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**Pool choice in a vertical landscape: tadpole rearing site flexibility in
phytotelm-breeding frogs**

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24 **Abstract**

25 Many species of Neotropical frogs have evolved to deposit their tadpoles in small water
26 bodies inside plant structures called phytotelmata. These pools are small enough to exclude
27 large predators but have limited nutrients and high desiccation risk. Here, we explore
28 phytotelm use by three common Neotropical species: *Osteocephalus oophagus*, an arboreal
29 frog that periodically feeds eggs to its tadpoles; *Dendrobates tinctorius*, a tadpole-
30 transporting poison frog with cannibalistic tadpoles; and *Allobates femoralis*, a terrestrial
31 tadpole-transporting poison frog with omnivorous tadpoles. We found that *D. tinctorius*
32 occupies pools across the chemical and vertical gradient, whereas *A. femoralis* and *O.*
33 *oophagus* appear to have narrower niches that are restricted primarily by pool height, water
34 capacity, alkalinity, and salinity. *Dendrobates tinctorius* tadpoles are particularly flexible,
35 and can survive in a wide range of chemical, physical, and biological conditions, whereas *O.*
36 *oophagus* seems to prefer small, clear pools and *A. femoralis* occupies medium-sized pools
37 with abundant leaf litter and low salinity. Together, these results show the possible niche
38 partitioning of phytotelmata among frogs, and provide insight into stressors and resilience of
39 phytotelm breeders.

40

41 **Keywords:** parental care; poison frogs; phytotelmata; niche partitioning, competition;
42 tadpoles

43 **Short title:** Pool choice in a vertical landscape

44

45 **Introduction**

46 The survival of young often hinges on the quality of the rearing environments created or
47 chosen by their parents. Whether it is by building nests (birds: (Brown and Brown 1991);
48 mice: (Bult and Lynch 1997, Zhao et al. 2016), digging burrows (rodents: (Svendsen 1976,
49 Ebensperger et al. 2014), or depositing clutches/larvae (e.g., salamanders: (Ruano-Fajardo et
50 al. 2014), frogs: (Pettitt et al. 2018)), the ecology of rearing sites is fundamental in shaping
51 offspring success. For animals with external fertilization, breeding-site choice can be
52 especially important, as optimal conditions for egg clutches may differ from the optimal
53 environment for hatchlings and adults (fish: (Ottesen and Bolla 1998, Mikheev et al. 2001);
54 salamanders: (Nussbaum 1987, Sih and Moore 1993); frogs: (Vági et al. 2019)). Many of
55 these animals assess and prefer biotic and abiotic properties of breeding sites that can
56 enhance offspring survival (Marsh and Borrell 2001, Mokany and Shine 2003, Brown and
57 Shine 2005, Touchon and Worley 2015). Thus, characterizing the nurseries where offspring
58 occur can provide information on the qualities parents assess when making these critical
59 reproductive decisions.

60

61 The challenge of finding an optimal rearing site becomes especially apparent in terrestrial or
62 arboreal breeding animals, whose larval forms are aquatic. For example, some treefrogs lay
63 clutches overhanging water bodies. The placement of clutches is essential, as tadpoles from
64 poorly placed clutches risk hatching and falling onto the ground (Wells 2007, Warkentin
65 2011). One remarkable amphibian strategy adapted to changing habitats between egg and
66 larval stages involves parents that physically transport recently hatched tadpoles from
67 terrestrial oviposition sites to small water-holding plant structures (phytotelmata), ponds, or

68 streams (Summers and Tumulty 2014, Schulte et al. 2020). Unlike other terrestrial breeding
69 amphibians, the physical transport of young allows parents to select the ideal environment for
70 their offspring to develop. Although it is difficult to extensively characterize streams and
71 ponds, microhabitats like phytotelmata provide a unique opportunity to fully measure the
72 biological, chemical, and physical aspects of a nursery, creating an opportunity to interpret
73 deposition choices with a depth of ecological information that is rarely available for other
74 rearing sites. Here, we investigate the chemical and physical properties of aquatic nurseries
75 that predict the presence of Neotropical tadpoles in phytotelm-breeding frogs.

76

77 The use of phytotelmata as tadpole nurseries can seem counterintuitive, as their small volume
78 makes them prone to desiccation and limited in food (Summers and McKeon 2004, Summers
79 and Tumulty 2014). However, their small size provides protection from large predators, and
80 various species have evolved different strategies for their offspring to succeed in these pools
81 (substrate specialisation: (von May et al. 2009, Pettitt et al. 2018); trophic egg-feeding:
82 (Weygoldt 1980, Brown et al. 2010); larval aggression/cannibalism: (Poelman and Dicke
83 2007, Gray et al. 2009, Rojas 2014); pool choice based on specific physical or chemical cues:
84 (Lin et al. 2008, Schulte et al. 2011). Despite the widespread use of phytotelmata, and the
85 non-random site selection shown by many frog parents, few studies go beyond quantifying
86 basic pool dimensions and pool occupation to understand tadpole deposition decisions.
87 Further, the bulk of phytotelm studies are focused only on bromeliads, while work exploring
88 potential trade-offs associated with different phytotelmata (i.e., physical and chemical
89 properties as well as food- and predator-related pressures), and how these change across a
90 vertical gradient, has gone largely overlooked.

91

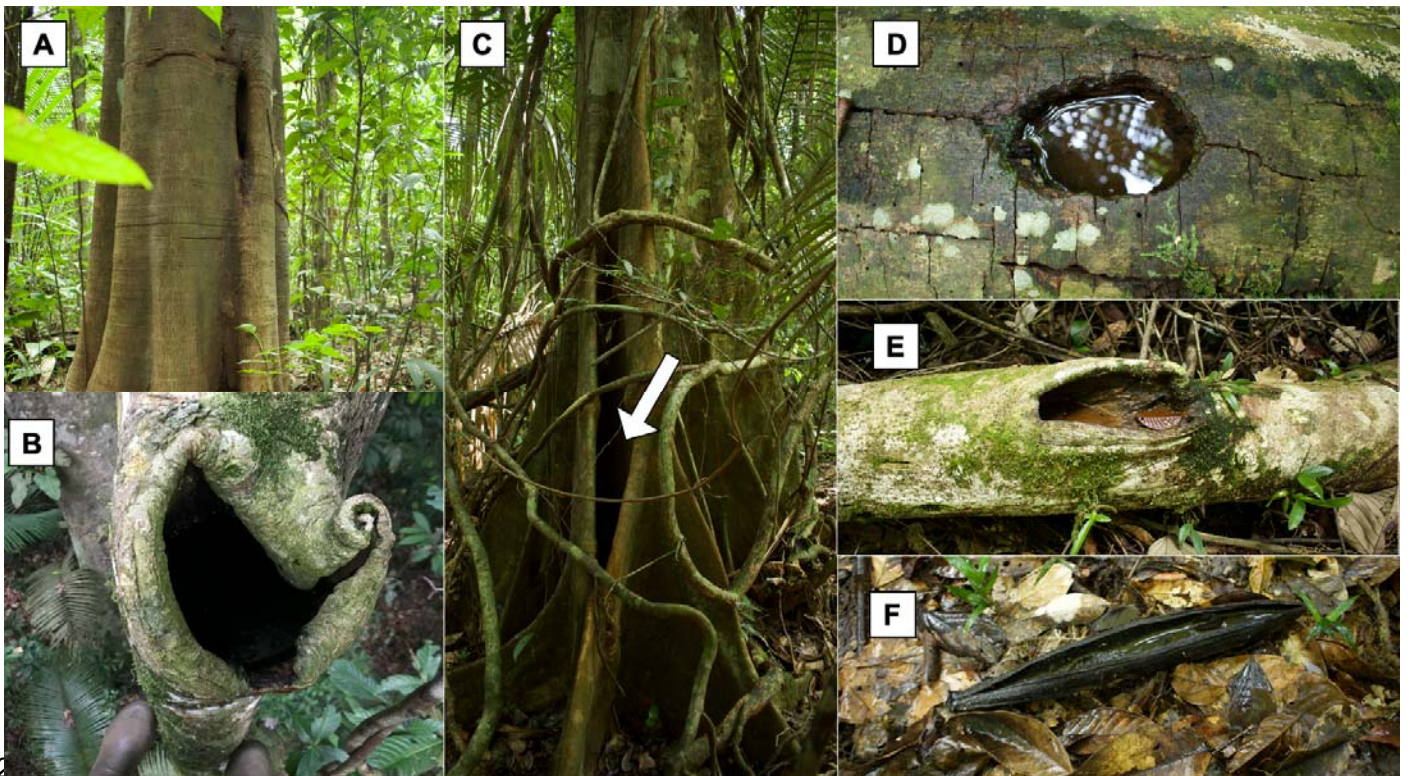
92 To understand what variables drive phytotelm selection, we compared pool occupation by
93 three Neotropical frogs (*Dendrobates tinctorius*, *Allobates femoralis*, and *Osteocephalus*
94 *oophagus*) that were most frequently detected in phytotelmata throughout field surveys at our
95 study site in French Guiana. *Osteocephalus oophagus* is a hylid treefrog with bi-parental care
96 and obligately oophagous tadpoles (Jungfer and Weygoldt 1999, Almendáriz et al. 2000). As
97 in our field site, adults have been found to call and breed in bromeliads, tree-holes, and palm
98 axils close to the forest floor (Jungfer and Weygoldt 1999). Tadpoles of this species develop
99 in the same pool in which the eggs are deposited. *Allobates femoralis* is a terrestrial frog
100 closely related to dendrobatids. Adult males aggressively defend territories during the rainy
101 season (Roithmair 1992, Narins et al. 2003), from which they carry recently hatched tadpoles
102 to a variety of terrestrial pools including phytotelmata close to the ground (Ringler et al.
103 2009, 2013). Tadpoles of this species are omnivorous (McKeon and Summers 2013), but not
104 cannibalistic (Summers and McKeon 2004).

