

1 **Emergent vulnerability to intensive coastal anthropogenic disturbances in** 2 **mangrove forests**

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8 9 **Abstract**

10 Mangrove forests, as one of the most productive coastal ecosystems in tropical and subtropical
11 areas, provide multiple valuable ecosystem services for human well-being. Mangrove coverage
12 has been declining dramatically across much of developing regions due to extensive coastal
13 anthropogenic disturbances such as reclamation, aquaculture, and seawall construction. As
14 coastal human activities increase, there is urgent need to understand not only the direct loss, but
15 also the vulnerability of mangroves to anthropogenic disturbances. In this study, we evaluated
16 spatial pattern of mangrove vulnerability based on the conceptual framework of “Exposure-
17 Sensitivity-Resilience” using geospatial datasets in mainland China. We find that within all
18 25,829 ha mangroves in five coastal provinces of mainland China in 2015, nearly 76% of
19 mangroves was exposed or threatened by anthropogenic disturbances. Coastal reclamation and
20 aquaculture were the key threats causing mangrove vulnerability. The overall distribution of high,
21 medium and low vulnerability was following similar trend of aquaculture distribution, which
22 suggests aquaculture was the greatest anthropogenic disturbance agent to mangroves. Hotspot
23 regions for mangrove vulnerability are located at the developing provinces such as Guangxi and

24 Hainan. This study provides the first spatially explicit evidence of the vulnerability of mangrove
25 forests to intensive coastal anthropogenic disturbances at national scale, could serve as a
26 benchmark for navigating coastal ecological redline management and coastal ecosystem
27 restoration.

28 **Keywords:** mangroves; human activities; reclamation; urbanization; aquaculture; seawall

29 **1 Introduction**

30 Global assessments indicate that coastal tropical and subtropical areas are threatened by the
31 greatest urbanization rates, suggesting an alarming to the conservation of coastal ecosystems in
32 this regions and their associated ecosystem services (Branoff, 2017; Friess et al., 2016; Rocha et
33 al., 2018; Seto et al., 2012). Mangrove provides essential ecosystem services to coastal
34 community, including products provision, cultural services, carbon sequestration, water
35 purification, shoreline protection/regulation and biodiversity conservation (Bell & Lovelock,
36 2013; Crooks et al., 2018; Estoque et al., 2018; Friess & Webb, 2014). However, intensive
37 anthropogenic disturbances caused by rapid coastal development exacerbates degradation of
38 coastal systems. Critical functions of coastal wetland ecosystems have been eroded by global-
39 scale anthropogenic pressures, which leads to irreversible land use and cover change, massive
40 biodiversity loss and overloaded land-based sources of pollution (Duke et al., 2014; Wang et al.,
41 2018).The world's area of mangrove forests has decreased by about 35% globally from 1980 to
42 2000 and experienced a decline of 1.97 % annually from 2000 to 2012 (Hamilton & Casey, 2016;
43 Valiela et al., 2001), about 3-5 times greater than that of terrestrial forest loss (Duke et al., 2014).
44 Coastal reclamation and conversion of mangroves to aquaculture or agriculture creates pressure
45 on mangrove ecosystems and has thus been a major driver of mangrove destruction (Flores-de-
46 Santiago et al., 2017; Mukherjee et al., 2014; Spalding, 2010). Numerous studies have quantified
47 increasing coastal reclamation and modifications over time, with demonstrated the deleterious
48 effects of urban, agricultural, aquaculture, and infrastructure development on mangrove
49 ecosystem (Hamilton & Friess, 2018; Rivera-Monroy et al., 2017). Global sustainability science
50 increasingly recognize the changes of structure and function taking place in ecosystem,

51 understanding which ecological functions are vulnerable to anthropogenic disturbances is key to
52 sustainable development (Richards & Friess, 2016).

53 Vulnerability defines the degree to which a system or system component is susceptible to
54 experience harm due to exposure to a perturbation or stressor which mostly associated with
55 environmental, social change, and from the absence of capacity to adapt (Adger, 2006; Turner II
56 et al., 2003; White, 1974). Vulnerability analysis in coupled social-ecological systems draws on
57 three major concepts: exposure, sensitivity and resilience (IPCC, 2014; Turner II et al., 2003).
58 Through many years of studies on mangroves under climate change to reveal its distinct
59 structures and process (Lovelock et al., 2015; Duke et al., 2014; Webb et al., 2013; Xu et al.,
60 2016), there remains limited conceptual framework or model for the vulnerability of mangroves
61 to anthropogenic disturbances (Branoff, 2017; Mukherjee et al., 2014; Ventura & Lana, 2014).
62 Predictions of mangrove vulnerability around cities are mostly reported in low- and lower-
63 middle-income regions of Africa, Latin America and India (DasGupta & Shaw, 2013; Elmqvist
64 et al., 2013; Nortey et al., 2016).

65 Urbanization, in terms of landscape pattern changes, is particularly detrimental to degradation
66 of mangrove in the Anthropocene. Human-induced coastal reclamation, as a typical land use
67 change, is one of the key drivers of mangrove deforestation (Richards & Friess, 2016).
68 Landscape changes caused by the human activities of coastal ecosystems can produce both direct
69 and indirect effects on the long term sustainability of mangrove wetlands and coastal
70 communities (Koh & Khim, 2014; Yim et al., 2018). These changes raise questions such as: how
71 to identify the pattern of vulnerable mangroves to multiple environmental changes, and what are
72 principal threats to local mangroves? This recognition requires revisions and enlargements in the
73 basic design of vulnerability assessments, including the capacity to treat coupled social-

74 ecological systems and those linkages within and without the systems that affect their
75 vulnerability.

76 Here, our aim was to evaluate the spatial pattern of mangrove vulnerability to coastal
77 anthropogenic disturbances in mainland China, the regions where experienced a rapid and
78 unprecedented process of urbanization. Our specific research questions were: (1) what is the
79 spatial pattern of mangrove distribution across mainland China? (2) what is the spatial pattern of
80 mangrove vulnerability to reclamation in mainland China? (3) what are the principle disturbance
81 agents to mangroves? For addressing these three questions, we firstly mapped the mangrove
82 extent for mainland China by using high-resolution Google Earth images and visual
83 interpretation, and subsequently developed a methodology to evaluate the vulnerability of
84 mangrove forests to three coastal anthropogenic disturbances (i.e., reclamation, aquaculture, and
85 seawall).

