Lung organoids and microplastic fibers: a new exposure model for emerging contaminants

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Abstract

Background: Three-dimensional (3D) structured organoids are the most advanced *in vitro* models for studying

human health effects, but they have not yet been applied to evaluate the biological effects associated with

microplastic exposure. Fibers from synthetic clothes and fabrics are a major source of airborne microplastics,

and their release from dryer machines is still poorly understood.

Objectives: In this study, we aimed to establish an *in vitro* organoid model of human lung epithelial cells to

evaluate its suitability for studying the effects of airborne microplastic contamination on humans. Furthermore,

we aimed to characterize the microplastic fibers (MPFs) released in the exhaust filter of a household dryer and

to test their interactions and inflammatory effects on the established lung organoids.

Methods: The polyester fibers emitted from the drying of synthetic fabrics were collected. Morphological

characterization of the fibers released into the air filter was performed by optical microscopy and scanning

electron microscopy (SEM)/energy dispersive x-ray spectroscopy (EDS). The organoids were exposed to

various MPF concentrations (1, 10, and 50 mg L^{-1}) and analyzed by optical microscopy, SEM, and confocal

microscopy. Gene expression analysis of lung-specific genes, inflammatory cytokines, and oxidative stress-

related genes was achieved by quantitative reverse transcription–polymerase chain reaction (qRT-PCR).

Results: We successfully cultured organoids with lung-specific genes. The presence of MPFs did not inhibit

organoid growth, but polarized cell growth was observed along the fibers. Moreover, the MPFs did not cause

inflammation or oxidative stress. Interestingly, the MPFs were coated with a cellular layer, resulting in the

inclusion of fibers in the organoid.

Discussion: This work could have potential long-term implications regarding lung epithelial cells undergoing

repair. This preliminary exposure study using human lung organoids could form the basis for further research

regarding the toxicological assessment of emerging contaminants such as micro- or nanoplastics.

Introduction

Atmospheric contamination through airborne particles and fibers is a long-existing and growing research topic of environmental pollution. Several epidemiological studies have already linked air pollution through ambient atmospheric particles to many adverse human health effects, including respiratory illness, cardiovascular disease, and carcinogenic effects (Churg and Bauer 2000; Loomis et al. 2013; Valavandis et al. 2008). In addition to the well-known air pollutants such as combustion particles from fuel-burning emissions and other particulate organic matter and aerosols (Nel 2005), microplastics (MPs) are now considered as emerging components of air pollution (Zhang et al., 2020 and authors therein).

The human health effects of MPs and nanoplastics (NPs) are usually deduced from *in vivo* exposure tests in animal models, i.e., mammalian or nonmammalian models such as mice, rats, *Xenopus laevis* (De Felice et al. 2018), and zebrafish (Bhagat et al. 2020), and from *in vitro* models such as human cell cultures. After exposure, i.e., inhalation or ingestion, MPs/NPs may pass through biological barriers, leading to translocation in the body tissue. Indeed, the transfer of MPs/NPs into human cells has been demonstrated using a variety of human cell lines, including epidermal cells (Triebskorn et al. 2019), lung epithelial cells (Lim et al. 2019; Schirinzi et al. 2017; Xu et al. 2019), endothelial cells (Barshtein et al. 2016), and intestinal cells (Cortés et al. 2020; Domenech et al. 2020). Human two-dimensional (2D) cell cultures are common *in vitro* models used to evaluate the biological effects associated with MP/NP exposure; however, these systems have some limitations such as an inaccurate representation of the *in vivo* tissue (Costa et al. 2016). Very recently, organoids, which are three-dimensional (3D) cellular structures generated from induced pluripotent stem cells, embryonic stem cells, or adult tissue-resident stem cells (Shamir and Ewald 2014), have been shown to be a powerful tool to overcome the limits of 2D culture. Indeed, organoids, such as kidney (Takasato et al. 2016), brain (Lancaster et al. 2013), intestine (Sato et al. 2011), liver (Huch et al. 2015), pancreas (Dossena et al. 2020), and lung (Sachs et al. 2019) organoids, have recently emerged as attractive model systems that contain key aspects of *in vivo* tissue and organ complexity while being more experimentally manageable than model organisms. Moreover, these 3D structures have been recently applied in nanosafety research (Kämpfer et al. 2020) and modeling diseases such as cancer (Tuveson and Clevers 2019). Nevertheless, organoids have not yet been applied to evaluate the biological effects associated with MP/NP exposure.

MPs are ubiquitous in the environment (Prata et al. 2020). In the atmosphere, a large proportion of the MPs consists of microplastic fibers (MPFs) derived from various sources, including synthetic clothes, textiles, and upholstery (Henry et al. 2019). Undeniably, synthetic clothing mostly made of polyethylene terephthalate (namely, polyester) is considered to be a major source of airborne MPs (Dris et al., 2016). Recent studies also

have demonstrated that a major pathway for MPFs is the atmosphere, through which they can reach very

remote areas such as glaciers (Ambrosini et al. 2019), the Arctic region (Bergmann et al., 2017; Lusher et al.,

2015), and Antarctica (Reed et al., 2018; Waller et al., 2017). Furthermore, Dris et al. (2017) and Liu et al.

(2019) have investigated and compared MPFs in indoor and outdoor air. They found that indoor air contains

considerably higher amounts of natural and synthetic fibers than outdoor air and dust.

MPFs can be released into the wastewater through the washing of synthetic textiles and clothes, as

demonstrated previously in several studies (Browne et al., 2011; De Falco et al., 2019; among others). In

addition to the washing process, the drying of synthetic textiles with a household clothes dryer further emits

MPFs into the atmosphere via the exhaust air. However, the air contamination caused by the drying of synthetic

clothes and fabrics is still poorly understood. Recently, O'Brien et al. (2020) have reported and quantified for

the first time the amount of MPFs released from a clothes dryer into the ambient air. The issue was further

addressed by Kapp and Miller (2020), who studied the spatial distribution of MPFs emitted from the vent of a

clothes dryer directly into the environment.

Human inhalation of atmospheric MPFs has been demonstrated *in vivo* (human lung biopsies) by

Pauly et al. (1998) and Pimentel et al. (1975). The deposition of MPs, however, depends on the properties and

lung anatomy (Churg and Bauer 2000; Lippmann et al. 1980). Currently, it is well known that humans in certain

exposure scenarios, such as industrial workers, are particularly susceptible to diseases caused by airborne

MPs in the lung (Eschenbacher et al. 1999; Hours et al. 2007). In the attempt to estimate the human inhalation

of indoor airborne MPs, Vianello et al. (2019) set up a mannequin simulating the human metabolic rate and

breathing (male person with light activity). The simulated human exposure revealed an average inhaled

concentration of 9.3 ± 5.8 MPs m^{-3} (or 272 MPs per day), corresponding to the value of indoor airborne fibers

reported by Dris et al., 2017 (median value of 5.4 fibers m^{-3}) – disregarding the problem of comparing the

results of studies using different analytical techniques.

