Title:  Hierarchical trait-based model reveals positive and negative effects of land

abandonment on butterfly communities across climatic regions in Japan

Running title:    Land abandonment and butterfly diversity

Naoki Sugimoto[1], Keita Fukasawa[2,*], Akio Asahara[3], Minoru Kasada[4,5], Misako

Matsuba[2] and Tadashi Miyashita[1]

[1] Graduate School of Agricultural and Life Sciences, University of Tokyo, 1-1-1 Yayoi,

Bunkyo-ku, Tokyo 113-8657, Japan

[2] Biodiversity Division, National Institute for Environmental Studies, 16-2 Onogawa,

Tsukuba, Ibaraki 305-8506, Japan

[3] Team HEYANEKO, 3-22-18-702 Minami-Urawa, Minami-ku, Saitama, Saitama

336-0017, Japan

[4] Graduate School of Life Sciences, Tohoku University 6-3, Aoba, Sendai 980-8578

Japan

[5] Department of Experimental Limnology, Leibniz-Institute of Freshwater Ecology and

Inland Fisheries, Alte Fischerhuette 2, 16775, Stechlin, Germany

* Corresponding author

Email: k.fukasawa37@gmail.com; fukasawa@nies.go.jp

Tel.: +81-29-850-2676

22                                                                Abstract

23                Land abandonment may decrease biodiversity but also provides an

opportunity for rewilding. It is therefore necessary to identify areas that may benefit

from traditional land management practices and those that may benefit from a lack of

human intervention. In this study, we conducted comparative field surveys of butterfly

occurrence in abandoned and inhabited settlements in 18 regions of diverse climatic

zones in Japan to test the hypotheses that species-specific responses to land

abandonment correlate with their climatic niches and habitat preferences. Hierarchical

models that unified species occurrence and their habitat characteristics revealed that

negative responses to land abandonment were associated with species having cold

climatic niches utilizing open habitats, suggesting that traditional land use by human

had long mitigated climate warming since the last glacier maximum through creating

non-forest open habitats. Maps representing species gains and losses due to land

abandonment, which were created from the model estimates, showed similar geographic

patterns, but some parts of areas exhibited high species losses relative to gains. A

hierarchical trait-based approach presented here revealed to be useful for scaling up

local-scale effect of land abandonment to a macro-scale assessment, which is crucial to

developing spatial conservation strategies in the era of depopulation.

40                                                                Keywords

# Introduction

Traditional agricultural landscapes, with high spatial heterogeneity and moderate levels of disturbance, harbor a variety of organisms worldwide [1]. However, recent social changes, such as depopulation, population aging, and population concentration in urban areas, have resulted in an increase in abandoned fields [2,3]. While land abandonment is expected to decrease biological diversity at the national scale [4,5], it may also provide an opportunity for rewilding by vegetation succession [6,7]. To establish a long-term, broad-scale strategy for biodiversity conservation, it is necessary to identify areas where restoration by human management is effective for enhancing biodiversity as well as areas where rewilding is more effective.

Many studies have examined the relationship between land abandonment and biodiversity, but most have targeted a particular region within a nation [8,9]. These results cannot necessarily be extrapolated to other regions with different climates and land cover types because species responses to environmental changes may differ in different climatic zones [10,11]. Moreover, it is generally not possible to predict the responses of species that are not included in a field survey. Meta-analysis may be a promising tool to evaluate the effect of land abandonment [12,13], but it has shortcomings of heterogeneity in study design and inconsistent outcomes [14].

A practical approach to resolve these issues is to collect biodiversity and landuse data that are relevant to the spatial scale of land abandonment, which are taken from multiple climate zones (ecoregions) with a standardized sampling strategy. This makes it possible to evaluate how habitat traits of organisms (i.e., typical habitat types inhabited by particular species) affect the relationship between land abandonment and species

occurrence. Historically, abandoned settlements (or deserted villages) have emerged

ubiquitously due to variety of causes such as economic environmental changes and

disasters [15–17], and they offer us unique opportunities to evaluate long-term effect of

land abandonment and recovery of ecosystems and to compare them among different

climate. Species-specific sensitivity to land abandonment is determined by ecological

traits [18,19] related to habitat preferences. Such trait-based relationships can be used to

estimate the effects of abandonment on species that are not included in census data.

Traditional agricultural landscapes have been altered in Japan, with the

increases in both urbanization and land abandonment. For instance, semi-natural

grasslands areas have decreased by 87% during the last century [20] and agricultural

fields have decreased by 26% [21]. These trends will continue in the future; 30–50% of

currently inhabited settlements in rural areas are expected to become uninhabited by 2050

[22]. Abandonment of agricultural landscapes consisting of various elements, such as

paddy field, dry crop field and grassland, has been causing decline in species diversity of

various taxa [13,23–26]. As a result, many organisms are expected to become endangered

[27], and butterflies are a typical example. In particular, grassland butterfly species are

decreasing substantially; they account for approximately 70% of all Red List species of

butterflies in Japan [28]. These grassland species were common in the late Pleistocene,

when the climate was much cooler than at present and natural grasslands were

predominant [29,30]. Thus, the effect of land abandonment is expected to be more severe

in northern Japan, since more grassland species are likely included in the regional species

pool. Meanwhile, land abandonment may increase forest-inhabiting species during the

process of vegetation succession. Although fewer butterfly species are designated as

endangered in forests in Japan, there appears to be a tradeoff in butterfly conservation

strategies with respect to whether we should manage abandoned land to maintain open

habitats by human intervention or let succession proceed. Thus, the effect of land

abandonment should be evaluated for each species across regions, and spatially explicit

planning is needed to optimize conservation strategies. For nation-wide evaluations,

butterflies are suitable subjects owing to extensive studies of their geographical

distributions, preferred habitats, and larval host plants [31].