105

106 Following broad species-wide comparisons, we focus on a more detailed analysis of pool
107 choice in *D. tinctorius*, a phytotelm specialist with predatory and cannibalistic tadpoles
108 which are deposited in a range of phytotelm types (e.g., palm bracts, tree-holes, fallen trees;
109 Fig. 1 and Fig. 2) that occur from the forest floor to more than 20 m in vertical height
110 (Gaucher 2002, Rojas 2014, 2015). The use of the high canopy pools is perplexing because
111 *D. tinctorius* is commonly successful in terrestrial pools (Rojas 2014). It is known that pool
112 chemistry can change drastically depending on substrate ('dead' or 'live'; see Fig. 1), height,
113 and community composition (Ruano-Fajardo et al. 2014, Pettitt et al. 2018). Thus, better

114 understanding the ecology of high arboreal pools and characterizing phytotelmata across the
115 vertical gradient could help explain both the apparent success of *D. tinctorius* in a wide range
116 of pools, and why parents sometimes carry their offspring such heights. Further, *D. tinctorius*
117 tadpoles predate on other species of phytotelm-breeders, making it a key amphibian species
118 for understanding the niche partitioning amongst communities of phytotelm breeders. To our
119 knowledge, this is the first study providing detailed biotic, physical, and chemical
120 comparisons of phytotelm choice between Neotropical species.

121



123 **Figure 1. Visual overview of sampled pool diversity.** Photos show the diversity of pools
124 across the vertical gradient. Phytotelmata used by frogs include “live” substrate pools such
125 as tree holes (A), high arboreal pools (B), and buttresses (C). There were also commonly
126 occupied “dead” substrate pools such as fallen trees (D, E) and palm bracts (F).

127

128

129 **Material and methods**

130 The study was carried out in the primary lowland terra-firme forest near the Camp Pararé at
131 the CNRS Nouragues Ecological Research Station in the Nature Reserve Les Nouragues,
132 French Guiana (4°02'N, 52°41'W) over two field seasons: 1st February to 20th March 2019,
133 and 30th January to 26th February 2020. The study area was chosen specifically because of
134 the high abundance of *D. tinctorius* (Rojas and Pašukonis 2019). Pools were found with a
135 combination of field methods. We opportunistically searched for pools targeting suitable
136 microhabitats such as fallen trees, trees with buttresses and palm trees. In addition, pools
137 were discovered by using tracking to follow *D. tinctorius* during previous studies (Pašukonis
138 et al. 2019). We also used experimentally-induced tadpole transport in combination with
139 tracking (Pašukonis et al. 2017) to find additional pools used by *A. femoralis*. Trees with high
140 arboreal pools were discovered by locating calls produced by the treehole-breeding frogs
141 *Trachycephalus resinifictrix* and *T. hadroceps* during night surveys.

142

143 **Sampled pools**

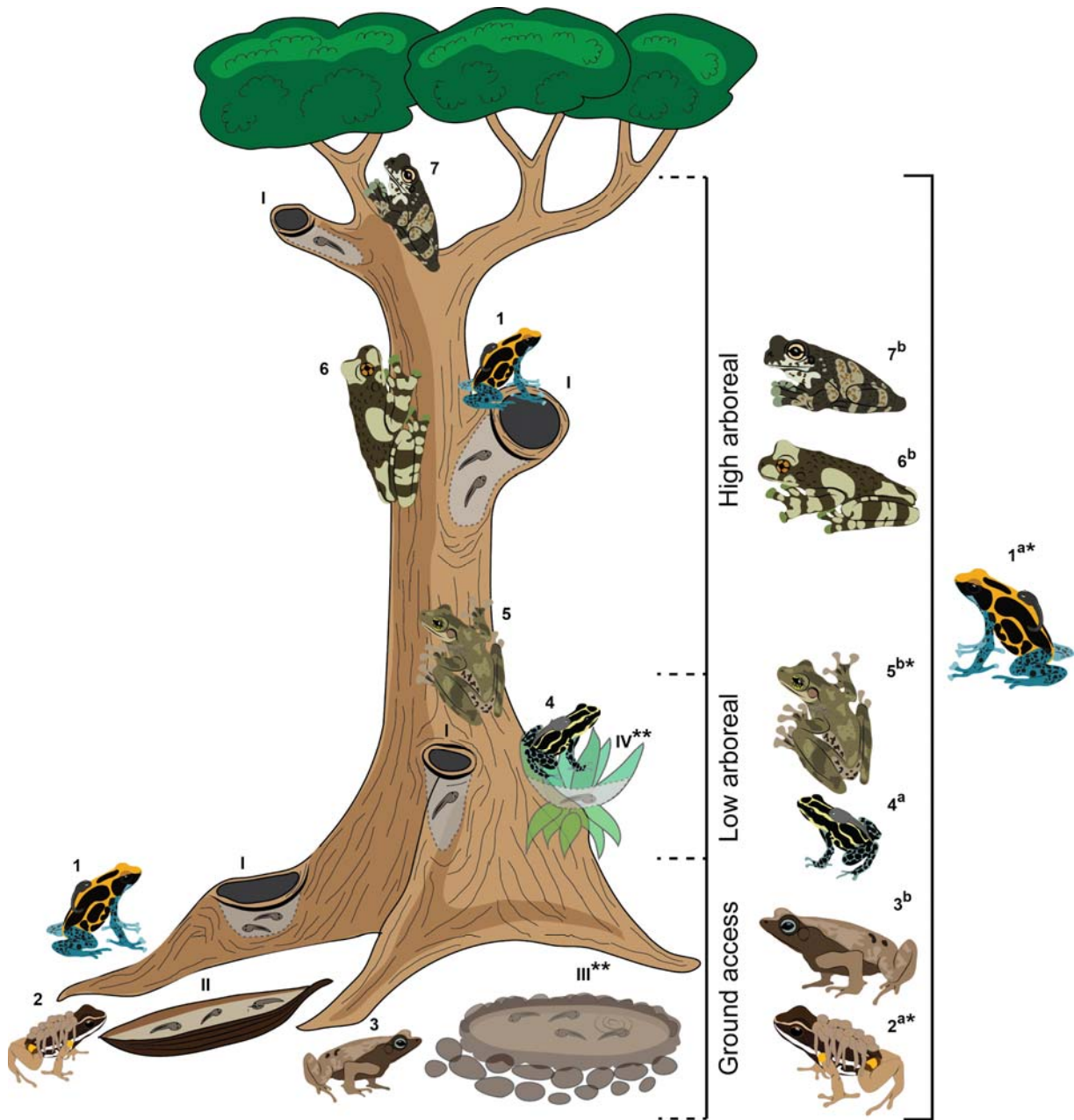
144 We exclusively considered phytotelmata throughout this study. Pools could be classified into
145 two categories: dead substrates, which included holes in dead branches, fallen trees, and
146 fallen *Oenocarpus* palm bracts; or live substrates which included live tree trunks, branches,
147 roots, and buttresses. We did not sample bromeliads and non-phytotelm pools as these pools
148 are not used by *D. tinctorius*. Based on the pools' height and accessibility to different frog
149 species, we termed the pools as “ground access”, “low arboreal” or “high arboreal” (Fig. 1,
150 2). Ground-access pools did not require vertical climbing ability to reach and included dead

151 fallen structures as well as pools in live roots or low buttresses. Low arboreal pools were
152 inside vertical structures low on the trunk or on high buttresses. High arboreal pools were
153 high on the trunk or in canopy branches and were accessed for sampling using rope-based
154 canopy access methods. There was a clear vertical separation between ground-access and
155 low-arboreal pools, which were all under 212 cm in height and between those and high
156 arboreal pools, which were all above seven meters in height. In total, we sampled 84 unique
157 pools across the 2019 and 2020 field seasons.

158

159 Several unique pools were sometimes found and sampled in the same tree. For all pools, we
160 recorded the pool type, location (latitude/longitude), height from the ground to the pool edge,
161 largest width and length parallel to the water surface, and the pool depth from the solid
162 sediment bottom to the maximum water level line. Based on these measurements, we
163 estimated the maximum water-holding capacity of each pool using the volume formula of a
164 semi-ellipsoid as in Rojas (2014). Other sampling methods differed between the two field
165 seasons.

166



167

168 *Figure 2. Vertical partitioning of the phytotelm-breeding anuran community in the*

169 *Nouragues Nature Reserve, French Guiana. Numbers indicate seven species detected*

170 *during this study: 1. *Dendrobates tinctorius*, 2. *Allobates femoralis* 3. *Rhinella castaneotica*,*

171 *4. *Ranitomeya amazonica*, 5. *Osteocephalus oophagus*, 6. *Trachycephalus resinifictrix*, 7. *T.**

172 *hadroceps. Letters indicate species with a) terrestrial or b) aquatic eggs. Latin numerals*

173 *indicate commonly used pool types: I. tree holes at various heights, II. fallen palm bracts, III.*

174 *ground puddles, IV. bromeliads. *Most commonly encountered species included in the*
175 *analysis; ** Pool types not sampled in this study (see methods).*

176

177 **2019 field season sampling**

178 In 2019, we quantified physical measures (height, pool dimensions, leaf-litter volume), biotic
179 measures (amphibian and invertebrate counts and diversity), as well as chemical
180 measurements (see Supp. Table 1 for description of all variables measured). For pools
181 accessible from the ground and smaller arboreal pools, we attempted to sample all tadpoles
182 and Odonata larvae (primary tadpole predators) in each pool. Initially, we carefully observed
183 the undisturbed pool and attempted to catch all tadpoles and Odonata larvae using a variety
184 of tools. We then syphoned the entire volume of the water and sediment from the pool,
185 emptied the leaf litter and searched for tadpoles and Odonata larvae. The volumes of water,
186 sediment, and leaf litter were measured. For deep arboreal pools, we repeatedly netted and
187 observed the pool until no more tadpoles were caught during five minutes of continuous
188 netting. We carefully scraped the inner walls of the pools and caught as many Odonata larvae
189 as possible. The leaf-litter volume could not be accurately measured for some deep arboreal
190 pools, but they typically were protected from falling leaves and had little leaf litter in them.