86

87 **2 Study area**

88 Since half of global mangrove areas are distributed within 25 km of urban centers inhabited by
89 dense human population (McLeod & Salm, 2006; Millennium Ecosystem Assessment, 2005),
90 more than 90% of the world's mangroves are destroyed or threatened by diverse forms of human
91 activities in recent decades (McLeod et al., 2011; Murray et al., 2018; Silliman et al., 2009).
92 Alarming losses of mangrove cover have occurred in most of developing areas of China due to
93 extensive coastal reclamation, deforestation, engineering and urbanization. Although mangroves
94 in China cover only 0.14% of the global mangrove area, one third of mangrove species can be

95 found in this region (Romañach et al., 2018), suggesting that valuable contribution of mangrove
96 conservation in China to biodiversity conservation in global scale.

97 Natural mangroves grow along the southeast coast from Fujian Province to Hainan Province
98 ($18^{\circ}12'-27^{\circ}20'N$). Planted mangroves are scattered within boundary of natural mangroves. To
99 protect local mangrove ecosystems, there are 35 conservation areas and several Ramsar
100 Convention sites in China, e.g., Dongzhaigang Mangrove Nature Reserve in Hainan Province,
101 Fujian Zhangjiangkou National Mangrove Nature Reserve.

102 Extensive coastal reclamation occurred in rapid urbanizing metropolitans of mainland China,
103 such as Bohai bay, Yangtze River delta and Pearl River delta, was responsible for approximately
104 950,000 ha coastal wetlands loss and consequently ecosystem service decrease in recent years
105 (Sajjad et al., 2018; Tian et al., 2016). According to the statistical data from the State Oceanic
106 Administration, People's republic of China in 2015, the impacts of coastal reclamation on
107 ecosystem services resulted in the loss in wetlands of about \$31,000 million, which accounts for
108 6% of ocean economy in China (State Oceanic Administration, 2016).

109

110 **3 Materials and Methods**

111 3.1 Mangrove mapping

112 For assessing the spatial vulnerability of mangroves, we firstly mapped the extent of mangrove
113 forest in mainland China using artificial visual interpretation. Processes of mangrove
114 interpretation included extracting mangroves with difference sources of base map (global
115 mangrove datasets, Landsat images, etc.) and modification with filed verification in each
116 province. Prior to interpretation, we imported the global mangrove distribution vector data in

117 2000 (<https://www.usgs.gov/>) as a base map and selected some mangrove samples in all five
118 provinces as test examples for mangrove identification. By modifying these samples under the
119 guidance of mangrove experts, mangroves are shown as dark green ribbon pattern and uniform
120 texture feature along coastline (Fig. S1). To eliminate visual interference from salt marsh species
121 (e.g., *Spartina alterniflora*), we selected Google Earth images in winter 2015 because mangroves
122 can be more easily identified during dormant season of salt marshes (Fig. S2). Meanwhile,
123 considering tidal inundation to mangroves, we compared the images of different time period in
124 same areas and selected images with low tide. This procedure helped to eliminate the error in
125 extracting submerged mangroves. Ultimately, we saved the mangrove layer in Google Earth and
126 exported to ArcMap 10.2. After interpretation with remote sensing images, we verified
127 mangrove area by field survey from February 2016 to March 2018 to examine interpretation
128 accuracy of mangrove forest map. We selected 56 mangrove validation sites and identified
129 mangrove distribution using Unmanned Aerial Vehicle (UAV) vertical photography and field
130 survey. About 758 UAV images were used to create a confusion matrix indicating the producer's
131 and user's accuracy. Availability of aerial photographs in our study area facilitated ground truth
132 process in evaluating accuracy of classified mangroves.

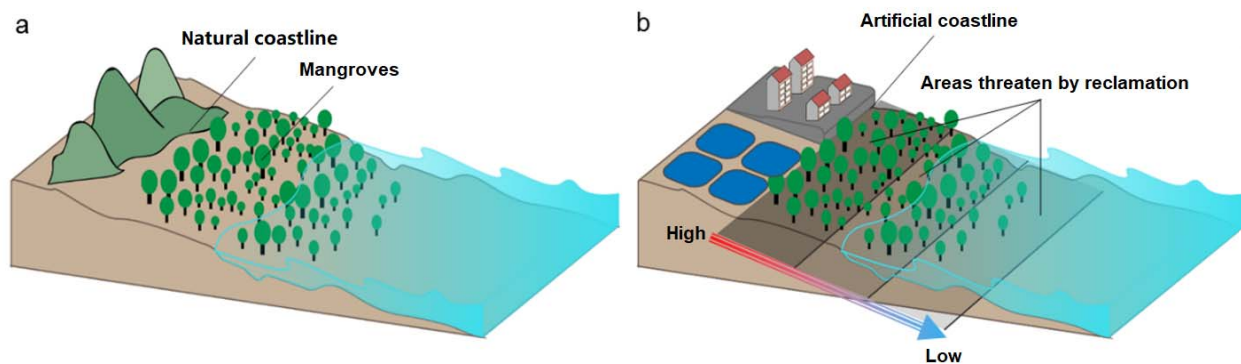
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134 3.2 Detection of anthropogenic disturbances

135 3.2.1 Reclamation activities

136 Reclamation is defined as the conversion of coastal land to agricultural, industrial, and urban
137 land use (Tian et al., 2016). We assume that the impacts of reclamation activities are decreasing
138 with distance to artificial coastline, which means the longer distance to coastline the less
139 negative influences of reclamation activities to local mangroves. We created a buffer of coastline

140 sourced from Sajjad et al., 2018, and the buffer radius is set as 900 m since this distance covers
141 98% of mangroves in mainland China. Then we conduct a stress-gradient analysis in three levels
142 of pressure based on 300 m intervals (Fig. 1, adapted from Sutton-Grier et al., 2015). The coastal
143 area within the first 300 m buffer zone adjacent to reclamation suggests the highest level of
144 pressure from reclamation activities on mangroves, which was assigned with a high-pressure
145 value.



