Soil respiration variation along an altitudinal gradient in Italian Alps: Disentangling forest
 structure and temperature effects

3

Aysan Badraghi¹, Maurizio Ventura¹, Andrea Polo¹, Luigimaria Borruso¹ and Leonardo
 Montagnani^{1,2*}

6 ¹ Faculty of Science and Technology, Free University of Bozen-Bolzano, Piazza Università 1,

7 39100 Bolzano, Italy

⁸ ² Forest Services, Autonomous Province of Bolzano, Via Brennero 6, 39100 Bolzano, Italy

9 *Corresponding author

10 E-mail: <u>Leonardo.montagnani@unibz.it</u> (LM)

11

12 [¶] These authors contributed equally to this work.

13

14

15 Abstract

To understand the main determinants of soil respiration (SR), we investigated the changes of soil 16 respiration and soil physicochemical properties, including soil carbon (C) and nitrogen (N), root 17 C and N, litter C and N, soil bulk densities and soil pH at five forest sites, along an 18 elevation/temperature gradient (404 to 2101 m a.s.l) in Northern Italy, where confounding 19 20 factors such as aspect and soil parent material are minimized, but an ample variation in forest 21 structure and composition is present. Our result indicated that SR rates increased with temperature in all sites, and about 55% - 76% of SR was explained by temperature. Annual 22 cumulative SR, ranging between 0.65 and 1.40 kg C m⁻² yr⁻¹, declined along the elevation 23 24 gradient, while temperature sensitivity (Q10) of SR increased with elevation. However, a high SR rate (1.27 kg C m⁻² yr⁻¹) and low Q10 were recorded in the old conifer forest stand at 1731 m 25 26 a.s.l., characterized by a complex structure and high productivity, introducing nonlinearity in the relations with elevation and temperature. Reference SR at the temperature of 10°C (SR_{ref}) was 27

not related to elevation. A significant linear negative relationship was found for bulk density 28 with elevation. On the contrary, soil C, soil N, root C, root N, pH and litter mass were better 29 30 fitted by nonlinear relations with elevation. However, it was not possible to confirm a significant correlation of SR with these parameters once the effect of temperature has been removed (SR_{ref}). 31 These results show how the main factor affecting SR in forest ecosystems along this Alpine 32 33 elevation gradient is temperature, but its regulating role can be strongly influenced by site biological characteristics, particularly vegetation type and structure. This study also confirms 34 that high elevation sites are rich in C stored in the soil and also more sensitive to climate change, 35 being prone to high carbon losses as CO₂. Conversely, forest ecosystems with a complex 36 structure, with high SR_{ref} and moderate Q10, can be more resilient. 37

38 Keywords: Soil respiration; temperature; forests; soil; environmental factors; elevation gradient

39 Introduction

Soil respiration (SR) is the largest biological carbon (C) flux after photosynthesis in terrestrial 40 41 ecosystems [1]. It largely determines the C balance between the terrestrial biosphere and the atmosphere [2,3,4] and assumes a decisive role in the carbon cycle and terrestrial carbon sink 42 43 capacity. The soil is the largest C pool in the terrestrial biosphere and has been increasingly recognized to play a crucial role in mitigating global warming resulting from climate change [5, 44 45 6, 7]. Small changes in soil CO2 efflux or soil organic C stocks could severely impact the global C cycle [8]. In this regard, SR is one of the fluxes that have received more attention by research 46 for a longer time. The study by Janssens et al. [9] evidenced the relevant role of forest 47 productivity in the determination of SR. Since then, other studies have investigated and 48 49 quantified the impact of productivity on SR modeling [10, 11]. Apart from productivity, SR is influenced by different abiotic and biotic factors such as soil temperature, moisture, and 50 microbial community, introducing a considerable uncertainty in SR estimates [12,13,14]. Among 51 these factors, the temperature has been the most often studied factor affecting respiratory 52 processes [15,16]. Predicting the SR response to increasing temperature (temperature sensitivity 53 of SR) has been one of the main objectives of research for years; therefore different equations 54 relating soil CO2 efflux with temperature [17,18,19] or with a combination of temperature and 55 soil humidity have been developed [20]. Nevertheless, the Q10 function [21] using the Q10 56

parameter to describe the temperature sensitivity of SR is one of the most widely used models,
still mainly employed to quantify the CO2 efflux from the soil in Earth system models.

On the other hand, the elevation is a key driver of climate properties. It plays an essential role in 59 the soil organic matter distribution and may dampen the effects of climate change [14,22,23,24, 60 61 25]. In general, temperature declines with elevation, thus elevation gradient has been used to assess soil respiration response to temperature in several studies [12,26,27,28]. These studies 62 indicate that CO2 exchange between soil and atmosphere varies along climatic gradients and that 63 temperature sensitivity (Q10) of SR increases with elevation. They also found a positive 64 relationship between soil organic matter (SOM) and elevation and reported that global soil 65 66 organic C stock at high elevation is more sensitive to climate change and is predicted to decrease in a warming climate [14,23,29,30,31,32]. However, several researchers have reported opposite 67 trends and found lower SOM and higher SR at a higher elevation [30,33,34]. This variability 68 may be partially due to confounding factors affecting SR other than temperature. Besides 69 70 elevation, mountain landscapes are, in fact, characterized by substantial changes of other site parameters such as slope and aspect, which can affect microclimatic conditions and, therefore 71 72 soil C dynamics [35]. Furthermore, due to the heterogeneity of geological substrates, soils of mountain regions are highly diverse over short spatial scales and this can generate marked 73 74 contrasts in soil biogeochemical functions [36]. Different results have also been found when SR is related to soil organic carbon [26,37]. 75

76 Besides, there is evidence that diverse plant biome types can influence SR rate differently. Therefore, the various plant communities can affect differently, microclimate, soil and litter 77 composition, and root distribution, therefore affecting soil respiration rate [18, 26,38,39,40]. 78 79 However, within the same plant biome, there is a high spatial heterogeneity of SR. Some authors 80 found a possible linkage between the topography, plant community structure (e.g., forest type 81 and speed of regeneration), and SR within the same forest ecosystems [18,38,39,40]. Further, forest management can also play a crucial role in SR [41]. For instance, tree removal can directly 82 influence soil respiration due to the removal itself (i.e., reduction of plant biomass) but even 83 indirectly changing the soil's physicochemical properties and micrometeorological conditions 84 85 [42].

Currently, the temperature dependency of SR and SOC decomposition is a major interest 86 regarding global climate change and the role of terrestrial ecosystems in regulating Earth's 87 88 climate [43,44]. Therefore, there is a need to better understand the interactions between temperature and soil CO2 efflux. The general goal of this study is to disentangle the possible 89 multi-effects on SR of soil properties, temperature, SOM, and vegetation structure (tree height in 90 particular) along a plant biome-elevation gradient. In particular, the existing differences in 91 vegetation structure allowed us to investigate the extent to which these biological variables and 92 the induced variation in microclimatology can alter the relation between elevation and SR. 93

Specifically, i), we tested the hypothesis that SR and SOM accumulation change linearly with 94 95 elevation. We also hypothesized that the Q10 value increases linearly with elevation as well. Furthermore, ii) we analyzed which are the main factors affecting SR other than temperature. To 96 97 better isolate the effect of temperature on SR, the study was conducted along an altitudinal/temperature gradient in Italian Alps, in conditions where confounding factors like 98 99 slope, aspect, and soil parent material are minimized. The differences in vegetation structure allowed us to investigate to which extent these biological variables and the induced variation in 100 101 microclimatology, can alter the relation between elevation and SR.

102 Material and Methods

103 Study areas

Five experimental sites were established between the top of the Rittner Horn mount and the city of Bolzano, Italy, on the southern side of the Alps (Fig 1a). The overall elevation gradient between the highest and the lowest site is 1700 m and the elevation separation between each site is approximately 420 ± 60 m. All sites are characterized by a soil evolved upon a glacial till laid on a porphyric bedrock and an SSE slope orientation. Annual precipitation ranges between 800 and 1000 mm.

