

# Structure of bacterial communities in Japanese-style bathrooms: Comparative sequencing of bacteria in shower water and showerhead biofilms using a portable nanopore sequencer

So Fujiyoshi<sup>a,b,c,d\*</sup>, Yukiko Nishiuchi<sup>a</sup>, Fumito Maruyama<sup>a,b,c,d\*\*</sup>

*a. Office of Industry-Academia-Government and Community Collaboration, Hiroshima University, Hiroshima, Japan.*

*b. Center for Holobiome and Built Environment (CHOBE), Hiroshima University, Hiroshima, Japan.*

*c. Graduate School of Biomedical and Health Sciences, Hiroshima University, Hiroshima, Japan.*

*d. Network for Extreme Environment Research (NEXER), Scientific and Technological Bioresource Nucleus (BIOREN), Universidad de La Frontera, Temuco, Chile.*

**\*Corresponding author:** So Fujiyoshi, Office of Industry-Academia-Government and Community Collaboration, Hiroshima University, 1-3-2 Kagamiyama, Higashi-Hiroshima, Hiroshima, 739-8511, Japan. E-mail: [fujiyoshi.so.62w@kyoto-u.jp](mailto:fujiyoshi.so.62w@kyoto-u.jp), Tel: +81(82)424-7048, Orchid ID: 0000-0003-4306-332X.

**\*\*Co-corresponding author:** Fumito Maruyama, Office of Industry-Academia-Government and Community Collaboration, Hiroshima University, 1-3-2 Kagamiyama, Higashi-Hiroshima, Hiroshima, 739-8511, Japan. E-mail: [fumito@hiroshima-u.ac.jp](mailto:fumito@hiroshima-u.ac.jp), Tel: +81(82)424-7048, Orchid ID: 0000-0003-2347-616X.

## 25 **Abstract**

26 Showers are one of the main exposure routes to diverse microbes for end users in built  
27 environments. Bacteria in water are responsible for biofilm formation on surfaces, and the inside  
28 of a showerhead is a specific niche. Here, for the purpose of microbial characterization, source  
29 estimation and possibility of infection, the bacterial compositions of both shower water and  
30 showerhead biofilms in the same bathroom were determined and compared using a portable  
31 nanopore sequencer. The results suggest that specific bacteria in source water would primarily  
32 adhere to the surface of the showerhead where they subsequently form biofilms, and the  
33 community compositions within biofilms largely vary depending on environmental factors. The  
34 relative abundance of several pathogenic bacterial genera in both water and biofilm samples was  
35 low. We suggest that it is important to manage risk of infection in each household, and rapid on-  
36 site analysis of microbial communities will allow the realization.

## 37 **Keywords**

38 Built environment, Bathroom, Showerhead biofilm, Shower water, Nanopore sequencer,  
39 Comparative study, Bacterial community, Risk of infection

40

## 41 **Introduction**

42 We are surrounded by numerous microbes in our daily lives, and the bathroom is one of the  
43 largest reservoirs of microbes in the built environment due to moderately high temperatures and  
44 humidity (Neu *et al.* 2018). The typical Japanese bathing style is to immerse the body in the  
45 bathtub and wash the body outside the bathtub, and many Japanese people are known to bath for  
46 longer times and with higher frequency (approximately 30 min, every day) than people in other  
47 countries (Tochihara, 1999; Odamaki *et al.* 2019).

48 Shower water is generally supplied by a drinking-water distribution system, so microbes found in  
49 the water usually pose no risk for healthy individuals. To control microbial contamination in  
50 water, different chemicals and chemical treatment methods (ozone, fluorine, chlorine, chlorine  
51 dioxide, monochloramine, and copper-silver ionization and photocatalysis) and physical  
52 treatments (thermal inactivation, UV, filtration, and precipitation) are continuously applied  
53 (Cervero-Aragó *et al.* 2015; Huang *et al.* 2020). However, some of the microbes in the water can  
54 be opportunistic pathogens (Ops) capable of causing serious and life-threatening infections in  
55 severely immunocompromised individuals. Some Ops in water cause pneumonia (e.g., non-  
56 tuberculous mycobacteria (NTM), *Legionella* spp. and *Mycoplasma* spp.), asthma, or allergies  
57 (Dannemiller *et al.* 2016; Montagna *et al.* 2016; Nishiuchi *et al.* 2017). Bacteria in water are  
58 important in biofilm formation on surfaces, as they provide the initial cells for attachment and  
59 further biofilm development, and the inside of a showerhead is a specific niche that is moist,  
60 warm, dark, and frequently replenished with low-level nutrient resources. Bacteria are more  
61 concentrated in the biofilm than in the feed water; the number of colony forming units (CFUs) in  
62 water after chlorination is  $10^2$  CFU/mL, while in biofilms, the number of CFUs is between  $10^4$ –  
63  $10^5$  CFU/mL (Kormas *et al.* 2010; Peng *et al.* 2020; Novak *et al.* 2020). In addition, biofilms  
64 provide protection against environmental stressors such as antimicrobial agents and disinfectants  
65 (Johnson 2008).

66 Previous microbiological studies of built environment have used culture methodology to detect  
67 and identify microbes and have focused primarily on *Legionella pneumophila* and  
68 *Mycobacterium avium* subsp. *hominissuis* (MAH) (Arikawa *et al.* 2019; Falkinham 2020a).  
69 Nishiuchi *et al.* reported predominant colonization of *M. avium* in bathtub inlets of bathrooms of  
70 patients in Japan using culture isolation (Nishiuchi *et al.* 2007; Nishiuchi *et al.* 2009).  
71 Subsequently, Iwamoto *et al.* reported a high degree of genetic relatedness between bacterial  
72 isolates from pulmonary MAH patients and bacterial isolates from their bathrooms based on  
73 variable numbers of tandem repeats (VNTR) analysis of 19 loci (Iwamoto *et al.* 2012). These  
74 reports implied that bathrooms are potentially a major source of MAH in Japan. However, the  
75 main limitation of those studies was that it was unclear whether the transmission of the pathogen  
76 originated from the patient or the drinking water distribution systems. Thus, understanding the  
77 composition of the microbiome in both water and biofilms is of great importance for health risk  
78 assessment.

79 In 2012, a portable nanopore sequencer MinION (Oxford Nanopore Technologies, London,  
80 UK) was developed, which is now considered a breakthrough in DNA sequencing technology.  
81 The MinION sequencer enables real-time, on-site analyses of any genetic material (Schmidt *et al.*  
82 *et al.* 2016; Parker *et al.* 2017; Nakagawa *et al.* 2019). Compared to the currently widely accepted  
83 MiSeq sequencer (Illumina, San Diego, CA, USA), the MinION sequencer can generate a much  
84 longer read length, although with lower accuracy. The increased information content inherent  
85 from longer read lengths helps researchers with alignment-based taxonomy assignments.  
86 Nygaard *et al.* analysed the building dust microbiome using MinION and MiSeq, and showed  
87 that MinION had a better taxonomic resolution than MiSeq at the genus and species levels. By  
88 optimizing the PCR conditions, MinION was shown to provide accurate microbial community  
89 composition and more accurate data than MiSeq in terms of species-level matches (Fujiyoshi *et al.*  
90 *et al.* 2020). Thus, nanopore sequencing is an available and useful tool for understanding the status  
91 of public health based on microbial genetic information.

92 In this study, we used a high-throughput portable nanopore sequencer to determine the bacterial  
93 community in shower water and in showerhead biofilms.

94

## 95 **Results**

### 96 **Sampling and Sequence reads**

97 Fifty samples from 25 residential bathrooms (showerhead biofilm and shower water paired  
98 samples, hereafter called biofilm and water samples) were collected from the Hokkaido to Kinki  
99 regions in Japan (Figure 1A and 1B). As illustrated in Figure 2, microbes from biofilms were  
100 generally clumped and embedded in extracellular material, consistent with biofilm morphology.  
101 The DNA yields from the biofilm and water samples were highly variable, and detectable  
102 amounts of DNA could not always be extracted. After PCR, 7 biofilm or water samples were not  
103 amplified; therefore, 7 pairs of samples were removed for analysis of the pairs. After sequencing,  
104 2 pairs of samples were removed due to a low number of sequences reads. Finally, 32 samples in  
105 16 pairs were analysed in this study. Among the 643,780 raw reads from water samples and  
106 489,734 raw reads from biofilm samples, 343,092 and 228,863 reads, respectively, were  
107 analysed after filtration (Supplemental Table S1). The mean operational taxonomic unit (OTU)  
108 counts from the water and biofilm samples were 400 and 279, respectively. The number of OTUs  
109 was not significantly different between water and biofilm samples ( $p$ -value = 0.16 > 0.05).  
110 Good's coverage values were greater than 98% for all the samples (Supplemental Table S1).

111

### 112 **Comparison of the bacterial community structure in water and biofilm samples**

113 The taxon richness and evenness level reflected by the Shannon index was significantly higher in  
114 the water samples than in the biofilm samples ( $p < 0.01$ ) (Figure 3 and Supplemental Table S1).  
115 In addition, the Chao1 index, which accounts for species richness, showed that phylogenetic  
116 diversity was significantly higher in water samples than in biofilm samples ( $p < 0.01$ )  
117 (Supplemental Table S1). The nMDS plots demonstrated that the bacterial communities in the  
118 water samples were more closely clustered than those in the biofilm samples (Figure 4).  
119 PERMANOVA results confirmed that the overall community structures were significantly  
120 different in both the water and biofilm samples ( $p < 0.01$ ; see Figure 4 for F and R-squared  
121 values). Venn diagrams illustrate the percentage of genera shared among each pair of samples. In  
122 the biofilm samples, an average of 66.1% of the genera were common to water samples, whereas  
123 in the water samples, an average of 36.1% of genera were common to biofilm samples (Figure  
124 5). The percentage of common genera in biofilms was significantly higher than that in water ( $p <$   
125  $0.01$ ), and the percentage of shared sequences was relatively higher in biofilms (84.1%) than in  
126 water (72.1%) ( $p > 0.05$ ).

127

## 128 **Bacterial community composition**

129 The top three genera in the water samples were *Sphingomonas* (18.8%), *Methylobacterium*  
130 (17.0%), and *Phreatobacter* (12.6%), which belong to Proteobacteria (alpha-subclass) (Figure 6),  
131 while three genera, *Sphingomonas*, *Methylobacterium*, and *Bradyrhizobium*, were found in all  
132 the water samples. The top three genera in the biofilm samples were *Methylobacterium* (25.2%)  
133 and *Sphingomonas* (22.4%), belonging to the phylum Proteobacteria (alpha-subclass), and  
134 *Brevibacterium* (8.2%), belonging to the phylum Actinobacteria (Figure 6, for more detail  
135 information see Supplemental Figure S1). Only *Methylobacterium* was commonly found in all  
136 biofilm samples. To investigate the differences between shower water and biofilm samples,  
137 Linear discriminant analysis (LDA) effect size (LEfSe) was performed, and the bacterial taxa  
138 significantly associated with each sample was identified. A non-parametric factorial Kruskal-  
139 Wallis sum-rank test embedded in LEfSe identified 15 bacterial genera that were characteristic  
140 only for water but not for biofilm samples (Figure 7).

