Title: Positive and negative effects of land abandonment on butterfly communities

revealed by a hierarchical sampling design across climatic regions

Running title: Land abandonment and butterfly diversity

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22   Abstract

Land abandonment may decrease biodiversity but also provides an

opportunity for rewilding. It is therefore necessary to identify areas that may benefit

from traditional land management practices and those that may benefit from a lack of

human intervention. In this study, we conducted comparative field surveys of butterfly

occurrence in abandoned and inhabited settlements in 18 regions of diverse climatic

zones in Japan to test the hypotheses that species-specific responses to land

abandonment correlate with climatic niches and habitat preferences. Hierarchical

models that unified species occurrence and habitat preferences revealed that negative

responses to land abandonment were associated with species that have cold climatic

niches and utilize open habitats, suggesting that species negatively impacted by land

abandonment will decline more due to future climate warming. Maps representing

species gains and losses due to land abandonment, which were created from the model

estimates, showed similar geographic patterns, but some areas exhibited high species

losses relative to gains. Our hierarchical modelling approach was useful for scaling up

local-scale effects of land abandonment to a macro-scale assessment, which is crucial to

developing spatial conservation strategies in the era of depopulation.

39   Keywords

40   depopulation, hierarchical model, rewilding, ridge regression, butterfly, climate change

41   adaptation

Introduction

Traditional agricultural landscapes, with high spatial heterogeneity and

moderate levels of disturbance, harbour a variety of organisms worldwide [1]. However,

recent social changes, such as depopulation and population ageing, have resulted in an

increase in abandoned fields [2,3]. While land abandonment is expected to decrease

biological diversity at the national scale [4,5], it may also provide an opportunity for

rewilding by vegetation succession [6,7]. To establish a long-term, broad-scale strategy

for biodiversity conservation, it is necessary to identify areas where restoration by human

management is effective for enhancing biodiversity as well as areas where rewilding is

more effective.

Many studies have examined the relationship between land abandonment and

biodiversity, but most have targeted a particular region within a nation [8,9]. These results

cannot necessarily be extrapolated to other regions with different climates and land cover

types because species responses to environmental changes may differ in different

climatic zones [10,11]. Moreover, a common approach for testing the effect of land

abandonment has been to analyse each species separately or analyse diversity metrics, but

such approaches cannot predict the responses of species in a community that are not

detected in a field survey. Meta-analysis may be a promising tool for evaluating the effect

of land abandonment [12,13], but it has shortcomings of heterogeneity in study design

and inconsistent outcomes [14].

A practical approach to resolve these issues is to collect biodiversity and

land-use data that are relevant to the spatial scale of land abandonment, with the data

collected from multiple climate zones (ecoregions) using a standardized sampling

strategy. This makes it possible to evaluate how the climatic niches or habitat preferences

of organisms (i.e. typical habitat types inhabited by particular species) affect the

relationship between land abandonment and species occurrence. Historically, abandoned

settlements (or deserted villages) have emerged globally due to a variety of causes such as

economic and environmental changes and disasters [15–17], and they provide unique

opportunities to evaluate the long-term effects of land abandonment and recovery of

ecosystems and to compare the effects among different climates. Species-specific

sensitivity to land abandonment is determined by ecological traits [18,19] related to

habitat preferences. Thus, information on habitat preferences which can be easily

obtained from the literature would be useful for estimating the effects of land

abandonment on species that are not included in field survey data.

Traditional agricultural landscapes have been altered in Japan with urbanization

and land abandonment, but land abandonment has become more serious in recent years

because of the country's declining human population. For instance, semi-natural

grasslands areas have decreased by 87% during the past century [20] and agricultural

fields have decreased by 26% [21]. These trends will continue in the future; 30%–50% of

currently inhabited settlements in rural areas are expected to become uninhabited by 2050

[22]. The abandonment of agricultural landscapes with various characteristics, such as

paddy fields, dry crop fields and grasslands, has caused declines in species diversity of

various taxa [13,23–26]. As a result, many organisms are expected to become endangered