110

Fig 1. (a) Map showing the research site locations selected for this study (b) Scheme of the elevation and vegetation gradient present along the studied Alpine slope.

Site A was established in shrubland vegetation of Dwarfing Mountain pine (Pinus mugo Turra) 114 115 near the summit of the Rittnern Horn/Corno del Renon mount (see the scheme in Fig 1b). Site B was established in a Norway spruce stand (Picea abies (L.) Karst.) at the long-term Fluxnet 116 research station of Renon- Mittelgrünwald/Selva Verde (https://doi.org/10.18140/flx/1440173;). 117 The site is characterized by an unvenaged distribution of tree diameters, approaching the 118 119 structure of old-growth forest stands (45,46]. Site C was located near the location of Riggermoos 120 (Oberbozen/Sopra Bolzano), in a low-density Scots pine (*Pinus sylvestris* L.) stand. Site D was 121 established in a mixed stand of Sessile oak (Quercus petrea (Matt.) Liebl.) and Chestnut (Castanea sativa L:) and the presence of Scots pine near the village of Signat/Signato. Site E 122 123 was located in a stand dominated by Downy oak (Quercus pubescens Willd.) and Flowering ash (Fraxinus ornus L.) on the hill slope just close to the city of Bolzano (Sankt Magdalena/Santa 124 125 Maddalena). All sites except site A are managed as high forest mainly for wood harvesting. Site A is managed as natural vegetation with occasional harvesting only at forest margins to avoid 126 127 expanding the pines in the adjacent pasture areas. Tree age was assessed in 2018 by tree ring count: it was found that site B had the oldest trees, slightly above 200 years, while other stands 128 were in the range of 50-100 years. Tree height was assessed in 2020 with the TruePulse sensor 129 (TruPulse 360 B laser range-finder, Laser Tech, Colorado, USA). Details on tree heights and the 130 main characteristics of the research sites are reported in Table 1. 131

Characteristics	Site A	Site B	Site C	Site D	Site E
Elevation	2101	1731	1354	865	404
(m a.s.l)					
Mean annual Temperature	4	4	12	11	14

132 Table 1. General characterization of the study sites.

(C)					
Land use	Shrubland	Forest	Forest	Forest	Forest
Mean dominant tree height (m)	1.5	29	18	15	5
Dominant species in the overstory	- Dwarfing Mountain pines (<i>Pinus</i> <i>mugo</i>)	 Norway spruce (<i>Picea</i> <i>abies</i>) Swiss stone pine (<i>Pinus</i> <i>cembra</i>) Larch (<i>Larix</i> <i>decidua</i>) 	- Scots pine (<i>Pinus</i> sylvestris)	 Sessile oak (Quercus petrea) Scots Pine (Pinus sylvestris) Chestnut (Castanea sativa) 	 Downy oak (Quercus pubescens) Flowering ash (Fraxinus ornus)
Main species in the understory	- Intervening grasses (Festuca halleri)	- Rusty leaved alprose (<i>Rhododendron</i> <i>ferrugineum</i>)	- Heather (<i>Erica carnea</i>)	Understory almost absent	 Smoke-bush (<i>Cotinus</i> <i>coggyria</i>) Succulent plants (<i>Opuntia</i> <i>humifusa</i>)

133

134 Soil respiration measurements

To quantify SR, ten iron collars (10 cm height, 20 cm diameter) were inserted in the soil, three weeks before the first measurements at each site. Measurements were performed with an opaque

survey chamber (Li-8100-104, LI-COR Biosciences, Nebraska, USA) connected to an LI-8100 137 analyzer (LI-COR Biosciences, Nebraska, USA). On each collar, the measurement period was 138 139 set to 120 s the first 20 s of the measurement were considered dead-band, so the flux computation was limited to 80 s. See Montagnani et al. [8] for further details about the 140 measurement settings. Starting on July 21, 2017, SR measurements were performed periodically, 141 about once per month, until July 20, 2018, for a total of 17 measurement days. The first 4 142 measurement series were performed every three weeks at all the sites. During the winter period, 143 SR measurements were performed only in the locations at a lower altitude because of snow in the 144 high-elevation locations. The measurement calendar for the different sites is provided in S1. 145 During measurements, the air temperature was measured inside the survey chamber (at 0.1 m 146 above ground, RHT Plus, Skye Instruments, UK). A soil temperature profile was installed at 147 control site B according to ICOS protocol [47]. Specifically, we used the -5 cm soil T data 148 provided by a CS605 probe, Campbell Scientific, USA. We recorded soil temperature 149 continuously at 30 min intervals for the whole experimental period (July 2017 - July 2018). In 150 addition, during the period May-July 2018, we placed at all the sites iButton sensors (Maxim 151 152 integrated, USA) at 5 cm below the soil surface.

153 Soil sampling and analysis

154 At the end of the last measurement session (July 2018), leaf litter present in each collar was sampled. The soil in each collar was sampled using a 4.8 split-corer (Eijkelkamp, NL), until 20 155 156 cm depth. In the laboratory, soil samples were weighed and sieved at 2 mm mesh size to separate roots, stones, and coarse organic matter fragments. Collected leaf litter and fine roots (< 2 mm 157 158 diameter) were weighted after oven-drying at 105 ± 5 °C. Soil bulk density was determined by dividing the weight of sieved soil by the core volume. Soil pH was measured using a pH-meter 159 (CRISON pH-Meter Basic 20+Electrod: Hach 50 10T CRISON, Barcelona, Spain) in water. 160 Root, litter, and soil samples were analyzed for organic C and N content using a FlashEATM 1112 161 162 Elemental Analyzer (Thermo Fisher Scientific, Waltham, MA, USA).

163 Data elaboration and statistical analysis

164 The mean amount $(\pm SD)$ of accumulated C and N stock in the soil, root, and litter were 165 computed for each site. Soil and C and N stocks were obtained as follows [48]:

166 Soil stock
$$(kg \, dm^{-3}) = C/100 \times BD_{soil} (kg \, m^{-3}) \times 0.2 (m)$$
 Eq.1

167 Where C is the mean soil organic C or total N content, BD_{soil} is the soil bulk density (kg dm⁻³) 168 and 0.2 m is the sampling depth.

Root C and N stocks were determined with the same computational approach, using root density (kg dm⁻³) in place of soil BD and root C and N content in place of soil C and N content. Litter C and N amounts were obtained by multiplying the litter mass by C or N content and dividing by the collar area.

Soil respiration data collected from each measurement point (collar) were related to chamber airtemperature using a logistic model [26]:

175
$$SR = \frac{a}{(1+b \cdot \exp(-k \cdot T_a))}$$
 Eq. 2

Where SR is soil respiration, a is the maximum value of SR, b determines the elongation of the
SR curve along the x-axis, k is the logistic growth rate or steepness of the SR curve along the xaxis, Ta is air temperature. Furthermore, SR data were also fitted with a Q10 model [49,50]:

179
$$SR = SR_{ref} \cdot Q_{10}^{\left(\frac{(Ta-Tref)}{10}\right)}$$
 Eq.3

Where SR is the soil respiration, SR_{ref} is the fitted SR at the reference temperature of 10°C (T_{ref}), Q10 is the temperature sensitivity of SR, defined as the factor by which soil respiration increases with a 10°C temperature increase, and Ta is chamber air temperature.

Models were fitted to SR data using the nls package in R software. Model fitness was evaluated 183 184 based on Akaike's Information Criterion (AIC), R-squared (R²), Mean Absolute Error (MAE), and Root Mean Squared Error (RMSE). The Q10 model was used to obtain the SR_{ref} and the Q10 185 value for every collar. Linear regression was used to compare the air temperature measured 186 continuously in the reference plot (site B) and air temperature inside the chamber during SR 187 188 measurements for each collar in each site. The obtained linear regression models were then used to predict chamber air temperature for the whole experimental period, for each collar, with a 30 189 min time resolution. Therefore, the predicted chamber air temperature was used to predict SR 190

values simultaneously for the whole experimental period, based on the logistic models relating SR with chamber temperature. For some collars, it was not possible to obtain a good fitness of the SR data using the logistic model; for these collars, the prediction soil respiration data from temperature was performed using the Q10 model developed for the same collar. Finally, the total cumulative SR for the whole experimental period was determined for each collar at each site.

