

1 **Parallel and non-parallel divergence within polymorphic populations of brook stickleback,**
2 ***Culaea inconstans* (Actinopterygii: Gasterosteidae)**

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11 Running title: Divergence within polymorphic brook stickleback populations

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13

14 **Abstract**

15 Studying parallel evolution allows us to draw conclusions about the repeatability of
16 adaptive evolution. Whereas populations likely experience similar selective pressures in similar
17 environments, it is not clear if this will always result in parallel divergence of ecologically
18 relevant traits. Our study investigates the extent of parallelism associated with the evolution of
19 pelvic spine reduction in brook stickleback populations. We find that populations with parallel
20 divergence in pelvic spine morphology do not exhibit parallel divergence in head and body
21 morphology but do exhibit parallel divergence in diet. In addition, we compare these patterns
22 associated with pelvic reduction in brook stickleback to well-studied patterns of divergence
23 between spined and unspined threespine stickleback. Whereas spine reduction is associated with
24 littoral habitats and a benthic diet in threespine stickleback, spine reduction in brook stickleback
25 is associated with a planktonic diet. Hence, we find that pelvic spine divergence is associated
26 with largely non-parallel ecological consequences across species.

27

28 **Keywords**

29 Adaptation, balancing selection, Gasterosteidae, pelvic spines, polymorphism, North America,
30 trophic morphology, lakes, diet, isotopes

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32

33 **Introduction**

34 Understanding the causes of adaptive phenotypic divergence is a central goal in
35 evolutionary biology. The genetic basis of adaptive divergence has received much needed
36 attention recently, but the ecological causes of selection on divergent phenotypes remain
37 unknown in many cases. Whereas understanding the genetic basis of phenotypic variation is
38 central to many questions in evolutionary biology, understanding the ecological correlates of
39 adaptive variation is key to answering questions about the causes of selection that drive
40 diversification. In populations with persistent adaptive phenotypic dimorphism, selection may
41 involve several ecological factors simultaneously, including predation, competition, or
42 parasitism, and selection driving divergence in one trait may affect divergence in other traits due
43 to pleiotropy. Alternatively, a suite of traits may change in concert with a particular adaptive
44 polymorphism due to plastic or evolutionary responses in resource use, habitat use, behaviour, or
45 predator interactions that arise as a consequence of the trait polymorphism. For example, cliff
46 nesting in some kittiwake gulls (*Rissa tridactyla* Linnaeus 1758) results in reduced predation risk
47 and reduced predator avoidance behaviours (Cullen 1956). The diurnal life history of butterflies
48 likely originated as a strategy to avoid bat predation and is associated with the loss of ultrasonic
49 hearing (Yack and Fullard 2008). In threespine stickleback (*Gasterosteus aculeatus* Linnaeus
50 1758), predators induce selection for longer protective spines, which are part of a suite of traits
51 associated with habitat divergence within some freshwater lakes (Rennison et al. 2019). In all of
52 these examples, predator-related differences in phenotype are associated with other ecological or
53 behavioural differences. Sticklebacks, in particular, have emerged as a model system in the study
54 of adaptive divergence, particularly with respect to the role of predation in driving adaptive
55 divergence.

56 Three species in the stickleback family (Gasterosteidae), in three different genera
57 (*Gasterosteus*, *Pungitius*, and *Culaea*), exhibit heritable variation in their pelvic or dorsal spines
58 (Nelson 1969, Nelson 1977, Chan *et al.* 2010). There is good evidence that spines are involved,
59 at least to some extent, in defence against predators (Hoogland *et al.* 1957, Hall 1956, Reisman
60 and Cade 1967). For example, gape-limited predators (e.g. some birds, trout, and other small or
61 medium-sized fishes) are deterred by spines, and these predators select for increased armor and
62 longer spines in stickleback (Reist 1980a, Reimchen 1992, Vamosi and Schluter 2004,
63 Marchinko 2008, Lescak and von Hippel 2011, Miller *et al.* 2017). Spined stickleback are bolder
64 and will tolerate being closer to predators than unspined stickleback (Reist 1980a, Reist 1980b,
65 Reist 1983). Large fish, such as pike (*Esox lucius* Linnaeus 1758), which are not gape-limited,
66 are less deterred by armor and spines. Hence, predators that are not gape-limited likely select for
67 spine reduction or spine loss (Nelson and Atton 1971, Andraso and Barron 1995, Leinonen *et al.*
68 2011). Anecdotal evidence has suggested that invertebrate predators, such as dragonfly nymphs
69 (Odonata: Anisoptera) and giant water bugs (Hemiptera: Belostomatidae), use spines and other
70 external bony structures to grasp their prey, suggesting that invertebrates may select for spine
71 reduction or loss (Reimchen 1980, Reist 1980b, Vamosi 2002, Lescak *et al.* 2012), but a meta-
72 analysis of invertebrate selection experiments, found little support for selection against spines
73 (Miller *et al.* 2017). Several other predators are known to prey on stickleback (Reimchen 1994),
74 including loons (*Gavia immer* Brunnich 1764; Reimchen 1980), muskrats (*Ondatra zibethicus*
75 Linnaeus 1766; Nelson 1977), and conspecific stickleback (Foster *et al.* 1988), but the influence
76 of each of these predators on stickleback spines is difficult to predict and has not been evaluated.
77 The evidence that predation influences selection on pelvic phenotypes in stickleback is further
78 supported by the observation that the size of pelvic girdle and pelvic spines in stickleback is

79 proportional to the density of predatory fish in a region (Miller *et al.* 2017). Insofar as different
80 predators use different habitats within lakes (Reimchen 1994), predator-mediated selection may
81 drive spined and unspined stickleback into different habitats.

82 Pelvic spines have received much more attention than dorsal spines in the stickleback
83 literature. Ancestrally, all species of stickleback had a pelvic structure composed of a pelvic
84 girdle and two spines (Bell 1974, Ward and McLennan 2009). Loss of the pelvic spines and
85 pelvic girdle has evolved in hundreds of populations, and many individuals have intermediate or
86 ‘vestigial’ pelvic structures such as half a girdle with only one spine or a complete pelvic girdle
87 with no spines (Klepaker 2013). Divergence in pelvic phenotype within stickleback populations
88 may be associated with resource competition leading to different habitat use (Nelson 1977,
89 Schulte 1994), but balancing selection caused by different predators that select for different
90 traits has been implicated as the main driver of pelvic divergence within and among stickleback
91 populations (Nelson 1969, Nelson and Atton 1971, Reimchen 1980, Reist 1980a, Reist 1980b,
92 Reimchen 1994, Marchinko 2009). In the absence of predators, pelvic reduction is associated
93 with low-calcium environments (Bell *et al.* 1993), but it is unlikely that variation in calcium
94 availability among habitats within a population would be sufficient to drive within-population
95 spine polymorphism. Populations with pelvic reduction are more often found in lakes which lack
96 an outlet (Nelson and Atton 1971), suggesting, perhaps, that the presence of multiple habitats
97 within a lake are required for pelvic spine divergence. Regardless of whether predation was the
98 initial, or is the primary, cause of the pelvic phenotype divergence within stickleback
99 populations, predators may drive different pelvic phenotypes to use different habitats within a
100 population, and individuals that use different habitats are likely exposed to different
101 environmental effects (see, for example, Rennison *et al.* 2019). The maintenance of within-

102 population divergence may, therefore, be dependent on multiple ecological factors, and adaptive
103 phenotypic divergence in pelvic morphology may have consequences on a variety of ecologically
104 relevant traits.

