# Passive sampling of environmental DNA in aquatic environments using 3D-printed hydroxyapatite samplers

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# 32 Abstract

The study of environmental DNA released by aquatic organisms in their habitat offers a
 fast, non-invasive and sensitive approach to monitor their presence. Common eDNA
 sampling methods such as filtration and precipitation are time consuming, require
 human intervention and are not applicable to a wide range of habitats such as turbid
 waters and poorly-accessible environments. To circumvent these limitations, we
 propose to use the binding properties of minerals to create a passive eDNA sampler.

- We have designed 3D-printed samplers made of hydroxyapatite (HAp samplers), a
   mineral known for its high binding affinity with DNA. The shape and the geometry of the
   samplers have been designed to facilitate their handling in laboratory and field. Here
   we describe and test the ability of HAp samplers to recover artificial DNA and eDNA.
- We show that HAp samplers efficiently recover DNA and are effective even on small
   amounts of eDNA (<1 ng). However, we also observed large variations in the amount</li>
   of DNA recovered even under controlled conditions.
- 46 4. By better understanding the physico-chemical interactions between DNA and the HAp
   47 sampler surface, one could improve the repeatability of the sampling process and
   48 provide an easy-to-use eDNA sampling tool for aquatic environments.
- 49
- 50 Key-words : DNA Binding, Environmental DNA, Hydroxyapatite, Passive sampling, 3D51 printing
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# 55 1 | INTRODUCTION

At a time of unprecedented threats on freshwater biodiversity, it is crucial to develop rapid, 56 accurate and minimally invasive tools to monitor aquatic ecosystems. About a decade ago, 57 58 methods based on the sampling of environmental DNA (eDNA) were proposed as a 59 revolutionary way to survey aquatic macro-organisms (Deiner et al., 2017). Macro-organisms 60 release DNA in their environment through different processes (e.g. faeces, excretion, 61 shedding cells, gametes) and this eDNA can take different forms (tissues, cells, organites, 62 nucleo-proteic complexes, ...). The direct sampling of eDNA coupled with molecular analysis 63 methods such as NGS (Shokralla et al., 2012) or quantitative polymerase chain reaction 64 (qPCR) (Langlois et al., 2020) allow the detection and identification of aquatic species while overcoming organism capture. Although eDNA offers many promising applications, several 65 66 methodological challenges remain.

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One of the most challenging aspects of eDNA-based approaches is the sampling method. 68 69 eDNA is present in very small quantities and is heterogeneously distributed in aquatic environments (Goldberg et al., 2016). To maximise its recovery, sampling methods must be 70 71 able to concentrate eDNA (Hinlo et al., 2017). Active filtration of a large volume of water is the 72 most commonly-used method to recover eDNA in aquatic systems. However, filtration has significant methodological limitations. Firstly, it is a long and tedious process requiring human 73 intervention, sometimes difficult to carry out in poorly-accessible habitats. Secondly, the 74 75 clogging of the filters is a recurrent problem which reduces the volume of water that can be 76 sampled (Williams, Huyvaert and Piaggio, 2017).

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To limit clogging, filtration membranes with large porosity (greater than 0.45 µm) are often 78 79 used. However, eDNA particles can be present in various forms (intra or extracellular), states (free or complexed with other particles) and sizes (from > 180 to  $< 0.2\mu$ m but most abundant 80 between 0.2 and 10 μm) (Turner et al., 2014; Moushomi et al., 2019 ; Wilcox et al., 2015). As 81 82 filtration is based on particle size sorting, the use of membranes with large porosity will 83 overlook smaller DNA particles, even though they may be an important source of eDNA 84 (Moushomi et al., 2019). Finally, given the complex dynamic of eDNA in aquatic environments 85 (i.e. pulsed emission, transport, retention, degradation), one filtration sample will provide an 86 instantaneous snapshot which is likely to be poorly integrative of the overall eDNA signals 87 (Pilliod et al., 2013; Spear et al., 2015).

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Passive eDNA sampling using natural substrates is a promising solution to overcome filtration 89 90 challenges. Organisms such as marine sponges (Mariani et al., 2019), molluscs (Der 91 Sarkissian et al., 2020) and biofilms (Rivera et al., 2021) can trap and accumulate eDNA particles in water. Minerals can also accumulate and protect DNA from enzymatic degradation 92 93 (Alvarez et al., 1998; Levy-Booth et al., 2007). Indeed, a sample of sediment can contain more eDNA than a water sample (Turner, Uy and Everhart, 2015). Recently, Kirtane and 94 colleagues (Kirtane, Atkinson and Sassoubre, 2020) have shown that montmorillonite and 95 96 coal-based mineral powders can be used as passive eDNA samplers in aquatic environments. Thanks to good DNA capture and preservation rates (up to 200 µg genomic 97 DNA / g) (Gardner and Gunsch, 2017), sediments and commercial mineral powders may very 98 99 well be more integrative eDNA substrates than filtration methods. Yet, these substrates are 100 difficult to handle and deploy in the environment, particularly in aquatic systems.

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In this study, as an alternative to filtration, we developed proof of concept 3D-printed passive 102 eDNA samplers. 3D printing allows control of the shape and composition of an object. The 103 shape of the samplers were designed with optimised surface/volume -ratio and a shape 104 easing handling in the field and in the lab. The samplers were made of pure hydroxyapatite 105 106 (HAp), a calcium phosphate mineral naturally present in bones and known for its high binding affinity toward DNA (Okazaki et al., 2001; Brundin et al., 2013). Here we describe the 107 development of hydroxyapatite samplers (HAp samplers) and test their ability to sample 108 109 eDNA in fresh waters. Two prototypes of samplers will be presented: a first test-version, with 110 which the concept and material will be tested, and a second version which shape and design 111 have been optimised for eDNA sampling. Using controlled laboratory experiments, our 112 objectives are to (i) quantify the HAp samplers DNA binding and release capacity, (ii) assess the range of DNA fragment size sampled, (iii) quantify the repeatability of DNA sampling 113 114 across several cycles of use of the HAp samplers, and (iv) evaluate the samplers capacity to 115 sample eDNA released by organisms in microcosm.