141

## 142 Risk of infection in bathroom environments based on the presence of pathogenic bacteria

143 Based on the BSL2 and BSL3 list from National Institute of Infectious Diseases (NIID), a total  
144 of 13 genera were identified in the water and biofilm samples (Figure 8). The genus present at  
145 the highest frequency was *Moraxella* with a prevalence of 81% (26/32) and a median value of  
146 0.40%, and the genus with the lowest frequency was *Nocardia* with a prevalence of 3% (1/32)  
147 and a median of 0.00%. Only *Moraxella* and *Staphylococcus* had mean values above 1%, and  
148 both were detected at high rates in some residences. The median value was 0.00% for most  
149 genera. *Mycobacterium*, which includes Ops that cause pulmonary disease, was detected in 34%  
150 (11/32) of the samples, with a median of 0.00%. *Legionella* was not detected in this study. Even  
151 samples taken from the same bathroom had an Op genus present in the biofilm but not in the  
152 water, or vice versa. There was no significant difference in the frequency, mean or median of  
153 genera that include pathogens in the biofilm and water samples (p-value > 0.05). At the species  
154 level, 5 and 7 BSL2 or BSL3 species were found in biofilm and water samples, respectively.  
155 These species accounted for less than 0.30% of the total community (Supplemental Table S2). In  
156 addition, these species had a lower percent identity (80.4% average value of identity) when  
157 analysed against a more refined and updated BLASTn database (Supplemental Table S3). The  
158 species that appeared in both analyses were *Staphylococcus aureus* (87.98% maximum value of  
159 identity) and *Vibrio parahaemolyticus* (87.86% maximum value of identity). The genera that  
160 appeared in both analyses were *Corynebacterium* (86.26% maximum value of identity) and  
161 *Mycobacterium* (86.87% maximum value of identity). For the four species of *Brucella* found in  
162 this study, most of them were assigned to different genera with low percent identity (76.34%  
163 average value of identity).

164

## 165 Discussion

### 166 Biofilm bacteria are derived from the water

167 The biofilm formed on the inner surface of the shower head interacts with the shower water  
168 (Rasmus *et al.* 2002; Peng *et al.* 2020). The results of our Venn diagram also show that some  
169 genera are shared between biofilm and water samples (Figure 5). However, the bacterial  
170 community structures of biofilm and water samples were significantly different, even when the  
171 samples obtained from the same bathroom (Figures 3 and 4). The bacterial community in  
172 biofilms had a lower alpha-diversity than the community in the water samples, suggesting that  
173 some, but not all of the bacteria in the water probably adhere to the showerhead surface and  
174 subsequently form a biofilm. Beta diversity analysis revealed significant differences between  
175 water and biofilm samples (Figure 4). The bacterial communities in the water were more closely  
176 clustered than those in the biofilm samples. The quality of drinking water in Japan is  
177 standardized by the Ministry of Health, Labour and Welfare, based on a total of 51 items, such  
178 as general bacterial CFU, chlorine and cadmium concentration, and each standard has a  
179 maximum value below which health is not affected (Nishiuchi *et al.* 2017; Novak *et al.* 2020).  
180 All the water samples collected in this study were supplied from drinking water distribution  
181 systems (not well or river water), and the well-controlled water quality in Japan may affect the  
182 similarity of the bacterial communities in the water samples. However, the bacterial communities  
183 in the biofilms differed greatly among samples. This probably suggests that the bacterial  
184 community in a biofilm is influenced by factors other than the water, such as the building type,  
185 the method of use, and the residents (Ji *et al.* 2017; Proctor *et al.* 2018). In this study, there was  
186 no correlation between the material of the showerhead and the bacterial community structure  
187 (data not shown).



188 The bacterial communities examined in this study are consistent with other studies showing the  
189 dominance of similar genera in freshwater systems and biofilms in bathrooms (Figure 6) (Feazel  
190 *et al.* 2009). *Methylobacterium* spp. are famous for forming pink-coloured biofilms and are  
191 prevalent in domestic water-associated environments, such as drinking water distribution  
192 systems, shower curtains, and showerheads worldwide (Gallego *et al.* 2005; Zhou *et al.* 2021). In  
193 Japan, *Methylobacterium* are also common bacteria with known culture methods (Yano *et al.*  
194 2013; Kawai *et al.* 2019). *Sphingomonas*, generally the most abundant genus detected in the  
195 biofilm samples, are ubiquitous in the environment, such as in soil, water, and sediments, as well  
196 as on shower curtains and are known to be constantly and consistently present in biofilms even  
197 after chlorination (White *et al.* 1996; Kelley *et al.* 2004; Douterelo *et al.*, 2018). *Phreatobacter*  
198 has been identified as a dominant genus in drinking water distribution systems in China (Li *et al.*  
199 2020; Ma *et al.* 2020; Jing *et al.*, 2021). *Brevibacterium* from biofilms are not as commonly  
200 detected in bathrooms as the other three genera. *Brevibacterium* species are isolated from various  
201 habitats, such as fermented food, animal and human skin, insects, soil, and mural paintings.  
202 Some species, such as *B. casei* and *B. epidermidis*, encountered in human skin, wounds, and  
203 blood can cause rare human infections, making them Ops (Denis and Irlinger, 2008). It should be  
204 investigated in the future whether the high percentage of *Brevibacterium* in showerhead biofilms  
205 is unique to Japan or due to environmental selective pressure. *Sphingomonas*, *Methylobacterium*,  
206 and *Bradyrhizobium* were found in all the water samples, while only *Methylobacterium* was  
207 found in all the biofilm samples. This shows that the bacteria in the water samples were  
208 relatively consistent, but the biofilm constituent bacteria varied among samples. LEfSe analysis  
209 identified significant differences in taxa between biofilm and water samples. A comparison of  
210 sample sources indicated that there were 15 taxa only from the water samples with LDA scores  
211 greater than 4.0 (Figure 7). No taxa identified were considered indicators of biofilms. This  
212 suggests that the bacteria in the biofilm samples are of water origin. A total of 66.1% and 36.1%  
213 of genera were common between the biofilm and water samples, respectively (Figure 5). These  
214 results suggest that specific bacteria in the original water would primarily adhere to the surface  
215 of the showerhead and subsequently form biofilms.

216 **Potentially pathogenic bacteria were present in shower water and biofilm samples, but only**  
217 **a low percentage of them were present in Japanese bathrooms**

218 The Ops found in the bathroom samples are well reported (Kelley *et al.* 2013; Novak *et al.*  
219 2020; Falkinham. 2020b). *Acinetobacter*, *Stenotrophomonas*, *Pseudomonas*, *Mycobacterium* and  
220 *Legionella* are commonly presented Ops worldwide. In this study, we encountered two pathogen-  
221 like species, *Staphylococcus aureus* (87.98% maximum value of identity) and *Vibrio*  
222 *parahaemolyticus* (87.86% maximum value of identity). *S. aureus* is known as a commensal  
223 bacterium and as a pathogen that causes opportunistic infections mainly of soft tissues, skin, and  
224 wounds (Kozajda *et al.* 2019). Studies have shown that the detection rate is higher in areas with  
225 high hand and foot contact (Ojima *et al.* 2002). Infection with *V. parahaemolyticus* can cause  
226 gastroenteritis, septicaemia, and infection. It is a seafood-associated and a water-borne food-  
227 mediated bacterium (Liu *et al.* 2015). This bacterium has been detected in freshwater in some  
228 cases (Maje *et al.* 2020; Silva *et al.* 2018). In addition to the presence of these bacteria, we also  
229 detected some Op genera that are closely related to *Corynebacterium* and *Mycobacterium*.  
230 *Corynebacterium*, a known human-associated bacterium, is commonly found in bathrooms,  
231 according to meta-analysis data from high-throughput amplicon analysis as well as from culture  
232 methods (Allen *et al.* 2004; Adams *et al.* 2015). In this study, *Legionella*, a problem in water  
233 systems around the world, was not detected. Some studies show that planktonic and biofilm  
234 *Legionella* concentrations are reduced by the presence of residual disinfectants within water  
235 material (Waak *et al.* 2018; Fish *et al.* 2020). The standard value of residual chlorine in water in  
236 Japan is set at 1 mg/L or less, is often 0.1 mg/L or less and is strictly controlled. The fact that  
237 *Legionella* was not detected in this study may indicate the high-performance level of the water  
238 quality management system in Japan. *Mycobacterium* was detected in 34% (11/32) of the  
239 samples, and the mean and median values were 0.40% and 0.00%, respectively (Figure 8).  
240 Arikawa *et al.* surveyed *Mycobacterium avium* subsp. *hominissuis* (MAH) in Japan, and the  
241 detection rate of MAH was 16.1% (Arikawa *et al.* 2019). Using a culture-independent method,  
242 quantitative PCR, Ichijo, *et al.* detected the frequency of *Mycobacterium* spp. in showerheads.  
243 The detection rate was lower than that in our study (13%) (Ichijo *et al.* 2014). According to  
244 Gebert *et al.*, the detection rate of *Mycobacterium* spp. is 37% across households in the United  
245 States and Europe, which is similar to the results of this study. However, the relative abundances  
246 of mycobacteria were much higher than those in this study, at approximately 15% and 7%,  
247 respectively (Gebert *et al.* 2018). Additionally, Feazel *et al.* reported that *M. avium* was  
248 identified in 20% of showerhead swabs in the US (Feazel *et al.* 2009). *Mycobacterium* was not

249 present at a higher rate in Japan than in the US and Europe; however, the incidence rate of  
250 nontuberculous mycobacterial pulmonary disease in Japan was reported to be 14.7 cases per  
251 100,000 people in 2014 (Namkoong *et al.* 2016), and the number of infected people is increasing  
252 rapidly compared to other countries (Yano *et al.* 2017). Non-tuberculous mycobacteria include  
253 nearly 200 species that can differ with respect to their ecology and pathogenicity (Tortoli 2014;  
254 Falkinham 2020a). Thus, to obtain more detailed information on specific mycobacterial species,  
255 it is necessary to sequence the gene using mycobacterium-specific primers, such as *hsp65* (65-  
256 kDa heat shock protein), as well as performing identification through culture methods (Telenti *et*  
257 *al.* 1993). In this study, we encountered the assignation of *Brucella* to other genera.  
258 Misidentification of *Brucella* species has been reported in many studies (Elsaghir and James  
259 2003; Horvat *et al.* 2011; Carrington *et al.* 2012), except for in those using near-full-length  
260 1,412-bp nucleotide sequences of 16S RNA genes and Sanger sequencing (Gee *et al.* 2004).  
261 Because of these recurring misidentifications, an increasing number of laboratories are now  
262 relying on molecular methods such as Gram staining, semisolid-medium motility tests and  
263 flagellar staining to identify *Brucella* (Yang *et al.* 2013). In the results of the sequence analysis  
264 in this study, even if *Brucella* was detected, it was necessary to validate the results by molecular  
265 methods using cultured strains.

