

1 **Area-based conservation planning in Japan: protected area network effectiveness to**
2 **the post-2020 Global Biodiversity Framework**

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12

13 **ABSTRACT**

14 To reframe the imperfect review processes of nation-scale actions on area-based
15 conservation through protected area (PA) networks, we first created novel infrastructure
16 to visualize nation-level biodiversity information in Japan. We then assessed the
17 performance of the existing PA network relative to land exploitation pressure and
18 evaluated conservation effectiveness of PA expansion for the post-2020 Global
19 Biodiversity Framework. The Zonation algorithm was used to spatially prioritize
20 conservation areas to minimize biodiversity loss and the extinction risk for 8,500
21 Japanese vascular plant and vertebrate species under constraints of the existing PA
22 network and land use. The spatial pattern of the identified priority areas, which were
23 considered candidate areas for expansion of the current PA network, was influenced by
24 land-use types according to the mask layers of non-PAs, and low-, middle-, and high-
25 ranked PAs. The current PA network reduced the aggregate extinction risk of multiple

26 species by 36.6%. Indeed, the percentage of built-up areas in the existing PAs was in
27 general smaller than that in the areas surrounding PAs. Notably, high-ranked PAs fully
28 restrained built-up pressure (0.037% per 10 years), whereas low-ranked PAs in the
29 national park and wild-life protection areas did not (1.845% per 10 years). Conservation
30 effects were predicted to substantially improve by expansion of high-ranked (legally
31 strict) PAs into remote non-PAs without population/socio-economic activities, or
32 expansion of medium-ranked PAs into agriculture forestry satoyama and urban areas. A
33 30% land conservation target was predicted to decrease extinction risk by 74.1% when
34 PA expansion was implemented across remote areas, satoyama, and urban areas;
35 moreover, PA connectivity almost doubled compared with the existing PA network. In
36 contrast, a conventional scenario showed that placing national parks in state-owned and
37 non-populated areas would reduce extinction risk by only 4.0%. The conservation
38 prioritization analyses demonstrated an effectiveness of using a comprehensive
39 conservation approach that reconciles land-sparing protection and land-sharing
40 conservation in other effective area-based conservation measures (OECM) in satoyama
41 and urban green spaces. Our results revealed that complementary inclusion of various
42 PAs interventions related to their governance and land-use planning plays a critical role
43 in effectively preventing biodiversity loss and makes it more feasible to achieve ambitious
44 conservation targets.

45

46 **1. Introduction**

47 Protected areas (PAs) are fundamental safeguards of biodiversity (Watson et al., 2014).
48 Therefore, area-based conservation measures have been debated for several decades
49 (Brundtland Commission, 1987; IUCN, 1994; Maxwell et al., 2020), but it remains
50 unclear how much area or coverage and what biodiversity should be included within PAs
51 (e.g., Chape et al., 2005). PA targets are recommended from two perspectives, biological
52 science and policy-making, based on conservation effects (needs) and socio-economic
53 costs (Svancara et al., 2005). Conservation scientists advocated that the 25–75% of areas
54 should be placed within PAs to preserve biodiversity on the planet (Noss et al., 2012);
55 alternatively, the Convention on Biological Diversity (CBD) set more feasible,
56 conservative goals across 1992 to 2010, calling for 4–17% protection, to reduce
57 biodiversity loss. Global Biodiversity Outlook 5 determined that the current Aichi Target
58 11's 17% protection of the terrestrial surface was partly achieved in many countries by
59 2020; this is the most successful of all the 20 Aichi targets (CBD Secretariat, 2020a).
60 Nevertheless, there is still rapid biodiversity loss: one million species are threatened with
61 extinction (IPBES, 2019), which demonstrates a need for renewed, bold conservation
62 targets to produce a biodiversity-positive outcome (Maron et al., 2018; Mace et al., 2018).
63 Therefore, the post-2020 Global Biodiversity Framework proposes a target of 30 % land
64 conservation (CBD Secretariat, 2020b).

65 Most importantly, area-based conservation targets should not include blindly placed
66 PAs and increased spatial percentage protection motivated by political considerations or
67 economic efficiency (Barnes et al., 2018; Magris and Pressey, 2018). Such targets should

68 fundamentally aim for spatially prioritized PA network planning; “quality, not quantity
69 (size), matters” (Pimm et al., 2018) means effectively capturing representativeness,
70 ecological connectivity, and areas of importance for biodiversity patterns that are
71 underpinned by ecological and evolutionary processes (Klein et al., 2009; Arponen, 2012;
72 Arponen and Zupan, 2016). Therefore, biologically oriented PA expansion is a delicate
73 mission in social equity when balancing conservation interventions and land-use policies
74 (Moilanen and Arponen, 2011). Increasing the number of strict PAs to focus on global
75 biodiversity, such as the ambitious suggestion of a “half-earth” initiative to protect 50%
76 of the surface and capture approximately 85% of species (Wilson, 2016), may be
77 infeasible in most countries with highly complicated land-use policies that involve
78 stakeholder interaction (Büscher et al., 2016). In fact, only one-fifth of the Key
79 Biodiversity Areas, which are defined as “sites of importance for the global persistence
80 of biodiversity” (IUCN, 2016), are fully covered by PAs (Butchart et al., 2015), which
81 indicates a compromised outcome behind conservation priorities.

82 “Other effective area-based conservation measures” (OECM) is an alternative
83 solution to bridge the gap between socio-economic feasibility and bold conservation
84 targets (Dudley et al., 2018; Bhola et al., 2020; Alves-Pinto et al., 2021). Notably, OECM
85 is defined by conservation outcomes through community-led or private efforts (Jonas et
86 al., 2014) when landscape management is shared with agricultural/forestry or indigenous
87 traditional land use, and has less strict protection or lacks legal restriction. This
88 conceptually may reflect the “third-of-third” conservation initiative by Hanski (2011): a
89 third of the land area is moderately managed as multi-objective conservation landscapes,

90 within which a third of the area is protected. For example, the rural landscape
91 management of satoyama and urban greening in Japan provides ancillary benefits to
92 biodiversity conservation through wildlife habitat diversification related to traditional
93 social–ecological productions (Washitani, 2001; Kobori and Primac, 2003; Kadoya and
94 Washitani, 2011). Therefore, a comprehensive approach that reconciles land-sparing and
95 -sharing conservation would be a practical solution to effectively achieve more ambitious
96 PA targets with the effectiveness (Maxwell et al., 2020). Moreover, in land-sharing
97 OECM, securing land exploitation prevention is key for fulfilling the post-2020 Global
98 Biodiversity Framework (Jonas et al., 2018; Donald et al., 2019); in fact, less strict PAs
99 in satoyama and urban areas are prone to exploitation under land-use pressures (Kim et
100 al., 2021).

101 Spatial conservation prioritization (SCP) supports the analytical need for PA targets
102 in the CBD (Moilanen et al., 2014; Kullberg and Moilanen, 2014). SCP is a well-
103 established method for identifying irreplaceable priority areas for long-term persistence
104 of potential biodiversity and essential ecosystem services, and is based on biological and
105 socio-economic viewpoints (Moilanen et al., 2009; Knight et al., 2011; Kukkala and
106 Moilanen, 2013). In the SCP analysis, PA network effectiveness for biodiversity
107 conservation is evaluated by area-dependent reduction of species extinction risks
108 (Lehtomäki et al., 2019; Hannah et al., 2020) and spatial priority ranking across the
109 existing PAs from intact landscape to managed (or private) landscape; this optimizes
110 competing land uses and is computed for each area associated with stakeholder groups
111 using their spatial layers (Moilanen et al., 2011; Honeck et al., 2020). Therefore, the SCP

112 output facilitates outcome-based assessments of existing PA networks and guides PAs
113 expansion by implementing top-down national-scale land-sparing and bottom-up private-
114 led land-sharing conservation. However, the SCP-based effectiveness assessments of the
115 PA networks have been rarely conducted in individual nations (e.g., Bicknell et al., 2017;
116 Di Minin et al., 2017).