105 We studied populations of brook stickleback (*Culaea inconstans* Kirtland, 1840) to
106 investigate the hypothesis that divergent pelvic phenotypes are associated with divergence in
107 habitat. Pelvic spine polymorphism in threespine stickleback (*Gasterosteus aculeatus* Linnaeus
108 ,1958) has been studied extensively, and, in a few well-studied cases, threespine stickleback with
109 dimorphic pelvic phenotypes are reproductively isolated sympatric ecomorphs (McPhail 1984,
110 Ridgeway and McPhail 1984, McPhail 1992, Schluter and McPhail 1992, Nagel and Schluter
111 1998, Rundle et al. 2000). In the majority of polymorphic threespine stickleback populations,
112 however, vestigial pelvic phenotypes are more abundant than either the fully spined morph or the
113 unspined morph (Klepaker et al. 2013). In contrast, the vestigial morphs are absent or rare in
114 most polymorphic brook stickleback populations (Nelson and Atton 1971, Nelson 1977,
115 Klepaker et al. 2013). Also, unlike dimorphic threespine stickleback populations, pelvic
116 phenotypes in brook stickleback populations are not reproductively isolated, but, nonetheless,
117 have persisted over multiple generations at stable frequencies except where anthropogenic
118 environmental disturbances have occurred (Lowey et al. 2020). Dorsal and pelvic spines in brook
119 stickleback are longest in the southern parts of their distribution and shortest in the north (Nelson
120 1969), whereas clinal variation in spine length is not present in threespine. An additional notable
121 difference between brook stickleback and other stickleback species with respect to spine
122 reduction is that there are no marine populations of brook stickleback, and, therefore, spine
123 reduction in brook stickleback is not associated with freshwater colonization (Nelson 1969)

124 We assessed body shape variation associated with pelvic spine polymorphism in brook
125 stickleback because organisms' body shapes can be substantially influenced by being exposed to
126 different environments or habitats, and variation in body shape can reflect important ecological
127 and behavioural differences among individuals (Bell and Foster 1994, Reimchen et al. 1985,
128 Webster 2011). In addition, it is possible that the gene or genes involved in pelvic polymorphism
129 may have pleiotropic effects on other morphological traits. We also investigated habitat
130 divergence among brook stickleback pelvic phenotypes by analyzing stable isotope signatures. If
131 the divergent pelvic morphs of brook stickleback use different habitats, then diet is likely to
132 differ as a consequence. According to Cutting *et al* (2016), fish muscle biochemistry reflects a
133 long-term average diet (over a few months) and is a good indicator of diet source. Analysis of
134 carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$) isotopes in fish tissues is a common method to evaluate
135 variation in habitat and resource use in freshwater environments (Post 2002). Different
136 photosynthetic organisms (e.g. plants vs phytoplankton) fix carbon isotopes in different ratios.
137 For example, terrestrial plants tend to fix more ^{13}C and, as a result, have higher $\delta^{13}\text{C}$ values
138 than phytoplankton. When photosynthetic organisms are consumed, their carbon isotope ratios
139 are assimilated and reflected within consumers' tissues, and $\delta^{15}\text{N}$ values tend to increase with
140 trophic level due to preferential assimilation of ^{15}N (Jardine et al. 2003; Eloranta et al. 2010). If
141 phenotypically different brook stickleback forage in different habitats (e.g. limnetic vs. benthic),
142 they may have different isotopic signatures if primary producer composition is unique to either
143 habitat and if their diet shifts to higher-order consumers. Based on patterns of ecological
144 divergence between spined and unspined threespine stickleback (Reimchen 1980, McPhail
145 1992), we predict that brook stickleback with reduced pelvises will be associated with benthic or
146 littoral habitats and will have a more benthic diet, which would lead to higher $\delta^{13}\text{C}$ values and

147 higher $\delta^{15}\text{N}$ values among unspined individuals. We do not, however, have a specific prediction
148 about how body shape may change between spined and unspined brook stickleback morphs that
149 use different habitats because, unlike diet and stable isotope signatures, morphological
150 differences between stickleback in different habitats show little parallelism among lakes
151 (Kaeuffer et al. 2012).

152

153 **Methods**

154

155 *Sample preparation and collection*

156 We collected adult brook stickleback from two lakes in Alberta, Canada, in 2017 and
157 2019 (with UTF-8 encoded WGS84 latitude and longitude): Muir Lake (53.627659, -
158 113.957524) and Shunda Lake (52.453899, -116.146192). Shunda Lake was previously known
159 as Fish Lake (as in Nelson and Atton 1971, Nelson 1977). These lakes were selected because,
160 among the lakes with polymorphic brook stickleback populations in the region, they had a
161 relatively high abundance of spined and unspined pelvic morphs (Nelson 1977, Lowey et al.
162 2020). Both lakes have fish survey and stocking records indicating the potential presence of
163 several stickleback predators, including brown trout (*Salmo trutta* Linnaeus 1758), rainbow trout
164 [*Ocorhynchus mykiss* (Walbaum 1792)], brook trout (*Salvelinus fontinalis* Mitchill 1814),
165 northern pike (*Esox lucius* Linnaeus 1758), and yellow perch [*Perca flavescens* (Mitchill, 1814)],
166 although recent surveys and stocking records (i.e. since 1990) list only rainbow trout and brown
167 trout, suggesting that these salmonids are likely the dominant predatory fish (Alberta
168 Environment and Parks 2021). In Muir Lake, brook stickleback also coexist with fathead
169 minnow [*Pimephales promelas* (Rafinesque, 1820)], whereas in Shunda Lake the fish

170 community includes longnose sucker (*Catostomus catostomus* Forster 1773), white sucker
171 [*Catostomus commersonii* (Lacépède 1803)], and northern pearl dace [*Margariscus nachtriebi*
172 (Cox 1896)]. We observed loons [*Gavia immer* (Brunnich 1764)], dragonfly nymphs
173 (Gomphidae and other unidentified families), giant water bugs [*Lethocerus americanus* (Leidy
174 1847)], and backswimmers (Notonectidae) at both lakes. The distributions and abundances of
175 stickleback predators and competitors across habitats in these lakes has not been evaluated. Muir
176 lake has a maximum depth of 6.5m, water conductivity of 236 μ S/cm, and pH of 8.7 (measured in
177 the summer with water temperature of 18 degrees C), and, although undeveloped native
178 woodlands and residential development surround the lake, the predominant land use in the area is
179 agriculture (Alberta Environment and Parks 2021). Shunda Lake has a maximum depth of 6.2m,
180 water conductivity of 264.5 μ S/cm, and pH of 8.6 (measured in the summer with water
181 temperature of 16 degrees C), and is surrounded entirely by native woodland (Alberta
182 Environment and Parks 2021). Shunda Lake has an outlet stream, whereas Muir Lake does not.

