRODUCTION:

Freshwater lakes are sensitive to climate change (Jenny et al., 2020; Woolway et al., 2020), and one of the greatest threats to lakes from climate change on a global scale is rising water temperatures (Austin & Colman, 2007; O'Reilly et al., 2015; Woolway et al., 2017; Maberly et al., 2020). The greatest seasonal increase in water temperature of seasonally ice-covered lakes is projected to take place during the spring (Schindler et al., 1990; Winslow et al., 2017), concurrent with expected seasonal increases in air temperature during winter in the northern hemisphere (Christensen et al., 2007; Sharma et al., 2019). Such changes are likely to challenge aquatic organisms with life-history traits that include critical life stages during the winter-to-spring transition in these regions.

Warming spring water temperatures and increases in the length of the frost-free season can prolong annual growing seasons for aquatic organisms through warmer summers, longer and warmer autumns, and shorter ice-cover duration (Sharma et al., 2019, 2020). Temperature is an abiotic master factor for aquatic ecosystems because water temperature directly affects the physical and chemical properties of water, and phenology, reproductive events, metabolic rates, growth, and survival of aquatic organisms (Fry, 1964; Brett, 1979; Gillooly et al., 2002; Brown et al., 2004; Ohlberger et al., 2007; Busch et al., 2012; Cline et al., 2013; Little et al., 2020). The responses of many lake organisms to climate-derived changes in lake ecology are projected to be inadequate to counter the speed and magnitude of climate change, leaving some species vulnerable to decline and extirpation (Hoffmann & Sgrò, 2011). These pressures present challenges for biodiversity conservation and sustainability of ecosystem services. To navigate challenges, a foundational understanding of the primary threats to aquatic ecosystems and

organisms across a range of spatial scales from local to global is needed (Vörösmarty et al.,

2010; Halpern et al., 2015; Langhans et al., 2019).

The effects of increasing temperature on lake fishes are predicted to lead to declines in cold-

water species and increases in warm-water species (Comte et al., 2013; Hansen et al., 2017).

Species that possess narrow optimal thermal ranges, live near their thermal limits, or have long

development times at cold temperatures are at-risk under warming climate scenarios as

temperature can have strong direct and indirect effects at early-life stages (Blaxter, 1991; Pepin,

1991; Ficke et al., 2007; Mari et al., 2016; Lim et al., 2017; Dahlke et al., 2020). Unlike their

marine counterparts, most freshwater fishes are restricted to their lake system, where their ability

to evade the effects of climate change is impeded due to the isolated nature of lakes (Ficke et al.,

2007). The ability of lake fish populations living close to their upper thermal limits to respond

either through physiological adaptation or physiological plasticity will be required if species are

to persist under increasingly stressful thermal conditions (Somero, 2009; Woolsey et al., 2015;

Howells et al., 2016). Additionally, the amount of genetic variability within a population may

constrain the ability to cope with environmentally-induced changes (*i.e.,* phenotypic plasticity;

Somero, 2009; Schindler et al., 2015, 2010).

Freshwater whitefishes, Salmonidae Coregoninae (hereafter coregonines), are of great socio-

economic value (Nyberg et al., 2001; Ebener et al., 2008; Vonlanthen et al., 2009, 2012; Lynch

et al., 2015, 2016), and are also considered to be critically sensitive to the effects of climate

change because they are cold, stenothermic fishes (Stockwell et al., 2009; Elliott & Bell, 2011;

Jeppesen et al., 2012; Isaak, 2014; Jonsson & Jonsson, 2014; Karjalainen et al., 2015, 2016a).

Coregonine fisheries worldwide have experienced population declines due to highly variable and

weak year-class strengths (Nyberg et al., 2001; Vonlanthen et al., 2012; Anneville et al., 2015;

Myers et al., 2015). In the 20th century, causes of decline included fishing and stocking practices

(Anneville et al., 2015) and eutrophication causing poor incubation conditions (Müller, 1992;

Vonlanthen et al., 2012). Today, in the northern hemisphere, the trophic state of lakes and

fisheries management practices are improving, but coregonines continue to be the focus of

reintroduction, restoration, and conservation efforts in many lakes (Favé & Turgeon, 2008;

Zimmerman & Krueger, 2009; Bronte et al., 2017). Reasons for declining recruitment are

unknown, but climate change, increasing water temperatures, and habitat degradation are

hypothesized as the main causal factors (Nyberg et al., 2001; Marjomäki et al., 2004; Jeppesen et

al., 2012; Anneville et al., 2015; Karjalainen et al., 2015, 2016a).

Coregonines generally spawn during late-autumn, embryos incubate over winter, and hatch in

early- or late-spring (Stockwell et al., 2009; Karjalainen et al., 2015). The time between

fertilization and hatching is inversely related to water temperature (Colby & Brooke, 1970, 1973;

Luczynski & Kirklewska, 1984; Pauly & Pullin, 1988; Karjalainen et al., 2016a). Rising spring

water temperatures (*e.g.,* $> 4°C$) trigger hatching of autumn-spawned coregonine embryos

(Häkkinen et al., 2002; Urpanen et al., 2005; Karjalainen et al., 2015). The length of the newly-

hatched larvae is negatively correlated with the temperature of incubation (Colby & Brooke,

1970, 1973; Luczynski & Kirklewska, 1984; Karjalainen et al., 2015). The long period between

spawning and hatching exposes coregonines to a variety of thermal conditions, potentially

resulting in a wide range of environmentally-induced phenotypes or plastic responses

(Karjalainen et al., 2015, 2016a, 2016b). Coregonines are known to be behaviorally and

developmentally plastic (Muir et al., 2013) and some species (*e.g.,* vendace *Coregonus albula*

and European whitefish *C. lavaretus*) can respond to short- and long-winter conditions within the

limits of phenotypic plasticity and through genetic adaptive changes, such as different embryo

developmental rates (Karjalainen et al., 2015, 2016a).