117 In this study, we measured the conservation effectiveness of the Japanese PA network:
118 the evaluation involved the spatial design of the existing PAs, the effect of threat reduction,
119 and the ecological outcome of PA expansion in relative to the Aichi targets and post-2020
120 Global Biodiversity Framework (Maxwell et al., 2020). Japan is geographically located
121 in a global biodiversity hotspot: notably, strict PAs placed in state-owned land currently
122 represent only 3.8% of total land areas, and most PAs (e.g., the national park) include
123 residential, private lands, such as agriculture and forestry areas, and urban areas, which
124 allow land exploitation; these areas differ in legal strictness and policies from those of
125 western countries (Tanaka et al., 2019).

126 We first created a novel system of nation-wide biodiversity features (spatial
127 distributions of 18,126 species, which included all vascular plants and vertebrates) and
128 conservation cards with priority scores at the 1 km-grid resolution across the Japanese
129 terrestrial and marine ecosystems. In this study, we used data from 8,500 terrestrial
130 species. This information provides scientific evidence for incorporating national/regional
131 biodiversity conservation strategies into basic plans for the conservation and sustainable
132 use of biodiversity. Then, we investigated i) effects of exploitation pressure prevention
133 on the existing PAs, ii) reduction of relative extinction risks by the existing PAs, iii)

134 changes in conservation effectiveness and connectivity among the PA network by nation-
135 level conservation actions (including the Aichi targets) over the past 20 years, and
136 predicted iv) predicted improvements under the PA expansion scenarios related to the
137 post-2020 target. These assessments were conducted by accounting for the existing PAs,
138 which have different legal strictness regarding protection and different land-use types that
139 suffer from various exploitation pressures. Finally, we propose an approach for land-
140 sparing and -sharing conservation across landscape that includes different land-uses types,
141 and demonstrated the importance of SCP-based PA assessment in formulating
142 conservation targets and their measurability.

143

144 **2. Material and methods**

145 *2.1. Study site*

146 Japan mainly consists of the main islands of Japan, the Ryukyu Islands and the Bonin
147 Islands, which are located off the eastern coast of Asia. These regions range from
148 hemiboreal to subtropical climatic regions: the mean annual temperature ranges from –
149 5.3–24.2°C and the annual precipitation ranges from 650–4538 mm. This warm and wet
150 monsoon climate contributes to development of diverse biota in the Japanese archipelago.
151 Moreover, these islands are biogeographically located from the Holarctic to the
152 Palaeotropical regions. The insularity during the Pliocene and Pleistocene intermittently
153 divided one large regional biota into local communities on individual islands, with
154 periodic connections reforming through land-bridge corridors. Such complicated
155 geographical and climatic conditions shaped region-specific endemic biodiversity pattern

156 through dispersal limitation and climate-related energy availability (Kubota et al., 2014,
157 2015, 2017). Therefore, Japan, which harbors ecologically and/or evolutionarily
158 distinctive biota ranks among the top 35 global biodiversity hotspots (Mittermeier et al.,
159 2011). Indeed, recent human impacts have destroyed most lowland habitats and posed
160 threats to biodiversity hotspot conservations. Consequently, Japan serves as an ideal
161 region where conservation biogeographers can develop spatial prioritization measures for
162 the PA network for capturing ecological and evolutionary potential.

163

164 *2.2. Biodiversity features*

165 We created the Japan Biodiversity Mapping Project (J-BMP) database based on
166 species occurrence information (n = 13,366,641 for 8,500 vascular plants and vertebrate
167 species (Table S1). For decades, substantial knowledge of Japanese natural history has
168 been accumulating through research activities and environmental assessments
169 individually run by researchers, local governments, environmental assessment companies,
170 and citizen scientists. Species occurrence, functional traits, and phylogeny data have been
171 thoroughly compiled in J-BMP. Detailed descriptions of these data can be found on the J-
172 BMP website (<https://biodiversity-map.thinknature-japan.com/index.html>).

173 Species distribution models were developed using Maxent version 3.4.1 (Phillips et
174 al., 2006) with 52 environmental variables, which included climatic, soil, geological,
175 topographical, and geographical conditions, as the predictor variables (Table. S2); these
176 variables were previously proposed to be potentially important factors that explain
177 biodiversity distributions in this region (Kubota et al., 2015, 2017). We binarized (1/0)

178 the predicted suitability at the 1-km grid cell-level using the sensitivity–specificity sum
179 maximizer threshold (Jiménez-Valverde and Lobo, 2007), and confirmed the accuracy of
180 each model using the area under the receiver operating characteristic curve (Fig. S1).
181 Moreover, we corrected commission errors of binarized presence/absence predictions
182 with *ex post facto* design by validating the observed presence/absence data, such as
183 regional species checklists, northern/southern limits of individual species distributions,
184 and species compositions sampled at local communities (e.g., a number of vegetation
185 samples and monitoring sites). The predicted species distributions were stacked, and
186 species richness maps were visualized for individual taxa (e.g., plants, vertebrates, insects,
187 stony corals, crustaceans, and shellfish etc.) on the J-BMP web site.

188

189 *2.3.PA data*

190 We extracted the spatial data of all PAs from Digital National Land Information
191 (<https://nlftp.mlit.go.jp/ksj/>) and Natural Environmental Information GIS
192 (http://www.biodic.go.jp/trialSystem/top_en.html) (Table S3). We classified these PA
193 categories for biodiversity conservation into three ranks on the basis of strictness of legal
194 protection: high = economic activities are strictly forbidden; medium = public permission
195 is required for economic activities, and small-scale land exploitation and agriculture or
196 forestry in private lands are allowed; and low = public permission is required for
197 economic activities such as large-scale land exploitation and also agriculture or forestry
198 in private lands are allowed. Subsequently, we created binary (1/0) maps of PAs for each
199 rank at the 1-km grid cell level by defining PAs for which more than 50% protected of

200 100 100-m cells (100 ha) were protected. In total, approximately 20.3% of the land area
201 was designated as PAs (high = 3.8%, medium = 6.9% , and low = 9.6%).

202

203 2.4.Land-use data

204 We compiled the data on land-use types at 1-km grid cell resolution. We first
205 categorized individual cells into residential land-use types on the basis of the land-use
206 subdivision mesh of the National Land Information Division, National Spatial Planning
207 and Regional Policy Bureau, MLIT of Japan
208 (<https://nlftp.mlit.go.jp/ksj/gml/datalist/KsjTmplt-L03-b.html>) and human population
209 density (<https://nlftp.mlit.go.jp/ksj/gml/datalist/KsjTmplt-mesh1000h30.html>). These
210 residential land-use types included built-up areas; road and rail-way construction areas;
211 agricultural land types, such as orchards; or forestry/grassland types, such as coastal
212 vegetation and urban areas.

213 Urban areas were defined by dominance of residential land. Agriculture areas were
214 defined as farmland areas (e.g., rice fields) in lowland areas. Satoyama areas were defined
215 as areas for which more than 45% of the area included at least two of the following three
216 land types based on National Actual Vegetation map
217 (https://www.biodic.go.jp/kiso/vg/vg_kiso.html): i) agricultural land that contained
218 natural vegetation and for which naturalness was ranked as 2 or 3, ii) secondary grass
219 lands for which naturalness was ranked as 4 or 5, and/or iii) secondary forested lands with
220 a predominance of *Castanopsis* or *Quercus* species and naturalness ranked as 7 or 8. The
221 areas that were not populated, which was not applicable to urban, agricultural, and

222 satoyama areas, were defined as remote areas, which were mostly located in mountain
223 areas. Based on these land type classifications, we categorized all 1-km grid cells across
224 Japan into four types of urban areas (5.8% of total land area), agricultural areas (18.0%),
225 satoyama (45.0%), and non-populated remote mountain areas (31.2%).

