# Population structure and environmental niches of *Rimicaris* shrimps from the Mid Atlantic Ridge

## Authors and Affiliations

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## Key words

Hydrothermal vents, life cycle, population structure, environmental niches, juvenile mortality, *Rimicaris* shrimps

## Abstract

Among the endemic and specialized fauna from hydrothermal vents, *Rimicaris* shrimp surely constitutes one of the most important and emblematic component of these ecosystems. In the Mid Atlantic Ridge, two species affiliated to this genus co-occur: *Rimicaris exoculata* and *Rimicaris chacei* that differ by their morphology, their trophic regime as well as by their abundance. The first forms large and dense aggregations on active vent chimney walls in close proximity to vent fluid emissions, whereas the second is recognized as much less conspicuous, living mostly in scattered groups or solitary further away from the fluids. However, the recent revision of *Rimicaris* juvenile stages from the Mid Atlantic Ridge show that *R. chacei* abundance would be higher than expected at these early life stages. Here, we describe and compare the population structures of *R. exoculata* and *R. chacei* at the Snake Pit and TAG vent fields. We show widely distinct population demographics between the two co-occurring shrimps with a large post settlement collapse in *R. chacei* population suggesting a large juvenile mortality for this species. We also observe important spatial segregation patterns between the two species and their different life stages. Additionally, our results highlight distinct niches for the earliest juvenile stages of both *R. exoculata* and *R. chacei*, compared to all the other life stages. Finally, we discuss the potential factors - predation and competitive interactions among others - that could explain the differences we observe in the population structure of these two species.

## **Introduction**

Fueled by the chemosynthetic energy arising from the mixing between vent fluids and surrounding seawater, hydrothermal vent ecosystems host dense and lush communities of endemic fauna which are organized in distinct species assemblages along steep thermal and chemical gradients. Distributed patchily along mid oceanic ridges, these environments constitute isolated but also ephemeral oases of life within the deep sea, either because of catastrophic events leading to the collapse of entire populations (Fornari et al., 2012) or simply because of the general lifespan of a vent field, as fluid activity is doomed to ultimately wane and cease over geological timescales (Van Dover, 2019). Therefore, vent species survival over successive generations is greatly interlinked with their ability to successfully settle in and colonize new vent fields, implying adaptations at each stage of their lifecycle.

As in other marine benthic ecosystems (Menge, 1991), population demographics of species from hydrothermal vents depend on the number of competent larvae reaching the vent site, the probability of success of their settlement and any mortality events occurring between their successive life stages (Kelly and Metaxas, 2007). Therefore, any biotic or abiotic factor affecting one of these three steps will likely impact the capacity of individuals to reach sexual maturity and the size of the adult populations. These environmental factors are acting differentially depending on the life stages considered as each of them can occupy its own ecological niche and differ by its vulnerability to the different mortality factors (Werner and Gilliam, 1984).

Whereas a high juvenile mortality, mostly related to size selective predation, is often characteristic of the post settlement phase in shallow water species (Gosselin and Qian, 1997), a similar generalization has not yet been established in hydrothermal vent ecosystems. This can be attributed to the difficulty of conducting long term population survey and temporal observations along appropriate timescales in such remote places. Settlement and recruitment patterns of these species have then mostly been inferred from cohort monitoring. According to these studies, populations of some vent species such as *Bathymodiolus azoricus* mussels or some gastropods are largely dominated by small size individuals (Comtet and Desbruyères, 1998; Kelly and Metaxas, 2007) implying either an important juvenile mortality or large variations between the different recruitment events. Conversely, a limited number of juveniles has been reported for several others including alvinellid polychaetes and many peltospirid or lepetodrilid gastropods (Zal et al., 1995; Faure et al., 2007; Matabos and Thiebaut, 2010; Marticorena et al., 2020) suggesting a frequent arrival of larvae at the vent field followed by a fast growth rate and an accumulation within the adult populations. Still, spatial or temporal limitation in sampling could also bias our understanding of post settlement processes in these populations, as different life stages sometimes inhabit different parts of the field. Important spatial zonation between juveniles and adults has been reported for instance in *Kiwa tyleri* crabs in which adult aggregates locate near the fluid exits while juveniles remain restricted at the periphery of the field (Marsh et al., 2015). Similarly, a segregation according to the organism size was observed along the thermal gradient in *Bathymodiolus azoricus* mussels with juveniles and small mussels preferentially found in areas with lower temperatures (Husson et al., 2017).

*Rimicaris* shrimps are important components of hydrothermal vent ecosystems as they constitute some of the most visually dominant megafauna in several regions such as in the Central Indian Ridge (Watanabe and Beedessee, 2015), the Mid Cayman Rise (Plouviez et al., 2015) or the Mid Atlantic Ridge (Zbinden and Cambon-Bonavita, 2020). In the northern part of the Mid Atlantic Ridge,

two species of this genus co-occur: *Rimicaris exoculata*, forming large aggregations close to vent fluid emissions, with densities of several thousands of individuals per m$^2$, and *Rimicaris chacei* supposedly less abundant than its congener (several tens to hundreds of individuals per m$^2$) as they are less conspicuous within the vent fields (Segonzac et al., 1993).

It is assumed that population demography of congeneric species should be comparable especially when they evolve in similar habitat conditions (Münzbergová, 2013). Observations of contrasting population sizes and distribution patterns in many biologically similar congeners suggest, however, that distinct population dynamics can exist even in cases of syntopy when species inhabit jointly the same habitat at the same time (Münzbergová, 2013; Jones and Ricciardi, 2014; Bouchemousse et al., 2016). A stable co-occurrence at a small spatial scale of closely related species usually presupposes some differentiation in resource use as well as mechanisms that prevent interbreeding (Beermann and Franke, 2012). When congeners display similar biological traits and ecological niches, competitive processes are expected to be intensified (Franke et al., 2007; Rius et al., 2009), although they do not necessarily constitute a major driver of their population dynamics (Bouchemousse et al., 2016). The two *Rimicaris* species differ by their trophic regime, highly dependent on the dense and diversified chemosynthetic episymbiotic communities hosted in their enlarged cephalothorax *for R. exoculata* (Gebruk et al., 2000; Ponsard et al., 2013; Zbinden and Cambon-Bonavita, 2020) and relying on both symbiotrophy, bacterivory and scavenging for *R. chacei* (Gebruk et al., 2000; Apremont et al., 2018). This mixotrophic regime of *R. chacei* has been hypothesized to result from the competition for space with *R. exoculata* that would maintain *R. chacei* at a distance from vent fluids essential for fueling their symbionts (Apremont et al., 2018).

Zonation patterns between *R. exoculata* life stages have been reported by different studies (Segonzac et al., 1993; Shank et al., 1998; Gebruk et al., 2010) which highlighted the existence of red juvenile patches adjacent to the dense aggregations of adults in several MAR vent fields. More recently, spatial segregation between males, found at the periphery away from the fluid mixing zones, and females, mainly occurring within the dense aggregates on active chimney walls, have also been observed (Hernández-Ávila et al., 2021). In addition, dense aggregations of small juveniles were also collected close to peripheral and diffuse fluid emissions arising from cracks at the TAG vent field (Methou et al., 2020; Hernández-Ávila et al., 2021). Surprisingly, these individuals were revealed to be juvenile stages of *R. chacei* unveiling a misidentification in the previous species affiliation of these life stages(Methou et al., 2020; Hernández-Ávila et al., 2021). Such a large abundance of *R. chacei* early life stages was unexpected considering the recognized low abundance of the adults, and implied for *R. chacei* either a low post-settlement survival, or a larger adult population than previously stated, or both. Therefore, a full characterization of the population structures of *R. exoculata* and *R. chacei* is required to properly assess the demography of each species and whether the differences in the relative abundance of life stages could be related to their different trophic strategies or to other external factors such as predation, or physiological stress arising from their environment.