Geographic variation is also important to consider with phenotypic plasticity. Many fishes in

high-latitudes are adapted to relatively colder waters, extensive periods of ice cover, and strong

seasonal daylight variations (Reist et al., 2006). Thus, high-latitude populations can show

differential long-term adaptation to climates, compared to low-latitude population, across a

latitudinal gradient (Conover & Present, 1990; Yamahira & Conover, 2002; Chavarie et al.,

2010; Wilder et al., 2020). For example, a number of fishes have demonstrated an inverse

relationship between length of the growing season and reproductive characteristics (*i.e.,*

countergradient variation; Billerbeck et al., 2000; Chavarie et al., 2010; Conover & Present,

1990; Conover & Schultz, 1995; Jonassen, 2000; Schultz et al., 1996, 1998; Yamahira &

Conover, 2002). Fishes at high latitudes experience lower temperatures overall and shorter

growing seasons and should exhibit lower standard metabolic rates, growth rates, and smaller

size-at-age than individuals at low latitudes (Reist et al., 2006; White et al., 2012). However, for

cold-water stenothermic fishes, water temperatures at low latitudes may exceed their optimal

range for significant portions of the growing season, or the amount of optimal thermal habitat

decrease, while water temperatures at high latitudes may remain near the optimum for maximal

growth efficiency throughout the growing season (Conover & Schultz, 1995). Because water

temperature has a great influence on fish physiology and varies across latitudes, a wide range of

adaptive responses by populations to increasing temperatures across latitudes is possible (Reist et

al., 2006). However, in the absence of genetic information, phenotypic changes are difficult to

distinguish as genetically based or the result of phenotypic plasticity (Merilä & Hendry, 2014;

Fox et al., 2019). Large-scale experimental studies across geographic regions can aid our

understanding of the role of phenotypic plasticity relative to adaptive evolution by leveraging

long-term adaptations to distinct geographic locations and environments (Hoffmann & Sgrò,

2011; Merilä & Hendry, 2014; Fox et al., 2019). Coregonines occur broadly across northern

latitudes and are an ideal group to test how cold-water lake fishes may respond to climate-driven

shifts in environmental variables, such as water temperature.

Our objective was to compare the reaction norms of coregonine embryos within and among

species from multiple sampling locations across North America and Europe to a standardized

thermal gradient during incubation. We hypothesized that coregonines have differential levels of

phenotypic plasticity in developmental and morphological traits of embryos in response to

warming winter incubation conditions based on putative adaptation to their local environments.

We predicted coregonines that share the same thermal environment respond similarly and

geographically distinct groups with different thermal environments respond dissimilarly to

increasing incubation temperatures.

MATERIALS AND METHODS:

<u>Study Species and Locations</u>

We used a cross-lake, cross-continent, cross-species approach to evaluate the responses and

thermal tolerances of coregonine embryos to changing thermal regimes. Wild-caught populations

of cisco (*C. artedi*) in Lake Superior (LS-Cisco; USA/Canada) and Lake Ontario (LO-Cisco;

USA/Canada), and vendace (LK-Vendace) and European whitefish (LK-Whitefish) in Lake

Southern Konnevesi (Finland; Fig. 1) were sampled.

Cisco is one of the most widespread of the North American species of coregonines (Eshenroder

et al., 2016) and were one of the most abundant fish in the Great Lakes (Yule et al., 2013). Cisco

is found in north-central to eastern United States and throughout most of Canada, with the lower

Great Lakes close to its southernmost extent (Scott & Crossman, 1973). Cisco spawning is

initiated when water temperatures decrease to 4-5°C in late Autumn (Pritchard, 1931;

Eshenroder et al., 2016) and occurs at different spawning depths. Spawning can occur at depths

ranging from 1-5 m in Lake Ontario (Pritchard, 1931; Paufve, 2019) and 10-64 m in Lake

Superior (Dryer & Beil, 1964; Paufve, 2019). Experimental thermal optima for normal cisco

embryo development is between 2 and 8°C (Colby & Brooke, 1970; Brooke & Colby, 1980).

However, temperature data at historical spawning grounds indicate that *in-situ* incubations

typically occur between 1 and 4°C, with Lake Ontario warmer than Lake Superior (Fig. 2;

unpublished data). Optimal growth and upper lethal temperatures for cisco adults are estimated

to be 18° and 26°C, respectively (Edsall & Colby, 1970; McCormick et al., 1971; Jobling, 1981).

Vendace and European whitefish are widely distributed in Northern Europe (Sipponen et al.,

2006). Vendace spawning is initiated when water temperatures decrease to 6°C at the end of

October in Lake Southern Konnevesi and lasts 2-3 weeks (Karjalainen & Marjomäki, 2018).

European whitefish spawn 2-3 weeks later in November when water temperatures decrease to 4-

5°C (Karjalainen et al., 2015). Spawning of vendace occurs in the littoral and sublittoral zone of

lakes and eggs dispersed widely at depths less than 20 m (Heikkilä et al., 2006; Karjalainen et

al., In press). Whitefish are known to spawn at shallower depths than vendace. Embryos are

incubated at 1-2°C until the beginning of April, when water temperatures have gradually

increased, and hatching peaks after ice-out at 4-6°C (Fig. 2, Karjalainen et al., 2015). Although

whitefish spawn later than vendace, their hatching occurs at the same time or slightly earlier than

vendace (Karjalainen et al., 2015). Optimal growth and upper lethal lethal temperatures for

adults are estimated to be 18° and 27°C for immature vendace and 18° and 29°C for European

whitefish, respectively (Tapaninen et al., 1998, Vielma et al. 2002, Siikavoipio et al. 2013).

<u>Adult Collections</u>

Adults were sampled using multi-mesh gillnets (51-89 mm stretched mesh) in Lake Superior

(46.85°, -90.55°), trap nets in Lake Ontario (44.05°, -76.20°), and seines (vendace) and multi-

mesh gillnets (European whitefish; 27-35 mm knot to knot mesh) in Lake Southern Konnevesi

(62.58°, 26.58°). Adult field collections occurred during coregonine spawning periods for Lake

Ontario and Lake Superior. On Lake Southern Konnevesi, adults were collected prior to

spawning and held in aquaculture pools with water fed directly from the lake until spawning was

initiated. Demographic data (*e.g.,* total length, mass, and egg diameter) were collected on

sampled adults. All sampling, fertilization, and experimental work for study groups on each

continent were conducted at a single laboratory in North America (University of Vermont

(UVM), USA) and Europe (University of Jyväskylä (JYU), Finland). Experiments were

performed during the 2018-19 season in Finland and the 2019-20 season in the USA.

For clarity, our operational use of a study group is to represent a single species haphazardly

sampled from a single location within a single lake (*e.g.,* cisco from the Apostle Islands region in

Lake Superior). Our sampling efforts represent a single location within large lakes and likely do

not capture all possible genetic variation within a species or population.

<u>Crossing Design and Fertilization</u>

Eggs and milt were stripped from 12 females and 16 males from each study group and artificially

fertilized under a blocked, nested full-sib, half-sib fertilization design to create a maximum of 48

full-sibling families nested within half-siblings per group. The crossing design maximized the

amount of genetic variation and minimized the potential loss of multiple families if a female or

male produced poor quality gametes, for a given total number of families, compared to a full-

factorial design. Adults used in the experiment were divided into three or four fertilization

blocks. A single block consisted of four males each paired to three females, where all offspring

of a given female were full siblings. Fertilizations were performed block by block to ensure germ

cell survival.