Soil respiration response to biological variables (soil C, root C, litter C, root dry weight, soil N, 196 197 root N, litter N, litter dry weight) was examined using Spearman's Correlation Test and linear mixed-effects models (LMMs) fitted by restricted maximum likelihood (REML). Before 198 applying LMMs, to avoid statistical errors, Variance Inflation Factor (VIF) was determined for 199 biological variables, and variables with high VIF values were excluded from the model 200 201 assessment. LMMs were built using the lme4 R package [50,51,52]. The models consisted of both fixed and random effects: biological variables were considered as fixed effects, and 202 203 sampling plots (collars) nested in each site were used in the random-effects formula. R² was used to summarize model goodness-of-fit together with AIC [53,54]. Since computed R² by LMMs 204 205 are a pseudo- R^2 and technically incorrect, the r2glmm R package was used to computing R^2 [54]. To exclude the confounding effect of temperature from LMMs and correlations tests, 206 207 environmental variables were related to SR_{ref} instead of SR [20,26,55]. Furthermore, to assess GPP and SR's correlation, tree height was used as a covariate in LMMs and Spearman test. 208 209 Statistical comparisons of average soil C and N, root C and N, litter C and N, soil bulk density, soil pH, and soil respiration in the different sites were performed by the Kruskal-Wallis test 210 (Dunn test, p < 0.05) for the non-normally distributed data, and a one-way ANOVA for the 211 normally distributed data (Tukey test, p < 0.05). The normality of the data and homogeneity of 212 213 variance was checked by the Shapiro-Wilk test and Levene's test, respectively [56,57]. To check linearity changes of SR, Q10, SOM with elevation, linear and nonlinear polynomial regressions 214 were applied between elevation and environmental variables (soil C and N, root C and N, litter C 215 and N, soil bulk density, soil pH). The linearity changes of these variables with elevation were 216 detected based on the lowest AIC and the highest R². The association of Q10, soil C and soil N 217 with environmental variables were determined using Spearman's Correlation Test. All statistical 218 analyses were performed using R version 3.6.0 ([59], www.r-project.org). 219

221 **Results**

222 Environmental factors variability along the altitudinal gradient

A significant difference in soil C stock was found only between site E (3891 ± 2756 g C m⁻²), at 223 the lowest altitude, where the C stock was smaller in comparison to sites A, B, and C (Fig 2a). 224 No significant differences were found for soil N stock in the different sites (Fig 2e). Root 225 biomass and root C and N stocks in site A were significantly higher than other sites (Fig 2b, d, f) 226 and litter mass in site B was significantly higher than site E (Fig 2h). However, the accumulated 227 C and N in the litter were not significantly different along the altitudinal gradient (Fig 2c, g). 228 Significant differences were found between pH values in the different sites: the lowest value of 229 soil pH was measured in site B (3.8 ± 0.1) and the highest value in site E $(5.9 \pm 0.2, \text{ Fig 2i})$. The 230 highest bulk density value was found in site D (1.01 ± 0.27 g cm⁻³, Fig 2i) and the lowest was 231 found in site B (0.26 ± 0.16 g cm⁻³; Fig 2j). 232

233

234

Fig 2. Stocks of (a), soil C (b,) root C (c), litter C (d), root dry weight (e,) soil N (f), root N (g), litter N (h), litter dry weight (i), soil pH, and soil bulk density (j), estimated in the different sites (A-E). Different lowercase letters indicate significant differences between sites according to the Kruskal-Wallis or ANOVA tests. Vertical bars represent the standard deviation of the mean for each site.

Based on the AIC and R², the linear relationship with elevation appeared in model selection only for soil bulk density i.e., soil bulk density resulted to be linearly related to elevation (Table 2). On the contrary, soil C, soil N, fine root mass, and root C, root N, pH and litter mass data were fitted better with nonlinear relations with elevation (Table 2; more detail about equations can be found in Table S1; supplementary material). Furthermore, a significant negative correlation was found between soil C and soil N with soil pH, mean dominant tree height, and bulk density (Table 3).

Table 2. linear and polynomial regressions between the different parameters tested and elevation. Asterisks indicate significance levels: $* - p \le 0.05$, $**- p \le 0.01$, and $*** - p \le$ 0.001, n.s. – nonsignificant. BD – soil bulk density; Q10 – temperature sensitivity of soil respiration; SR- cumulated soil respiration; SR_{ref} – respiration at the 10°C reference temperature; Im – linear regression model; Poly-Im – polynomial regression model. The model selection was based on the lowest AIC and the highest R² (in bold).

254										
	Parameters		Elevation							
255		Signific	cance	Multiple R ²		AIC				
256		lm	Poly-lm	lm	Poly-lm	lm	Poly-lm			
	Soil C (g C m ⁻²)	***	***	0.28	0.40	764	758			
257	Soil N (g N m ⁻²)	*	*	0.15	0.20	510	509			
258	SR (kg C m ⁻² yr ⁻¹)	***	***	0.33	0.36	34	34			
259	Q10 (kg C m ⁻² yr ⁻¹)	***	**	0.28	0.41	77	67			
	$SR_{ref}(kg C m^{-2} yr^{-1})$	n.s.	n.s.	< 0.01	0.08	139	139			
260	Fine root mass (g m ⁻²)	*	**	0.10	0.24	717	712			
261	Root C (g C m ⁻²)	*	***	0.39	0.80	199	186			
262	Root N (g N m ⁻²)	n.s.	***	0.28	0.80	102	88			
202	Litter mass (g m ⁻²)	n.s.	*	0.07	0.12	774	772			
263	Litter C (g C m ⁻²)	n.s.	n.s.	< 0.01	0.04	222	223			
264	Litter N (g N m ⁻²)	n.s.	n.s.	< 0.01	< 0.01	112	114			
265	рН	**	***	0.63	0.88	22	8			
265	BD (kg dm ⁻³)	***	***	0.32	0.29	9	11			

266

267

Table 3. Spearman rank coefficients for the correlations between different variables (RDW
- root dry weight, LDW – litter dry weight, BD – soil bulk density, Q10 – temperature
sensitivity of soil respiration). Asterisks indicate significance levels: * - p ≤ 0.05, **- p ≤
0.01, and *** - p ≤ 0.001; n.s. – nonsignificant.

Variable	Soil	Soil	RD	Root	Root	LD	Litter	Litte	рН	BD	Tree
	С	Ν	W	C	Ν	W	С	r N			height
Soil C	1.00	0.94	0.14	-	-	0.02	-0.17	-	-	-	0.33
	-	***	n.s.	0.11	0.07	n.s.	n.s.	0.09	0.62	0.52	*
				n.s.	n.s.			n.s.	*	*	
Soil N	0.94	1.00	0.11	0.36	0.32	0.01	-0.36	-	-	-	0.21
	***	_	n.s.	n.s.	n.s.	n.s.	n.s.	0.23	0.54	0.39	n.s.
								n.s.	*	*	
Q10	0.15	0.13	0.06	0.37	0.46	-	0.01	-	-	-	-0.29
	n.s.	n.s.	n.s.	n.s.	n.s.	0.08	n.s.	0.44	0.29	0.23	*
						n.s.		n.s.	n.s.	n.s.	

272

273 Soil respiration

Both logistic and Q10 models confirmed that soil respiration rates increased with temperature in
all sites (Fig 3), and the seasonal pattern of SR was similar to that of air chamber temperature
(S1 Fig).

277

278

279

Fig 3. Rates of soil respiration against chamber air temperature in the different sites along the elevation gradient (A-E). The regression lines in the plots were built using the mean values of model parameters (R_{ref} and Q10 value for the Q10 model and a, b, and k value for the logistic model) obtained for different replicate collars of each site (n = 10).