The available data or information providing evidence of the negative human health effects of inhaled

MPFs are still rare and insufficient (Prata 2018). To date, the human exposure risk still remains unclear, and

the consequences of MPF exposure are not yet well understood (Prata et al. 2020). Accordingly, there is an

increasing demand for interdisciplinary research between environmental and human health sciences (Dries et

al. 2017). In this study, we propose the application of the innovative human organoid model for human

exposure tests for MPs/NPs and other particulate contaminants. Considering the described exposure risk for

humans to airborne MPs, the use of the human lung as an organoid model in this study is an innovative

experimental approach.

Importantly, exposure tests should not only apply environmentally relevant concentrations but also

environmentally relevant MP properties (Prata et al. 2020), such as nonstandard, fibrous MPs. Therefore, to

test the effects of airborne MPFs on human lung airway organoids (hAOs), polyester fibers emitted from the

drying process of synthetic clothes and fabrics were collected and used as a test contaminant. The specific

aims of this paper were as follows: i) to characterize the release of MPFs in the exhaust filter of a household

dryer machine; ii) to analyze the effect of these environmentally relevant MPFs on hAOs as a possible target

of airborne contamination from synthetic clothes; iii) to characterize the established hAOs; and iv) to evaluate

the use of human organoid models to assess the effects of pollutants to humans.

Material and methods

hAO isolation and culture

The herein applied hAOs were generated from tissue-resident adult stem cells. Healthy human lung tissues

(airway space) were obtained from the Thoracic Surgery and Lung Transplant Unit, IRCCS Ca' Granda

Ospedale Maggiore Policlinico, Milan, Italy. The use of human specimens was approved by the Institutional

Review Board (CE 0001977). Young patients underwent minimal invasive wedge lung resection for

spontaneous pneumothorax.

For the processing of solid lung tissue, the biopsies (0.5–1 cm^3) were minced in small pieces and

rinsed with wash medium (Dulbecco's modified Eagle medium, DMEM) supplemented with 1% fetal bovine

serum, 1% penicillin/streptomycin and 1× Glutamax (see Table S1 for the source and ID numbers of all

reagents used). Fragments were digested in wash medium containing 0.125 mg mL^{-1} collagenase I, 0.125 mg

mL^{-1} dispase II, and 0.1 mg mL^{-1} DNase I for 90 min at 37 °C. The digestion was stopped by adding cold wash

medium, and the suspension was filtered through a 70-μm cell strainer and then spun for 5 min at 400 × g. In

case of visible red pellets, erythrocytes were lysed in 5 mL of red blood cell lysis buffer for 15 min at room

temperature before the addition of wash medium and centrifugation at 400 × g. The cell pellet was mixed with

Matrigel, and 40 μL of Matrigel-cell suspension was allowed to solidify on prewarmed nontissue culture 24-

well plates for 20–30 min. After complete Matrigel solidification, culture medium containing AdDMEM/F12

supplemented with 10 mM Hepes, 1× penicillin/streptomycin, 1× Glutamax, 1% B27, 1.25 mM N-

acetylcysteine, 25 ng mL^{-1} FGF7, 500 ng mL^{-1} RSPO1, 100 ng mL^{-1} FGF10, 100 ng mL^{-1} Noggin, 5 mM

nicotinamide, 500 nM A83.01, and 500 nM SB202190 was added. For establishment of the organoids, the

culture medium was supplemented with 10 μM Rock inhibitor Y27632 during the first three days. The culture

medium was changed every 3–4 days. After 10–14 days, the organoids were removed from the Matrigel, mechanically dissociated into small fragments, and then split 1:4–1:6 in fresh Matrigel, enabling the formation of new organoids. All cell cultures were routinely tested for mycoplasma contamination by quantitative reverse transcription–polymerase chain reaction (qRT-PCR).

hAO characterization

Total RNA from organoids was isolated using TRIzol reagent, according to the manufacturer's instructions. The RNA concentration and purity were verified using a NanoDrop ND-100 spectrophotometer (NanoDrop Technologies). For the qRT-PCR assay, cDNA was synthesised from 200 ng of total RNA with SuperScript IV VILO. The cDNA was diluted 10-fold, and 1 µL of the sample was used as a template for qRT-PCR analysis with SYBR Select Master Mix on a CFX96 thermal cycler (Bio-Rad). The relative expression levels of the selected target genes were determined using the $\Delta\Delta C_t$ method and normalized to the geometric mean of the ACTB and TBP mRNA levels using the primers listed in Table S2.

To characterize the hAO, gene expression analysis by qRT-PCR was performed after 10–14 days of standard culture and again after MPF exposure to investigate possible changes in the lung cell compartments.

Sampling and characterization of MPFs from a dryer machine

To generate environmentally relevant MPFs for the human organoid exposure experiment, polyester fibers emitted from the drying of synthetic clothes and fabrics were collected. Specifically, one polyester t-shirt, six sweatshirts, and two blankets of different colors (dry weight: 5,427 g) were washed in a washing machine and subsequently dried in a common domestic tumble dryer. The dryer filter was previously cleaned by a vacuum cleaner, and all fibers derived from the first drying cycle were discarded. The same items were then washed again, and the synthetic fibers from the filtered exhaust air of the tumble dryer were collected (Figure S1), wrapped in aluminum foil, and transported to the laboratory for subsequent analyses.

The MPFs were first morphologically analyzed with a Leica EZ4D stereomicroscope and then with a Leica DMRA2 light microscope equipped with a Leica DC300 F digital camera. The detailed morphology and the elemental composition of the fibers were studied with a Zeiss LEO 1430 scanning electron microscope (SEM) coupled with a Centaurus detector for energy dispersive x-ray spectroscopy (EDS) analysis. A subsample of MPFs was mounted onto standard SEM stubs and gold-coated. The elemental analysis was

performed using Oxford Instruments INCA version 4.04 software (Abingdon, UK). The operating conditions

were as follows: accelerating voltage, 20 kV; probe current, 360 pA; and working distance, 15.0 mm.