[27], and butterflies are a typical example. In particular, grassland butterfly species are

decreasing substantially; they account for approximately 70% of all Red List butterfly

species in Japan [28]. These grassland species are elements of the eastern Eurasian

temperate steppe biome [29], which is located in a cooler region than the Mediterranean

and European steppes [30]. These species were common in Japan in the late Pleistocene

when the climate was much cooler than at present and natural grasslands were

predominant [31,32]. Later, in the warm Holocene when forests became predominant,

open lands maintained by human activity offered refugia for the Pleistocene relic

grassland fauna. Thus, the effect of land abandonment is expected to be more severe for

species that favour cold climates, since grassland species in eastern Eurasia are likely to

be cold adapted. Meanwhile, land abandonment may increase forest-dwelling species

during the process of vegetation succession. Although relatively few forest-dwelling

butterfly species in Japan are designated as threatened, there appears to be a trade-off in

butterfly conservation strategies with respect to whether we should manage abandoned

land to maintain open habitats by human intervention or let succession proceed. Thus, the

effect of land abandonment should be evaluated for each species across regions, and

spatially explicit planning is needed to optimize conservation strategies. For nation-wide

evaluations, butterflies are suitable subjects owing to extensive studies of their

geographical distributions, preferred habitats and larval host plants [33].

The aim of this study was to clarify the effect of land abandonment on butterfly

communities at a national scale, including different climatic regions, and to determine

how responses to land abandonment differ among species with different habitat

preferences. Here, we chose both inhabited and abandoned settlements in each region

across Japan to estimate the effect of land abandonment, defined by inhabitation. We

focused on three land-use types, namely dry crop field, paddy field and built-up area (i.e.

houses and gardens around them), because they can be identified from past terrain maps

and visual inspection in the field. The following hypotheses are addressed in this study: (1) species preferring colder temperatures are more likely to show negative responses to land abandonment; and (2) open habitat species (inhabiting grassland/agricultural lands) generally show negative responses to land abandonment, whereas forest-dwelling species exhibit positive responses. Based on the estimated coefficients of land abandonment for butterflies with different habitat preferences, we created a nation-wide map reflecting the risks and benefits of land abandonment for butterfly diversity; this map could provide a guideline for identifying regions that may benefit from continuous human intervention.

Materials and methods

Study area

We selected 18 regions that covered most parts of mainland Japan (Fig. 1) and differed in terms of mean annual temperature (MAT) (Table S1). The range of MAT in the selected regions was 4.5°C-15.7°C, which covers most of the MAT range in areas inhabited by humans in mainland Japan (the range of MAT of all the meteorological stations except Mt. Fujiwas was 6.0°C-18.8°C) [34]. In each region, we selected both abandoned and nearby inhabited settlements in approximately equal number, with a total of 2–5 settlements, including both types. We selected abandoned settlements from databases of abandoned settlements in Japan [35–37] according to the following criteria: (1) the settlement was inhabited in the past but uninhabited at present; (2) agricultural land was present before abandonment; and (3) the year of abandonment was known. As a result, we surveyed 34 abandoned and 30 inhabited settlements across all regions in Japan. The time elapsed since abandonment varied among settlements, ranging from 8 to 53

134   years (Table S1). Pearson's correlation coefficient between years since abandonment and

135   MAT was -0.16 (95% CI: -0.39, 0.09), and the difference in MAT between abandoned and

136   inhabited settlements was not significant ($p = 0.44$, Student's t-test). The mean areas ($\pm$

137   standard error) of the abandoned and inhabited settlements were, respectively, 10.6 ± 3.1

138   ha and 12.2 ± 4.4 ha, and these areas did not differ significantly.

140   *Butterfly survey*

We visited each settlement during late May to early August in 2015 and in 2016.

To account for differences in butterfly phenology among regions with different

temperatures, the visits started in the southern region and ended in the northern region in

each year. In 2016, we revisited 16 settlements which we surveyed in 2015 to examine the

effect of survey season on butterfly communities. In each visit, we established 4 to 16

plots (each 100 m^2) for the butterfly survey, depending on the area of the settlements. The

survey plots were located at dry crop fields, paddy fields and built-up areas, and the land

use category at each plot was also recorded. All the survey plots in inhabited settlements

were under active land use; crop fields and paddy fields were cultivated, and houses were

occupied. In contrast, in the abandoned settlements, all human activity was absent and

some houses had even collapsed. The butterfly survey was conducted on sunny or slightly

cloudy days from 8:00 am to 12:30 pm. A 5-minute census was conducted from roads or

rural paths adjacent to the survey plots, and the presence/absence of all species of

butterflies found at the plot was recorded. As we did not capture individual butterflies, but

instead recorded individuals by sight, some groups of species were difficult to identify to

the species level during flight. Thus, the following taxonomic groups were omitted from

further analysis: (1) four species of *Papilio* (*P. protenor*, *P. macilentus*, *P. memnon* and *P.*

helenus); (2) six species of the tribe Argynnini (*Fabriciana adippe*, *Speyeria aglaja*,

Nephargynnis anadyomene, *Argynnis paphia*, *Argyreus hyperbius* and *Damora sagana*);

(3) Hesperiidae; and (4) three genera of the tribe Theclini (*Neozephyrus*, *Favonius* and

Chrysozephyrus). We also excluded the non-native species *Pieris rapae* [38] and *P.*

brassicae.