Approximately 200 eggs per female were fertilized with an equal amount of milt (5-15 μl) from

each male in the block. After the addition of milt, water was added to activate the germ cells and

gently mixed for one minute. The embryos were rinsed with water 2-3 times until the water was

clear. Temperature of the water used during all fertilizations was ca. 4°C. Reconstituted

freshwater medium (ISO 6341, 2012) was used during fertilizations to standardize the chemical

properties of the water used among study groups and between labs. Embryos were transported in

coolers either by shipping overnight for Lake Superior or driven same-day for Lake Ontario. A

temperature logger (HOBO® Water Temperature Pro v2) recorded air temperature inside the

cooler during transport (Lake Superior: mean (SD) = 2.80°C (0.21); Lake Ontario: mean (SD) =

3.28°C (0.37)). No embryo transport was required for Lake Southern Konnevesi. Fertilization

success was determined by haphazardly taking 10 embryos from each family and assessing under

microscopy within 72-hours post-fertilization (Oberlercher & Wanzenböck, 2016). If fertilization

was low (<30%), the family was removed from the experimental setup.

<u>Rearing Conditions</u>

Embryos from successfully fertilized families were individually distributed into 24-well cell

culture microplates and incubated in 2 ml of reconstituted freshwater. Reconstituted freshwater

was used during incubation to maintain sterility, prevent bacterial growth in the wells, and

eliminate the need for fungicide treatments on the embryos. A total of 36 embryos per family

were used for Lake Southern Konnevesi species and 48 embryos per family for each of Lake

Ontario and Lake Superior cisco. Families were randomly distributed across three or four

microplates (*i.e.,* 12 eggs per family per microplate and two families per 24-well microplate).

Microplates from each study group were incubated at target constant temperatures of 2.0

(coldest), 4.5 (cold), 7.0 (warm), and 9.0°C (warmest) and randomly placed in climate-controlled

chambers at UVM (Memmert[®] IPP260Plus) and climate-controlled rooms at JYU (Huurre[®]).

The range of experimental incubation temperature treatments was chosen to mimic *in-situ* mean

temperatures and to exceed optimum embryonic development temperatures. Forced airflow was

used in both the climate-controlled chambers and rooms to ensure equal air circulation around

the microplates. All microplates were covered to minimize evaporation. Microplate orientation

and position were rotated weekly to eliminate any temperature heterogeneity within the

chambers and rooms. HOBO[®] TMCx-HD temperature probes were used at UVM to record water

temperatures hourly in 50-ml beakers placed inside each climate chamber, while Escort iMini

temperature loggers were used directly inside the well of a subset of microplates for each

temperature treatment at JYU. Daily mean water temperatures were calculated. Incubations took

place in the dark, with the exception of short (< 60 mins) maintenance periods. Microplates were

checked weekly for dead eggs and the eye-up stage. During the hatch period, microplates were

checked on a two-day cycle for newly hatched larvae. For cisco, all newly hatched larvae were

photographed for developmental and morphological traits (Nikon® D5600 and Nikon® AF-S DX

18-55mm lens). Egg size, total length, and yolk-sac axes were measured from images using

Olympus® LCmicro. For Lake Southern Konnevesi, the larvae were preserved in ethanol at hatch

and flushed and soak in distilled water for 15 min before measuring the total length and fresh

mass under the microscope (Karjalainen, 1992).

Mean water temperature during incubations was maintained near the target incubation

temperature for the cold and warm treatments at each lab. Mean incubation water temperatures

for the cold and warmest treatments were lower than the target incubation temperature at JYU,

but not at UVM (Table 1).

<u>Developmental and Morphological Traits</u>

Embryo survival was estimated as the percent of embryos surviving between the eye-up and

hatch stages to rule out unfertilized eggs which can bias survival estimates. Incubation period

was assessed by two variables: the number of days from fertilization to hatching (days post-

fertilization; DPF) and the sum of the degree-days (accumulated degree-days; ADD). Total

length-at-hatch (mm) and yolk-sac volume (YSV; mm^3) were measured from five individuals per

family at, or as close as possible to, 50% hatching for each family. Yolk-sac volume was

calculated assuming the shape of an ellipse (Blaxter & Hempel, 1963):

$$YSV = \frac{\pi}{6} ab^2 \, ,$$

where a = length of the yolk sac (mm) and b = height of the yolk sac (mm).

<u>Statistical Analyses and Estimation of Variance Components</u>

Embryo survival was analyzed as a binomial response variable, and incubation period, length-at-

hatch, and yolk-sac volume at hatching as continuous response variables. Early embryo mortality

and variable fertilization success induced from fertilization failure produced inequalities in the

number of offspring among families. The sample size for incubation period is a function of

embryo survival and subsequently resulted in an unbalanced design. All non-proportional data

were checked visually for approximate normality using histograms and Q-Q plots before the

analysis with parametric models. A cubic transformation was applied to LAH for cisco and a

cubic root transformation was applied to DPF, ADD, and YSV to normalize the distributions.

Therefore, binary data (*i.e.,* embryo survival) were analyzed with binomial generalized linear

mixed-effects models (LMM) and transformed variables (*i.e.,* DPF, ADD, LAH, and YSV) were

analyzed with restricted maximum likelihood LMMs with the *lme4* package v.1.1-26 (Bates et

al., 2015). To eliminate any confounding effects between continents, cisco from lakes Superior

and Ontario were analyzed independently from vendace and European Whitefish from Lake

Southern Konnevesi. Population (for cisco only), species (for Lake Southern Konnevesi only),

and incubation temperature were included as fixed effects and female, male, family (female and

male combination), microplate, and fertilization block as random effects. Because embryos were

raised independently, the replication unit in the statistical models is the individual embryo.

Although incubation temperature was treated as a fixed variable based on our experimental

design (*i.e.,* few treatment levels with high replication), we acknowledge that temperature is a

continuous, independent variable in nature. All traits were examined for population or species,

depending on the continent, and incubation temperature effects in addition to individual parental

effects (female, male, and family effects), microplate, fertilization block, and all possible

interactions with backward, stepwise effect-selection using the *buildmer* package v.1.7.1

(Voeten, 2020). The maximal model for each trait was selected by comparing a model including

or lacking the term of interest to the reference model based on changes in log-likelihood, Akaike

information criterion, Bayesian information criterion, and change in explained deviance.

Significance values for the mixed-effects model parameters (*i.e.,* population, species, incubation

temperature, interaction effects, and any random-effects selected) were determined using a

likelihood ratio test between the maximal model and reduced models with the model parameter

of interest removed (Myers et al. 1995; Luke 2017). All statistical tests used $\alpha = 0.05$ to

determine significance.