In this study, we describe and compare the population structures of *R. exoculata* and *R. chacei* from different types of shrimp assemblages at two mid-Atlantic vent fields, TAG and Snake Pit, using length-frequency analysis and the recently revised identification of *Rimicaris* shrimp life stages from these areas (Methou et al., 2020). Taking into account the significant spatial zonation of the different life stages of these two species, we also characterized their realized environmental niches and

112 discussed the role of potential biotic and abiotic factors affecting the demography of these shrimp
113 populations.

## **Materials & Methods**

### **Field Sampling**

*Rimicaris* shrimps were collected on board the R/V *Pourquoi Pas?* during the BICOSE 2 cruise
(http://dx.doi.org/10.17600/18000004) between January 26th and March 10th, 2018 at the TAG
(26°08.2'N, 44°49.5'W, 3620 m depth) and the Snake Pit (23°22.1'N 44°57.1'W, 3470 m depth)
hydrothermal vent fields using the suction sampler of the HOV *Nautile*. Samples were collected all over
the active venting areas at the TAG vent field and on the three active edifices of Snake Pit: the Beehive,
the Moose and the Nail (Figure S1). In total, 3919 *R. exoculata* shrimps (1533 from TAG and 2386 from
Snake Pit) and 1201 *R. chacei* shrimps (886 from TAG and 315 from Snake Pit) were collected.

Spatially discrete samples were targeted to explore the fine scale variations in the distribution
of populations of each species – from 10's of centimeters for high densities areas to few meters in
sparsely populated ones – with 13 different spatial samples at TAG and 17 at Snake Pit (Table S1).
During the dives, each sample was kept separate using the different chambers of the suction sampler
carousel and the closeable bioboxes of the HOV *Nautile*. All shrimps were stored in 80% EtOH. For 20
of the sampling points, temperature measurements were carried out prior to shrimps sampling,
moving the HOV temperature probe across several points (0.1°C precision for a temperature below
50°C) within the targeted sampling area, for 3 to 6 min 45s with a sampling frequency of 1Hz.

Additionally, a more precise environmental characterization of the vent fluid chemistry was
obtained for 14 of these sampling points with pH, iron – $Fe^{2+}$ and total iron ($Fe_{tot}$) – and $H_2S$
concentrations measurements. Water samples were collected with the PEPITO sampler implemented
on the HOV Nautile (Sarradin et al., 2007) and equipped with blood bags (Terumo®). The suction inlet
is coupled to the HOV temperature probe. Prior to use, all the equipment used for sampling was
rigorously washed 3 times with diluted hydrochloric acid (pH 2, Suprapur, Merck) and then thoroughly
rinsed with ultrapure water (Mili-Q element system). Immediately after the recovery of the ROV,
samples were processed in the chemical lab onboard (clean lab, P 100 000; ISO8) and pH was measured
on a subsample using a Metrohm pH-meter. Measurements were carried out after calibration of the
pH_meter with NBS buffers (pH 4, 7 and 10) at 25°C. Total iron ($Fe_{tot}$) concentration were determined
onboard on water samples taken by PEPITO using the Ferrozine spectrophotometric method (Stookey,
1970). Free inorganic sulfides ($\Sigma S = HS^- + S^{2-} + H_2S$; (Le Bris et al., 2000) were also determined onboard on
recovered water samples using the Cline spectrophotometric method (Cline, 1969).

In addition to the PEPITO sampling, total dissolvable $Fe^{2+}$ were measured *in situ* using a
chemical miniaturized analyzer (CHEMINI; (Vuillemin et al., 2009). The *in situ* measurement was based
on flow injection analysis with a colorimetric detection (methylene blue method). A calibration of the
analyzer was performed *in situ* at the beginning and at the end of each HOV dive and Fe(II) stock
solutions. Hydrothermal samples were pumped without any filtration and the signal acquisition
(~ 3 min) was initiated at the same time as the water sampling since the same inlet is used for both
PEPITO and CHEMINI. Finally, the *in situ* concentration of dissolved oxygen was monitored during each
sampling using an optode (Aanderaa).

### Life stage identifications and measurements

For each individual, carapace length (CL) was determined to the nearest 0.1 mm by Vernier calipers from the rear of the eye socket to the rear of the carapace in the mid-dorsal line. Each shrimp was identified and sorted according to its life stage and sex. Sex was identified in adults by the occurrence of the appendix masculina on the second pleopod in males, and by the shape of the endopod of the first pleopod (Komai and Segonzac, 2008). Since these sexual characters appear only at the adult stage, sex of juvenile/subadult specimens could not be determined.

Juvenile stages of *R. exoculata* and *R. chacei* were identified according to (Methou et al., 2020). Because pleopod morphologies are identical between subadults and small females, the distinction between them was made according to the Onset of Sexual Differentiation (OSD) size defined previously for *R. exoculata* and *R. chacei*, respectively for a CL of 10 mm and 5.98 mm (Methou et al., 2020) (Hernández-Ávila et al., 2021). All the individuals without male characteristics were sorted as females if their size exceeded OSD, whereas individuals with a smaller size than OSD were sorted as subadults. Subadults and juvenile stages were sometimes considered together as immature individuals in the analyses.

### Statistical analysis of the population structure

*Rimicaris* shrimps were grouped first by species and then, by life stage for each vent field. Visual examination of our dataset and Shapiro-Wilk normality tests revealed that the size distributions of the shrimps collected during the BICOSE 2 expedition were not following a Gaussian distribution. Therefore, non-parametric tests were used for intergroup comparisons, using a Mann Whitney test to compare differences in size between vent fields (Snake Pit vs TAG), species (*R. exoculata* vs *R. chacei*) or sex (females vs males). Size frequency distributions between groups were compared with the Kolmogorov-Smirnov two-sample goodness of fit tests. Significant variations in the proportions of the two *Rimicaris* species or in proportions of immature over adult individuals for each of these two *Rimicaris* species were tested using $\chi^2$ test with Yate's correction for one degree of freedom. Frequencies of males and females in samples were tested for significant variation from a 1:1 sex ratio using $\chi^2$ test with Yate's correction for one degree of freedom (Detailed p-values for these tests are given in electronic supplementary material, table S2). The different spatial samples were grouped by k-means hierarchal clustering based on Euclidean distance matrix and displayed with a heatmap. The optimal number of groups was determined both visually and statistically using the gap statistic method (Tibshirani et al., 2001). All tests were performed using R version 3.6.1.

### Cohort analysis of the population structure

Histograms of the size frequency distributions were analyzed as mixtures of normal distributions for samples containing at least 40 individuals. Size-frequency histograms were plotted using a 1 mm size class interval. This interval was chosen to meet the following criteria described by (Jollivet et al., 2000): (1) most size-classes have at least five individuals each, (2) adjacent empty size-classes have been minimized and (3) an interval clearly larger than the estimated error rate on measurements. We determined the overlapping component distribution that gives the best fit to the histogram with the "mixdist" package (Macdonald and Du, 2012) in R version 3.6.1. Our modal decompositions were considered as valid according to the test from the "mixdist" package, which is based on the $\chi^2$ approximation to the likelihood ratio statistic. This allowed the identification of gamma components and their parameters – mean, sigma, estimated and proportion – each corresponding

194  respectively to the mean CL size, the standard deviation on the CL size and the proportion of a defined
195  cohort.