191

192 We used visually apparent morphological traits to identify tadpoles, except for *Allobates*
193 *femoralis*, *A. granti*, and *Ameerega hahneli*, which we could not reliably differentiate in the
194 field. Because *Allobates femoralis* was more common in our study area than *A. granti* and
195 *Am. hahneli* and we never observed *A. granti* and *Am. hahneli* directly at the pools we
196 classified all *A. femoralis*-like tadpoles as such. Is it important to note that some *A. granti*

197 and *Am. hahneli* tadpoles may have been misclassified as *A. femoralis*. However, this does
198 not affect the interpretation of our results as all three species are cryptic terrestrial poison
199 frogs similar in appearance, ecology and behavior. We also opportunistically recorded all
200 species of adult frogs heard or seen at each pool throughout the sampling period. This was
201 used as an amphibian diversity index between 0 and 8 species observed at each pool.
202 Tadpoles of only three out of seven recorded species, namely *D. tinctorius*, *O. oophagus* and
203 *A. femoralis*, were detected in pools with sufficient frequency for further analysis (N = 34
204 (2019), N = 7 and N = 10 pools, respectively).

205

206 Sampled invertebrates were counted, photographed, and classified only to a group level
207 (usually order or class) apparent in the field. To estimate the predation pressure on tadpoles
208 we used the total count and average size of all Odonata larvae detected in the procedure
209 described above. To estimate density and diversity of aquatic invertebrates, we carefully
210 searched and counted invertebrates in a sample of up to 10 liters of water and up to one liter
211 of sediment in proportion to the total estimated pool volume. For each liter of the water
212 volume sampled, we sampled ~ 100 mL of sediment from the bottom of the pool. When the
213 water volume was less than one liter or the amount of sediment was less than 100 mL, we
214 sampled the entire pool and recorded the exact volumes. In the final analysis, we used the
215 invertebrate density (count divided by the volume sampled) and the diversity index
216 corresponding to our classification (between 0 and 12). The following 12 categories were
217 used to quantify invertebrate diversity: Odonata Zygoptera larvae, Odonata Anisoptera
218 larvae, surface Coleoptera adults, diving Coleoptera adults, Coleoptera Scirtidae larvae,
219 Trichoptera larvae, Diptera Culicidae larvae, Diptera Chironomidae larvae, Diptera Tipulidae

220 larvae, other Diptera larvae, small red Annelida, other unidentified larvae. All water,
221 sediment, tadpoles and invertebrates were released back into the pool after sampling.

222

223 We measured water conductivity, salinity and total dissolved solids (TDS), dissolved oxygen
224 and temperature with electronic sensors (EZDO 7200 and pHenomenal OX4110H). Water
225 chemistry (KH (also known as alkalinity), hardness and NO₃) was recorded using aquarium
226 water testing strips (JBL EasyTest). All measures were taken from the undisturbed surface
227 water of the pool.

228

229 **2020 field season sampling**

230 The 2020 dataset focused solely on *D. tinctorius* tadpole counts and pH measurements of
231 weekly resampled ground access phytotelmata (N = 26) over the time period of a month
232 (February 2020). Rainfall data were provided by the Nouragues Ecological Research Station
233 from an above-canopy weather station in the study area. High arboreal pools (N = 8 (2020))
234 were only measured once. pH was recorded using a pH meter (AMTAST Waterproof pH
235 Meter). The pH meter was calibrated once per day, prior to pool sampling, using both acidic
236 (pH = 4) and neutral (pH = 7) calibration solutions. The pH of ground access pools was taken
237 directly by submerging the pH probe into the pool, and the measurement was recorded once
238 read-out stabilized. For arboreal pools, a sample of water was collected using a syringe,
239 which was then sealed at both ends. Once on the ground, one end of the syringe was opened,
240 and the pH was measured by submerging the pH probe into the syringe. Syringes were never
241 reused. Between pool sampling, the pH probe was wiped with a clean cloth and rinsed with
242 aquifer water.

243

244 **Statistical Analyses**

245 The presence of *D. tinctorius* in pools was analyzed using 2019 field data. As a result of the
246 high collinearity between variables in the 2019 dataset (see Supp Fig 1), we used a principal
247 components regression to analyze phytotelm ecology data. We first checked data for a non-
248 random structure following Björklund (Björklund 2019); then, we established that the
249 correlation matrices were significantly different from random ($\Psi = 10.22, p = 0; \phi = 0.238, p$
250 < 0.001) to ensure they were suitable for a PCA. Based on each PC's difference from random
251 matrices, we selected the first three principal components as predictors of probability for *D.*
252 *tinctorius* tadpole presence as a binomial response in the principal component regression
253 (PC1-3 explained about 53% of the variation proportion of variance explained \pm SE : PC1 =
254 $0.24 \pm 0.48, PC2 = 0.17 \pm 0.40, PC3 = 0.11 \pm 0.33$). We evaluated the fit of negative
255 binomial GLM models based on AIC ranks ((Akaike 1974); see Supp. Table 2). Models
256 within two AIC scores of each other were further evaluated by assessing the significance of
257 interactions between model terms.

258

259 To better understand which variables contributed significantly to each principal component,
260 we calculated which variables had index loadings larger than random data. Following the
261 methods outlined by Björklund (2019) and Vieira (2012), we randomized the data and
262 calculated new correlation matrices which we permuted 1,000 times. We then compared the
263 index of loadings ($IL_{ij} = u_{ij}^2 \times \lambda_i^2$, see Vieira (2012) for details) between each observed PC
264 and the randomly generated data to see which variables contributed significantly to each
265 principal component.

266

267 The 2020 dataset consisted of weekly resampled pools throughout February 2020. Models
268 took repeated measures of pool ID into account as a random effect. Both the presence of *D.*
269 *tinctorius* tadpoles (count; negative binomial family) and pH (Gaussian family) from
270 resampled pools were modeled using a mixed effects generalized linear model in the package
271 “glmmTMB” (Magnusson et al. 2020). Predictor structure for both pH and *D. tinctorius*
272 models were built based on biologically relevant variables (pool substrate, time, *D. tinctorius*
273 tadpole count (for pH model), water capacity, surface area:depth ratio). Using these
274 variables, models were first fit with relevant interactions (see Supp. Table 3, 4), which were
275 then removed if they did not contribute significantly to the model using single term deletions
276 (drop1). Residuals were diagnosed using the package “DHARMA” (Hartig 2020); final
277 models were then checked for overdispersion and zero-inflation and corrected as necessary.
278 All code was done in R (Core 2015).

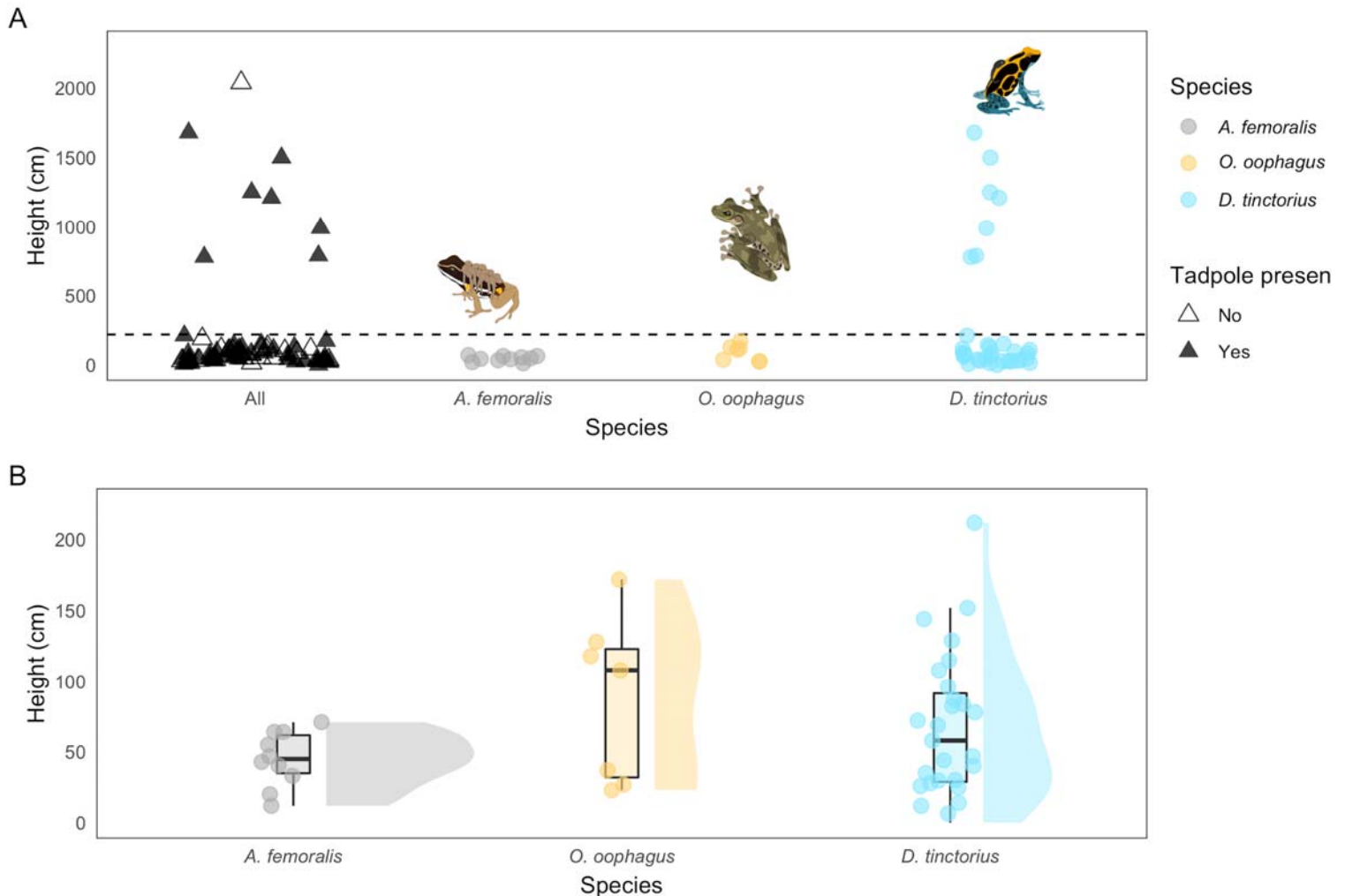
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280 **Results**

281 **Species-wide trends**

282 We detected tadpoles or adults of 7 species from 4 families using surveyed phytotelmata for
283 breeding (Fig. 2). Dendrobatidae: *Dendrobates tinctorius*, *Ranitomeya amazonica*,
284 Aromobatidae: *Allobates femoralis*; Hylidae: *Osteocephalus oophagus*, *Trachycephalus*
285 *resinifictrix*, *T. hardroceps*; Bufonidae: *Rhinella castaneotica*. The tadpoles of only three
286 species (*D. tinctorius*, *O. oophagus* and *A. femoralis*, present in N = 34, N = 7 and N= 10
287 pools, respectively) were detected frequently enough for further analysis. Although the data
288 on *D. tinctorius* are more robust, trends for *O. oophagus* (N = 125 tadpoles) and *A. femoralis*

289 (N = 117 tadpoles) emerge despite a smaller dataset. The species-wide dataset is based on the
290 sampling of 70 unique pools in 2019.