183 Brook stickleback were collected in June and July using unbaited minnow traps (5 mm
184 mesh). To sample brook stickleback from the littoral zone, traps were set adjacent to the shore at
185 0.5-2 m depths. To sample brook stickleback from the limnetic zone, traps were set at least 50 m
186 from shore and suspended from floats at a depth of 1-2m. Traps were retrieved one to twelve
187 hours after being set. All brook stickleback samples were anesthetized and euthanized in an
188 overdose mixture of lake water and eugenol. In 2019, the posterior portion (posterior to the
189 pelvic girdle) of each individual from Muir Lake and Shunda Lake was frozen on dry ice then
190 preserved at -20°C for stable isotope analysis. All other tissues were preserved in 70% EtOH for
191 morphometric analysis. Samples were collected under a fisheries research license issued by the
192 Government of Alberta. Collection methods and the use of animals in research was approved by

193 the Animal Care Committee at Mount Royal University (Animal Care Protocol ID 101029 and
194 101795). Spined and unspined individuals were initially identified at the site of capture based on
195 close visual inspection and prodding with fine-tipped tweezers. Sex was also assigned at the site
196 of capture by examining gonads and by noting the presence of nuptial colouration. In 2019,
197 benthic invertebrates were collected from the littoral zones of Muir Lake and Shunda Lake by
198 rinsing and sorting through mud samples, and plankton was collected from the pelagic zone of
199 these two lakes using a Wisconsin Plankton Sampler. Benthic invertebrates were identified to
200 family or species (if possible) immediately after capture. Plankton and benthic invertebrate
201 samples were frozen on dry ice immediately, then preserved at -20°C for stable isotope analysis.

202

203 *Geometric Morphometrics*

204 The brook stickleback specimens were bleached and dyed using alizarin red following the
205 protocol from Xie *et al.* (2019). After bleaching and staining, we captured ventral and left-lateral
206 photographs of each fish against a 1x1cm grid using a Canon EOS Rebel T6i mounted above
207 each specimen at a height of 15 cm with two SV SlimPanel LED high-intensity illuminators. We
208 verified the pelvic phenotype for each individual by examining the ventral photograph, and,
209 following Kepakker *et al.* (2013), each individual was classified as having a “normal pelvis” with
210 a complete pelvic girdle and both spines (hereafter “spined”), a “lost pelvis” with complete
211 absence of pelvic girdle and spines (hereafter “absent”), or a “vestigial pelvis” wherein one or
212 more spines or pelvic girdle elements (the ascending branch, anterior process, or posterior
213 process) is missing. The population from Muir Lake contains very few vestigial pelvic
214 phenotypes (2 to 3% of the population), and, to allow comparisons across populations, we
215 combined the absent and vestigial pelvic phenotypes into a single category that we called

216 “reduced”. Analyses involving the three pelvic phenotypes (i.e., spined, vestigial, and absent)
217 were not possible in all instances (i.e. due to the lack of vestigial individuals in Muir Lake
218 samples), and, when they were possible, did not yield substantial differences in statistical results
219 or overall conclusions relative to the analyses with two categories (i.e. spined and reduced)
220 presented below. Specimens that were severely bent or distorted, and those whose pectoral fins
221 obscured their operculum, were excluded from the analysis. Landmarks placed on the left-lateral
222 photos were used to build a Tps file using tpsUtil version 1.79 (Rohlf, 2019).

223 To quantify two-dimensional body shape, we digitized 27 anatomical landmarks on each
224 of the left-lateral photographs using tpsDIG2w32, version 2.31 (Rohlf, 2018). The landmark
225 selection was based on previously established landmarks (Krisjansson 2005, Taugbol 2014). All
226 landmarks were visible from the lateral side of the fish (Figure 1). The number of dorsal spines
227 varies among individuals and populations from four to six (Nelson 1969). For this reason, instead
228 of recording the location of each dorsal spine, the location of the first and last dorsal spine were
229 used as landmarks. In case the dorsal spine landmarks are, in fact, not homologous, we also
230 analyzed whole-body shape data without the posterior dorsal spine landmark and without any
231 dorsal spine landmarks. Regardless, for samples collected in 2019, it was only possible to place
232 landmarks on the head (see Figure 1) because the posterior portion of each fish (i.e. posterior to
233 the pelvic girdle) was destroyed for use in the stable isotope analysis. Hence, whole-body
234 morphometric analysis used only samples collected in 2017, whereas analysis of head
235 morphology used samples from 2017 and 2019.

236

237 *Stable Isotope Analysis*

238 Stickleback, zooplankton, and benthic macroinvertebrate tissues collected in 2019 were
239 thawed then rinsed with distilled H₂O to clean off any lake-debris or mud. We then dried each
240 sample at 65°C for 48 hrs in an incubator oven, ground it into a powder in liquid nitrogen using a
241 mortar and pestle, then packed the ground tissue into 4 x 6 mm tin capsules for isotope analysis.
242 Stable isotope analysis was performed on the packed capsules using a Carbon and Nitrogen Ratio
243 Mass Spectrophotometer at the University of Calgary Geosciences Isotope Analysis Laboratory.
244 Stable isotope ratios are expressed as a delta notation (δ) which is defined as the parts per
245 million (‰) difference from a universal standard (Zanden *et al.* 1999). The standard material for
246 $\delta^{13}\text{C}$ is Pee Dee belemnite (PDB) limestone, and for $\delta^{15}\text{N}$ it is atmospheric nitrogen (both ‰
247 values arbitrarily set at 0 ‰; Zanden *et al.* 1999, Ben-David and Flaherty 2012). To assess
248 accuracy and repeatability of our isotope ratio measurements, we recorded triplicate or duplicate
249 isotope measurements for 22.5% of samples (including all plankton samples).

250

251 *Statistical Analysis*

252 To evaluate the hypothesis that pelvic phenotype is associated with other morphological
253 changes, we assessed shape variation between pelvic phenotypes using the GEOMORPH
254 package in R (Adams and Collyer 2020). We first performed a Generalized Procrustes analysis
255 to estimate a scaling factor that compensated for the natural variation in fish size and applied this
256 to all samples. This assured that all landmarks from the samples were placed on comparable
257 locations and avoided wide dispersion of landmark coordinates. We used the `procD.lm` function
258 to conduct Procrustes ANOVA (with type III sums of squares to allow for unbalanced data – see
259 Table 1), which uses a linear model to evaluate the morphological variation attributable to the
260 following factors: lake, sex, year, pelvic phenotype, and body size. We expected that the effect of

261 pelvic phenotype on morphology might be different between lakes or between sexes, and that the
262 effect may be influenced by allometry (Aguirre et al. 2008, Reimchen et al. 2016). So, we
263 included two-way interactions in our linear model. The probability of the observed effect for
264 each factor was evaluated by comparison to a null distribution generated by 10,000 resampling
265 permutations. We used a reverse stepwise approach for model selection, starting with the full
266 model (main effects and two-way interactions) and removing any body size (i.e. allometric)
267 interaction terms that were not significant.

268 We performed a visual evaluation of variation in isotopic signatures among fish,
269 plankton, and benthic macroinvertebrate samples using a bi-plot of $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ signatures.
270 To evaluate the hypothesis that the spined and unspined brook stickleback pelvic phenotypes use
271 different habitats and forage on different food sources, we tested the association of sex, lake,
272 pelvic phenotype, and fish size with $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ signatures using generalized linear models
273 with gaussian error distributions. We expected that the association between pelvic phenotype and
274 isotope signatures might be different between lakes or between sexes, and that the association
275 may be influenced by allometry (Reimchen et al. 2016). So, as with the analysis of
276 morphological variation, we included two-way interactions in our model. We set contrasts
277 among factors using the `contr.sum` function, and we used the CAR package (Fox and Weisberg
278 2019) to generate an ANOVA table (with type III sums of squares to allow for unbalanced data –
279 see Table1) to evaluate the variation in $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ signatures attributable to each factor and
280 their interactions terms. We used a reverse stepwise approach for model selection, starting with
281 the full model (main effects and two-way interactions) and removing any body size (i.e.
282 allometric) interaction terms that were not significant. In addition, to avoid over-fitting our
283 model (with a sample size of only 105 brook stickleback $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ isotope signatures), we

284 further reduced the number of interaction terms in our analyses of stable isotope variation by
285 removing any two-factor interactions that were not significant.