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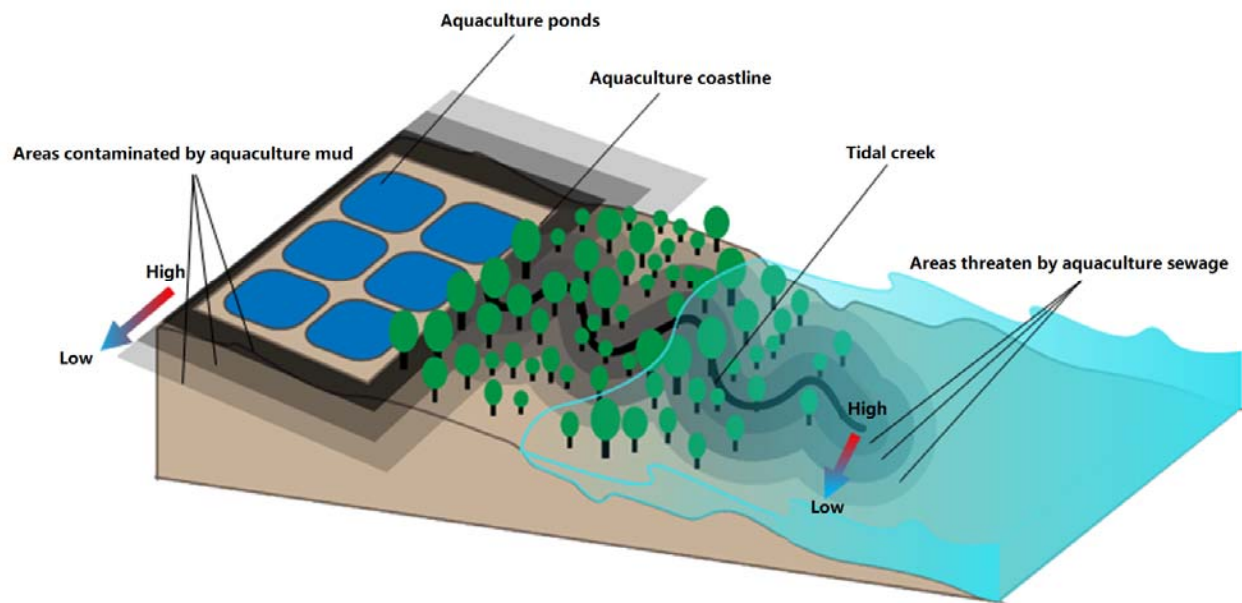
147 **Fig. 1.** A schematic diagram of the spatial distribution of the reclamation pressure.

148

149 3.2.2 Aquaculture pollution

150 Aquaculture pollution dataset contains two layers-aquaculture ponds and tidal creek, which are
151 the main sources of pollutants in local ecosystem. Muddy sediments and wastewater of
152 aquaculture ponds produce highly concentrated pollutants that lead to degradation of mangrove
153 system (Xin et al., 2014). Aquaculture ponds data was extracted from the layer of coastline in
154 our previous study (Sajjad et al., 2018), tidal creek was detected by visual image interpretation
155 based on Google Earth. We created buffer zones for both layers with distance of 70 m, 140 m
156 and 210 m separately to model stress-gradient analysis regarding aquaculture pollution (Fig. 2,

157 adapted from Sutton-Grier et al., 2015). High threats from aquaculture pollution was assigned in
158 the coastal area of 70 m buffer zone, while relative low threats within the 210 m buffer zone.



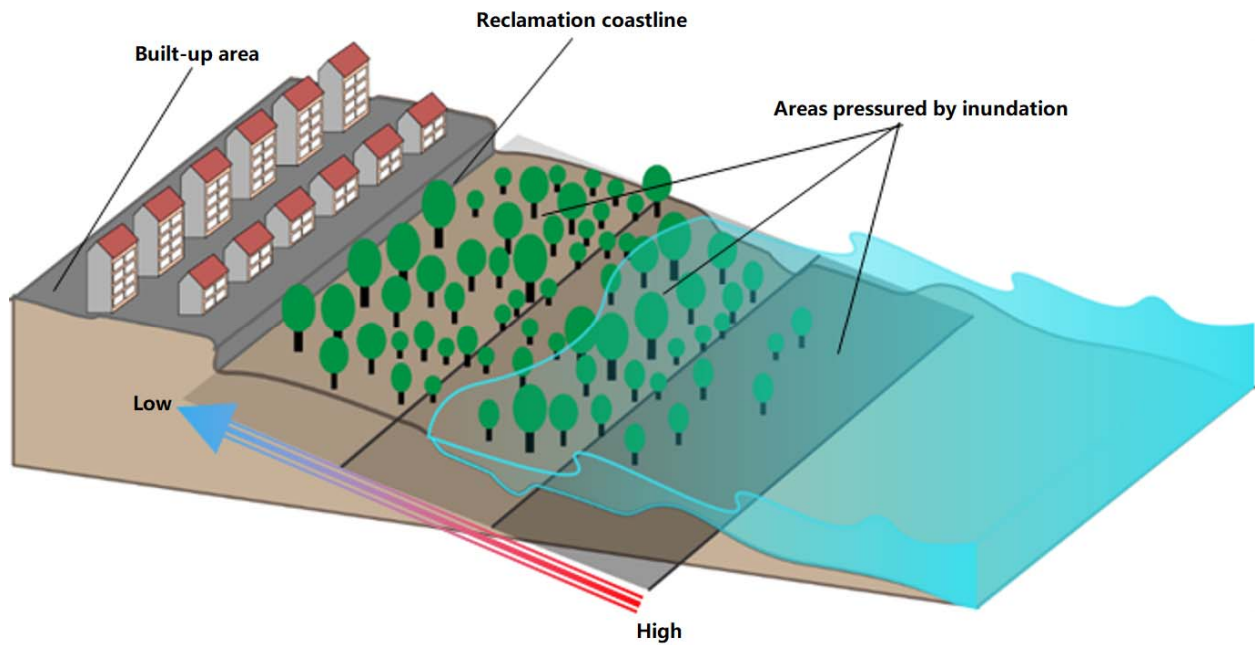
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160 **Fig. 2.** A schematic diagram of the spatial distribution of the aquaculture pollution pressure.