MPF exposure to hAOs

The MPFs collected from the filter of the dryer machine were resuspended at a final concentration of 500 mg

L^{-1} in AdDMEM/F12 supplemented with 10 mM Hepes, 1× penicillin/streptomycin, and 1× Glutamax; then the

sample was sonicated at a high intensity (3 × 5 min) using a Bioruptor sonicator (Diagenode). To obtain

organoid-MPF cocultures, the hAOs were split as described above, and then the organoid fragments were

embedded in Matrigel drops and mixed with MPFs at a ratio of 70% Matrigel with organoids and 30% culture

medium containing the MPFs, at different concentrations (1, 10, and 50 mg L^{-1}). Upon solidification of the

Matrigel, the organoid-MPF coculture was supplemented again with the culture medium containing the MPFs

at the respective concentrations. After ten days, the organoids reached maturation (ca. 200–300 μm in

diameter). At this point, in order to improve the interaction between the organoids and MPFs, the hAOs were

carefully removed from the Matrigel and cultured in suspension with only culture medium containing the MPFs

at the same concentrations for 1 week with gentle agitation. An organoid coculture was cultivated under the

same conditions without MPF exposure as a control. After exposure, the lung organoid-MPF cocultures were

collected for gene expression analysis and fixed for confocal microscopy and SEM analysis, as described

below.

SEM

To study the effects of MPFs on organoid growth, the control and exposed samples (1, 10, and 50 mg L^{-1})

were fixed in a mixture of 4% paraformaldehyde and 2% glutaraldehyde in 0.1 M sodium cacodylate-buffered

solution at pH 7.4. After several washes in the same buffer, the samples were post-fixed in 1% OsO_4 for 1.5 h

at 4°C and then dehydrated in a graded ethanol series. As a final step, the organoids were treated with

hexamethyldisilazane for complete chemical dehydration. All samples were mounted onto standard aluminum

stubs, gold sputtered, and analyzed under a Zeiss LEO 1430 SEM at 20 kV.

Immunofluorescence staining and confocal microscopy

For confocal microscopy analysis, the organoids were washed with 0.01 M phosphate-buffered saline, pH 7.4

(PBS) and fixed for 15 min in 4% paraformaldehyde, incubated in 0.05 M NH_4Cl in PBS for 30 min,

permeabilized for 15 min in PBS containing 1% bovine serum albumin (BSA) and 0.2% Triton X-100, and

blocked for 30 min in 1% BSA. The organoids were incubated overnight at 4 °C with a mouse monoclonal

antibody against α-tubulin diluted in 0.1% BSA/PBS. The samples were then rinsed in PBS and incubated for

3 h at room temperature with rhodamine phalloidin (cytoskeleton) and the secondary Alexa Fluor 488-

conjugated goat anti-mouse antibody. After several washes with PBS, the organoids were finally incubated for

5 min at room temperature with the DNA dye Hoechst 33342 (1:5000). At the end of the staining procedure,

the organoids were washed with PBS (3 × 5 min), mounted in 1:2 PBS/glycerol, and observed under a Nikon

A1 confocal microscope.

Inflammatory cytokine and oxidative stress evaluation

To investigate the possible changes in the inflammatory cytokine expression in hAOs after MPF exposure,

gene expression analysis by qRT-PCR was performed as described above at the end of the coculture with the

highest MPF concentration using the primers listed in Table S2. Considering that the exposure of MPFs could

induce oxidative stress in human tissues (Hu and Palić 2020), we analyzed the expression of genes involved

in oxidative stress pathways reported in relation to MPF exposure, including superoxide dismutase family

genes (SOD1 and SOD2), glutathione detox-related genes (GSTA1 and GPX1), catalase (CAT), and ROS-

controlling genes (NOX2, COX1, and ND1). In addition, the capacity of our 3D structures to respond to

standard inflammatory stimuli was evaluated after treatment with poly(I:C) (50 µg mL^{-1}) as a positive control.

Data analysis

For statistical analysis of the gene expression data, a two-way analysis of variance followed by Bonferroni's

multiple comparison test (α = 0.05) was performed. Photoshop was used to export the image graphs.

213   **Results**

hAO isolation and characterization

The hAOs were successfully generated from the biopsies of three healthy lung donors. A heatmap of selected

genes (Figure 1) demonstrated that the hAO lines derived from the three different donors have similar gene

expression profiles (compared to human pancreas organoids as a control).

The organoids were analyzed by optical microscopy, confocal 3D construction, and SEM imaging. The

normal organoids (control group) exhibited a spherical shape with a typical diameter of 200–300 µm and an

inner cavity (Figure 2A–B). Confocal microscopy confirmed that the cells were differentiated, showing the

presence of ciliated cells and nonciliated cells, including club cells (identified by the presence of microvilli), that

were irregularly distributed on the surface and corresponding to their *in vivo* position (Figure 2C–E).

Immunofluorescence experiments with the actin cytoskeleton marker revealed the different shapes and

dimensions of the cells containing nuclei, which were visible in the surface section (Figure 2C–E) and

transverse plane (Figure 2D). Moreover, the inner surface of the organoids established a differentiated cell

structure, including ciliated cells facing the organoid cavity, as demonstrated in Figure 2D. The presence of

cilia at the inner surface of the organoids could be explained by the rotating motion of the organoid cavity,

which can be seen in a video recording of the organoids (Video S1). High-resolution SEM imaging of the

organoid surface showed the same irregularly shaped ciliated and nonciliated cells, displaying the dimensions

of the cilia and microvilli (Figure 2F).

MPFs in the air filter of a dryer machine

The total weight of the MPFs removed from the air filter of a dryer machine was 2.3 g (Figure S1), which was

released from 5.4 kg of 100% polyester fabrics (0.42 g of dry MPFs kg^{-1} dry fabric). The procedure was

repeated once again with the same fabrics (rewashing and redrying), and the release of MPFs was similarly

high. Optical microscopy experiments demonstrated fibers of different colors, reflecting the color mixture of the

dried materials (Figure 3A–B). For a more detailed analysis, SEM (Figure 3C–E) revealed that the MPFs varied

in size and morphology. The fiber length ranged from 139 μm to 1,535 μm (mean: 663 ± 333 μm). Interestingly,

the shape of these MPFs differed at the transverse surface; no MPFs had a round profile along the entire

length, but they exhibited a varying profile from flat and twisted to tattered (Figure 3C–E). Flat MPFs exhibited

maximum widths of 21–24 μm along their widest dimension and a minimum height of 1–3 μm along their

thinnest dimension. Furthermore, we observed that some fibers showed a rough surface, where small pieces

of <10 μm flaked off from the fiber (Figure 3D) or a chapped ending was revealed (Figure 3E), which could

lead to the release of very small MPs from the fibers.