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#### 117 2 | MATERIALS AND METHODS

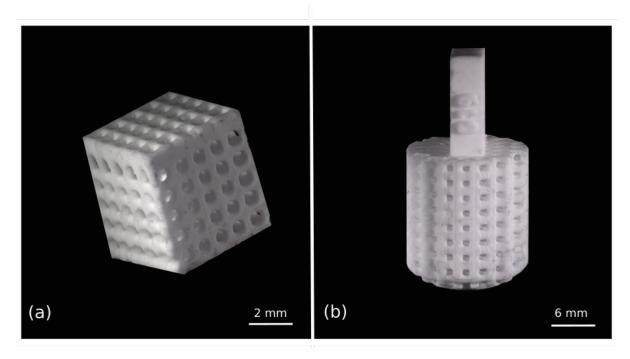
118 2.1 | 3D-printed HAp samplers design

119 2.1.1 | Raw material and printing setup

120 A photopolymerizable organic resin (3D Mix, 3DCeram Company, HAP, Bonnac-la-Côte, 121 France) containing 40-60 % (w/v) of hydroxyapatite powder ( $Ca_{10}(PO_4)_6(OH)_2$ , stoichiometric 122 hydroxyapatite), a synthetic calcium phosphate with Ca/P atomic ratio of 1.67, was the raw 123 material used to fabricate the samplers. The samplers were built from this hydroxyapatite-

enriched resin using a 3D stereolithographic printer (CERAMAKER C900, 3DCeram
Company, with 55mW laser power and 100μm layer thickness).

Two types of prototypes of HAp samplers were produced. The first prototype (P1) is a test version corresponding to 10 pieces cut out of a 3D-printed mesh (Fig. 1a). P1 prototypes have an exposed surface of 240 mm<sup>2</sup> and a macroporosity of 500  $\mu$ m in diameter. A second more elaborate prototype (P2) was then produced with optimized geometry and porosity, and printed in 25 copies (Fig. 1b). P2 has a total surface of 480 mm<sup>2</sup> and a macroporosity of 400  $\mu$ m in diameter.



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FIGURE 1: Images of 3D-printed hydroxyapatite samplers prototype P1 (a) and P2 (b) obtained with a confocal microscope (objective x0.5, LEICA Z16 APO, camera LEICA DMC5400).

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137 2.1.2 | Debinding and sintering steps

Once printed, cleaned with a specific solvent (Ceracleaner, 3DCeram Company, Bonnac-la-Côte, France) and dried, the HAp samplers underwent debinding and sintering steps. Debinding aims at removing all organic components (in particular the organic resin) and was conducted in a conventional oven following the thermal cycle described in Table 1.

142 **TABLE 1** : Process parameters for debinding HAp samplers

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Step	Temperature (°C)	Heating rate	Dwell (min)
		(°C/min)	
1	20-200	0,2	120
2	200-300	0,1	120
3	300-380	0,1	120
4	380-550	0,1	120
5	550-950	1	0
6	950-20	2	-

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Sintering aims at consolidating the samplers by densifying them (creation of necks and reduction of the porosity between the individual ceramic particles) (Rahaman, 2017), and is achieved by a thermal treatment at higher temperature (1°C/min up to 1150°C, 60 min. at 1150°C, followed by a second step at 3°C/min up to 1250°C, 60 min at 1250°C, finally cooling to room temperature at 3°C/min). After sintering, no additional processing (i.e. finishing or polishing) was performed. The presence of pure HAp was confirmed by X-ray diffraction (XRD) performed on as-sintered samples.

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153 2.2 | Expected DNA recovery from HAp samplers

We used the term "DNA recovery" to define the quantity of DNA that binds and is released 154 from the HAp samplers. We estimated the theoretical maximum DNA recovery (DNA<sub>max</sub>) 155 based on the hypothesis that a single layer of DNA molecules would bind on the HAp surface 156 of the samplers. According to equation 1, the number of DNA molecules that can bind to the 157 surface is obtained by dividing the exposed surface (Se) of a sampler (P1 = 240 mm<sup>2</sup>, P2 = 158 480 mm<sup>2</sup>) by the surface of a DNA base pair (Sd = 6.46E-10 mm<sup>2</sup>). The surface of a DNA 159 base pair was calculated according to Mandelkern et al (1981) (diameter = 2 nm, length = 3.4 160 nm). The number of DNA molecules per sampler is then divided by Avogadro's constant (NA = 161 162  $6.02214076 \times 1023$  mol - 1) to give the number of DNA moles per sampler. The number of 163 moles of DNA is then divided by the molar mass of a DNA base pair (W = 650 daltons) to 164 obtain the total mass of DNA that can bind to a sampler. Being smaller, P1 has a maximum 165 theoretical recovery capacity of 400 ng of DNA per sampler, while P2 has a capacity of 800 166 ng.