The habitat preferences for each butterfly species were determined based on the

habitat categorization established by the Japan Butterfly Conservation Society [33]. The

categorization consists of the following 11 types: forests, forest edges, open forests,

grasslands, crop fields, gardens, built-up areas, rivers, wetland, alpine and rocky habitats.

The habitat preferences of a species often comprised several habitat types (Table S1).

168   *Statistical analysis*

We applied three statistical approaches: (1) evaluating the correlation between

species-specific responses to temperature and land abandonment; (2) evaluating the

relationship between habitat preferences of species and susceptibility to land

abandonment; and (3) developing a predictive model for species-specific responses to

land abandonment based on habitat preferences. Our dataset had a hierarchical structure

in which settlements were nested within a region, and a settlement contained multiple

land-use types. Survey plots with the same land-use type were generally spatially

aggregated in a settlement. Therefore, we incorporated region-, settlement- and

land-use-type-level random effects into the analyses to accommodate the spatially

hierarchical sampling design. Then, we treated the presence/absence of a species in a

plot as the unit of analysis. These models were implemented by a hierarchical Bayesian

approach.

(1) Correlation between the species-specific response to temperature and land

182   abandonment

We applied generalized linear mixed models (GLMM) [39], incorporating

heterogeneous responses to environmental properties among species as a random slope

to determine whether species that prefer colder temperatures are more likely to respond

negatively to land abandonment. The presence and absence of the ith species at the jth

survey plot, Y_{ij}, was assumed to follow a Bernoulli distribution, Bernoulli(Y_{ij}; p_{ij}), where

p_{ij} is the probability of occurrence. The parameter p_{ij} was assumed to be expressed by a

logistic regression model. We used land abandonment ($ABAN_j$) and mean annual

temperature ($TEMP_j$) as explanatory variables. We also included month surveyed

($MONTH_j$) and a categorical variable of three land-use types (LU$_j$: dry field, built-up

area and paddy field [baseline category]) as confounding factors. $TEMP_j$ was obtained

from the Climate Mesh Data 2000 [40]. $MONTH_j$ and LU$_j$ were confounding factors

controlling for the possible influence on the other parameters of seasonality and

land-use type at each survey plot, respectively. We also considered region-, settlement-

and land-use-type-level random effects, namely $\varepsilon_{reg(j)}$, $\varepsilon_{set(j)}$ and $\varepsilon_{lu(j)}$, following normal

priors with mean zero and variance σ_r^2, σ_s^2 and σ_l^2, respectively, to accommodate the

hierarchical sampling scheme with unbalanced sample size [41]. Overall, p_{ij} was

modelled as follows:

$\text{logit}(p_{ij}) = \beta_{0i} + \beta_{1i}ABAN_j + \beta_{2i}TEMP_j + \beta_{3i}MONTH_j + \beta_{4i}I(LU_j = \text{"dry field"}) +$

$\beta_{5i}I(LU_j = \text{"built-up area"}) + \varepsilon_{reg(j)} + \varepsilon_{set(j)} + \varepsilon_{lu(j)},$ Eqn. 1

where β_{0i} is a species-specific fixed intercept, and β_{1i}, β_{2i}, β_{3i}, β_{4i} and β_{5i} are regression

coefficients for the ith species. $I()$ is an indicator function for the dummy variable LU_j.

To infer species-specific effects of land abandonment and the correlation with

their preferred temperature, a GLMM with a random slope was applied to estimate

regression coefficients for each species. In a random slope model, regression

coefficients of species are treated as random effects subject to the same prior

distribution. As a prior distribution of $\boldsymbol{\beta}_i = (\beta_{1i}, \beta_{2i})$, the bivariate normal distribution

$\text{MVN}(\boldsymbol{\beta}_i; \boldsymbol{\mu_\beta}, \boldsymbol{\Sigma_\beta})$ was applied. $\boldsymbol{\mu_\beta} = (\mu_{\beta1}, \mu_{\beta2})$ and $\boldsymbol{\Sigma_\beta}$ are the mean vector and

covariance matrix, respectively. The covariance matrix was decomposed to standard

deviations and Pearson's correlation coefficient as follows:

$$\boldsymbol{\Sigma_\beta} = \begin{bmatrix} \sigma_{\beta1}^2 & \rho\sigma_{\beta1}\sigma_{\beta2} \\ \rho\sigma_{\beta1}\sigma_{\beta2} & \sigma_{\beta2}^2 \end{bmatrix}$$