Considering the average dimensions of the fibers described by SEM analyses, we obtained a fiber

volume of 28,485 μm^3 fiber^{-1} (633 μm × 22.5 μm × 2 μm) or 2.8 × 10^{-5} mm^3 fiber^{-1}, indicating that 59 × 10^6 fibers

were collected in the filter of a dryer machine after washing 5,427 g of polyester fabrics (considering that the

density of polyester is 1.38 mg mm^{-3}). Elemental analysis showed a C:O ratio of 74:26, which corresponds to

that of polyester (see Figure S2 for the EDS spectrum), thus confirming that the fibers were made of polyester.

Effects of MPFs on hAOs

The hAOs exposed to MPFs were affected by all concentrations of MPFs (1, 10, and 50 mg L^{-1}). Optical microscopy, confocal 3D construction, and SEM image analyses revealed that the organoids exposed to MPFs were not inhibited in their growth and did not exhibit any cellular abnormalities compared to the control group, independent of the MPF concentration (Figure 4A–F). Most organoids did not interact with the fibers, or only to a small extent, and maintained their radial (almost spherical) architecture (Figure 4D–F). In addition, no difference in the expression of epithelial to mesenchymal transition markers was observed (Figure 5A). Moreover, fluorescence staining of the surface (Figure 6A–B) and transverse (Figure 6C) sections of the organoids as well as high-resolution SEM images (Figure 6D–H) showed the same cellular structure as the controls, including ciliated and nonciliated cells, even when the organoids were in contact with MPFs. To verify that the lung heritage of hAOs was not affected by their exposure to MPFs, we performed gene expression analysis of the typical epithelial lung markers NKX2.1 and CLDN1 as well as the specific airway lung markers SFTPA1 and SFTPC (AT2 cells), SCGB1A1 (club cells), NPHP1 and DNAH5 (ciliated cells), and KRT5 (basal cells); no significant differences were observed between the organoids exposed to MPFs and the control organoids (Figure 5B). The only notable observation was a higher density of polarization of the cytoskeleton near the fiber contact site, visible as an intensification of the red color in Figure 6B. Owing to an organoid that had broken apart (likely during the transfer of the dehydrated organoid onto the aluminum stub), we were able to take SEM images of the inner surface of the organoids and could observe the same cellular differentiation (multiciliated and nonciliated cells) as the outer surface (Figure 6E). Rather than depending on the MPF concentration, the observed effects varied with the degree of MPF-organoid contact, which presumably depended on the development phase of the organoids. The organoids that made contact with fibers during maturation grew around the fibers and integrated them into their body (Figure 6F–H).

Interestingly, the organoids closely bonded to an MPF formed an organoid-fiber interlacement or fully encompassed the MPF. We further observed that some organoids grew polarized around the fibers. For example, Figure 7A–G shows cell growth along the length of a fiber. The organoids seemingly included and phagocytosed fibers by developing netlike extensions of cells (clearly visible in Figure 7D–F), resulting in a full internalization of the MPFs. While the cells of the organoid body growing along the length of a fiber exhibited a differentiated structure (Figure 7A–B), the thin cell extension covering the surface of a fiber exiting the organoid body (Figure 7D–F) was rather smooth and uniform. These observations of cells covering a fiber with a thin layer stands in contrast to the images depicted in Figure 6F–H showing that cells did not cover the fibers

281 once they left the organoid. We also noticed that organoids in suspension with MPFs can be damaged by

282 fibers piercing the cell tissue (Figure 7C).

Inflammation and oxidative stress evaluation

To investigate possible changes in the expression of inflammatory cytokines and oxidative stress-related

genes, gene expression analysis of hAOs was performed at the end of their coculture with the highest MPF

concentration. The results revealed no significant differences in the gene expression of cytokines or

oxidative stress-related genes in hAOs exposed to MPFs in relation to the control hAOs (Figure 5C–D). In

contrast, an increase of inflammatory markers was observed after poly(I:C) treatment, as reported by others

(Jose et al. 2020) (Figure 5C).

Discussion

The first objective of the present study was to set up an innovative *in vitro* model based on the use of

3D structures that could adequately represent the native lung, where emerging atmospheric contaminants

such as synthetic materials could have adverse human health effects. And secondly, we wanted to use this

model to investigate the potential effects of polyester fibers isolated from a dryer machine on human cell

biology.

Organoids have been recently exploited to overcome the limitations of 2D cell culture systems. Due to

their capacity to self-organize into minimal biological units and thus potentially recapitulate the functionality

and complexity of the tissue of origin, they have emerged as powerful models for studying human development

and disease (Fatehullah et al., 2016). Although lung organoids, including those presented herein, do not yet

recapitulate all of the complex structures and cellular interactions associated with the different regions of the

lung (e.g., trachea, bronchi, bronchioles, alveoli and all mesenchymal and vascular compartments), they

contain specific cell types present in the epithelium of the original organ. Therefore, they can be applied for

basic and translational research (Barkauskas et al. 2017). We showed that lung-specific cell types, including

basal, AT2, and ciliated cells, are represented in our human lung organoids (Figure 7A).

For the first time, this study presents an attempt to use human lung organoids for toxicological

assessment of the effects of MPFs. Synthetic textiles have been identified and extensively described as one

of the major sources of MPFs in wastewater (Browne et al., 2011; De Falco et al., 2019). The current research

is the first to report the amount and features of MPFs released from synthetic fabrics in a dryer machine,

addressing potential atmospheric contamination. In fact, a recent paper by O'Brien et al. (2020) has shown

that the use of dryer machines causes an increase of the MPF concentration in the ambient air; however, they

did not measure the total amount of MPFs released by the drying process. Washing and drying textiles in

household machines present two possible MPF emission pathways: one being through the wastewater

(residual water in textiles after the centrifugation cycle of the washing machine), and one being via the air

passing by an integrated air filter. The former goes directly into domestic wastewater without filters. Here, we

focused on the air pathway; MPFs released by textiles and forced by a strong air flux into the air exhaust are

retained by the air filter and then specifically collected at the end of each drying cycle. Considering the evidence

reported by O'Brien et al. (2020), the air filter does not remove all MPFs coming from the dryer machine. Thus,

the total amount of MPFs released in the air flux can be even higher.

The effects of MPs/NPs on human health have been tested previously using a variety of 2D cell lines.