The aim of this study was to clarify the effect of land abandonment on butterfly

communities at a national scale, including different climatic regions, and to determine

how responses to land abandonment differ among species with different habitat traits.

Here, we chose both inhabited and abandoned settlements in each region across Japan to

estimate the effect of land abandonment, defined by inhabitation. We focused on three

land use types, dry crop field, paddy field and built-up area (i.e. houses and garden around

them), because they can be identified from past terrain map and visual inspection in the

field. The hypotheses addressed in this study are (1) species preferring colder

temperatures are more likely to show negative responses to land abandonment, and (2)

open habitat species (inhabiting grassland/agricultural lands) generally show negative

responses to land abandonment, whereas forest-dwelling species exhibit positive

responses. Based on parameters related to the responses to land abandonment in

butterflies with different habitat traits, we created a nation-wide map for the risks and

benefits of land abandonment for butterfly diversity; this map could provide a guideline

for identifying regions that may benefit from continuous human intervention.

                              Materials and methods

                                    *Study area*

We selected 18 regions that covered most parts of mainland Japan (Fig. 1), with

various annual mean temperatures (Table S1). In each region, we selected both

abandoned and nearby inhabited settlements of approximately equal numbers, with a total

of 2–5 settlements, including both types. We selected abandoned settlements from the

databases of abandoned settlements in Japan [32–34] by the following criteria; 1.

Settlement was inhabited in the past while uninhabited at present, 2. agricultural land

existed before abandonment, 3. The year of abandonment was determined, and 4.

Settlement was not submerged by a dam. As a result, we surveyed 34 abandoned and 30

inhabited settlements across all regions in Japan. The time elapsed since abandonment

varied among settlements, ranging from 8 to 53 years (Table S1). The mean areas ($\pm$

standard error) of the abandoned and inhabited settlements were, respectively, $10.6 \pm 3.1$

ha and $12.2 \pm 4.4$ ha, and these areas did not differ significantly.

126                                    *Butterfly survey*

We visited each settlement during late May to early August in 2015 and in 2016.

Accounting for the difference in butterfly phenology among regions with different

temperatures, the first visitations started in the southern region and ended in the northern

region in each year. In 2016, we revisited 16 settlements which we surveyed in 2015 to

examine the effect of survey season on butterfly communities. In each visitation, we

established 4 to 16 spots (each 100 m$^2$) for the butterfly survey, depending on the area of

the settlements. The survey spots were located at dry crop field, paddy field and built-up

area, and the land use category at each spot was also recorded. All the survey spots in

inhabited settlements were under active land use; crop fields and paddy fields were

cultivated, and houses were used by residents. In contrast, all these human activities were

absent, and some houses were even collapsed. The area of each spot was about 100 m$^2$.

The butterfly survey was conducted on sunny or slightly cloudy days from 8:00 am to

12:30 pm. A 5-minute census was conducted from roads or rural paths adjacent to the

survey spots, and the presence/absence of all species of butterflies found at the plot was

recorded. As we did not capture individual butterflies, but instead recorded individuals by

sight, some groups of species were difficult to identify to the species level during flight.

Thus, the following taxonomic groups were omitted from further analysis: (1) four

species of *Papilio* (*P. protenor*, *P. macilentus*, *P. memnon,* and *P. helenus*), (2) six species

of the tribe Argynnini (*Fabriciana adippe*, *Speyeria aglaja*, *Nephargynnis anadyomene*,

*Argynnis paphia*, *Argyreus hyperbius,* and *Damora sagana*), (3) Hesperiidae, and (4)

three genera of the tribe Theclini (*Neozephyrus*, *Favonius,* and *Chrysozephyrus*). We also

excluded nonnative species, *Pieris rapae* [35] and *P. brassicae*.

The habitat traits for each butterfly species were determined based on the habitat

categorization established by the Japan Butterfly Conservation Society [31]. The

categorization consists of the following 11 types: forests, forest edges, open forests,

grasslands, crop fields, gardens, built-up areas, rivers, wetland, alpine, and rocky habitats.

The habitat trait of a species often comprised several habitat types (Table S1).

*Statistical analysis*

We applied the three statistical approaches to (1) evaluate the correlation

between species-specific responses to temperature and land abandonment, (2) evaluate

the relationship between habitat traits of species and susceptibility to land abandonment,

and (3) develop a predictive model for species-specific responses to land abandonment

based on habitat traits. We incorporated region- and settlement-level random effects into

the analyses because of the spatially clustered sampling design. These models were

implemented by a hierarchical Bayesian approach.

(1) Correlation between the species-specific response to temperature and land

abandonment

We applied generalized linear mixed models (GLMM) [36], incorporating

heterogeneous responses to environmental properties among species as a random effect

to determine whether species that prefer colder temperatures are more likely to respond

negatively to land abandonment. The presence and absence of $i$th species at $j$th survey

spot, $Y_{ij}$, was assumed to follow a Berunoulli distribution, Bernoulli($Y_{ij}$; $p_{ij}$), where $p_{ij}$ is

the probability of occurrence. The parameter $p_{ij}$ was assumed to be expressed by a

logistic regression model. We used land abandonment ($ABAN_j$) and mean annual

temperature ($TEMP_j$) as explanatory variables. We also included month surveyed

($MONTH_j$) and categorical variable of three land use types (paddy field, dry field and

built-up area, $LU_j$) as confounding factors. $TEMP_j$ was obtained from the Climate Mesh

Data 2000 [37]. $MONTH_j$ and $LU_j$ were confounding factors controlling the possible

influence of seasonality and land use type at each survey spot on the other parameters,

respectively. We also considered region- and settlement-level random effects, $\varepsilon_{reg(j)}$ and

$\varepsilon_{set(j)}$, following normal prior with mean zero and variance $\sigma_r^2$ and $\sigma_s^2$, respectively.

