1 **Extracting spatial networks from capture-recapture data reveal structured**

2 **habitat use within a marine mammal's spatial range**

6 Tyler R. Bonnell [a,b,c], Robert Michaud [d], Angélique Dupuch [a,b], Véronique Lesage[e], Clément

7 Chion [a,b]

[a] Department of Natural Sciences, Université du Québec en Outaouais, Gatineau, Québec,

Canada

[b] Institut des Sciences de la Forêt Tempérée, Université du Québec en Outaouais, Ripon,

Québec, Canada

[c] Department of Psychology, University of Lethbridge, 4401 University Drive, Alberta T1K

3M4, Canada

[d] Groupe de Recherche et d'Éducation sur les Mammifères Marins (GREMM), 870 avenue

Salaberry, Bureau R24, Québec, QC G1R 2T9, Canada

[e] Maurice Lamontagne Institute, Fisheries and Oceans Canada, P.O. Box 1000, 850 route de

la Mer, Mont-Joli, QC G5H 3Z4, Canada

Corresponding author: Tyler Bonnell (tyler.bonnell@uleth.ca)

Running headline: Spatial networks from capture-recapture data

Abstract:

1. Estimating the impacts of anthropogenic disturbances requires an understanding of the

habitat use patterns of individuals within a population. This is especially the case

when disturbances are localized within a population's spatial range, as variation in

habitat-use within a population can drastically alter the distribution of impacts.

2. Here, we illustrate the potential for multilevel multinomial models to generate spatial

networks from capture-recapture data, a common data source use in wildlife studies to

monitor population dynamics and habitat use. These spatial networks capture which

regions of a population's spatial distribution share similar/dissimilar individual usage

patterns, and can be especially useful for detecting structured habitat use within the

population's spatial range.

3. Using simulations and 18 years of capture-recapture data from St. Lawrence Estuary

(SLE) beluga, we show that this approach can successfully estimate the magnitude of

similarities/dissimilarities in individual usage patterns across sectors, and identify

sectors that share similar individual usage patterns that differ from other sectors, i.e.,

structured habitat use. In the case of SLE beluga, this method identified multiple

clusters of individuals, each preferentially using restricted areas within their summer

range of the SLE.

4. *Synthesis and applications*. Multilevel multinomial models can be effective at

estimating spatial structure in habitat use within wildlife populations sampled by

capture-recapture of individuals. Our finding of a structured habitat use within the

SLE beluga summer range has direct implications for estimating individual exposures

to localized stressors, such as underwater noise from shipping or other activities.

Keywords: capture-recapture data, *Delphinapterus leucas*, habitat use, Multinomial Model,

network community detection, photo identification, spatial networks.

1. Introduction

An understanding of the spatial and temporal distribution of a species of concern is of central importance to conservation and management (Evans & Hammond, 2004). The existence of spatial structuring within populations can have important ecological and management implications. If a population as a whole can be considered as highly mixed, i.e., with individuals showing no strong patterns of home range use or sub-structuring within the wider population, then all individuals are equally likely to feel the impacts of local changes in the environment. In contrast, if the population cannot be considered to be highly mixed, and shows strong sub-structuring and site-fidelity patterns, local stressors might have a disproportionate impact on segments of the population. For example, if noise pollution increased in only one sector, in a highly mixed population all individuals would be lightly impacted, but in a spatially structured population a subset of the population would be highly impacted. These differences in spatial structuring of populations can lead to biased estimation of the likelihood and magnitude of impacts from local stressors both at the individual and population levels (DeFur et al., 2007).

Capture-recapture methods are commonly used to monitor individuals within populations, providing information on vital rates, demography, and insights into within population social mixing and habitat use (e.g., Koivuniemi et al., 2016). Photo-identification is a long-recognized method to 'capture' individuals with distinct markings (hereafter photo-ID data) (Urian et al., 2015), and digital photography along with high-resolution video and machine learning models to identify individuals has led to large capture-recapture datasets (Schneider et al., 2019). Novel statistical and computational methods applied to these capture-recapture datasets have enhanced the potential for quantifying within population structures through the use of social network analysis (Perryman et al., 2019; Schilds et al., 2019; Silk et al., 2021).