A strong linear relationship was found between observed and predicted SR ($R^2 = 0.73$; S2 Fig; supplementary material). Temperature explained between 55 % and 76% of the variance in soil respiration at the experimental sites (Table 4). The Q10 and SR_{ref} values obtained for the different sites ranged between 1.75 and 2.96, and between 2.17 and 4.49, respectively (Table 4). The Q10 value recorded in site A (highest elevation) was significantly different from the others (Table 4). A significant linear relationship was identified between Q10 and elevation (Table 2). However, the trend of Q10 against temperature was better described by a nonlinear relation (Table 2). Furthermore, a significant negative correlation was found between Q10 and mean dominant tree height (Table 3). No significant relationship between SR_{ref} and altitude was found (Table 2). However, significant differences were found between experimental sites, as the highest SR_{ref} value was recorded in site B and the lowest values in site A, C, and E (Table 4).

Table 4. Mean values of Q10 (temperature sensitivity) and SR_{ref} (soil respiration at the temperature of 10°C) for each site; and the mean value of MAE (Mean Absolute Error), Root Mean Squared Error (RMSE), R² (R-square), and AIC (Akaike information criterion) for each model (logistic and Q10). Different letters indicate significant differences between sites. Values are expressed as mean \pm SD.

	A	В	С	D	Е
Q10	2.96 ± 0.72^{b}	1.83 ± 0.25^{a}	1.90 ± 0.24^{a}	1.75 ± 0.15^{a}	1.75 ± 0.32^{a}
SR _{ref}	2.17 ± 0.60^{a}	$4.49 \pm 0.71^{\circ}$	2.24 ± 0.58^a	3.17 ± 1.03^{b}	2.98 ± 0.41^{ab}
AIC logistic model	32.95 ± 10.45	34. 25 ± 6.45	34.19 ± 11.65	48.00 ± 14.55	49.76 ± 6.70
AIC Q10 model	30.42 ± 13.69	34.07 ± 7.31	35.39 ± 11.53	48.44 ± 14.13	49.11 ± 6.60
R ² logistic model	0.76 ± 021	0.75 ± 0.20	0.75 ± 0.16	0.67 ± 0.11	0.58 ± 0.10
R ² Q10 model	0.75 ± .23	0.71 ± 0.23	0.68 ± 0.21	0.61 ± 0.11	0.55 ± 0.11
MAE logistic model	1.06 ± 0.65	1.06 ± 0.45	0.68 ± 0.32	1.26 ± 0.62	1.29 ± 0.25
MAE Q10 model	1.07 ± 0.68	1.14 ± 0.52	0.79 ± 0.36	1.46 ± 0.76	1.39 ± 0.26
RMSE logistic model	1.36 ± 0.94	1.27 ± 0.50	0.86 ± 0.44	1.70 ± 0.89	1.71 ± 0.36
RMSE Q10 model	1.38 ± 1.0	1.36 ± 0.55	0.99 ± 0.49	1.87 ± 1.02	1.77 ± 0.38

The cumulative SR in site A was significantly lower than in the other sites (Fig 4). The results of the statistical analysis also confirmed a nonlinear relationship between cumulative SR and elevation (Table 2).

304

305

Fig 4. Total cumulative soil respiration (kg C m⁻² yr⁻¹) calculated for the different sites (A-E). Values are indicated on the bars. Error bars indicate standard deviation. Different letters on the bars indicate significant differences between sites according to the ANOVA test.

Soil C, mean dominant tree height, and litter dry weight resulted in the best variables to explain SR (SR_{ref}; at 10 C) in LMMs (VIF<10; Table 5). According to the model, about 0.22 of SR was explained by tree height ($R^2 = 0.22$, Table 5). Meanwhile, a positive association between SR and mean dominant tree height and a negative association between SR and root C and N were found by Spearman's Correlation Test (Fig 5).

315

Table 5. Results of linear mixed-effects models testing biological variables impact on the SR_{ref}; at 10 C. VIF – Variance Inflation Factor. Parameters in bold show significant correlations

Model variables	Value	VIF	p-value	R ²
Intercept	1.61		0.003	0.34
Soil C	< - 0.01	1.20	0.92	< 0.01
Tree height	0.06	1.28	0.002	0.22
Litter dry weight	< 0.01	1.07	0.14	0.06

319

- 321 322
 - 323

Fig 5. Relations of reference soil respiration at 10°C (SR_{ref}) and different site properties.

325 **Discussion**

Organic C and N content along the altitudinal gradient

Overall, our data confirms an increase of the soil organic C stock in soil with increasing 327 elevation [7,24, 25,27,59,60,61,62]. The low temperature can limit the decomposition of organic 328 matter at high altitudes, as the temperature is the main driver for the loss of the soil organic C. 329 For this reason, altitude could induce a significant increase in SOM [23,59,62,63,63]. However, 330 331 SOM increment was not linear along the elevation gradient and therefore, our first hypothesis was not confirmed. The Nonlinear relationship of SOM with elevation and the high value of soil 332 C in site C (with a mean temperature of 12 C) suggests that other factors other than temperature 333 have influenced SOM accumulation. Different microclimatic or micromorphological conditions 334 caused by differences in slope and aspect can influence C storage in soils [33,35,65]. However, 335 in the present study, all the sites are characterized by similar slopes and by the same south or 336 south-east facing; therefore, we tend to exclude an influence of these factors on soil C 337 accumulation in the examined sites. According to recent studies, SOM is not consistently related 338 to variation in climatic conditions along elevation gradients; it is also strongly affected by 339 productivity or by vegetation type/composition [30,33,66,67]. Our data confirms a significant 340 positive correlation between mean dominant tree height and soil C (Table 3). Therefore, the 341 higher amount of soil C found in the present study at intermediate altitude could be explained by 342 higher site productivity in sites B and C in particular, which is also suggested by the mean 343 344 dominant tree height.

Soil pH and bulk density are considered two of the main variables influencing other soil properties, soil microbial activity and soil respiration [68,69]. Generally, at high elevation, the higher precipitation and lower evapotranspiration rate decrease soil pH by increasing the leaching of basic cations [65,70,71,72,73]. This is confirmed by the strong relationship between

elevation and soil pH found in the present study (Table 2, Fig 2). In addition, soil bulk density was significantly diminished by increasing elevation (Fig 2; Table 2). One of the main factors affecting soil bulk density is SOM content [74]. Therefore, the lowest values of soil bulk density at high elevation could be explained by the high amount of soil C, as confirmed by the negative association found between soil C and soil bulk density (Table 3), previously reported in other studies [7, 61,62,75].

355 Factors affecting soil respiration

The total SR observed in the sites is within the range reported for similar forests [26,60]. The 356 decrease in SR along the elevation gradient observed in the present study could be explained by 357 the reduction of temperature along elevation. In fact, temperature resulted in the main controlling 358 factor on SR, explaining most of the variability of SR. This result is in agreement with other 359 studies performed along altitudinal gradients, reporting that temperature can explain between 360 55% and 76% of the SR variability [6, 9, 14,27,63, 76]. On the other hand, the annual cumulative 361 362 SR in site B was 2 times larger than in site A, which has the same mean annual temperature. This finding, together with nonlinear SR concentrations with elevation/temperature, suggests that 363 364 other environmental factors can have a role in regulating SR [34,60]. For instance, Grand et al. [36] reported that soil and vegetation heterogeneity strongly affect soil carbon efflux in complex 365 366 geomorphic terrain. In the present study, all five sites were established on relative homogeneity of soil substratum; therefore a high SR rate in site B could not be resulted by the confounding 367 368 role of the soil parent material. Site B is an uneven-aged dense forest stand, and the mean 369 dominant tree height is approximately 29 m (Table 1). Since tree height can be used as a proxy 370 of gross primary production (GPP), the high SR rate in site B could be attributed to high GPP [9,29,66,77,78,79], which can provide substrates for root and microbial respiration through 371 photosynthesis [66,80]. This finding is confirmed by LMMs and correlation test evidencing a 372 significant positive relation between SR_{ref} and mean dominated tree height, therefore indicating 373 374 that, after removing the effect of temperature, productivity results in one of the main factors affecting SR. 375

At a global scale, SR has been related to soil C, litter production and pH, and negatively correlated with soil bulk density; therefore, a high value of soil C and litter accumulation could lead to an increase in soil respiration [20,66,81,82,83]. In the present study, the highest amount

of soil C and dry litter weight were also observed in site B. However, we could not find a significant correlation between SR and soil C and litter dry weight (Fig 5; Table 5).