Overall, $p_{ij}$ was modeled as follows:

$\text{logit}(p_{ij}) = \beta_{0i} + \beta_{1i}ABAN_j + \beta_{2i}TEMP_j + \beta_{3i}MONTH_j + \beta_{4i}I(LU_j = \text{"dry field"}) +$

$\beta_{5i}I(LU_j = \text{"built-up area"}) + \varepsilon_{reg(j)} + \varepsilon_{set(j)},$     Eqn. 1

where $\beta_{0i}$ is an intercept, and $\beta_{1i}$, $\beta_{2i}$, $\beta_{3i}$, $\beta_{4i}$ and $\beta_{5i}$ are regression coefficients for the $i$th

species. $I()$ is an indicator function indicating dummy variables for $LU_j$.

To infer species-specific effects of land abandonment and the correlation with

their suitable temperature, a GLMM with a random slope was applied to estimate

regression coefficients for each species. In a random slope model, regression

coefficients of species are treated as random effects subject to the same prior

distribution. As a prior distribution of $\boldsymbol{\beta}_i = (\beta_{1i}, \beta_{2i})$, the bivariate normal distribution

$MVN(\boldsymbol{\beta}_i; \boldsymbol{\mu_\beta}, \boldsymbol{\Sigma_\beta})$ was applied. $\boldsymbol{\mu_\beta} = (\mu_{\beta 1}, \mu_{\beta 2})$ and $\boldsymbol{\Sigma_\beta}$ are the mean vector and

covariance matrix, respectively. Covariance matrix was decomposed to standard

deviations and Pearson's correlation coefficient as follows:

$$\boldsymbol{\Sigma_\beta} = \begin{bmatrix} \sigma_{\beta 1}^2 & \rho\sigma_{\beta 1}\sigma_{\beta 2} \\ \rho\sigma_{\beta 1}\sigma_{\beta 2} & \sigma_{\beta 2}^2 \end{bmatrix}$$

The $\rho$ is Pearson's correlation coefficient of $\beta_{1i}$ and $\beta_{2i}$; a positive value indicates that a

species preferring cooler temperatures is more likely to be negatively impacted by land

abandonment. The priors of regression coefficients of confounding factors, $\beta_{3i}$, $\beta_{4i}$ and

$\beta_{5i}$ were normal distributions with mean $\mu_{\beta 3}$, $\mu_{\beta 4}$ and $\mu_{\beta 5}$, variance $\sigma_{\beta 3}$, $\sigma_{\beta 4}$ and $\sigma_{\beta 5}$,

respectively.

We calculated posterior distributions of the model by the Bayesian method

using vague or weakly informative priors for the hyperparameters of the model. For $\beta_{0i}$,

$\mu_{\beta 1}$, $\mu_{\beta 2}$, and $\mu_{\beta 3}$, $\mu_{\beta 4}$ and $\mu_{\beta 5}$, a normal prior with mean 0 and variance 100 was used. A

half-Cauchy distribution with a scale parameter set to 5 was used for the standard

deviations as a weakly informative prior; $\sigma_r$, $\sigma_s$, $\sigma_{\beta 1}$, $\sigma_{\beta 2}$, $\sigma_{\beta 3}$, $\sigma_{\beta 4}$ and $\sigma_{\beta 5}$ [38,39], was

used. Uniform distribution with range (-1, 1) was used for the prior of correlation

coefficient, $\rho$. Samples from the posterior distribution were obtained using No-U-Turn

Sampler implemented in RStan 2.21.1 [40] (three chains, 1000 iterations after 1000

burn-in iterations).

For the explanatory variable $ABAN_j$, we considered two candidate variables,

"abandoned or not" and "years since abandonment", because it is unknown whether the

response is immediate or gradual. We compared the performance of models using

"abandoned or not" and "years since abandonment" on the basis of a widely applicable

Bayesian information criterion (WBIC) [41]. WBIC is a generalization of Bayesian

information criterion that is applicable to both regular and singular statistical models; it

asymptotically approximates the minus logarithm of the marginal likelihood. A Bayes

factor (the ratio of posterior probability of competing models) was calculated by the

exponential of the difference in WBICs. The abandonment variable with a higher WBIC

was used for further analyses.

(2) Relationship between habitat traits of species and responses to land

abandonment

To test whether species with different habitat traits respond differently to

abandonment, we applied a hierarchical model incorporating a trait-species hierarchy

[42]. The model is an extension of the species-level model described in section (1), in

which the average effect of abandonment $\mu_{\beta1}$ was replaced with the linear predictor of

habitat traits as follows:

222             $\mu_{\beta1i} = \alpha_0 + \alpha_1 H_{ik}$,                                    Eqn. 2

where $H_{ik}$ is a binary variable indicating whether the $i$th species utilizes the $k$th ($k$ = 1, 2,

…, 11) habitat type or not. $\alpha_0$ is intercept, and $\alpha_1$ indicates the difference in the average

effects of abandonment between habitat types, where positive and negative values of $\alpha_1$

indicate positive and negative effects of abandonment for species utilizing the $k$th

habitat type relative to species utilizing other habitat types. The same structure was used

to determine the effect of $TEMP_j$; that is, $\mu_{\beta2i} = \alpha_{T0} + \alpha_{T1} H_{ik}$. As in the model of section

(1), slopes of each species vary following Multivariate normal distribution with mean

vector of ($\mu_{\beta1i}$, $\mu_{\beta2i}$).

We calculated posteriors of $\alpha_0$, $\alpha_1$, $\alpha_{T0}$, and $\alpha_{T1}$ separately for the 11 habitat

types. We used vague priors of a normal distribution with mean 0 and variance 1000 for

these parameters. Other prior settings were the same as those for the species-level model

(Section (1)). Samples from the posterior distribution were obtained using RStan 2.21.1

[40], with the same settings as in Section (1).