161

162 3.2.3 Seawall

163 With the increase of dam construction and large reservoir projects in developing countries
164 (Moran et al., 2018; Shi et al., 2018), large area of landscape transformed from natural coastline
165 to seawall. The newly built impervious surface areas squeeze coastal zone toward ocean with
166 increasing risk of flood (Doody, 2013; Ramesh et al., 2015). Three buffer zones were created
167 along the claimed coastline with the distances of 300 m, 600 m and 900 m to create stress-
168 gradient layer of flood risk (Fig. 3, adapted from Sutton-Grier et al., 2015). In this case, 300 m
169 buffer zone is the area with the lowest flood threat, which is opposite to the threat distribution of
170 reclamation threat.



171

172 **Fig. 3.** A schematic diagram of the spatial distribution of the seawall barrier pressure.

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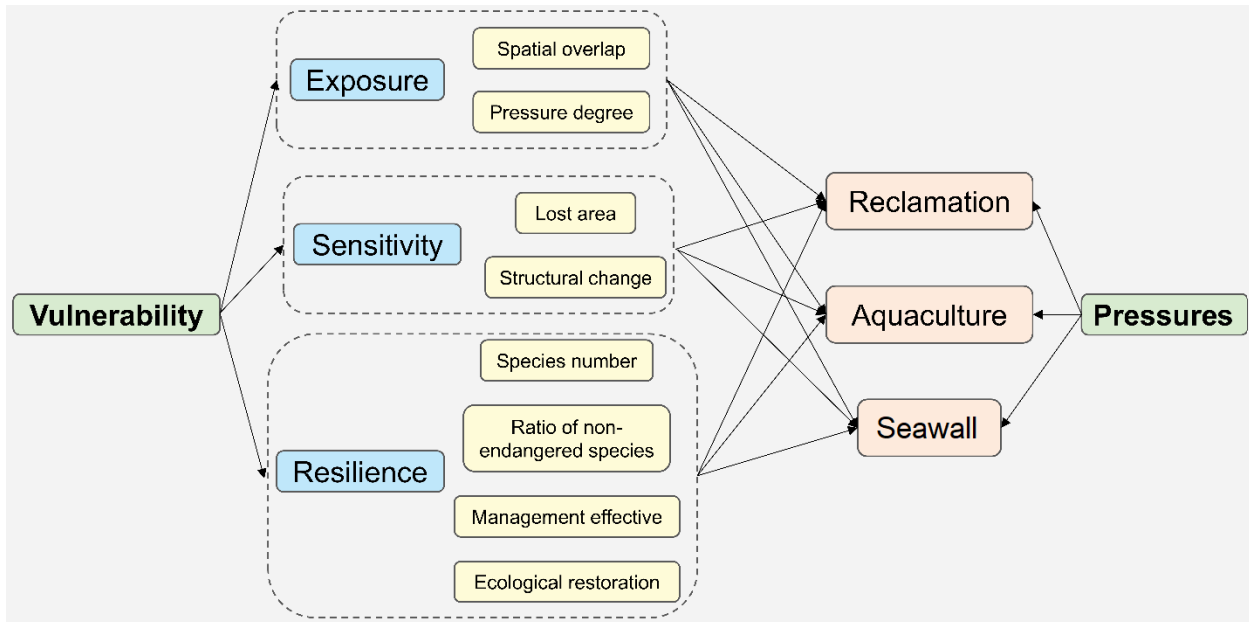
174 3.3 Vulnerability assessment

175 Referring to the vulnerability framework proposed by Turner II et al. (2003), we proposed a
176 spatial assessment framework to evaluate mangrove vulnerability using InVEST model, which is
177 a spatially-explicit modelling tool for assessing ecosystem services and would return a suite of
178 results in a raster format to evaluate ecosystem states. (Sharp et al., 2015). We used the Habitat
179 Risk Assessment submodule from InVEST model with inputs of all anthropogenic disturbance
180 agents to local mangroves. Vulnerability of mangroves was characterized by integrating three
181 subsystems: exposure, sensitivity and resilience (Fig. 4; Table S1).

182 Exposure index represents the amount and intensity of anthropogenic stressors to mangroves.

183 Exposure value was divided by the overlapped area of mangroves and reclamation and

184 reclamation pressure intensity. Sensitivity index indicates negative impacts of reclamation on
185 mangroves, and was quantified by mangrove loss area and mangrove species change. Resilience
186 characterizes the ability of mangroves to maintain their structural and functional stability in the
187 process of and after reclamation disturbances. And it was quantified by mangrove species
188 number, ratio of non-endangered species, regional ecological management, and potential
189 restoration areas for mangroves. We then clustered these indexes to corresponding vulnerability
190 ranking of low (value=1), moderate (value=2) and high (value=3) using *K*-means algorithm (See
191 Supplementary material for more details).



192

193 **Fig. 4.** Indexes of mangrove vulnerability to anthropogenic disturbance pressures.

194

195 Exposure indicators represent risks of mangrove while experiencing vast reclamation activities,
196 which describes characteristics and components of exposure to reclamation. Inputs of exposure
197 included polygons of mangroves and reclamation with attributes about mangrove distribution

198 and 12 types of reclamation pressure. Sensitivity is a dose-response of mangrove system to
199 reclamation, it is an interaction between reclamation and mangrove conditions. Sensitivity, as an
200 interlinked factor between the other two aspects, was calculated with potential loss and potential
201 structure change of mangroves to different reclamation pressures. Resilience refers to recovery
202 capacities of mangrove system to reclamation activities, resilience was represented by both two
203 internal indicators (species numbers and ratio of non-threaten species) and external resilience
204 indicator (ecological management).

205 Vulnerability of mangroves can be characterized by degree to which mangrove is likely to
206 experience damage during coastal development. Vulnerability of mangroves is predicated on
207 trade-offs between economic development and coastal ecosystem services maintenance, as they
208 are affected by interconnections operating at different spatiotemporal and functional scales.
209 Based on vulnerability framework, the Habitat Risk Assessment model in InVEST was used to
210 calculate spatial vulnerability of mangroves in each province while adapting to coastal
211 reclamation (Sharp et al., 2015). Combination of vulnerability framework and InVEST model
212 produces spatial qualitative estimate of potential risks in terms of vulnerability value while
213 exposed to reclamation, which differentiates areas with relatively high or low exposure to
214 reclamation threats.