226

227 *2.5. Land exploitation prevention assessment by PA*

228 In high-, medium-, and low-ranked PAs, we compared the composition of land-use
229 types with their surrounding areas; the surrounding areas within a 5-km radius of existing
230 PAs were defined as non-PAs. Moreover, to assess prevention effects of the PA network
231 on land exploitation, we investigated built-up areas from 2006 to 2016 for each PA rank
232 and category (e.g., natural parks, natural conservation areas, and wild-life conservation
233 areas). To calculate naturalness score, we re-categorized land types into residential,
234 agricultural, or forestry/grassland areas on the basis of land-use information, which
235 included built-up areas, road and rail-way construction areas, cultivated areas, and
236 orchards. We scored residential, agricultural, and forestry/grassland areas as 0, 1, and 2,
237 respectively, and then counted total naturalness scores at a 100-m scale for each PA. The
238 decrease of naturalness score from 2006 to 2016 was defined as built-up pressure: if
239 forestry/grassland areas were replaced with residential areas, the change of naturalness
240 score was calculated as a 100% decrease; if land-use type did not shift during this period,
241 it was calculated as 0%. However, because we only focused on prevention of built-up
242 pressure by PAs, we ignored increments of naturalness score (e.g., secondary succession
243 from agricultural areas to forest/grasslands). Additionally, we evaluated the spatial

244 distribution of solar power systems (their number and area at 1 km-grid resolution), which
245 has recently been considered typical land exploitation in satoyama and agricultural areas
246 with semi-natural environments (Kim et al., 2021). Spatial distribution of solar power
247 facilities was derived from information on certification of renewable energy power
248 generation business plans in October 2020. In this assessment, we focused on the amount
249 of solar power facilities that generate >20 kw per 1 km² in the existing high-, medium-,
250 and low-ranked PA and non-PAs.

251

252 *2.6. Spatial conservation priority analysis*

253 We used Zonation software, which produces a balanced, complementarity-based
254 ranking of conservation priority across the area of interest based on a set of input spatial
255 features (Moilanen et al., 2009; Moilanen et al., 2014; Kullberg and Moilanen, 2014;
256 Lehtomäki et al., 2019). Starting from the full landscape, Zonation iteratively ranks and
257 removes cells that lead to the smallest aggregate loss of conservation value while
258 accounting for total and remaining distributions of (potentially) weighted features and
259 other relevant factors such as connectivity, costs, or habitat condition. For this analysis,
260 we used 8,500 species distribution layers at the 1 km-grid resolution (377,589 cells),
261 which included J-BMP-derived data from 7,346 vascular plants, 77 amphibians, 103
262 reptiles, 607 birds, 136 mammals, and 231 freshwater fishes. Because these groups widely
263 cover foundation and umbrella taxa with diverse taxon-dependent biogeographic patterns
264 (Lehtomäki et al., 2019), we assumed that they well represent the biodiversity pattern in
265 Japan.

266 We conducted spatial prioritization using the continuous suitability predictions of
267 individual species within their presence cells. In this computation, we included all species
268 across all taxa to detect priority areas for expansion of the existing PAs using the additive
269 benefit function (ABF) methods; this allowed achievement of cost-efficient species
270 coverage (Moilanen et al., 2014). The ABF minimizes aggregated extinction risk and
271 implicitly places higher priority on species-rich cells because the algorithm sums
272 conservation values over all species in a cell (Moilanen, 2007). For individual taxa, we
273 weighted each species by a compound weight on the basis of the Japanese Red List
274 category (RLC; least concern = 1, near threatened = 2, vulnerable = 4, endangered = 6,
275 critically endangered = 8, data deficient = 2). To aggregate all species across different
276 taxa (j) with different number of species (N_j), we normalized the aggregate weight (AGG_j)
277 for each taxon with N_j to account for the rather different numbers of species that belonged
278 to different taxa. The weight w of species i in taxon j was then defined as follows: $w_{i,j} =$
279 $(RLC_i \times AGG_j)/N_j$.

280 To identify the most cost-effective complementary additions to the current PA
281 network, we used hierarchical mask analysis in Zonation to account for high-, medium,
282 low-ranked PAs, remote non-PAs, satoyama areas, and urban areas. This analysis forced
283 the highest priorities into the high-ranked PAs, followed by the medium and low-ranked
284 PAs, finally the non-PAs of the landscape (Moilanen et al., 2011; Lehtomäki et al., 2019).
285 Therefore, the hierarchical mask was used to implement gap analysis on the current PA
286 network and revealed cost-efficient gap filling for future PA expansion: the top-priority,
287 non-PAs are the most cost-effective complementary additions to the PA network

288 (Lehtomäki et al., 2009).

289

290 *2.7. Aggregated relative extinction risk*

291 Using the Zonation analysis output (relationship between land loss and species' range
292 loss), we evaluated the relative extinction risk across multiple taxa. In the spatial
293 prioritization of Zonation, the relative reduction (V) of distribution (R) for i -th species is
294 defined as $V_i = R_i^z$; in our analysis, the parameter z was set as 0.25. Therefore, species-
295 specific extinction risk (ER_i) relative to the proportion of PA coverage in total land areas
296 was calculated as $ER_i = 1 - R_i^z$. We calculated the aggregated ER over species of multi-
297 taxon as the sum of individual species' remaining extinction risk across all species relative
298 to the spatial design of the PA network (Hannah et al., 2020).

299

300 *2.8. Connectivity*

301 PA connectivity was evaluated by the ProtConn index (Saura et al., 2017).
302 ProtConn is defined as the percentage of a country that is covered by protected and
303 connected lands, and includes the four fractions that account for intra- and inter-PA
304 connectivity by moving between different PAs. ProtConn increases as area increases
305 among connected PAs or as distance decreases between the disconnected PAs, and the
306 maximum value of ProtConn corresponds to PA coverage (%) in total land areas. In
307 ProtConn computation, distance thresholds were set to 10 km to represent the potential
308 dispersal distance for species between PAs, and the probability of direct dispersal to other
309 separated PAs was defined as 50% (Saura et al., 2018).

310

311 *2.9. Improvement assessment: past and future PA expansion*

312 We retrospectively and prospectively evaluated the conservation effectiveness
313 of PA expansion retrospectively and prospectively. First, we delineated spatial
314 distributions of PAs every 10 years during 2000–2020, and evaluated the reduction of the
315 aggregated ER by the proportion of PA coverage and the connectivity (ProtConn) of the
316 PA network.

317 Second, we simulated PA expansion following a post-2020 target (i.e., increasing PA
318 up to 30%) based on spatial prioritization with the mask layer. To examine how the
319 existing PA network should be expanded under the hierarchical protection system (high-,
320 medium-, and low-ranked PAs) and heterogeneous land-use types, we built two scenarios.
321 In the OECM scenario, we selected candidate areas from high-ranked PAs in remote areas,
322 medium-ranked PAs in satoyama and agricultural areas, and low-ranked PAs in urban
323 areas; these were selected based on the saturating patterns of the relative extinction curves.
324 These expansion scenarios of new PAs across national lands in remote areas and private
325 lands in satoyama, agricultural, and urban areas are based on the concept of combining
326 land-sparing protection with a land-sharing approach through OECM. Alternatively, in
327 the conventional (business-as-usual) scenario, we selected candidate expansion areas
328 only from national land areas (national forests). In both scenarios, we evaluated their
329 conservation outcomes as the changes in the aggregated ER and ProtConn.

330

331 **3. Results**

332 *3.1. Built-up pressure prevention in PAs*

333 The compositions of land-use type across the inside to outside of existing PAs
334 corresponded to their legal strictness of protection (Table 1): high-ranked PAs were
335 mostly comprised of remote areas, and medium/low-ranked PAs and the areas
336 surrounding PAs included satoyama and agricultural areas. The percentage of built-up
337 areas in the existing PAs was generally smaller than that in the areas surrounding PAs
338 (Table 1): high-ranked PAs fully restrained built-up pressure, whereas low-ranked PAs in
339 the national park and wildlife protection areas did not, especially in comparison with
340 high- and medium-ranked PAs. In fact, the amount of solar power facilities in low-ranked
341 PAs and/or the areas surrounding PAs was greater than that in high/middle-ranked PAs
342 (Table 1).