(3) Predictive model of species-specific responses to land abandonment

A trait-based approach is a powerful tool to predict species-specific responses

to environmental changes [43], and allows us to evaluate the broad-scale distribution of

a species. Here, we developed a predictive model based on the habitat trait of a species.

As most of the species have a habitat trait comprising multiple habitat types (i.e., they

inhabit several habitat types), the predictive model was a multivariate extension of Eqn.

2:

$$\mu_{\beta 1 i} = \alpha_0 + \Sigma_k \alpha_k H_{ik}.$$    Eqn. 3

However, models that are too complex can lead to overfitting and tend to have

a large predictive error. In such a situation, ridge regularization is useful to constrain

overfitting and improve predictive accuracy [44]. In a Bayesian manner, ridge

regularization can be simply implemented using the same normal prior, $N(0, \sigma_\alpha^2)$, for

each regression coefficient. This prior works as a penalty for the divergence of squared

regression coefficients from zero, and the standard deviation, $\sigma_\alpha$, corresponds to the

weight of the penalty. The prior of $\sigma_\alpha$ was set to a weakly informative prior,

half-Cauchy(0, 5), and $\sigma_\alpha$ was determined by calculating the posterior distribution. The

prior of $\alpha_0$ was a normal distribution with mean 0 and variance 100. Samples from the

posterior distribution were obtained by RStan 2.21.1 [40], with the same MCMC

sampling settings as in Sections (1) and (2).

To make projection maps for the effect of land abandonment on butterfly

assemblages, we defined the loss (CL) and gain (CG) in butterfly species due to land

abandonment as follows:

$$CL_j = \sum_i^{N_{sp}} \delta_{ij} \left| min\left(\mu_{\beta 1 i}, 0\right) \right|,$$

258    and

$$CG_j = \sum_i^{N_{sp}} \delta_{ij} max\left(\mu_{\beta 1 i}, 0\right),$$

259 where $N_{sp}$ is the number of species, and $\delta_{ij}$ is a binary indicator of whether the *i*th

260 species is present in the *j*th grid cell. Operators $min(\mu_{\beta 1i}, 0)$ and $max(\mu_{\beta 1i}, 0)$ are

261 filters to identify species affected negatively and positively, respectively. CL and CG

262 can be interpreted as log-odds ratios of occurrence probabilities across species after land

263 abandonment. Large values of CL or CG indicate a large loss or gain in species in the

264 local butterfly assemblage due to land abandonment, respectively.

265  We used nation-wide range maps of 70 butterfly species in Japan (Table S3),

266 which were projected by Kasada et al. [45] using distribution models (MaxEnt) from

267 butterfly records obtained from the fourth and fifth National Surveys on the Natural

268 Environment. Prior to calculate CL and CG, we excluded seven alpine butterflies,

269 *Carterocephalus palaemon*, *Erebia neriene*, *E. ligea*, *Oeneis norna*, *Colias palaeno*,

270 *Anthocharis cardamines* and *Aporia hippia*, because their ranges are irrelevant to land

271 abandonment. The range maps comprise presence and absence data for species in each

272 grid cell of a "1 × 1 km mesh (the third mesh)," a Japanese national grid system whose

273 unit cell size is $30''$ in latitude and $45''$ in longitude, approximately 1 × 1 km. We

274 determined the expected coefficient of land abandonment, $\mu_{\beta 1}$, using Eqn. 3, for 70

275 species. We then separately evaluated CL and CG for all the species and for species

276 ranked "near threatened (NT)", "vulnerable (VU)", "endangered (EN)" or "critically

277 endangered (CR)" in the red list of Japan in 2019

278 (https://www.env.go.jp/press/files/jp/110615.pdf, visited 10/19/2020). Losses and gains

279 of species were estimated separately, rather than as a combined gain-minus-loss,

280 because separate estimation is more informative for decisions regarding conservation

281 practices.

## Results

A total of 49 butterfly species were recorded during our field surveys (Table

S2). The top 10 species with the highest average occurrence (i.e., number of presence

spots/total number of spots) were *Pieris melete*, *Colias erate*, *Celastrina argiolus*,

*Ypthima argus*, *Libythea celtis*, *Plebejus argus*, *Lycaena phlaeas*, *Eurema mandarina*,

*Papilio machaon* and *Apatura metis*, in decreasing order, and they represented 73.4% of

the total occurrence records.

The model selection by WBIC indicated that the model including "abandoned

or not" (WBIC = 3575.77) outperformed the model for "years since abandonment"

(WBIC =3608.51), suggesting relatively immediate responses to land abandonment. The

Bayes factor for the former model against the latter was $1.65 \times 10^{14}$. Among 49 species

of butterflies analyzed, 12 showed a significant negative response (i.e., the 95% CI of

the regression coefficient of "abandoned or not" was less than zero), while 3 species

showed a positive response (Table S4). *Lycaena phlaeas, Papilio machaon* and *Papilio*

*xuthus* exhibited the largest negative effect of abandonment. The three species that

showed positive response were *Limenitis camilla*, *Cyrestis thyodamas* and *Graphium*

*sarpedon*. Effects of abandonment and temperature were positively correlated (Fig. 2);

Pearson's correlation coefficient for the effect of land abandonment and annual mean

temperature was 0.402 with a 95% CI of (0.0414, 0.678).

There was a significant relationship between habitat types of species and

responses to land abandonment for four habitat types; species utilizing farmland,

grassland and built-up areas responded negatively, while those utilizing forest

responded positively (Fig. 3, Table S5).