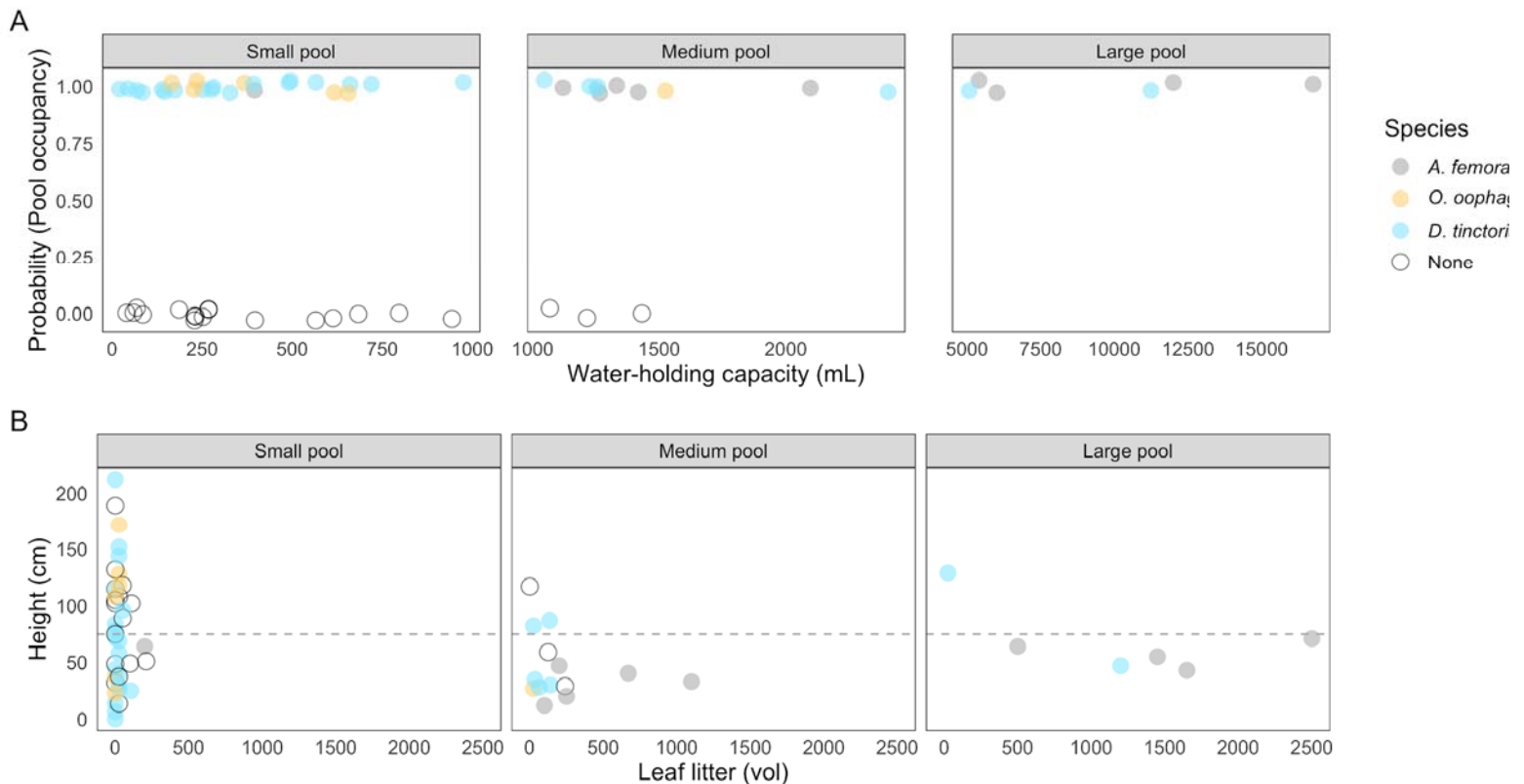
292 **Figure 3. Tadpole presence across the vertical landscape.** Panel (A) shows all sampled
293 pools. For the “All” category, colored/empty triangles represent presence/absence data of all
294 three species. Dashed line is drawn at 220cm; pools above this limit are classified as high
295 arboreal pools. Panel (B) highlights occupied pools below 220 cm. *Dendrobates tinctorius* (N
296 = 34) tadpoles occur in pools across the vertical landscape. Distribution of *O. oophagus* (N =
297 7) and *A. femoralis* (N = 10) tadpoles shows possible vertical niche partitioning. Boxplot

298 *whiskers extend 1.5 * interquartile range. Violin plots represent density distribution for*
299 *species occurrence. Data is from the 2019 field season.*

300

301 Differences in pool accessibility are highlighted by Fig. 3. Compared to *A. femoralis* and *O.*
302 *oophagus*, one of the most striking aspects of *D. tinctorius* ecology is its flexibility with
303 respect to site choice on a vertical axis. *Dendrobates tinctorius* tadpoles were found in pools
304 from the forest floor to more than 15 meters in the canopy. For *O. oophagus*, a strictly
305 arboreal frog in its adult stage, tadpoles were detected only in low arboreal pools, ranging
306 from 20 cm to 1.7 m in height. In *A. femoralis*, tadpoles were only found in ground-access
307 pools, and occurred at a maximum height of 71 cm.

308

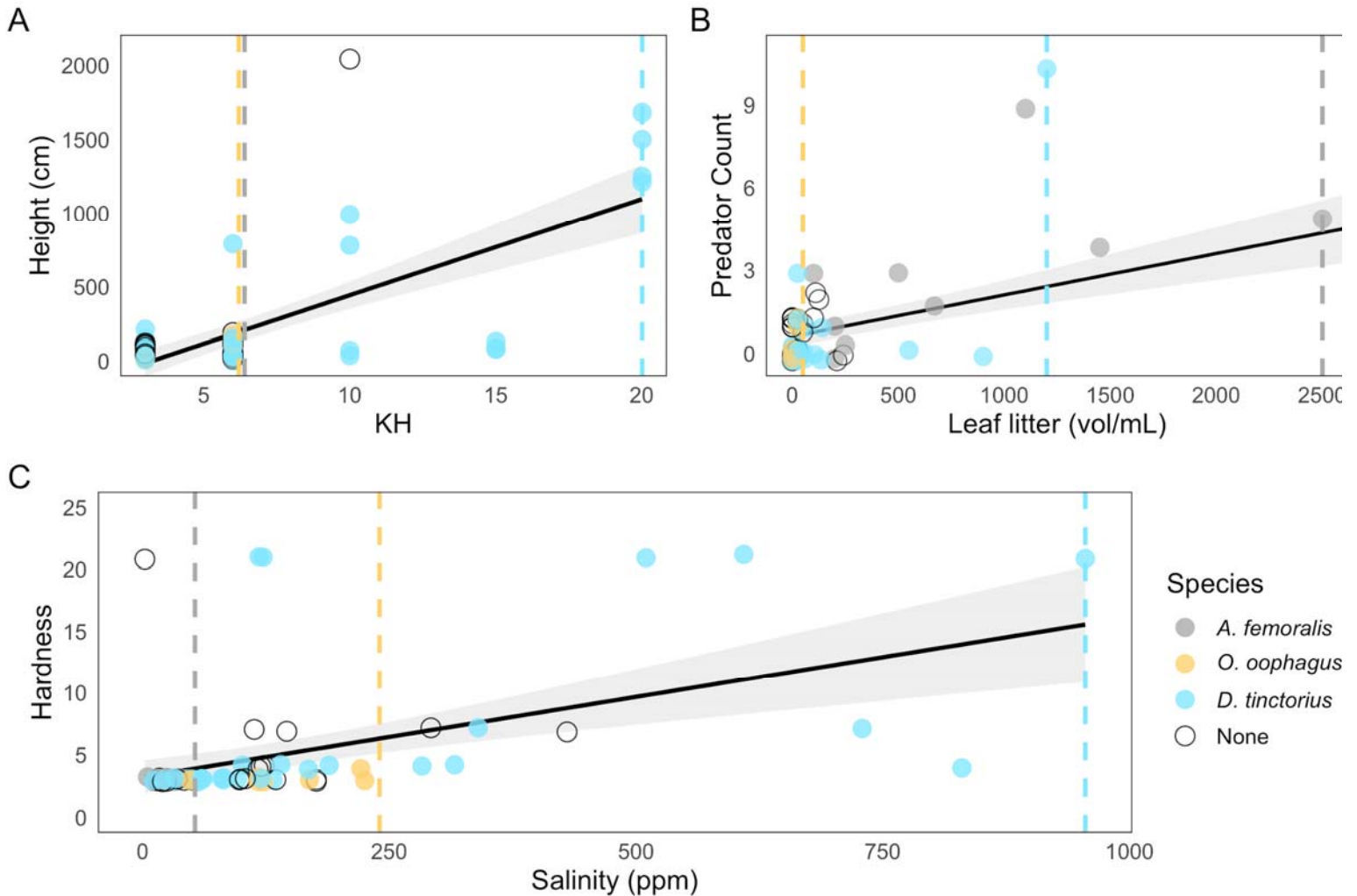


310 **Figure 4. Pool occupancy based on water capacity and leaf-litter volume of phytotelmata.**
311 All data are subsetted for low arboreal and ground-access pools (< 220 cm). Panel (A) is the
312 probability of pool occupancy (binomial, 0/1) based on water capacity; data are faceted
313 based on relative pool size (Small = < 1000 mL, Medium = < 5000 mL, and Large = > 5000
314 mL). Points are plotted with a small amount of random noise on the y-axis to facilitate
315 visualization of overlapping data. Panel (B) illustrates the correlation between leaf litter and
316 height, faceted by the same pool categories as Panel A. Points are colored by species
317 presence. Dashed line indicates the vertical limit of *A. femoralis* (< 75 cm). Out of the 62
318 ground-access and low arboreal pools observed, *D. tinctorius* co-occurred once with *A.*
319 *femoralis* and once with *O. oophagus*; *O. oophagus* and *A. femoralis* tadpoles were never
320 found in the same pool.