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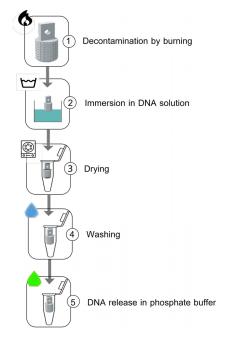
DNAmax = (St/Sd)/NAxW Equation 1

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#### 169 2.3 | Protocol of DNA binding and release

The HAp sampler DNA binding and release protocol is composed of 5 steps. First, HAp samplers are decontaminated before each experiment by a thermal treatment in air at 550 ° C for 3 hours (Thermolyne model 30400 furnace), a procedure typically used to decontaminate glassware. Second, DNA is bound to the HAp samplers by immersing them in an aqueous solution (varying composition upon the present study) containing DNA. Third, samplers are transferred to Eppendorf tubes and centrifuged for 1 minute at 3000 rpm to dry them. Fourth, samplers are washed with 1 mL of sterile ultrapure water. Finally, DNA is released from the

- 177 samplers by immersing them in 1 mL of 0.1 M phosphate buffer pH 8, vortexed for 30
- 178 seconds and incubated at room temperature for 1 hour.



179

180 **FIGURE 2** : DNA binding and release protocol

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182 2.4 | DNA sampling experiments by HAp samplers

183 2.4.1 | Experiment 1: DNA binding and release

A DNA sampling experiment with concentrated DNA fragments of various sizes (i.e. a DNA 184 size marker) was conducted to validate the binding and release protocol and to assess 185 whether DNA fragments of different sizes have different binding efficiencies. After 186 decontamination, one batch of HAp samplers (P1 and P2) was incubated in tubes (1 sampler/ 187 tube) containing 2 mL of a solution of large DNA fragments at 5000 ng / mL (λ DNA / BstEII 188 Digest, 117-8450 bp). A second batch of HAp samplers was incubated in tubes containing 2 189 mL of a solution of short DNA fragments at 2000 ng / mL (PCR 20 bp Low Ladder, 20-2000 190 191 bp). Both batches were incubated for 17 hours on a rotary shaker (IKA Roller 6 Digital, 40

192 rpm). Controls were tubes with 2 mL of solution and devoided of samplers. Residual DNA in 193 the supernatants was quantified by taking 60 µl aliquots of the supernatant after 45 min and 194 17 h of incubation. After incubation, HAp samplers were removed from the DNA solutions 195 using sterile clamps and DNA was released according to the protocol in section 2.3. All 196 supernatants aliquots and released DNA solutions were stored at -20°C prior to analysis.

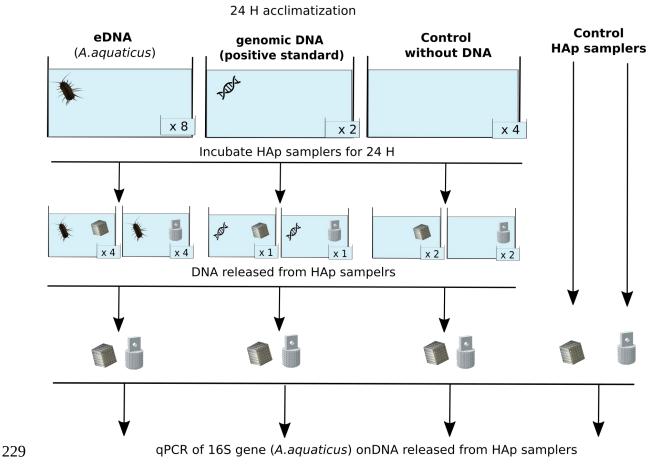
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- 198 2.4.2 | Experiment 2: repeatability

199 A quantification of repeatability was conducted to test whether HAp samplers can be reused 200 after several cycles of use. A cycle of use is defined here as a thermic treatment phase 201 followed by a DNA binding and release phase. For this purpose, 5 prototypes 1 and 25 202 prototypes 2 HAp samplers were incubated in a concentrated solution of DNA size marker 203 (λDNA/BstEII Digest 117-8450 pb) at a concentration of 2.8 µg/mL on a rotary shaker (Roller 10 Digital IKA) for 17 H. This experiment was carried out three times in a row (hereafter called 204 205 experiments A, B and C) under strictly identical conditions, at room temperature ( $24^{\circ}C \pm 2^{\circ}C$ ) 206 with decontamination through thermic treatment between each use. After incubation, HAp samplers were removed from the DNA solution with sterile clamp, washed and DNA was 207 released with 1 mL of 0.1 M phosphate buffer pH 8 according to the protocol section 2.3 DNA 208 samples were stored at -20°C prior to analysis. 209

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- 211 2.4.3 | Experiment 3: microcosm experiment

Asellus aquaticus, a small freshwater isopod, was used as a target organism to test the capacity of the HAp samplers to collect eDNA. 40 organisms of *A.aquaticus* sampled from a natural pond (Lyon, France) in april 2019 were divided into 8 glass microcosms (5 individuals / microcosm) containing 500 mL of synthetic water (Peltier and Weber, 1985). Positive controls

216 correspond to microcosms where we injected genomic DNA (final microcosm at 1 ng/mL) extracted from a pool of 10 A.aquaticus. Negative controls were of two types: control 217 218 microcosms containing water without DNA and a sampler, and control samplers from which 219 the DNA was released just after the decontamination step (i.e. without incubation in DNA solution). After 24 hours of A. aquaticus acclimatization, the two prototypes of HAp samplers 220 221 were incubated in microcosms (1 sampler / microcosm) for 24 hours. All microcosms were 222 placed in a cold room at 18°C, spaced 0.5 m apart and covered to limit the risk of 223 contamination. The organisms were not fed during the experiment to reduce the amount of 224 allochthonous DNA. After incubation, the HAp samplers were collected with sterile clamps 225 and the DNA was released according to section 2.3 of the protocol. Released DNA was purified (Macherey-Nagel <sup>™</sup> NucleoSpin <sup>™</sup> gel and PCR cleaning kit) to avoid potential 226 227 inhibition of the downstream qPCR by the phosphate buffer (see next section), following the manufacturer's recommendations. Purified eDNA was stored at -20 ° C prior to analysis. 228



230

- FIGURE 3 : Experimental design testing HAp samplers efficacy to recover eDNA from Asellus
- 232 aquaticus in microcosms.