**Thermal and Environmental niches of each species, sex and life stage**

Four temperature descriptors were calculated for each of the 20 population samples where
temperature was measured: the mean (Avg.T), the minimum (Min.T), the maximum (Max.T) and the
standard deviation (Std.T) (Table S3). Four additional environmental measures – pH, $H_2S$, $Fe^{2+}$ and $Fe_{tot}$
concentration – were used for 14 of these samples (Table S3) for a second more refined niche analysis.
Values below the detection limit of the measurement devices were considered as null for the niche
analysis. Pearson's correlations were used to select non-redundant variables.

The niches of males, females, subadults and juveniles of each shrimp species were examined
similarly to the study of thermal niches of species encountered in *B. azoricus* mussel beds at the Lucky
Strike vent field (Husson et al., 2017). The niche of a set of specimens (i.e. males, females, subadults
or juvenile stages) was studied using the outlier mean index (OMI) created by (Dolédec et al., 2000),
and computed using the "niche" function of R package ade4 (Dray and Dufour, 2007). On a PCA of
environmental parameters, sampling points corresponding to the different population samples
collected are weighted by the specimen's number. The center of gravity of each of these weighted
points is the average position of a group of individuals – defined here by species and life stage or sex
– in the scatterplot defined by the PCA, i.e. the average environmental conditions in which this group
thrives in. Different indexes are given by the OMI analysis including the OMI index, also called
"marginality". A permutation test is used to check the significance of this index, indicating if species
marginality is significantly higher than expected by chance. If this test is not significant, this means that
the species distribution is independent of the thermal conditions. The OMI analysis also gives a
measure of the niche breath: the tolerance index (Tol), which corresponds to the variance around the
centroid. Associated to the tolerance (Tol), the residual tolerance (RTol) indexes, defined as the part
of the variance that is not explained by the environmental variables used in the PCA, indicates whether
the chosen variables are suitable for the niche analysis. These three index (OMI, Tol and RTol)
constitute the total inertia of the niche, and each can be expressed as a percentage of this inertia. At
the same time, OMI analysis also identifies the thermal variables that best differentiate the thermal
niches of the studied group of individuals.

# **Results**

## **Population structures of *Rimicaris* shrimps at TAG & Snake Pit in February-March 2018**

At each vent field, overall size-distribution of the two *Rimicaris* species were drastically
different from each other with significantly distinct size-frequency distributions both at TAG
(Kolmogorov-Smirnov 2-samples test, D = 0.949, p < 0.001) and at Snake Pit vent fields (Kolmogorov-
Smirnov 2-samples test, D = 0.545 p < 0.001). Unlike *R. exoculata*, *R. chacei* size distribution was indeed
skewed towards the smaller sizes, corresponding to the juveniles and subadult life stages (Figure 1).
This results in a significantly higher proportion of immature individuals for *R. chacei* compared to *R.*
*exoculata*, either at TAG (*R. chacei*: 91.9%, *R. exoculata*: 23.8%; $\chi^2$ = 1036.4, p < 0.001; Table 1) or Snake
Pit (*R. chacei*: 52.1%, *R. exoculata*: 32.6%; $\chi^2$ = 45.74, p < 0.001; Table 1). Interestingly, immature
individuals represented a significantly higher proportion of the whole *R. chacei* population at TAG
compared to Snake Pit ($\chi^2$ =240.3, p < 0.001), whereas immature individuals were significantly more
represented in Snake Pit populations of *R. exoculata* compared to TAG ($\chi^2$ = 34.2, p < 0.001).

Nonetheless, a similar proportion of *R. chacei* stage A juveniles were observed in the populations of
both vent fields (TAG: 41.2%, Snake Pit: 38.1%; $\chi^2$ = 0.79, p = 0.373). *R. exoculata* stage A juveniles
however, were still in lower proportions at the TAG vent field (TAG: 3.7%, Snake Pit: 13.9%; $\chi^2$ = 108.03,
p < 0.001).

Overall, larger sizes were observed for *R. exoculata* compared to *R. chacei* at any life stage,
either for adults, subadults or juveniles stages (Table S4; Mann-Whitney test, p < 0.001, all cases).
Largest recorded adults of the two species were nonetheless of similar size at TAG (20.1 mm and 20.5
mm CL respectively for *R. exoculata* and *R. chacei*) but were bigger for *R. exoculata* at Snake Pit (24.4
mm and 21.1 mm CL respectively for *R. exoculata* and *R. chacei*). Males of *R. exoculata* exhibited
significantly larger average sizes than females in both vent fields (TAG: W= 24460, p < 0.001; Snake Pit:
W= 80537, p < 0.001). However, the largest recorded *R. exoculata* adults were females both at TAG
and at Snake Pit (20.1 mm and 24.4 mm CL respectively). In *R. chacei*, the average size of males and
females was not significantly different at Snake Pit (W= 2180, p = 0.065) and males were slightly larger
than females in average at TAG (W= 249.5, p = 0.001). Like for *R. exoculata*, largest recorded *R. chacei*
adults were females at both vent fields (20.5 mm and 21.8 mm CL respectively).

At each vent field, MIX modal decomposition detected four cohorts in the populations of *R.*
*exoculata* and three cohorts in the populations of *R. chacei* (Figure S2). In *R. exoculata*, the second and
third cohorts had the highest proportions of individuals, gathering roughly two third of the
populations, both at TAG and Snake Pit (Table 2). These 2 cohorts mostly gathered adults with some
immatures stages below the OSD in the second cohort. The first cohort only comprised immature
individuals and represented at most 1/5 of the total population. The fourth cohort comprised the larger
adults, and represented at most 1/5 of the total population. In *R. chacei* however, the first cohort, that
gathered mostly immature individuals, comprised half of the population at Snake Pit and up to 92.5%
of the population at TAG.  The second cohort of *R. chacei* comprised small adults mostly, with
proportion varying between 3 to nearly 40% of the whole population. The third cohort comprised large
adults only representing only 4 to 12% of the total population (Table 2).

Populations of the two *Rimicaris* species differed also slightly by their sex ratio. As in 2014
(Hernández-Ávila et al., 2021), sex ratio of *R. exoculata* collected in 2018 from TAG and Snake Pit were
strongly biased towards females at both vent fields (TAG: $\chi^2$ = 690.41, p < 0.001; Snake Pit: $\chi^2$ = 926.56,
p < 0.001; Table 1). Similarly, sex ratio of the *R. chacei* collected at these same vent fields deviated
significantly from 1:1 (TAG: $\chi^2$ = 16.06, p < 0.05.; Snake Pit: $\chi^2$ = 10.07, p < 0.001; Table 1). However,
this deviation from a 1:1 sex ratio was less important in *R. chacei* populations compared to *R. exoculata*
ones with a much larger proportions of males among adult individuals at TAG (*R. chacei*: 26.4%, *R.*
*exoculata*: 11.6% ; $\chi^2$ = 12.38, p < 0.001) and at Snake Pit (*R. chacei*: 37.1%, *R. exoculata*: 12.1% ; $\chi^2$ =
68.92, p < 0.001). No significant variation in the sex ratio of TAG and Snake Pit populations could be
observed as the two vent field harboured similar proportions of males, whether for *R. chacei* ($\chi^2$ = 0.12,
p = 0.732) or for *R. exoculata* ($\chi^2$ = 2.04, p = 0.153).