343

344 **Table 1.** Percentage of land types on the inside and outside of the existing protected
345 areas (PAs), built-up pressure on the existing PAs, and the amount of solar power
346 facilities with >20kW per 1km² in the existing PAs. These were assessed for high-
347 ranked PAs (High), medium-ranked PAs (Medium), low-ranked PAs (Low), and
348 surrounding non- PAs (Surrounding) based on land-use changes between 2006 and
349 2016.

Legal Strictness on land use in protected areas (PA)	Remote area ratio (%)	Satoyama area ratio (%)	Agricultural area ratio (%)	Urban area ratio (%)	Built-up pressure (%)	Amount of solar power facility (kW/1km ²)
High-ranked PA	90.0	6.2	3.3	0.5	0.037	0.2

Medium-ranked PA	51.4	32.9	12.6	3.2	0.518	34.7
Low-ranked PA	35.6	41.3	17.7	5.3	1.845	296
Surrounding PAs	24.7	49.6	18.8	6.9	2.588	557

350

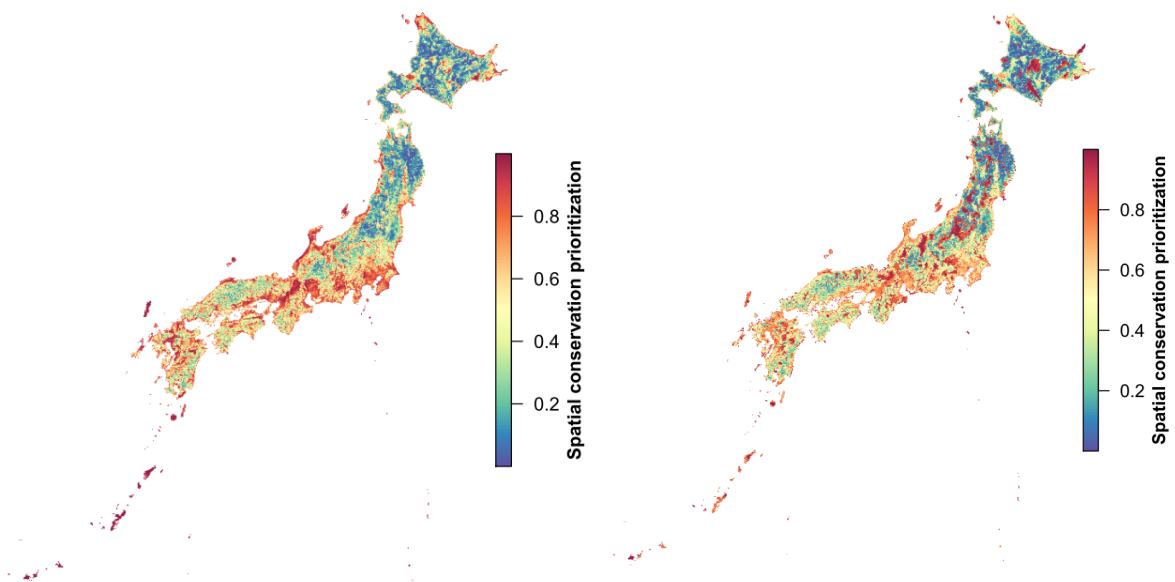
351 3.2. Extinction risk reduction in the existing PA network

352 Identified conservation priority areas of terrestrial biodiversity, which included
353 vascular plants and vertebrates were heterogeneous across Japan, especially in the middle
354 of Honshu Island and across the Ryukyu Islands and Bonin Islands (Fig. 1): these priority
355 areas were consistent regardless of mask layers.

356

357 1) without mask layer

2) with mask layer



358

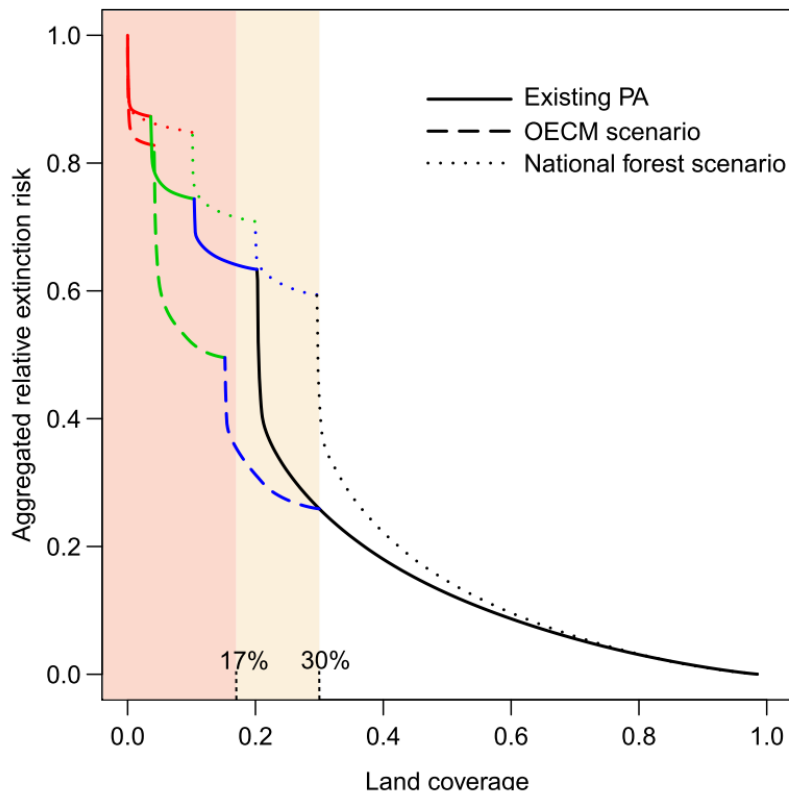
359 **Fig. 1.** Priority areas of biodiversity conservation in Japan. Spatial conservation
360 prioritization was based on biodiversity features across multi-taxon (8,500 species in
361 total) and the related analysis was conducted by using the Zonation software: 1-km grid
362 cell-level information was provided by the Japan Biodiversity Mapping Project (J-BMP;

363 Fig. 1). The mask layer analysis accounted for high-, medium-, low-ranked protected
364 areas (PAs), and remote non-PAs, satoyama areas, and urban areas. This analysis forces
365 the highest priorities into the high-ranked PAs, followed by the medium- and low-ranked
366 PAs, and finally the non-PAs of the landscape.

367

368 The existing PA network contributed to reducing the aggregate extinction risk
369 of multiple species by 36.6% (Fig. 2): high-ranked PAs (3.73% of total land area) reduced
370 extinction risk by 12.7%, and high- and medium-ranked PAs (10.9% of total land areas)
371 reduced extinction risk by 25.6%. The slope of the relative extinction curve was relatively
372 gentle within the current PAs but became steep in non-PAs. For the individual taxa
373 (mammal, bird, reptile, amphibian, freshwater fish, and vascular plant), the aggregated
374 relative extinction risk exponentially decreased with increasing land coverage across
375 priority areas (Fig. S2). For all taxa, extinction risk decay rates became substantially
376 smaller if more than 15% of land coverage was considered priority areas, and relative
377 extinction risk was expected to reduce 30–40% (Fig. S2).