Here, ρ is Pearson's correlation coefficient between β_{1i} and β_{2i}; a positive value indicates

that a species preferring cooler temperatures is more likely to be negatively impacted by

land abandonment. The priors of regression coefficients of confounding factors, namely

β_{3i}, β_{4i} and β_{5i}, followed normal distributions with mean $\mu_{\beta3}$, $\mu_{\beta4}$ and $\mu_{\beta5}$ and variance $\sigma_{\beta3}$,

$\sigma_{\beta4}$ and $\sigma_{\beta5}$, respectively.

We calculated posterior distributions of the model by the Bayesian method

using vague or weakly informative priors for the hyperparameters of the model. For β_{0i},

$\mu_{\beta1}$, $\mu_{\beta2}$, $\mu_{\beta3}$, $\mu_{\beta4}$ and $\mu_{\beta5}$, a normal prior with mean 0 and variance 100 was used. A

half-Cauchy distribution with a scale parameter set to 5 was used as a weakly

informative prior for the standard deviations: σ_r, σ_s, $\sigma_{\beta 1}$, $\sigma_{\beta 2}$, $\sigma_{\beta 3}$, $\sigma_{\beta 4}$ and $\sigma_{\beta 5}$ [42,43]. A

uniform distribution with range (-1, 1) was used for the prior of the correlation

coefficient, ρ. Samples from the posterior distribution were obtained using the

No-U-Turn Sampler implemented in RStan 2.21.1 [44] (three chains, 1000 iterations

after 1000 burn-in iterations with no thinning). We obtained 3000 posterior samples.

Convergence of the Markov chain Monte Carlo (MCMC) algorithm was evaluated

using $\hat{R}$ [45], and we adopted $\hat{R} < 1.1$ as a threshold of successful convergence [46].

We also evaluated the goodness of fit by using the posterior predictive p-value [47] with

the following summary statistics: (1) proportion of presence observations over the

dataset; (2) proportion of presence observations at abandoned settlements; (3)

proportion of presence observations at inhabited settlements; (4) the Gini coefficient of

the proportion of presence observations among the settlements; and (5) the Gini

coefficient of the proportion of presence observations among species.

For the explanatory variable $ABAN_j$, we considered two candidate variables,

"abandoned or not" and "years since abandonment", because it is unknown whether the

response is immediate or gradual. We compared the performance of models using

"abandoned or not" and "years since abandonment" on the basis of the widely

applicable Bayesian information criterion (WBIC) [48]. WBIC is a generalization of the

Bayesian information criterion which is applicable to both regular and singular

statistical models; it asymptotically approximates the negative logarithm of the marginal

likelihood. The Bayes factor (the ratio of posterior probabilities of competing models)

was calculated as the exponential of the difference in WBICs. The abandonment

variable with a higher WBIC was used for further analyses.

(2) Relationship between habitat preferences of species and responses to land
abandonment

To test whether species with different habitat preferences respond differently to
land abandonment, we applied a hierarchical model incorporating a hierarchy of habitat
type and species [49]. The model is an extension of the species-level model described in
section (1), in which the average effect of land abandonment $\mu_{\beta 1}$ was modelled by the
linear predictor of habitat preferences as follows:

251            $\mu_{\beta 1i} = \alpha_0 + \alpha_1 H_{ik},$ Eqn. 2

where H_{ik} is a binary variable indicating whether the ith species utilizes the kth ($k = 1, 2,$
…, 11) habitat type or not. α_0 is the intercept, and α_1 indicates the difference in the
average effects of land abandonment between habitat types, where positive and negative
values of α_1 indicate positive and negative effects for species utilizing the kth habitat
type relative to species utilizing other habitat types. The same structure was used to
determine the effect of $TEMP_j$; that is, $\mu_{\beta 2i} = \alpha_{T0} + \alpha_{T1} H_{ik}$. As in the model of section (1),
slopes for each species vary following a multivariate normal distribution with mean
vector ($\mu_{\beta 1i}, \mu_{\beta 2i}$).

We calculated posteriors of α_0, α_1, α_{T0} and α_{T1} separately for the 11 habitat
types. We used vague priors of a normal distribution with mean 0 and variance 1000 for
these parameters. Other prior settings were the same as those for the species-level model
(Section (1)). Samples from the posterior distribution were obtained using RStan 2.21.1
[44], with the same MCMC settings and posterior diagnostics as in Section (1).