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- 234 2.5 | DNA quantification and analysis
- 235 2.5.1 | Quantification of DNA size marker

In the first experiment, DNA binding and release by HAp samplers were evaluated by following the DNA concentration and fragment sizes in three compartments: (1) in the supernatant (i.e. residual DNA), (2) in the washing solution, and (3) in the releasing solution (Fig. 4). DNA was quantified using a QuBit  $\circledast$  3.0 fluorometer (Invitrogen) with the dsDNA BR kit (broad range, 2 to 1000 ng/µL) according to the manufacturer's protocol. The binding of

large DNA fragments (117-8450 pb) was evaluated using gel electrophoresis (1.3% agarose),
and the binding of small fragments (35-2000 bp) using a 2100 Bioanalyzer with an Agilent
high-sensitivity DNA chip (Agilent Technologies).

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For the second experiment (repeatability), the amount of DNA released from the HAp 245 246 samplers was measured by fluorescence (excitation at 480 nm and emission at 520 nm) 247 using an Infinite M200 Pro microplate fluorometer (TECAN, Switzerland). A QuantiFluor® dsDNA kit was used according to the manufacturer's protocol, with a DNA sample volume of 248 249 10  $\mu$ L and 190  $\mu$ L of working solution. A five-fold dilution series (1500-0 ng/ $\mu$ L) of standard 250 DNA (Lambda DNA Standard, 100µg/mL) was used to build the standard curve and calculate 251 the sample DNA concentration in  $\mu g/\mu L$ . The results are reported in percentage of recovered 252 DNA (i.e. DNA bound and released).

253 *DNAr* is the measured concentration of released DNA ( $\mu$ g/mL) per HAp sampler and DNAtot 254 is the initial DNA concentration added in each tube (DNAtot = 2.8  $\mu$ g/mL).

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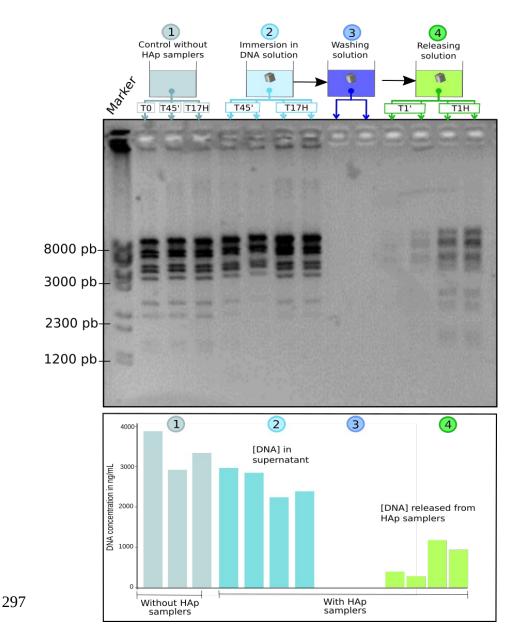
256 2.5.2 | Quantitative PCR assay for eDNA from *A.aquaticus* 

For the third experiment (microcosm experiment), quantitative PCR (qPCR) was used to 257 quantify the amount of A.aquaticus eDNA recovered by the samplers. We designed a pair of 258 259 primers to specifically amplify a 110 bp fragment of the mitochondrial 16S gene of A. aquaticus (5' GGTTTAAATGGCTGCAGTATCC 3', 5' CTTGTGTAATAAAAAGCCTACCTC 3'). 260 The amplification specificity of the primers was tested in silico using primer-BLAST (NCBI) 261 and assessed experimentally through PCR and electrophoresis gel analysis. The qPCR 262 263 reaction volume was 20 µL consisting of 1X SsoAdvanced Universal SYBR Green Supermix 264 (Bio-Rad Laboratories Inc., Hercules, CA), 0.5  $\mu$ M of primers and 2  $\mu$ L of DNA released from

265	samplers. All gPCRs assays were run in duplicate in 96 wells plate on a CFX96 Touch™ Real-
266	Time PCR Detection System (Bio-Rad Laboratories, Inc., Hercules, CA). qPCR cycle started
267	with an incubation at 95 °C for 10 min followed by 45 cycles of denaturation at 95 °C for 10
268	sec and an annealing/extension step at 60 °C for 20 sec before a final melt curve from 65-95
269	°C (0.5 °C increments). Each qPCR plate included a five-fold dilution series of the genomic
270	DNA at a concentration between 0 and 2.5 ng/ $\mu$ L quantified by a QuBit 3.0 assay.
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277	2.6   Statistical analysis
277 278	2.6   Statistical analysis Linear mixed-effect models (LMMs) were used to test the influence of the prototype version
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288 8450 bp) or short DNA fragments (35 bp to 2000 bp). In the solution containing long DNA

