A Multiple Regression Assessment of the Biomineral Urease Activity from Urine

Drainpipes of California Rest Areas

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Abstract

Clogging and odor is strongly associated with ureolytic biomineralization in

waterless and low-flow urinal drainage systems in high usage settings. These blockages

continue to hinder widespread waterless and low-flow urinal adoption due to

subsequent high maintenance requirements and hygiene concerns. Through field

observations, hypothesis testing, and multiple regression analysis, this study attempts

to characterize, for the first time, the ureolytic activity of the biomineralization found in

alternative technologies located at 9 State-owned restrooms. Multiple regression

analysis ($n = 55$, df = 4, $R^2 = 0.665$) suggests that intrasystem sampling location ($\hat{\beta} =$

1.23, $p < 0.001$), annual users per rest area ($\hat{\beta} = 0.5$, $p = 0.004$), and the

organic/inorganic mass fraction ($\hat{\beta} = 0.59$, $p = 0.003$), are statistically significant

influencers of the ureolytic activity of biomineral samples ($p < 0.05$). Conversely, *ureC*

gene abundance ($p = 0.551$), urinal type ($p = 0.521$) and sampling season ($p = 0.956$)

are not significant predictors of biomineral ureolytic activity. We conclude that high

concentrations of the urease alpha subunit, *ureC*, which can be interpreted as proxy

measure of a strong, potentially ureolytic community, does not necessarily mean that

the gene is being expressed. Future studies should assess *ureC* transcriptional activity

to measure gene expression rather than gene abundance to assess the relationship

between environmental conditions, their role in transcription, and urease activities. In

sum, this study presents a method to characterize biomineral ureolysis and establishes

baseline values for future ureolytic inhibition treatment studies that seek to improve the

usability of urine collection and related source separation technologies.

1. Introduction

Waterless and low-flow urinals reduce water consumption, improve hygiene with

touchless operation, and can be used for source separation of urine; additionally,

waterless systems require less plumbing than conventional systems. However, these

source-separation technologies are susceptible to biomineralization [1,2].

Biomineralization, usually of a mixed composition of struvite, calcium phosphate,

calcium oxalate, and calcium carbonate, has plagued urine diversion projects since the

earliest projects were studied, leading to clogging, odor, and overall user dissatisfaction

[1–4].

Researchers have described the formation of biomineralization in terms of (a)

cellular activities, (b) passive formation of crystals caused by biofilms, and (c) biological

and chemical facilitation of crystal supersaturation conditions [5–7]. Biomineralization in

urine source-separation contexts is likely governed by a combination of mechanisms.

Urease and its ureolytic activity are measures of biomineralization potential

because the rate of precipitation is dependent, in part, on the rate of increase of media

pH, which depends on the rate of ureolysis. The elevated pH resulting from ureolysis

plays a critical role in the supersaturation crystal formation process. Because urinals are

subject to intermittent supplements of a urea and an ion source, urinals and urine

drainage traps become a selective breeding ground for ureolytic organisms that cause

an increase in the pH of collected urine and facilitate mineral precipitation as has been

observed in urological devices[8]. Ureolytic bacteria responsible for the

biomineralization use the nickel-dependent metalloenzyme, urease, to catalyze the

hydrolysis of urea into ammonia and bicarbonate which in turn raises the pH and

creates conditions favorable of precipitation [4]. Broomfield et al. (2009), in their

catheter study, demonstrated that rates of calcium and magnesium encrustation caused

by various ureolytic bacteria isolates is correlated with an increase in ureolytic activity

[9] An elevated pH promotes calcium phosphate and oxalate stone formation due to a

shift in phosphate speciation from HPO_4^{2-} to PO_4^{3-} and the decomposition of ascorbic

acid into oxalate—both cases represent an increase in ion concentrations that lead to

elevated encrustation rates found in catheters [10]. Ureolysis also results in carbonate

and bicarbonate ion formation which can further contribute to biomineralization as the

urine becomes supersaturated [11]. Researchers similarly showed that greater ureolytic

rates from bacterial urease are correlated with greater rates of calcium carbonate

precipitation [12–14]. Studies using *Proteus mirabilis* have shown that urease defective

mutants fail to form crystalline biofilms in laboratory models, demonstrating the key role

of pH and urease activity in crystal formation [15]. In dental plaque studies, researchers

suggest that ammonia generating capacity in a mixed-species model of ureolytic oral

biofilms is essential for the stabilization of microbial communities in ureolytic

environments [16]. Losses of sufficient quantities of urease resulted in the acidification

of biofilms and a decrease in community diversity [16].