286

287 **Results**

288 A complete summary of all samples analyzed in this study, categorized by year, lake,
289 habitat, sex, and pelvic phenotype, is presented in Table 1. We caught no fish in the limnetic
290 zone in Shunda Lake. The limnetic-caught fish from Muir Lake were significantly smaller than
291 littoral-caught fish (two-sided $t = -2.618$, $df = 53.437$, $p = 0.01148$), and we only caught three
292 female fish in the Muir Lake limnetic zone – none of which were spined. We excluded limnetic-
293 caught fish from subsequent statistical analyses.

294

295 *Morphology*

296 We organized our analyses of morphological variation based on which landmarks were
297 available among samples from different years: one analysis involving sixteen head-only
298 landmarks that were present in all samples (from 2017 and 2019: number of observations = 295),
299 and one analysis involving all twenty-seven landmarks, some of which (i.e. posterior to the
300 pelvic girdle) were not present in 2019 samples because of destructive sampling for stable
301 isotope analysis (number of observations = 179). For simplicity, Figure 2 shows only the whole-
302 body morphological associations. Brook stickleback head-body morphology was significantly
303 associated with fish size and differed significantly between sexes (2017 whole-body: Table 2;
304 2017-2019 heads only: Table 3). In addition, head morphology varied significantly among years
305 (Table 3). Females had more elongated abdominal regions, narrower bodies, and shorter heads,
306 which is a pattern observed in threespine stickleback as well (Aguirre et al. 2008). The

307 association between pelvic phenotype and head-body morphology was different in each lake. In
308 Muir Lake, pelvic reduction was most-noticeably associated with a deeper body, larger head, and
309 anterior-shifted pectoral fins, whereas in Shunda Lake, pelvic reduction was associated with a
310 compression of the ventral portion of the head (including a less up-curved mouth), posterior-
311 shifted pectoral fins, and a longer, narrower body (Figure 2). These patterns are consistent with
312 our hypothesis that brook stickleback morphology is affected by pelvic phenotype, and these
313 results are consistent with previous observations that morphological divergence is non-parallel
314 (Kaeuffer et al. 2012).

315

316 *Stable isotopes*

317 In both lakes, brook stickleback had higher $\delta^{15}\text{N}$ isotopic signatures than benthic
318 invertebrates (Figure 3) suggesting that, as expected, fish occupy a higher trophic niche relative
319 to the macroinvertebrates. In Shunda Lake, the plankton had the lowest $\delta^{15}\text{N}$ signature (Figure
320 3) suggesting the expected hierarchy in trophic position, with fish at the top, plankton at the
321 bottom, and macroinvertebrates in the middle. In Muir Lake, the $\delta^{15}\text{N}$ isotopic signature was not
322 substantially lower than the average for macroinvertebrates, but the $\delta^{13}\text{C}$ signature for plankton
323 in Muir Lake was substantially higher than any of the other samples. The higher-than-expected
324 $\delta^{13}\text{C}$ signature and higher-than-expected $\delta^{15}\text{N}$ signature for plankton in Muir Lake is an
325 unexpected result, and further investigation is needed for a satisfactory explanation.

326 Stickleback $\delta^{13}\text{C}$ isotope signatures were significantly associated with fish size (Table
327 4), but none of the interactions terms with fish size were significant. This suggests that the effect
328 of size on $\delta^{13}\text{C}$ signature is the same regardless of pelvic phenotype, sex, or lake. The
329 association between pelvic phenotype and $\delta^{13}\text{C}$ signature was different in each lake, and Muir

330 Lake had a much higher $\delta^{13}\text{C}$ signature than Shunda Lake (Table 4, Figure 3). In Muir Lake,
331 pelvic reduction was associated with a higher $\delta^{13}\text{C}$ signature, whereas in Shunda Lake, pelvic
332 reduction was associated with a lower $\delta^{13}\text{C}$ signature (Figure 4). In both lakes, however, the
333 shift in brook stickleback $\delta^{13}\text{C}$ signature associated with pelvic reduction was towards the
334 planktonic $\delta^{13}\text{C}$ signature (Figure 3). In Muir Lake, limnetic-caught fish tended to have a higher
335 $\delta^{13}\text{C}$ signature than littoral-caught fish (Figure 3). These patterns are consistent with our
336 hypothesis that brook stickleback with different pelvic phenotypes forage in different habitats,
337 but these patterns are inconsistent with our prediction that, as in threespine stickleback, pelvic
338 reduction would be associated with littoral or benthic habitats. In fact, contrary to our prediction,
339 these results suggest that pelvic reduction is associated with limnetic (or planktonic) feeding in
340 both lakes.

341 The effect of size on stickleback $\delta^{15}\text{N}$ isotopic signature was dependent on lake (Table
342 5). In Muir Lake, larger fish had a lower $\delta^{15}\text{N}$ isotopic signature, whereas larger fish had a
343 higher $\delta^{15}\text{N}$ isotopic signature in Shunda Lake (Figure 5). The effect of pelvic phenotype on
344 stickleback $\delta^{15}\text{N}$ isotopic signature was dependent on sex. In males, pelvic reduction is
345 associated with a higher $\delta^{15}\text{N}$ signature, whereas pelvic reduction is associated with a lower
346 $\delta^{15}\text{N}$ signature in females (Figure 5). These patterns are consistent with our hypothesis that
347 brook stickleback pelvic morphs forage in different habitats.

348

349 **Discussion**

350 The hypothesis that pelvic spine polymorphism in brook stickleback is associated with
351 divergence in habitat use was supported by our results. Based on carbon isotope signatures,
352 brook stickleback with pelvic reduction (i.e. with either no pelvic structure or a vestigial pelvic

353 structure) likely feed on planktonic as opposed to benthic macroinvertebrate food sources and,
354 therefore, likely use more limnetic as opposed to littoral or benthic habitat. This result is,
355 however, contrary to our prediction based on the well-established association between pelvic
356 reduction and benthic habitats in threespine stickleback (Reimchen 1980, McPhail 1992). There
357 were also significant changes in head and body morphology associated with pelvic reduction,
358 but, as expected (Kaeuffer et al. 2012), the specific nature of this morphological difference was
359 lake dependent, and the magnitude of the morphological difference between pelvic phenotypes
360 was small. In fact, the proportion of the variation in morphology, $\delta^{13}\text{C}$ isotope signature, or
361 $\delta^{15}\text{N}$ signature attributable to pelvic spine variation was small relative to individual-level
362 variation and between-lake variation (Figures 3 and 4, Tables 2, 3, 4, and 5).

363 There is ample evidence supporting the role of stickleback pelvic spines in predator
364 interactions. It is likely that balancing selection driven by multiple and varied predators is
365 involved in maintaining the spine polymorphism, and that predation drives associations between
366 spine morphology and diet, habitat, and body morphology. But, we do not know the mechanism
367 via which predation and other ecological factors cause the unexpected association between pelvic
368 reduction and a more planktonic diet (as opposed to the predicted association between pelvic
369 reduction and a more benthic diet). It is possible that, in these systems, stocked trout (or other
370 gape-limited predatory fish) forage in littoral zones, thereby selecting for spined stickleback in
371 these habitats. Stocked rainbow trout have been observed ambushing minnows from under
372 floating docks in the littoral zones of other nearby lakes (J. Mee personal observation).