Using a hierarchical model with ridge regularization, we obtained the following

predictive model for the effect of land abandonment based on the habitat trait of each

species:

[predicted effect of land abandonment]

=    –0.44 – 0.39 × [built-up area] – 0.99 × [grassland] – 0.40 × [garden]

– 0.74 × [farmland] – 0.38 × [alpine area] – 0.36 × [rocky area] + 0.34 × [wetland]

+ 0.06 × [riverside] + 0.54 × [open forest] + 0.58 × [forest edge]

+ 0.56 × [forest].        Eqn. 4

By applying this model, 37 of 70 species showed negative responses. Using the

predicted values thus obtained, we drew maps of CL (Fig. 4a) and CG (Fig.4b) of

species throughout Japan's mainland. The geographic distributions of loss and gain were

similar in a coarse scale and the Pearson's correlation coefficient was 0.56; that is,

larger values of both CL and CG were found in inland middle Kyushu, inland western

Honshu, inland middle Honshu, northern Honshu and Hokkaido, while smaller values

were found in high elevation areas. However, there was some contrast between CL and

CG in lowland where small CG relative to CL was found. Standard deviations of CL

and CG (Fig. S1ab) were almost proportional to the CL and CG themselves (Pearson's

correlation coefficient was 0.86 and 0.99, respectively).

The 10 and 4 red list species were evaluated as victims and beneficiaries to

land abandonment, respectively. Distribution patterns of CL and CG for red list species

of Japan were shown in Fig. 4cd. The correlation between CL and CG for red list

species was weaker than that of all the species (Pearson's correlation coefficient = 0.36),

and their spatial patterns are different in a coarse scale. For instance, CL was small

while CG was large in northern Honshu. On the contrary, CL was large but CG was

small in eastern Hokkaido. Similar to the maps for all the species, standard deviations

(Fig. S1cd) were almost proportional to the CL and CG themselves (Pearson's

correlation coefficient was 0.95 and 0.99, respectively).

## Discussion

As expected, the responses of butterflies to land abandonment were associated

with climatic niches and habitat traits of species. Our study approach that compared

community assemblages between inhabited and abandoned areas using a standard survey

protocol enabled us to estimate the effects of land abandonment over broad range of

climatic condition. In combination with trait-based approach, we were able to draw

nation-wide potential maps of loss and gain of species richness through land

abandonment.

Legacy of past climate and land use can affect current functional species

composition within regional species pool [46,47], and different functional compositions

among region can result in different community-level responses to current

environmental change [48]. Such a difference would also reflect historical background of

organisms inhabiting seminatural environments in Japan. Grassland was more abundant

in Pleistocene due to cold and dry climate than at present [30]. After the last glacier

period, open lands created by human activity have offered refugia for the Pleistocene

relics [49], and such refugia become unlikely to be sustainable if they are abandoned [50].

Our results also raise a conservational concern that species declined by land abandonment

will decline more due to climate warming in the future, because the species susceptible to

land abandonment tend to prefer low temperatures (Fig.2). In turn, this means that

continuing land management in rural areas will be an effective measure of climate change

adaptation for biodiversity conservation in Japan.

Decrease in open habitat species (grassland, crop field, and residential species)

(Figure 4) would be  related to the decreased species richness of herbaceous plants [23],

leading to the loss or decline of host plants for these butterflies. Indeed, we often

observed tall grasses, such as *Miscanthus sinensis*, as predominant species in abandoned

paddy, dry fields, and built-up area in uninhabited settlements. The monodominance of

wind-pollinated *Miscanthus* would also decrease abundance and diversity of nectaring

flowers that are important for adult butterflies. In contrast, forest species tended to

increase in response to abandonment, although the proportion of such species was low.

Development of old-growth forest in Japan requires approximately 150 years [51], and

some tree species require more than a hundred years to recover [52]; accordingly,

recovery of forest butterfly species depending on such trees would also require very long

time as well.

Comparing with land abandonment and temperature, local land use (paddy field,

dry crop field and built-up area) had minor impact on butterfly occurrences. This is

probably due to our focus on adult butterflies which were often affected by

landscape-scale effect of land use [53]. In our study, the effect of landscape scale

abandonment was actually more prominent than local factors.

It should be noted that there was a positive correlation between losses and gains

in butterfly richness both for all species and for red list species due to land abandonment

at the level of grid cells, indicating a tradeoff between the benefits of the active

management of abandoned sites and letting succession proceed. As endangered

butterflies in Japan include more grassland species than forest species, human

intervention should be prioritized, in principle, to conserve such species, but deliberate

decision-making is required for areas where the benefit of abandonment is high. Our

results indicated that conservation of butterflies by active management is desirable in

lowland and eastern Hokkaido, in which loss of species preferring abandoned landscape

is small by such management. The identification of highly cost-effective areas for habitat

management based on species ranges and habitat preferences, with a high resolution, is

necessary, because maintaining secondary natural environments is generally

labor-intensive [54]. It is of course necessary to conduct similar analyses using taxonomic

groups other than butterflies to enable more comprehensive assessments as effect of

abandonment differs among taxonomic groups [13].

Here, we evaluated the risks and benefits of land abandonment using butterfly

occurrence data; however, it is also important to clarify how the abundance of larval host

plants and adult nectar plants change over time after abandonment. Previous studies have

shown that many grassland species are still found even 10 years after abandonment [55],

indicating the presence of host and nectar plants during that period. In our study, the

effects of abandonment did not show a clear increase over time, as the model with a

dichotomous variable (inhabited vs. abandoned settlements) outperformed the model

with a continuous variable (years since abandonment). As Japan has a monsoon climate

with a relatively high temperature and rainfall in comparison with Europe, vegetation

succession may proceed within a few years, resulting in a rapid response to abandonment.

This suggests the necessity to restore habitats as early as possible after abandonment. The

potential maps of effect of abandonment will contribute to prioritize habitat management

plans in areas where settlement abandonment is expected in the immediate future and a

large decrease in biodiversity after abandonment.