Through multiple linear regression modelling, this study will be the first of its kind

to: (a) model biomineral enzyme activity in terms of both categorical and quantitative

predictors, (b) examine biomineral enzyme activity from urine source-separation

technology, and (c) do so on a geographic scale with a sufficiently large sample size.

This study also builds upon previous works describing soil or biofilm ureolytic activity

that (a) use small sample sizes in parametric hypothesis tests (n=6) or multiple

regression (n=4), (b) neglect discussion of model validation beyond the coefficient of

determination (R^2), (c) do not discuss whether their data fits assumptions required for

application of a statistical test, and (d) mention statistical significance, but not practical

significance, i.e. the magnitude of effect [17–20].

Finding a link between environmental parameters such as intrasystem sampling

location, usage frequency, seasonality, gene abundance found through qPCR, and

urinal types with the enzymatic activity of the biomineral samples will be useful in

understanding the effects of restroom configuration on ureolytic activity. Understanding

the effect of seasonality and sampling locations within a urine drainage system where

ureolytic activity is highest may be insightful when predicting locations and times of year

where the components of the urine collection system are most susceptible to biomineral

fouling.

2. Materials and Methods

The coming subsections will describe the sampling procedures and locations

followed by methods used in downstream analyses to quantify the environmental

variables used in the statistical analyses. The R Markdown HTML output containing the

script can be found in the Online Resources section. The raw environmental data can

be found in the Dryad repository (DOI:10.25338/B82906) as an .RDS file.

2.1 Sample Collection

Rest areas were categorized by the types of urinals installed: conventional ~ 1

gal/flush, low-flow ~0.125 gal/flush, and waterless no flush. Biomineralization deposits

were scraped into sterile 50 mL conical tubes from fouled fixtures and drainage systems

when available. A total of 2 conventional, 2 low-flow, and 5 waterless public restrooms

along California highways, also known as rest areas, were observed in this study. A

summary of sites, drainpipe configurations, and characteristic samples are shown in

Figure 1 as: (a) location of sampling sites with respect to urinal type used in this study,

(b) biomineralization formation on a waterless urinal cartridge at Erreca on 16 Sep

2019, (c) a view of reduced internal pipe diameter by biomineralization in a urine

drainage pipe at the Dunnigan northbound oriented public rest area on 12 Dec 2019,

and (d) general drainage system layout consisting of the drains directly connected to the

urinals, which collectively flows into a main drain also connected to the sink drains. The

men's restrooms were typically fitted with two urinals at two different heights to conform

to the American Disability Act (ADA).

All samples were stored in an ice chest after collection and processed within

three days of sample collection. Previous work monitoring the ureolysis rate in soils

have found that a distinct slowdown in ureolytic rate was not detected until 8 months of

cooled storage [21]. As such, the sampling preservation measures were deemed

adequate.

Fig. 1 Sampling sites, characteristic samples, and typical drainage system

119   configurations

2.2 Biomineral Ureolytic Enzyme Activity Characterization

To compare enzymatic activities of biomineral samples between various sites *in*

vitro, a known wet mass of the biomineral samples was suspended and mixed in a 100

mL volume of 7.3 pH 200 mM (4-(2-hydroxyethyl)-1-piperazineethanesulfonic acid)

(HEPES) buffer containing 2.5% urea m/m. The rate of increase in conductivity is

proportional to that of urea hydrolysis and can be used as a surrogate measure for

enzymatic activity [22]. As a comparative basis between samples, one unit of specific

activity is defined as uS cm^{-1}min^{-1} per gram of volatile solids (VS).