It is often the case, however, that efforts when collecting capture-recapture data are

not evenly distributed. This is especially the case when the population under study occupies a

large spatial extent, and where capture methods are not static as in the case with fixed camera

traps. This variation in sampling effort can heavily bias estimates of social and spatial

networks (Farine & Whitehead, 2015; Hupman et al., 2018; Whitehead, 2008). Data-stream

permutations have been used to assess potential biases in network estimates from capture-

recapture data when estimating networks directly from counts of individuals seen together or

in the same regions (Farine, 2017; Silk et al., 2021). Alternatively, state-space models have

been applied to capture-recapture data to include potential sampling biases in estimated

networks when based on counts of individuals seen together (Gimenez et al., 2019). Both of

these approaches build networks where individuals are the nodes, and the edges represent

links between individuals. Here we propose a multilevel multinomial model approach that

uses capture-recapture data to estimate spatial networks, i.e., where the nodes are spatial

regions and the edges between nodes represent the magnitude of similarity in the individual

using those regions. By taking this approach, it then becomes possible to quantify spatial

structure in habitat use within a population's spatial distribution.

The multilevel multinomial modeling approach that we propose to use here, does not

have a large body of literature to draw on for use with capture-recapture data, but presents

unique advantages (Koster & McElreath, 2017). If sampling efforts varies by region within a

population's spatial distribution, sighting probability of individuals could be greatly inflated

or deflated. The use of a multilevel structure, however, allows for sighting probabilities to be

nested per region and expressed in relative terms, i.e., as deviations from the mean

probability of sighting. This allows the approach to identify the relative magnitude of use of a

particular region for each individual. This generates a particular usage profile for each region,

i.e., which individuals highly/lowly use that region ("high users" and "low users" hereafter).

It is then possible to quantify how correlated the usage profiles between regions are,

providing information about which regions share similar/dissimilar usage profiles. We

suggest that this approach can successfully generate effort-corrected spatial networks within

populations, and can help identify differential patterns in habitat use among individuals and

regions.

To evaluate the performance of multilevel multinomial models at identifying spatial

structuring within animal populations, we first tested the approach with simulated datasets

with and without population spatial structuring. We then applied the method to observed data,

using a long-term (18 years) photo-ID dataset of beluga from the St. Lawrence Estuary,

Canada, and quantified spatial structuring within the population's summer range in the St.

Lawrence Estuary (SLE). Finally, we discuss how these estimates of population spatial

structuring provide important information for understanding current local stressors and their

potential impacts on this endangered, and declining population (Lesage, In Press).

2. Material and Methods

2.1 Data

Photo-identification boat surveys were conducted from June to October in 1989—2007 as

part of an ongoing long-term study on beluga social organization. The choice of survey area

on a given day was selected in a way to avoid resampling areas covered the previous days,

and also according to weather conditions. When beluga were encountered, the GPS position

of the research vessel was noted, and a herd follow was undertaken to photograph as many

individuals as possible within the herd using a handheld camera. A herd follow was limited to

three hours, with GPS location noted at least every 30 min. A detailed description of the

photo-ID survey protocol is available in (Michaud, 2014). Surveys were neither systematic

nor random in design, but covered various sectors of a large portion of the population's

summer distribution and a broad range of habitats. However, sampling effort was unevenly

distributed across the 14 defined sectors of the summer range (Fig. 1).

Each photograph was treated using standard protocols for image selection, scoring and

matching (Urian et al., 2015). Each uniquely-identified individual was attributed a

resightability index ranging from 1 to 3 based on the degree of distinctiveness of markings.

This resulted in a photo-identification collection of 821 unique individuals that were

recaptured on average 9 times over the study period (range 1—90), and which were each

associated with a GPS position and sector of initial encounter with the beluga herd.

Genetically-determined sexes were available for only 29%, of the individuals included to the

catalogue. From individuals with known sex, females had an estimated mean of 12 photo-

IDs, and males 17 photo-IDs suggesting some bias in terms of capturability (Table S1).

Given, age classes were not available for most individuals, and the low percentage of

individuals with known sex, the remaining analysis focused on the population as a whole.

Figure 1: Spatial distribution of each of the 7525 photo identifications (red dots) with the 821

uniquely-identified beluga from the St. Lawrence Estuary, Canada (red square in the inset

map) over our study period (1989—2007). The 14 sectors are outlined and labeled in white,

and cover the summer range of the population.