266 One of the major advantages of using targeted metagenomic techniques, such as 16S rRNA gene  
267 sequencing, is that these techniques are culture-independent and can theoretically recover almost  
268 all bacterial taxa in any habitat. Therefore, it is possible to identify the microorganisms,  
269 including the percentage of potential pathogens, in the entire community. Moreover, MinION  
270 can generate long read length. The increased information content inherent from longer read  
271 lengths assists researchers with alignment-based taxonomy assignment (Wommack *et al.* 2008).  
272 Mitsuhashi *et al.* and Nakagawa *et al.* reported that 5-min and 3-min running times on MinION,  
273 respectively, were sufficient to detect specific bacteria (Mitsuhashi *et al.*, 2017; Nakagawa *et al.*,  
274 2019). Furthermore, by optimizing the PCR conditions, MinION was shown to provide accurate  
275 microbial community composition and more accurate data than Miseq in terms of species-level  
276 matches (Fujiyoshi *et al.* 2020). However, because of the widespread use of DNA sequencing  
277 technology, it is necessary to carefully investigate the analysis methods, databases, and results of  
278 these methods before applying them. The tendency is to use top hits to classify microorganisms,  
279 but in general, cut-offs of 95% and 98.7% are used to classify bacterial isolates at the genus and  
280 species levels, respectively (Stackebrandt and Goebel 1994). This point needs to be discussed  
281 carefully, especially when mentioning Op bacteria in samples. Additionally, DNA does not  
282 provide information whether a microorganism is alive, dead, or infectious. Therefore, it is  
283 necessary and efficient to conduct epidemiological studies using both metagenomic approaches  
284 and culture methods simultaneously. In 2020, we published a paper on a "suitcase lab" in which  
285 a single suitcase can contain all the necessary equipment from sampling to detection by the  
286 LAMP method and showed that the work can be completed on-site within two hours if the  
287 specific target species is decided upon (Fujiyoshi *et al.* 2021). In the future, if such a tool can be  
288 used to analyse the microbiome on-site quickly and easily, it will contribute to the improvement  
289 of microbial risk assessment and control not only of water quality but also of the built  
290 environment.

291 Our comprehensive analysis of the bacterial community in the built environment revealed that  
292 specific bacteria in source water would primarily adhere to the surface of the showerhead where  
293 they subsequently form biofilms. Moreover, the bacterial communities in biofilms differed  
294 greatly among samples. This probably suggests that the bacterial community in a biofilm is  
295 influenced by factors other than the water. The findings of this study are the first data in Japan  
296 for assessing microbial communities in built environments. This is also the first study to analyse  
297 the microbial communities of both water and its biofilm obtained from same bathrooms, by  
298 which the microbial similarity and difference of water and biofilm were emphasized clearly. We  
299 suggest that it is important to manage risks of infection in each household and that this requires  
300 rapid and easy on-site identification of microbial communities.

301

## 302 **Methods**

### 303 **Sampling**

304 A total of 50 showerhead biofilms and shower water samples from 25 independent bathrooms  
305 were collected from five regions in Japan: 3 from Tohoku, 2 from Chubu, 7 from Kanto and 12  
306 from Kinki, as shown in Figure 1B and Supplemental Table S1 (the sampling sites were  
307 classified according to Arikawa *et al.* 2019). The showerhead biofilm and shower water samples  
308 in the same bathroom were aseptically collected. Showerhead biofilms were swabbed with  
309 sterilized nylon swabs (FLOQ swab 552C; Becton, Dickinson and Company, Tokyo, Japan), and  
310 2 L of shower water was filtered on-site using a portable peristaltic pump (Sentino microbiology  
311 pump, Pall Life Science, MI) and a 0.2- $\mu$ m filter cartridge (Sterivex, Millipore, MA, USA)  
312 (Figure 1). The cartridge was put into a sterile tube, and then the samples were immediately  
313 placed in a cool box at 4°C, transported to the lab within a few hours and kept at -20°C until use.

### 314 **Scanning electron microscope (SEM)**

315 The biofilms on coverslips and filters were prefixed in a solution of 2.5% glutaraldehyde in 0.1  
316 M phosphate buffer (PB; pH 7.4) for 10 min and rinsed three times with PB. Samples were then  
317 fixed again with 2.5% glutaraldehyde for 1 h and rinsed three times with PB. Another fixation  
318 reagent, 1% (w/v) osmium tetroxide in PB, was added to the samples, followed by incubation for  
319 1 h. Subsequently, the samples were rinsed three times with PB and dehydrated with increasing  
320 concentrations of ethanol (30%, 50%, 70%, 90%, 99%, and 100%). Dehydrated samples were  
321 soaked in isoamyl acetate, successively critical-point-dried with an HCP-2 (Hitachi Ltd., Tokyo,  
322 Japan), and coated with 8:2 platinum–palladium alloy using an E-1030 ion sputter (Hitachi Ltd.,  
323 Tokyo, Japan). The resultant coating was 12 nm thick. Samples were observed using an S4700  
324 scanning electron microscope (Hitachi Ltd.).

325

326 **DNA extraction and PCR conditions**

327 All samples were recovered using aseptic techniques and appropriate negative and positive  
328 controls. A swab and a filter, picked out from the cartridge, were directly placed in a bead tube  
329 of a DNeasy PowerBiofilm Kit (QIAGEN, Germantown, MD, USA) under a laminar flow  
330 cabinet, and DNA was extracted according to the manufacturer's protocol with some  
331 modifications (Arai *et al.* 2018): instead of glass beads in a PowerBiofilm bead tube, 400  $\mu$ L of  
332 sterilized  $\phi$ 0.5 mm zirconia beads (TORAY, Tokyo, Japan) and two gains of  $\phi$ 5 mm zirconia  
333 beads (TORAY) were used for homogenization. The samples were bead beaten with a bead  
334 crusher (TissueLyser II, QIAGEN) at 3,200 rpm for 10 min. The DNA was eluted in 100  $\mu$ L of  
335 elution buffer and then purified and condensed with a Dr. GenTLE precipitation carrier (Takara  
336 BIO, Tokyo, Japan). The concentration and purity of the DNA were measured with a DS-11FX+  
337 Spectro/Fluorometer (DeNovix, Wilmington, USA) and a QuantiFluor<sup>TM</sup> dsDNA System  
338 (Promega, Madison, USA). PCR amplification and barcoding of 16S rRNA genes were  
339 conducted using the 16S Barcoding Kit (SQK-RAB204; Oxford Nanopore Technologies,  
340 Oxford, UK) containing the 27F/1492R primer set and MightyAmp DNA polymerase Ver.3  
341 (Takara Bio). PCR was performed according to a previous report (Fujiyoshi *et al.* 2020). A  
342 reaction containing no template served as the negative control. The amplified fragments were  
343 separated in a 2% agarose gel, stained with Safelook Load Green (Wako Chemicals Co. Ltd,  
344 Osaka, Japan), and checked with a FAS Nano Gel Document System (NIPPON Genetics, Tokyo,  
345 Japan).

346

### 347 **Nanopore sequencing library construction**

348 After purifying a PCR product (50  $\mu$ l) with 30  $\mu$ l of Agencourt AMPure XP (Bechman Coulter,  
349 Tokyo, Japan), the amount and purity of the DNA eluted with 10  $\mu$ l of buffer solution (pH 8.0,  
350 10 mM Tris-HCl with 50 mM NaCl) were determined as above. One hundred or 50 fmol of  
351 purified amplicon DNA was used as input DNA for MinION-compatible libraries. The  
352 amplicons were added to 1  $\mu$ l of rapid adapter (Oxford Nanopore Technologies) and incubated at  
353 room temperature for 5 min.

354

### 355 **Sequencing and data analysis**

356 Each nanopore sequencing library was run on FLO-MIN106 R9.4 flow cells (Oxford Nanopore  
357 Technologies) after performing platform QC analysis. The amplicon library (11 µl) was diluted  
358 with running buffer (35 µl), with 3.5 µl of nuclease-free water, and with 25.5 µl of loading beads.  
359 A 48-h sequencing protocol was initiated using MinION control software MinKNOW v.1.11.5 or  
360 1.14.1. MinION sequence reads (i.e., FAST5 data) were converted into FASTQ files using  
361 Albacore v.2.1.3 or 2.3.3 software (Oxford Nanopore Technologies). FASTQ files were analysed  
362 as described previously (Fujiyoshi *et al.* 2020). The files examined the sequence read length  
363 distribution using FastQC (v 0.11. 2) (Andrews. 2010), and Seqkit 0.8.0  
364 (<https://bioinf.shenwei.me/seqkit/>) (Shen *et al.* 2016) was used to filter the sequence data by the  
365 lengths of 1,400–1,600 to include 1,500–base reads. After filtering, the sequence reads were  
366 mapped using bwa-mem (v. 0.7. 17) (Li 2013), with the MinION analysis option (-x ont2d) (Jain  
367 *et al.* 2015), to a database derived from the Ribosomal Database Project (RDP Release 11,  
368 Update 5, Seqt. 30. 2016) (Cole *et al.* 2014), and the top hit was used for genus and species  
369 assignment. The RDP hierarchy browser ([http://rdp.cme.msu.edu/hierarchy/hb\\_intro.jsp](http://rdp.cme.msu.edu/hierarchy/hb_intro.jsp)) was  
370 used with the following filters: strain = "Type"; source = "isolates"; size ">= 1,200"; quality =  
371 "Good"; and taxonomy = "Nomenclatural" (for analytical scheme, see Figure 1C). Sequences  
372 from all the samples were normalized to the sample containing the lowest number of reads (5115  
373 reads). After the removal of singleton OTUs, all data analysis was carried out with R (v. 3.3.1)  
374 (R Core Team 2018). The R package vegan (v. 2.5-5) was applied for diversity and community  
375 analyses. Two metrics of alpha diversity were used in this study, Shannon diversity index  
376 (richness and evenness) and Chao1 index (richness). Significant differences in the number of  
377 OTUs and alpha diversity metrics between samples were determined using paired t-tests. Beta  
378 diversity was explored by non-metric multidimensional scaling (nMDS) of Bray-Curtis  
379 dissimilarity among samples. Statistical significance was calculated by permutational analysis of  
380 variance (PERMANOVA). Venn diagrams were visualized with the R package 'VennDiagram'  
381 (Chen and Boutros 2011), illustrating the shared and unique genera among each bathroom paired  
382 sample. Linear discriminant analysis (LDA) effect size (LEfSe) was applied to identify specific  
383 bacterial genera between the samples (Segata *et al.* 2011). Taxa were considered significant  
384 based on LDA scores greater than 4.0 and p-values smaller than 0.05.