An increase in SR has also been observed as a consequence of increasing soil pH between 4 and 381 7, because of the positive effect of pH on soil microbial activity within this range 382 383 [1,68,84,85,86]. In contrast, SR and bulk density are generally negatively correlated, as a low SR indicates increasing rates of SOM accumulation and therefore a decrease in bulk density [82]. 384 Furthermore, SR declines with increasing bulk density due to the lower soil porosity and oxygen 385 availability for microbial activity in compacted soils [18,59,87]. However, our analysis did not 386 387 confirm a significant correlation of SR with pH and bulk density (Table 5; Fig 5). Prediction of 388 SR is difficult because of a range of factors such as aspect, slope, and soil types [35,36,60,65]. In the present study, by minimizing the confounding role of these parameters, we conclude that the 389 most important controlling factors on SR along an Alpine elevation gradient were temperature 390 and vegetation type/composition or GPP. 391

Temperature sensitivity of soil respiration (Q10)

The temperature sensitivity of SR is an important ecological model parameter, and according to 393 previous studies, its value is mainly controlled by temperature [14,17,88,89]. The O10 and SRref 394 values found at site B substantially confirm the values found by Acosta et al. [90] at the same 395 site (2.0 and 4.09, respectively). Although at a smaller spatial scale, Acosta et al. [90] also found 396 an increasing SRref as a function of stand age (and consequently height). The significant trend of 397 Q10 with elevation in this research confirms the results of previous studies and the higher 398 sensitivity of high elevation ecosystems to global warming [14,91,92,93,94]. However, the only 399 significant difference was found between the Q10 value at site A (higher elevation) and the other 400 sites (Table 4). The Q10 value in site A was also significantly higher than site B, characterized 401 by a similar mean temperature. According to different studies, Q10 is negatively related to pH 402 and positively dependent on soil C [55,95]. However, the amount of soil C and pH could not be 403 the reasons for the lower value of Q10 in site B, which is characterized by a lower pH value and 404 405 by a similar amount of soil C. Temperature sensitivity of SR can also be affected by forest structure [96]. Dense forest stands with a closed canopy can create a specific understory 406 microclimate by providing a cool shelter during heat waves, which can decrease daily maximum 407 air temperature by up to 5.1 °C [97,98]. Therefore, we can hypothesize that the dense forest stand 408 409 site B, with the highest mean dominant height, is less sensitive to global warming. This feature was also described by Niu et al. [99] for the same site and confirmed by a significant negative 410 411 correlation between Q10 and mean dominating tree height (Table 3). Whether a sylviculture that maintains high and old forests with a complex structure also achieves a lower sensitivity to 412 413 climate change by reducing the depletion of C stored in the soil, is a novel research question emerging from this research. If the findings obtained here will be confirmed, it would imply that 414 more conservative forest management can not only maintain current C stocks in the biomass but 415 also leads to a reduced sensitivity to the temperature of the C stored in the soil. 416

417 **Conclusions**

419

In this study, a significant nonlinear relationship between SR, SOM, Q10, and elevation was 420 421 detected along the examined Alpine altitudinal gradient, rejecting our initial linearity hypothesis. Our data confirmed a negative trend between SR and elevation. On the contrary, SOM and Q10 422 423 showed a positive trend with elevation. These results lead us to conclude that temperature is the major controlling factor on the annual soil respiration, Q10, and SOM, but its regulating role 424 425 may be strongly affected by site biological characteristics, particularly by GPP or vegetation type/composition. The high value of Q10 detected at high elevation confirmed a higher potential 426 427 vulnerability of high mountain ecosystems to climate change, where small temperature changes can induce stronger increase CO2 emissions. However, the site with the highest tree age and 428 429 height and more complex structure showed high SRref and moderate Q10, indicating that the length in the life cycle and related changes in forest structure can dampen, to some extent, the 430 effects of climate change on ecosystems and decrease the positive feedback due to soil CO2 431 emissions to the atmosphere. 432

433

434 **References**

Luo Y, Zhou X. Soil respiration and the environment. Sandiego, CA: Academic Press;
 2006.

437 2. Valentini R, Matteucci G, Dolman AJ, Schulze ED, Rebmann C, Moors EJ, et al.
438 Respiration as the main determinant of carbon balance in European forests. Nature. 2000; 404:
439 861-865. DOI: 10.1038/3500908.

3. Scharlemann JP, Tanner EV, Hiederer R, Kapos V. Global soil carbon: understanding
and managing the largest terrestrial carbon pool. Carbon Manag. 2014; 5: 81–91.

Makita N, Kosugi Y, Sakabe A, Kanazawa A, Ohkubo S, Tani M. Seasonal and diurnal 442 4. patterns of soil respiration in an evergreen coniferous forest: Evidence from six years of 443 observation with automatic chambers. **PLoS** ONE, 2018; 13(2): e0192622. 444 https://doi.org/10.1371/journal.pone.0192622 445

5. Cong WF, Van Ruijven J, Mommer L, De Deyn GB, Berendse F, Hoffland E. Plant
species richness promotes soil carbon and nitrogen stocks in grasslands without legumes. Journal
of Ecology. 2014; 102(5): 1163–1170. DOI 10.1111/1365-2745.12280.

Luo W, Li MH, Sardans J, Lü XT, Wang C, Peñuelas J, Wang Z, Han XG, Jiang Y.
Carbon and nitrogen allocation shifts in plants and soils along aridity and fertility gradients in
grasslands of China. Ecol Evol. 2017; 7(17): 6927–6934, DOI 10.1002/ece3.3245.

452 7. Devi SB, Sherpa SSSS. Soil carbon and nitrogen stocks along the altitudinal gradient of 453 the Darjeeling Himalayas, India. Environ Monit Assess. 2019: 191 (361): https://doi.org/10.1007/s10661-019-7470-8 454

8. Montagnani L, Badraghi A, Speak AF, Wellstein C, Borruso L, Zerbe S, Zanotelli D.
Evidence for a non-linear carbon accumulation pattern along an Alpine glacier retreat
chronosequence in Northern Italy. PeerJ. 2019; 7:e7703, DOI 10.7717/peerj.7703.

Janssens I.A, Lankreijer H, Matteucci G, Kowalski AS, Buchmann N, Epron D.
Productivity overshadows temperature in determining soil and ecosystem respiration across
European forests. Glob Chang Biol. 2001; 7: 269–278.

Hopkins F, Gonzalez-Meler MA, Flower CE, Lynch DJ, Czimczik C, Tang J, Subke JA.
Ecosystem-level controls on root-rhizosphere respiration, New Phytol. 2013; 199: 339e351,
http://dx.doi.org/10.1111/nph.12271.

Scandellari F, Zanotelli D, Ceccon C, Bolognesi M, Montagnani L, Cassol P, Melo GW,
Tagliavini M. Enhancing prediction accuracy of soil respiration in an apple orchard by
integrating photosynthetic activity into a temperature-related model, Eur J Soil Biol. 2015; 70:
77-87. Doi: 10.1016/j.ejsobi.2015.07.006.