The observed effects were very diverse and depended on the properties of the particle (e.g., size and polymer),

the duration and concentration of exposure, and the tested cell line. Interestingly, most of the existing studies

analyzing the cytotoxicity of virgin NPs have demonstrated a lack of adverse effects in intestinal, cerebral, and

epithelial cell lines. A concentration-dependent significant decrease in cell viability (Lim et al. 2019) and

disruption in gene transcription and protein expression after exposure to polystyrene NPs (25 and 70 nm) were

only detected for lung cells (Xu et al., 2019), highlighting the need to study the effects of inhaled MPs/NPs.

Moreover, with only a few exceptions (Magrì et al. 2018), all previous studies used spherical polystyrene NPs,

which are limited in a realistic exposure scenario. In particular, tests representing the human lung should use

MPs/NPs in a size and shape that have been shown to be present in the air. Since a substantial portion of

airborne environmental MPs consist mostly of MPFs between 200 and 600 µm (Dris et al. 2016; Henry et al.

2019) and fibers >250 µm have been found in human lungs (Pauly et al. 1998), the MPFs applied in the current

study represent an environmentally relevant and realistic test material.

Based on a review of studies regarding atmospheric MPs, Prata et al. (2018) calculated that a person's

lungs could be exposed to 26–130 airborne MPs a day. In addition, Vianello et al. (2019) have shown that half

of the inhaled and identified MPs are smaller than 50 µm. It has already been reported that inhalability,

deposition, and persistence of MPs in the lungs depend on their properties and the different regions of the

respiratory tract (Prata 2018). For fibers, in general, deposition in the respiratory tract is well summarized by

Mossman et al. (2011). For example, it has been assumed that the length of MPs is of secondary importance

because the fibers can align themselves in parallel with the respiratory flow axis and can penetrate to the

periphery of the lung and into the alveoli, as already observed in lung biopsies by Pauly et al. (1998). Once

they enter deep into the lung, particles (especially the longer fibers) tend to avoid clearance since the

magnitude of retention in the lungs depends on the fiber diameter and length (Timbrell 1982). If not cleared by

biological mechanisms, MPFs persist in the lung.

The existing database of fiber toxicity studies (Mossman et al., 2011) suggests that human exposure

to respirable fibers may induce significant and persistent pulmonary inflammation, cell proliferation, or fibrosis.

Immune cells are not the only source of pro- and anti-inflammatory cytokines in the lung; the epithelial cells

themselves are known to produce numerous cytokines directly in response to injury or pathogens (Feng et al.

2020), and they contribute to the impressive innate immunity functions of the lung (Whitsett and Alenghat

2015).

Despite the mild and somewhat controversial effects obtained in 2D systems exposed to MPs/NPs,

we did not find striking results. Indeed, our findings did not show any significant activation of the inflammatory

or oxidative stress pathway at a molecular level. Similarly, confocal microscopy and SEM images did not

provide any evidence of organoid growth inhibition or cellular differentiation; these findings were the same as

those of the control group. This observation was not completely surprising, as the field of biomaterials is

increasingly demonstrating how plastic can be applied as an inert scaffold to support cell growth for human

tissue engineering without altering the biological features (Jubeli et al. 2019). Although the biocompatibility of

several plastic materials has already been demonstrated, polyester fibers can contain additives that can be

added to textiles in order to impart desirable physical, chemical, and biological (i.e., antimicrobial) properties

(Sait et al. 2021). Since additives are typically not covalently bonded to the polymer matrix and can therefore

leach, they are considered more harmful than the polymer itself, inducing several effects such as endocrine

disruption (Kitamura et al., 2005), genotoxicity (Zhao et al., 2013), immunotoxicity (Qiu et al., 2018), and

developmental toxicity (Mu et al., 2018). Our results did not show inflammation- or oxidative stress-induced

effects of MPFs on hAO development, suggesting that additives, whenever they were present, did not cause

adverse effects.

Of note, the most relevant result related to MPF exposure was the polarization of the cell growth along

the fibers such that the organoid was induced to envelop the encountered fibers by a cellular layer. This

behavior was clearly observed in different cases and at different MPF concentrations, indicating that it should

be carefully considered as a possible adverse effect of MPFs. This effect is particularly alarming if we consider

repair processes that could follow MPF inhalation, as it is known that fibers can be retained in the lungs.

Based on the findings of this study, during the repair phase of a damaged lung epithelium, the

presence of nonbiodegradable fibers may lead to their inclusion in the repaired tissue with unknown long-term

effects. This scenario is clearly only a speculation based on an experimental observation performed on a

surrogate of the real organ, but it could be the basis for further studies using this promising tool.

References

Ambrosini R, Azzoni RS, Pittino F, Diolaiuti G, Franzetti A, Parolini M. 2019. First evidence of microplastic contamination in the supraglacial debris of an alpine glacier. Environ Pollut 253:297–301, PMID: 31323612, https://doi.org/10.1016/j.envpol.2019.07.005.

Barkauskas CE, Chung M-I, Fioret B, Gao X, Katsura H, Hogan BLM. 2017. Lung organoids: current uses and future promise. Development 144(6):986–997, PMID: 28292845, https://doi.org/10.1242/dev.140103.

Barshtein G, Livshits L, Shvartsman LD, Shlomai NO, Yedgar S, Arbell D. 2016. Polystyrene Nanoparticles Activate Erythrocyte Aggregation and Adhesion to Endothelial Cells. Cell Biochem Biophys 74(1):19–27, PMID: 26972298, https://doi.org/10.1007/s12013-015-0705-6.

Bergmann M, Wirzberger V, Krumpen T, Lorenz C, Primpke S, Tekman MB, et al. 2017. High Quantities of Microplastic in Arctic Deep-Sea Sediments from the HAUSGARTEN Observatory. Environ Sci Technol 51(19):11000–11010, PMID: 28816440, https://doi.org/10.1021/acs.est.7b03331.

Bhagat J, Zang L, Nishimura N, Shimada Y. 2020. Zebrafish: An emerging model to study microplastic and nanoplastic toxicity. Sci Total Environ 728:138707, PMID: 32361115, https://doi.org/10.1016/j.scitotenv.2020.138707.

Browne MA, Crump P, Niven SJ, Teuten E, Tonkin A, Galloway TS, et al. 2011. Accumulation of Microplastic on Shorelines Wordwide: Sources and Sinks. Environ Sci Technol 45(21):9175–9179, PMID: 21894925, https://doi.org/10.1021/es201811s.

Churg A, Bauer M. 2000. Ambient Atmospheric Particles in the Airways of Human Lungs. Ultrastruct Pathol 24(6):353–361, PMID: 11206332, https://doi.org/10.1080/019131200750060014.