378



379

380 **Fig. 2.** Conservation effectiveness of the current protected area (PA) network and 30%
381 land conservation scenarios in the post-2020 Global Biodiversity Framework target that
382 expands PAs into non-PAs. PAs effectiveness was evaluated by the reduction of
383 aggregated relative extinction risk of all vascular plants and vertebrate species distributed
384 in Japan. Red, green and blue curves showed conservation effectiveness of the high-,
385 medium-, and low-ranked PAs, respectively, as increasing PA coverages across mask
386 layers of land-use types. The solid line shows the conservation effectiveness of the current
387 protected area. The dashed line shows the expected conservation effectiveness of a
388 comprehensive approach that combines land-sparing protection and other effective area-
389 based conservation measures (OECM) in satoyama and urban areas: specifically, high-
390 ranked PAs expanded into remote state-owned areas (national forests), medium-ranked
391 PAs into satoyama and agricultural areas, and low-ranked PAs into urban areas. The
392 dotted line shows the expected conservation effectiveness of a conventional approach that
393 only expanded PAs into national land areas (national forests).

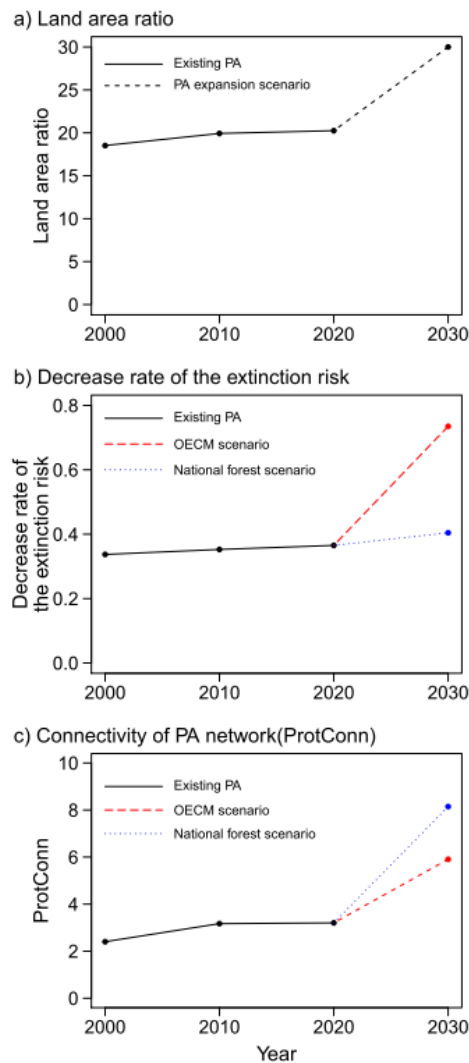
394

395

396 *3.3. Conservation effectiveness improvement: past and future*

397 The total PA area increased from 74,741 to 75,892 km² during 2010 to 2020.
398 The existing PA network in 2010 reduced 35.2% of the aggregate extinction risk. The PAs
399 expansion from 2010 to 2020, which was implemented to meet the Aichi targets, reduced
400 the aggregate extinction risk by 1.4% and improved the PA network connectivity (Fig. 3).
401 The spatial pattern of priority areas, which were assumed to be candidate areas to expand
402 the current PA network, was influenced by land-use types according to the mask layers of
403 non-PAs, and low-, middle-, and high-ranked PAs (Fig. 4). Under PA expansion up to
404 30% of total land area, the OECM scenario, which expands the PAs in priority areas
405 across remote, satoyama, and urban areas, predicted 74.1% reduction in total aggregate
406 extinction risk (Figs. 2 and 3b) and almost doubled connectivity with the current PA
407 network (Fig. 3). Specifically, expansion of high-ranked (legally strict) PAs in remote
408 non-PAs (0.57% of land area) without population/socio-economic activities, or expansion
409 of medium-ranked PAs in agriculture–forestry satoyama and expansion of low-ranked
410 PAs in urban areas (6.34% and 2.76% of land area, respectively) substantially improved
411 conservation effectiveness. In contrast, the conventional scenario, which solely assumed
412 PA expansion in remote, state-owned areas (e.g., national forests) reduced extinction risk
413 by only 4.0%.

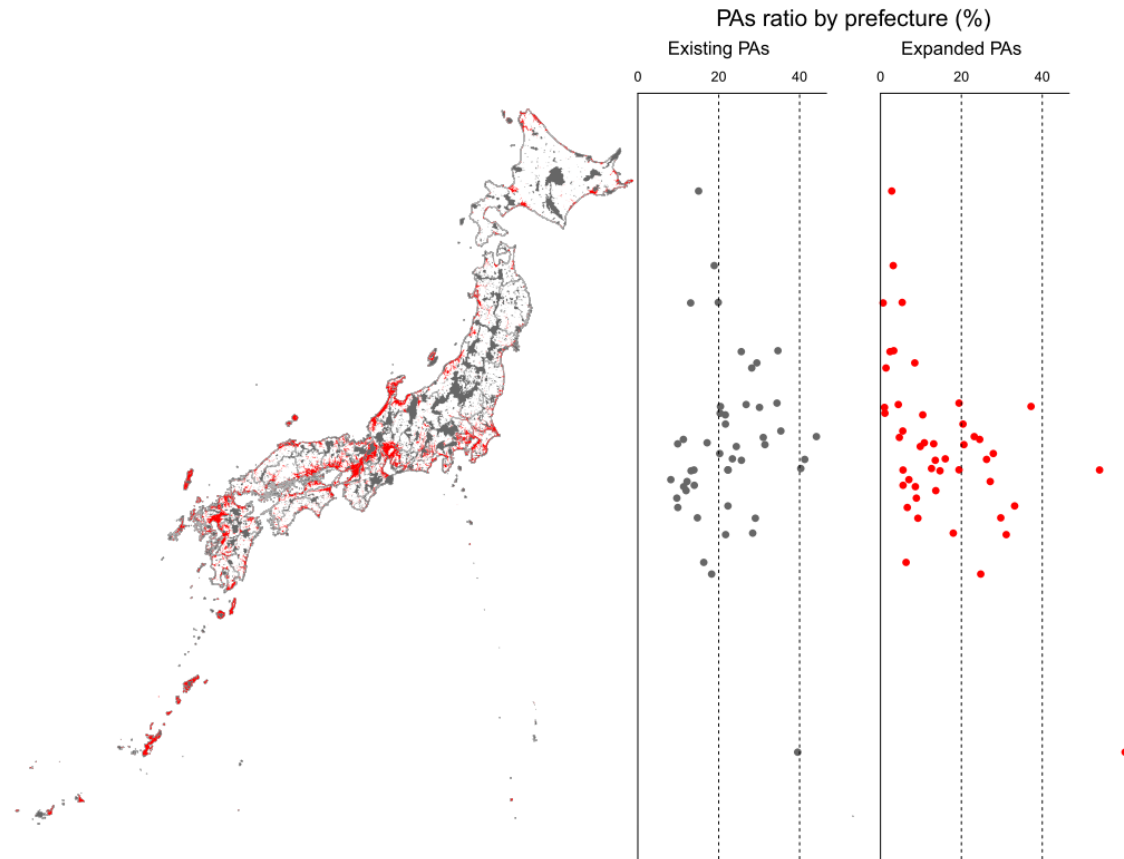
414



415

416 **Fig. 3.** Nation-level progress of the protected area (PA) network and conservation
417 effectiveness over the past 20 years. Conservation effectiveness was evaluated by land
418 area ratio (%) of PAs, reduction of aggregated extinction risk, and connectivity of PAs
419 every 10 years during 2000 to 2020. The conservation outcome of the post-2020 target
420 (30% PAs of total land area) was predicted based on the PA expansion in priority areas
421 across remote, satoyama, and urban areas that were identified in Fig. 2. The OECM
422 scenario represented a comprehensive approach that combines land-sparing protection
423 and land-sharing conservation in satoyama and urban areas, whereas national forest
424 scenario represented a conventional approach that only expands PAs into state-owned
425 forests.

426



427

428 **Fig. 4.** A protected area (PA) network that achieved the post-2020 Global Biodiversity
429 Framework's target of 30% PA coverage. Red and black areas show the expanded and
430 existing PAs, respectively.