2.2 Multilevel Multinomial Model

The probability of seeing an individual in each delineated sector of the St. Lawrence

Estuary was estimated using a multilevel multinomial model, where the dependent variable

was the number of times an individual was captured photographically (i.e., photo-identified)

in each sector. The multilevel structure of the model allowed for the estimation of both the

mean probability of photo-identifying an individual in each sector, and the individual level

differences in this probability by using individual ID as a random intercept. If we take, as an

example, a case where the study area comprises only two sectors, then the log-odds of finding

individual i in a sector other than the reference sector can be modeled using a multilevel

multinomial following (Koster & McElreath, 2017) as:

$$log\left(\frac{p_{1,i}}{p_{r,i}}\right) = \mu_1 + v_{1i}$$
$$log\left(\frac{p_{2,i}}{p_{r,i}}\right) = \mu_2 + v_{2i}$$

Where $p_{1,i}$ is the probability of seeing beluga i in sector 1, $p_{r,i}$ is the probability of

seeing beluga i in the reference sector, μ_1 and μ_2 are the intercepts, i.e., the mean probability

of seeing a beluga in sectors 1 and 2, and v_{1i} and v_{2i} are the estimated individual differences

(i.e., random intercepts) from the mean probability of capture in sectors 1 and 2, respectively.

The mean probabilities μ_1 and μ_2 represent preference/avoidance of the specified sector,

while v_{1i} and v_{2i} are the sector-specific individual deviations from the mean probability of

capture. As a result, it is possible to model the covariance of the individual differences

between two sectors using a multivariate normal distribution:

$$\begin{bmatrix} v_{1i} \\ v_{2i} \end{bmatrix} \sim Multi\ Normal(0, \Omega_v)$$

$$\Omega_v = \begin{bmatrix} \sigma_{1,1} & \sigma_{1,2} \\ \sigma_{2,1} & \sigma_{2,2} \end{bmatrix}$$

This multivariate normal distribution has a mean of 0 and a covariance matrix Ω_v.

Here the diagonal entries in the covariance matrix ($\sigma_{1,1}$ and $\sigma_{2,2}$) represent the magnitude of

individual differences within a sector. This magnitude of individual differences identifies

whether there are individual differences in the probability of being seen in a sector (i.e., high

values of $\sigma_{1,1}$ and $\sigma_{2,2}$), or whether all individuals are equally likely to be seen (i.e., low

values of $\sigma_{1,1}$ and $\sigma_{2,2}$). The off-diagonal entries ($\sigma_{2,1}$ and $\sigma_{1,2}$) are the covariance estimates

between sectors, i.e., identifying sectors that share similar user profiles. By converting

covariance of individual differences between sectors to correlations, this multilevel modeling

approach quantifies how much information individual differences in the probability of being

seen in one sector can provide about another sector. Positive correlations suggest that the

high/low users in one sector are similarly high/low users in another sector, while negative

correlations suggest high/low users in one sector are the low/high users in another sector.

This model can be fit with a Bayesian approach (brms package in the R environment)

using a multivariate syntax: bf(y | trials(n) ~ 1 + (1|q|ID)) + multinomial(). Here, y is a set of

column vectors where each column is a sector and each row is an individual. The values in

this column vector indicate how many times each individual was seen in each sector. While n

is the total number of times an individual was captured photographically, q represents an

arbitrary character choice that allows correlations between the estimates of random intercepts

for each sector (Bürkner, 2017).

2.2.1 Addition of a common reference sector

We added a preset reference category to the data (i.e., a fifteenth sector used as a

reference sector) to standardize estimates of relative deviations from mean capture

probability between individuals. To ensure that all individuals had the same baseline

probability of being seen in the reference sector, we added the total number of captures of

each individual to the new reference sector. Doing so set the probability of a capture in the

reference sector to 0.5 for all individuals, i.e., equal to the probability of being captured

outside of it. As the parameters in the multinomial model measured deviations away from the

reference sector, we obtained standardized estimates of the relative deviations between

individuals and more accurate spatial networks (Fig. S1).

2.2.2 Dealing with biases in capture-recapture datasets

This multilevel multinomial model approach accounted for repeated sampling of

individuals, and provided an estimate of whether some individuals were seen more or less

often than the mean probability of capture in each sector (i.e. v_{1i} and v_{2i}). Unlike the mean

probabilities μ_1 and μ_2 that represent preference/avoidance of a specified sector, these

estimates of individual differences from the mean probability of capture were not impacted

by differential sampling among sectors. This was not the case for estimates of the mean

probability of capture for each sector (i.e. μ_1 and μ_2), which were expected to increase in

highly sampled sectors. For example, the over-sampling of the Saguenay River compared to

other sectors (SAG in Fig. 1) increased the mean probability of capturing individuals in that

sector. However, over-sampling of this sector was unlikely to affect the relative probability of

being captured among individuals given that all individuals' chances of being captured were

likely to go up or down equally.