**Spatial variation in the population structure of *Rimicaris* shrimps from TAG & Snake Pit**

Discrete spatial variations in the structure of the shrimp assemblages were explored. Our30
samples were collected into six visually distinct types of assemblages (Figure 2) according to the
dominant *Rimicaris* life stage/species/sex and the proximity of the sampled assemblage to the
hydrothermal activity. More specifically, these assemblages were classified as close to active emissions

for those adjacent to black smokers, close to diffuse emissions for those in a close proximity to
transparent shimmering flows, or in inactive areas when no vent fluids could be detected visually in
their surroundings. A number of six clusters was also determined statistically to be the most optimal
number of clusters for our population dataset using the gap statistic method (Figure 3).

The most visually dominating assemblage type was the dense aggregates observed close to
the active vent fluid emissions (Figures 2A and S3). These assemblages comprised almost exclusively
*R. exoculata*, with *R. chacei* sometimes also present but in very low proportions ($\chi^2$ = 93.5 – 714.1, p <
0.001). Most of the dense aggregates were dominated by adults of *R. exoculata*, with a larger
proportion of adults in all samples ($\chi^2$ = 47.1 – 501.1, p < 0.001), except in one from Snake Pit (The Nail
edifice) that exhibited similar proportions of adults and immature individuals ($\chi^2$ = 0.46, p = 0.049).
Immatures from this population sample were in majority at an advanced juvenile stage (i.e. subadults
or stage B juvenile), early juveniles only representing 2.6% of all the *R. exoculata* individuals. Large
variations were observed in the proportions of immature stages between the different aggregates
from 5.8% to 32.6% of all the *R. exoculata* individuals. For nearly all the dense aggregates, sex ratio of
*R. exoculata* strongly deviated from 1:1 ($\chi^2$ = 31.2 – 311, p < 0.001), with a large bias toward females.
One exception though, was found in a dense aggregate from TAG exhibiting an equilibrated sex ratio
for *R. exoculata* ($\chi^2$ = 0.69, p > 0.05).

At the periphery of the dense aggregates, in parts less impacted by the turbulent mixing
between hydrothermal fluid and seawater, or where fluid exits are dampened, a second type of
assemblage was observed at Snake Pit (The Beehive), also largely dominated by *R. exoculata* (Figures
2B and S4A). Unlike dense aggregates, these two population samples exhibited large proportions of
immatures ($\chi^2$ = 202.2, p < 0.001; 231.2, p < 0.001) representing 90.4% and 91.3% of all the *R. exoculata*
individuals collected there. These were called nurseries of *R. exoculata*. Most of these immature
individuals were at an early juvenile stage, with stage A juvenile representing 70.4% and 74.5% of all
the immature individuals. In addition, sex ratio of the few *R. exoculata* adults from these assemblages
did not significantly deviated from 1:1 ($\chi^2$ = 3.2, p = 0.07) or only slightly toward females ($\chi^2$ = 5.4, p =
0.02). As for the dense aggregates, *R. chacei* shrimps were nearly absent from *R. exoculata* nurseries
with only two *R. chacei* juveniles collected in total.

Two other types of shrimp assemblages observed close to diffuse vent fluid emissions were
dominated by *R. chacei* ($\chi^2$ = 16.5 – 888.1, p < 0.001) representing 77.4% to 99.5% of all the *Rimicaris*
shrimps collected. These were generally not directly adjacent to the dense aggregates, but clearly
separated even if the distance was not necessarily important. One of these two types of assemblages
was collected in a hole between rocks at TAG or between *Bathymodiolus puteoserpentis* mussels at
Snake Pit (The Moose edifice). Although individuals were rather densely packed, they were hard to see
due to the relatively small size of the assemblages and their "hidden" localization. These assemblages
were called "hidden aggregates" of *R. chacei* (Figure 2C and S4B). These hidden aggregates clearly
exhibited larger proportions of *R. chacei* adults than immatures ($\chi^2$ = 35.6 – 60.1, p < 0.001), adults
representing 71.7% to all the *R. chacei* individuals collected there. Sex ratio of *R. chacei* from these
assemblages was strictly equilibrated for the sample collected at TAG ($\chi^2$ = 0, p = 0.99) but slightly
deviated toward females for the one collected at Snake Pit ($\chi^2$ = 7.45, p = 0.006).

The second type of assemblages dominated by *R. chacei*, was observed close to diffuse
emissions exiting from cracks at the bottom of vent chimneys or in flat areas of the vent fields (Figures
2D and S4C). Contrary to the hidden aggregates, significantly larger proportions of immature

individuals were observed in these nurseries of *R. chacei* ($\chi^2$ = 36.8 – 868.5, p < 0.001), representing at
least 96.7% of all the individuals of the species. In general, most of these immature individuals were at
an early juvenile stage, with stage A juvenile representing 69.9% to 82.6% of all the immature
individuals. Though, one *R. chacei* nursery from TAG exhibited a larger proportion of *R. chacei*
subadults than stage A juveniles.

A last assemblage type found in areas of moderate fluid influence was observed at TAG in the
vicinity of dense aggregates (Figures 2E and S4D). These assemblages were characterized by similar
proportions of the two *Rimicaris* species ($\chi^2$ = 2.9, p > 0.05) or only slightly larger proportions of one of
the two species ($\chi^2$ = 7.5 – 7.9, p < 0.05). Larger proportions of *R. exoculata* adults ($\chi^2$ = 24 – 30.1, p <
0.001) and larger proportions of *R. chacei* immatures ($\chi^2$ = 56.4 – 208.9, p < 0.001) were usually
observed in these assemblages although similar proportions of adults and immature individuals of both
species were found for one of them ($\chi^2$ = 0.4 – 2.5, p > 0.05). In all of them, most of the immature
shrimps were at an advanced juvenile stage for both species, stage A juveniles representing at most
7.3% and 2.4% of all the immature individuals respectively for *R. exoculata* and for *R. chacei*. In general,
sex ratio of *R. exoculata* from these assemblages did not significantly deviated from 1:1 ($\chi^2$ = 0.1 – 3, p
> 0.05), except in one with a large bias toward females ($\chi^2$ = 117.4, p < 0.001).

At last, *Rimicaris* shrimps were also collected at the periphery of the vent chimneys, on inactive
sulfides away from the vent fluid activity (Figures 2F and S4E). These assemblages of scattered
individuals exhibited either similar proportions of the two *Rimicaris* species ($\chi^2$ = 0.1 – 2, p > 0.05) or
larger proportions of *R. exoculata* individuals ($\chi^2$ = 26.4 p < 0.001). As in dense aggregates, larger
proportions of *R. exoculata* adults over immature individuals were observed in all of these assemblages
($\chi^2$ = 16.2 – 44.6, p < 0.001). On the other hand, both larger proportion of *R. chacei* adults ($\chi^2$ = 20.2, p
< 0.001) and larger proportions of *R. chacei* immature individuals ($\chi^2$ = 28.9, p < 0.001) were reported
in these two different population samples. Sex ratio of *R. exoculata* from these assemblages deviated
from 1:1 ($\chi^2$ = 6.4 – 16.3, p < 0.001) with a large bias toward males. Unlike *R. exoculata*, sex ratio of *R.*
*chacei* from these assemblages did not significantly deviated from 1:1 ($\chi^2$ = 3, p = 0.11).