431

432 **3. Discussion**

433 *4.1. Conservation priority areas reduce extinction risk*

434 Japanese nation-scale conservation priorities have been evaluated based on

435 biological irreplaceability and the human influence index to determine

436 threats/vulnerability (Kusumoto et al., 2017; Kubota et al., 2017; Lehtomäki et al., 2019).

437 Additionally, the importance of area-based measures have been argued by accounting for

438 land-use types, such as natural forests, managed forests, agricultural areas, and urban

439 areas, which are associated with the distribution of conservation benefits and cost among
440 local stakeholders to balance their social equity (Kubota et al., 2019). Therefore, the
441 present SCP-based analysis measured the outcomes/effectiveness of the PA network
442 based on the prevention of built-up pressure, the response curves of species, or aggregated
443 extinction risk when increasing land conservation according to higher priority scores
444 within different land-use types, which included current PAs, non-populated remote areas,
445 satoyama (agro-forestry), agricultural areas, and urban area. Importantly, extinction risk
446 reduction was gradual (Fig. 2), which indicated no clear thresholds such as a biologically
447 appropriate safe limit of PA size, although the protection of 40% habitat (e.g., forest lands)
448 to promote biodiversity persistence was empirically suggested from the relationship
449 between habitat loss and ecological responses (Swift and Hannon, 2010; Arroyo-
450 Rodríguez et al., 2020). Our SCP results provide scientific guidelines for nation-level
451 actions in the framework of international (or national) conservation targets, and revealed
452 that PAs should be maximized as much as possible to halt biodiversity loss.

453

454 *4.2. Post-2020 PA network balancing land-sparing and -sharing conservation*

455 The total coverage of all PAs in Japan has already achieved 20.3%, which is
456 more than Aichi Target 11. Nonetheless, the existing Japanese PA network partially failed
457 to prevent human impacts on priority areas: indeed, low-ranked less-strict PAs that
458 accounted for 47.3% of total PAs were inadequate for restraining built-up pressure (Table
459 1), such as solar power plant deployment (Kim et al., 2021), although the current PA
460 network reduced the aggregate extinction risk of multiple species by 36.6% (Fig. 2). This

461 ineffectiveness of the current PA network is because high-ranked PAs are mostly located
462 in areas with limited economic value (Kusumoto et al., 2017), as reported in
463 nation/global-scale analyses (Venter et al., 2014; Jenkins et al., 2015). Kadoya et al.
464 (2014) also showed that the current PA coverage is not sufficiently effective for halting
465 extinction rate of endangered plant populations. Therefore, the conservation effectiveness
466 of PA networks can never be achieved alone with a government-led top-down approach
467 that focuses on nationally, owned land areas. In fact, higher ranked priority areas were
468 substantially found privately managed lands used for agriculture or forestry and
469 around/within urban areas that resulted in leakage from current area-based conservation
470 planning (Ewers & Rodrigues, 2008).

471 Notably, the SCP outputs predicted that 30% land conservation according to
472 priority score should reduce extinction risk by 74.1% (Fig. 2). Specifically, the expansion
473 of high-ranked (strict) PAs in remote areas mostly without human populations and socio-
474 economic activities (Fig. 2), and expansion of medium-rank PAs in managed satoyama
475 and agricultural areas substantially improved conservation effectiveness by halting
476 biodiversity loss (Fig. 2). Alternatively, the current low-ranked PAs had few restrictions
477 on preventing land exploitation (Table 1). In fact, a number of priority areas was identified
478 in private lands including urban areas across low-ranked PAs and non-PAs, which were
479 distributed across Tokyo, Osaka, Kyoto, Aichi, and Fukuoka (Fig. 2), were seriously
480 exposed to built-up pressure (Table 1). The PA expansion strategy across remote mountain,
481 satoyama, agricultural, and urban areas contrasted with that of conventional conservation
482 based solely on land-sparing approaches (e.g., placing national parks in state-owned

483 areas); only expanding PAs solely in remote, non-populated areas reduced extinction risk
484 by only a few percentage points (Fig. 2). Most national forests distributed in remote
485 mountain areas are sufficiently captured by the existing PAs; thus, they are ineffective as
486 candidate areas for PA expansion.

487 Several previous studies reported that well-managed traditional land uses are
488 not necessarily in conflict with biodiversity conservation (e.g. Katoh et al., 2009; Jiao et
489 al., 2019). Those studies suggested that the network of rural land areas enlarges the size
490 of wildlife habitats and potentially plays a complementary role in area-based multi-use
491 conservation measures. Therefore, combining private sectors' actions in non-nationally
492 owned areas, such as sustainable managements of rural/peri-urban landscapes, with the
493 PA network through legal means is key to sufficiently protecting biological priority areas
494 and implement the Satoyama initiative
495 (http://www.env.go.jp/nature/satoyama/syuhourei/practices_en.html) (Kozar et al., 2019).
496 Our analysis demonstrated that satoyama, agricultural, and urban areas represent
497 substantial biodiversity, such as species in lowland waterside environments. Moreover, in
498 a 30% area-based conservation scenario, which balances land-sparing PAs in state-owned
499 land and OECM in satoyama, the PAs connectivity almost doubled compared with the
500 existing PA network (Fig. 4). Therefore, a comprehensive approach that combines land-
501 sparing protection in state-owned areas and land-sharing OECM in private land areas
502 contributes to managing leakage and promoting resilience of the current PA network.
503 Implementing various complementary PAs interventions related to their governance and
504 land-use planning plays a critical role in preventing species extinction and biodiversity

505 loss, and may makes it feasible to achieve ambitious conservation targets. Consequently,
506 a more flexible approach across hard/soft protection measures, which include OECM
507 (Jonas et al., 2014), community-based conservation (Brooks et al., 2012), and privately
508 PAs (Bingham et al., 2017), is required to enhance the PA effectiveness in the post-2020
509 strategic goals (Bingham et al., 2017; Bhola et al., 2020).

510

511 *4.3. Importance of nation-scale rigorous assessments on conservation effectiveness*

512 There is a critical need for scientific assessments of the PA effectiveness in
513 individual nations (Geldmann et al., 2013; Zafra-Calvo et al., 2017; Xu et al., 2021). All
514 practices for reversing biodiversity decline are locally implemented by experience-based
515 solutions and the responsibility lies within each political boundary (Arponen, 2012).
516 Consequently, translating international commitments into individual nations' actions is
517 problematic because global-scale protection of biodiversity mostly depends on the details
518 of how individual countries set spatial priorities for designing PA networks (Pimm et al.,
519 2018). In this regard, parochialism in conservation actions should also be reconciled with
520 transnational prioritization on the planet (Hunter and Hutchinson, 1994; Dudley, 1995;
521 Wells et al., 2010). Nation-specific prioritization measures are more feasible and
522 increases the local benefits of conservation, which is a positive aspect of parochialism,
523 by accounting for region-specific constraints, such as land-use planning. However,
524 individual nation-led collective PA expansion is relatively ineffective compared with
525 coordinated actions promoted by cosmopolitan SCP outputs (Montesino Pouzols et al.,
526 2014). Such an efficiency gap (negative aspects of parochialism) between globally

527 identified priority areas and nationally prioritized/designated PAs could be
528 collaboratively improved by SCP-based review mechanisms (Kullberg et al., 2019).
529 Moreover, pressures on PAs are expected to intensify because of climate change and
530 human impacts within PA boundaries or in their surrounding areas (de la Fuente et al.,
531 2020); notably, these are side effects of land-sparing conservation or the OECM approach.
532 Therefore, nation-specific effectiveness of land conservation is increasingly unclear or
533 limited (Geldmann et al., 2013; Coad et al., 2019). In the post-2020 Global Biodiversity
534 Framework, the performance measurement of PA networks should be conducted more
535 rigorously to assist nation-specific hard/soft measures on general land-use planning,
536 which directly supports decision-making processes of the PA expansion and management.