### 385 **Possibility of pathogens in the bathroom**



386 The presence or absence of genera on the National Institute of Infectious Diseases (NIID) in  
387 Japan list of Bio Safety Level (BSL) 2 and 3 were checked in each sample (for the genus in NIID  
388 list, see Supplemental Table S4, and for the list of species, see the following link:  
389 <https://www.nite.go.jp/nbr/mrinda/list/risk/bacteria/ALL>, updated 2010-06). As a result, each  
390 read corresponding to BSL2 and BSL3 species was collected, and then sequences longer than  
391 1,000 bp were filtered by seqkit (v0.8.0). The sequences were clustered with psi-cd-hit.pl to  
392 identify representative sequences (Li and Godzik 2006; Fu *et al.* 2012). The settings in psi-cd-  
393 hit.pl were the default except for -prog blastn. For further analysis, the sequences with the  
394 highest number of reads in each cluster were used. Taxonomy was identified with BLASTn (E  
395 value = 1.0e-10, word size = 7, reward = 2, penalty = -3, gap open = 5, gap extend = 2, filter:  
396 unmarked “low complexity regions”) (Camacho *et al.* 2009) against 16S ribosomal RNA from  
397 curated type strain sequences from bacteria and archaea in GenBank (update date: 2021/01/11,  
398 number of sequences: 21,699). The analysis scheme is shown in Figure 1C.

## 399 **Data availability**

400 The authors declare that all the data supporting the findings of this study are available within the  
401 paper (and its Supplemental files) and that raw data were presented where possible. The raw  
402 MiSeq data reported in the paper (Figure 6, Supplemental Figures and Tables) have been  
403 uploaded to the DDBJ database under the accession number DRA010182. It will be open to  
404 public when this manuscript is accepted for publication

405

## 406 **Author contributions**

407 **S.F.** and **F.M.** conceived and designed the experiment. **S.F.** and **F.M.** collected samples. **S.F.**  
408 performed the experiments and analysed the data. **S.F.**, **Y.N.**, and **F.M.** interpreted the data. **S.F.**  
409 wrote the manuscript. **S.F.**, **Y.N.**, and **F.M.** reviewed drafts of the manuscript.

410

## 411 **Acknowledgements**

412 This work was supported by The Kyoto University Research Funds for Young Scientists and the  
413 Japan Society for the Promotion of Science under Grants-in-Aid for Scientific Research  
414 (KAKENHI) (grant number 20K18903 to S.F.). KAKENHI (grant numbers  
415 18K19674/18KK0436/20H00562) and the Japan Agency for Medical Research and  
416 Development (grant number 20wm0225012h0001/21fk0108129h0502) awarded grants to F.M.  
417 This manuscript was edited for English language by American Journal Experts (AJE).

418

## 419 **Competing interests**

420 The authors declare that the research was conducted in the absence of any commercial or  
421 financial relationships that could be construed as a potential conflict of interest.

422

## 423 **References**

424 Adams, R. I., Bateman, A. C., Bik, H. M., Meadow, J. F., 2015. Microbiota of the indoor  
425 environment: a meta-analysis. *Microbiome*, 3, 49. <https://doi.org/10.1186/s40168-015-0108-3>.

426 Allen, M. J., Edberg, S. C., Reasoner, D. J., 2004. Heterotrophic plate count bacteria —what is  
427 the significance in drinking water?. *Int J Food Microbiol*, 1, 265–74.  
428 <https://doi.org/10.1016/j.ijfoodmicro.2003.08.017>.

429 Andrews, S., 2010. *Fastqc: A quality control tool for high throughput sequence data*. Babraham  
430 Bioinformatics, Babraham Institute, Cambridge, United Kingdom.

431 Arai, S., Kim, H., Watanabe, T., Tohya, M., Suzuki, E., Ishida-Kuroki, K., Maruyama, F.,  
432 Murase, K., Nakagawa, I., Sekizaki, T., 2018. Assessment of pig saliva as a *Streptococcus suis*  
433 reservoir and potential source of infection on farms by use of a novel quantitative polymerase  
434 chain reaction assay. *Am J Vet Res*, 79, 941–948. <https://doi.org/10.2460/ajvr.79.9.941>.

435 Arikawa, K., Ichijo, T., Nakajima, S., Nishiuchi, Y., Yano H., Tamaru, A., Yoshida, S.,  
436 Maruyama, F., Ota, A., Nasu, M., Starkova, D. A., Mokrousov, I., Narvskaya, O. V., Iwamoto,  
437 T., 2019. Genetic relatedness of *Mycobacterium avium* subsp. *hominissuis* isolates from  
438 bathrooms of healthy volunteers, rivers, and soils in Japan with human clinical isolates from  
439 different geographical areas. *Infect Genet Evol*, 74, 103923.  
440 <https://doi.org/10.1061/j.meegid.2019.103923>.

- 441 Boe-Hansen, R., Albrechtsen, H. J., Arvin, E., Jørgensen, C., 2002. Bulk water phase and biofilm  
442 growth in drinking water at low nutrient conditions. *Water Res*, 36, 4477–86.  
443 [https://doi.org/10.1016/s0043-1354\(02\)00191-4](https://doi.org/10.1016/s0043-1354(02)00191-4).
- 444 Camacho, C., Coulouris, G., Avagyan, V., Ma, N., Papadopoulos, J., Bealer, K., Madden, T. L.,  
445 2009. Blast+: architecture and applications. *BMC Bioinformatics*, 10, 421.  
446 <https://doi.org/10.1186/1471-2105-10-421>.
- 447 Cervero-Aragó, S., Rodríguez-Martínez, S., Puertas-Bennasar, A., Araujo, R. M., 2015. Effect of  
448 common drinking water disinfectants, chlorine and heat, on free *Legionella* and amoebae-  
449 associated *Legionella*. *PLoS One*, 10, e0134726. <https://doi.org/10.1371/journal.pone.0134726>.
- 450 Chen, H., Boutros, P. C., 2011. VennDiagram: a package for the generation of highly-  
451 customizable Venn and Euler diagrams in R. *BMC Bioinformatics*, 12, 35.  
452 <https://doi.org/10.1186/1471-2105-12-35>.
- 453 Cole, J. R., Wang, Q., Fish, J. A., Chai, B., McGarrell, D. M., Sun, Y., Brown, C. T., Porras-  
454 Alfaro, A., Kuske, C. R., Tiedje, J. M., 2014. “Ribosomal Database Project: data and tools for  
455 high throughput rRNA analysis. *Nucleic Acids Res*, 42 (Database issue): D633–42.  
456 <https://doi.org/10.1093/nar/gkt1244>.
- 457 Dannemiller, K. C., Gent, J. F., Leaderer, B. P., Peccia, J., 2016. Indoor microbial communities:  
458 Influence on asthma severity in atopic and nonatopic children. *J. Allergy Clin Immunol*, 138, 76–  
459 83.e1. <https://doi.org/10.1016/j.jaci.2015.11.027>.
- 460 Denis, C., Irlinger, F., 2007. Safety assessment of dairy microorganisms: aerobic coryneform  
461 bacteria isolated from the surface of smear-ripened cheeses. *Int J Food Microbiol* 126: 311–5.  
462 <https://doi.org/10.1016/j.ijfoodmicro.2007.08018>.
- 463 Douterelo, I., Fish, K. E., Boxall, J. B., 2018. Succession of bacterial and fungal communities  
464 within biofilms of a chlorinated drinking water distribution system. *Water Res*, 15, 74–85.  
465 <https://doi.org/10.1016/j.watres.2018.04.058>.
- 466 Elsaghir, A. A. F., James, E. A., 2003. Misidentification of *Brucella melitensis* as *Ochrobactrum*  
467 *anthropi* by Api 20NE. *J Med Microbiol*, 52, 441–2. <https://doi.org/10.1099/jmm.0.05153-0>.
- 468 Falkinham, J. O., III, 2020a. *Mycobacterium avium* Complex (MAC) in water distribution  
469 systems and household plumbing in the United States. *Water*, 12, 3338.  
470 <https://doi.org/10.3390/w12123338>.
- 471 Falkinham, J. O., III, 2020b. Living with *Legionella* and other waterborne pathogens.  
472 *Microorganisms*, 8, 2026. <https://doi.org/10.3390/microorganisms8122026>.

- 473 Feazel, L. M., Baumgartner, L. K., Peterson, K. L., Frank, D. N., Harris, J. K., Pace, N. R., 2009.  
474 Opportunistic pathogens enriched in showerhead biofilms. *Proc Natl Acad Sci U S A*, 106,  
475 16393–9. <https://doi.org/10.1073/pnas.0908446106>.
- 476 Fish, K. E., Reeves-McLaren, N., Husband, S., Boxall, J., 2020. Uncharted waters: the  
477 unintended impacts of residual chlorine on water quality and biofilms. *npj Biofilms and*  
478 *Mirobiomes*. 6, 34. <https://doi.org/10.1038/s41522-020-00144>.
- 479 Fu, L., Niu, B., Zhu, Z., Wu, S., Li, W., 2012. Cd-Hit: accelerated for clustering the next-  
480 generation sequencing data. *Bioinformatics*, 28, 3150–2.  
481 <https://doi.org/10.1093/bioinformatics/bts565>.
- 482 Fujiyoshi, S., Muto-Fujita A., Maruyama, F., 2020a. Evaluation of PCR conditions for  
483 characterizing bacterial communities with full-length 16S rRNA genes using a portable nanopore  
484 sequencer. *Sci Rep*, 10, 12580. <https://doi.org/10.1038/s41598-020-69450-9>.
- 485 Fujiyoshi, S., Yarimizu, K., Miyashita, Y., Rilling, J., Acuña, J. J., Ueki, S., Gajardo, G.,  
486 Espinoza-González, O., Guzmán, L., Jorquera, M. A., Nagai, S., Maruyama, F., 2021. Suitcase  
487 Lab: new, portable, and deployable equipment for rapid detection of specific harmful algae in  
488 Chilean coastal waters. *Environ Sci Pollut Res Int*, 28, 14144–55.  
489 <https://doi.org/10.1007/s11356-020-11567-5>.
- 490 Gallego, V., García, M. T., Ventosa A., 2005. *Methylobacterium isbiliense* sp. nov., isolated  
491 from the drinking water system of Sevilla, Spain. *Int J Syst Evol Microbiol*, 55, 2333–7.  
492 <https://doi.org/10.1099/ijs.0.63773-0>.
- 493 Gebert, M. J., Delgado-Baquerizo, M., Oliverio, A. M., Webster, T. M., Nichols, L. M., Honda,  
494 J. R., Chan, E. D., Adjemian, J., Dunn, R. R., Fierer, N., 2018. Ecological analyses of  
495 Mycobacteria in showerhead biofilms and their relevance to human health. *mBio*, 9, e01614-18.  
496 <https://doi.org/10.1128/mBio.01614-18>.
- 497 Gee, J. E., De, B. K., Levett, P. N., Whitney, A. M., Novak, R. T., Popovic T., 2004. Use of 16S  
498 rRNA gene sequencing for rapid confirmatory identification of *Brucella* isolates. *J Clin*  
499 *Microbiol*, 42, 3649–54. <https://doi.org/10.1128/JCM.42.8.3649-3654.2004>.
- 500 Horvat, R. T., El Atrouni, W., Hammoud, K., Hawkinson, D., Cowden, S., 2011. Ribosomal  
501 RNA sequence analysis of *Brucella* infection misidentified as *Ochrobactrum anthropi* infection.  
502 *J Clin Microbiol*, 49, 1165–8. <https://doi.org/10.1128/JCM.01131-10>.
- 503 Huang, G., Ng, T. W., Chen, H., Chow, A.T., Liu, S., Wong, P. K., 2020. Formation of  
504 assimilable organic carbon (AOC) during drinking water disinfection: A microbiological  
505 prospect of disinfection byproducts. *Environ Int*, 135, 105389.  
506 <https://doi.org/10.1016/j.envint.2019.105389>.