One population sample, collected in a flat area of moderate fluid emissions away from an
active venting chimney at Snake Pit (The Nail edifice), did not exhibited a clear structure corresponding
to one of the six assemblages defined above, and clustered with the hidden aggregates of *R. chacei*
adults (Figure 3). However, unlike samples from this last assemblage, this population showed similar
proportions of both *Rimicaris* species ($\chi^2$ = 0.2, p > 0.05). Moreover, proportion of adults over immature
individuals were similar ($\chi^2$ = 0.1 – 1.4, p > 0.05) and sex ratio did not significantly from 1:1 ($\chi^2$ = 0.1 –
0.11, p > 0.05) both for *R. exoculata* and *R. chacei*. This assemblage also differs from the others by
being dominated by *Mirocaris fortunata* (58.9% of all the shrimps sampled), another co-occurring
alvinocaridid at these sites that were only observed in 10 of our samples in very low proportions.

## Thermal & environmental niches of *Rimicaris* shrimps and their life stages from TAG & Snake Pit

358  Average temperatures around *Rimicaris* assemblages were comprised between 3.4°C and
359  22.5°C in average, with temperature reaching up to 38.1°C and minimum records down to 3.1°C.
360  Standard deviations that can be considered as the thermal variability around the assemblages varied
361  from 0.3°C to 6.9°C. Concerning the vent fluid chemistry, $H_2S$ concentrations around *Rimicaris*
362  assemblages were comprised between 0 µM and 8.5 µM in average with pH variations between 6.86

and 7.9 as well. It is worth mentioning that total sulfides concentrations were measured onboard on recovered samples and may be well underestimated compared to the actual in situ conditions. In situ iron concentrations around *Rimicaris* assemblages were highly variables with $Fe^{2+}$ concentrations comprised between 0 µM and 239 µM and $Fe_{tot}$ concentrations between 0 µM and 316 µM. As the minimal temperatures were strongly correlated with average temperature ($r^2$=0.97, p < 0.05), only average temperature (Avg.T), temperature standard deviation (Std.T) and maximum temperature (Max.T) were used in the two PCA. No clear significant correlations were observed between any of the thermal descriptors and the concentrations of the different reduced chemical elements (Pearson correlations, p > 0.01 in all cases).

For the thermal niches (Figure 3A and 3B; 20 sampling points), most of the variance was explained by the two first axes of the PCA representing respectively 93.6% and 5.3% of the inertia. All three temperature metrics were very well represented on the PCA ($\sum cos^2$ = 0.99, 0.99 and 0.98 for Avg.T, Std.T and Max.T respectively) and were strongly correlated with the first axis (cor = 0.96, 0.95 and 0.99 for Avg.T, Std.T and Max.T respectively). Avg.T, Std.T and Max.T almost contributed equally to the construction of the first axis (contribution = 32.7%, 32.4% and 34.9% respectively). Similarly, Avg.T and Std.T also contributed equally to the construction of the second axis (contribution = 46.4% and 53.4% respectively).

In the case of the environmental niches (Figure 3C and 3D; 14 sampling points), a large part of the variance was explained by the two first axes of the PCA representing respectively 64.7% and 22.4% of the inertia. All the environmental parameters were relatively well represented on the two first axes of the PCA with $\sum cos^2$ > 0.85 for all parameters except for Std.T and pH ($\sum cos^2$ = 0.81 and 0.70 respectively). Thermal descriptors as well as pH correlated more with the first axis (cor = -0.97, -0.88, -0.95 and 0.83 for Avg.T, Std.T, Max.T and pH respectively) whereas $H_2S$ measurements were more correlated with the second axis (cor = -0.83). Iron concentrations were equally correlated with the first (cor = -0.75 and -0.67 for $Fe^{2+}$ and $Fe_{tot}$ respectively) and the second axis of the PCA (cor = 0.62 and 0.64 for $Fe^{2+}$ and $Fe_{tot}$ respectively). The thermal descriptors contributed the most to the construction of the first axis (contribution = 21%, 17% and 20.1% respectively for Avg.T, Std.T and Max.T), followed by pH and $Fe^{2+}$ and $Fe_{tot}$ (contribution = 15.4%, 12.5% and 10% respectively). On the other hand, $H_2S$ contributed in large part to the construction of the second axis followed by iron concentrations (contribution = 24.3% and 25.8% respectively for $Fe^{2+}$ and $Fe_{tot}$).

As revealed by the OMI analysis, stage A juveniles of *R. exoculata* and *R. chacei* occupied distinct thermal niches compared to their older counterparts (Figure 4A). The thermal niches of these early life stages differed both in terms of niche breadth and overall position (Figure 4B). Indeed, stage A juveniles of *R. exoculata* exhibited a much lower tolerance (Tol) than the other adults or immature stages of this species (Figure 4B). Similarly, stage A juveniles of *R. chacei* also displayed a lower tolerance (Tol) than the other *R. chacei* life stages, as well as a mean niche position that significantly deviated from the average sampled thermal conditions (Permutation test, p < 0.05). Additionally, permutation tests also showed a significant marginality for *R. exoculata* females and subadult stages (p < 0.05; Table 3) despite a rather similar tolerance (Tol) compared to *R. exoculata* males and stage B juveniles (Table 3). Conversely, niche position of *R. chacei* adults and subadults did not significantly departed from the average sampled thermal conditions. Overall, every *R. exoculata* life stages presented a larger niche breadth than any of the *R. chacei* life stages (Figure 4A). The same trends were also observed in the more detailed environmental niches with distinct niches for the earliest

juvenile stages of both *R. exoculata* and *R. chacei*, compared to all the other life stages (Figure 4C and
4D) and a lower niche breadth for *R. chacei* compared to any *R. exoculata* life stages (Figure 4C and
4D), although significant marginality were only found for *R. exoculata* females ($p < 0.05$; Table 3). For
the thermal niches, residual tolerance (RTol) ranged from 0.3% to 10.2% of total inertia; thus the
variables used to describe these niches explained 89.8% to 99.7% of the niche distribution of the
different *Rimicaris* life stages (Table 3). On the other hand, residual tolerance (RTol) for the
environmental niches ranged from 26.6% to 49% of total inertia; thus the variables used to describe
these niches only explained 51% to 74.4% of the niche distribution of the different *Rimicaris* life stages
(Table 3).

## Discussion

### Small scale distribution of *Rimicaris* life stages along the environmental gradient

Our temperature measurements indicate that thermal conditions around the different *R.*
*exoculata* assemblages of TAG and Snake Pit are comprised between 3.4°C and 22.5°C with maximal
peaks of temperatures up to 38.1°C (Table 3). These results slightly extend the thermal range occupied
by these shrimps compared to previous *in situ* temperature measurements around *R. exoculata*
aggregates, which reported average thermal conditions between 4.6°C and 16°C and up to 25°C at
maximum (Desbruyères et al., 2001; Geret et al., 2002; Zbinden et al., 2004; Schmidt et al., 2008).
Although *R. exoculata* reach 100% of mortality when exposed during 1h at 33°C only in closed
pressurized chambers, these shrimps tolerate better shorter exposures of 10 min with only 35% of
mortality at 39°C (Ravaux et al., 2019). With a defined critical thermal maximum (CTmax) of 38°C for
this species, it thus appears that *R. exoculata* shrimps occupy *in situ* most of its fundamental thermal
niche. We also noticed important movements of shrimps in aggregates where average temperature
exceed 20°C – PL06A1, PL15A3 & PL20A1 – whereas the others tend to remain steady confirming the
*in vivo* behavioral observations of Ravaux et al. (Ravaux et al., 2019). On the other hand, we observed
that *R. chacei* inhabits a narrower thermal niche, comprised between 3.4°C and 14.5°C – up to 20.7 °C
at maximum – whatever the life stage considered (Table 3). In the absence of experimental work on *R.*
*chacei* thermotolerance, it is unclear if their more limited distribution within the hydrothermal vent
gradient stems from a physiological limitation or from biological interactions such as competition with
its congeneric species. Part of the answer can be brought by a comparison with *M. fortunata*, another
alvinocaridid shrimp from the MAR, with a thermal tolerance similar to *R. exoculata* (Shillito et al.,
2006) but occupying a more restricted thermal habitat (Desbruyères et al., 2001) (Methou, personnal
observations). The narrower distribution of *R. chacei* could also simply be related to a lower thermal
preference, as seen in many vent species that preferentially selects habitats with cooler temperatures
than their upper thermal limit (Bates et al., 2010). This all suggests that temperature *per se* is not a
sufficient limiting factor to explain differences in alvinocaridid shrimp distributions.