289 fragments, quantification of DNA concentration (Fig. 4, bottom panel) shows a clear reduction 290 in DNA concentration in the supernatant after 17 hours of exposure to the HAp samplers for both sampler replicates. Once immersed in the releasing solution, and after only 1 minute, the 291 292 HAp samplers started to release DNA. The amount of released DNA then tripled after 1 hour 293 of incubation. By examining the DNA band profiles in the supernatants and in the releasing solution, we found that P1 bound all DNA fragment sizes from 2000 to 8450 bp. Fragments 294 below 2000 bp were not visible on the electrophoresis gel (Fig. 4, top panel). The same 295 observations were made on P2 (see supporting information). 296



**FIGURE 4** : DNA binding and release by two replicates of the HAp samplers prototype 1. DNA fragment size (agarose electrophoresis gel, top panel) and concentration (bottom panel) are shown in the following order (left to right): 1) in the control solution without HAp samplers after 0 minute, 45 minutes and 17 hours, 2) in the DNA solution 45 min and 17 H after addition of HAp samplers, 3) in the washing solution and 4) in the releasing solution 1 min and 1 H after immersion of the HAp samplers.

#### 304

Repeating the same experiment but using this time short DNA fragments (35-2000 pb) and a microfluidics-based automated electrophoresis system does not show an effect of fragment size on DNA binding (Fig. 5): both samplers prototypes bound and released DNA fragments ranging from 35 to 2000 bp although the resolution of the marker for fragments above 600 bp was not optimal.

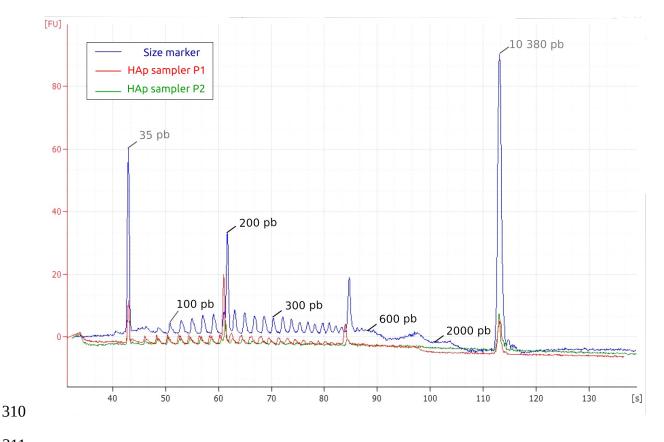


FIGURE 5 : Electropherograms of the DNA fragments bound and released by the prototype 1 (red curve) and prototype 2 (green curve). As a standard, the blue curve represents the profile of the initial DNA marker. The horizontal axis represents the migration time of DNA fragments in seconds, and the vertical axis represents fluorescence. The left-most (35 bp) and rightmost (10380 bp) peaks are internal markers.

#### 317 3.2 | Experiment 2: Reusing HAp samplers over time

A repeatability experiment was conducted to test the hypothesis that HAp samplers can be 318 319 reused and that their binding efficacy is stable after several cycles of use. We performed three 320 consecutive cycles of use (experiment A, B and C), each composed of a decontamination, DNA binding and release steps. The percentage of DNA recovered by the samplers was lower 321 322 in experiment A compared to experiments B and C, with an average of 8%, 17% and 15%, 323 respectively (Fig. 6). In the meantime, experiment A showed a disproportion of samplers (18) 324 out of 30, against 0 for experiment B and C) which failed to recover any DNA compared to the 325 other experiments (Fisher exact test, p < 1E-10). After removing the samplers which failed to 326 capture any DNA, we tested the influence of the experiment and prototype on the percentage 327 of DNA recovered using a linear mixed-effect models with experiments (A, B and C) and 328 sampler prototypes (P1 or P2) as the fixed effects, and samplers as random effect on the intercept. The experiment and the sampler prototype had no significant effect on the 329 percentage of DNA recovered (LRT, sampler prototype :  $\chi^2_{df=1}$ =4.96, p=0.08, experiment : 330  $\chi^2_{df=2}=1.16$ , p=0.28). Nonetheless, while not associated with any experiment in particular, the 331 332 percentage of DNA recovered was highly variable. The coefficient of variation of DNA 333 recovered was on average 65% considering all the samplers and 34% when excluding the samplers which failed to recover any DNA. Altogether, while we found that the samplers can 334 335 still recover DNA after several cycles of use, we also discovered that the capacity of HAp 336 samplers to recover DNA is variable and unpredictable: at some times it may not work at all. 337 while at others it may recover a large amount of DNA.

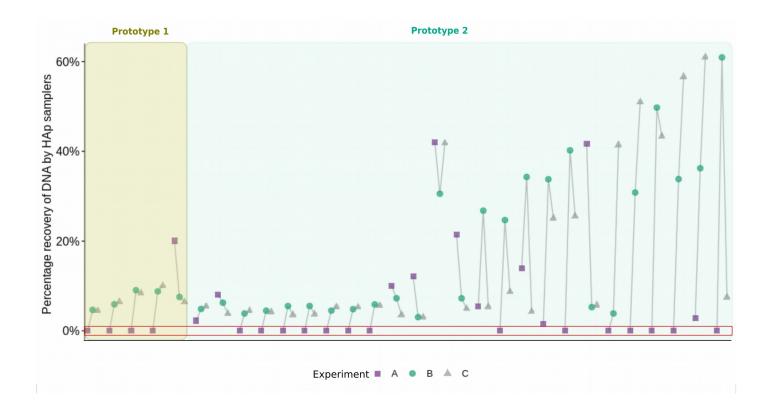


FIGURE 6 : Percentage of DNA recovered by two prototypes (P1 and P2) of HAp samplers in
three consecutive experiments (A, B and C). 5 P1 and 25 P2 samplers are sorted according
to their variance of recovered DNA. Samplers in the red box did not recover any DNA.