Cincinelli A, Scopetani C, Chelazzi D, Lombardini E, Martellini T, Katsoyiannis A, et al. 2017. Microplastic in the surface waters of the Ross Sea (Antarctica): Occurrence, distribution and characterization by FTIR. Chemosphere 175:391–400, PMID: 28236709, https://doi.org/10.1016/j.chemosphere.2017.02.024.

Cortés C, Domenech J, Salazar M, Pastor S, Marcos R, Hernández A. 2020. Nanoplastics as a potential environmental health factor: effects of polystyrene nanoparticles on human intestinal epithelial Caco-2 cells. Environ Sci Nano 7(1):272–285, https://doi.org/10.1039/C9EN00523D.

Costa EC, Moreira AF, de Melo-Diogo D, Gaspar VM, Carvalho MP, Correia IJ. 2016. 3D tumor spheroids: an overview on the tools and techniques used for their analysis. Biotechnol Adv 34(8):1427–1441, PMID: 27845258, https://doi.org/10.1016/j.biotechadv.2016.11.002.

De Falco F, Di Pace E, Cocca M, Avella M. 2019. The contribution of washing processes of synthetic clothes to microplastic pollution. Sci Rep 9(1):6633, PMID: 31036862, https://doi.org/10.1038/s41598-019-43023-x.

De Felice B, Bacchetta R, Santo N, Tremolada P, Parolini M. 2018. Polystyrene microplastics did not affect body growth and swimming activity in Xenopus laevis tadpoles. Environ Sci Pollut Res 25(34):34644–34651, PMID: 30317408, https://doi.org/10.1007/s11356-018-3408-x.

Domenech J, Hernández A, Rubio L, Marcos R, Cortés C. 2020. Interactions of polystyrene nanoplastics with in vitro models of the human intestinal barrier. Arch Toxicol 94(9):2997–3012, PMID: 32592077, https://doi.org/10.1007/s00204-020-02805-3.

Dossena M, Piras R, Cherubini A, Barilani M, Dugnani E, Salanitro F, et al. 2020. Standardized GMP-compliant scalable production of human pancreas organoids. Stem Cell Res Ther 11(1):94, PMID: 32127043, https://doi.org/10.1186/s13287-020-1585-2.

Dris R, Gasperi J, Mirande C, Mandin C, Guerrouache M, Langlois V, et al. 2017. A first overview of textile fibers, including microplastics, in indoor and outdoor environments. Environ Pollut 221:453–458, PMID: 27989388, https://doi.org/10.1016/j.envpol.2016.12.013.

Dris R, Gasperi J, Saad M, Mirande C, Tassin B. 2016. Synthetic fibers in atmospheric fallout: A source of microplastics in the environment? Mar Pollut Bull 104(1-2):290–293, PMID: 26787549, https://doi.org/10.1016/j.marpolbul.2016.01.006.

Eschenbacher WL, Kreiss K, Lougheed MD, Pransky GS, Day B, Castellan RM. 1999. Nylon Flock–Associated Interstitial Lung Disease. Am J Respir Crit Care Med 159(6):2003–2008, PMID: 10351952, https://doi.org/10.1164/ajrccm.159.6.9808002.

Fatehullah A, Tan SH, Barker N. 2016. Organoids as an in vitro model of human development and disease. Nat Cell Biol 18(3):246–254, PMID: 26911908, https://doi.org/10.1038/ncb3312.

Feng S, Achoute L, Margolskee RF, Jiang P, Wang H. 2020. Lipopolysaccharide-Induced Inflammatory Cytokine Expression in Taste Organoids. Chem Senses 45(3):187–194, PMID: 31993633, https://doi.org/10.1093/chemse/bjaa002.

Henry B, Laitala K, Klepp IG. 2019. Microfibres from apparel and home textiles: Prospects for including microplastics in environmental sustainability assessment. Sci Total Environ 652:483–494, PMID: 30368178, https://doi.org/10.1016/j.scitotenv.2018.10.166.

Hernandez E, Nowack B, Mitrano DM. 2017. Polyester Textiles as a Source of Microplastics from Households: A Mechanistic Study to Understand Microfiber Release During Washing. Environ Sci Technol 51(12):7036–7046, PMID: 28537711, https://doi.org/10.1021/acs.est.7b01750.

Hours M, Févotte J, Lafont S, Bergeret A. 2007. Cancer mortality in a synthetic spinning plant in Besançon, France. Occup Environ Med 64(9):575–581, PMID: 17227838, https://doi.org/10.1136/oem.2006.028282.

Hu M, Palić D. 2020. Micro- and nano-plastics activation of oxidative and inflammatory adverse outcome pathways. Redox Biol 37:101620, PMID: 32863185, https://doi.org/10.1016/j.redox.2020.101620.

Huch M, Gehart H, van Boxtel R, Hamer K, Blokzijl F, Verstegen MMA, et al. 2015. Long-Term Culture of Genome-Stable Bipotent Stem Cells from Adult Human Liver. Cell 160(1-2):299–312, PMID: 25533785, https://doi.org/10.1016/j.cell.2014.11.050.

Jose SS, De Zuani M, Tidu F, Hortová Kohoutková M, Pazzagli L, Forte G, et al. 2020. Comparison of two human organoid models of lung and intestinal inflammation reveals Toll-like receptor signalling activation and monocyte recruitment. Clin Transl Immunol 9(5), PMID: 32377340, https://doi.org/10.1002/cti2.1131.

Jubeli E, Khzam A, Yagoubi N. 2019. Cells integration onto scaffolds prepared from polyester based polymers – importance of polymer thermal properties in addition to hydrophilicity. Int J Polym Mater Polym Biomater 68(17):1068–1077, https://doi.org/10.1080/00914037.2018.1525549.

Kämpfer AAM, Busch M, Schins RPF. 2020. Advanced *In Vitro* Testing Strategies and Models of the Intestine for Nanosafety Research. Chem Res Toxicol 33(5):1163–1178, PMID: 32383381, https://doi.org/0.1021/acs.chemrestox.0c00079.

Kapp KJ, Miller RZ. 2020. Electric clothes dryers: An underestimated source of microfiber pollution. PLOS ONE 15(10):e0239165, PMID: 33027292, https://doi.org/10.1371/journal.pone.0239165.

Lancaster MA, Renner M, Martin C-A, Wenzel D, Bicknell LS, Hurles ME, et al. 2013. Cerebral organoids model human brain development and microcephaly. Nature 501(7467):373–379, PMID: 23995685, https://doi.org/10.1038/nature12517.