Similarly, our dataset also point out that *R. exoculata* inhabits in a wider range of reduced
compound concentrations (traced through *in situ* total dissolvable iron)  than *R. chacei* for a given life
stage, although this trend was much less marked between the earliest juvenile stages of the two
species. Our *in situ* chemistry measurements around biological communities indicates significant
deviation from the conservative dilution model as it has already been reported in previous studies (Le
Bris et al., 2006, 2019; Nakamura and Takai, 2014). This deviation from a conservative dilution of the
endmember originates from the highly reactive nature of the sulfides and iron species and the
consumption or production of these reduced compounds by microbial activities.

At both fields, an important zonation between adults and juvenile stages was also found for *R.*
*exoculata* and for *R. chacei*. This spatial segregation along the thermal gradient is reflected in the
distinct thermal niches observed between the earliest juveniles and the other life stages for the two
*Rimicaris* species. Similar differences in distribution of bathymodiolin mussel life stages have been
hypothesized to be related to intraspecific competition for space between adults and juveniles (Comtet
and Desbruyères, 1998; Husson et al., 2017). This was further supported by evidence of recruitment
inhibition by high density assemblages of adult mussels from the East Pacific Rise (Lenihan et al., 2008).
Still, an influence of intraspecific competitive interactions for *Rimicaris* shrimps is tempered by their
high motility as well as the non-negligible presence of early juveniles within several assemblages of
high densities dominated by adult stages. A progressive acquisition of a high thermal tolerance of these
recruiting stages, as seen in post-larvae stages of the intertidal shrimp *Palaemon elegans* (Ravaux et
al., 2016), could also limit their upper thermal distribution. Additionally, distribution of *R. exoculata*
and *R. chacei* juveniles further differ by their habitat topography, on steep chimney walls for *R.*
*exoculata* nurseries and within flat areas or smooth slopes for *R. chacei* nurseries. As topography
strongly influences the overall current hydrodynamics, with low topography areas generally associated
with low mixing rates (Girard et al., 2020), these structural differences in habitats might reflect the
relative dependency of each species to the vent fluid mixing gradient essential to fuel their symbiosis.
Although our measurements were limited, this variability of the fluid chemistry between the nursery
habitats of the two species could be observed in our dataset with lower $H_2S$ and Fe concentrations in
*R. chacei* nurseries compared to *R. exoculata* ones.

We also confirmed the segregation pattern between males and females of *R. exoculata*
observed at TAG in January-February 2014 (Hernández-Ávila et al., 2021)for both of the studied vent
fields. Such pattern was not observed in *R. chacei*, which also exhibited a more equilibrated sex ratio,
more similar to what has been observed for its sister species *R. hybisae* (Nye et al., 2013).

Despite an apparent relative stability with the maintained presence of an assemblage type in
the same area over the years, the spatial zonation of the different *Rimicaris* life stages also seems
dynamic with a succession of assemblages in other areas on short temporal scales. During one of our
dives, we witnessed at TAG, the entire disappearance of a dense aggregation assemblage of *R.*
*exoculata* adults replaced by an aggregate of *R. chacei* hidden between rocks in just about ten days
(Figure S5). Unfortunately, we are currently unable to confirm if this succession of assemblages is
related to a modification of the environmental conditions, as we could not obtain temperature
measurements in the first place, prior to the replacement of the dense aggregate, due to technical
difficulties during the dive.

In addition to the assemblages dominated by a particular life stage of one of the two *Rimicaris*
species, different types of mixed assemblages were also observed in the areas a little further away
from the fluids. Interestingly, these assemblages with a mixed proportion of the two species were also
the assemblages with the lowest densities of *Rimicaris* individuals, visually countable unlike within the
other assemblage types (Figure 2.). These more peripheral assemblages with less marked structuration
could correspond to assemblages with much lower densities due to suboptimal habitat conditions for
*Rimicaris* spp. Lower supply of reduced compounds to feed their cephalothoracic symbionts, or an
absence of a symbiont pool to renew their symbiotic communities along their molt cycle would limit
the establishment of these species in these areas. A finer environmental characterization of these
particular areas would be required to better understand the subtle ecological differences that could
exist between these habitats and the more dense shrimp assemblage types. In any cases, these density

dependent trends suggest that competitive processes might occur between *R. chacei* and *R. exoculata*
life stages.

**Population dynamics of *Rimicaris* shrimps from the MAR are driven by biotic interactions**

As most species from hydrothermal vents, *R. exoculata* and *R. chacei* exhibited multimodal
size-frequency distribution suggesting a discontinuous recruitment pattern in both species, thus
confirming the previous observations made for *R. exoculata* in 2014 (Hernández-Ávila et al., 2021).
Still, population structures of the two congeneric *Rimicaris* species from the Mid Atlantic Ridge differed
widely with a large post recruitment collapse of *R. chacei* populations at both vent fields. This collapse
was even more pronounced for the TAG population where immature individuals represented more
than 90% of all the individuals collected. This implies either a high mortality rate of *R. chacei* early post-
settlement stages in comparison to *R. exoculata* and/or a large variability in the magnitude of the
different larval settlement events. Although temporal tracking of these populations could not be
performed, important aggregations of *R. chacei* juveniles in nurseries were already reported between
January and February 2014 (Hernández-Ávila et al., 2021) as well as between March and April 2017
(Methou et al., 2020). It is then unlikely that the observed variations between *R. chacei* life stages
abundance would be related to variations in larval supply and/or settlement success between years.

Therefore, we hypothesize that the post recruitment collapse observed in *R. chacei*
populations result more from a higher juvenile mortality that could be explained by different factors,
not necessarily mutually exclusive from each other (Figure 5). Among them, predation could have a
major impact on *R. chacei* demography, as it has been recognized as one of the most important cause
of juvenile mortalities in shallow water ecosystems (Gosselin and Qian, 1997). The more peripheral
distribution of *R. chacei* juveniles within the vent fields could expose them to a higher predation
pressure by populations of *Maractis rimicarivora* anemones found in these areas and particularly
abundant at TAG (Copley et al., 2007). Indeed, we could witness the catch of one juvenile from the *R.*
*chacei* nursery area by these anemones during a dive at TAG in 2014 (Supplementary video material
1). Predation by these carnivorous anemones on *R. exoculata* adults have already been observed (Van
Dover et al., 1997) and was proposed to affect more largely male individuals from the peripheral areas
(Hernández-Ávila et al., 2021) which could explain, at least partially, the large sex ratio bias toward
females of this species. Additionally, a larger vulnerability to predation by macrourid fishes could also
affect preferentially *R. chacei* juveniles. Despite a limited observation time on these vent field areas,
we reported twice predation events by these fishes within or close to *R. chacei* nurseries assemblages
during our samplings at TAG, both in 2014 and in 2018 (Supplementary video material S1). This higher
predatory pressure at TAG would also be in agreement with the greater collapse observed in the
populations of *R. chacei* at this site compared to those from Snake Pit (Figure 1.). Contrastingly, the
distribution of *R. exoculata* early juveniles, sheltered on the flank of large chimney structures or within
the dense aggregates of *R. exoculata* adults probably protect them from this high predatory pressure.