Gravimetric analyses followed standard methods for the examination of water

and wastewater [23]. A mass balance was performed by comparing the wet solid mass

with the dry mass following 105°C desiccation and fixed mass after 550°C ashing. Total

solids (TS) represent the inorganic matter in the sample while VS represents organic

matter. Each biomineral sample was analyzed in triplicate and then averaged.

2.3 Quantifying Gene Abundance using Real-time Polymerase Chain Reaction

(qPCR)

To examine the relationship between *in vitro* ureolytic activity and the genetic

predispositions for ureolysis, the genomes of phylotype representatives for the presence

of urease genes were examined by qPCR. A similar protocol was described previously

[24]. The urease associated gene were designed on the urease alpha subunit encoding

gene (*ureC*). Primer sequences were obtained from the literature [25]. Sensitivity and

efficiency were established from the y-intercept and slope of the standard curve, which

was created by running triplicate, 10-fold serial dilutions of plasmid DNA containing the

ligated amplicon of each gene (Eurofins Genomics LLC, Louisville, KY). The sensitivity

of ureC-F (TGGGCCTTAAAATHCAYGARGAYTGGG) and ureC-R

(SGGTGGTGGCACACCATNANCATRTC) was <4,000 copies/qPCR reaction and the

efficiency was 80.6% (R^2 = 0.9974). Poor sensitivity and low efficiency for *ureC* is

expected due to the nature of SYBR degenerative primers. Biomineral samples were

kept frozen at -20°C prior to DNA extraction. DNA was manually extracted from 0.25 g

of sample using a commercially available kit following manufacturer recommendations

and eluted in 100 µL of diethylpyrocarbonate (DEPC) treated water (Qiagen DNeasy

Power Soil Kit, cat # 12888-50). Each 12 µL reaction contained 6 µL SYBR master mix

(Applied Biosystems SYBR Green PCR Master Mix, cat # 4309155), 0.48 µL of a

primer-water mixture (primers at final concentration of 400 nM), 4.52 µL of DEPC-

treated water, and 1 µL of extracted DNA. qPCR was performed using an automated

fluorometer (ABI PRISM 7900 HTA FAST, Thermo Fisher Scientific). Standard

amplification conditions were used: 95°C for 3 min, 40 cycles of 95°C for 15 s, 52°C for

30 s, and 72°C for 30 s, with a melting curve at 95°C for 15 s, 52°C for 15 s, and 95°C

for 15 s. Data was analyzed using Applied Biosystems SDS software, version 2.4.

Fluorescent signals were collected during the annealing phase and C_q values extracted

with a threshold of 0.2 and baseline values of 3–10 for the *ureC* assay . Amplification

specificity was verified using the dissociation temperature (T_m) of the qPCR amplicons

specific to each gene. Acceptable T_m ranges were determined to be +/- 2% of the

positive controls. For *ureC*, the acceptable T_m range was 80.8°C - 84.1°C. Samples with

detectable amplification but with T_m's outside of the acceptable ranges were considered

false positives and were deemed negative for the gene of interest. The absolute copy

numbers were also normalized in terms of volatile solids (VS) mass present in the

biomineral samples.

2.4 Statistical Analyses

All statistical work and data visualization was done using R version 4.0.2. An a-

priori power analysis was first used to inform the design of this study, suggesting that a

linear model can sufficiently capture a large effect size (f=0.35) at a level of significance

of 0.05 for a power of 0.8 using 1 tested dependent variable and 5 total predictors with a

minimum sample size of 25 [26]. After excluding sample rows missing data from low

quality qPCR reads and samples that did not have enough mass for gravimetric

analysis or biomineral enzyme activity, this randomly sampled, complete case analysis

included a sample size of 55 from 9 different facilities. In the regression analysis,

conventional urinals were aggregated with low-flow urinals because both urinal types

include flush water. A stepwise forward variable selection method was used. A

corrected Akaike information criterion (AICC) was also used to validate model selection

[27]. The ordinary least squares (OLS) multiple regression analysis was performed

assuming that a natural log-log transformed linear model is an adequate descriptor of

the system, whereby normality was verified in the Supplementary Information section. A

natural log-log transformed dataset enables for a practical interpretation of the effect

size as a percent change, or in this case, the elasticity between two biological variables

[28]. Regression coefficients were interpreted as natural log-level for categorical

variables. For the coming subsections, unless specified, variables will be discussed in

terms of natural logarithms.