To allow for interspecific comparisons, the response to temperature for each trait was

standardized to the optimal temperature for each study group. Based on literature data (cf.

above), the coldest incubation temperature treatment (2.0°C and 2.2°C; Table 1) was assumed to

be the optimal incubation temperature. For each trait, the within-family mean was calculated for

each temperature treatment and the percent change from the optimal temperature estimated.

Standard error was calculated as the among-family variation in percent change.

The phenotypic variance components were partitioned into random effects for female, male,

female x male, and unexplained or random residual variance components using mixed-effects

models with the *fullfact* package v.1.3 (Houde & Pitcher, 2016) for each study group and

incubation temperature treatment. Negative variance components were treated as zero (Neff &

Pitcher, 2005). The percent of total phenotypic variation was used to calculate the Pearson

correlation coefficient between each variance component and the increase in incubation

temperature for each study group. A threshold of ±0.7 (*i.e.,* an R^2 of 0.49) was used to categorize

correlations either positive or negative, with all values in between as no correlation. European

whitefish from Lake Southern Konnevesi were removed from this analysis due to a low embryo

survival and a low number of families.

All analyses were performed in R version 4.0.3 (R Core Team, 2020).

RESULTS:

<u>Spawning Adults</u>

Total lengths and fresh mass of spawning adults used for gamete collection varied widely among

study groups (Table 2). LK-Vendace were notably smaller than all other study groups. The

remaining study groups varied less in size, but LK-Whitefish were smaller than LS-Cisco and

LO-Cisco.

The LK-Vendace females had the smallest egg diameters and LO-Cisco females had the largest

egg diameters among the study groups (Table 3). LK-Whitefish and LS-Cisco egg diameters

were similar (Table 3).

 Developmental and Morphological Traits and Variance Components

All cisco traits had significant interaction effects between population and incubation temperature

(maximum $P < 0.001$; Tables 4 and 5). All vendace and European whitefish traits had significant

interaction effects between species and incubation temperature (maximum $P = 0.002$; Tables 4

and 5). The interaction effects precluded any interpretation of main effects, but did suggest

different norms of reaction for the model groups. Below we describe the interaction effects. All

random effects (*i.e.,* female, male, and female x male) were significant (maximum $P = 0.038$)

except male for LAH and female x male for LAH and YSV in cisco (Tables 4 and 5). All

statistical model results can be found in Tables 4 and 5.

Embryo Survival

Embryo survival was highest among all study groups at the coldest temperature and lowest at the

warmest temperature (Fig. 3). The effect of temperature for cisco depended on population

because embryo survival was higher for LO-Cisco (99.3%) than LS-Cisco (80.0%) at the coldest

temperature but not different between populations (difference < 0.1%) at the warmest

temperature (Fig. 3). For Lake Southern Konnevesi, the effect of temperature depended on

species because the difference in embryo survival between LK-Vendace and LK-Whitefish was

less pronounced at the coldest temperature (29.0%) than at the warmest temperature (50.5%; Fig.

3). LK-Vendace and LK-Whitefish embryo survival had a differential temperature response as

LK-Whitefish had a greater decrease (74.4%) than LK-Vendace (17.7%) from the coldest to

warmest incubation temperatures. LK-Whitefish had the strongest, decreasing response to

increasing incubation temperatures compared to all other study groups (Fig. 3).

In the phenotypic variance component analysis, the residual error was the largest component of

phenotypic variation in embryo survival (means > 55.2%) for all study groups (Fig. 4, Online

Resource 1). The mean female variance had the highest percentage, excluding error, of the

phenotypic variation in embryo survival for LK-Vendace (17.4%), LS-Cisco (24.1%), and LO-

Cisco (19.9%; Fig. 4, Online Resource 1). The female variance component correlations for

embryo survival had either negative or no correlations to increasing temperature; however, male

and error variances had positive and no correlations suggesting that as the female component

decreases at higher temperatures the importance of the male component and error increases

(Table 6).

Incubation Period (days post-fertilization)

The number of days post-fertilization to hatching was highest for all study groups at the coldest

temperature and decreased as temperature increased (Fig. 3). For cisco, DPF was higher for LO-

Cisco (179.2 days) than LS-Cisco (154.3 days) at the coldest temperature and the difference

between populations was less pronounced at the warmest temperature (difference = 5.0 days;

Fig. 3). For Lake Southern Konnevesi, the effect of temperature depended on species because the

difference in DPF between LK-Vendace and LK-Whitefish was less pronounced at the coldest

temperature (8.9 days) than at warmest temperature (27.3 days; Fig. 3). All study groups had

similar responses to temperature, with between 54.2 to 68.3% decreases in DPF from the coldest

to warmest treatments. However, LS-Cisco, LO-Cisco, and LK-Whitefish had a greater decrease

in DPF (66.1, 68.3, 65.3%, respectively), than LK-Vendace (54.2%; Fig. 3).

In the phenotypic variance component analysis, the residual error was the largest component of

phenotypic variation in DPF (means >60.8%) for LK-Vendace and LS-Cisco (Fig. 4, Online

Resource 1). The mean female variance was the largest phenotypic variation component in DPF

for LO-Cisco (47.1%). LK-Vendace and LS-Cisco had similar mean female variances for DPF

across all temperatures, with 28.1 and 21.0%, respectively (Fig. 4, Online Resource 1). The DPF

correlations for female effect had a negative response to temperature for LS-Cisco and LO-Cisco

(Table 6).

Incubation Period (accumulated degree-days)

Accumulated degree-days were highest for all study groups at 6.9°C (Fig. 3). The effect of

temperature for cisco depended on population because ADD was higher for LO-Cisco (531.9 and

547.7 ADD) than LS-Cisco (461.0 and 492.5 ADD) at the cold and warm temperatures,

respectively, and the differences between populations were less pronounced at the coldest and

warmest temperatures (differences = 49.2 and 41.3 ADD, respectively; Fig. 3). LS-Cisco and

LO-Cisco ADD responded similarly to increasing incubation temperature. For Lake Southern

Konnevesi, the effect of temperature depended on species because the difference in ADD

between LK-Vendace and LK-Whitefish was less pronounced at the coldest temperature (7.7

ADD) than at the warm temperature (198.1 ADD; Fig. 3). LK-Vendace and LK-Whitefish ADD

had a differential temperature response as LK-Vendace had a greater increase (198.4%) than LK-

Whitefish (159.4%) from the coldest to warm treatment. LK-Vendace had the strongest,

increasing response to increasing incubation temperatures compared to all other study groups

(Fig. 3).

In the phenotypic variance component analysis and correlations, ADD had a similar response as

DPF among all study groups as the data only had a different temperature scaling factor (Fig. 4,

Online Resource 1, Table 6).

Length-at-Hatch

All study groups had a common, decreasing response in LAH as temperature increased (Fig. 5).