265   (3) Predictive model of species-specific responses to land abandonment

266   A hierarchical modelling approach is a powerful tool for predicting

267 species-specific responses to environmental changes [50], and allows us to evaluate the

268 broad-scale distribution of a species. Here, we developed a predictive model based on

269 the habitat preferences of a species. As most of the species inhabit several habitat types,

270 the predictive model was a multivariate extension of Eqn. 2:

271             $\mu_{\beta 1 i} = \alpha_0 + \Sigma_k \alpha_k H_{ik}.$ Eqn. 3

However, models that are too complex are vulnerable to overfitting and tend to

have large prediction error. In such a situation, ridge regularization is useful to constrain

overfitting and improve prediction accuracy [51]. In a Bayesian manner, ridge

regularization can be simply implemented using the same normal prior, $N(0, \sigma_\alpha^2)$, for

each regression coefficient. This prior works as a penalty for the divergence of squared

regression coefficients from zero, and the standard deviation, σ_α, corresponds to the

weight of the penalty. The prior of σ_α was set to a weakly informative prior,

half-Cauchy(0, 5), and σ_α was determined by calculating the posterior distribution. The

prior of α_0 was a normal distribution with mean 0 and variance 100. Samples from the

posterior distribution were obtained using RStan 2.21.1 [44], with the same MCMC

sampling settings and posterior diagnostics as in Sections (1) and (2).

To make projection maps for the effect of land abandonment on butterfly

assemblages, we defined the loss (CL) and gain (CG) in butterfly species due to land

abandonment as follows:

$$\mathrm{CL}_j = \sum_i^{N_{sp}} \delta_{ij} \left| min\left(\mu_{\beta 1i}, 0\right) \right|,$$

286   and

$$\mathrm{CG}_j = \sum_i^{N_{sp}} \delta_{ij} max\left(\mu_{\beta 1i}, 0\right),$$

where N_{sp} is the number of species and δ_{ij} is a binary indicator of whether the ith species

is present in the jth grid cell. Operators $min\left(\mu_{\beta 1i}, 0\right)$ and $max\left(\mu_{\beta 1i}, 0\right)$ are filters to

identify species affected negatively and positively, respectively. CL and CG can be

interpreted as log-odds ratios of occurrence probabilities across species after land

abandonment. Large values of CL or CG respectively indicate a large loss or gain in

species in the local butterfly assemblage due to land abandonment.

We used nation-wide range maps of 70 butterfly species in Japan (Table S3),

which were projected by Kasada et al. [52] using distribution models (MaxEnt) from

butterfly records obtained from the fourth and fifth National Surveys on the Natural

Environment. Before calculating CL and CG, we excluded seven alpine butterflies,

Carterocephalus palaemon, *Erebia neriene*, *E. ligea*, *Oeneis norna*, *Colias palaeno*,

Anthocharis cardamines and *Aporia hippia*, because land abandonment is not relevant

in their ranges. The range maps comprise presence and absence data for species in each

grid cell of a "1 × 1 km mesh (the third mesh)," a Japanese national grid system whose

unit cell size is $30''$ in latitude and $45''$ in longitude, approximately 1 × 1 km. We

determined the expected coefficient of land abandonment, $\mu_{\beta 1}$, using Eqn. 3, for 70

species. We then evaluated CL and CG separately for all species and for species

classified as "near threatened (NT)", "vulnerable (VU)", "endangered (EN)" or

"critically endangered (CR)" in the Red List of Japan in 2019 [53]. Then, we calculated

CG – CL to evaluate whether loss of species richness due to land abandonment

outweighed the gain.

308   Results

A total of 49 butterfly species were recorded during our field surveys (Table

S2). The top 10 species with the highest average occurrence (i.e. number of presence

plots/total number of plots) were *Pieris melete*, *Colias erate*, *Celastrina argiolus*,

Ypthima argus, *Libythea celtis*, *Plebejus argus*, *Lycaena phlaeas*, *Eurema mandarina*,

Papilio machaon and *Apatura metis*, in decreasing order, and they represented 73.4% of

the total occurrence records.

The parameters of all models successfully converged ($\hat{R} < 1.1$). In posterior

predictive checking, no summary statistics of the observed data showed significant

divergence from the posterior predictive distributions generated from all estimated

models (Table S4). Model selection using the WBIC indicated that the model including

"abandoned or not" (WBIC = 3577.68) outperformed the model for "years since

abandonment" (WBIC = 3611.28), suggesting relatively immediate responses to land

abandonment. The Bayes factor for the former model against the latter was 3.89×10^{14}.