2.5 Characterizing the Ureolytic Activity in Urinal Traps *in situ*

In situ urinal trap testing was conducted to characterize the ureolytic rate at the

time of sample collection within the urine drain trap. *In situ* biomineral ureolytic activity

was used to support the regression analysis derived from *in vitro* urease assays. The

project team developed a method using pH and conductivity meters to characterize the

baseline ureolytic rates. A description of the trap testing is shown graphically in Figure

2.

Fig. 2 Schematic of *in situ* trap activity test procedure: (a) using portable vacuum and

in-line liquid sample collector to remove trap contents, (b) application of standard urea

solution to empty trap, and (c) testing of urinal liquid trap to determine relative activity

The in-situ urinal trap procedure was conducted as follows:

First, the urine drain trap is vacuumed out as shown in Figure 2. Once emptied, a

200 mM 7.3 pH HEPES buffer containing 2.5% m/m urea is added until the drain trap is

full. Logging pH and EC meters were submerged in the trap opening and recorded for a

total of 10 minutes from which the ureolytic rate could be estimated using the rate of EC

formation.

3. Results and Discussion

After evaluating and selecting the most parsimonious multiple linear regression

model composed of categorical and quantitative environmental variables, the observed

influence, or lack thereof, of these variables will be discussed in the context of

biomineral ureolytic activity.

3.1 Multiple Linear Model and Validation

The multiple linear model composed of 55 observations is described in Tables 1 and 2. As shown in correlation heatmaps and residual analysis from Supplementary Figures 1 and 2, the linear model is in agreement with the Gauss-Markov OLS regression assumptions, which require that: a) the expected value of the regression residuals tends towards zero, b) the residuals are homoscedastic c) there is no autocorrelation between the regressors and the residuals such that exogeneity is upheld, d) the predictors are not multicollinear, and e) the residuals are also normal [28]. The residuals shown in Supplementary Figure 1 do not appear to have a trend based on the index plot, do not exhibit any correlation with each other from the autocorrelation plot, and appear homoscedastic from the fitted values vs. residuals plot. Finally, the residuals also appear normally distributed from the quantile-quantile plot in Supplementary Figure 1. It was concluded that the natural log-log linear model appropriately describes natural logarithmically transformed data and that the model fits well with the data. The AICC model selection results are shown in Supplementary Table 2, suggesting that the most parsimonious and probable model is Model 3 [27,29].

The regression results describing the most probable model (Model 3) is shown in Table 1 and 2, which also depicts the regression results from other tested models. The results presented in Tables 1 and 2 suggest that *ureC* gene concentrations (Model 4, p = 0.551), sampling season (Model 5, p = 0.956), and urinal types were statistically insignificant predictors of ureolytic activity ($p > 0.05$) and of low practical significance as indicated by the relatively small regression coefficients (see Table 6). From Table 1, the strongest predictor of biomineral ureolytic activity was the sampling location, namely,

those sampled from the main urinal drainage pipes exhibited the greatest enzymatic

activity. In Model 3, the second strongest predictor was the organic to inorganic fraction.

Annual number of users at a given rest area also positively influenced urease activity

likely due to the increased loading and usage frequency resulting in a semi-constant

stream of nutrients and salts necessary for a strong ureolytic community to develop and

thrive.