For cisco, LAH was higher for LO-Cisco (11.32 and 9.75 mm) than LS-Cisco (10.21 and 8.68

mm) at the coldest and warmest temperatures, respectively, and the difference between

populations was less pronounced at the cold and warm temperatures (differences = 0.99 and 0.90

mm, respectively; Fig. 5). LS-Cisco and LO-Cisco responded to increasing incubation

temperature with a 15.9 and 13.8% respective decrease in LAH from the coldest to warmest

treatments. For Lake Southern Konnevesi, the effect of temperature depended on species because

the difference in LAH between LK-Vendace and LK-Whitefish was more pronounced at the cold

and warm temperatures (2.73 and 2.72 mm, respectively) than at the coldest and warmest

temperatures (2.68 and 2.61 mm, respectively; Fig. 5). LK-Vendace and LK-Whitefish each

responded similarly to temperature with a 9.0 and 9.2% respective decrease in LAH from the

coldest to warmest treatments. LS-Cisco and LO-Cisco LAH had a stronger, decreasing response

to increasing incubation temperatures than LK-Vendace and LK-Whitefish (Fig. 5).

In the phenotypic variance component analysis, the residual error was the largest component of

phenotypic variation in LAH (means >49.2%) for all study groups (Fig. 4, Online Resource 2).

The mean female variance had the highest percentage, excluding error, of the phenotypic

variation in LAH for LK-Vendace (40.6%), LS-Cisco (38.2%), and LO-Cisco (17.1%; Fig. 4,

Online Resource 2). The LAH correlations for each study group had a similar response to

temperature with negative or no female and male correlations and positive or no female x male

correlations (Table 6).

Yolk-sac Volume

Yolk-sac volume was highest for all study groups at 9.0°C and decreased as temperature

decreased (Fig. 5). For cisco, the difference in YSV was similar between populations at the

warmest incubation temperature (0.04 mm^3) but diverged as incubation temperature decreased;

YSV in LO-Cisco (0.40 mm^3) was smaller than LS-Cisco (0.64 mm^3) at the coldest temperature

(Fig. 5). Yolk-sac volume in LS-Cisco and LO-Cisco responded differently to incubation

temperature, with a 203.6 and 311.0% respective increase from the coldest to warmest treatment.

For Lake Southern Konnevesi, the effect of temperature depended on species because the

difference in YSV between LK-Vendace and LK-Whitefish was less pronounced at the coldest

temperature (0.20 mm^3) than at the warmest temperature (1.07 mm^3; Fig. 5). LK-Vendace and

LK-Whitefish had the strongest response to temperature with an increase in YSV of 445.0 and

536.6% from the coldest to warmest treatment, respectively. LK-Vendace and LK-Whitefish had

a stronger, increasing response to increasing incubation temperatures in YSV than LS-Cisco and

LO-Cisco (Fig. 5).

In the phenotypic variance component analysis, the residual error was the largest component of

phenotypic variation in YSV (means >53.9%) for all study groups (Fig. 4, Online Resource 2).

The mean YSV female variance was the highest percentage, excluding error, of the phenotypic

variation for LK-Vendace (23.9%), LS-Cisco (20.5%), and LO-Cisco (23.9%; Fig. 4, Online

Resource 2). The YSV correlations for female, female x male, and error variance components

had differential responses to temperature with positive female, positive female x male, and

negative error correlations for LK-Vendace, while LS-Ciso had inverse correlations to LK-

Vendace (Table 6). All LO-Cisco variance components had no correlation to temperature (Table

6).

DISCUSSION:

Our incubation experiments demonstrated both similar and dissimilar reaction norms to

temperature for developmental and morphological traits in coregonines. First, we found different

responses to temperature in embryo survival within and among study groups. Second, incubation

periods (both DPF and ADD) responded similarly to increasing temperature (negative response

for DPF and positive response for ADD) among study groups, however, LK-Vendace had the

strongest response and longest incubations across all temperatures. Third, all study groups had

similar negative responses to temperature for LAH and positive responses in YSV, with the

strongest responses for LAH in LS-Cisco and LO-Cisco and for YSV in LK-Vendace and LK-

Whitefish. Lastly, differential levels of parental effects were found within and among study

groups and traits.

Embryo survival had an overall negative correlation with increasing temperature among all study

groups. However, cisco and vendace had different levels of response to temperature compared to

European whitefish. LK-Whitefish embryo survival had the strongest, negative response to

temperature (74.4% survival loss) and all other study groups were impacted less ($< 26\%$ survival

loss) by increasing temperatures. Temperature is a strong driver of coregonine embryo

development (Karjalainen et al., 2015) and survival (Colby & Brooke, 1970; Brooke & Colby,

1980; Luczynski & Kirklewska, 1984) but study groups showed different levels of sensitivity to

increased incubation temperatures reflecting specific physiological adaptations.

Additionally, our experiment, temperature aside, provided near-optimal incubation conditions to

individually reared embryos and these conditions are idealized compared to what occurs in the

wild. For instance, embryos in the wild are deposited on the substratum and are exposed to

deposited sediment that can impact survival (Müller, 1992). Interaction between temperature and

sediments are likely, and temperature increases may act as a catalyzer of embryo sensitivity to

sediment stress (Mari et al., 2016, 2021). Even though temperature did negatively impact embryo

survival in our experiment, the effect of temperature in the wild, in combination with other

factors, could be even stronger. Similarly, both study groups from North America required

transportation, which could have had negative effects on embryo survival. Lake Superior had a

greater transportation distance and time than Lake Ontario, which could explain some of the

difference observed in embryo survival between the two study groups. Lake Ontario embryos

had > 99% survival at optimal temperatures which suggests that transportation did not impact

them.

The longer incubation periods from LK-Vendace and LK-Whitefish and strong response to

increasing temperatures, even at the warmest incubation temperatures, support previous findings

that vendace and European whitefish from Lake Southern Konnevesi have a high degree of

developmental flexibility (Karjalainen et al., 2015). The different response between vendace and

European whitefish was likely due to species-specific differences (Karjalainen et al., 2015) and

ecotypes (*i.e.,* benthic versus pelagic; Mcphee et al., 2012). Additionally, the different magnitude

of temperature responses among all study groups suggests a differential level of developmental

plasticity to increasing incubation temperatures among species and locations. Long, relatively

cold incubations may require a shorter period of spring warming for individuals to initiate

hatching, while short, relatively warm incubations may require a longer period of warmer spring

conditions to hatch (Karjalainen et al., 2015). If winter water temperatures rise as embryos

incubate, the ability to match optimal spring nursery conditions may be weakened (Cushing,

1990; Karjalainen et al., 2015; Myers et al., 2015). Populations that are more resilient to

increasing or variable winter incubation temperatures may have a better opportunity to regulate

ontogeny and adjust the time of hatching.