- 507 Ichijo, T., Izumi, Y., Nakamoto, S., Yamaguchi, N., Nasu, M., 2014. Distribution and respiratory  
508 activity of Mycobacteria in household water system of healthy volunteers in Japan. PLoS One, 9,  
509 e110554. <https://doi.org/10.1371/journal.pone.0110554>.
- 510 Iwamoto, T., Nakajima, C., Nishiuchi, Y., Kato, T., Yoshida, S., Nakanishi, N., Tamaru, A.,  
511 Tamura, Y., Suzuki, Y., Nasu, M., 2012. Genetic diversity of *Mycobacterium avium* subsp.  
512 *hominissuis* strains isolated from humans, pigs, and human living environment. Infec Genet  
513 Evol, 12, 846–52. <https://doi.org/10.1061/j.meegid.2011.06.018>.
- 514 Jain, M., Fiddes, I. T., Miga, K. H., Olsen, H. E., Paten, B., Akeson, M., 2015. Improved data  
515 analysis for the MinION nanopore sequencer. Nat Methods, 12, 351–6.  
516 <https://doi.org/10.1038/nmeth.3290>.
- 517 Ji, P., Rhoads W. J., Edwards, M. A., Pruden, A., 2017. Impact of water heater temperature  
518 setting and water use frequency on the building plumbing microbiome. ISME J, 11, 1318–30.  
519 <https://doi.org/10.1038/ismej.2017.14>.
- 520 Jing, Z., Lu, Z., Mao, T., Cao, W., Wang, W., Ke, Y., Zhao, Z., Wang, X., Sun, X. 2021.  
521 Microbial composition and diversity of drinking water: A full scale special-temporal  
522 investigation of a city in northern China. Sci Total Environ, 776, 145986.  
523 <https://doi.org/10.1016/j.scitotenv.2021.145986>.
- 524 Johnson L. R., 2008. Microcolony and biofilm formation as a survival strategy for bacteria. J  
525 Theor Biol, 7, 24–34. <https://doi.org/10.1016/j.jtbi.2007.10.039>.
- 526 Kawai, M., Ichijo, T., Takahashi, Y., Noguchi, M., Katayama, H., Cho, O., Sugita, T., Nasu, M.,  
527 2019. Culture independent approach reveals domination of human-oriented microbes in a  
528 pharmaceutical manufacturing facility. Eur J Pharm Sci, 137, 104973.  
529 <https://doi.org/10.1016/j.ejps.2019.104973>.
- 530 Kelley, S.T., Gilbert, J.A., 2013. Studying the microbiology of the indoor environment. Genome  
531 Biol, 14, 202. <https://doi.org/10.1186/gb-2013-14-2-202>.
- 532 Kelley, S. T., Theisen, U., Angenent, L. T., St Amand, A., Pace, N. R., 2004. Molecular analysis  
533 of shower curtain biofilm microbes. Appl Environ Microbiol, 70, 4187–92.  
534 <https://doi.org/10.1128/AEM.70.7.4187-4192.2004>.
- 535 Kozajda, A., Ježak, K., Kapsa, A., 2019. Airborne *Staphylococcus aureus* in different  
536 environments—a review. Environ SciPollut Res Int, 26, 34741–53.  
537 <https://doi.org/10.1007/s11356-019-06557-1>.

- 538 Kormas, K. A., Neofitou, C., Pachiadaki, M., Koufostathi, E., 2010. Changes of bacterial  
539 assemblages throughout an urban drinking water distribution system. *Environ Monit Assess*, 165,  
540 27–38. <https://doi.org/10.1007/s10661-009-0924-7>.
- 541 Li, H., 2013. Aligning sequence reads, clone sequences and assembly contigs with BWA-MEM.  
542 arXiv preprint, arXiv, 1303.3997.
- 543 Li, W., Godzik, A., 2006. Cd-hit: a fast program for clustering and comparing large sets of  
544 protein or nucleotide sequences. *Bioinformatics*, 22, 1658–9.  
545 <https://doi.org/10.1093/bioinformatics/btl158>.
- 546 Li, W., Tan, Q., Zhou, W., Chen, J., Li, Y., Wang, F., Zang, J., 2020. Impact of substrate  
547 material and chlorine/chloramine on the composition and function of a young biofilm microbial  
548 community as revealed by high-throughput 16S rRNA sequencing. *Chemosphere*, 242, 125310.  
549 <https://doi.org/10.1016/j.chemosphere.2019.125310>.
- 550 Liu, Y., Hung, Y. H., Yuan, J., Chen, F, Cai, W., Liu, J., Ma, X., Xie, C., Zheng, C., Zhuo, L.,  
551 Cao, X., Tan, H., Li, B., Xie, H., Liu, Y., Ip, D., 2015. A foodborne outbreak of gastroenteritis  
552 caused by *Vibrio parahaemolyticus* and Norovirus through non-seafood vehicle. *PLoS One*, 10,  
553 e0137848. <https://doi.org/10.1371/journal.pone.0137848>.
- 554 Ma, X., Li, G., Chen, R., Yu, Y., Tao, H., Zhang G., Shi B., 2020. Revealing the changes of  
555 bacterial community from water source to consumers tap: A full-scale investigation in eastern  
556 city of China. *J Environ Sci*, 87: 331–40. <https://doi.org/10.1016/j.jes.2019.07.017>.
- 557 Maje, M. D., Kaptchouang Tchatchouang, C. D., Manganyi, M. C., Fri, J., Ateba, C. N., 2020.  
558 Characterisation of *Vibrio* species from surface and drinking water sources and assessment of  
559 biocontrol potentials of their bacteriophages. *Int J Microbiol*, 2020, 8863370.  
560 <https://doi.org/10.1155/2020/8863370>.
- 561 Mitsuhashi, S., Kryukov, K., Nakagawa, S., Takeuchi, J. S., Shiraishi, Y., Asano, K., Imanishi,  
562 T., 2017. A portable system for rapid bacterial composition analysis using a nanopore-based  
563 sequencer and laptop computer. *Sci Rep*, 7, 5657. <https://doi.org/10.1038/s41598-017-05772-5>.
- 564 Montagna, M. T., Cristina, M. L., De Giglio, O., Spagnolo, A. M., Napoli, C., Cannova, L.,  
565 Deriu, M. G., Delia, S. A., Giuliano, A., Guida, M., Laganà, P., Liguori, G., Mura, I., Pennino,  
566 F., Rossini, A., Tardivo, S., Torre, I., Torregrossa, M. V., Villafrate, M. R., Albertini, R.,  
567 Pasquarella, C., 2016. Serological and molecular identification of *Legionella* spp. isolated from  
568 water and surrounding air samples in Italian healthcare facilities. *Environ Res*, 146, 47–50.  
569 <https://doi.org/10.1016/j.envres.2015.12.015>.

- 570 Nakagawa, S., Inoue, S., Kryukov, K., Yamagishi, J., Ohno, A., Hayashida, K., Nakazwe, R.,  
571 Kalumbi, M., Mwenya, D., Asami, N., Sugimoto, C., Mutengo, M. M., Imanishi, T., 2019. Rapid  
572 sequencing-based diagnosis of infectious bacterial species from meningitis patients in Zambia.  
573 *Clin Transl Immunology*, 8, e01087. <https://doi.org/10.1002/cti2.1087>.
- 574 Namkoong, H., Kurashima, A., Morimoto, K., Hoshino, Y., Hasegawa, N., Ato, M., Mitarai, S.,  
575 2016. Epidemiology of pulmonary nontuberculous mycobacterial disease, Japan. *Emerg Infect*  
576 *Dis*, 22, 1116–7. <https://doi.org/10.3201/eid2206.151086>.
- 577 Neu, L., Bänziger, C., Proctor, C. R., Zhang, Y., Liu W. T., Hammers, F., 2018. Ugly  
578 ducklings—the dark side of plastic materials in contact with potable water. *npj Biofilms*  
579 *Microbiomes*, 4, 7. <https://doi.org/10.1038/s41522-018-0050-9>.
- 580 Nishiuchi, Y., Tamura, A., Kitada, S., Taguri, T., Matsumoto, S., Tateishi, Y., Yoshimura, M.,  
581 Ozeki, Y., Matsumura, N., Ogura, H., Maekura, R., 2009. *Mycobacterium avium* complex  
582 organisms predominantly colonize in the bathtub inlets of patients’ bathrooms. *Jpn J Infect Dis*,  
583 62, 182–6.
- 584 Nishiuchi, Y., Maekura, R., Kitada, S., Tamaru, A., Taguri, T., Kira, Y., Hiraga, T., Hirotani, A.,  
585 Yoshimura, K., Miki, M., Ito, M., 2007. The recovery of *Mycobacterium avium-intracellulare*  
586 complex (MAC) from the residential bathrooms of patients with pulmonary MAC. *Clin Infect*  
587 *Dis*, 45, 347–51. <https://doi.org/10.1086/519383>.
- 588 Nishiuchi, Y., Iwamoto, T., Maruyama, F., 2017. Infection sources of a common non-  
589 tuberculous mycobacterial pathogen, *Mycobacterium avium* complex. *Front Med*, 4, 27.  
590 <https://doi.org/10.3389/fmed.2017.00027>.
- 591 Novak Babič, M., Gostinčar, C., Gunde-Cimerman, N., 2020. Microorganisms populating the  
592 water-related indoor biome. *Appl Microbiol Biotechnol*, 104, 6443–62.  
593 <https://doi.org/10.1007/s00253-020-10719-4>.
- 594 Odamaki, T., Bottacini, F., Mitsuyama, E., Yoshida, K., Kato, K., Xiao, J. Z., van Sinderen, D.,  
595 2019. Impact of a bathing tradition on shared gut microbes among Japanese families. *Sci Rep*, 9,  
596 4380. <https://doi.org/10.1038/s41598-019-40938-3>.
- 597 Ojima, M., Toshima, Y., Koya, E., Ara, K., Tokuda, H., Kawai, S., Kasuga, F., Ueda, N., 2002.  
598 Hygiene measures considering actual distributions of microorganisms in Japanese households. *J*  
599 *Appl Microbiol*, 93, 800–9. <https://doi.org/10.1046/j.1365-2672.2002.01746.x>.
- 600 Parker, J., Helmstetter, A. J., Devey, D., Wilkinson, T., Papadopoulos, A. S. T., 2017. Field-based  
601 species identification of closely-related plants using real-time nanopore sequencing. *Sci Rep*, 7,  
602 8345. <https://doi.org/10.1038/s41598-017-08461-5>.