Lim SL, Ng CT, Zou L, Lu Y, Chen J, Bay BH, et al. 2019. Targeted metabolomics reveals differential biological effects of nanoplastics and nanoZnO in human lung cells. Nanotoxicology 13(8):1117–1132, PMID: 31272252, https://doi.org/10.1080/17435390.2019.1640913.

Lippmann M, Yeates DB, Albert RE. 1980. Deposition, retention, and clearance of inhaled particles. Occup Environ Med 37(4):337–362, PMID: 7004477, https://doi.org/10.1136/oem.37.4.337.

Liu C, Li J, Zhang Y, Wang L, Deng J, Gao Y, et al. 2019. Widespread distribution of PET and PC microplastics in dust in urban China and their estimated human exposure. Environ Int 128:116–124, PMID: 31039519, https://doi.org/10.1016/j.envint.2019.04.024.

Loomis D, Grosse Y, Lauby-Secretan B, Ghissassi FE, Bouvard V, Benbrahim-Tallaa L, et al. 2013. The carcinogenicity of outdoor air pollution. Lancet Oncol 14(13):1262–1263, PMID: 25035875, https://doi.org/10.1016/S1470-2045(13)70487-X.

Lusher AL, Tirelli V, O'Connor I, Officer R. 2015. Microplastics in Arctic polar waters: the first reported values of particles in surface and sub-surface samples. Sci Rep 5:14947, PMID: 26446348, https://doi.org/doi:10.1038/srep14947.

Magrì D, Sánchez-Moreno P, Caputo G, Gatto F, Veronesi M, Bardi G, et al. 2018. Laser Ablation as a Versatile Tool To Mimic Polyethylene Terephthalate Nanoplastic Pollutants: Characterization and Toxicology Assessment. ACS Nano 12(8):7690–7700, PMID: 29944342, https://doi.org/10.1021/acsnano.8b01331.

Mossman BT, Lippmann M, Hesterberg TW, Kelsey KT, Barchowsky A, Bonner JC. 2011. Pulmonary Endpoints (Lung Carcinomas and Asbestosis) Following Inhalation Exposure to Asbestos. J Toxicol Environ Health Part B 14(1-4):76–121, PMID: 21534086, https://doi.org/10.1080/10937404.2011.556047.PMID: 21534086.

Munari C, Infantini V, Scoponi M, Rastelli E, Corinaldesi C, Mistri M. 2017. Microplastics in the sediments of Terra Nova Bay (Ross Sea, Antarctica). Mar Pollut Bull 122(1-2):161–165, PMID: 28645759, https://doi.org/10.1016/j.marpolbul.2017.06.039.

Napper IE, Thompson RC. 2016. Release of synthetic microplastic plastic fibres from domestic washing machines: Effects of fabric type and washing conditions. Mar Pollut Bull 112(1-2):39–45, PMID: 27686821, https://doi.org/10.1016/j.marpolbul.2016.09.025.

Nel A. 2005. Air Pollution-Related Illness: Effects of Particles. Science 308(5723):804–806, PMID: 15879201, https://doi.org/10.1126/science.1108752.PMID: 15879201.

O'Brien S, Okoffo ED, O'Brien JW, Ribeiro F, Wang X, Wright SL, et al. 2020. Airborne emissions of microplastic fibres from domestic laundry dryers. Sci Total Environ 747:141175, PMID: 32781315, https://doi.org/10.1016/j.scitotenv.2020.141175.

Pauly JL, Stegmeier SJ, Allaart HA, Cheney RT, Zhang PJ, Mayer AG, et al. 1998. Inhaled cellulosic and plastic fibers found in human lung tissue. Cancer Epidemiol Prev Biomark 7(5):419–428, PMID: 9610792.

Peeken I, Primpke S, Beyer B, Gütermann J, Katlein C, Krumpen T, et al. 2018. Arctic sea ice is an important temporal sink and means of transport for microplastic. Nat Commun 9(1):1505, PMID: 29692405, https://doi.org/10.1038/s41467-018-03825-5.

Pimentel JC, Avila R, Lourenço AG. 1975. Respiratory disease caused by synthetic fibres: a new occupational disease. Thorax 30(2):204–219, PMID: 1179318, https://doi.org/10.1136/thx.30.2.204.

Prata JC. 2018. Airborne microplastics: Consequences to human health? Environ Pollut 234:115–126, PMID: 29172041, https://doi.org/10.1016/j.envpol.2017.11.043.

Prata JC, da Costa JP, Lopes I, Duarte AC, Rocha-Santos T. 2020. Environmental exposure to microplastics: An overview on possible human health effects. Sci Total Environ 702:134455, PMID: 31733547, https://doi.org/10.1016/j.scitotenv.2019.134455.

Reed S, Clark M, Thompson R, Hughes KA. 2018. Microplastics in marine sediments near Rothera Research Station, Antarctica. Mar Pollut Bull 133:460–463, PMID: 30041337, https://doi.org/10.1016/j.marpolbul.2018.05.068.

Sachs N, Papaspyropoulos A, Zomer-van Ommen DD, Heo I, Böttinger L, Klay D, et al. 2019. Long-term expanding human airway organoids for disease modeling. EMBO J 38(4), PMID: 30643021, https://doi.org/10.15252/embj.2018100300.

Sait STL, Sørensen L, Kubowicz S, Vike-Jonas K, Gonzalez SV, Asimakopoulos AG, et al. 2021. Microplastic
fibres from synthetic textiles: Environmental degradation and additive chemical content. Environ Pollut
Barking Essex 1987 268:115745, PMID: 33065478, https://doi.org/10.1016/j.envpol.2020.115745.

Sato T, Stange DE, Ferrante M, Vries RGJ, van Es JH, van den Brink S, et al. 2011. Long-term Expansion of
Epithelial Organoids From Human Colon, Adenoma, Adenocarcinoma, and Barrett's Epithelium.
Gastroenterology 141(5):1762–1772, PMID: 21889923, https://doi.org/10.1053/j.gastro.2011.07.050.

Schirinzi GF, Pérez-Pomeda I, Sanchís J, Rossini C, Farré M, Barceló D. 2017. Cytotoxic effects of commonly
used nanomaterials and microplastics on cerebral and epithelial human cells. Environ Res 159:579–587,
PMID: 28898803, https://doi.org/10.1016/j.envres.2017.08.043.

Shamir ER, Ewald AJ. 2014. Three-dimensional organotypic culture: experimental models of mammalian
biology and disease. Nat Rev Mol Cell Biol 15(10):647–664, PMID: 25237826,
https://doi.org/10.1038/nrm3873.PMID: 25237826.