321

322 Despite small sample sizes we found clear trends: *O. oophagus* tadpoles are heavily biased
323 towards small, clear pools and *A. femoralis* is present in medium and large pools with large
324 amounts of leaf litter, whereas *D. tinctorius* occurs throughout the sampled range (Fig. 4).

325



327 **Figure 5. Chemical and physical predictors of tadpole presence in Neotropical tadpoles.**

328 We find that *D. tinctorius* tadpoles are tolerant to a wide range of KH, hardness, height, and
329 salinity values, but appear to be limited with respect to high leaf litter volumes. Plots are
330 based on variables with varied species limits (dashed lines). Colored points represent species
331 presence. Black lines are fit with a GLM smoother, with 95% CI highlighted in light grey.

332

333 As opposed to *A. femoralis* and *O. oophagus*, *D. tinctorius* can occupy chemically diverse
334 pools, showing remarkable flexibility with respect to KH, salinity, and hardness that clearly
335 limits the other species. *Allobates femoralis* and *O. oophagus* appear to exist in similar KH

336 ranges (KH = 3-6), while *D. tinctorius* appears more tolerant of extreme values (KH = 3-20).
337 *Allobates femoralis* tadpoles occurred in pools with a salinity range from 5 to 37 ppm, while
338 *O. oophagus* tadpoles occupied pools with a range from 48 to 225 ppm (Fig. 5, Panel C).
339 *Dendrobates tinctorius* again appears to have no functional limitation, occupying pools with
340 salinity from 11 ppm up to 955 ppm.

341

342 **Deposition site decisions: *Dendrobates tinctorius***

343 Because we detected *D. tinctorius* tadpoles much more frequently, we were able to conduct a
344 more thorough analysis of the variables predicting tadpole presence in this species (see Supp.
345 Table 1). We used principal components as predictors for *D. tinctorius* presence. Based on an
346 AIC model comparison, we did not detect any significant interactions between components
347 (Supp. Table 2). A negative binomial GLM only detected PC1 to play a significant role in
348 predicting tadpole presence (Table 1, CI: 0.08-0.42, $p = 0.003$), where an increase in
349 component value increased the probability of detecting tadpoles.

350

Tadpole presence (Y/N)			
<i>Predictors</i>	<i>Estimates</i>	<i>CI</i>	<i>p</i>
(Intercept)	-0.85	-1.27 – -0.49	<0.001
PC1	0.25	0.08 – 0.42	0.003
PC2	0.09	-0.16 – 0.40	0.528
PC3	0.21	-0.07 – 0.48	0.144

351

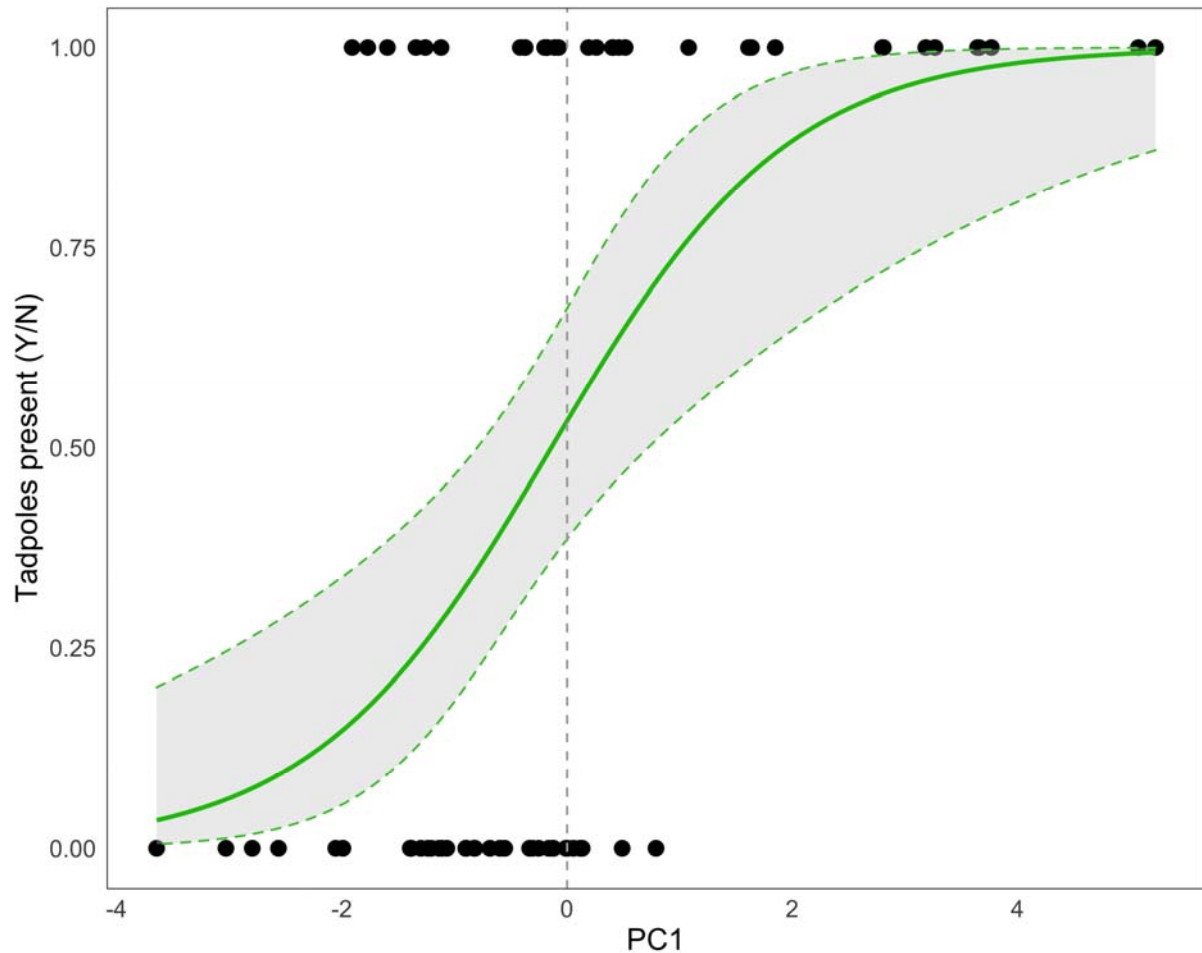
352 *Table 1. Principal component regression of tadpole presence in phytotelm pools. Using a*
353 *negative binomial GLM, we found that only the first principal component is significant in*
354 *predicting tadpole presence.*

355

356 Following Björklund (2019), we found that, when compared to randomly generated matrices,
357 five out of the original 14 traits (see Supp. Table 1 for trait definitions) contributed
358 significantly to the first principal component. The significant traits can be broadly
359 categorized using three descriptors: (1) chemical (KH, $p < 0.001$; IL = 1.50, hardness, $p =$
360 0.001 , IL = 1.30; salinity, $p < 0.001$, IL = 1.62); (2) physical (height, $p = 0.013$, IL = 1.06);
361 and (3) biological (invertebrate diversity, $p < 0.001$, IL = 1.20) (see Fig. 6). Altogether, these
362 results show that *D. tinctorius* tadpoles were found significantly more frequently in pools
363 with higher levels of hardness, KH, and salinity; higher in the vertical gradient; and with
364 more diverse invertebrate communities (Fig. 6).

365

366



367

368 **Figure 6. Binomial response of *Dendrobates tinctorius* presence to the first principal**
369 **component.** Dashed lines represent the 95% confidence interval. Significant variables within
370 this component are detailed in Supp. Table 1.