A major contribution of our study was the ability to apply the estimated

relationship between habitat traits and responses to land abandonment to butterfly species

that were not observed in the field survey. The utility of such a trait-based approach is

widely recognized [56], but there are some caveats, as responses of habitat generalists

may be difficult to predict. To resolve this problem, models incorporating responses of

host plants to abandonment and resultant butterfly responses should be explicitly

constructed, as has been demonstrated in climate change prediction models [57].

In conclusion, we demonstrated that a hierarchical trait-based approach proves

to be useful to scale up the results of comparative field studies to a macro-scale evaluation

of species losses and gains by land abandonment. Land abandonment had both positive

and negative effects on different functional groups of butterflies in Japan, and our

approach could be used to identify areas exhibiting high species losses relative to gains

after land abandonment. We believe this approach is promising for large-scale and

comprehensive biodiversity assessments regarding the risks and benefits of land

abandonment, which are urgently needed in view of ongoing depopulation in developed

countries.

Ethics

Our field survey was conducted from roads and rural paths, and no permission was

required in compliance with Japanese law. All information of site locations is coarse

enough to protect landowners' privacy. To avoid negative consequences on conservation,

we kept specific records of endangered species closed.

Data accessibility

The datasets and program codes are available from the Dryad Digital Repository at

https://datadryad.org/stash/share/8u3_mEXi3SsWzm_ZElaD7iq7wWB8FSQ45czwu8u

_V24, except specific records of endangered species.

Author contributions

NS, KF, and TM conceived the ideas; NS conducted the field study; AA prepared part of

the data for the analysis; MK and MM prepared the projected national distribution maps

of butterflies; NS and KF conducted the analysis; and NS, KF, and TM wrote the

manuscript with input from AA, MK and MM.

Competing interest

We have no competing interest.

Funding

This work was supported in part by the Environment Research and Technology

Development Fund (S9) of the Ministry of the Environment, Japan, and JSPS

KAKENHI Grant Number 16K16223 and 21H03656.

References

1.    Fischer J, Hartel T, Kuemmerle T. 2012 Conservation policy in traditional farming

landscapes. *Conserv. Lett.* **5**, 167–175. (doi:10.1111/j.1755-263X.2012.00227.x)

2.    Benayas JMR, Martins A, Nicolau JM, Schulz JJ. 2007 Abandonment of

agricultural land: an overview of drivers and consequences. *CAB Rev. Perspect.*

*Agric. Vet. Sci. Nutr. Nat. Resour.* **2**, 1–14.

3.    Fukamachi K, Oku H, Nakashizuka T. 2001 The change of a satoyama landscape

and its causality in Kamiseya, Kyoto Prefecture, Japan between 1970 and 1995.

*Landsc. Ecol.* **16**, 703–717.

4.    Schmitt T, Rákosy L. 2007 Changes of traditional agrarian landscapes and their

conservation implications: A case study of butterflies in Romania. *Divers. Distrib.*

**13**, 855–862. (doi:10.1111/j.1472-4642.2007.00347.x)

5.    Henle K *et al.* 2008 Identifying and managing the conflicts between agriculture

and biodiversity conservation in Europe–A review. *Agric. Ecosyst. Environ.* **124**,

60–71.

6.    Corlett RT. 2016 Restoration, Reintroduction, and Rewilding in a Changing World.

*Trends Ecol. Evol.* **31**, 453–462. (doi:10.1016/j.tree.2016.02.017)

7.    Svenning J-C *et al.* 2016 Science for a wilder Anthropocene: Synthesis and future

directions for trophic rewilding research. *Proc. Natl. Acad. Sci.* **113**, 898–906.

(doi:10.1073/pnas.1502556112)

8.    Scozzafava S, De Sanctis A. 2006 Exploring the effects of land abandonment on

habitat structures and on habitat suitability for three passerine species in a highland

area of Central Italy. *Landsc. Urban Plan.* **75**, 23–33.

459   (doi:10.1016/j.landurbplan.2004.10.006)

460 9. Herrando S, Brotons L, Anton M, Paramo F, Villero D, Titeux N, Quesada J,

461   Stefamescu C. 2015 Assessing impacts of land abandonment on Mediterranean

462   biodiversity using indicators based on bird and butterfly monitoring data. *Environ.*

463   *Conserv.* **FirstView**, 1–10. (doi:doi:10.1017/S0376892915000260)

464 10. Randin CF, Dirnböck T, Dullinger S, Zimmermann NE, Zappa M, Guisan A. 2006

465   Are niche-based species distribution models transferable in space? *J. Biogeogr.* **33**,

466   1689–1703. (doi:10.1111/j.1365-2699.2006.01466.x)

467 11. Wenger SJ, Olden JD. 2012 Assessing transferability of ecological models: An

468   underappreciated aspect of statistical validation. *Methods Ecol. Evol.* **3**, 260–267.

469   (doi:10.1111/j.2041-210X.2011.00170.x)

470 12. Queiroz C, Beilin R, Folke C, Lindborg R. 2014 Farmland abandonment: Threat or

471   opportunity for biodiversity conservation? A global review. *Front. Ecol. Environ.*

472   **12**, 288–296. (doi:10.1890/120348)

473 13. Koshida C, Katayama N. 2018 Meta-analysis of the effects of rice-field

474   abandonment on biodiversity in Japan. *Conserv. Biol.* **32**, 1392–1402.

475   (doi:10.1111/cobi.13156)

476 14. Stewart G. 2010 Meta-analysis in applied ecology. *Biol. Lett.* **6**, 78–81.

477   (doi:10.1098/rsbl.2009.0546)

478 15. Fukasawa K, Mishima Y, Yoshioka A, Kumada N, Totsu K, Osawa T. 2016

479   Mammal assemblages recorded by camera traps inside and outside the evacuation

480   zone of the Fukushima Daiichi Nuclear Power Plant accident. *Ecol. Res.* **31**, 493.