Similarly, potential biases due to ease of recognition, e.g., some individuals or age

classes might bear more distinctive markings than others, are minimized using a multilevel

multinomial approach as it focuses mainly on differences in the probability of being seen

between sectors. For example, if juveniles are five times less likely to be successfully photo-

identified than adults, then they might be less often represented in the photo-ID database

compared to other age classes. However, the difference in distribution of these fewer photo-

identified juveniles across sectors is unlikely to be impacted. For instance, if we successfully

photo identified all adults 15 times and all juveniles 3 times, and if both spent twice as much

of their time in the Saguenay River compared to all other sectors, then the photo-ID

distribution (seen in versus outside of the Saguenay River) for adults and juveniles would be

expected to be 10:5 and 2:1 respectively. In this example, the capturability varies by age class

but in both age classes the probability of being captured in the Saguenay River would be

twice that of the remaining sector. The adaptive partial pooling properties of multilevel

models, however, leads individuals with few photo-IDs, and thus which contain less

information, to be less likely to show measurable deviations from the mean probability of

being captured. This means that if an age- or sex-class has very little chance of being

identified by photo-ID (e.g., newborn calves or very young individuals), then they are likely

to contribute less to the estimated spatial structures estimated by the multilevel multinomial

approach.

By using a multilevel modeling approach, we also reduced the chance of false

positives when making comparisons between many different individuals in many different

sectors (i.e., problem of multiple comparisons). For example, if we were to estimate the

differences in the probability of being seen in each sector separately for each individual, the

risk of false positives, i.e., detecting differences where there is none, would be increased.

Instead, if a multilevel approach is used to estimate the differences in probability of being

seen it is possible to make effective use of partial pooling to reduce extreme values,

especially in cases where the number of recaptures is not equal between individuals. Finally,

by running this analysis in a Bayesian framework, we were able to place priors on the

individual differences within sectors. In our case, the model was initiated assuming that there

were no differences between individuals in their use of each sector, i.e., student_t(3,0,1).

2.3 Network analysis

Social networks are often used when visualizing and quantifying social structures

within populations, with individuals often represented as nodes and their interactions as edges

between these nodes (Croft et al., 2008; Farine & Whitehead, 2015). In our case, we used

sectors as nodes, and the similarities in user profiles between sectors as edges (i.e., $\sigma_{1,2}$ and

$\sigma_{2,1}$). The correlations between sectors estimated from the multilevel multinomial model can

be used to create a network where the posterior predictions of each correlation parameter

corresponds to an edge weight in the network. In this way, each edge has a posterior

distribution and can be used to create many networks from which a distribution of network

metrics can be generated, e.g., the distribution of node strength values can be calculated for

each sector. The advantage of having distributions of network measures is that the measures

can be readily compared, e.g., does one sector have a higher node strength than another? It is

also possible to use the distribution of edge weights, and a chosen threshold (e.g., 95%

credible interval), to highlight only the edges where the sign of the correlation is known with

a particular range of certainty. In this paper, we used this latter approach to generate a signed

network (i.e., a network with positive and negative edges), and used a simple signed-edge

rule to define network communities: where a distinct community was a set of nodes that

shared positive edges but no negative edges. We also made use of signed blockmodeling, an

algorithm that can also be used to identify blocks of nodes, and that maximized within block

positive edges and minimized within block negative edges (Doreian & Mrvar, 2015). While

the signed-edge rule generally provides relatively intuitive results with simple networks,

signed blockmodeling is likely to be particularly advantageous when dealing with more

complex networks. The network communities detected using these two algorithms where

used to identify population spatial structures.