In addition to their distinct spatial distribution, the juvenile stages of the two *Rimicaris* shrimps
differ in their average settlement size, twice as large for *R. exoculata* compared to *R. chacei* (Methou
et al., 2020). Numerous studies on marine fishes and shrimps have reported body size as a key factor
in the survival of early juveniles with a lower overall mortality for larger individuals (Meekan et al.,
2006; Gagliano and McCormick, 2007; Mace and Rozas, 2018). Thus survival of *Rimicaris* juveniles could
fit into the "bigger is better" hypothesis, whether larger body size would provide them a lower
vulnerability against predation, a better physiological state and/or a higher competitive ability.

As an alternative to a high juvenile mortality, important emigration of *R. chacei* juveniles –
secondary migration – in a different location following their initial settlement at the vent fields could
also lead to similar patterns in their population structure. A long-term monitoring with video imagery
of the areas of *Rimicaris* nurseries or individual tracking early stage shrimps could help to refine our
hypotheses and determine more precisely the relative influence of these different factors on their
population dynamics.

Interspecific competition for space between the two species, already suggested by their spatial
segregation patterns, could also have a major influence on their demography. It is commonly accepted
that abundant species within an assemblage outcompete less common ones. This general statement,
however, has been contradicted by various manipulative studies which did not reported any effect of
being naturally rare or common on the competitive ability of a species (Angel et al., 2006; Matias et
al., 2012). Despite this, we still argue that the low abundance of *R. chacei* adults might result from a
lower competitive ability of their early juvenile stages to access areas of active fluid emissions, as
suggested by their narrower thermal niche compared to similar stage of *R. exoculata* (Figure 4.).
Marine lobsters for instance also experienced important demographic bottleneck before recruitment,
due to the more specific requirements of their early benthic juvenile stages for sheltered areas, which
are limited in space (Wahle and Steneck, 1991).

In hydrothermal vents, competitive processes for space are confounded with competition for
food in symbiotrophs whose nutrition depend entirely on their access to the vent fluid reduced
elements. Both *R. exoculata* and *R. chacei* are fueled by large communities of filamentous
ectosymbionts (Gebruk et al., 2000; Ponsard et al., 2013; Apremont et al., 2018), which progressively
make up during their ontogeny an increasing part of their diet after recruitment (Methou et al., 2020).
A more limited access to areas of active fluids for *R. chacei* juveniles could potentially lead to a greater
limitation to primarily acquire their episymbionts and develop a sufficient symbiotic community within
their cephalothorax, which could thus cause mass mortality in these populations. In this scenario, the
mixotrophic behavior of *R. chacei* would be the result of a past competition with *R. exoculata* leading
to a trophic niche shift in *R. chacei* adults over evolutionary timescales allowing the stable coexistence
of the two species in the northern MAR vent fields. Additionally, the more limited local diversity at TAG
(Desbruyères et al., 2001), likely related to the absence of *Bathymodiolus* mussel beds known to house
large macrofaunal diversity (Desbruyères et al., 2001; Rybakova and Galkin, 2015), enhance even more
the limitation in alternative food resources for *R. chacei* thus possibly explaining the larger post
recruitment collapse of their population at this site. These shrimps are not the first case of population
demographics driven by competitive interactions for food in hydrothermal vent ecosystems.
Populations of *Paralvinella pandorae* from Juan de Fuca Ridge exhibited indeed similar post settlement
collapses when found in sympatry with their congeneric species *P. palmiformis*, whose isotopic niches
largely overlap (Levesque et al., 2003).

This link between nutrition modes and population demographics of these shrimps is also
further supported when their population structures are compared with those of *R. hybisae* from the
Mid Cayman Spreading Center. The latter is extremely close phylogenetically to *R. chacei*, insofar as
the two species cannot be distinguished on the sole basis of the genetic markers used so far
(Vereshchaka et al., 2015). In this way, *R. hybisae* populations from the Mid Cayman Rise can almost
be seen as a case of *R. chacei* allopatry with no known competitors like *R. exoculata*. However, *R.*

*hybisae* size-frequency structures largely differ from those of *R. chacei* and appear to be more similar
to those of *R. exoculata* (Nye et al., 2013).

Unlike *R. exoculata*, *R. chacei* adults were more difficult to catch with the suction sampler
during our attempts to collect them as they were sometimes escaping the tip of the sampler when we
approached them. This escape behavior was not observed for the juvenile specimens of the same
species inhabiting the nurseries. It is therefore possible that our dataset might slightly underestimate
the exact proportion of *R. chacei* adults in the field. Nonetheless, it is highly unlikely that this potential
underestimation of *R. chacei* adult number would affect the strong patterns of population structure
observed for this species at both vent field. Similarly, we believe that the disequilibria in sampling
effort between the different assemblage types, with an overrepresentation of the dense aggregates
against the other assemblage types, is representative of their larger importance within the
hydrothermal vent fields. A detailed cartography giving the distribution of these assemblages, with
shrimp densities estimations when possible, over the entire vent fields would help to better assess our
dataset representativeness and if our sampling strategy could have affected partially our conclusions
on the population dynamics of these two species.

The mechanisms by which *R. chacei* populations maintain a high supply of settlers despite a
limited adult number comparatively to *R. exoculata* are still unclear. Larger *R. chacei* populations from
yet unknown vent fields in the MAR representing source populations – with TAG and Snake Pit as sink
populations – at the metapopulation level or differences in reproductive output between *R. exoculata*
and *R. chacei*. Additional work on the reproductive biology of *R. chacei* will be needed to better
understand differences in the life history biology of these species. Further experimental studies on the
thermal physiology and tolerance to the vent fluid toxic compounds of both *R. chacei* adults and
*Rimicaris* juveniles stages would also be required to understand if the distribution patterns observed
between these different life stages are related to physiological limitations or other biotic factors.

## Ethics

604 The animals used in this study were collected on the French contract area for the exploration of polymetallic
605 sulfur deposits and are not considered as endangered species. Each individual were immediately preserved in
606 ethanol after recovery and no live experiments were carried out.

## Author's contributions

P.M. collected and sorted the specimens in 2018, carried out the measures and identification of specimens, lead
data analysis, participated in the conception and design of the study and drafted the manuscript; I.H-A. helped
to collect and sort the specimens in 2018 and critically revised the manuscript; C.C. carried out the collection of
environmental measurement in the field, provide her help in the analysis of these data and critically revised the
manuscript; M-A.C-B. coordinated the study and helped draft the manuscript; F.P. helped to collect and sort the
specimens in 2018, conceived and designed the study, coordinated the study and helped draft the manuscript.
All authors gave final approval for publication and agreed to be held accountable for the work performed therein.

## Competing interests

The authors declare they have no competing interests.

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**<u>Figures</u>**

**Figure 1.** Size-frequency distribution of *Rimicaris exoculata* and *R. chacei* from two vents fields,
denoting life stages and sexes. **A.** *R. exoculata* from TAG. **B.** *R. exoculata* from Snake Pit. **C.** *R. chacei*
from TAG. **D.** *R. chacei* from Snake Pit.