Table 1 Summary of effect sizes of significant predictors on biomineral ureolytic activity

Significant Predictor Variables	$\hat{\beta}$	CI (95%)	Effect on Biomineral Activity per g VS as Elasticity[a]
Annual Users per Rest Area	0.5	0.17, 0.82	A 25% increase in annual users per rest area corresponds to a 11.7 (3.9, 20.1) % increase in biomineral activity
VS/TS (g/g)	0.59	0.21, 0.97	A 25% increase in VS/TS (g/g) corresponds to a 14.1 (4.8, 24.2) % increase in biomineral activity
Intrasystem Location: Main Drain	1.24	0.83, 1.64	Compared to samples obtained from cartridges, those obtained from the gallery main drain had a 245 (129, 416) % larger geometric mean in biomineral activity

241   [a] Parenthetical contents represent effect sizes at limits of confidence intervals

Table 2 Multiple regression summary of model predicting biomineral ureolytic activity

Predictor Variables	Model 1 Estimates	Model 2 Estimates	Model 3 Estimates	Model 4 Estimates	Model 5 Estimates	Model 6 Estimates
Intercept	**-7.35 (-12.80 – -1.90)****	-2.18 (-6.92 – 2.57)	-0.55 (-5.05 – 3.94)	-0.46 (-5.00 – 4.08)	-0.56 (-5.12 – 3.99)	-0.36 (-4.92 – 4.20)
Annual Users per Rest Area	**0.96 (0.56 – 1.37)*****	**0.57 (0.22 – 0.92)****	**0.50 (0.17 – 0.82)****	**0.47 (0.14 – 0.81)****	**0.49 (0.16 – 0.83)****	**0.49 (0.16 – 0.82)****
Intrasystem Location: Gallery Drain		-0.28 (-0.62 – 0.07)	-0.19 (-0.52 – 0.13)	-0.15 (-0.51 – 0.21)	-0.19 (-0.53 – 0.15)	-0.14 (-0.50 – 0.22)
Intrasystem Location: Gallery Main Drain		**1.02 (0.61 – 1.43)*****	**1.24 (0.83 – 1.64)*****	**1.26 (0.85 – 1.68)*****	**1.24 (0.83 – 1.65)*****	**1.23 (0.82 – 1.64)*****
VS/TS (g/g)			**0.59 (0.21 – 0.97)****	**0.57 (0.19 – 0.96)****	**0.58 (0.17 – 0.99)****	**0.56 (0.18 – 0.95)****
ureC Concentration (copy #/g VS)				0.01 (-0.02 – 0.04)		
Sampling Season					0.01 (-0.31 – 0.33)	
Urinal Type						-0.12 (-0.49 – 0.25)
Observations	55	55	55	55	55	55
R^2 / R^2 adjusted	0.299 / 0.286	0.595 / 0.571	0.662 / 0.635	0.665 / 0.630	0.662 / 0.628	0.665 / 0.631

 [a] Significance codes: 0 ' *** ' 0.001 ' ** ' 0.05 * ' . ' 0.1 ' ' 1

3.2   The Influence of Organic Matter on Ureolytic Activity

That the organic content is significant ($p = 0.003$) and of sizeable effect ($\hat{\beta} = 0.59$) in predicting ureolytic activity, as shown in Table 1, is consistent with past findings from soil research that found correlations between organic matter concentrations and urease activity [13,14,30]. Others also observed that increased carbohydrate availability at neutral pH was correlated with increased *Actinomyces naeslundii* biofilm urease activity [14,17]. Liu et al. (2008), however, noticed that carbohydrate availability had no effect on *ureC* gene expression marked by through reverse-transcriptase quantitative real-time PCR (RT-qPCR) mRNA transcripts. Liu et al. (2008) hypothesizes that these observations were due to carbohydrate availability and pH modulation affecting the expression of genes other than *ureC* responsible for urease synthesis or apoenzyme activation [17].

Increasing the biomass of the inoculum by providing a carbon source in microbial induced calcite precipitation studies has been reported to promote the ureolytic activity [14]. Tobler et al. (2011) concluded that molasses supplementation selected for a larger microbial community that obtains their nitrogen from ureolysis, though there is no nitrogen limitation in urinals [14]. Others, who studied the environmental factors affecting microbially induced calcium precipitation concluded that increasing biomass may also increase ureolytic activity as there could be more active cells present [31]. Extracellular urease has also been suggested to be stabilized by adsorption to soil colloids, particularly organic matter, which may be similar to that observed in biomineral samples obtained from urine drain pipes [19].