371

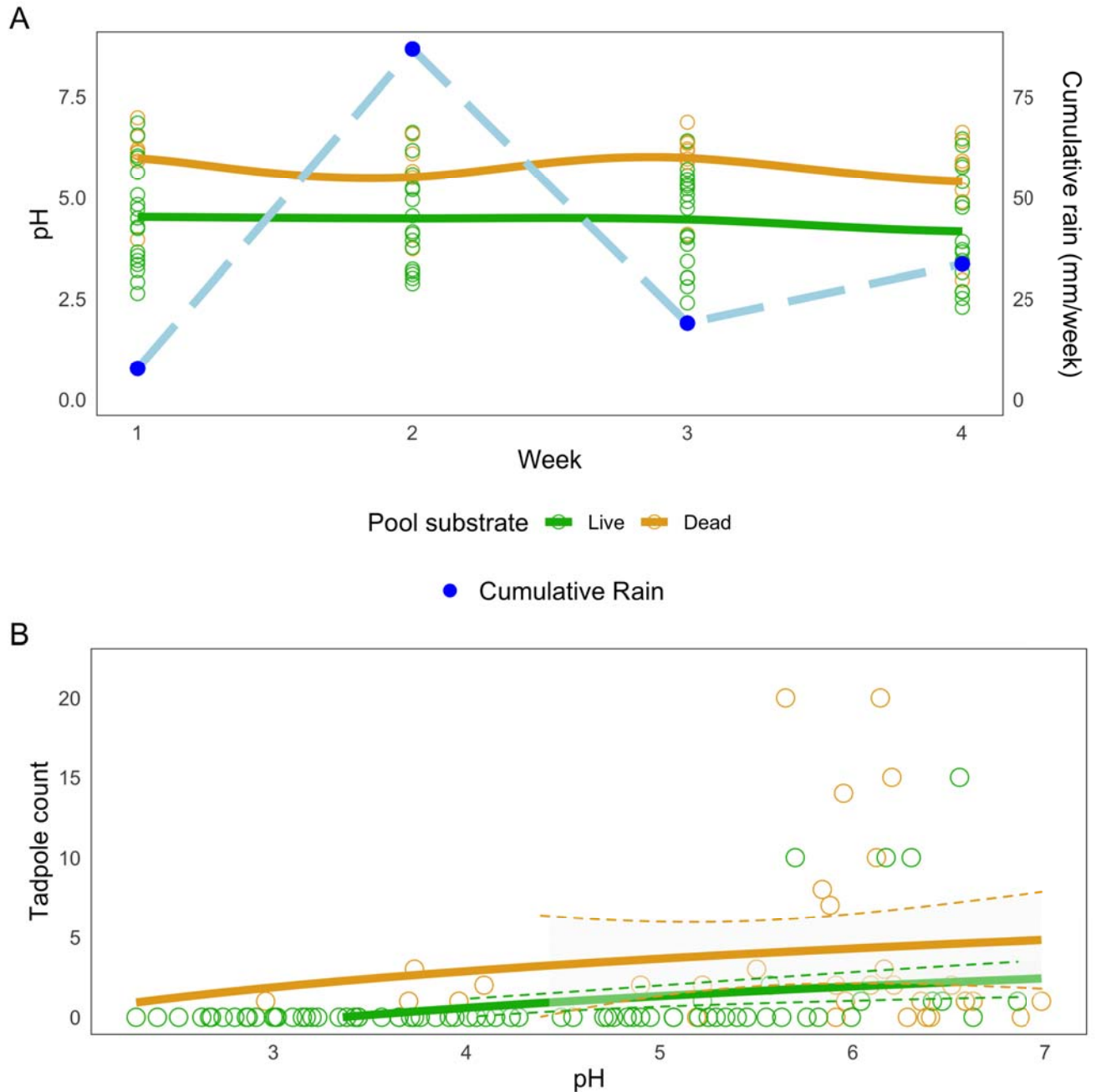
372 *Dendrobates tinctorius* across temporal scales

373 Using both 2019 and 2020 datasets, we were able to follow phytotelmata across multiple
374 timescales: 13 weekly resampled ground-access and low arboreal pools, 13 annually
375 resampled ground-access and low arboreal pools, and 7 annually resampled high arboreal
376 pools. Overall, we found that pools can persist over multiple sampling seasons. High arboreal
377 pools appear to be the most stable with respect to both tadpole count and tadpole density

378 compared to low arboreal and ground-access pools sampled both years (\bar{x}_{High} (2019) = 13.14
379 tadpoles, \bar{x}_{High} (2020) = 10 tadpoles versus \bar{x}_{Low} (2019) = 0.92 tadpoles, \bar{x}_{Low} (2020) = 1.31
380 tadpoles). High arboreal pools also had the highest average pH and KH (pH_{High} = 6.73,
381 KH_{High} = 15.14) compared to averages of other pool substrates ($\text{pH}_{(\text{Low})\text{Live}}$ = 4.35, $\text{pH}_{(\text{Low})\text{Dead}}$
382 = 5.68; $\text{KH}_{(\text{Low})\text{Live}}$ = 5.69, $\text{KH}_{(\text{Low})\text{Dead}}$ = 5.88). Due to difficult accessibility, high arboreal
383 pools were sampled only once per year, and thus were excluded from further analysis
384 involving repeated sampling.

385

386 When considering pools resampled weekly over the course of a month, two trends emerge:
387 (1) pH is consistently higher in pools contained in ‘dead’ substrates than in ‘live’ substrates
388 (*Odds Ratio*= 1.47, *Table 2*). For all substrate types, however, pH values decreased over the
389 one-month sampling period (Fig. 7A; *Odds Ratio* = -0.11, *Table 2*, which may be related to
390 rainfall levels throughout the month); and (2) the number of *D. tinctorius* tadpoles can be
391 predicted, in part, by the interaction between pool substrate and pH (Fig.7B). Dead pools
392 have higher numbers of *D. tinctorius* tadpoles (\bar{x}_{Live} = 0.72, \bar{x}_{Dead} = 4.03, 2020 data). This
393 pH/substrate relationship is clearly important, as tadpoles occur in higher numbers in high pH
394 pools. Time (in weeks) was not detected as an important variable in determining *D. tinctorius*
395 tadpole numbers throughout the month.



396

397 **Figure 7. Phytotelm stability and tadpole presence across a month. Panels (A, B) are**

398 *subsetted to exclude high arboreal pools, and emphasize how pH is related to (A) pool*

399 *substrate and (B) predicting *D. tinctorius* tadpole presence in resampled pools. Phytotelmata*

400 *made of dead substrate have higher pH values than live substrate ($N_{Dead} = 8$, $N_{Live} = 18$;*

401 *each resampled four times) (Table 2); more tadpoles are found in pools with a higher pH*

402 (Table 3). Dashed lines represent 95% CI. Blue points in (A) indicate the weekly
403 accumulation of rainfall (February 2020). Tan lines are dead substrates and light green lines
404 are live substrates; lines in A are plotted with a “LOESS” fit and B with a “GLM” fit.
405

<i>Predictors</i>	<i>Estimates</i>	pH		
		<i>CI</i>	<i>z</i>	<i>p</i>
(Intercept)	4.66	4.13 – 5.20	17.09	<0.001
Pool type [Dead]	1.47	0.56 – 2.37	3.17	0.002
Week	-0.11	-0.18 – -0.04	-2.93	0.003
Random Effects				
σ^2	0.20			
τ_{00} pool_id	1.13			

406
407 Table 2. Negative binomial generalized mixed model predicting pH over time based on pool
408 substrate. Data include low arboreal and ground access pools. Pool type is a 2-level
409 categorical variable (“Dead”, “Live”). Where pH is significantly higher in dead pool types
410 versus live pools type, and decreases over time.
411

<i>D. tinctorius</i> tadpoles (count)				
<i>Predictors</i>	<i>Estimate</i>	<i>CI</i>	<i>z</i>	<i>p</i>
(Intercept)	-10.35	-16.65 – -4.05	-3.22	0.001
Pool type [Dead]	11.08	2.96 – 19.21	2.67	0.008
pH	1.59	0.48 – 2.71	2.80	0.005
Week	-0.16	-0.39 – 0.06	-1.40	0.160
Pool type [Dead]: pH	-1.50	-2.93 – -0.07	-2.05	0.040
Random Effects				
σ^2	2.62			
τ_{00} pool_id	1.87			

412

413 *Table 3. Negative binomial generalized mixed model predicting D. tinctorius tadpoles in*
 414 *resampled pools in February 2020.*

415

416 Several pools (N = 5, 2020) dried out completely during the month-long observation period,
 417 three of which were fallen palm bracts. Thus, although dead phytotelmata tend to have higher
 418 pH values, and have a remarkable buffering capacity when filled, they also appear to dry out
 419 more easily during dry periods.

420

421 **Discussion**

422 Juvenile stages of development are particularly deadly for animals with external fertilization.
 423 Thus, the decision of where to breed and raise young is vital to offspring survival. In this
 424 study, we investigated the tadpole rearing-site ecology of Neotropical phytotelm-breeding
 425 frogs. Out of seven detected species, five showed some form of parental care, three out of

426 which were sufficiently common to infer patterns of pool choice. Broadly, we found that the
427 deposition choices of two of the three species are characterized by height and pool size, such
428 that *O. oophagus* tadpoles occur in small, low arboreal pools below ~ 2 m and *A. femoralis*
429 tadpoles occur in medium and large pools below ~ 1 m and with access only from the
430 ground, which is not surprising considering that adults are poor climbers (Roithmair 1992).
431 *Dendrobates tinctorius* tadpoles, in contrast, occur in pools from the ground to the canopy,
432 and of sizes ranging from 19.6 mL to 270L. When focusing on *D. tinctorius* pool choice, we
433 found that despite being able to tolerate an impressive range of physical/chemical factors,
434 tadpoles are more likely to be found in higher pools with greater levels of KH, salinity, and
435 hardness, and higher invertebrate diversity.

436

437 **Inter-specific comparison of rearing-site choices**

438 In line with previous research, we found that the preference for water capacity varies among
439 species, and that some of this variation is explained by parental behavior (Lin et al. 2008,
440 McKeon and Summers 2013, Summers and Tumulty 2014). For example, *O. oophagus*
441 provisions their tadpoles with trophic eggs (Jungfer and Weygoldt 1999), which allows
442 parents to choose very small arboreal phytotelmata with rather clear water and little food
443 (Brown et al. 2008a, b, 2010, Summers and Tumulty 2014). Despite the desiccation risk
444 associated with the notably small pools chosen by *O. oophagus*, their nurseries are predator-
445 free, which is often suggested as the primary factor driving the invasion of phytotelmata
446 (Crump 1974, Magnusson and Hero 1991, Gomez-Mestre et al. 2012, Summers and Tumulty
447 2014).

448

449 In contrast, *Allobates femoralis* does not provision tadpoles and preferably transports them to
450 medium-to-large ground access pools. These pools tend to have both high leaf-litter
451 concentrations and a high number of predators (Fig. 5B; concurrent with McKeon and
452 Summers (2013)). The effect of leaf litter on Neotropical tadpoles is unclear, but large
453 amounts have been found to have both positive effects (increased growth rate in African
454 tadpoles; (Lehtinen 2004)) and negative effects (decreased growth rate and development in
455 temperate-region tadpoles; (Williams et al. 2008)). Because *A. femoralis* are confined to
456 ground access phytotelmata due to their inability to climb, choosing to deposit their tadpoles
457 in pools containing high amounts of leaf litter may be their best option: despite the higher
458 predation risk (which *A. femoralis* fathers try to minimize (Ringler et al. 2018)), leaf litter
459 provides a source of food and shelter/predator protection to tadpoles that do not exist in clear
460 pools. Interestingly, the turbid leaf litter pools occupied by *A. femoralis* were functionally
461 available to *D. tinctorius*, who do not appear to use them. As *D. tinctorius* tadpoles are
462 predatory, clear pools may be better suited for visual foraging.