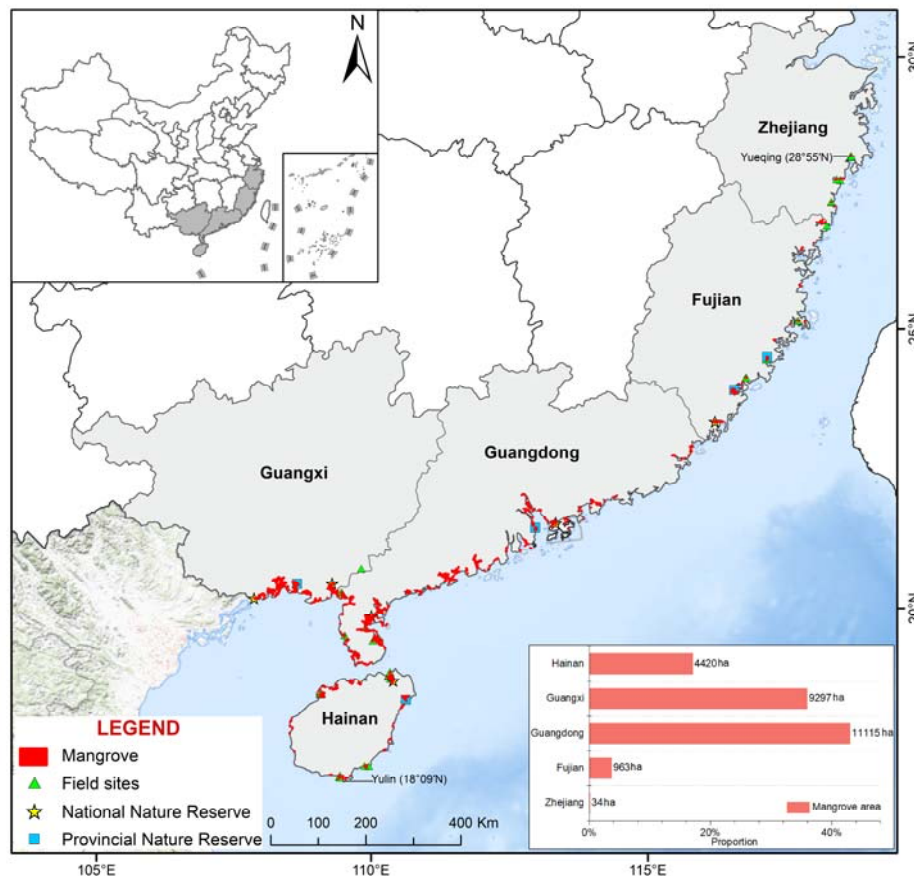
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216 **4 Results**

217 4.1 General distribution of mangroves

218 In total, 25,829 ha mangroves were identified in five coastal provinces of mainland China in
219 2015 (Fig. 5), 96.14% of which were distributed in southern part of coastal provinces (Hainan,

220 Guangxi and Guangdong Province), mainly located in the Beibu Gulf Economic Rim (the
221 economic region surrounding around China's southwestern coastal area). About 11,115 ha (43%)
222 mangroves was classified in Guangdong Province, where many coastal areas are dominated by
223 sedimentary environments. Followed by Guangxi Province with a number of 9,297 ha
224 mangroves, the third largest area of mangroves was 4,420 ha in Hainan Province. About 963 ha
225 and 34 ha mangroves were detected in Fujian Province and Zhejiang Province respectively.



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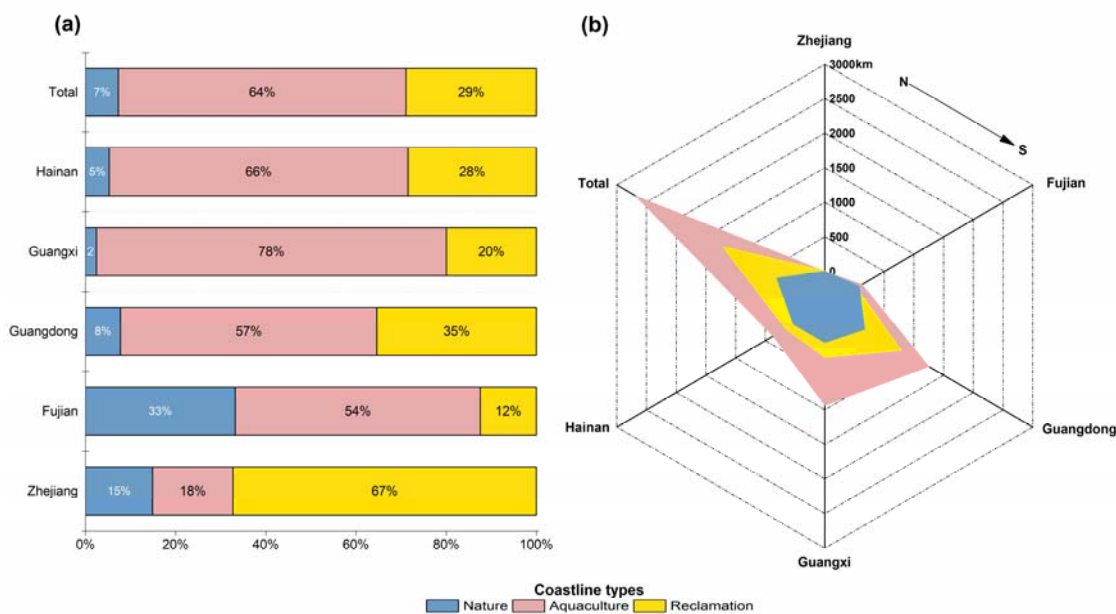
Fig. 5. Mangroves distribution in five coastal provinces.

228 4.2 Principal disturbance agents to mangroves

229 Landscape conversion from natural coastline to artificial coastline was the key driver of
230 mangrove degradation. By overlapping coastline and mangroves, we found that only 7% of
231 mangrove was adjacent to natural coastline types (Fig. 6). Most of mangroves were exposed to

232 anthropogenic disturbances caused by redeveloped coastlines, where 64% of mangrove was
233 located near aquaculture ponds and 29% of mangrove was threatened by filled land. It suggests
234 that 93% of mangrove was squeezed by human-built bank and dam toward ocean, and they were
235 threatened by floods on the marine side.