Takasato M, Er PX, Chiu HS, Little MH. 2016. Generation of kidney organoids from human pluripotent stem
cells. Nat Protoc 11(9):1681–1692, PMID: 27560173, https://doi.org/10.1038/nprot.2016.098.

Timbrell V. 1982. Deposition and retention of fibres in the human lung. In: Inhaled Particles *V* (W.H. Walton,
ed). Pergamon. 26(1-4):347–369, PMID: 7181276, https://doi.org/10.1007/978-3-0348-7491-5_36.

Triebskorn R, Braunbeck T, Grummt T, Hanslik L, Huppertsberg S, Jekel M, et al. 2019. Relevance of nano-
and microplastics for freshwater ecosystems: A critical review. TrAC Trends Anal Chem 110:375–392,
https://doi.org/10.1016/j.trac.2018.11.023.

Tuveson D, Clevers H. 2019. Cancer modeling meets human organoid technology. Science 364:952–955,
PMID: 31171691 https://doi.org/10.1126/science.aaw6985.

Valavanidis A, Fiotakis K, Vlachogian T. 2008. Airborne Particulate Matter and Human Health: Toxicological
Assessment and Importance of Size and Composition of Particles for Oxidative Damage and Carcinogenic
Mechanisms. J Environ Sci Health Part C 26(4):339–362, PMID: 19034792
https://doi.org/10.1080/10590500802494538.

Vianello A, Jensen RL, Liu L, Vollertsen J. 2019. Simulating human exposure to indoor airborne microplastics
using a Breathing Thermal Manikin. Sci Rep 9(1):8670, PMID: 31209244, https://doi.org/10.1038/s41598-
019-45054-w.

Waller CL, Griffiths HJ, Waluda CM, Thorpe SE, Loaiza I, Moreno B, et al. 2017. Microplastics in the Antarctic
marine system: An emerging area of research. Sci Total Environ 598:220–227, PMID: 28441600,
https://doi.org/10.1016/j.scitotenv.2017.03.283.

Whitsett JA, Alenghat T. 2015. Respiratory epithelial cells orchestrate pulmonary innate immunity. Nat Immunol
16:27–35, PMID: 25521682, https://doi.org/10.1038/ni.3045.

Xu M, Halimu G, Zhang Q, Song Y, Fu X, Li Y, et al. 2019. Internalization and toxicity: A preliminary study of
effects of nanoplastic particles on human lung epithelial cell. Sci Total Environ 694:133794, PMID:
31756791, https://doi.org/10.1016/j.scitotenv.2019.133794.

Zhang Y, Kang S, Allen S, Allen D, Gao T, Sillanpää M. 2020. Atmospheric microplastics: A review on current
status and perspectives. Earth-Sci Rev 203:103118, https://doi.org/10.1016/j.earscirev.2020.103118.

Figures

Figure 1. Gene expression heatmap of human lung airway organoids (hAOs) generated from the biopsies of

three healthy lung donors and a human pancreas organoid (hPO) for comparison.

Figure 2. Characterization of human lung organoids (control group). (A) Stereo microscopy image and (B) optical microscopy image of organoids. (C–E) Immunofluorescence imaging of the organoid surface and the transverse section generated by confocal microscopy showing the cellular organization with a cytoskeletal marker (anti-acetylated tubulin; green) and counterstaining of the actin cytoskeleton (phalloidin 565; red) and nuclei (Hoechst 33342; blue). (F) Pseudo-colored SEM image of the organoid's surface showing irregularly shaped multiciliated (green) and nonciliated cells with microvilli (purple).

559 **Figure 3.** Synthetic fibers obtained from the air filter of a dryer machine. (A–B) Optical microscopy images

560 demonstrating MPFs of different sizes and colors. (C–E) SEM images of MPFs at different magnifications

561 showing varying morphologies.

Figure 4. Human lung organoids exposed to MPFs. (A–C) Optical microscopy images of organoids in suspension with MPFs at increasing concentrations (top to bottom: 1, 10, and 50 mg L^{-1}). (D) Phase contrast image and (E) immunolabeled 3D reconstruction of the same organoid (MPF concentration: 10 mg L^{-1}) showing the interaction between the organoid and a fiber. The organoid was stained for cytoskeletal proteins: microtubules (anti-acetylated tubulin; green), F-actin (phalloidin 565; red), and nuclei (Hoechst 33342; blue). Also visible in blue is a synthetic fiber. (F) SEM image of an organoid (MPF concentration: 50 mg L^{-1}) in contact with a fiber, demonstrating a spherical body and a differentiated cell structure.

Figure 5. qRT-PCR analysis of the control organoids (blue) compared to organoids exposed to MPFs (red;

MPF concentration: 50 mg L^{-1}) for selected target genes: (A) epithelial to mesenchymal transition markers, (B)

general epithelial lung markers and specific airway markers, (C) inflammatory cytokines compared to the

response of positive control organoids to standard inflammatory stimuli (green; poly(I:C) 50 µg mL^{-1}), and (D)

oxidative stress-related genes.

Figure 6. Human lung organoids exposed to MPFs. (A–B) Surface and (C) transverse sections of an

immunolabeled organoid (MPF concentration: 50 mg L^{-1}) showing the cell organization. The organoid was

stained to show ciliated cells (anti-acetylated tubulin; green), F-actin cytoskeleton (phalloidin 565; red), and

nuclei (Hoechst 33342; blue). Also visible in blue is a synthetic fiber. (D) SEM image of an organoid's outer

surface showing ciliated and nonciliated cells (MPF concentration: 10 mg L^{-1}). (E) SEM image of the inner

surface of an organoid that had broken apart, demonstrating cell differentiation facing the organoid's cavity

(MPF concentration: 10 mg L^{-1}). (F–H) SEM images of organoids (MPF concentration: 50 mg L^{-1})

demonstrating a differentiated cell structure and integration of a fiber into the organoid; the image in panel G

is partially pseudo-colored for better visualization (cell tissue, yellow; fiber, blue).

Figure 7. SEM images of human lung organoids exposed to MPFs, partially pseudo-colored for better visualization (cell tissue, yellow; fibers, purple/blue). (A–B) Oriented cell growth of organoids along the length of a fiber; the cell structure and partially coated fibers are shown (MPF concentration: 10 mg L^{-1}). (C) Complete internalization of a fiber inside an organoid (MPF concentration: 50 mg L^{-1}). (D) Net-like extension of cells growing on a fiber. (E and its magnification F) Thin cell extension covering the surface of a fiber exiting the organoid body (MPF concentration: 50 mg L^{-1}). (G) Merged image of three continuous images showing the polarized growth of organoids with the length of multiple fibers (MPF concentration: 50 mg L^{-1}).