481   (doi:10.1007/s11284-016-1366-7)

16.  Dyer C, Jones R. 2010 *Deserted Villages Revisited*. University of Hertfordshire

Press. See https://books.google.co.jp/books?id=JLcfFqRel3MC.

17.  Webb RH, Wilshire HG. 1980 Recovery of soils and vegetation in a Mojave desert

ghost town, Nevada, U.S.A. *J. Arid Environ.* **3**, 291–303.

(doi:10.1016/s0140-1963(18)31634-3)

18.  Ohwaki A, Ogawa H, Taketani K, Tomisawa A. 2014 Butterfly responses to

cultivated field abandonment are related with ecological traits in a temperate

Japanese agricultural landscape. *Landsc. Urban Plan.* **125**, 174–182.

(doi:10.1016/j.landurbplan.2014.02.024)

19.  Slancarova J, Bartonova A, Zapletal M, Kotilinek M, Fric ZF, Micevski N, Kati V,

Konvicka M. 2016 Life history traits reflect changes in mediterranean butterfly

communities due to forest encroachment. *PLoS One* **11**, 1–18.

(doi:10.1371/journal.pone.0152026)

20.  Ogura J. 2006 The transition of grassland area in Japan. *J. Kyoto Seika Univ.* **30**,

159–172.

21.  Ministry of Agriculture, Forestry and Fisheries. 2016 On the present condition of

abandoned farmland and the countermeasures. See

http://www.maff.go.jp/j/nousin/tikei/houkiti/pdf/2804_genjo.pdf (accessed on 16

February 2016).

22.  Ministry of Land, Infrastructure, Transport and Tourism. 2014 On the estimation

of future population in Japan by 1 km$^2$ grid. See

http://www.mlit.go.jp/kokudoseisaku/kokudoseisaku_tk3_000044.html (accessed

on 13 June 2017).

23.   Uchida K, Ushimaru A. 2014 Biodiversity declines due to abandonment and

506           intensification of agricultural lands: Patterns and mechanisms. *Ecol. Monogr.* **84**,

507           637–658. (doi:10.1890/13-2170.1)

508    24.   Uchida K, Ushimaru A. 2015 Land abandonment and intensification diminish

509           spatial and temporal ??-diversity of grassland plants and herbivorous insects

510           within paddy terraces. *J. Appl. Ecol.* **52**, 1033–1043.

511           (doi:10.1111/1365-2664.12443)

512    25.   Uchida K, Takahashi S, Shinohara T, Ushimaru A. 2016 Threatened herbivorous

513           insects maintained by long-term traditional management practices in semi-natural

514           grasslands. *Agric. Ecosyst. Environ.* **221**, 156–162.

515           (doi:10.1016/j.agee.2016.01.036)

516    26.   Uchida K, Koyanagi TF, Matsumura T, Koyama A. 2018 Patterns of plant

517           diversity loss and species turnover resulting from land abandonment and

518           intensification in semi-natural grasslands. *J. Environ. Manage.* **218**, 622–629.

519           (doi:10.1016/j.jenvman.2018.04.059)

520    27.   Koyanagi TF, Furukawa T. 2013 Nation-wide agrarian depopulation threatens

521           semi-natural grassland species in Japan: Sub-national application of the Red List

522           Index. *Biol. Conserv.* **167**, 1–8. (doi:10.1016/j.biocon.2013.07.012)

523    28.   Nakamura Y. 2011 Conservation of butterflies in Japan: Status, actions and

524           strategy. *J. Insect Conserv.* **15**, 5–22. (doi:10.1007/s10841-010-9299-x)

525    29.   Hiura I. 1971 An analysis of butterfly fauna of Japan from the standpoint of

526           historical biogeography. *Spec. Bull. Lepidopterol. Soc. Japan* **5**, 73–88.

30.   Yamaura Y *et al.* 2019 Genomic reconstruction of 100 000-year grassland history

in a forested country: population dynamics of specialist forbs. *Biol. Lett.* **15**,

20180577. (doi:10.1098/rsbl.2018.0577)

31.    Japan Butterfly Conservation Society. 2012 *Field guide of butterflies in Japan*.

Tokyo: Seibundo Shinko-sha.

32.    Asahara A. 2010 *Landscape of abandoned and depopulated settlements 4*.

Saitama: Heyaneko.

33.    Asahara A. 2011 *Landscape of abandoned and depopulated settlements 5*.

Saitama: Heyaneko.

34.    Murakage Y. 2009 Murakage Yataro's travelogue of the villages [Murakage

Yataro no Shuraku Kikou]. See http://www.aikis.or.jp/~kage-kan/ (accessed on 27

February 2020).

35.    Ryan SF *et al.* 2019 Global invasion history of the agricultural pest butterfly Pieris

rapae revealed with genomics and citizen science. *Proc. Natl. Acad. Sci. U. S. A.*

**116**, 20015–20024. (doi:10.1073/pnas.1907492116)

36.    Bolker BM, Brooks ME, Clark CJ, Geange SW, Poulsen JR, Stevens MHH, White

J-SS. 2009 Generalized linear mixed models: a practical guide for ecology and

evolution. *Trends Ecol. Evol.* **24**, 127–135.

(doi:http://dx.doi.org/10.1016/j.tree.2008.10.008)

37.    Japan Meteorological Business Support Center. 2002 *Mesh climatic data of Japan*.

Japan Meteorological Agency.

38.    Gelman A. 2006 Prior distributions for variance parameters in hierarchical models.

*Bayesian Anal.* **1**, 515–534. (doi:10.1214/06-BA117A)

39.    Van Dongen S. 2006 Prior specification in Bayesian statistics: Three cautionary

tales. *J. Theor. Biol.* **242**, 90–100. (doi:http://doi.org/10.1016/j.jtbi.2006.02.002)

40.   Stan Development Team. 2016 RStan: the R interface to Stan. R package version

2.14.1.