2.4 Testing the multilevel multinomial modelling approach

The accuracy of the multilevel multinomial modelling approach was assessed by

generating test datasets from the observed photo-ID data. We ensured that the test datasets

contained the same number of unique individuals, distribution of sightings (i.e., some

individuals were seen more than others), and overall number of photo-IDs as the observed

dataset. We, however, varied the spatial location of individual photo-ID captures in two

ways. First, to test if the proposed method correctly detected no pattern when none existed,

we created a completely random test dataset by permuting the sector associated with each

photo-ID in the observed dataset. The expected result was to find no correlations between

sectors, given that the sectors for each photo-ID had been randomly permuted. To then test

whether the proposed method could also correctly identify patterns when a known pattern

existed, we generated a structured test dataset by randomly assigning each uniquely identified

individual to four equally populated spatial clusters with the following and hypothetical home

range of adjacent sectors: cluster 1-BSM, SAG, CTN, cluster 2- CTN, CTO, AMN, cluster 3-

AVO, AVS, AVN, and cluster 4-AME, CTS, CTE. Following this, we altered the sector of

where the individual photo-IDs were taken so as to fall within sectors associated with an

individual's clusters, i.e., one of their home range sectors. We did this by choosing a sector

for each photo-ID based on the individual's assigned clusters 80% of the time; a random

sector was chosen for the other 20% of the time, introducing noise in the assignment of

sectors. We then tested whether the model correctly identified the correlations between

sectors that defined the home range of each clusters.

3. Results

3.1 Testing the multilevel multinomial modelling approach

When the multilevel multinomial model was fit to the data with sectors randomly

permuted between all photo-IDs, the model found as expected no evidence for

positive/negative correlations between sectors (Fig. 2a), and when we artificially created

spatially distinct clusters, the model accurately estimated the correlations between sectors that

defined these artificial population spatial structures (Fig. 2b). The simple signed-edge rule

and blockmodeling algorithm applied to the simulated datasets both revealed the four

artificially generated spatial clusters, though the blockmodeling algorithm had difficulty with

the multi-membership node as it could not assign a node to two blocks (i.e., the CTN node

that was shared between cluster 1 and 2).

Figure 2: Similarity and dissimilarity between sectors in the simulated datasets: a) randomly

permuted data, where there is no population spatial structure, and b) spatially structured data,

where there are four distinct clusters within the population. In b) the simulated clusters are

represented by color codes for each of their sectors (Note: CTN is part of the orange and

yellow clusters). The green edges (lines) between two sectors signify that the sectors share

high/low users, while red edges (lines) signify that they have dissimilar high/low users. The

lack of an edge signifies that the high/low users of one sector does not provide information

about the high/low users of other sectors.

3.2 Quantifying individual variation in habitat use among sectors from observed data

The multilevel multinomial model approach applied to the 18 years of observed

photo-ID data indicated differences between high and low users in all sectors, though the

magnitude of these individual differences varied between them (Table 1). The model also

found that these individual differences were correlated between sectors (Table S2), indicating

a high magnitude of similarity/dissimilarity between sectors in terms of which beluga used

those sectors heavily or rarely. Taking two sectors as examples, e.g., the SAG and CTE

sectors, the top 10 estimated high users of the SAG (i.e., individuals with a relatively high

probability of being found there, blue dots in Fig 3a), were low users of the CTE sector (blue

dots in Fig 3b).

Figure 3: Estimate of the relative use of the SAG (a) and CTE sectors (b) by each photo-

identified individual (i.e., deviation from mean use, v_{SAGi} and v_{CTEi}). The values are

deviations (black points) from the mean probability of recapturing individuals within a sector

(red dashed line) and are on a logit scale. The horizontal grey lines represent the 95%

credible interval. The estimated top 10 users of the SAG sector are represented by blue dots

(panel a), and those same individuals are also highlighted in blue in the CTE sector (panel b),

illustrating how correlations between sectors were estimated.

Our model indicated that the CTN sector was relatively uniformly used by all

individuals (i.e., low "sd" value; Table 1) compared to other sectors. In contrast, individual

differences in usage were the largest in the AME sector, with some very high/low users of

that sector (Table 1).

Table 1: Parameter estimates from the multilevel multinomial model predicting the

probability of capturing an individual by sector. Estimated magnitudes of within sector

individual differences in usage (sd; e.g., $\sigma_{1,1}$) are presented for each sector. Higher estimates

indicate higher contrast between high users and low users of that sector, whereas lower

estimates indicate a greater homogeneity in usage. To facilitate interpretation we have

ordered the table by lowest to highest estimates of individual differences in usage, and

provide the lower and upper 95% credible intervals for each estimate (e.g., l-95% CI, u-

95%CI). As the number of parameters in the model is large, the overall mean by sector (i.e.