One limitation of this study is that it is unclear what component of the organic

fraction is correlated with increased ureolytic activity as VS is a bulk measurement

encompassing any organic mass. Within the biomineral/stone matrix is also an organic

fraction composed of carbohydrates, proteins, lipids, and dead cell mass that binds the

mineral fraction of the precipitate [4]. Therefore, future research could evaluate different

organic components such as proteins and exopolysaccharide substances (EPS).

3.3   The Non-effect of Urinal Type and Seasonality on Ureolytic Activity

In addition to the linear regression results, Kruskal-Wallis testing for biomineral

ureolytic activity between waterless and low-flow urinals provides evidence that

waterless and low-flow are likely identical in population in terms of biomineral activity (p

= 0.47). While urinal type is not a statistically significant predictor of ureolytic activity,

biomineral samples from waterless urinals have exhibited a greater maximum ureolytic

activity than any biomineral sample obtained from low-flow urinals in this study, as

shown on Figure 3.

Finally, sampling season (as shown in Table 2) demonstrated no statistical ($p <$

0.001) or practical significance ($\hat{\beta} = 0.01$) in predicting biomineral activity. This may

explain why fouling is a year-round phenomenon, as the biomineral ureolytic activity

remains unaffected by seasonality, as the high urease activities year-round facilitate

conditions necessary for precipitation to occur. Because seasonality does not seem to

impact biomineral activity, future observations on the ureolytic activity of urine

drainpipes may be performed without temporal confounding effects. Though, future

microbial ecology studies are needed to understand the bacterial community structure

of the biomineral samples and should include sampling events from different seasons.

289
Fig. 3 Descriptive statistics on the effects of urinal type on natural log-transformed *ureC*

gene copies and biomineral activity

292
3.4   Effects of Intrasystem Sampling Location on Ureolytic Activity

While the ureolytic activity of biomineral samples obtained from the drainage

pipes immediately following the drain traps were not significantly different from those

corresponding to samples obtained from waterless urinal cartridges (Pairwise Wilcoxon

Rank Sum: $p = 0.053$), samples taken from the main drain lines which contacts

handwashing water were significantly non-identical in terms of ureolytic activity

(Kruskal-Wallis: $P < 0.001$; Pairwise Wilcoxon Rank Sum: $p < 0.001$). Within one

system, cartridges and gallery drain lines immediately succeeding the urinal are

exposed to the same urine feed without mixing with potable water and thus face similar

environmental conditions that influence ureolytic activity [13]. Because drain line

samples directly follow cartridge samples and are exposed to the same urine, the

relative similarity in environmental conditions between cartridge and drain line samples

may explain their different ureolytic rates compared to main drainpipe samples but not

with each other.

3.5   Biomineral Ureolytic Activity may be Predicted by Transcriptional Activity

more than by *ureC* gene abundance

Kruskal-Wallis testing results suggest that the *ureC* abundance between low-flow

and waterless urinals are significantly nonidentical ($p < 0.001$), but there was no

detected significant effect on biomineral ureolytic activity as suggested by the multiple

regression results shown in Table 2. The lack of statistical significance describing the

relationship between *ureC* gene copies and ureolytic activities disagrees with bivariate

correlation studies done by Fisher et al. (2016) and Sun et al. (2019), where it was

found that soil ureolysis rates were significantly correlated with *ureC* gene copies.

Notably, neither studies discussed effect size and used a small sample size (n < 12) for

analyses describing individual soil horizons [32]. Conversely, other soil urease studies

have also found that ureolytic activities are correlated with total nitrogen (TN), total

carbon (TC), and soil organic carbon (SOC) concentrations, but not the abundance of

ureC genes as in agreement with our study [33]. The regression results suggest that

ureolytic gene abundance is insufficient in predicting ureolytic activity in a linear model.