41.   Watanabe S. 2013 A widely applicable Bayesian information criterion. *J. Mach.*

*Learn. Res.* **14**, 867–897.

42.   Pollock LJ, Morris WK, Vesk PA. 2012 The role of functional traits in species

distributions revealed through a hierarchical model. *Ecography (Cop.).* **35**,

716–725. (doi:10.1111/j.1600-0587.2011.07085.x)

43.   Öckinger E *et al.* 2010 Life-history traits predict species responses to habitat area

and isolation: A cross-continental synthesis. *Ecol. Lett.* **13**, 969–979.

(doi:10.1111/j.1461-0248.2010.01487.x)

44.   Hooten MB, Hobbs NT. 2015 A guide to Bayesian model selection for ecologists.

*Ecol. Monogr.* **85**, 3–28. (doi:10.1890/14-0661.1)

45.   Kasada M, Matsuba M, Miyashita T. 2017 Human interest meets biodiversity

hotspots: A new systematic approach for urban ecosystem conservation. *PLoS One*

**12**, e0172670. (doi:10.1371/journal.pone.0172670)

46.   Blonder B *et al.* 2018 Late Quaternary climate legacies in contemporary plant

functional composition. *Glob. Chang. Biol.* **24**, 4827–4840.

(doi:10.1111/gcb.14375)

47.   Perring MP, De Frenne P, Baeten L, Maes SL, Depauw L, Blondeel H, Carón MM,

Verheyen K. 2016 Global environmental change effects on ecosystems: the

importance of land-use legacies. *Glob. Chang. Biol.* **22**, 1361–1371.

(doi:10.1111/gcb.13146)

48.  Svenning J-C, Normand S, Skov F. 2009 Plio-Pleistocene climate change and

geographic heterogeneity in plant diversity–environment relationships. *Ecography*

*(Cop.).* **32**, 13–21. (doi:10.1111/j.1600-0587.2008.05732.x)

49.  Ushimaru A, Uchida K, Suka T. 2018 Grassland biodiversity in Japan: threats,

management and conservation. In *Grasslands of the world: diversity, management*

*and conservation* (eds VR Squires, J Dengler, L Hua, H Feng), pp. 197–218. Boca

Raton: CRC press. (doi:10.1201/9781315156125)

50.  Fukasawa K, Akasaka T. 2019 Long-lasting effects of historical land use on the

current distribution of mammals revealed by ecological and archaeological

patterns. *Sci. Rep.* **9**, 1–11. (doi:10.1038/s41598-019-46809-1)

51.  Yamaura Y, Lindenmayer D, Yamada Y, Gong H, Matsuura T, Mitsuda Y, Masaki

T. 2020 A spatially explicit empirical model of structural development processes

in natural forests based on climate and topography. *Conserv. Biol.* **34**, 194–206.

(doi:https://doi.org/10.1111/cobi.13370)

52.  Ohtani S, Koike F, Townsend P. 2005 Implications of 19th century landscape

patterns for the recovery of Fagus crenata forests. *Appl. Veg. Sci.* **8**, 125–132.

(doi:10.1658/1402-2001(2005)008[0125:IOTCLP]2.0.CO;2)

53.  Bergman K-O, Askling J, Ekberg O, Ignell H, Wahlman H, Milberg P. 2004

Landscape effects on butterfly assemblages in an agricultural region. *Ecography*

*(Cop.).* **27**, 619–628. (doi:https://doi.org/10.1111/j.0906-7590.2004.03906.x)

54.  Takeuchi K, Brown RD, Washitani I, Tsunekawa A, Yokohari M. 2012 *Satoyama:*

*the traditional rural landscape of Japan.* Springer Science & Business Media.

55.  Krauss J *et al.* 2010 Habitat fragmentation causes immediate and time-delayed

biodiversity loss at different trophic levels. *Ecol. Lett.* **13**, 597–605.

(doi:10.1111/j.1461-0248.2010.01457.x)

56.   Webb CT, Hoeting JA, Ames GM, Pyne MI, LeRoy Poff N. 2010 A structured and

dynamic framework to advance traits-based theory and prediction in ecology. *Ecol.*

*Lett.* **13**, 267–283. (doi:10.1111/j.1461-0248.2010.01444.x)

57.   Araújo MB, Luoto M. 2007 The importance of biotic interactions for modelling

species distributions under climate change. *Glob. Ecol. Biogeogr.* **16**, 743–753.

(doi:10.1111/j.1466-8238.2007.00359.x)

Fig. 1 Location of study sites. Each site contained both rural abandoned villages and inhabited villages.

Fig. 2 Relationship between coefficients of mean annual temperature and land abandonment. Black dots and gray error bars indicate posterior means and 95% credible intervals of species, respectively. The dashed ellipsoid and line are 95% equiprobable ellipsoid and principal axis of coefficients.

Fig. 3 Effects of land abandonment on the occurrence of butterflies with different habitat characteristics. The 1 and 0 of x axis indicate the species does and does not utilize the habitat type, respectively. Gray dots represent coefficients of abandonment ($\beta_{1i}$) for each species. Black dots and error bars represent posterior means and 95% credible interval (95% CI) of the expected value of the coefficient of abandonment ($\mu_{\beta 1i}$ in eqn. 1), respectively. Four habitat types with significant effect on $\mu_{\beta 1i}$ (i.e. 95% CI of $\alpha_1$ in eqn. 2 did not overlap 0) were shown.

Fig. 4 Maps of cumulative loss (CL) and gain (CG) in butterfly species richness due to land abandonment for all the species(a and b, respectively) and for the Red List species (c and d, respectively). The boxed graphs are hexagonal binning plots showing their relationship with elevation.

Fig. 4 *Continued.*