236 In a local scale, reclamation activities in Zhejiang Province resulted in the highest density of
237 filled coastline among all coastal provinces. Followed by Guangxi Province, high percentage
238 (98%) of artificial coastline was adjacent to mangrove, due to large areas of aquaculture feeding
239 ponds. In all these five provinces, the highest ratio (30%) of natural coastline adjacent to
240 mangroves appeared in Fujian Province. Even though the coverage of aquaculture in Fujian,
241 Guangdong and Hainan Province were lower than that in Guangxi, but aquaculture land in all
242 provinces covered more than that of natural land. Guangdong was the province which has the
243 longest coastline both artificial and nature, followed by Guangxi, demonstrating that these two
244 provinces need to be given priority (Fig. 6b).



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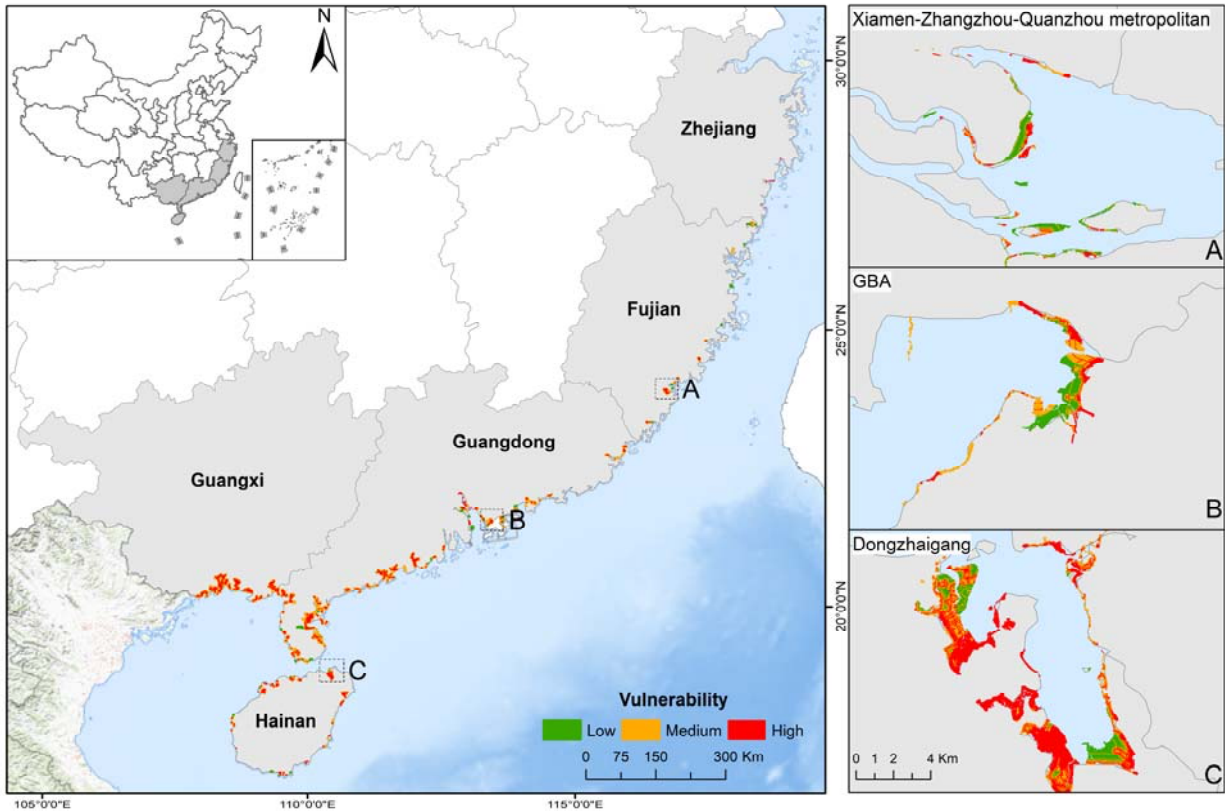
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Fig. 6. Proportions (a) and lengths (b) of different coastline types in 2015.

247

248 4.3 Spatial distribution of mangrove vulnerability

249 In this study, spatial distribution of vulnerability in mangroves was derived by combining
250 vulnerability framework and Habitat Risk Assessment (HAR) model in InVEST (Fig. 7). We
251 found that 29% of mangrove areas was identified under high vulnerability, only 24% of them
252 was estimated as area with low vulnerability. The average value of vulnerability in all five
253 provinces were ranked from high to low as Zhejiang, Hainan, Guangxi, Guangdong, and Fujian.
254 As can be seen from Fig. 8, all mangroves in Zhejiang Province were under high vulnerability,
255 which indicates high risk of degradation in this region. Mangroves in Zhejiang Province were
256 planted species, due to its climate conditions and high coverage of artificial coastline, there is
257 limited suitable natural coastal areas for mangroves. About 64.41% of mangroves in Hainan
258 Province were observed under high vulnerability, only 9.44% of them were remained in low
259 human influences. The main reason of such high vulnerability in this region was caused by
260 aquaculture pollution and high ratio of endangered species. Nearly 25% of mangroves in both
261 Guangxi and Fujian Province was classified with high vulnerability, it was because of marine
262 aquaculture and coastal redevelopment, all of which have led to significant degradation and
263 ecological disturbances to mangroves.