Among the 49 butterfly species analysed, 13 showed a significant negative response (i.e.

the upper 95% CI of the regression coefficient of "abandoned or not" was less than

zero), while 3 species showed a positive response (Table S5). *Lycaena phlaeas, Papilio*

machaon and *Papilio xuthus* exhibited the largest negative effect of land abandonment.

The three species that showed positive responses were *Limenitis camilla*, *Cyrestis*

thyodamas and *Graphium sarpedon*. Effects of land abandonment and temperature were

positively correlated (Fig. 2); Pearson's correlation coefficient between the effect of

land abandonment and annual mean temperature was 0.4 with a 95% CI of (0.07, 0.7).

There was a significant relationship between habitat types of species and

responses to land abandonment for four habitat types; species utilizing farmland,

grassland and built-up areas responded negatively. Although those utilizing forest

showed a significant difference from the other habitat types, the 95% CI of the average

response to land abandonment overlapped with zero (Fig. 3, Table S6).

Using a hierarchical model with ridge regularization, we obtained the

predictive model for the effect of land abandonment based on the habitat preferences of

each species (Table S7). Although the parameter uncertainty of each regression

coefficient obtained by ridge regression was large due to the correlation between habitat

preferences, the model retained discriminative ability for species which respond

negatively to land abandonment (Fig. S1). By applying this model, posterior means of

37 and 33 species were negative and positive, respectively, and the upper 95% CIs of 7

species were negative. Using the predicted values thus obtained, we drew maps of CL

(Fig. 4a) and CG (Fig.4b) of species throughout Japan's mainland. The geographic

distributions of loss and gain were similar on a coarse scale and the Pearson's

correlation coefficient was 0.55; specifically, larger values of both CL and CG were

found in inland central Kyushu, inland western Honshu, inland central Honshu, northern

Honshu and Hokkaido, while smaller values were found in high elevation areas.

However, CL outweighed CG in 99.6% of grid cells (358045 of 359429) and CL tended

to be higher in lowlands (Fig. 4c,d,e). Standard deviations of CL and CG and the

proportion of CG (Fig. S2a,b,d) were almost directly proportional to their posterior

means (Pearson's correlation coefficient $r = 0.86$, 0.99 and 0.96, respectively). The

standard deviations of the proportion of CL and CG – CL (Fig. S2c,e) had stronger

positive correlations with CG ($r = 0.87$ and 0.95, respectively) than did their posterior

means ($r = -0.96$ and 0.51), reflecting the large uncertainty of CG.

Of the Red List species, 10 were evaluated as victims of land abandonment and

4 were evaluated as beneficiaries. Distribution patterns of CL and CG for Red List

species of Japan are shown in Fig. 5a,b. The correlation between CL and CG was

weaker for Red List species than for all species ($r = 0.36$), and their spatial patterns

were different on a coarse scale. For instance, CL was small while CG was large in

northern Honshu. Conversely, CL was large but CG was small in eastern Hokkaido (Fig.

5c,d,e). Similar to the maps for all species, the standard deviations of the CL and CG

and the proportion of CG (Fig. S3a,b,d) were almost directly proportional to their

posterior means ($r = 0.95$, 0.99 and 0.92, respectively), and the standard deviations of

the proportion of CL and CG – CL, rather than their posterior means, were correlated

with CG (Fig. S3c,e).

366   Discussion

As expected, the responses of butterflies to land abandonment were associated

with the climatic niches and habitat preferences of species. Through our approach in this

study, which compared community assemblages between inhabited and abandoned areas

using a standard survey protocol, we were able to estimate the effects of land

abandonment over a broad range of climatic condition. In combination with the

hierarchical modelling approach, we were able to draw nation-wide potential maps of the

loss and gain of species richness through land abandonment.

374   The legacies of the past climate and past land use can affect current functional

species compositions within regional species pools [54,55], and different functional

compositions among regions can result in different community-level responses to

current environmental changes [56]. Such a difference would also reflect the historical

background of organisms inhabiting seminatural environments in Japan. Grassland was

more abundant in the Pleistocene because the climate was colder and drier than at

present [32]. After the last glacial period, open lands created by human activity have

offered refugia for the Pleistocene relics [57], and such refugia are unlikely to be

sustainable if they are abandoned [58]. Our results also raise a conservational concern

that species negatively impacted by land abandonment will decline more due to climate

warming in the future, because the species susceptible to land abandonment tend to

prefer low temperatures (Fig. 2). This means that continuing land management in rural

areas would be an effective measure of climate change adaptation for biodiversity

conservation in Japan.