Fish spawning strategies are variable, ranging in frequency from daily to once in a lifetime and

in timing from the same time each year to across all seasons (McBride et al., 2015). For many

species, spawning strategies and breeding patterns are constrained by the adult body condition,

gonadogensis, and the environment (Jørgensen et al., 2006; van Damme et al., 2009; Muir et al.,

2014; McBride et al., 2015). In this context, the short duration of cisco embryo incubation

periods when exposed to high temperatures was notable. High-latitude populations typically

spawn earlier in autumn and may have the opportunity to delay spawning later into the season,

while still providing an adequate incubation period for embryo development, if water

temperatures continue to rise and do not inhibit ovulation as a result of climate change. However,

low-latitude coregonine populations already spawn in late-autumn and early-winter (Stockwell et

al., 2009; Eshenroder et al., 2016), which begs the question: do low-latitude populations have

capacity to spawn later in the winter if temperatures continue to rise? Winter spawning may lead

to less vulnerability to contemporary climate change. For instance, Atlantic herring (*Clupea*

harengus) have both autumn- and winter-spawning stocks in the North Sea that share the same

summer feeding grounds and start oocyte development at the same time (van Damme et al.,

2009). Similarly, coregonines can exhibit varying spawning strategies.

Sympatric coregonine species with autumn, winter, and spring-spawning stocks co-occur in

several northern- and central-Eurasian lakes (Eronen & Lahti, 1988; Schulz & Freyhof, 2003;

Schulz et al., 2006; Ohlberger et al., 2008) and allopatric spring-spawning stocks of cisco are

found in Lac des Écorces (southwestern Quebec; Hénault & Fortin, 1989, 1991; Pariseau et al.,

1983). Winter and spring spawners continue oocyte development through autumn which results

in a lower number of larger eggs compared to the autumn-spawning stocks (Eronen & Lahti,

1988; Hénault & Fortin, 1991). Oocyte development is driven by body energy content, and

winter- and spring-spawning stocks may give iteroparous females the chance to mitigate the

disproportionate energy demand toward somatic growth during the summer when metabolic

demands are higher. Consequently, changes in the environment and the condition of an

individual spawning adult could affect future coregonine spawning strategies. Our results suggest

that the cisco embryos examined may not have the developmental plasticity to mitigate the

effects of increased water temperatures on incubation period. In this context, adjusting the time

of spawning may be a more efficient long-term life-history strategy than the embryos adapting to

increased temperatures. Research on the reproductive plasticity of coregonine adults *(e.g.,*

changes to fecundity of females and size of eggs from temperature) in the face of climate change

is a logical next step.

Lake morphology is also important to consider for the question of a winter- or spring-spawning

adaptation; deeper lakes could sufficiently provide cold thermal refuges at greater depths if

suitable spawning habitat is available and level of oxygen sufficient (Jane et al., 2020). Spring-

spawning ciscos in Lac des Écorces, where a 4°C summer stratum does not exist, initiate

spawning when spring water temperatures reach 6°C at depths ranging from 20-30 m (Hénault &

Fortin, 1989, 1991). This strategy of spawning in deeper, colder water allows for normal

embryogenesis throughout the summer to mitigate high water temperatures during the summer

period at shallow depths. Model projections of suitable thermal and oxythermal habitat for cisco

indicate deeper and less eutrophic lakes will likely provide the best cold-water habitat as water

temperatures and land uses change (Jacobson et al., 2010; Herb et al., 2014; Schmitt et al., 2020).

While deep lakes may possess acceptable thermal refugia for coregonines, access to and

requirements for suitable spawning and incubation habitat is unknown for most populations.

In addition to lower survival and shorter incubations as temperature increases, we also found

both similar and different responses to temperature in morphological traits (*i.e.,* length-at-hatch

and yolk-sac volume) among study groups. The dissimilarity in morphological traits among

study groups is likely related to different egg sizes at fertilization. Smaller eggs will produce

smaller larvae, requiring a lower growth and development rate and less demand on maternal yolk

than larger eggs. The demand for yolk and egg size are positively related and temperature during

embryogenesis is positively related to metabolic rate (Hodson & Blunt, 1986; Kamler, 2008). In

all groups, LAH slightly decreases with increasing temperatures. In LK-Vendace and LK-

Whitefish, however, YSV increases more strongly with temperature, than in cisco. This suggests

a higher yolk conversion efficiency across all temperatures in LK-Vendace and LK-Whitefish

than in cisco. All study groups had a decrease in YSV with time as basal metabolism consumed

yolk as a function of the length of incubation. Regardless of the mode, our results suggest a

synergistic relationship among species, location, egg size, incubation period, and incubation

temperature in determining the phenotype of LAH and YSV.

The trade-off between LAH and YSV is well documented in larval fish physiology (Blaxter,

1991). Climate change impacts may only exacerbate the importance of each morphological trait

in determining either a match or mismatch between larval coregonines and their zooplankton

prey. While our experiments used constant incubation temperatures due to logistical constraints,

the impact different spring warming rates can have on the time of hatching and the size of larvae

should not be ignored. Lake Southern Konnevesi vendace and European whitefish previously

exhibited flexibility in embryo development rates and feeding windows under different warming

scenarios (Karjalainen et al., 2015). Such complex responses challenge our ability to predict the

downstream impacts changing autumn, winter, and spring water temperatures may have on

embryo and larval phenotypic plasticity.

Traits of embryos depend not only on species, population, and incubation temperature but also

on parental and transgenerational effects (Blaxter, 1963, 1991; Kekäläinen et al., 2018). Our

results suggest that both female and male effects controlled a portion of early-stage offspring

trait phenotypes in coregonines. The variability in phenotypes induced by parental effects can

provide more flexibility for a population to cope with changing inter-annual environmental

conditions, prevent full year-class failure, and ensure population persistence (Wright & Trippel,

2009; Oomen & Hutchings, 2015; Karjalainen et al., 2016a). In fishes, the female effect is

usually more pronounced than male and female x male interaction effects, and is stronger in

traits directly related to egg size (Nagler et al., 2000; Kennedy et al., 2007; Huuskonen et al.,

2011), and our results support this conclusion. Residual error estimates, however, remained high.

Intersexual selection and mate pairing has been proposed as an important component affecting

coregonine offspring fitness (Wedekind et al., 2008; Huuskonen et al., 2011; Karjalainen &

Marjomäki, 2018), and may play a role in conserving natural biodiversity within populations

(Anneville et al., 2015). The long-term stability of commercially exploited stocks, which can

experience fisheries-induced evolution, has been linked to population diversity (Schindler et al.,

2010; Freshwater et al., 2019). Spawning stocks that comprise individuals of variable sizes and

ages (*e.g.,* portfolio effect; D. E. Schindler et al., 2010) may contribute differently to spawning,

offspring performance, and recruitment (Luck et al., 2003; Figge, 2004), and are likely an

important consideration as the rapid rate of climate change adds additional stressors on

populations.