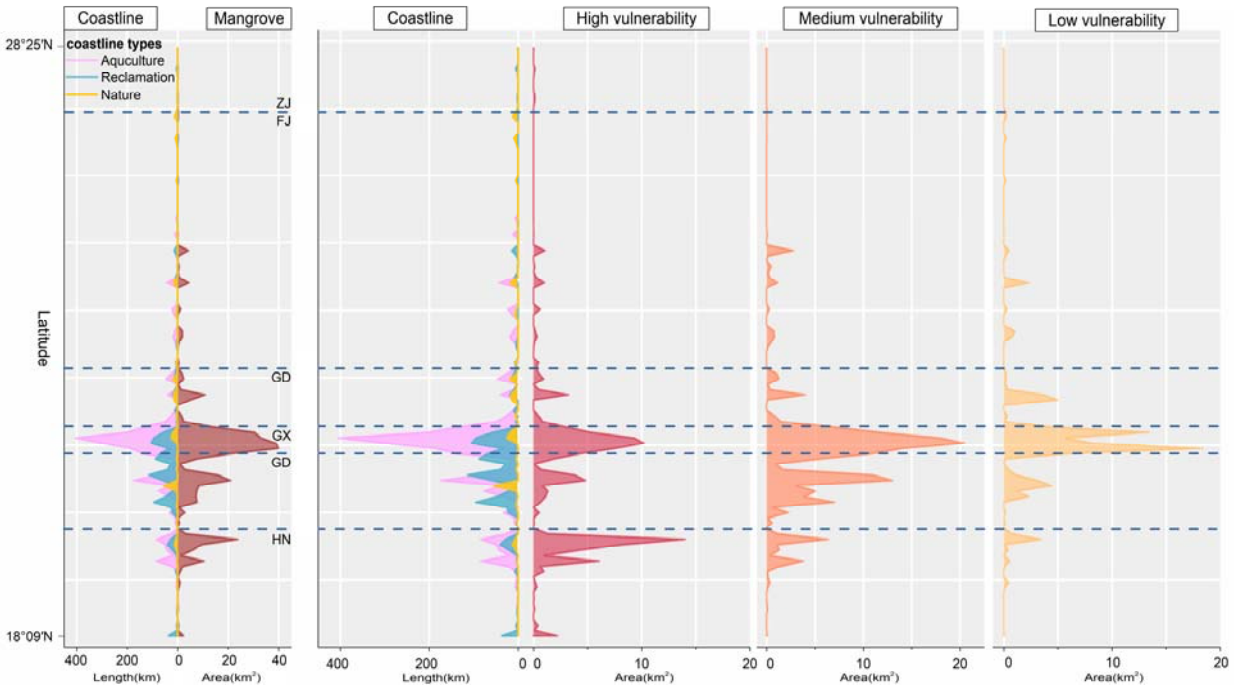
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265 **Fig. 7.** Distribution of mangrove vulnerability to anthropogenic disturbances along China's
266 coastal area. (GBA: Guangdong-Hongkong-Macao Greater Bay Area).

267

268 Among all these selected vulnerability indicators, aquaculture and reclamation coastlines were
269 the key drivers of mangroves degradation. Areas with high vulnerability in Guangxi and Hainan
270 Province was found near large areas of aquaculture and reclamation. The overall distribution of
271 high, medium and low vulnerability was following similar trend (Fig. 8), the largest groups of
272 vulnerability (peak values) appeared in Guangxi, Guangdong and Hainan Province. However,
273 peak values of three levels of vulnerability were slightly different, while that of medium and low
274 vulnerability were in Guangxi Province, and mangroves in Hainan had the largest area of high

275 vulnerability. In general, based on the distribution of coastline and vulnerability, we found that
276 aquaculture was primarily threats to high risk of mangroves degradation in these five provinces.



278 **Fig. 8.** Spatial distribution of coastline with mangrove and different level of vulnerability. The
279 left side of Y axis shows the latitude of different types of coastline (aquaculture, reclamation,
280 and natural) from Zhejiang Province to Hainan Province (ZJ: Zhejiang Province; FJ: Fujian
281 Province; GD: Guangdong Province; GX: Guangxi Province; HN: Hainan Province), the right
282 side of Y axis represents area of mangroves and different level of vulnerability spatially. The left
283 X axis is the length of mangroves and right X axis is the area of different level of vulnerability.

284

285 In a local scale, mangroves in urban agglomerations experienced higher vulnerability than that
286 rest of mangrove (Fig. 7). As the Guangdong-Hong Kong-Macao Greater Bay Area (GBA)
287 becomes an important economic zone in China, rapid urbanization in coastal areas lead to high
288 level of disturbance caused by human activities, which is the principal threat to local mangrove

289 ecosystem. In the meantime, Guangdong Province has the largest area of mangrove in all five
290 provinces, which includes 7 natural protected areas for mangrove conservation. These protected
291 areas help to maintain key functions of local mangrove ecosystems under coastal reclamation.
292 The northern boundary of natural mangroves is located in Fujian Province, the Xiamen-
293 Zhangzhou-Quanzhou metropolitan in this Province was also exposed to relative high
294 vulnerability due to extensive reclamation activities. Even if protected in a national nature
295 reserve, mangroves in Dongzhaigang also faced serious threats due to the substantially increased
296 aquaculture ponds and construction area.

297 4.4 Accuracy assessment

298 Our results show that the map of mangrove forests in 2015 showed a high user's accuracy of
299 97.4% and producer's accuracy of 82.2%. We also compared our results with similar research in
300 China to quantify the reliability of our interpretation (Jia et al., 2018). Detected total area of
301 mangroves (25,829 ha) in our research was comparable with their results derived from Landsat
302 archive (22,419 ha). Due to the coarse resolution of Landsat archive (30 m), small mangrove
303 fragments in urban agglomeration identified in our study was not able to be classified by using
304 Landsat images. Therefore, the identified area of mangrove in our study is larger than their result.
305 The main difference between these two datasets was in the Pearl River Estuary and Xiamen-
306 Zhangzhou-Quanzhou metropolitan. In order to improve our accuracy of mangrove identification
307 in those regions, field investigation was carried out in summer time. Overall, the validation
308 results suggest that our mangrove map is mainly consistent with ground truth and similar
309 research in China.

310

311 **5 Discussion**

312 5.1 Impacts of principal threats on vulnerability distribution to local ecosystems

313 Due to continuous influences of coastal reclamation and aquaculture farming, mangroves are
314 exposed to dynamic and extensive threats in local region. Conversion of mangroves to
315 aquaculture ponds was encouraged by local governments, which becomes main income of
316 coastal regions. Among all types of coastal development activity in terms of land use change, the
317 most significant replacement of land use was aquaculture (64% of total area), which leads to high
318 vulnerability of local mangroves.

319 As comparing all three levels of vulnerability distribution, we found that mangroves with
320 medium level of vulnerability was the most sensitive area to aquaculture factor. The overall
321 distribution pattern of medium vulnerability was coordinated with that of aquaculture. Except in
322 Zhejiang Province, aquaculture was the domain land use type in local mangrove areas, which
323 threatened more than half of mangroves in our study area. Threat of aquaculture in driving
324 mangrove deforestation is not only a problem in China, but also was found in southeast Asia
325 countries, such as Thailand, Indonesia, Vietnam, and the Philippines (Primavera, 2000; Richards
326 & Friess, 2016).