388   Habitat degradation for openland butterflies due to land abandonment occurred

quickly, whereas forest butterflies would require a longer period of time to experience a

positive effect. We suggest that land abandonment resulted in a decrease in the richness

of herbaceous plant species [23], leading to the loss or decline of host plants for open

habitat butterflies (grassland, crop field, and residential species). Indeed, we often

observed tall grasses, such as *Miscanthus sinensis*, as predominant species in abandoned

paddy, dry fields, and built-up area in uninhabited settlements. The monodominance of

wind-pollinated *Miscanthus* would decrease the abundance and diversity of

nectar-producing flowers that are important for adult butterflies. In contrast, forest

species tended to increase in response to land abandonment, although the proportion of

such species was low. Development of old-growth forest in Japan requires

approximately 150 years [59], and some tree species require more than a hundred years to

recover [60]; accordingly, recovery of forest butterfly species that depend on such trees

would also require very long time as well.

It should be noted that there was a positive correlation between losses and gains

in butterfly richness both for all species and for Red List species due to land abandonment

at the level of grid cells, indicating a trade-off between the benefits of active management

of abandoned sites and letting succession proceed. As threatened butterflies in Japan

include more grassland species than forest species, human intervention should be

prioritized, in principle, to conserve such species, but deliberate decision-making is

required for areas where the benefit of land abandonment is high. Our results indicate that

conservation of butterflies by active management is desirable in lowland and eastern

Hokkaido, where the loss of species that prefer abandoned landscapes would small under

such management. Highly cost-effective areas for habitat management based on species

ranges and habitat preferences need to be identified with high resolution, because

maintaining secondary natural environments is generally labour-intensive [61]. The maps

of gain and loss of butterfly diversity presented in this study did not incorporate spatial

heterogeneity in land abandonment and instead assumed that all locations in Japan were

abandoned. In actuality, the current human population density and the speed of decline is

spatially heterogeneous, which will create heterogeneous land-use patterns in the future

[62,63]. However, our predictions would still be useful for evaluating biodiversity loss

and gain in rural landscapes because human population decline is expected to be severe in

most of those areas.

Here, we evaluated the risks and benefits of land abandonment using butterfly

occurrence data; however, it is also important to clarify how the abundance of larval host

plants and adult nectar plants changes over time after abandonment. Previous studies

have shown that many grassland species are still found even 10 years after abandonment

[64], indicating the presence of host and nectar plants during that period. In our study, the

effects of land abandonment did not show a clear increase over time, as the model with a

dichotomous variable (inhabited vs. abandoned settlements) outperformed the model

with a continuous variable (years since abandonment). As Japan has a monsoon climate

with relatively high temperatures and rainfall in comparison with Europe, vegetation

succession may proceed within a few years, resulting in a rapid response to land

abandonment. This suggests the need to restore habitats as soon as possible after

abandonment. The potential maps of the effect of land abandonment will contribute to

prioritizing habitat management plans in areas where settlement abandonment in the

immediate future and a large decrease in biodiversity after abandonment are expected.

A major contribution of our study was the ability to apply the estimated

relationship between habitat preferences and responses to land abandonment to butterfly

species that were not observed in the field survey. Potential drawbacks of our survey

design are sparse occurrence records and low completeness of species in the dataset due

to the small amount of survey effort per sample. Hierarchical modelling could overcome

the problems of sparse data, such as model unidentifiability, by introducing a hierarchy of

superspecies that stabilize the species-level estimates [65]. With this hierarchical

structure, we could predict the responses of butterfly communities even when not all

species were detected in the field survey. The utility of such a hierarchical modelling

approach is widely recognized [66], but there are some caveats, as responses of habitat

generalists may be difficult to predict. To resolve this problem, models incorporating

responses of host plants to land abandonment and resultant butterfly responses should be

explicitly constructed, as has been done for climate change prediction models [67].