373 We found that, in unspined fish only, males had significantly higher $\delta^{15}\text{N}$ signatures
374 than females. This suggests that unspined females occupy a lower trophic position than unspined
375 males, which may indicate that unspined females have a more planktivorous diet, whereas

376 unspined males have a more benthic diet. This is consistent with patterns observed in male
377 threespine stickleback, which are more benthic-associated than female threespine stickleback as
378 a consequence of their breeding behaviour (Spoljaric and Reimchen 2008, Aguirre and Akinpelu
379 2010, McGee and Wainwright 2013). Male threespine stickleback migrate from the limnetic to
380 the littoral habitats to build nests, spawn, and defend their eggs. Brook stickleback also build
381 nests and defend their eggs, but, whereas threespine stickleback build their nests on the substrate,
382 brook stickleback males build nests on vertical rocky surfaces or vegetation (Reisman and Cade
383 1967, McLennan 1995). Hence, it is not clear that the breeding behaviour of brook stickleback
384 males should cause them to be more benthic-associated than females. Also, if the difference in
385 $\delta^{15}\text{N}$ signature between unspined males and females is due to male breeding behaviour, it is not
386 clear why spined males do not have a higher $\delta^{15}\text{N}$ signature than spined females unless the
387 breeding behaviour of spined and unspined males is different. It seems likely that differences
388 between brook stickleback and threespine stickleback life history and reproductive behaviour
389 obscure any simple comparison between the species regarding the interactions among diet,
390 habitat, and morphology.

391 Morphological variation among threespine stickleback populations is associated with
392 habitat specialization (Reimchen et al. 1985, Webster 2011). Even subtle changes in body
393 morphology can be associated with fitness parameters like foraging patterns, body condition, and
394 growth rate (Webster et al. 2011). Different selection pressures in different habitats may favour
395 different morphological traits (Reimchen et al. 1985). If traits that are well-adapted for a certain
396 environment are ill-suited for another, there may be fitness trade-offs among habitats (Webster et
397 al. 2011). The morphological variation among pelvic phenotypes may be due to differing water
398 depth, water chemistry (e.g. pH), predation risk, and parasite prevalence in different habitats

399 (Reimchen et al. 1985, Webster 2011). There is, therefore, good reason to assume that
400 differences in habitat use and diet between spined and unspined brook stickleback cause the
401 observed morphological differences between morphs in the present study. However, it is not
402 clear if the morphological differences between spined and unspined brook stickleback constitute
403 a plastic response or a heritable response to divergent habitat use. Additionally, if the
404 morphological divergence is heritable, we do not know whether it has functional significance
405 driven by different selective regimes in different habitats, or if it results from pleiotropic effects
406 of the genes underlying the pelvic divergence.

407 There was an obvious difference in $\delta^{13}\text{C}$ signature between the population in Muir Lake
408 and the population in Shunda Lake. This could be the result of different diet preferences between
409 populations (Jardine *et al.* 2003, Eloranta et al. 2010), or it may reflect chemical differences in
410 the environment (e.g. weather, sediment, or human impact, pH). Muir Lake is in Alberta's
411 Central Parkland natural region, which is a prairie landscape, and its riparian area is dominated
412 by marsh vegetation such as cattails (*Typha* spp.) and sedges (*Carex* spp.), and likely has
413 relatively high pH (AWA 2020). Shunda Lake is in the Upper Foothills natural region, which is
414 mountainous, and its riparian area is characterized by conifer stands and understory shrubs, and
415 likely has relatively low pH (AWA 2020, Ross and Kyba 2015). This difference between lakes
416 may be a reason for the difference in isotopic signatures among the plankton and invertebrate
417 samples in these two lakes and may represent an ecological basis for the differences in
418 morphology and isotopic signature between these two brook stickleback populations.

419 The isotopic signatures of brook stickleback, benthic invertebrates, and zooplankton
420 relative to one another (Figure 3) were generally consistent with the linear ascending relationship
421 of $\delta^{15}\text{N}$ in food-web and trophic position studies of freshwater fish and their food sources (Post

422 2002). The zooplankton signature from Muir Lake did not, however, fit neatly within this
423 paradigm. We lack any evidence to support an explanation for this unexpected isotopic signature
424 in Muir Lake, but we can offer some conjecture. It is possible that Muir lake, at the time of
425 sampling, was dominated by a single species or relatively few species of zooplankton (n.b. we
426 did not identify zooplankton to species in our study). The expectation that zooplankton should
427 have a lower $\delta^{15}\text{N}$ and a higher $\delta^{13}\text{C}$ isotopic signature than their consumers (Post 2002)
428 assumes an average isotopic signature among a community of zooplankton. If our sample
429 constituted only a few species (or even a single species), it may have been a species with a
430 particularly ^{15}N -enriched and ^{13}C -enriched isotopic signature. Further sampling and analysis of
431 isotopic signatures in this lake would be required to explain this result.

432 The observation of significant differences in morphology and stable isotope signature
433 between lakes suggests another future avenue of inquiry. If differences between lakes are
434 heritable, there may be implications related to parallelism in the evolution of pelvic spine
435 reduction. It is unknown whether the genetic basis of pelvic spine polymorphism is the same in
436 all brook stickleback populations. If differing body morphologies between lakes are associated
437 with different genetic bases for pelvic reduction in different lakes (e.g. via pleiotropy), there may
438 be less phenotypic and genetic parallelism in this system than previously assumed. Investigations
439 of the heritability and genetic bases of polymorphism in these lakes are ongoing.

440

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456

457 **Shared Data**

458 All data and analysis scripts have been uploaded to the Dryad Data repository:

459 <https://datadryad.org/stash/share/Vho8zW5Aqhj5yACQtLSKf0Dm9BWcpGXIHlIVJshTeQ0>

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665

666 **Table 1.** Summary of samples analyzed in this study. Morphological analyses focused on
 667 samples from 2017, whereas analysis of stable isotopes only included samples from 2019. For all
 668 analyses described in the text, the vestigial and absent pelvic phenotypes were combined into a
 669 single “reduced” phenotypic category. All statistical analyses used type III sums of squares to
 670 account to unbalanced sampling.

	Spined	Pelvic Phenotype	
		Vestigial	Absent
2017			
Muir Lake			
Littoral			
female	20	1	24
male	24	1	18
Shunda Lake			
Littoral			
female	20	24	9
male	29	6	3
2019			
Muir Lake			
Littoral			
female	3	0	4
male	16	0	15
Limnetic			
female	0	0	3
male	6	1	10
Shunda Lake			
Littoral			
female	14	11	4
male	18	9	2

671

672 **Table 2.** ANOVA table for the linear model fitted to whole-body 2D morphology for brook
 673 stickleback collected in 2017.

	df	Type III SS	F	p
Log(length)	1	0.004158	4.4343	0.0008
Lake	1	0.008727	9.3065	< 0.0001

Sex	1	0.014800	15.7835	< 0.0001
Pelvic phenotype	1	0.002982	3.1796	0.0086
Lake * Sex	1	0.001971	2.1021	0.0532
Lake * Pelvic phenotype	1	0.003046	3.2482	0.0075
Sex * Pelvic phenotype	1	0.000754	0.8043	0.5432
Residuals	171	0.160345		