327 Land redevelopment and infrastructure construction was another direct factor leading to
328 mangrove loss. Even it influences on vulnerability was not that significant as aquaculture, it
329 shows high correlations with high level of vulnerability, especially in the areas dominated by
330 reclamation coastlines in Hainan and Fujian Province. In those areas adjacent to long artificial
331 coastline, it also addresses high level of influence on vulnerability. Extensive and large area of
332 land reclamation occurred in Pearl River Delta (Guangdong Province) and Beibu Gulf (Guangxi
333 Province), these regions were shown with high vulnerability of mangrove degradation. Coastal

334 artificial infrastructures, such as large industrial complexes, coastal interstate highways, and
335 airports, occupied areas suitable for mangrove growth, and accounted for 29% of total mangrove
336 areas in mainland China. Local government historically aimed to increase terrestrial land through
337 expansion towards ocean, especially in the most developed metropolises, such as GBA, Bohai
338 Bay (Tianjin Province), Hangzhou Bay (Zhejiang Province) and Min River estuary (Fujian
339 Province). High intensity of human disturbances toward ocean poses potential stress to mangrove
340 ecosystem, which reduce the buffer zone of mangroves in adapting to sea level rise.

341 5.2 Threats to local ecosystem services: habitats and biodiversity

342 Mangroves degradation and loss is not only bounded with mangrove ecosystem itself,
343 furthermore, spillover effects of mangroves degradation have significant influences on
344 ecosystem processes and services that they provide. Vulnerability of habitats and biodiversity
345 increased as structure and functionality of local ecosystems were challenged by anthropogenic
346 disturbances. Within areas of high vulnerability in mangroves, the most challenges to local
347 ecosystems were habitats degradation and biodiversity loss (Arkema et al., 2014). Nearly 35 ha
348 of mangrove disappeared in Futian mangrove conservation area because of city construction,
349 which leads to 39% of bird density loss and 45% of species reduction (Wang & Wang, 2007).
350 Species extinctions can be followed by loss in functional diversity, particularly in mangroves
351 with high vulnerability. Further degradation in mangroves is likely to be followed by accelerated
352 functional losses. New top-down policies adopted by China's central government in 2016 has
353 imposed a conception of 'Redline', a boundary delineating a coastal strip in which to preserve
354 existing natural habitats, both for maintaining ecosystem function and to reduce conflicts
355 between natural processes (exacerbated by accelerated sea-level rise) and human settlements and
356 infrastructure. The Central Leading Group for Comprehensively Deepening Reforms of China

357 approved a series of decisions of Redline policy in 2016. It comprises “Opinions on Delimiting
358 and Guarding Ecological Protection Redline”, “Program for Wetland Protection and Restoration”,
359 and “Rules on Coastal Line Protection and Utilization”, which all emphasize the integration of
360 “One Redline Map” in China’s coastal area. China’s coastal Redline would be challenged by
361 2020 in maintaining 35% (6,300 km) of coastal line, which is the goal set by the central
362 government in protecting coastal line (Larson, 2015). Consequently, transitioning the coastline
363 from natural to artificial, through such methods as land claim or armoring, makes coastal areas
364 more vulnerable in terms of lost biodiversity and increased natural hazards (Arkema et al., 2015;
365 Steffen et al., 2015). Mangroves, as a typical coastal ecosystem included in the Redline area,
366 maintain key functions of local ecosystems, they also contribute to biodiversity conservation.

367 In 2018, China’s governments announced to halt all coastal land reclamation that related to
368 business land use. But degradation of mangroves and irreversible landscape changes are still
369 challenging in maintain key functions of coastal ecosystems. As mangrove habitats are smaller or
370 fragmented, their essential ecosystem services may be lost within and beyond its boundary. The
371 central government should act quickly to nominate and protect vulnerable coastal sites, because
372 this would be consistent with the national Redline program and reclamation restoration. The
373 local governments of coastal provinces and cities should also proactively adopt effective
374 measures of their own, consistent with the national Redline policies, to protect vulnerable coastal
375 ecosystem.

376 5.3 Limitations

377 There are some uncertainties in this study. The 900-m-buffer of coastal line and the 300 m
378 interval in the buffer zone were chosen based on general distribution of mangroves in our study
379 area, which may different while implying to other regions. Moreover, in the current analysis, all

380 8 indicators in vulnerability assessment were treated with equal weight, which was set according
381 to current states of coastal region in China. This may various due to spatial heterogeneity in
382 different area, and need to adjust while applying to other cases.

383 Spatial outputs of mangrove vulnerability in this research provide a qualitative representation
384 of the relative contributions of anthropogenic disturbances to coastal vulnerability and highlight
385 the potential threats of reclamation in degrading mangrove ecosystems. Further information on
386 social, economic and coastal disaster are helpful in giving a comprehensive assessment of
387 mangroves ecosystem and better informing development strategies and permitting in setting
388 conservation priorities, monitoring deforestation and forest degradation.

389

390 **6 Conclusion**

391 Mangroves are under threats from different types of anthropogenic disturbances, coastal rapid
392 development, in terms of aquaculture and reclamation, has been a principal driver of mangrove
393 loss and degradation in China, and it becomes a growing concern worldwide. These two factors
394 were the main drivers of medium and high level of vulnerability. The continuously altered,
395 destroyed or transformed mangroves with high vulnerability is an early warning signal of
396 dramatic functionality loss of such major coastal ecosystem, which initiates more research of
397 mangrove in coastal urban environments.

398 Mangrove areas of high vulnerability suggest relative high level of combinate anthropogenic
399 disturbances from conversion for alternative uses (aquaculture, urban construction), which can
400 be prioritized for the development of conservation strategies. Distribution of mangrove
401 vulnerability also located the hotspots of key functionality loss and the vulnerable urban areas

402 with high risk of storm surges or climate-related disasters. As we estimated, the major
403 disturbance agent to mangroves was aquaculture ponds, especially in Guangxi and Hainan
404 Provinces. These findings indicate an urgent need for implementing adaptive strategies to serve
405 mangroves from anthropogenic disturbances, and provides a spatially-explicit assessment for
406 navigating mangrove conservation.

407

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