- 603 Peng, H., Zhang, Y., Wang, R., Liu, J., Liu, W. T., 2020. Assessing the contribution of biofilm to  
604 bacterial growth during stagnation in shower hoses. *Water Supply*, 20, 2564–76.  
605 <https://doi.org/10.2166/ws.2020.161>.
- 606 Proctor, C. R., Reimann, M., Vriens, B., Hammes, F., 2018. Biofilms in shower hoses. *Water*  
607 *Res*, 131, 274–86. <https://doi.org/10.1016/j.watres.2017.12.027>.
- 608 R Core Team, 2018. R: a language and environment for statistical computing. R Foundation for  
609 Statistical Computing, Vienna, Austria. <http://www.R-project.org>.
- 610 Schmidt, K., Mwaigwisya, S., Crossman, L. C., Doumith, M., Munroe, D., Pires, C., Khan, A.  
611 M., Woodford, N., Saunders, N. J., Wain, J., O’Grady, J., Livermore, D. M., 2016. Identification  
612 of bacterial pathogens and antimicrobial resistance directly from clinical urines by nanopore-  
613 based metagenomic sequencing. *J Antimicrob Chemother*, 72, 104–14.  
614 <https://doi.org/10.1093/jac/dkw397>.
- 615 Segata, N., Izard, J., Waldron, L., Gevers, D., Miropolsky, L., Garrett, W. S., Huttenhower, C.,  
616 2011. Metagenomic biomarker discovery and explanation. *Genome Biol*, 12, R60.  
617 <https://doi.org/10.1186/gb-2011-12-6-r60>.
- 618 Shen, W., Le, S., Li, Y., Hu, F., 2016. Seqkit: a cross-platform and ultrafast toolkit for FASTA/Q  
619 file manipulation. *PLoS One*, 11, e0163962. <https://doi.org/10.1371/journal.pone.0163962>.
- 620 Silva, I. P., Carneiro, C. S., Saraiva, M. A. F., Oliveira, T. A. S., Sousa, O. V., Evangelista-  
621 Barreto, N. S., 2018. Antimicrobial resistance and potential virulence of *Vibrio*  
622 *parahaemolyticus* isolated from water and bivalve mollusks from Bahia, Brazil. *Mar Pollut Bull*,  
623 131, 757–62. <https://doi.org/10.1016/j.marpolbul.2018.05.007>.
- 624 Stackebrandt, E., Goebel, B. M., 1994. Taxonomic note: a place for DNA-DNA reassociation  
625 and 16S rRNA sequence analysis in the present species definition in bacteriology. *Int J Syst Evol*  
626 *Microbiol*, 44, 846–9. <https://doi.org/10.1099/00207713-44-4-846>.
- 627 Telenti, A., Marchesi, F., Balz, M., Bally, F., Böttger, E. C., Bodmer, T., 1993. Rapid  
628 identification of mycobacteria to the species level by polymerase chain reaction and restriction  
629 enzyme analysis. *J Clin Microbiol*, 31, 175–8. <https://doi.org/10.1128/JCM.31.2.175-178.1993>.
- 630 Tochiara, Y., 1999. Bathing in Japan: A Review. *J Human Environment System*, 3, 27–34.
- 631 Tortoli, E., 2014. Microbiological features and clinical relevance of new species of the genus  
632 *Mycobacterium*. *Clin Microbiol Rev*, 27, 727–52. <http://doi.org/10.1128/CMR.00035-14>.



- 633 Waak, M., Lapara, T. M., Halle, C., Hozalski, R. M., 2018. Occurrence of *Legionella* spp. In  
634 water-main biofilms from two drinking water distribution systems. *Environ. Sci. Technol*, 52,  
635 7630–9. <https://doi.org/10.1021/acs.est.8b01170>.
- 636 White, D. C., Sutton, S. D., Ringelberg, D. B., 1996. The genus *Sphingomonas*: physiology and  
637 ecology. *Curr Opin Biotechnol*, 7, 301–6. [https://doi.org/10.1016/s0958-1669\(96\)80034-6](https://doi.org/10.1016/s0958-1669(96)80034-6).
- 638 Wommack, K. E., Bhavsar, J., Ravel, J., 2008. Metagenomics: read length matters. *Appl Environ*  
639 *Microbiol*, 74, 1453–63. <https://doi.org/10.1128/AEM.02181-07>.
- 640 Yang, J., Ren, X. Q., Chu, M. L., Meng, D. Y., Xue, W. C., 2013. Mistaken identity of *Brucella*  
641 infection. *J Clin Microbiol*, 51, 2011. <https://doi.org/10.1128/JCM.03222-12>.
- 642 Yano, H., Iwamoto, T., Nishiuchi, Y., Nakajima, C., Starkova, D. A., Mokrousov, I., Narvskaya,  
643 O., Yoshida, S., Arikawa, K., Nakanishi, N., Osaki, K., Nakagawa, I., Ato, M., Suzuki, Y.,  
644 Maruyama, F., 2017. Population structure and local adaptation of MAC lung disease agent  
645 *Mycobacterium avium* subsp. *hominissuis*. *Genome Biol Evol*, 9 2403–17.  
646 <https://doi.org/10.1093/gbe/evx183>.
- 647 Yano, T., Kubota, H., Hanai, J., Hitomi, J., Tokuda, H., 2013. Stress tolerance of  
648 *Methylobacterium* biofilms in bathrooms. *Microbes Environ*, 28, 87–95.  
649 <https://doi.org/10.1264/jsme2.me12146>.
- 650 Zhou, Z., Xu, L., Zhu, L., Liu, Y., Shuai, X., Lin, Z., Chen, H., 2021. Metagenomic analysis of  
651 microbiota and antibiotic resistome in household activated carbon drinking water purifiers.  
652 *Environ Int*, 148, 106394. <https://doi.org/10.1016/j.envint.2021.106394>.  
653

## 654 **Figure legends**

655 **Figure 1.** Sampling and sequencing procedure of showerhead biofilms and shower water.  
656 Sampling place and method (A). Out of the 25 samples collected from five regions in Japan, 16  
657 samples were analysed. The denominator indicates the number of samples collected, and the  
658 numerator indicates the number of samples available for analysis (B). The analysis scheme is  
659 divided into two parts: community (left) and pathogen analysis (right) (C).

660

661 **Figure 2.** SEM image of a biofilm on the inner surface of a showerhead obtained from one water  
662 distributor.

663

664 **Figure 3.** Alpha-diversity measured using the Shannon index. Each box plot represents the  
665 diversity distribution of samples from water and biofilm; \*\*  $p < 0.01$ .

666

667 **Figure 4.** Non-metric multidimensional scaling (nMDS) analysis based on Bray-Curtis  
668 dissimilarity metrics of samples labelled according to the sample sources. A clear separation  
669 between biofilm bacterial communities and water communities was revealed (PERMANOVA, F-  
670 value=3.416, R-squared=0.096, p-value=< 0.01). Ellipses represent the 95% confidence interval.  
671 Pink: Biofilm, Blue: water

672

673 **Figure 5.** Comparison of the shared genera in biofilm and water samples from the same  
674 bathroom. Venn diagrams show a comparison of the percentage of shared genera for biofilms  
675 (pink) and water (blue) collected from the same bathroom. Parentheses indicate the percentage of  
676 the shared genus sequences in each sample.

677

678 **Figure 6.** Taxonomic composition of the bacterial community in the biofilm and water samples  
679 at the genus level using a stacked bar plot (top 10 genera).

680

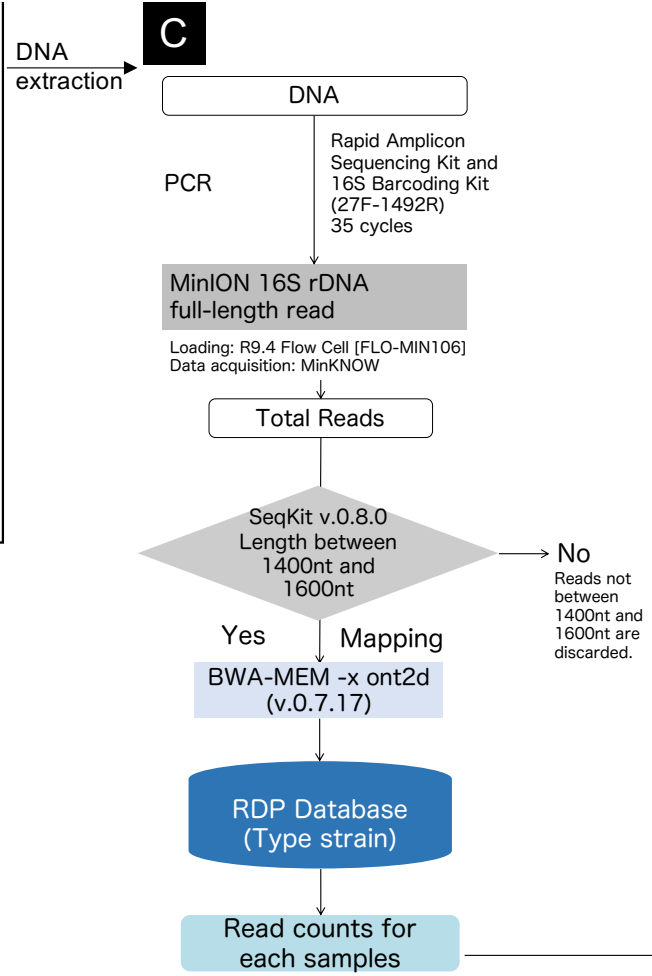
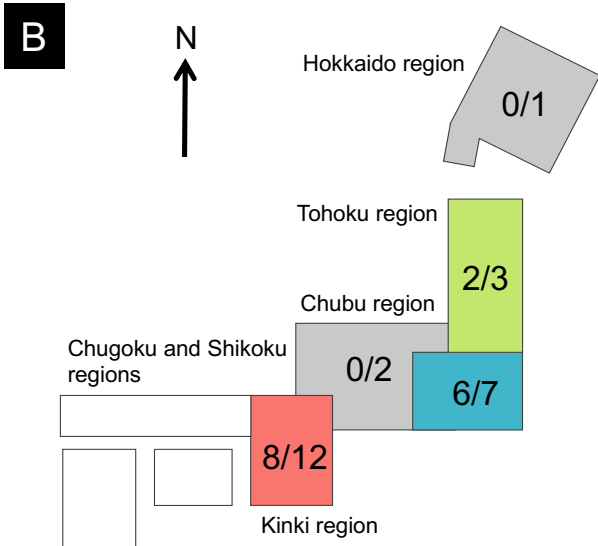
681 **Figure 7.** LEfSe analysis of the bacterial community of water and biofilm. Both water and  
682 biofilm samples were analysed, but no indicator genera were detected in biofilm. Features are  
683 significant based on their adjusted p-value and LDA score. The adjusted p-value cut-off=0.05  
684 and LDA score=4.0.