Greater abundances of potentially ureolytic bacteria indicated by proxy of sample

ureC gene concentrations, may not be correlated with biomineral ureolytic rates as

suggested by the regression results. That *ureC* was detectable indicates that part of the

bacterial community in the biomineral samples has the urease-positive genotype, but

not all bacteria with the *ureC* may be displaying a urease-positive phenotype [34]. This

is because urease activity may not be expressed under the growth conditions found in

urine drain pipes, and may explain why urease activities did not differ significantly when

grouped by urinal type [34]. Expression of the urease-positive genotype and the

eventual translation into the urease protein is regulated at the transcriptional level rather

than at the genomic level [35–37].

That *ureC* gene abundance is not a statistically significant predictor of biomineral

ureolytic activity is likely due to the need for environmental conditions that would induce

certain microbial transcriptional responses that cause an increase in urease activity.

When comparing *ureC* copies per g VS, values grouped by intrasystem sampling

location differed significantly between cartridge vs. gallery drain (Kruskal-Wallis: $p <$

0.001; Wilcoxon Rank Sum: $p < 0.001$) and cartridge vs. gallery main drain (Kruskal-

Wallis: $P < 0.001$; Wilcoxon Rank Sum: $p < 0.001$). However, Figure 3 reinforces

hypothesis testing results in that samples from the main drain with the lowest functional

gene concentrations exhibited maximal ureolytic activity of all samples as predicted by

the multiple regression model. One possible explanation is that the main drains and low-

flow urinal drain lines are exposed to flush and sink water, which leads to a decrease in

nitrogen concentrations in the stream contacting the biofilm due to dilution. In response,

the ureolytic ammonia oxidizing bacterial community may be upregulating *ureC*

transcription to produce more urease to convert the urea into ammonia at a faster rate

for pH regulation or to acquire ammonia for biomass production or energy generation

[38]. This hypothesis is in agreement with the literature, as researchers have shown that

ureC mRNA transcripts were 10-fold higher in *Ruminococcus albus* cultures grown on

peptides than those grown using an ammonia or urea-based media [39].

From Figure 4, the observation that conventional and low-flow urinals can have

similar *in situ* ureolytic rates with those from waterless urinals is consistent with the

regression results where it was found that urinal type is neither a significant ($p = 0.521$)

and practical ($\hat{\beta} = -0.12$) predictor of the *in vitro* biomineral ureolytic activity. Figure 4

demonstrates that Dunnigan northbound, a waterless urinal site, exhibited the greatest

maximum *in situ* ureolytic rates. Conversely, Tejon Pass, an SRRA fitted with

conventional urinals, ranked 2nd of all sites screened for *in situ* ureolytic rate.

Our findings indicate that flush water alone may not be an adequate preventative

measure for preventing ureolytic biomineralization, as urease activity can be as strong

in conventional and low-flow biomineralization as it is in waterless biomineralization,

even if there is a smaller ureolytic community in flush type urinals as indicated by low

relative *ureC* gene concentrations shown in Figure 3. It is also possible that flush water

may also influence the precipitation chemistry in drain lines, as flush water containing

elevated magnesium and calcium concentrations may contribute to crystallization. While

the smaller abundance of *ureC* gene concentrations in low-flow urinal samples is

insufficient in accounting for the similar ureolytic activities exhibited by the two urinal

types and intrasystem sampling locations, the differences in *ureC* gene concentrations

grouped by urinal type shown in Figure 3 may likely be due to a difference in community

structures. Future next-generation-sequencing and microbial ecology studies should

visualize the potentially ureolytic microbial community structure by sequencing the *ureC*

gene in addition to 16S rRNA to visualize the total bacterial community to find

relationships between the bacterial community, environmental factors, and ureolytic

activity.

Figure 4 Comparison of *in situ* trap urea conversion rate for various SRRA with trap-

375   type urinals

In conjunction with measuring bulk parameters such as pH, future studies should

incorporate RT-qPCR to determine the effects of nutrient concentrations on urease

gene expression at the transcriptional level. A future RT-qPCR experiment on ureolytic

biomineral samples can reveal how the effects of varying dilution rates between low-

flow and waterless urinals affects the transcriptional activity of a gene of interest and its

relationship with ureolytic activity.

Conflicts of Interest

383 The authors declare no competing financial interest.

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