537

538 **Declaration of competing interest**

539 The authors declare no competing financial interests or personal relationships that could
540 have appeared to influence the work reported in this paper.

541

542 **Appendix. Supplementary materials**

543 Supplementary results to this article can be found online.

544

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551

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843 **Supplemental materials**

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845 **Table S1.** Amount of species occurrence data used to predict species distribution on the

846 J-BMP database.

847

Taxon	Number of species	Number of occurrences
Vascular plants	7,346	7,520,572
Mammals	136	2,049,450
Birds	607	3,011,635
Reptiles	103	71,627
Amphibians	77	112,109
Freshwater fish	231	601248
Total	8,500	13,366,641

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850 **Table S2.** List of environment variables used to infer species distribution.
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No.	Environmental value	source
1	Annual Mean Temperature	JMA's Mesh Climate Data 2000
2	Mean Diurnal Range (Mean of monthly (max temp - min temp))	
3	Isothermality (* 100)	
4	Temperature Seasonality (standard deviation *100)	
5	Max Temperature of Warmest Month	
6	Min Temperature of Coldest Month	
7	Temperature Annual Range	
8	Mean Temperature of Wettest Quarter	
9	Mean Temperature of Driest Quarter	
10	Mean Temperature of Warmest Quarter	
11	Mean Temperature of Coldest Quarter	
12	Annual Precipitation	
13	Precipitation of Wettest Month	
14	Precipitation of Driest Month	
15	Precipitation Seasonality (Coefficient of Variation)	
16	Precipitation of Wettest Quarter	
17	Precipitation of Driest Quarter	
18	Precipitation of Warmest Quarter	
19	Precipitation of Coldest Quarter	
20	Maximum snow depth	
21	Average annual solar radiation	
22	Total annual sunshine hours	
23	Natural forest area	Ministry of the Environment,
24	Secondary forest area	1/50,000 Vegetation survey
25	Planted forest area	(2nd - 5th) and 1/25,000
26	Natural grassland area	vegetation survey (6th and 7th)
27	Secondary grassland area	
28	Natural forest area within 5km	
29	Secondary forest area within 5 km	
30	Planted area within 5km	
31	Natural grassland area within 5 km	

32	Secondary grassland area within 5 km	
33	Agricultural land area	Japanese National Land
34	Paddy area	Numerical Information,
35	City area	Landuse Subdivision Mesh
36	Wasteland area	
37	Beach area	
38	Agricultural land area within 5km	
39	Paddy field area within 5km	
40	Urban area within 5km	
41	Wasteland area within 5km	
42	Beach area within 5km	
43	Land area	
44	Water area	
45	Elevation	Japanese National Land Numerical
46	Standard deviation of elevation	Information, Elevation and slope
47	Slope angle	angle tertiary mesh
48	Distance from the sea	Japanese National Land Numerical
		Information, coastline
49	Geology	1:500,000 Fundamental Land
		Classification Survey soil map
50	Cation exchange capacity of surface soil	SoilGrid
51	Amount of organic carbon in surface soil	
52	Surface soil pH	

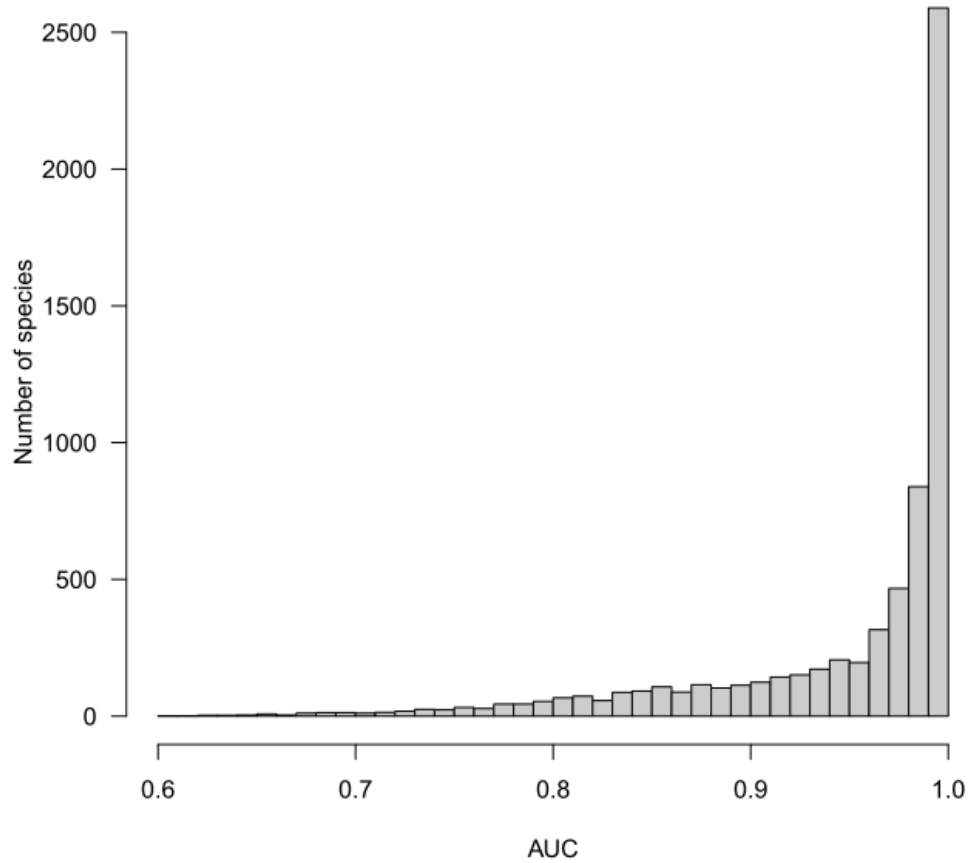
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854 **Table S3.** Types of Japanese protected areas and their legal strictness regarding land
855 use.