674

675 **Table 3.** ANOVA table for the linear model fitted to head-only 2D morphology for brook

676 stickleback collected in 2017 and 2019.

	df	Type III SS	F	p
Log(length)	1	0.02024	6.5175	< 0.0001
Year	1	0.01697	5.4632	0.0004
Sex	1	0.01019	3.2824	0.0055
Lake	1	0.02118	6.8201	< 0.0001
Pelvic phenotype	1	0.00174	0.5593	0.8024
Year * Sex	1	0.00179	0.5761	0.7905
Year * Lake	1	0.02082	6.7046	< 0.0001
Year * Pelvic phenotype	1	0.01321	4.2533	0.0013
Sex * Lake	1	0.00565	1.8186	0.0819

Sex * Pelvic phenotype	1	0.00239	0.7698	0.5975
Lake * Pelvic phenotype	1	0.00721	2.3214	0.0322
Residuals	283	0.87888		

677

678 **Table 4.** ANOVA table for the generalized linear model fitted to $\delta^{13}\text{C}$ isotopic signature for
679 brook stickleback collected in 2019.

	df	Type III SS	F	p
Lake	1	527.5	848.4586	< 0.0001
Pelvic phenotype	1	0.29	0.4613	0.49873
Sex	1	0.64	1.0329	0.31214
Length	1	4.69	7.5448	0.00724
Lake * pelvic phenotype	1	2.97	4.7848	0.03125
Residuals	92	57.2		

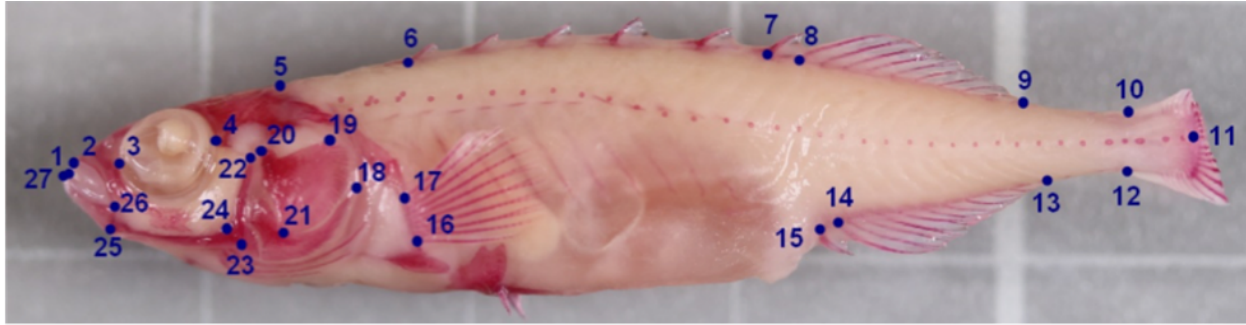
680

681 **Table 5.** ANOVA table for the generalized linear model fitted to $\delta^{15}\text{N}$ isotopic signature for
682 brook stickleback collected in 2019.

	df	Type III SS	F	p
Lake	1	3.009	5.4621	0.021628
Pelvic phenotype	1	0.299	0.542	0.463479
Sex	1	1.965	3.5674	0.06211
Length	1	0.944	1.7137	0.193798

Lake * length	1	2.872	5.2133	0.024742
Pelvic phenotype * Sex	1	4.477	8.1275	0.005394
Residuals	91	50.13		