μ_i), and estimated correlations between individual differences (e.g., $\sigma_{2,1}$) are presented in the

supplementary section (Table S2).

Parameter	Estimate	SD	l-95% CI	u-95% CI
sd(mu_CTN)	0.33	0.03	0.27	0.39
sd(mu_AMO)	0.65	0.38	0.05	1.41
sd(mu_AMN)	0.76	0.29	0.14	1.31
sd(mu_AVO)	0.78	0.05	0.68	0.88
sd(mu_CTE)	0.78	0.05	0.68	0.89
sd(mu_AVS)	0.83	0.07	0.70	0.97
sd(mu_AVE)	1.16	0.20	0.78	1.57

sd(muy_SAG)	1.16	0.08	1.01	1.31
sd(mu_CTS)	1.18	0.11	0.97	1.40
sd(mu_CTO)	1.27	0.14	1.00	1.55
sd(mu_AVN)	1.30	0.17	0.98	1.65
sd(mu_BSM)	1.30	0.11	1.09	1.52
sd(mu_AMS)	1.53	0.22	1.10	1.98
sd(mu_AME)	2.00	0.19	1.64	2.38

3.3. Characterizing the population spatial network

Spatial patterns emerged from using the between sector correlations to generate a

signed network overlaid on top of the sectors in the St. Lawrence Estuary. Applying the

simple-signed rule and the blockmodeling algorithm to delineate network communities, both

indicate that there are three distinct spatial clusters of individuals within the beluga summer

range: the lower St. Lawrence Estuary (AVO, AVS, AVE, AVN), the Saguenay River and

mouth (BSM, SAG, CTN), and the upper Estuary and eastern portion of the lower St.

Lawrence Estuary (CTE, CTS, CTO, AME, AMS) (Fig. 4). In the case of AVS, however, the

simple sign-rule suggested multi-membership for this sector, while the blockmodeling

algorithm found AVS to be either: 1) part of the cluster containing (AVO, AVN, AVE), or 2)

that the two clusters (orange and purple in fig. 4) merged into one depending on the choice of

weighting parameter (i.e., emphasizing positive or negative edges).

Figure 4: Population spatial structure characterized by similarity and dissimilarity in user

profiles between sectors in the St. Lawrence Estuary beluga population. The green edges

between two sectors signify that the sectors share high/low users, while red edges signify that

they have dissimilar high/low users. The lack of an edge signifies that the high/low users of

one sector does not provide information about the high/low users of other sectors. Nodes

represent sectors, and are coloured based the cluster they belong to: i.e., shared green edges,

and no shared red edges. Node sizes represent the magnitudes of individual differences in use

within the sector, i.e., larger nodes suggest specialized use by a subset of the population.

4. Discussion

Here we've shown that using capture-recapture data with a multilevel multinomial modelling approach, it is possible to estimate spatial networks that can identify spatial structures within populations while controlling for sampling biases. Applying this method to data from beluga in the St. Lawrence Estuary suggests a non-random habitat use within the summer range of this population.

In particular, the use of the multilevel multinomial model provided information about within sector usage patterns of individual beluga. Our results showed that some sectors were predominantly used by a subset of individuals, while other sectors were used more uniformly by all individuals in the population. The CTN sector for instance appeared as a potential high mixing zone for the population, whereas the AME sector seemed to be used by a specific subset of the population. These findings align well with a recent study on habitat connectivity in this population exploiting a different dataset, and suggesting that the CTN sector interconnects strongly with other sectors (Ouellet et al., 2021).

The use of the multilevel multinomial model also provided information about the similarity in usage profiles across sectors. Our results across sectors add to the evidence that the beluga population cannot be assumed to be randomly mixing within its summer habitat, suggesting instead the existence of multiple spatial clusters of individuals that make use of particular sectors of the St. Lawrence Estuary and the Saguenay River. This result suggests, for example, that individuals that are repeatedly seen in the SAG sector, are also repeatedly seen in the BSM and CTN sectors, but are seen very little in the CTE and CTS sectors (Fig. 4). Given that these habitat-use patterns are constructed with all age and sex categories of beluga, and that beluga are known to segregate by age- and sex-classes during at least the summer period in the SLE and elsewhere (Loseto et al., 2006; Michaud, 1993; Smith et al.,

1994), more work will be required to quantify additional population spatial structures within

age/sex categories. This is crucial given that juveniles and adult females tend to show less

distinctive markings compared to adult males, making captures by photo ID more difficult for

this age/sex classes and reducing the amount of information they can provide when

estimating population spatial networks. Ongoing photogrammetric and machine learning

work to identify sex, estimate age, and facilitate individual identification will allow a more

sex/age specific estimate of habitat use in this population.