463

464 Contrary to our expectations, the water capacity of pools was not a key variable in predicting
465 *D. tinctorius* tadpole presence, corroborating Rojas' (2014) findings. While a higher water-
466 holding capacity is expected to decrease desiccation risk, pool volume and depth are not
467 always reliable measures for water-holding persistence in phytotelmata (Rudolf and Rödel
468 2005), making frogs adjust their preference based on other pool traits (see below). The
469 presence of large conspecifics, for instance, may be used as a cue of pool stability and thus
470 influence pool choice by males (Rojas 2014).

471

472 An unexpected variable that segregated all three frog species was pool salinity, which tends to
473 increase with height. Some high arboreal pools were particularly deep and a low turnover of
474 stagnant water could explain high salinity levels, where ions (and similarly, salts) concentrate
475 in pools over a long period of time (Sawidis et al. 2011). These pools were mostly occupied
476 by *D. tinctorius* tadpoles, which appear to withstand salinity conditions of up to 1,000 ppm.
477 *Allobates femoralis* tadpoles, in contrast, were only found in low-salt environments. Low
478 salinity is likely a byproduct of the high amount of leaf litter, which appears to buffer salt
479 concentrations (Roache et al., 2006; see Supp. Fig 3). As microbial activity is limited by high
480 salinities, low-salt pools are ideal for the production of detritus (Roache et al. 2006), the main
481 food source of *A. femoralis* tadpoles. *Osteocephalus oophagus* tadpoles were found in salinity
482 ranges from 48 to 225 ppm (Fig. 5C), and the relatively saline pools (> 700 ppm) that occur
483 within *O. oophagus* vertical ranges were only occupied by *D. tinctorius*. Adult *O. oophagus*
484 deposit their egg clutches in the water as opposed to the other two species, which only use
485 phytotelmata as tadpole-rearing sites. The saltier conditions of pools higher in the canopy may
486 not be suitable for eggs (Christy and Dickman 2002, Albecker and McCoy 2017), which may
487 limit the suitable conditions for oviposition in *O. oophagus*. Poison frogs, in contrast, can
488 escape these limitations because their clutches are terrestrial. Although the small sample size
489 does not allow any stronger interpretation, it appears that both chemical and physical variables
490 are important in shaping ideal pool conditions in *A. femoralis* and *O. oophagus*.

491

492 **Pool choice and flexibility in *Dendrobates tinctorius***

493 Over the two sampling seasons, the bulk of our study focused on the factors that shaped *D.*
494 *tinctorius* presence and persistence. Despite having sampled over 80 unique pools and found

495 350 *D. tinctorius* tadpoles (N = 208, 2019; N = 142, 2020), understanding the critical
496 variables that drive *D. tinctorius* pool choice is difficult because of the wide range of
497 physical and chemical properties in which these tadpoles occur. Further, the interactions
498 between physical, chemical, and biological characteristics in phytotelmata are complex and
499 collinear. Based on both our principal component regression and analysis, we found that the
500 probability of detecting *D. tinctorius* tadpoles increases positively with specific physical
501 (height), chemical (KH, salinity, and hardness), and biological (invertebrate diversity)
502 properties. Interestingly, we found that salinity, hardness, and KH also tend to increase with
503 increasing height (Supp. Fig 4). Overall, these chemical components tend to vary in the same
504 direction when moving up the vertical axis, suggesting a positive relationship between these
505 chemical and physical traits. Invertebrate diversity of occupied pools, in contrast, tends to
506 stay relatively constant across heights and might serve as an important food source for
507 predatory *D. tinctorius* tadpoles.

508

509 In this study, we found that KH increases with height, and pools with high KH are more
510 likely to have tadpoles in them. KH is a measure of a solution's buffering capacity or, in other
511 words, a solution's resistance to pH changes (Yang et al. 2008). KH values in low arboreal
512 and ground-access pools usually ranged from 3 to 6 KH, while average KH in high canopy
513 pools was 15. Interestingly, two of the five lower pools with a KH above 8 had *D. tinctorius*
514 counts of over 10 tadpoles, demonstrating that, when these conditions are available
515 terrestrially, *D. tinctorius* tadpoles take advantage of them. The apparent preference for high
516 alkalinity environments is interesting, as work studying the formation of fungal granules has
517 established that high-alkalinity conditions inhibit fungal growth (Yang et al. 2008). The

518 potential relationship of KH limiting the growth of fungi in phytotelm conditions is
519 noteworthy as amphibian fungal pathogens such as *Batrachochytrium dendrobatidis* (*Bd*)
520 spread aquatically (Rosenblum et al. 2010), and *Bd* presence (prevalence of ~5%) has been
521 reported for *D. tinctorius* in our study area (Courtois et al. 2015). Thus, the consistent
522 detection of *D. tinctorius* tadpoles in high KH pools could indicate that fathers are selecting
523 environments less prone to fungal contamination. Although we are unsure of the proximate
524 mechanisms driving the fathers' choice of particular chemical conditions in phytotelmata, we
525 establish here that KH, hardness, and salinity play an important role in shaping *D. tinctorius*
526 pool choice, and suspect that these chemical conditions may be linked to the long turn-over
527 time of high arboreal pools.

528

529 **The stability of ephemeral pools**

530 In 2020, we were able to follow a subset of low arboreal and ground access pools over a
531 month, recording the pH and *D. tinctorius* tadpole presence on a weekly basis. We found that
532 pools made of dead substrate (fallen palm bracts, dead trees) had a higher pH than live
533 substrates (tree holes). The gross average pH of dead phytotelmata across our sample was
534 5.68, which is almost exactly the value of unpolluted rainwater (pH = 5.65 when saturated
535 with atmospheric CO₂; (Koshy et al. 1997)). In contrast to most live substrates (average pH =
536 4.35), dead phytotelmata are usually in canopy gaps, where rain falls directly into the pools.
537 When reported, the pH of most phytotelmata is acidic (Kitching 2001, von May et al. 2009,
538 Poelman et al. 2013, Ramos et al. 2017); but see Lehtinen 2004, which shows that *Pandus*
539 bromeliads were close to neutral pH). However, most studies on phytotelmata are biased

540 towards living plants and trees, and assess chemical/biological variables of pools at a single
541 time point.

542

543 Throughout the month, all pool types decreased in pH (Table 2); a similar trend was also
544 found in bamboo phytotelmata in Peru (von May et al. 2009), suggesting a time-dependent
545 process causing pools to become increasingly acidic over time. Remarkably, some pools
546 dried out multiple times during our sampling period, and when refilled by rain were
547 approximately at the same pH as before the drying event (ex. palm bract originally pH 6.98
548 (Week 1), dried out (Week 2), refilled pH 6.87 (Week 3); live tree hole pH 2.91 (Week 1),
549 dried out (Week 2), pH 3.02 (Week 3)). This indicates that pool substrate may play an
550 important role in establishing pool pH. Three out of the five pools that dried out were dead
551 palm bracts, suggesting that this pool type, despite having favourable chemical conditions
552 when filled, may also be at a higher risk for desiccation and decomposition.

553

554 Surprisingly, pools in dead substrates, such as palm bracts and fallen trees, contained more
555 tadpoles than other pool types despite drying out more regularly across our month survey
556 (Table 3). Such pools tend to occur in forest gaps, which makes them more prone to
557 dessication. However, pools in these lit areas may also have more microbial activity and less
558 food limitation (Kitching 2001, Rudolf and Rödel 2005), making them attractive deposition
559 sites for tadpoles. Suitable pools are a limiting resource for frogs and other animals
560 (Donnelly 1989a, b, Fincke 1992, Poelman and Dicke 2007, Ringler et al. 2015) and new
561 pools for *D. tinctorius*, such as those in tree-fall gaps, appear unpredictably and are rapidly
562 occupied despite the high rates of competition and cannibalism (Rojas 2015). This strategy

563 can be particularly beneficial when parents arrive early to new pools, as it allows their
564 offspring to be predators rather than prey. Thus, the competition to be the first to deposit
565 tadpoles might make pools in dead substrates that occasionally dry out additionally attractive.

566

567 Interestingly, the size range of tadpoles in dead substrates is much more variable than in low
568 and high arboreal pools (CF, BR, AP pers. observ.), suggesting that the pools remain
569 attractive even when already occupied by larger cannibals. This pattern corroborates the
570 experimental evidence that *D. tinctorius* preferably deposit newly-hatched tadpoles in pools
571 already occupied by conspecifics (Rojas 2014, 2015); possibly, tadpole presence serves as an
572 indicator of pool stability, which might be more valuable to a father's deposition choice than
573 the risk of having his offspring cannibalized by conspecifics.

574

575 **High arboreal pools**

576 While most of our work focused on low arboreal and ground access pools, this study
577 provides one of the first thorough characterisations of high arboreal phytotelmata in the
578 Amazon. Gaucher (2002) unexpectedly found *D. tinctorius* tadpoles in canopy pools up to
579 25 meters high while studying the treefrog *Trachycephalus hadrocephalus*. Other poison frogs,
580 such as *D. auratus* have been reported to use arboreal tree holes as well (Summers 1990). We
581 found large numbers of tadpoles in arboreal pools up to 20 m in height, which suggests some
582 benefit of these pools given the presumed high energetic expense that fathers invest in
583 transporting their tadpoles from terrestrial oviposition sites.

584

585 One key advantage of high arboreal phytotelmata may be a regular food source provided by
586 *Trachycephalus* treefrogs that specialize in these pools. During this study, all of the suitable
587 high arboreal pools were found by locating nocturnal calls of *T. resinifictrix* and *T.*
588 *hadroceps*, indicating that these habitats were potentially used as breeding sites. Although the
589 breeding frequency of these treefrog species is sporadic (Gaucher 2002), successful breeding
590 events result in clutches that consist of hundreds to thousands of eggs and tadpoles, which *D.*
591 *tinctorius* tadpoles readily consume ((Gaucher 2002), personal obs. AP and BR). As
592 proposed by Gaucher (2002), it could be that *D. tinctorius* fathers cue on *Trachycephalus*
593 calls for locating high arboreal pools, but this warrants further investigation.