685

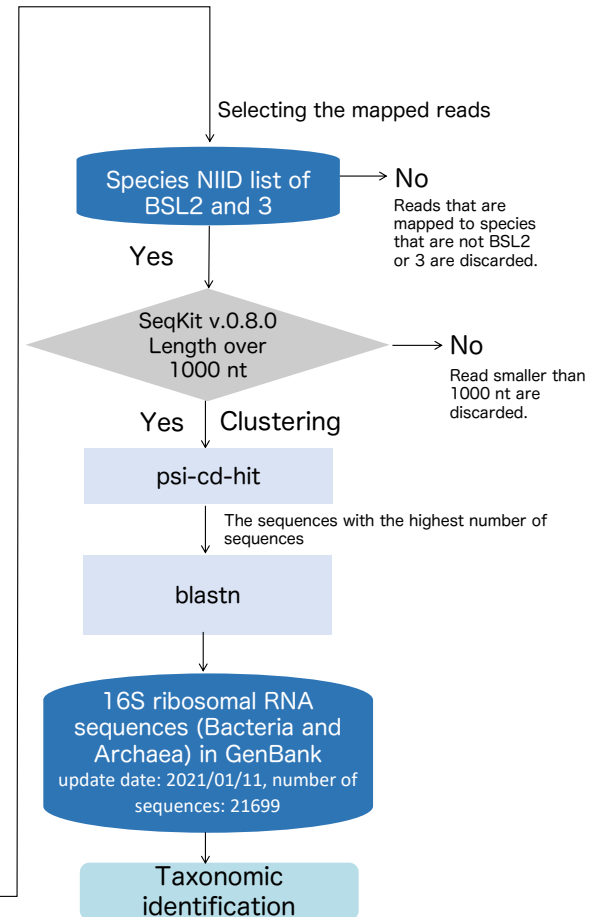
686 **Figure 8.** This heatmap table summarizes the results for all showerhead swab and shower water  
687 libraries pooled at the genus level and grouped by region in Japan. Figure footnotes: – signifies  
688 that the genera were not detected in the sample; SH indicates a biofilm sample and W indicates a  
689 water sample.

690

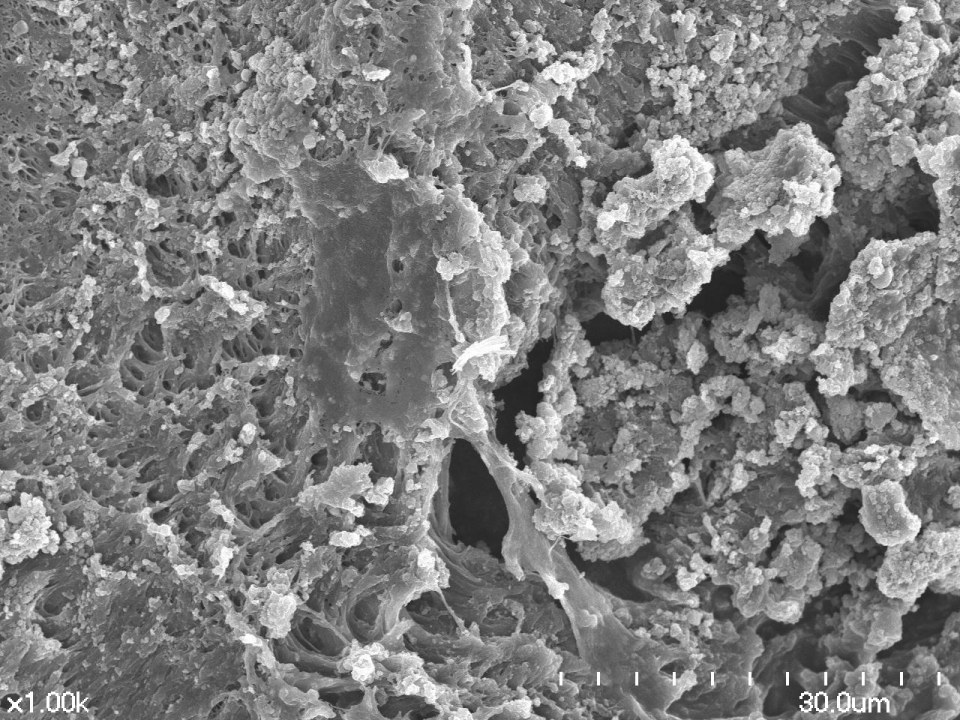
691 **Supplemental Figure S1.** Taxonomic composition of the bacterial community in each sample at  
692 the genus level using a stacked bar plot (top 30 genera).