12. Zhang ZS, Dong XJ, Xu BX, Chen YL, Zhao Y, Gao YH, Hu YG, Huang L. Soil
respiration sensitivities to water and temperature in a revegetated desert. J. Geophys. Res.
Biogeosci. 2015; 120: 773–787. doi:10.1002/2014JG002805

471 13. Chen D, Yu M, González G, Zou X, Gao Q. Climate Impacts on Soil Carbon Processes
472 along an Elevation Gradient in the Tropical Luquillo Experimental Forest. Forests. 2017. 8; 90:
473 doi:10.3390/f8030090

474 14. Ma M, Zang Z, Xie Z, Chen Q, Xu W, Zhao C, Shen G. Soil respiration of four forests
475 along elevation gradient in northern subtropical China. Ecol Evol. 2019; 9: 12846–12857.

476 15. Reichstein M, Beer C. Soil respiration across scales: the importance of a model-data
477 integration framework for data interpretation. J Plant Nutr Soil Sci. 2008; 171: 344e354.

478 16. Subke JA, Bahn M. On the 'temperature sensitivity' of soil respiration: Can we use the
479 immeasurable to predict the unknown? Soil Biol Biochem. 2008; 42 (9): 1653-1656.

Lloyd J, Taylor JA. On the temperature dependence of soil respiration, Funct Ecol. 1994;
8(3): 315–323. doi:10.2307/2389824.

18. Xu M, Qi Y. Soil surface CO₂ efflux and its spatial and temporal variation in a young
ponderosa pine plantation in California. Global Change Biol. 2001; 7: 667 – 677.
https://doi.org/10.1046/j.1354-1013.2001.00435.x

Janssens IA, Dore S, Epron D, Lankreijer H, Buchmann N, Longdoz B, Brossaud J,
Montagnani L. Climatic influences on seasonal and spatial differences in soil CO₂ efflux. In:
Valentini R (ed) Fluxes of carbon, water and energy of European forests. Berlin: Springer; 2003.

20. Reichstein M, Rey A, Freibauer A, Tenhunen J, Valentini R, Banza R. Modeling
temporal and large-scale spatial variability of soil respiration from soil water availability,
temperature and vegetation productivity indices. Global Biogeochem Cycles. 2003; 17: doi:
10.1029/2003GB002035.

Van't Hoff JH. Lectures on theoretical and physical chemistry. In Chemical Dynamics
Part I (pp. 224–229). London: Edward Arnold. Vereecken, H., Pachepsky, Y., Simmer, C.,
Rihani, J., Kunoth, A., Korres, W., et al., 2016. On the role of patterns in understanding the
functioning of soil-vegetation atmosphere systems. J Hydrol. 1898; 542: 63–86.
<u>https://doi.org/10.1016/j.jhydrol.2016.08.053</u>

497 22. Lomolino MV. Elevation gradients of species-density: historical and prospective views.
498 Glob Ecol Biogeogr. 2001; 10: 3–13.

Prietzel J, Zimmermann L, Schubert A, Christophel D. Organic matter losses in German
Alps forest soils since the 1970s most likely caused by warming. Nat Geosci. 2016; 1-8, <u>DOI:</u>
10.1038/NGEO2732

502 24. Shedayi AA, Xu M, Naseer L, Khan B. Altitudinal gradients of soil and vegetation
503 carbon and nitrogen in a high altitude nature reserve of Karakoram ranges. SpringerPlus. 2016;
504 5: 320. DOI 10.1186/s40064-016-1935-9

Jiang L, He Z, Liu J, Xing C, Gu X, Wei C, Zhu J, Wang X. Elevation Gradient Altered
Soil C, N, and P Stoichiometry of Pinus taiwanensis Forest on Daiyun Mountain. Forests. 2019;
10: 1089. doi:10.3390/f10121089

26. Rodeghiero M, Cescatti A. Main determinants of forest soil respiration along an
elevation/temperature gradient in the Italian Alps. Glob Chang Biol. 2005; 11: 1024–1041. doi:
10.1111/j.1365-2486.2005.00963.x

511 27. Shi Z, Wang JS, He R, Fang YH, Xu ZK, Quan W, Zhang ZX, Ruan HH. Soil

respiration and its regulating factor along an elevation gradient in Wuyi Mountain of Southeast
China. Chinese J Ecol. 2008; 27 (4): 563-568.

Luo S, Liu G, Li Z, Hu C, Gong L, Wang M, Hu H. Soil respiration along an altitudinal
gradient in a subalpine secondary forest in China. iForest. 2014; 8: 526-532.

516 29. Kane ES, Valentine DW, Schuur EAG, Dutta K. Soil carbon stabilization along climate
517 and stand productivity gradients in black spruce forests of interior Alaska. Can. J. For. Res.
518 2005; 35: 2118–2129.

519 30. Djukic L, Zehetner F, Tatzber M, Gerzabek MH. Soil organic-matter stocks and
520 characteristics along an Alpine elevation gradient. J Plant Nutr Soil Sci. 2010; 173: 30–38. DOI:
521 10.1002/jpln.200900027

522 31. Kunkel ML, Flores AN, Smith TJ, McNamara JP, Benner SG. A simplified approach for
523 estimating soil carbon and nitrogen stocks in semi-arid complex terrain. Geoderma. 2011; 165
524 (1): 1–11.

525 32. Tian Q, He H, Cheng W, Bai Z, Wang Y, Zhan X. Factors controlling soil organic
526 carbon stability along a temperate forest altitudinal gradient. Sci Rep. 2016; 6: 18783. DOI:
527 10.1038/srep18783

528 33. Garcia-Pausas J, Casals P, Camarero L, Huguet C, Sebastia MT, Thompson R, Romanya
529 J. Soil organic carbon storage in mountain grasslands of the Pyre-nees, effects of climate and

34. Zimmermann M, Meir P, Bird MI, Malhi Y, Ccahuana AJQ. Temporal variation and
climate dependence of soil respiration and its components along a 3000 m altitudinal tropical
forest gradient. Global Biogeochem Cy. 2010; 24: GB4012. doi: 10.1029/2010GB003787

35. Kobler J, Zehetgruber B, Jandl R, Dirnböck T, Schindlbacher A. Effects of slope aspect
and site elevation on seasonal soil carbon dynamics in a forest catchment in the Austrian
Limestone Alps. 19th EGU General Assembly, EGU2017, proceedings from the conference
held. 2017; 23-28 April, in Vienna, Austria, p.16691

36. Grand S, Rubin A, Verrecchia EP, Vittoz P. Variation in Soil Respiration across Soil and
Vegetation Types in an Alpine Valley. PLoS ONE. 2016; 11 (9): e0163968.
<u>https://doi.org/10.1371/journal.pone.0163968</u>

Migliavacca M, Reichstein M, Richardson AD, Colombo R, Sutton MA, Lasslop G.
Semiempirical modeling of abiotic and biotic factors controlling ecosystem respiration across
eddy covariance sites. Global Change Biology. 2011; 17(1): 390-409. DOI: 10.1111/j.13652486.2010.02243.x.

38. Bréchet L, Ponton S, Alméras T, Bonal D, Epron D. Does spatial distribution of tree size
account for spatial variation in soil respiration in a tropical forest?. Plant Soil. 2011; 347 (293):
https://doi.org/10.1007/s11104-011-0848-1

39. Xu X, Shi Z, Li D, Zhou X, Sherry RA, Luo Y. Plant community structure regulates
responses of prairie soil respiration to decadal experimental warming. Glob Change Biol. 2015;
21: 3846-3853. https://doi.org/10.1111/gcb.12940

Tian Q, Wang D, Tang Y, Li Y, Wang M, Liao C, Liu F. Topographic controls on the
variability of soil respiration in a humid subtropical forest. Biogeochemistry. 2019; 145: 177–
<u>https://doi.org/10.1007/s10533-019-00598-x</u>

Tonon G, Dezi S, Ventura M, Scandellari F. The Effect of Forest Management on Soil
Organic Carbon. In: Sauer TJ, Eiler JM, Sivakumar MVK (eds) Sustaining Soil Productivity in

Response to Global Climate Change: Science, Policy, and Ethics. John Wiley & Sons, Inc. 2011;
pp 225–238

Greiser C, Meineri E, Luoto M, Ehrlén J, Hylander K. Monthly microclimate models in a
managed boreal forest landscape. Agri For Meteorol. 2018; 250–251: 147–158.
https://doi.org/10.1016/j.agrformet.2017.12.252.