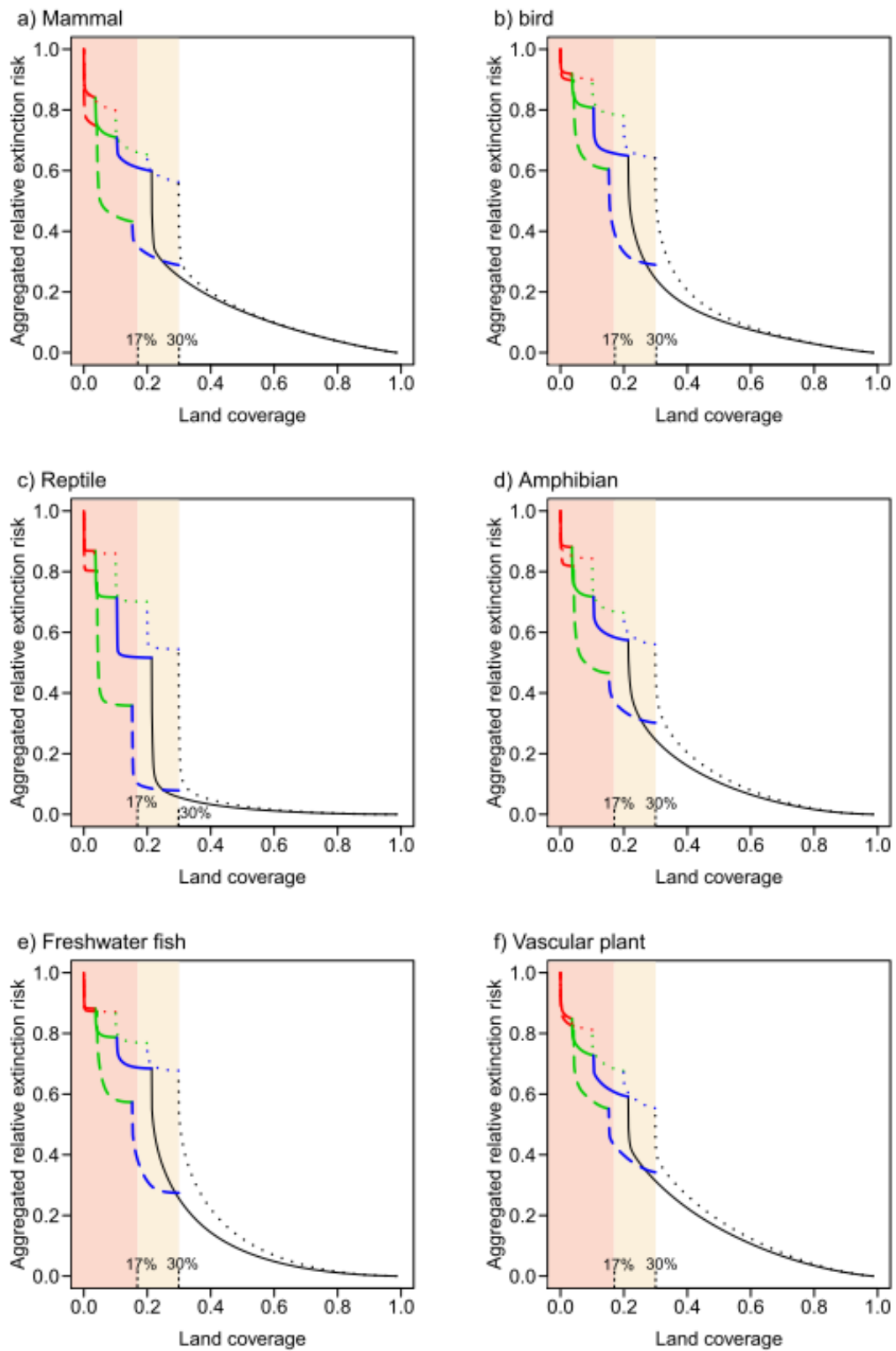
Basis laws	Types of protected areas	Zoning areas	Legal strictness
Natural Park Law	National park	Special Protection Area	Strict
		Classes I Special Zone	Strict
		Classes II Special Zone	Medium
		Classes III Special Zone	Medium
		Ordinary Area	Low
		Classification undecided	Low
	Quasi-national park	Special Protection Area	Strict
		Classes I Special Zone	Strict
		Classes II Special Zone	Medium
		Classes III Special Zone	Medium
	Prefectural natural park	Ordinary Area	Low
		Special Protection Area	Strict
		Classes I Special Zones	Strict
Classes II Special Zones		Medium	
Classes III Special Zones		Medium	
Natural Environment Conservation Law	Wilderness Areas		Strict
	Nature Conservation Areas	Special Area	Strict
		Ordinary Area	Low
	Prefectural Nature Conservation Areas	Special Area	Strict
		Ordinary Area	Low
Protection and Control of Wild Birds and Mammals and Hunting Management Law	National Wildlife Protection Areas	Special Protection Specified Area	Strict
		Special Protection Area	Medium
		Area except for Special Protection Area	Low
	Prefectural Wildlife Protection Areas	Special Protection Specified Area	Strict
		Special Protection Area	Medium
		Area except for Special Protection Area	Low
Law for the Conservation of Endangered Species of Wild Fauna and Flora	Natural Habitat Conservation Areas	Managed Area	Strict
		Monitoring Area	Low
Act Concerning Utilization of National Forest Land	Forest Reserve	Preservation Area	Strict
		Conservation and Utilization Zone	Medium
	Green corridor		Low

856 **Fig S1.** Species distribution prediction results for 8,500 species of all vascular plants
857 and vertebrates in Japan. Mean and median values of AUC were 0.949 and 0.983,
858 respectively.
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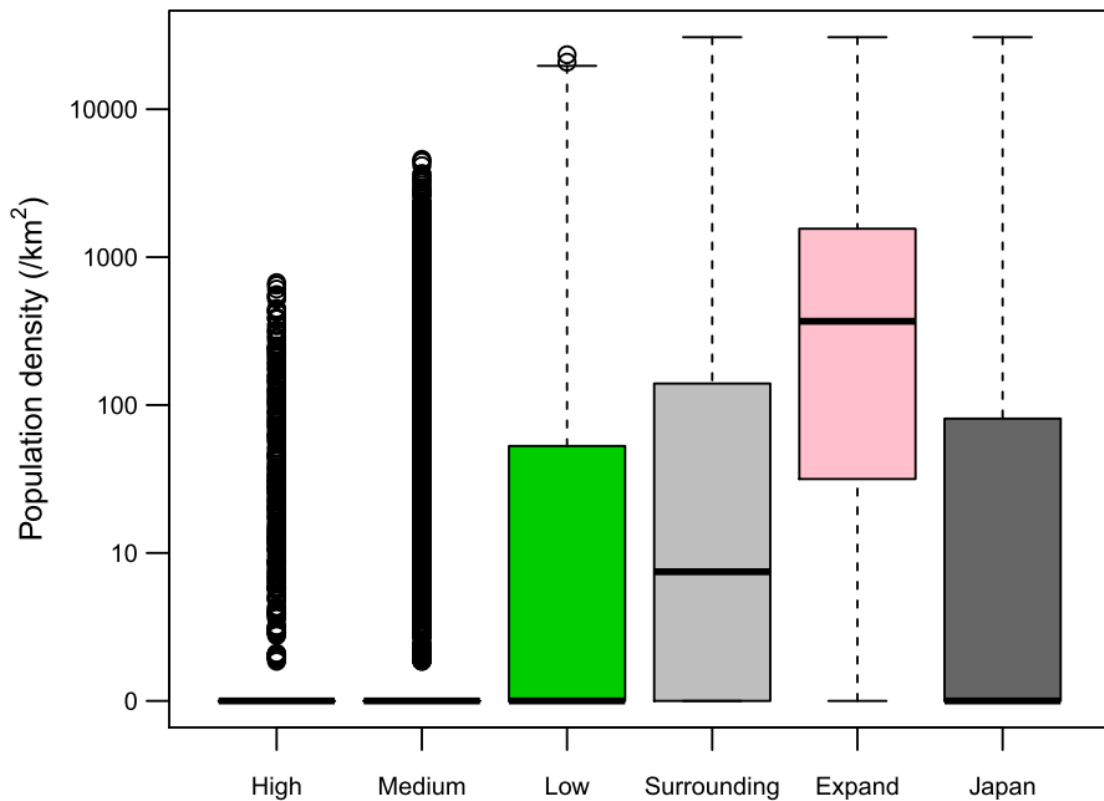
863 **Fig. S2.** Conservation effectiveness of the current protected area (PA) network and 30%
864 land conservation scenario in the post-2020 Global Biodiversity Framework target that
865 expands into non-PAs. PAs effectiveness was evaluated by the reduction of aggregated
866 relative extinction risk based on each taxon distributed in Japan. Red, green and blue
867 curves show conservation effectiveness of the high-, medium- and low-ranked PAs,
868 respectively, with increasing PA coverage across mask layers of land-use types. The solid
869 line shows the conservation effectiveness of the current PA network. The dashed line
870 shows the expected conservation effectiveness of a comprehensive approach that
871 combines land-sparing protection and other effective area-based conservation measures
872 (OECM) in satoyama and urban areas: specifically, high-ranked PAs expand into remote
873 state-owned areas (national forests), medium-ranked PAs expand into satoyama and
874 agricultural areas, and low-ranked PAs expand into urban areas. The dotted line shows
875 the expected conservation effectiveness of a conventional approach that only expands
876 PAs into national land areas (national forests).



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879 **Fig. S3.** Population density on the inside and outside of existing protected areas (PAs).
880 Population density (/km²) was evaluated in high-ranked PA (High), medium-ranked PA
881 (Medium), low-ranked PA (Low), surrounding non-PAs (Surrounding), expanded PAs
882 under the post-2020 Global Biodiversity Framework target (Expand), and all over Japan
883 based on land types (Japan).



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