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### 343 3.3 | Experiment 3: Environmental DNA sampling

We deployed the HAp samplers in a microcosm containing isopods (*Asellus aquaticus*) to test their ability to recover eDNA and used qPCR to quantify the amount of *A. aquaticus* 16S gene recovered by the samplers. In a microcosm with no organisms, we observed low levels of DNA that were similar or slightly above the amount of DNA observed in control samplers that were not immersed in a microcosm (Fig. 7). This is indicative of a slight level of crosscontamination between microcosms, and allowed us to determine an amount of 16S DNA below which we cannot differentiate between a contamination and a positive result (limit of

blank, LOB). As expected, using concentrated genomic DNA as a positive control, the samplers recovered large amounts of 16S DNA molecules (Fig. 7). In the microcosm that contained isopods, the amount of 16S DNA molecules was about 3 orders of magnitude lower, with 3 samplers out of 8 below the limit of blank. Overall, the two HAp prototypes recovered *A. aquaticus* eDNA with the same efficiency (Wilcox-test, p = 0.89).

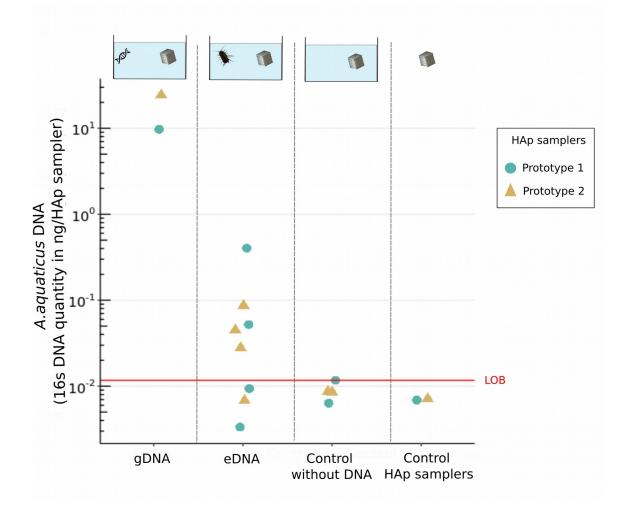


FIGURE 7 : Quantity (in log scale) of *A.aquaticus* 16S gene recovered by the two prototypes of HAp samplers (P1 = circle, P2 = triangle) after 24h of incubation in microcosm containing genomic DNA used as a positive control (gDNA) or five individuals of *A. Aquaticus* (eDNA).

361 Two types of negative controls were used: one control microcosm without DNA (control 362 without DNA), and HAp samplers without microcosm incubation (control HAp samplers). Red 363 line corresponds to the limit of blanks (LOB).

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#### 365 4 | Discussion

366 4.1 | HAp samplers recover artificial DNA and eDNA

367 Our results demonstrated that HAp samplers can passively recover artificial DNA and eDNA. 368 Using artificial DNA, DNA recovery was fast and optimal. In 17 hours, HAp samplers recovered up to 4 times more DNA (i.e. 1.75 µg) than the theoretical quantity we had 369 370 estimated using a projection of a DNA monolayer on the surface of the samplers (see 371 methods). These results confirm the high binding affinity between DNA and hydroxyapatite 372 reported in literature (Okazaki et al., 2001; Del Valle et al., 2014) and suggest that more than one layer of DNA molecules can bind to the HAp surface. To our knowledge, this is the first 373 374 time that these binding properties are tested and validated on 3D-printed objects. eDNA 375 experiments showed that HAp samplers recovered eDNA from living macro-organisms (A. aquaticus). 5 out of 8 HAp samplers allowed a positive detection of A. aquaticus after only 24 376 hours of incubation in microcosm. Given the low densities of these small isopods which, 377 unlike large organisms commonly used in eDNA microcosm experiments (e.g. fish, 378 amphibians; Maruyama et al., 2014; Jo et al., 2020), are likely to release very small amounts 379 380 of eDNA, and given the short experiment duration, this overall high rate of detection demonstrates the high sensitivity of HAp samplers to detect organisms. 381

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#### 385 4.2 | Binding of different DNA fragment size

While it was hypothesized that DNA fragment size would influence DNA binding (Ogram et al., 386 387 1994), we did not find any evidence that certain fragment sizes bind preferentially to the HAp samplers. The samplers recovered DNA fragments of various sizes (i.e. 35-8450 bp), 388 although bands below 2000 bp were not visible on the electrophoresis gel, possibly due to a 389 390 higher concentration of the larger fragments in the marker solution. However, the sensitive 391 microfluidics-based automated electrophoresis analysis showed that smaller fragments 392 (<2000 bp) were bound and released by HAp samplers. eDNA is a complex mixture of genetic 393 material ranging from cells to more or less degraded free DNA fragments (Wilcox et al., 394 2015). A sampling method that is not biased toward a given range of fragment sizes is a real 395 advantage for eDNA sampling, in particular in environments where eDNA could be rapidly 396 degraded into small free DNA fragments (Seymour et al., 2018). While free DNA binds to the HAp samplers, it remains to be tested whether other forms of eDNA such as proteo-nucleic 397 398 complexes or even larger particles can also be collected.