43. Carey JC, Tang J, Templer PH, Kroeger KD, Crowther T, et al. Temperature response of
soil respiration largely unaltered with experimental warming. PNAS. 2016; 113 (48): 13797–
13802.

565 44. Tang J, Cheng H, Fang C. The temperature sensitivity of soil organic carbon
566 decomposition is not related to labile and recalcitrant carbon. PLoS ONE. 2017; 12 (11):
567 e0186675. https://doi.org/10.1371/journal.pone.0186675

Montagnani L, Manca G, Canepa E, Georgieva E, Acosta M, Feigenwinter C, Janous D.
A new mass conservation approach to the study of CO₂ advection in an alpine forest. J. Geophys.
Res. Atmos. 2009; 114: D07306, DOI:10.1029/2008JD010650.

46. Xu X, Yi C, Montagnani L, Kutter E. Numerical Study of the Interplay between
Thermo-topographic Slope Flow and Synoptic Flow on Canopy Transport Processes, Agric For
Meteorol. 2018; 255: 3-16. <u>https://doi.org/10.1016/j.agrformet.2017.03.004</u>.

47. Gielen B, Acosta M, Altimir N, Buchmann N, Cescatti A, Ceschia E, Fleck S. Soilmeteorological measurements at ICOS monitoring stations in terrestrial ecosystems. Int
Agrophys. 2019; 32: 645-664. DOI: 10.1515/intag-2017-0048

Ventura M, Panzacchi P, Muzzi E, Magnani F, Tonon G. Carbon balance and soil carbon
input in a poplar short rotation coppice plantation as affected by nitrogen and wood ash
application. New Forests. 2019. <u>https://doi.org/10.1007/s11056-019-09709-w</u>

49. Richards FJ. A flexible growth function for empirical use. J Exp Bot. 1959; 10: 290 –
300.

582 50. Janssens IA, Pilegaard K. Large seasonal changes in Q10 of soil respiration in a beech
583 forest. Glob Chang Biol. 2003; 9: 911- 918.

584 50. Crawley M.J. The R Book. Chichester: John Wiley & Sons, Ltd.; 2007.

585 51. Zuur A, Ieno EN, Walker N, Saveliev AA, Smith GM. Mixed Effects Models and 586 Extensions in Ecology with R. New York: Springer; 2009.

587 52. Bates D, Maechler M, Bolker B, Walker S. Ime4: linear mixed-effects models using
588 Eigen and S4. R package version 1.1-7; 2014.

53. Nakagawa S, Holger S. A general and simple method for obtaining R² from generalized
linear mixed effects models. Methods Ecol Evol. 2013; 4: 133-142.

54. Jaeger BC, Edwards LJ, Das K, Sen PK. An R² Squared Statistic for Fixed Effects in the
Generalized Linear Mixed Model. J Appl Stat. 2016; 44 (6): 1086–1105

593 55. Meyer N, Welp G, Amelung W. The temperature sensitivity (Q10) of soil respiration:
594 Controlling factors and spatial prediction at regional scale based on environmental soil classes.
595 Global Biogeochem Cy. 2018; 32: 306–323. <u>https://doi.org/10.1002/2017GB005644</u>

596 56. Shapiro SS, Wilk MB. Analysis of variance test for normality. Biometrika. 1965; 52:
597 591–611.

598 57. Levene H. In Contributions to Probability and Statistics: Essays in Honor of Harold
599 Hotelling, I. Olkin et al. eds., Stanford University Press, 1960; 278-292.

600 58. R Core Team. R: A language and environment for statistical computing. R Foundation
601 for Statistical Computing, Vienna, Austria. 2019. Available online at https://www.R-project.org/.

59. Wang G, Zhou Y, Xu X, Ruan H, Wang J. Temperature Sensitivity of Soil Organic
Carbon Mineralization along an Elevation Gradient in the Wuyi Mountains, China. PLoS ONE.
2013; 8(1): e53914. doi:10.1371/journal.pone.0053914

60. Chatterjee A, Jenerette GD. Variation in soil organic matter accumulation and metabolic
activity along an elevation gradient in the Santa Rosa Mountains of Southern California, USA. J
Arid Land. 2015; 7(6): 814–819. doi: 10.1007/s40333-015-0085-1

608 61. Tsozuéa D, Nghonda JP, Tematio P, Basga SD. Changes in soil properties and soil
609 organic carbon stocks along an elevation gradient at Mount Bambouto, Central Africa. Catena.
610 2019; 175: 251–262. https://doi.org/10.1016/j.catena.2018.12.028

611 62. de la Cruz-Amo L, Bañares-de-Dios G, Cala V, Granzow-de la Cerda I, Espinosa CI,

Ledo A, Salinas N, Macía M.J, Cayuela L. Trade-offs among aboveground, belowground, and

soil organic carbon stocks along altitudinal gradients in Andean Tropical Montane Forests. Front.

614 Plant Sci. 2020; 11:106. doi: 10.3389/fpls.2020.00106

615 63. Kirschbaum MUF. The temperature dependence of organic-matter decomposition-still a
616 topic of debate. Soil Biol Biochem. 2006; 38: 2510-2518

617 64. He X, Hou E, Liu Y, Wen D. Altitudinal patterns and controls of plant and soil nutrient
618 concentrations and stoichiometryin subtropical China. Sci Rep. 2016; 6: 24261. DOI:
619 10.1038/srep24261

620 65. Griffiths PR, Madritch MD, Swanson AK. The effects of topography on forest soil 621 characteristics in the Oregon Cascade Mountains (USA): Implications for the effects of climate 622 change on soil properties. For Ecol Manag. 2009; 257(1): 1-7.

623 66. Bahn M, Rodeghiero M, Anderson-Dunn M, Dore S, Gimeno C, et al. Soil respiration in
624 European grasslands in relation to climate and assimilate supply. Ecosystems. 2008; 11: 1352–
625 1367. DOI: 10.1007/s10021-008-9198-0

626 67. Shi Y, Baumann F, Ma Y, Song C, Uhn PK, Scholten T, He JS. Organic and inorganic
627 carbon in the topsoil of the Mongolian and Tibetan grasslands: pattern, control and implications.
628 Biogeosciences. 2012; 9: 2287–2299. doi:10.5194/bg-9-2287-2012

- 629 68. Vanhala P. Seasonal variation in the soil respiration rate in coniferous forest soils. Soil
 630 Biol Biochem. 2002; 34: 1375–1379.
- 69. Zhang YY, Wu W, Liu H. Factors affecting variations of soil pH in different horizons in
 hilly regions. PLoS One. 2019; 14(6): e0218563. doi: 10.1371/journal.pone.0218563

633 70. Martí C, Badía D. Characterization and classification of soils along two altitudinal
634 transects in the Eastern Pyrenees, Spain. Arid Soil Res Rehabil. 1995; 9: 367-383. DOI:
635 10.1080/15324989509385905

636 71. Smith JL, Halvorson JJ, Jr HB. Soil properties and microbial activity across a 500 m
637 elevation gradient in a semi-arid environment. Soil Biol Biochem. 2002; 34: 1749–1757.

638 72. Seibert J, Stendahl J, Sørensen R. Topographical influences on soil properties in boreal
639 forests. Geoderma. 2007; 141(1–2): 139–148.

Badía D, Ruiz A, Girona A, et al. The influence of elevation on soil properties and forest
litter in the Siliceous Moncayo Massif, SW Europe. J Mt Sci. 2016; 13: 2155–2169.
https://doi.org/10.1007/s11629-015-3773-6

Klopfenstein ST, Hirmas DR, Johnson W. Relationships between soil organic carbon
and precipitation along a climosequence in loess-derived soils of the Central Great Plains, USA.
Catena. 2015; 133: 25-34. <u>http://dx.doi.org/10.1016/j.catena.2015.04.015</u>

5. Schrumpf M, Schulze ED, Kaiser K, Schumacher, J. How accurately can soil organic
carbon stocks and stock changes be quantified by soil inventories?. Biogeosciences. 2011; 8:
1193–1212. <u>https://doi.org/10.5194/bg-8-1193-2011</u>.