Our findings about SLE beluga being spatially structured in their habitat use during

summer have direct implications for estimating the impacts of anthropogenic stressors on this

population. Local stressors can have a disproportionate impact on particular subsets of the

larger population. Properly accounting for animal movements and population spatial structure

can lead to drastically different results about impacts of individual stressors or their

cumulative effects. For instance, predictions from an agent-based model of beluga and marine

traffic in the STE, found that if beluga spend more time within the Saguenay River Sector

they likely experience reduced exposure to noise pollution (Chion et al. Under Review). This

suggests that subsets of individuals within the larger population that use this sector will have

reduced noise exposure. This refuge effect, however, is predicted to be lost under scenarios

where additional marine traffic is added to the Saguenay River (Chion et al. in review).

When implementing the multilevel multinomial model on other capture-recapture

datasets, the use of test datasets should hold a prominent role in the analysis. The use of

permutation/randomization methods to both generate spatially structured and unstructured

datasets, while maintaining the sample size distribution of the original datasets, can be very

valuable in helping to set model priors and to interpret the final model results. The use of

permutation approaches is common in social network analysis (Croft et al., 2011; Farine,

2017), and is becoming more common in statistical workflows more generally (Gelman et al.,

2013; McElreath, 2020).

5. Conclusions

We have introduced the use of multilevel multinomial modeling to estimate spatial

networks from a capture-recapture approach that is gaining in applicability, i.e., photo-ID

data. We've shown, using test datasets, that the proposed method is effective at detecting

population spatial structures. When applied to 18 years of photo-ID data from an endangered

population of beluga in the St. Lawrence Estuary, our results provide evidence that the

population is composed of multiple spatial cluster of individuals with distinct habitat-use

patterns. We suggest that multilevel multinomial models can be effective at extracting spatial

structuring within animal populations monitored by capture-recapture sampling, contributing

to impact assessments with direct implications for conservation and management.

6. Acknowledgment

428 We would like to thank all those who worked with and supported the GREMM in

429 collecting these long term data.

7. Funding

432 Funding was provided by Ministère des Forêts, de la Faune et des Parcs du Québec and

433 Secrétariat à la stratégie maritime du Québec.

8. Data accessibility

436 The permuted and simulated photo ID datasets are available on github

437 (github.com/tbonne/photoID_multinomial), along with code used in the analysis.

9. Author Contributions

RM collected the data; TRB conceived the analytical methodology and performed the analysis; all authors contributed critically to the drafts and gave final approval for publication.

References:

Croft, D. P., James, R., & Krause, J. (2008). *Exploring Animal Social Networks*. Princeton

University Press.

Croft, D. P., Madden, J. R., Franks, D. W., & James, R. (2011). Hypothesis testing in animal

social networks. *Trends in Ecology & Evolution*, *26*(10), 502–507.

DeFur, P. L., Evans, G. W., Hubal, E. A. C., Kyle, A. D., Morello-Frosch, R. A., & Williams,

D. R. (2007). Vulnerability as a function of individual and group resources in

cumulative risk assessment. *Environmental Health Perspectives*, *115*(5), 817–824.

Doreian, P., & Mrvar, A. (2015). Structural balance and signed international relations.

Journal of Social Structure, *16*, 1.

Evans, P. G., & Hammond, P. S. (2004). Monitoring cetaceans in European waters. *Mammal*

Review, *34*(1–2), 131–156.

Farine, D. R. (2017). A guide to null models for animal social network analysis. *Methods in*

Ecology and Evolution, *8*(10), 1309–1320.

Farine, D. R., & Whitehead, H. (2015). Constructing, conducting and interpreting animal

social network analysis. *Journal of Animal Ecology*, *84*(5), 1144–1163.

https://doi.org/10.1111/1365-2656.12418

Gelman, A., Carlin, J. B., Stern, H. S., Dunson, D. B., Vehtari, A., & Rubin, D. B. (2013).

Bayesian data analysis. CRC press.