683

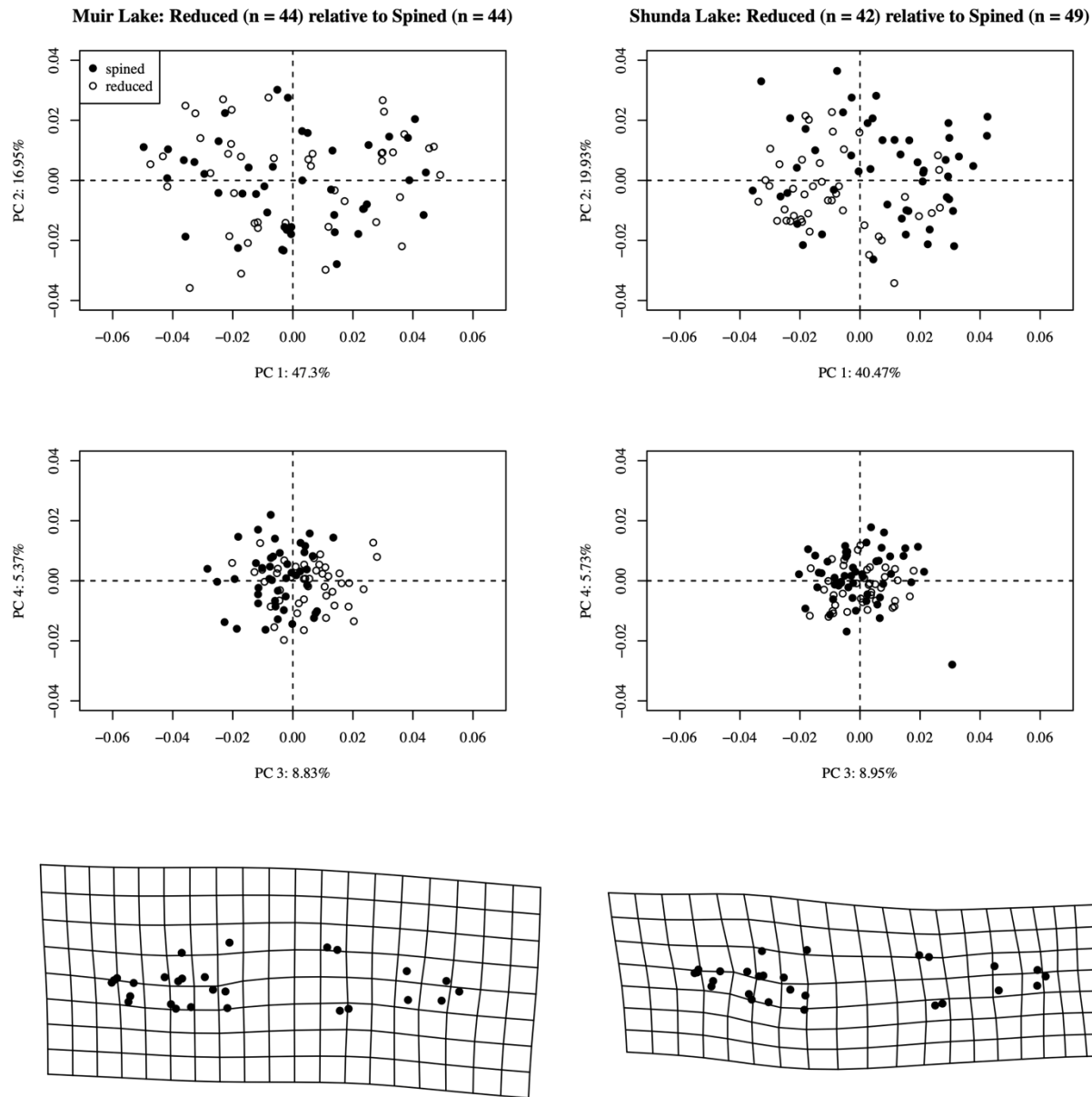


685

686 **Figure 1.** Position of 27 morphological landmark locations on a brook stickleback specimen (n =
687 105). 1) Anterior ventral tip of upper lip, 2) Anterior dorsal tip of upper lip, 3) Anterior border or
688 the eye, 3) Posterior border of the eye, 4) Posterior dorsal tip of skull, 5) Base of first dorsal
689 spine, 6) Base of last dorsal spine, 7) Anterior junction of the dorsal fin, 9) Posterior junction of
690 the dorsal fin, 10) Dorsal insertion of the caudal fin, 11) Posterior end of the hypural plate at the
691 midline, 12) Anterior insertion of the caudal fin, 13) Posterior junction of the anal fin, 14)
692 Anterior junction of the anal fin, 15) Base of anal spine, 16) Anterior tip of the pectoral fin, 17)
693 Dorsal tip of the pectoral fin, 18) Dorsal tip of the gill cover, 19) Posterior dorsal tip of
694 operculum, 20) Anterior dorsal tip of operculum, 21) Ventral tip of operculum, 22) Dorsal tip of
695 preoperculum, 23) Posterior angular tip of preoperculum, 24) Posterior jaw, 25) Anterior tip
696 ventral jaw, 26) Posterior tip of lips, 27) Anterior tip of lower lip. Landmarks 7 through 15 were
697 unavailable for samples collected in 2019.

698

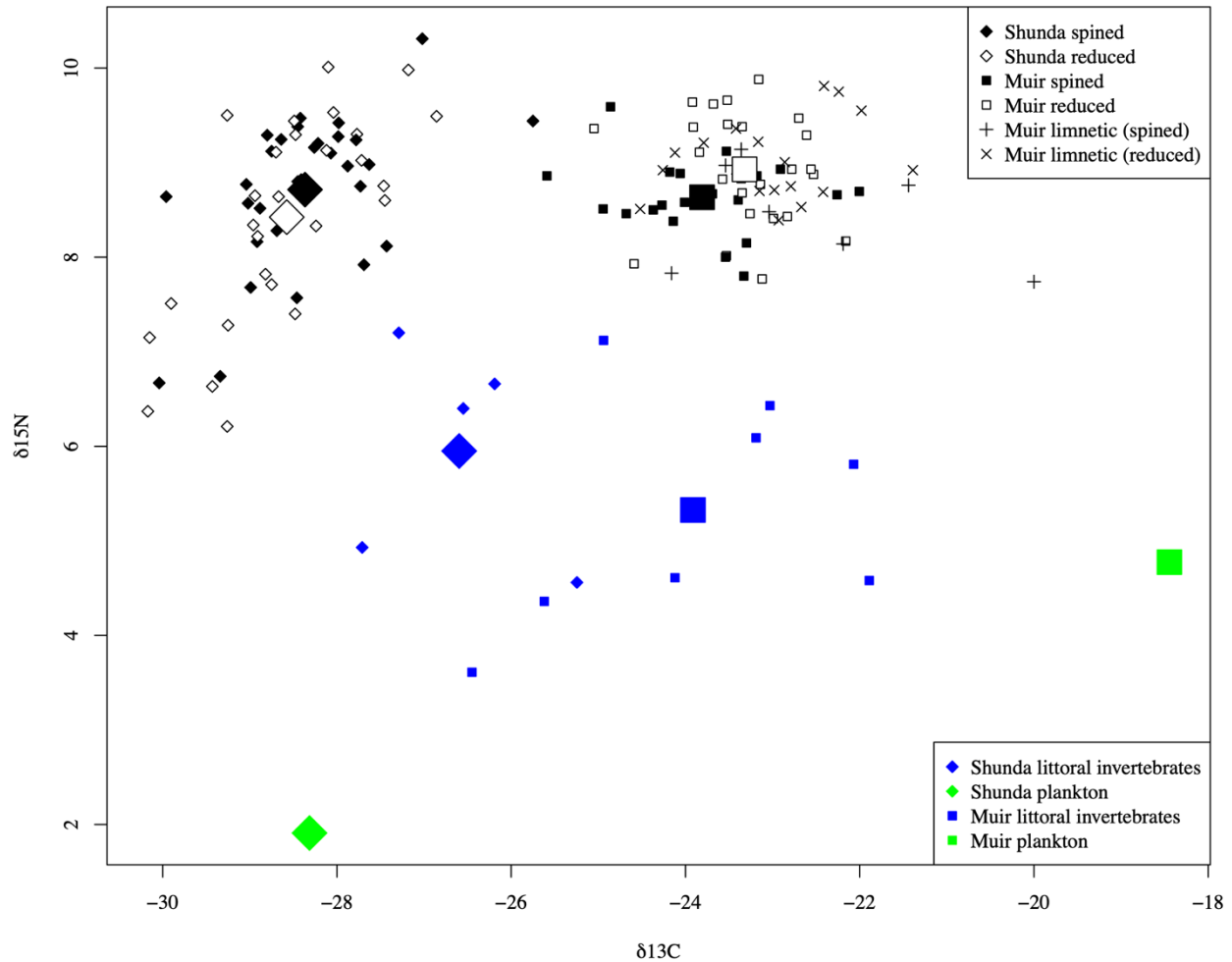
699



700

701 **Figure 2.** The effect of pelvic phenotype on body morphology in brook stickleback from Muir
702 Lake (left column) and Shunda Lake (right column). The top two plots in each column show the
703 results of principle components analyses for each lake. Values for each individual fish for the
704 first four principle components (PCs) are shown with the proportion variance in body
705 morphology explained by each PC shown on the axis labels. The bottom plot in each column
706 shows the distortion of a symmetrical (square) grid when the average shape of a brook
707 stickleback with a reduced pelvic phenotype is superimposed on the average shape of a brook
708 stickleback with a complete pelvic structure.