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#### 400 4.3 | Repeatability issues

Although HAp samplers show a great potential for DNA sampling, repeatability appears to be 401 a concerning issue. Many HAp samplers showed extreme variability in DNA recovery among 402 403 experiments carried out under strictly identical conditions (section 3.2). Given the high 404 number of samplers that did not recover any DNA during the first but not the later experiments (Fig. 6), one might have expected that DNA recovery would improve with cycle of use. 405 However, no effect of cycle of use or sampler prototype was found. In some cases, the DNA 406 407 recovery remained stable over time, in some it increased, and in other it decreased. 408 Surprisingly, while highly variable, there was not a set of samplers or one prototype in

409 particular which was less effective than the others to recover DNA. This unexplained the complexity of the binding mechanism between DNA and variability highlights 410 411 hydroxyapatite and the factor that controls it, and reinforces the necessity to better understand the evolution of the HAp surface after several DNA cycles of use. According to 412 Okazaki et al. (2001), the binding affinity is based on an electrostatic interaction between the 413 414 negative charges of the phosphate groups of DNA to the calcium ions at the surface of the 415 hydroxyapatite. This ionic interaction strongly depends on the physico-chemical properties of 416 the sampler surface and the solution in which the binding reaction takes place (Gallo et al., 417 2018). Among the surface properties, porosity, specific surface area, crystallinity and 418 stoichiometry of the HAp phase (calcium groups can be substituted by other ions) could play 419 a major role in DNA binding. The different manufacturing steps, such as the HAp densification 420 (i.e. sintering), can greatly influence most of these surface properties. In particular, ionic 421 substitution (e.g. carbonatation) and partial dehydration are known to occur frequently in HAp 422 during thermal treatment (Wang, Dorner-Reisel and Müller, 2004; Lafon, 2004) such as the 423 ones used here to decontaminate the samplers before and between experiments, and might be the source of the observed variability. Surface analyse needs to be carried out to identify 424 the physical (e.g. porosity, crystalline phases) and chemical (e.g. surface ionic groups) 425 parameters involved in DNA binding on the HAp surface and the extent to which these 426 parameters are influenced by the manufacturing and use of the sampler (e.g. sintering, 427 428 debinding, immersion in DNA solution).

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#### 433 **5 | CONCLUSION**

In view of the democratisation of the use of eDNA, tools are needed to easily and costeffectively sample eDNA. We demonstrate that 3D passive hydroxyapatite samplers can be designed and used to collect eDNA, albeit some repeatability issues. Provided we can get a better understanding and control of the interaction between eDNA and HAp, this approach offers an alternative sampling solution for eDNA-based biomonitoring. It also opens up an interdisciplinary field at the interface between engineering, surface science and molecular ecology.

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# 451 AUTHORS' CONTRIBUTIONS

452 TL and CM conceived the ideas and designed HAp samplers. Experimental design was 453 conceived by TL, LK and HV. HAp samples were thermal treated and characterized by HR, 454 ST and LG. Laboratory experiments were conducted by HV and LK. Data analysis was 455 conducted by HV and TL. HV and TL led the writing of the manuscript. All authors contributed 456 to the manuscript.

# 457 **REFERENCES**

- 458 Alvarez, A. J. *et al.* (1998) 'Amplification of DNA bound on clay minerals', *Molecular* 459 *Ecology*, 7(6), pp. 775–778. doi: 10.1046/j.1365-294x.1998.00339.x.
- Brundin, M. *et al.* (2013) 'DNA Binding to Hydroxyapatite: A Potential Mechanism for
  Preservation of Microbial DNA', *Journal of Endodontics*, 39(2), pp. 211–216. doi:
  10.1016/j.joen.2012.09.013.
- 463 Deiner, K. *et al.* (2017) 'Environmental DNA metabarcoding: Transforming how we survey
  464 animal and plant communities', *Molecular Ecology*, 26(21), pp. 5872–5895. doi:
  465 10.1111/mec.14350.
- 466 Del Valle, L. J. *et al.* (2014) 'DNA adsorbed on hydroxyapatite surfaces', *J. Mater.Chem. B*, 2(40), pp. 6953–6966. doi: 10.1039/C4TB01184H.
- 468 Der Sarkissian, C. *et al.* (2020) 'Unveiling the Ecological Applications of Ancient DNA From
   469 Mollusk Shells', *Frontiers in Ecology and Evolution*, 8, p. 37. doi:
   470 10.3389/fevo.2020.00037.
- Gallo, M. *et al.* (2018) 'Resorption of calcium phosphate materials: Considerations on the
  in vitro evaluation', *Journal of the European Ceramic Society*, 38(3), pp. 899–914.
  doi: 10.1016/j.jeurceramsoc.2017.07.004.
- Gardner, C. M. and Gunsch, C. K. (2017) 'Adsorption capacity of multiple DNA sources to
  clay minerals and environmental soil matrices less than previously estimated', *Chemosphere*, 175, pp. 45–51. doi: 10.1016/j.chemosphere.2017.02.030.
- Goldberg, C. S. *et al.* (2016) 'Critical considerations for the application of environmental DNA
  methods to detect aquatic species', *Methods in Ecology and Evolution*. Edited by
  M. Gilbert, 7(11), pp. 1299–1307. doi: 10.1111/2041- 210X.12595.
- 480 Hinlo, R. *et al.* (2017) 'Methods to maximise recovery of environmental DNA from water
  481 samples', *PLOS ONE*. Edited by H. Doi, 12(6), p. e0179251. doi:
  482 10 1271 (journal page 0170251)
- 482 10.1371/journal.pone.0179251.
- Jo, T. *et al.* (2020) 'Estimating shedding and decay rates of environmental nuclear DNA with
  relation to water temperature and biomass', *Environmental DNA*, 2(2), pp. 140–151.
  doi: 10.1002/edn3.51.
- 486 Kirtane, A., Atkinson, J. D. and Sassoubre, L. (2020) 'Design and Validation of Passive
  487 Environmental DNA Samplers Using Granular Activated Carbon and Montmorillonite
  488 Clay', *Environmental Science & Technology*, 54(19), pp. 11961–11970. doi:
  489 10.1021/acs.est.0c01863.
- 490 Lafon, J.-P. (2004) Synthèse, stabilité thermique et frittage d'hydroxyapatites carbonatées
   491 (pp. 218).
- 492 Langlois, V. S. *et al.* (2020) 'The need for robust qPCR based eDNA detection assays in
  493 environmental monitoring and species inventories', *Environmental DNA*, p. edn3.164.
  494 doi: 10.1002/edn3.164.
- Levy-Booth, D. J. *et al.* (2007) 'Cycling of extracellular DNA in the soil environment', *Soil Biology and Biochemistry*, 39(12), pp. 2977–2991. doi: 10.1016/j.soilbio.2007.06.020.
- 497 Mandelkern, M., Elias, J. G., Eden, D., & Crothers, D. M. (1981). The dimensions of DNA in
   498 solution. *Journal of molecular biology*, *152*(1), 153-161. doi: 10.1016/0022 499 2836(81)90099-1.
- 500 Mariani, S. *et al.* (2019) 'Sponges as natural environmental DNA samplers', *Current* 501 *Biology*, 29(11), pp. R401–R402. doi: 10.1016/j.cub.2019.04.031.
- 502 Maruyama, A. et al. (2014) 'The Release Rate of Environmental DNA from Juvenile and Adult