Gimenez, O., Mansilla, L., Klaich, M. J., Coscarella, M. A., Pedraza, S. N., & Crespo, E. A.

(2019). Inferring animal social networks with imperfect detection. *Ecological*

Modelling, *401*, 69–74.

Hupman, K., Stockin, K. A., Pollock, K., Pawley, M. D. M., Dwyer, S. L., Lea, C., & Tezanos-

Pinto, G. (2018). Challenges of implementing Mark-recapture studies on poorly

marked gregarious delphinids. *PLOS ONE*, *13*(7), e0198167.

https://doi.org/10.1371/journal.pone.0198167

Koivuniemi, M., Auttila, M., Niemi, M., Levänen, R., & Kunnasranta, M. (2016). Photo-ID as a

tool for studying and monitoring the endangered Saimaa ringed seal. *Endangered*

Species Research, *30*, 29–36.

Koster, J., & McElreath, R. (2017). Multinomial analysis of behavior: Statistical methods.

Behavioral Ecology and Sociobiology, *71*(9), 138.

Lesage, V. (In Press). The challenges of a small population exposed to multiple

anthropogenic stressors and a changing climate: The St. Lawrence Estuary beluga.

Polar Research.

Loseto, L. L., Richard, P., Stern, G. A., Orr, J., & Ferguson, S. H. (2006). Segregation of

Beaufort Sea beluga whales during the open-water season. *Canadian Journal of*

Zoology, *84*(12), 1743–1751.

McElreath, R. (2020). *Statistical rethinking: A Bayesian course with examples in R and Stan*.

CRC press.

Michaud, R. (1993). *Distribution estivale du béluga du Saint-Laurent: Synthèse 1986 à 1992*.

Ministère des pêches et des océans, Direction de la gestion des pêches et de ….

Michaud, R. (2014). *St. Lawrence Estuary beluga (Delphinapterus leucas) population*

parameters based on photo-identification surveys, 1989-2012. DFO Can. Sci. Advis.

Sec. Res. Doc. 2013/130. iv + 27 p.

Ouellet, J.-F., Michaud, R., Moisan, M., & Lesage, V. (2021). Estimating the proportion of a

beluga population using specific areas from connectivity patterns and abundance

indices. *Ecosphere*, *00(00)*(e03560). https://doi.org/10.1002/ecs2.3560

Perryman, R. J. Y., Venables, S. K., Tapilatu, R. F., Marshall, A. D., Brown, C., & Franks, D.

W. (2019). Social preferences and network structure in a population of reef manta

rays. *Behavioral Ecology and Sociobiology*, *73*(8), 114.

https://doi.org/10.1007/s00265-019-2720-x

Schilds, A., Mourier, J., Huveneers, C., Nazimi, L., Fox, A., & Leu, S. T. (2019). Evidence for

non-random co-occurrences in a white shark aggregation. *Behavioral Ecology and*

Sociobiology, *73*(10), 138. https://doi.org/10.1007/s00265-019-2745-1

Schneider, S., Taylor, G. W., Linquist, S., & Kremer, S. C. (2019). Past, present and future

approaches using computer vision for animal re-identification from camera trap data.

Methods in Ecology and Evolution, *10*(4), 461–470. https://doi.org/10.1111/2041-

210X.13133

Silk, M. J., McDonald, R. A., Delahay, R. J., Padfield, D., & Hodgson, D. J. (2021). CMRnet:

An r package to derive networks of social interactions and movement from mark–

recapture data. *Methods in Ecology and Evolution*, *12*(1), 70–75.

https://doi.org/10.1111/2041-210X.13502

Smith, T. G., Hammill, M. O., & Martin, A. R. (1994). Herd composition and behaviour of

white whales (Delphinapterus leucas) in two Canadian arctic estuaries. *Studies of*

White Whales (Delphinapterus Leucas) and Narwhals (Monodon Monoceros) in

Greenland and Adjacent Waters, *39*, 175–184.

Urian, K., Gorgone, A., Read, A., Balmer, B., Wells, R. S., Berggren, P., Durban, J., Eguchi,

T., Rayment, W., & Hammond, P. S. (2015). Recommendations for photo-

identification methods used in capture-recapture models with cetaceans. *Marine*

Mammal Science, *31*(1), 298–321.

Whitehead, H. (2008). *Analyzing animal societies: Quantitative methods for vertebrate social*

analysis. University of Chicago Press.

517