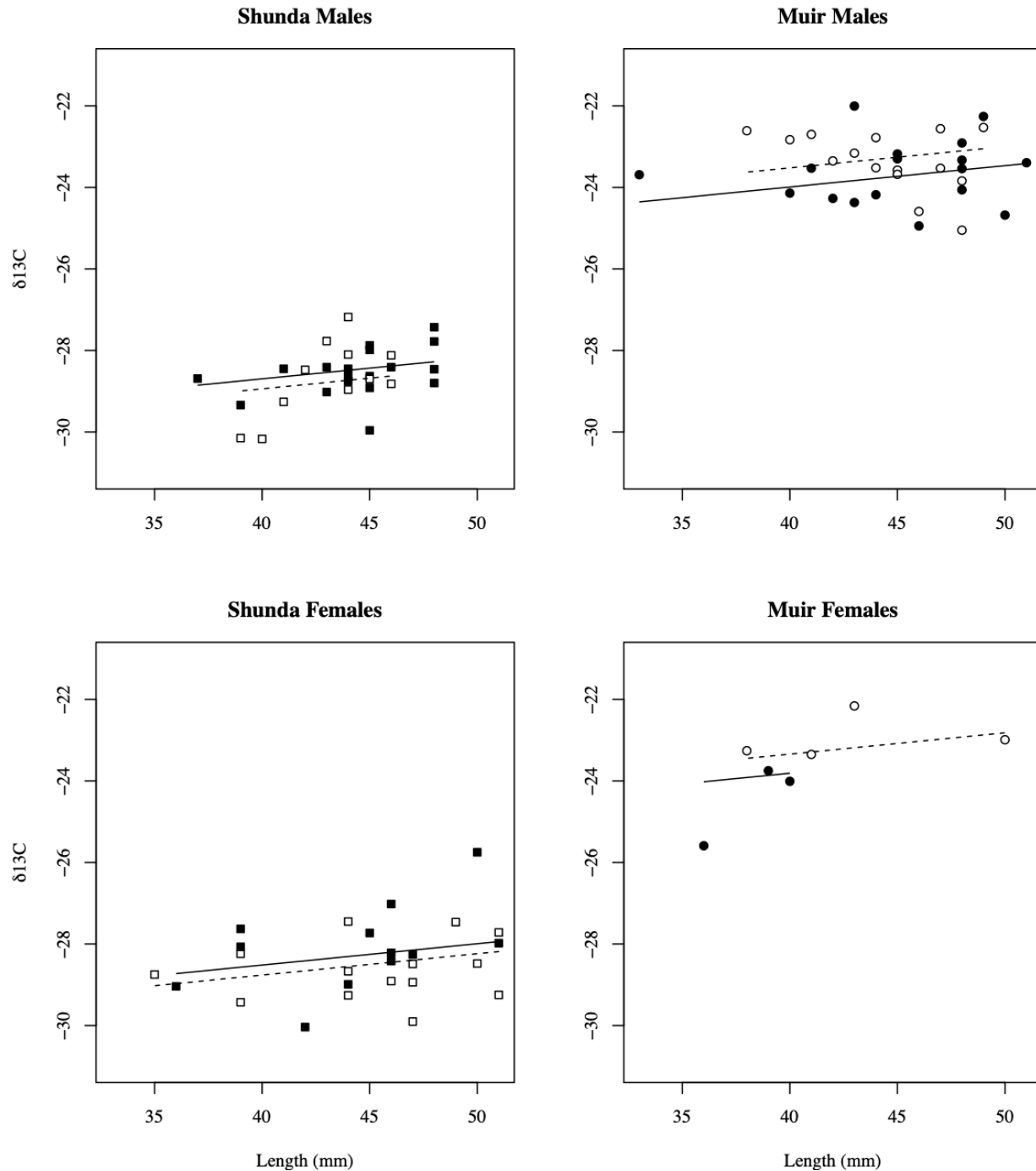
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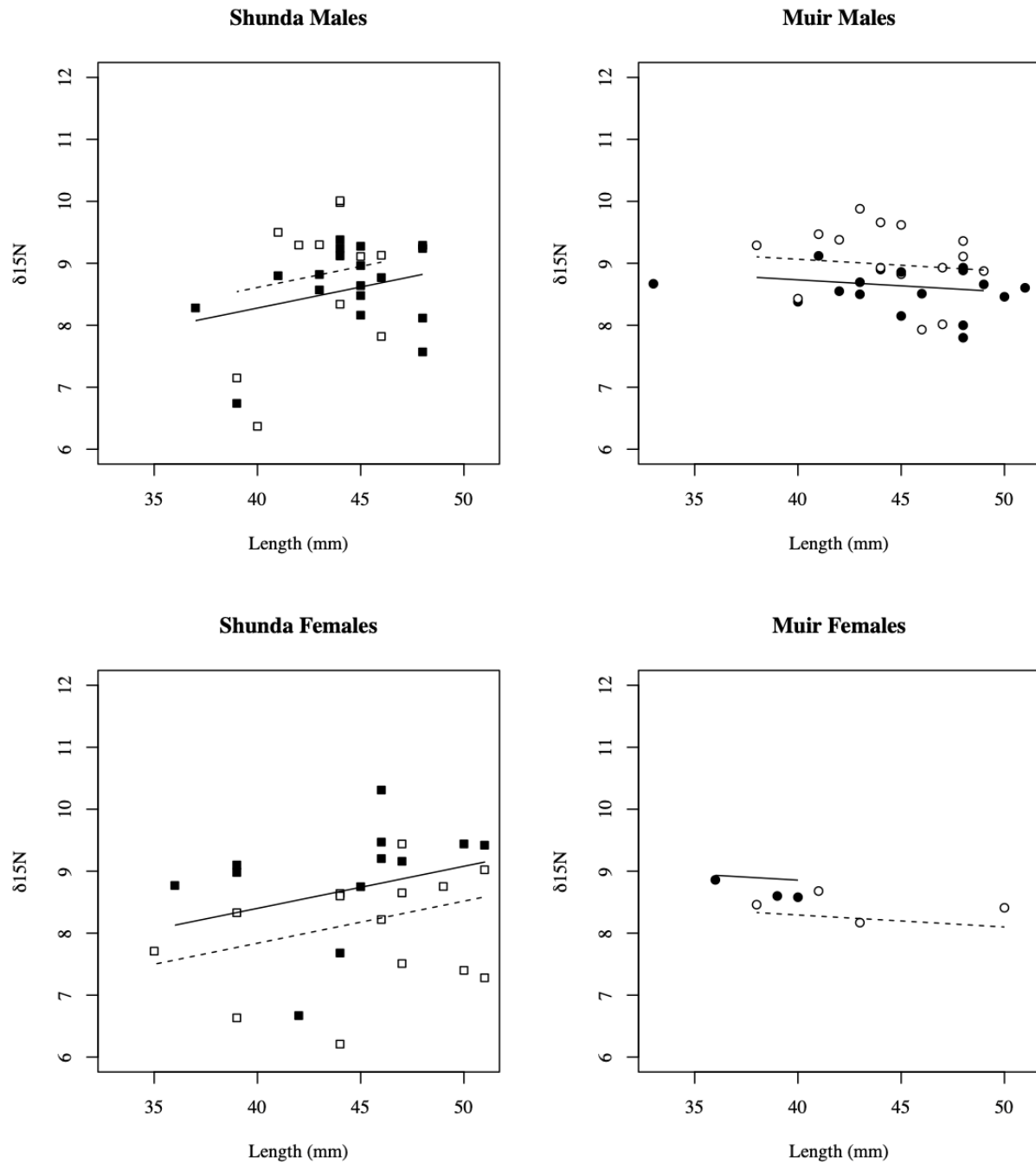
711 **Figure 3.** Carbon and nitrogen stable isotope signatures for spined and unspined brook
712 stickleback, zooplankton, and benthic littoral macroinvertebrates from Muir Lake and Shunda
713 lake. Macroinvertebrate taxa represented in the plot include caddisfly larvae in the family
714 Limnephilidae, unidentified caddisfly larvae, amphipods, leaches, dragonfly larvae in the family
715 Gomphidae, unidentified dragonfly larvae, unidentified snails, and unidentified damselfly larvae.
716 Also shown (with + and x symbols) are limnetic-caught brook stickleback from Muir Lake. All
717 fish were caught in the littoral zone unless otherwise indicated. The larger symbols represent
718 mean values for each subgroup (excluding limnetic-caught), whereas the smaller symbols
719 represent values for individuals.

720



721

722 **Figure 4.** Summary of the effects of sex, lake, pelvic phenotype, and fish size on $\delta^{13}C$ isotope signature in brook stickleback.
723 As in Figure 3, closed symbols represent spined fish and open
724 symbols represent fish with reduced pelvic phenotypes. Lines show the linear relationships
725 inferred from the model described in the text and in Table 4 (solid = spined, dashed = reduced).
726



727

728 **Figure 5.** Summary of effects of sex, lake, pelvic phenotype, and fish size on $\delta^{15}\text{N}$ isotope
729 signature in brook stickleback. As in Figure 3, closed symbols represent spined fish and open
730 symbols represent fish with reduced pelvic phenotypes. Lines show the linear relationships
731 inferred from the model described in the text and in Table 5 (solid = spined, dashed = reduced).
732