594

595 Unexpectedly, we also found that dead substrate pools share some characteristics with high
596 arboreal pools, particularly with respect to chemical qualities (a more basic pH), tadpole
597 abundance, and being a limited or hard-to-access resource. As such, both pool types offer
598 benefits that fathers may value: despite having a shorter life, novel pools (such as fallen palm
599 bracts and holes in fallen trees) are worth invading as deposition sites because they are easy
600 to access and have a high probability of having food, and a suitable chemical profile; high
601 arboreal pools, on the other hand, may have sporadic food and are hard to access, but they are
602 stable and less prone to chemical fluctuations. Together, these different pools are both
603 worthy deposition sites, as they provide different stable environments which create a range of
604 possible offspring success.

605

606 **Conclusions**

607 When comparing the occurrence of tadpole species in pools, one of the first trends that
608 emerges is the presence/absence on the basis of specific phytotelm characteristics. For
609 example, *A. femoralis* and *O. oophagus* vertical ranges technically overlap, yet tadpoles
610 never co-occur. In species that demonstrate a distinct morphological limitation or vertical
611 preference, it may be that tadpoles occur in pools because that is what is available to their
612 parents. These constraints play a role in the environment tadpoles are exposed to, and should
613 affect their physiology and behavior. But what about when parents are completely
614 unconstrained? *Dendrobates tinctorius* occur across the vertical gradient and occupy pools
615 that range from acidic (pH = 2.96) to neutral pH, with volumes from 19 mL to over 270L and
616 in pools that range from fresh to slightly saline (~1000 ppm), which hints at a remarkable
617 physiological flexibility that has been overlooked thus far. Therefore, physiological studies
618 comparing phytotelm-breeding tadpoles would be especially interesting to better understand
619 parental decisions. It is also warranted to measure *D. tinctorius* growth in pools with different
620 chemical compositions to see if (despite surviving) these tadpoles are paying a cost for the
621 deposition choices by their fathers.

622

623 In sum, natural history studies allow us to grasp species' flexibility; this is becoming
624 increasingly relevant when we consider the effects of climate change in the Amazon.
625 Forecasted changes in precipitation (Cochrane and Barber 2009, Silvestrini et al. 2011) will
626 have direct effects on animals dependent on plant-based water bodies, and surely the
627 chemical composition of the water bodies themselves. Further studies examining larval
628 anuran responses to challenging environmental conditions (especially KH and pH) are

629 necessary to better understand the potential resistance and adaptability of phytotelm-
630 dependent frogs and how this may shape species resilience in the future.

631

632 **Significance statement**

633 This two-year field project is a reflection of over a decade of natural history observation and
634 experiments in the Neotropics. In this study, we expand our knowledge of the ecology of
635 phytotelm-dependent frogs, this time focusing on a comparative overview between larval
636 species and the precise ecological factors that shape the microhabitats in which eggs and
637 larvae are deposited. The breadth of this work will surely engage a wide variety of readers
638 who are interested in ecological dynamics in the tropics.

639 Here, we synthesize the occurrence and interactions between three different amphibian
640 species, which contributes to understanding the dynamics of each species independently, but
641 this work is also a perspective into the interactions between larvae and parents within the
642 context of an ecological study. Species in this study are specifically plant-dependent as
643 tadpoles, which provides a unique opportunity to thoroughly survey and measure the
644 microhabitats in which they occur. Together, our findings show how biological, physical, and
645 chemical components interact in predicting larval presence in species with parental care,
646 which is a strategy present in ten percent of amphibians. We also provide the first detailed
647 account of the ecology of high arboreal breeding pools. Accessing trees more than 20 meters
648 in height is challenging, and conducting a thorough survey of these environments framed in
649 comparison to microhabitats across the vertical gradient is a testament to the scope of this

650 work. While our study is based on amphibians, it more broadly focuses on the ecology that
651 shapes larval deposition sites and the wide array of species' flexibility we observe in the
652 Neotropics.

653 Hopefully, a wide variety of researchers will be excited to learn more about the diversity of
654 microhabitats in the Amazon and a cross-species comparison of the amphibians that depend
655 on them.

656 **Data availability statement** All data will be archived in the data repository of the University
657 of Jyväskylä.

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681 *Ethics statement* — The study was approved by the scientific committee of the Nouragues
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683 the Nouragues Nature Reserve (No 01-2019) . We strictly adhered to the current French and
684 European Union law, and followed the Association for the Study of Animal Behaviour’s
685 (ASAB) Guidelines for the use of live animals in teaching and research (ASAB, 2017).

686

687 **Author contributions**

688 **Chloe Fouilloux:** Writing – original draft (lead); Investigation (equal); Data curation (equal);
689 Formal analysis (lead); **Shirley Jennifer Serrano-Rojas:** Investigation (equal); Data curation
690 (equal); Formal analysis (supporting); Writing - review and editing (supporting); **Juan David**
691 **Carvajal-Castro:** Investigation (equal); Writing - review and editing (supporting); **Marie-**
692 **Therese Fischer:** Investigation (supporting); Writing - review and editing (supporting);
693 **Janne Valkonen:** Methodology (supporting); Investigation (supporting); Writing - review
694 and editing (supporting); **Philippe Gaucher:** Investigation (supporting); **Andrius Pašukonis:**
695 Conceptualization (equal); Methodology (equal); Investigation (equal); Writing – review and
696 editing (equal); Supervision (equal); Funding acquisition (equal); **Bibiana Rojas:**
697 Conceptualization (equal); Methodology (equal); Investigation (supporting); Writing – review
698 and editing (equal); Supervision (equal); Funding acquisition (equal).

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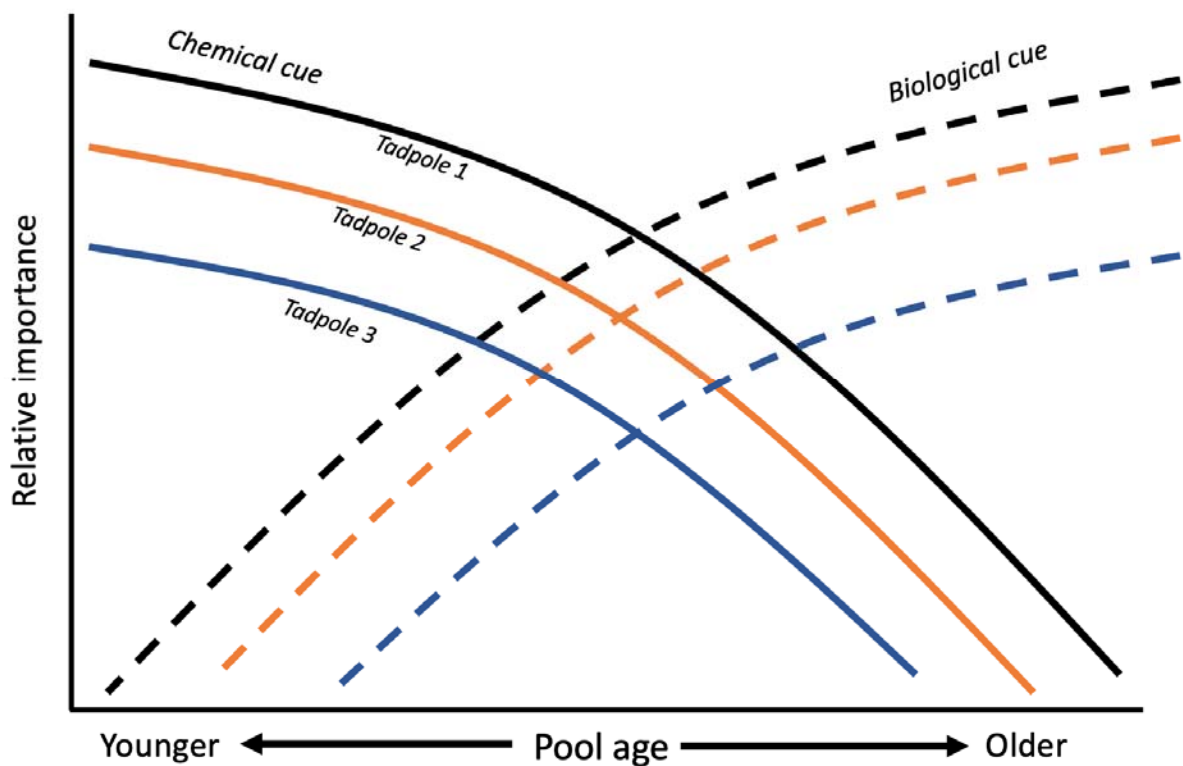
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856

857 **Speculations and alternative viewpoints**

858 *Dendrobates tinctorius* males typically father egg clutches of 2-5 tadpoles per clutch and
859 breed year-round (Rojas and Pašukonis 2019). As a result of the presumed high energetic
860 expense from carrying each tadpole from each clutch singly, we hypothesize that tadpoles
861 transported later may be subject to bet-hedging by fathers.



862

863 *Fig xx. Hypothetical succession of relative cue importance in phytotelmata across time.*

864

865 Combined with the important chemical aspects of pools shown from 2019 data, it seems that
866 fathers can cue on either chemical (KH, salinity, hardness) or biological components (the
867 presence of conspecifics) as information about pool stability. We speculate that the
868 importance of chemical and physical cues changes with respect to pool age. For example, in
869 new uninvaded pools, chemical cues of the pool may be more important (left-side of figure),
870 while in older, established pools that are more densely occupied, density serves as a main cue
871 for transporting fathers (right-side of figure). Finally, the value of these cues may vary with
872 the amount of offspring fathers are caring for. Hypothetically, fathers who must transport
873 more offspring are less discerning of where they transport latter tadpoles because they can
874 afford to make less “ideal” deposition decisions because of their large reproductive output.