503	Fish', PLoS ONE. Edited by M. Stöck, 9(12), p. e114639. Doi:
504	10.1371/journal.pone.0114639.
505	Moushomi, R. <i>et al.</i> (2019) 'Environmental DNA size sorting and degradation experiment
506	indicates the state of Daphnia magna mitochondrial and nuclear eDNA is subcellular',
507	Scientific Reports, 9(1), p. 12500. doi: 10.1038/s41598-019-48984-7.
508	Ogram, A. V. <i>et al.</i> (1994) 'Effects of DNA Polymer Length on Its Adsorption to Soils',
509	Applied and Environmental Microbiology, 60(2), pp. 393–396. doi:
510	10.1128/AEM.60.2.393-396.1994.
511	Okazaki, M. et al. (2001) 'Affinity binding phenomena of DNA onto apatite crystals',
512	Biomaterials, vol. 22, no 18, (pp. 2459-2464). doi : 10.1016/S0142-
513	9612(00)00433-6
514	Pilliod, D. S. et al. (2013) 'Estimating occupancy and abundance of stream amphibians
515	using environmental DNA from filtered water samples', Canadian Journal of Fisheries
516	and Aquatic Sciences. Edited by J. Richardson, 70(8), pp. 1123–1130. doi:
517	10.1139/cjfas-2013-0047.
518	Rahaman, M. N. (2017). Ceramic processing and sintering (Vol. 1). CRC press.
519	Rivera, S. F. et al. (2021) 'Exploring the capacity of aquatic biofilms to act as
520	environmental DNA samplers: Test on macroinvertebrate communities in rivers',
521	Science of The Total Environment, 763, p. 144208. doi:
522	10.1016/j.scitotenv.2020.144208.
523	Seymour, M. et al. (2018) 'Acidity promotes degradation of multi-species environmental DNA
524	in lotic mesocosms', Communications Biology, 1(1), p. 4. doi: 10.1038/s42003-017
525	0005-3.
526	Shokralla, S. <i>et al.</i> (2012) 'Next-generation sequencing technologies for environmental DNA
527	research: NEXT-GENERATION SEQUENCING FOR ENVIRONMENTAL DNA'
528	Molecular Ecology, 21(8), pp. 1794–1805. doi: 10.1111/j.1365-294X.2012.05538.x.
529	Spear, S. F. et al. (2015) 'Using environmental DNA methods to improve detectability in a
530	hellbender (Cryptobranchus alleganiensis) monitoring program', Biological
531	Conservation, 183, pp. 38–45. doi: 10.1016/j.biocon.2014.11.016.
532	Turner, C. R. <i>et al.</i> (2014) 'Particle size distribution and optimal capture of aqueous
533	macrobial eDNA', <i>Methods in Ecology and Evolution</i> . Edited by M. Gilbert, 5(7), pp.
534	676–684. doi: 10.1111/2041-210X.12206.
535	Turner, C. R., Uy, K. L. and Everhart, R. C. (2015) Fish environmental DNA is more
536	concentrated in aquatic sediments than surface water', <i>Biological Conservation</i> , 183,
537	pp. 93–102. doi: 10.1016/j.biocon.2014.11.017.
538	Wang, T., Dorner-Reisel, A. and Müller, E. (2004) 'Thermogravimetric and thermokinetic
539	investigation of the dehydroxylation of a hydroxyapatite powder', Journal of the
540	European Ceramic Society, 24(4), pp. 693–698. doi: 10.1016/S0955-2
541	219(03)00248-6.
542	Wilcox, T. M. <i>et al.</i> (2015) 'Environmental DNA particle size distribution from Brook Trout
543	(Salvelinus fontinalis)', <i>Conservation Genetics Resources</i> , 7(3), pp. 639–641. doi:
544 545	10.1007/s12686-015-0465-z.
545 546	Williams, K. E., Huyvaert, K. P. and Piaggio, A. J. (2017) 'Clearing muddled waters:
546 547	Capture of environmental DNA from turbid waters', <i>PLOS ONE</i> . Edited by H. Doi, 12(7), p. e0179282. doi: 10.1371/journal.pone.0179282.
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