The methods we used provide reproducible experimental conditions (*e.g.,* uniform water source

between laboratories, no moving water, minimal embryo disturbance, etc.) and standardized

results that can be compared to future experiments that examine temperature responses across a

wider range of populations. Additionally, efforts to include more northerly populations of cisco

from Canada were thwarted due to restrictions on transport of live embryos across an

international border. This further highlights the importance of the standardized experimental

methods we used to allow for future large-scale, cross-laboratory experiments. However, our

results do suggest that some form of latitudinal variation is likely present and promote fruitful

opportunities for future large-scale experimental research on coregonines and other cold,

stenothermic fishes.

Additionally, interpreting the impacts of parental responses within an environmental context

continues to be important to determine how parental effects may mitigate species' responses to

rapid climate change. The existence of varying parental responses raises questions concerning

possible causal mechanisms. Genomic studies will be needed to better understand what is

genetically impacted by increasing temperatures, how it is impacted, and when during

development (*i.e.,* when is temperature most critical; Chen et al., 2018; Narum et al., 2013). A

mechanistic understanding of thermal response from populations across latitudes will be

essential to predict the vulnerability of species and populations to climate change. Furthermore,

incorporating phylogenetic contrasts would highlight the shared phylogenetic history and genetic

architecture among our study groups and aid future comparative studies examining phenotypic

traits across species and genera (Garland et al., 2005). This study does contain species from the

same genus, but species are either reproductively isolated on different continents (*i.e.,* cisco and

vendace) or are distinct ecotypes (*i.e.,* vendace and European whitefish). Moreover, stocking

practices are important to consider, including supportive breeding, and may affect the adaptive

potential of populations through an artificial selection process (Ford, 2002; Christie et al., 2012;

Anneville et al., 2015) and their ability to respond to environmental changes.

CONCLUSION:

Water temperature is fundamental in regulating fish physiology, and environmental variation

during development can play a large role in generating variability in offspring through

phenotypic plasticity (Little et al., 2020). How coregonines respond during the critical embryonic

stage is important to determine whether the capacity to respond to climate change and the

projected increases in their incubation temperatures exists. Knowing how populations have

adapted historically to environmental variability will help us understand the range of possible

responses to climate change and assist managers to keep coregonines out of hot water.

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TABLES:

Table 1. Mean (SD) water temperatures during embryo incubations at the University of Vermont (UVM) and University of Jyväskylä (JYU).

	Incubation Temperature Treatment (°C)			
Laboratory	2.0 Coldest	4.5 Cold	7.0 Warm	9.0 Warmest
UVM	2.0 (0.5)	4.4 (0.2)	6.9 (0.2)	8.9 (0.3)
JYU	2.2 (1.5)	4.0 (0.7)	6.9 (0.5)	8.0 (0.6)

Table 2. Mean (SD) total length (TL) and fresh mass (FM) of the female and males from Lake Southern Konnevesi (LK-Vendace (*Coregonus albula*) and LK-Whitefish (*C. lavaretus*)), Lake Superior (LS-Cisco (*C. artedi*)), and Lake Ontario (LO-Cisco).

	LK-Vendace		LK-Whitefish		LS-Cisco		LO-Cisco	
Sex	TL (mm)	FM (g)	TL (mm)	FM (g)	TL (mm)	FM (g)	TL (mm)	FM (g)
Female	144.67 (16.51)	18.36 (5.95)	256.57 (11.63)	117.00 (19.16)	428.92 (44.40)	676.02 (181.51)	380.33 (24.18)	567.59 (122.89)
Male	140.83 (9.22)	13.85 (2.27)	285.75 (40.86)	171.34 (87.22)	400.25 (34.35)	523.82 (134.65)	366.56 (25.30)	443.29 (103.16)

Table 3. Mean (SD) egg diameter of females with the number of eggs measured (N) from Lake Southern Konnevesi (LK-Vendace (*Coregonus albula*) and LK-Whitefish (*C. lavaretus*)), Lake Superior (LS-Cisco (*C. artedi*)), and Lake Ontario (LO-Cisco).

Population	Egg Diameter (mm)	N
LK-Vendace	1.58 (0.11)	273
LK-Whitefish	2.13 (0.12)	70
LS-Cisco	2.14 (0.12)	140
LO-Cisco	2.30 (0.08)	240

Table 4. Likelihood ratio test output for each model selected for embryo survival and incubation period (number of days post-fertilization (DPF) and accumulated degree days (°C; ADD)) from lakes Superior and Ontario cisco (*Coregonus artedi*) and Lake Southern Konnevesi vendace (*C. albula*) and European whitefish (*C. lavaretus*). t indicates temperature, pop indicates population, and sp indicates species. The full model that was selected is bolded for each trait and species.

Trait	Species	Model	Effect Tested	df	χ^2	p-value
Embryo Survival	Cisco	**t + pop + t:pop + female:male + female**				
		pop + female:male + female	t	3	443.54	< 0.001
		t + female:male + female	pop	1	600.61	< 0.001
		t + pop + female:male + female	t:pop	3	198.56	< 0.001
		t + pop + t:pop + female	female:male	1	181.47	< 0.001
		t + pop + t:pop + female:male	female	1	23.36	< 0.001
	Vendace & European Whitefish	**t + sp + t:sp + female:male + female**				
		sp + female:male + female	t	3	223.54	< 0.001
		t + female:male + female	sp	1	993.43	< 0.001
		t + sp + female:male + female	t:sp	3	52.94	< 0.001
		t + sp + t:sp + female	female:male	1	1042.9	< 0.001
		t + sp + t:sp + female:male	female	1	1015.8	< 0.001
Incubation Period (DPF)	Cisco	**t + pop + t:pop + female:male + female + male**				
		pop + female:male + female + male	t	3	31,183.13	< 0.001
		t + female:male + female + male	pop	1	3,651.86	< 0.001
		t + pop + female:male + female + male	t:pop	3	315.04	< 0.001
		t + pop + t:pop + female + male	female:male	1	58.62	< 0.001
		t + pop + t:pop + female:male + male	female	1	65.04	< 0.001
		t + pop + t:pop + female:male + female	male	1	14.22	<0.001
	Vendace & European Whitefish	**t + sp + t:sp + female:male + female + male**				
		sp + female:male + female + male	t	3	7,178.37	< 0.001
		t + female:male + female + male	sp	1	904.95	< 0.001
		t + sp + female:male + female + male	t:sp	3	353.67	< 0.001
		t + sp + t:sp + female + male	female:male	1	13.55	< 0.001
		t + sp + t:sp + female:male + male	female	1	36.07	< 0.001
		t + sp + t:sp + female:male + female	male	1	4.77	0.029
Incubation Period (ADD)	Cisco	**t + pop + t:pop + female:male + female + male**				
		pop + female:male + female + male	t	3	15,519.48	< 0.001
		t + female:male + female + male	pop	1	3,586.18	< 0.001
		t + pop + female:male + female + male	t:pop	3	137.36	< 0.001
		t + pop + t:pop + female + male	female:male	1	54.58	< 0.001
		t + pop + t:pop + female:male + male	female	1	63.20	< 0.001
		t + pop + t:pop + female:male + female	male	1	15.34	< 0.001
	Vendace & European Whitefish	**t + sp + t:sp + female:male + female + male**				
		sp + female:male + female + male	t	3	3,571.72	< 0.001
		t + female:male + female + male	sp	1	763.19	< 0.001
		t + sp + female:male + female + male	t:sp	3	466.23	< 0.001
		t + sp + t:sp + female + male	female:male	1	12.70	< 0.001
		t + sp + t:sp + female:male + male	female	1	36.79	< 0.001
		t + sp + t:sp + female:male + female	male	1	4.33	0.038