**Figure 2.** Overview of the different *Rimicaris* assemblage types from TAG and Snake Pit **A.** Dense aggregate of *R. exoculata* adults observed near vent fluid exits. **B.** Nursery of *R. exoculata* adjacent to the dense adult aggregate on the flank of chimneys at Snake Pit. **C.** Hidden aggregate of *R. chacei* observed behind a mussel bed at Snake Pit. **D.** *R. chacei* nursery associated with low temperature diffusions. **E.** Low density alvinocaridid assemblage at the periphery of a dense aggregate of *R. exoculata* at TAG. **F.** Scattered *Rimicaris* individuals, in areas away from any visible vent fluid emissions.

**Figure 3.** Heat map displaying hierarchical clustering of the different samples collected in this study, according to their composition (in terms of species, sexes and life stages).

**Figure 4.** Niches of *Rimicaris* shrimps according to their species, sex and life stages. **A.** Two dimensional
plot of the OMI analysis. Each ellipse represents the thermal niche of a *Rimicaris* group (defined by its
species, life stage, and sex in adults). **B.** *Rimicaris* groups total marginality (OMI index) versus total
tolerance of their thermal niches. **C.** Two dimensional plot of the OMI analysis. Each ellipse represents
the environmental niche of a *Rimicaris* group (defined by its species, life stage, and sex in adults). **D.**
*Rimicaris* groups total marginality (OMI index) versus total tolerance of their environmental niches.

**Figure 5.** Schematic view of the different Rimicaris assemblage distribution at hydrothermal vents from
the Mid Atlantic Ridge with current hypotheses on the factors affecting their population dynamics.

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850

## Tables

| Species | Vent field | Total n. specimens | Adults | | | Immatures | | | | Sex Ratio | | |
|---|---|---|---|---|---|---|---|---|---|---|---|---|
| | | | Total (TA) | Female (TF) | Male (TM) | Total (TI) | Subadult | Juvenile B | Juvenile A | TM:TF | χ2 (1 df) | Significance |
| *Rimicaris exoculata* | TAG | 1533 | 1168 | 1033 | 135 | 365 | 155 | 153 | 57 | 0.13:1 | 690.41 | *** |
| | Snake Pit | 2386 | 1609 | 1415 | 194 | 777 | 232 | 212 | 333 | 0.14:1 | 926.56 | *** |
| *Rimicaris chacei* | TAG | 884 | 72 | 53 | 19 | 812 | 448 | - | 364 | 0.36:1 | 16.06 | *** |
| | Snake Pit | 315 | 151 | 95 | 56 | 164 | 44 | - | 120 | 0.59:1 | 10.07 | ** |

**Table 1.** Sample and population data of *Rimicaris exoculata* and *Rimicaris chacei* from TAG and Snake Pit.

| Species | Vent field | Cohort 1 | | | Cohort 2 | | | Cohort 3 | | | Cohort 4 | | |
|---|---|---|---|---|---|---|---|---|---|---|---|---|---|
| | | % | μ | σ | % | μ | σ | % | μ | σ | % | μ | σ |
| *Rimicaris exoculata* | TAG | 14.5 | 8.6 | 0.6 | 25.2 | 10.3 | 0.6 | 39.6 | 11.9 | 1.3 | 20.7 | 16.5 | 1.3 |
| | Snake Pit | 21.7 | 7.9 | 0.7 | 33.5 | 10.5 | 0.9 | 33.6 | 13.7 | 2.4 | 11.2 | 20.4 | 1.4 |
| *Rimicaris chacei* | TAG | 92.5 | 4.7 | 0.4 | 3.3 | 6.8 | 0.3 | 4.2 | 15.3 | 3.2 | - | - | - |
| | Snake Pit | 48.9 | 4.6 | 0.3 | 39.4 | 9.4 | 2.5 | 11.7 | 17.4 | 2.4 | - | - | - |

**Table 2.** Cohort analysis with proportions (%), average (μ) and standard deviations (σ) of the modal components estimated from the size–frequency distributions of *Rimicaris exoculata* and *Rimicaris chacei* collected from TAG and Snake Pit.

| | n | Average temperatures (°C) | | | Thermal variability (°C) | | | Thermal range (°C) | | H₂S (µM) | Fe²⁺ (µM) | Fe_tot (µM) | OMI analysis | | | | | | | Permutation |
|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|---|
| | | mean | min | max | mean | min | max | min | max | mean | mean | mean | inertia | OMI | Tol | Rtol | omi | tol | rtol | p.value |
| **R. exoculata females** | 1249 | 12.1 | 3.4 | 20.3 | 3.4 | 0.3 | 5.9 | 3.1 | 38.1 | 3.7 | 41 | 87.9 | 9.01 | 0.56 | 4.82 | 3.63 | 6.2 | 53.5 | 40.3 | NS |
| **R. exoculata male** | 188 | 9.2 | 3.4 | 20.3 | 2.2 | 0.3 | 5.9 | 3.1 | 38.1 | 2.7 | 25.6 | 71.6 | 8.17 | 0.12 | 4.58 | 3.47 | 1.5 | 56,0 | 42.5 | NS |
| **R. exoculata subadults** | 143 | 10.3 | 3.4 | 20.3 | 2.8 | 0.3 | 5.9 | 3.2 | 38.1 | 3.8 | 36.4 | 80.5 | 10.55 | 0.75 | 5.26 | 4.54 | 7.1 | 49.9 | 43,0 | NS |
| **R. exoculata stage B juveniles** | 263 | 10.1 | 3.4 | 20.3 | 2.9 | 0.5 | 5.9 | 3.2 | 38.1 | 2.7 | 35.2 | 79.2 | 9.75 | 0.52 | 5.45 | 3.78 | 5.4 | 55.9 | 38.8 | NS |
| **R. exoculata stage A juveniles** | 300 | 5.4 | 3.4 | 20.3 | 1 | 0.3 | 5.9 | 3.1 | 38.1 | 0.8 | 4.3 | 16.6 | 5.71 | 0.14 | 3.19 | 2.39 | 2.4 | 55.8 | 41.9 | NS |
| **R. chacei females** | 36 | 6.1 | 3.6 | 9.7 | 1.2 | 0.5 | 3.9 | 3.3 | 20.7 | 2.1 | 8.8 | 22.4 | 2.21 | 0.83 | 0.56 | 0.82 | 37.6 | 25.4 | 37,0 | NS |
| **R. chacei male** | 26 | 5.4 | 3.4 | 9.7 | 1.1 | 0.3 | 3.9 | 3.1 | 20.7 | 2.3 | 14.5 | 34.2 | 2.42 | 0.71 | 0.52 | 1.18 | 29.6 | 21.5 | 49,0 | NS |
| **R. chacei subadults** | 187 | 4.6 | 3.4 | 8.2 | 1.2 | 0.3 | 3.9 | 3.1 | 20.7 | 0 | 10.3 | 36.5 | 3.59 | 1.45 | 0.45 | 1.69 | 40.3 | 12.6 | 47.1 | NS |
| **R. chacei stage A juveniles** | 322 | 3.8 | 3.4 | 8.2 | 0.5 | 0.3 | 3.9 | 3.1 | 20.7 | 0.02 | 0.3 | 12.2 | 3.79 | 2.18 | 0.6 | 1.01 | 57.5 | 15.9 | 26.6 | NS |

**Table 3.** Thermal range and parameters of the OMI analysis for the different *Rimicaris* groups.