Our hierarchical modelling approach also allowed us to evaluate the correlation

between susceptibility to climate warming and land abandonment, which could provide

important information for the management of semi-natural habitats in order to ensure

biodiversity under climate change. However, we should be cautious about extrapolating

our results to other of in the world with different biogeographical factors. For example,

some grassland butterflies in Europe have expanded their range northward in response to

the warming climate [68], while such a range expansion is not expected in Japan. Such a

difference would arise from the different latitudes of stable grassland biomes. Grassland

butterflies in Japan have high commonality with Eurasian temperate steppe fauna ranging

from eastern Eurasia to Mediterranean region, and seminatural grassland in Japan can be

thought as the extrazonal remnant of the temperate stable grassland [29]. Interestingly,

the climatic limit of eastern Eurasian steppe is much cooler than the Mediterranean region

[30]. This is due to the difference in relative importance of coldness and dryness for the

formation of stable grassland; in eastern Eurasia, the rainy season occurs in the growing

season of plants and winter frost plays a more important role than in the Mediterranean

steppe. Europe and Japan are respectively located in the northern and southern regions of

the outer part of the Eurasian temperate steppe. Thus, a northward biome shift due to

climate warming can help grassland butterflies in Europe, but not in Japan. A

continental-scale comparative study of the responses of grassland butterflies to climate

warming is a potential topic for future research to test the hypothesis.

Although we focused on the effect of land abandonment in this study, the

hierarchical modelling approach would be also applicable to predicting the response of

biological community to climate change if climatic niche information of species [69] is

available. A database of climatic niche information of butterflies in Japan has not yet been

developed due to the difficulty in collecting the distribution data of butterflies which are

widespread in Asia, and this should be a topic for further research.

In conclusion, a hierarchical modelling approach that incorporates the

relationship between the habitat preferences of species and their responses to land

abandonment was useful for scaling up the results of comparative field studies to a

macro-scale evaluation of species losses and gains due to land abandonment. Land

abandonment had both positive and negative effects on different functional groups of

butterflies in Japan, and our approach could be used to identify areas exhibiting high

species losses relative to gains after land abandonment. We believe this approach is

promising for large-scale and comprehensive biodiversity assessments regarding the risks

and benefits of land abandonment, which are urgently needed in view of ongoing

depopulation in developed countries.

Ethics

Our field survey was conducted from public roads and rural paths, and no permission

was required in compliance with Japanese law. All information of site locations is

coarse enough to protect landowners' privacy. To avoid a negative impact on

conservation due to specimen collection, we have not released specific records of

endangered species.

Data accessibility

The datasets, program codes and model predictions shown in Fig.4, 5 and Supplementary

Fig. S2, 3 are available from the Dryad Digital Repository at

https://datadryad.org/stash/share/8u3_mEXi3SsWzm_ZElaD7iq7wWB8FSQ45czwu8u

_V24, except specific records of endangered species.

Author contributions

NS, KF, and TM conceived the ideas; NS conducted the field study; AA prepared part of

the data for the analysis; MK and MM prepared the projected national distribution maps

of butterflies; NS and KF conducted the analysis; and NS, KF, and TM wrote the

manuscript with input from AA, MK and MM.

Competing interest

We have no competing interest.

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Fig. 1 Location of study sites. Each site contained both rural abandoned villages and inhabited villages.

Fig. 2 Relationship between coefficients of mean annual temperature and land abandonment. Black dots and grey error bars indicate posterior means and 95% credible intervals of species, respectively. The dashed ellipsoid and line are 95% equiprobable ellipsoid and principal axis of coefficients.

Fig. 3 Effects of land abandonment on the occurrence of butterflies with different habitat characteristics. On the x axis, 0 indicates that the species does not utilize the habitat type and 1 indicates that it does. Gray dots represent coefficients of abandonment (β_{1i}) for each species. Black dots and error bars represent posterior means and 95% credible interval (95% CI) of the expected value of the coefficient of abandonment ($\mu_{\beta1i}$ in eqn. 1), respectively. Four habitat types with significant effect on $\mu_{\beta1i}$ (i.e. 95% CI of α_1 in eqn. 2 did not overlap 0) were shown.

Fig. 4 Maps of cumulative loss (CL, a), gain (CG, b), the proportions of CL (c) and CG (d) and the difference between CG and CL (CG - CL, e) in butterfly species richness due to land abandonment for all the species. The boxed graphs are hexagonal binning plots showing their relationship with elevation.

Fig. 4 *Continued.*

Fig. 4 *Continued.*

(e)

Fig. 5 Maps of cumulative loss (CL, a), cumulative gain (CG, b), the proportions of CL (c) and CG (d) and the difference between CG and CL (CG - CL, e) in butterfly species richness due to land abandonment for Red List species. The boxed graphs are hexagonal binning plots showing their relationship with elevation.

Fig. 5 *Continued.*

Fig. 5 *Continued.*

(e)