Table 5. Likelihood ratio test output for each model selected for length-at-hatch (mm) and yolk-sac volume (mm^3) from lakes Superior and Ontario cisco (*Coregonus artedi*), Lake Southern Konnevesi vendace (*C. albula*), and Lake Southern Konnevesi European whitefish (*C. lavaretus*). t indicates temperature and pop indicates population. The full model that was selected is bolded for each trait and species.

Trait	Species	Model	Effect Tested	df	χ^2	p-value
Length-at-Hatch	Cisco	**t + pop + t:pop + female**				
		pop + female	t	3	886.79	< 0.001
		t + female	pop	1	628.15	< 0.001
		t + pop + female	t:pop	3	19.03	< 0.001
		t + pop + t:pop	female	1	161.40	< 0.001
	Vendace & European Whitefish	**t + sp + t:sp + female:male + female**				
		sp + female:male + female	t	3	467.20	< 0.001
		t + female:male + female	sp	1	2091.33	< 0.001
		t + sp + female:male + female	t:sp	3	14.49	0.002
		t + sp + t:sp + female	female:male	1	13.83	< 0.001
		t + sp + t:sp + female:male	female	1	48.81	< 0.001
Yolk-sac Volume	Cisco	**t + pop + t:pop + female**				
		pop + female	t	3	1,163.87	< 0.001
		t + female	pop	1	246.56	< 0.001
		t + pop + female	t:pop	3	99.25	< 0.001
		t + pop + t:pop	female	1	410.42	< 0.001
	Vendace & European Whitefish	**t + sp + t:sp + female + male**				
		sp + family + female	t	3	980.02	< 0.001
		t + family + female	sp	1	805.34	< 0.001
		t + sp + family + female	t:sp	3	107.58	< 0.001
		t + sp + t:sp + female	family	1	33.33	< 0.001
		t + sp + t:sp + family	female	1	6.96	0.008

Table 6. Phenotypic variation component correlation directions from increasing incubation temperature for embryo survival (%), incubation period (number of days post-fertilization; DPF), incubation period (accumulated degree-days; ADD), length-at-hatch (mm), and yolk-sac volume (mm^3) from Lake Southern Konnevesi vendace (LK-Vendace (*Coregonus albula*)), Lake Superior cisco (LS-Cisco (*C. artedi*)), and Lake Ontario cisco (LO-Cisco). - indicates a negative correlation, + indicates a positive correlation, and 0 indicates no correlation.

Trait	Study Group	Correlation Direction			
		Dam	Sire	Dam:Sire	Error
Embryo Survival	LK-Vendace	-	+	0	+
	LS-Cisco	-	0	0	0
	LO-Cisco	0	+	0	+
Incubation Period (DPF)	LK-Vendace	0	0	+	0
	LS-Cisco	-	0	0	0
	LO-Cisco	-	0	0	+
Incubation Period (ADD)	LK-Vendace	0	-	+	0
	LS-Cisco	-	0	0	0
	LO-Cisco	-	-	0	+
Length-at-Hatch	LK-Vendace	0	-	+	0
	LS-Cisco	-	0	0	0
	LO-Cisco	-	0	+	0
Yolk-sac Volume	LK-Vendace	+	0	+	-
	LS-Cisco	-	+	-	+
	LO-Cisco	0	0	0	0

FIGURE CAPTIONS:

Fig. 1 Map showing the location of each lake (LS = Lake Superior; LO = Lake Ontario; LK = Lake Southern Konnevesi) sampled in North America (A) and Europe (B)

Fig. 2 Winter water temperatures from Lake Southern Konnevesi, Lake Superior, and Lake Ontario. Lake Southern Konnevesi and Lake Superior data were recorded using *in-situ* sensors on the lakebed (10-m deep). Lake Ontario data was recorded using remote sensing sea surface temperatures. X indicates *ca.* spawning and O indicates *ca.* hatching. The earlier spawning (X) in Lake Southern Konnevesi indicates vendace (*Coregonus albula*) and latter for European whitefish (*C. lavaretus*)

Fig. 3 Mean embryo survival (%) and incubation period (number of days post-fertilization (DPF) and accumulated degree days (°C; ADD)) at each incubation temperature (°C; left) and standardized temperature responses within each study group (%; right) from Lake Southern Konnevesi (LK-Vendace (*Coregonus albula*) and LK-Whitefish (*C. lavaretus*)), Lake Superior (LS-Cisco (*C. artedi*)), and Lake Ontario (LO-Cisco). Error bars indicate standard error

Fig. 4 Mean percent of total phenotypic variation across incubation temperatures for embryo survival, incubation period (number of days post-fertilization (DPF) and accumulated degree days (°C; ADD)), length-at-hatch (mm), and yolk-sac volume (mm^3) from Lake Southern Konnevesi vendace (LK-Vendace (*Coregonus albula*), Lake Superior cisco (LS-Cisco (*C. artedi*)), and Lake Ontario cisco (LO-Cisco). Error bars indicate standard error

Fig. 5 Mean length-at-hatch (mm; LAH) and yolk-sac volume (mm^3; YSV) at each incubation temperature (°C; left) and standardized temperature responses within each study group (%; right) from Lake Southern Konnevesi (LK-Vendace (*Coregonus albula*) and LK-Whitefish (*C. lavaretus*)), Lake Superior (LS-Cisco (*C. artedi*)), and Lake Ontario (LO-Cisco). Error bars indicate standard error