Keenan TF, Migliavacca M, Papale D, Baldocchi D, Reichstein M, Torn M, Wutzler T.
Widespread inhibition of daytime ecosystem respiration. Nat Ecol Evol. 2019; 3(3): 407–415.
doi: 10.1038/s41559-019-0809-2

77. Reichstein M, Ciais P, Papale D, Valentini R, Running S, Viovy N. Reduction of
ecosystem productivity and respiration during the European summer 2003 climate anomaly: a
joint flux tower, remote sensing and modelling analysis. Glob Change Biol. 2007; 13(3): 634–
651. DOI 10.1111/j.1365-2486.2006.01224.x

78. Yu GR, Zhu XJ, Fu YL, He HL, Wang QF, Wen XF. Spatial patterns and climate drivers
of carbon fluxes in terrestrial ecosystems of China Global Change Bio. 2013; 119: 798-810.
<u>https://doi.org/10.1111/gcb.12079</u>

659 79. Chen S, Zou J, Hu Z, Lu Y. Climate and vegetation drivers of terrestrial carbon fluxes: a
660 global data synthesis. Adv Atmos Sci. 2019; 36: 679–696.

80. Ma J, Liu R, Li C, Fan L, Xu G, Li Y. Herbaceous layer determines the relationship
between soil respiration and photosynthesis in a shrub-dominated desert plant community. Plant
Soil. 2020, 449: 193–207. https://doi.org/10.1007/s11104-020-04484-6

81. Raich JW, Tufekciogul A. Vegetation and soil respiration: Correlations and controls.
Biogeochemistry. 2000; 48: 71–90. https://doi.org/10.1023/A:1006112000616

666 82. Chen Q, Wang Q, Han X, Wan S, Li L. Temporal and spatial variability and controls of
667 soil respiration in a temperate steppe in northern China. Global Biogeochem Cycles. 2010; 24:
668 GB2010. doi:10.1029/2009GB003538.

669 83. Oertel C, Matschullat J, Zurba K, Zimmermann F, Erasmi S. Greenhouse Gas Emissions
670 From Soil - A review, Chemie der Erde. 2016; 76: 327–352, doi:10.1016/j.chemer.2016.04.002.

Andersson S, Nilsson SI. Influence of pH and temperature on microbial activity,
substrate availability of soil-solution bacteria and leaching of dissolved organic carbon in a mor
humus. Soil Biol Biochem. 2001; 33 (9): 1181–1191. <u>https://doi.org/10.1016/S0038-</u>
0717(01)00022-0

85. Reth S, Reichstein M, Falge E. The effect of soil water content, soil Temperature, soil
pH value and the root mass on soil CO₂ efflux- a modified model. Plant Soil, 2005; 268: 21-33.

677 86. Chappell C, Johnson A. Influence of pH and bulk density on carbon dioxide efflux in
678 tree urban wetland types. Professional Agricultural Workers Journal. 2015; 3 (1, 5).
679 <u>htp://tuspubs.tuskegee.edu/pawj/vol3/iss1/5</u>

680 87. Mordhorst A, Peth S, Horn R. Influence of mechanical loading on static and dynamic
681 CO₂ efflux on differently textured and managed Luvisols. Geoderma. 2014; 219–220: 1–13.

682 88. Chen B, Liu S, Ge J, Chu J. Annual and seasonal variations of Q10 soil respiration in the
683 sub-alpine forests of the Eastern Qinghai-Tibet Plateau, China. Soil Biol Biochem. 2010; 42:
684 1735-1742

Feng J, Wang J, Song Y, Zhu B. Patterns of soil respiration and its temperature
sensitivity in grassland ecosystems across China, Biogeosciences. 2018; 15: 5329–5341,
https://doi.org/10.5194/bg-15-5329-2018

Acosta M, Pavelka M, Montagnani L, Kutsch W, Lindroth A, Juszczak R, Janouš D. Soil
surface CO₂ efflux measurements in Norway spruce forests. Comparison between four different
sites acrossEurope — from boreal to alpine forest. Geoderma. 2013; 192: 295-303. DOI:
10.1016/j.geoderma.2012.08.027.

692 91. Zhou T, Shi P, Hui D, Luo Y. Global pattern of temperature sensitivity of soil
693 heterotrophic respiration (Q10) and its implications for carbon-climate feedback, J Geophys Res694 Biogeo. 2009; 114: 271–274, https://doi.org/10.1029/2008JG000850, 2009.

Song X, Peng C, Zhao Z, Zhang Z, Guo B, Wang W, Jiang H, Zhu Q. Quantification of
soil respiration in forest ecosystems across China. Atmos Environ. 2014; 94: 546–551,
https://doi.org/10.1016/j.atmosenv.2014.05.071

698 93. Zhou W, Hui D, Shen W. Effects of Soil Moisture on the Temperature Sensitivity of Soil
699 Heterotrophic Respiration: A Laboratory Incubation Study. PLoS ONE. 2014; 9(3): e92531.
700 doi:10.1371/journal.pone.0092531

701 94. Zhao J, Li R, Li X, Tian L. Environmental controls on soil respiration in alpine meadow
702 along a large altitudinal gradient on the central Tibetan Plateau. Catena. 2017; 159: 84–92.
703 https://doi.org/10.1016/j.catena.2017.08.007

704 95. Zhou Z, Guo C, Meng H. Temperature sensitivity and basal rate of soil respiration and
705 their determinants in temperate forests of North China. PLoS One. 2013; 8(12): e81793.
706 https://doi.org/10.1371/journal.pone.0081793

96. Quan Q, Wang C, He N, Zhang Z, Wen X, Su H, Wang Q, Xue J. Forest type affects the
coupled relationships of soil C and N mineralization in the temperate forests of northern China.
Sci Rep. 2014; 4: 6584, DOI: 10.1038/srep06584,

710 97. Rambo TR, North MP. Canopy microclimate response to pattern and density of thinning Nevada forest. For Ecol 2009; 257 (2): 435-442, 711 in а Sierra Manag. 712 https://doi.org/10.1016/j.foreco.2008.09.029

98. Arx GV, Dobbertin M, Rebetez M. Spatio-temporal effects of forest canopy on
understory microclimate in along-term experiment in Switzerland. Agr Forest Meteorol. 2012;
166–167: 144–155, https://doi.org/10.1016/j.agrformet.2012.07.018

99. Niu S, Luo Y, Fei S, Yuan W, Schimel D, Law BE, et al. Thermal Optimality of Net
Ecosystem Exchange of Carbon Dioxide and Underlying Mechanisms. New Phytol. 2012; 194
(3): 775-783, DOI: 10.1111/j.1469-8137.2012.04095.x.

719 Supporting Information

- 720 S1. Collected soil respiraton data by soil chambers. Used for soil respiration modeling
- 721 **S2. Measured soil physicochemical properties.** Used for the analysis.
- 722 S1 Fig. Predicted daily mean air temperature (C, closed circles) and soil CO₂ fluxes (µmol

723 $CO_2 C m^{-2} s^{-1}$, open circles) during the experimental period (July 2017- July 2018). Mean 724 annual air temperature (Tair, °C) and total cumulative soil respiration in the entire experimental

- period (SR, kg C m^{-2}) are indicated for each site on the top of the relative plots.
- 726 S2 Fig. Observed soil respiration vs. predicted soil respiration by model.

727 S1 Table. The developed equations by linear and polynomial regressions between the

728 different parameters tested and elevation. BD – soil bulk density; Q10 – temperature

sensitivity of soil respiration; SR- cumulated soil respiration; SRref – respiration at the 10°C

reference temperature; Elev – elevation.





Elevation (m a.s.l)



É

а

В

C Site

D

0.0

A

É

D

C Site

a T

Ε

ат

Ė

а

D

ab

D

Figure

A

В

2