Bacterial community analysis

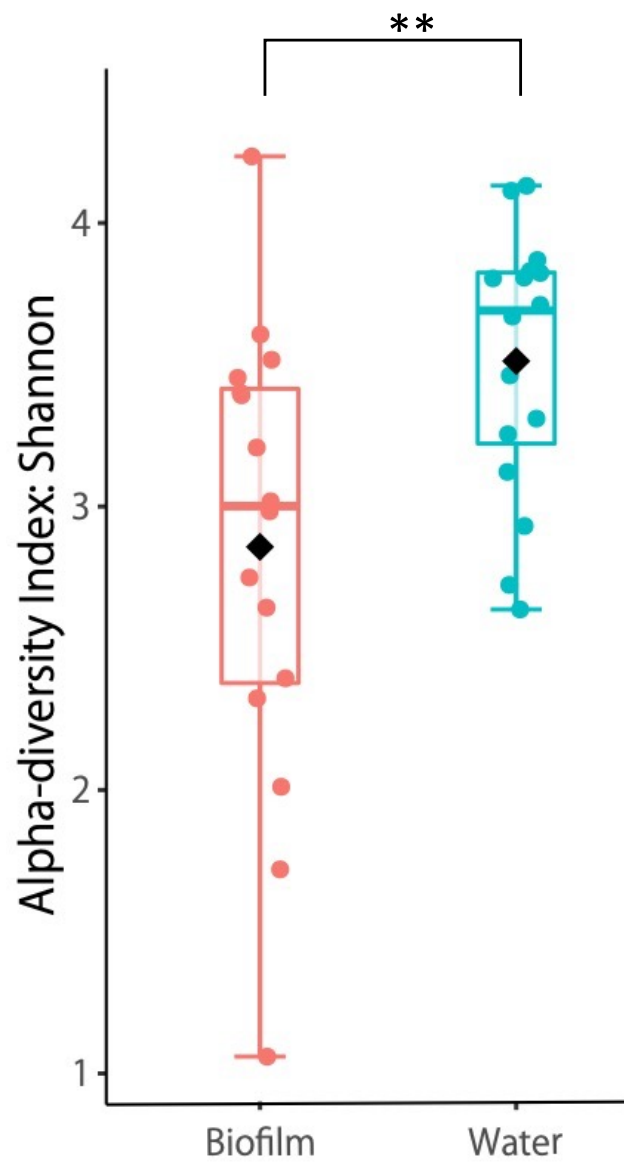


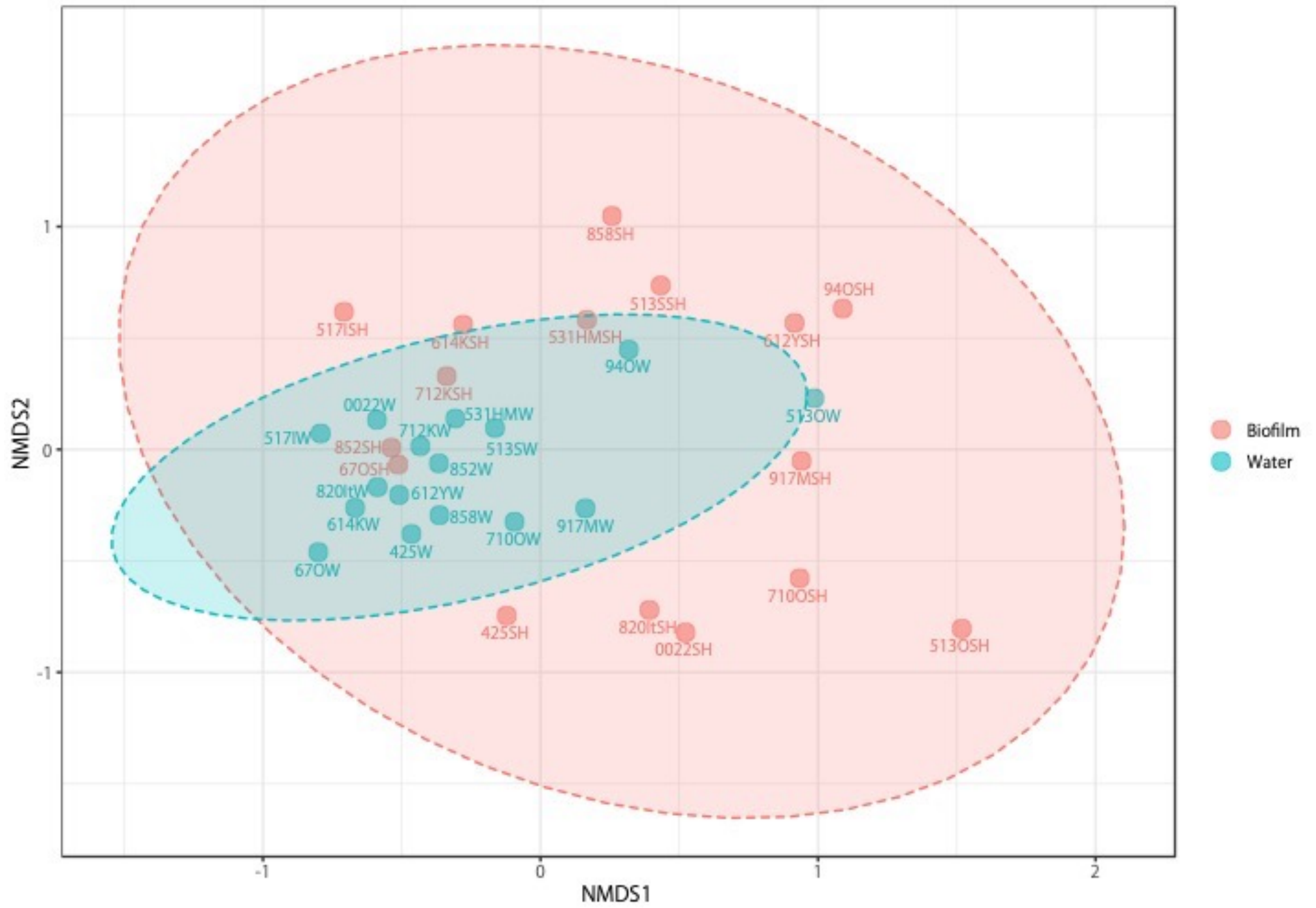
Analysis of pathogenic bacteria



x1.00k

30.0um

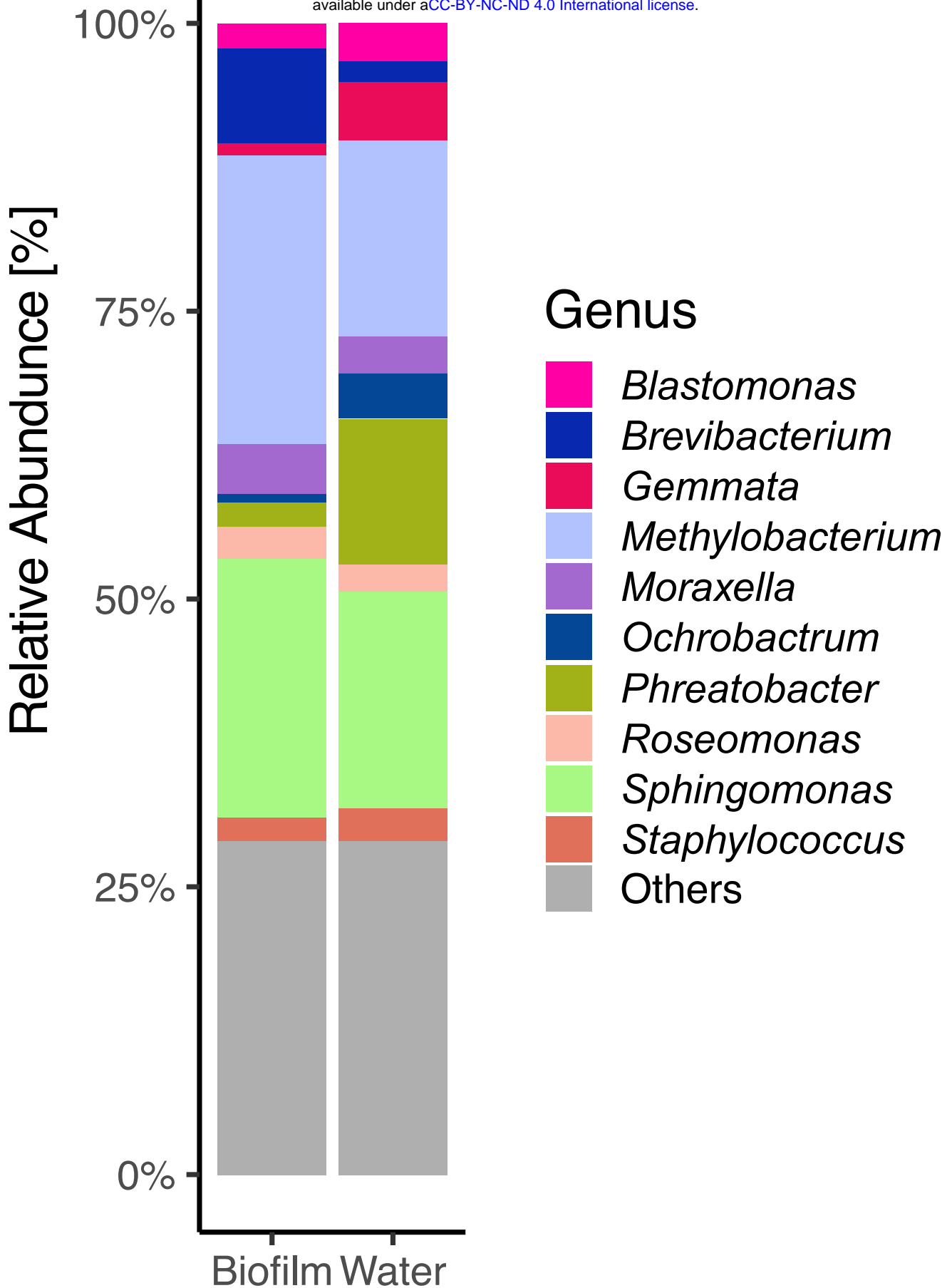


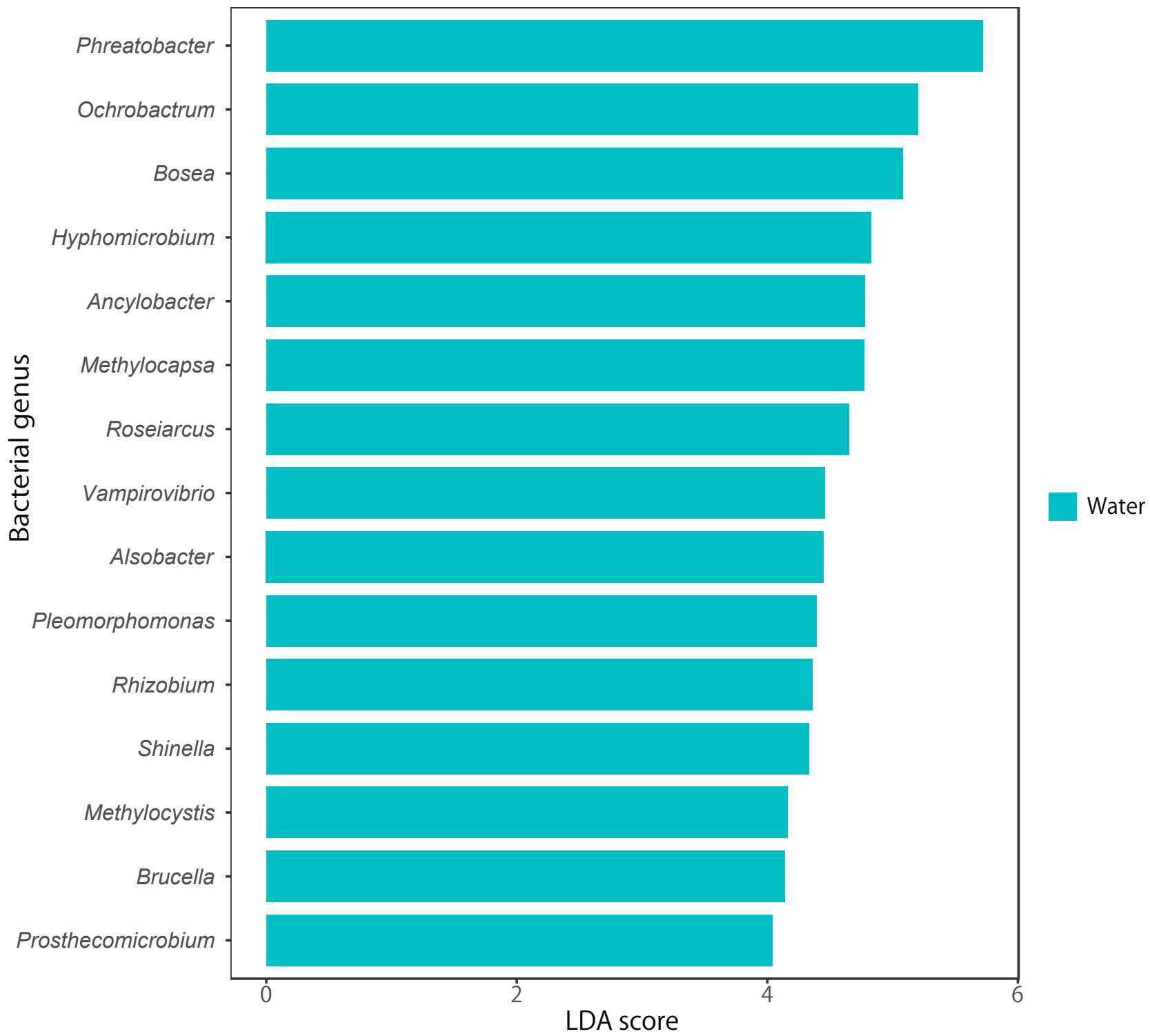


bioRxiv preprint doi: <https://doi.org/10.1101/2021.07.14.452346>; this version posted July 15, 2021. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under aCC-BY-NC-ND 4.0 International license.

Sample name	Biofilm [%] (%)	Shared genus /Total genus	Water [%] (%)
425SH-425W	58.7 (84.9)		23.9 (60.8)
858SH-858W	78.6 (99.8)		23.9 (50.5)
513OSH-513OW	36.8 (45.3)		38.9 (89.6)
94OSH-94OW	46.2 (98.4)		14.0 (76.3)
67OSH-67OW	76.3 (96.2)		77.6 (98.6)
710OSH-710OW	88.2 (99.8)		33.7 (72.5)
712KSH-712KW	86.8 (98.8)		52.3 (89.8)
820ItSH-820ItW	43.5 (25.2)		31.3 (50.9)
852SH-852W	93.6 (99.9)		51.2 (87.5)
513SSH-513SW	81.3 (99.9)		18.6 (79.2)
517ISH-517IW	68.9 (97.5)		42.5 (62.0)
531HMSH-531HMW	85.7 (99.8)		29.0 (76.0)
614KSH-614KW	52.8 (94.8)		36.4 (64.4)
612YSH-612YW	46.7 (86.3)		8.3 (43.8)
917MSH-917MW	82.6 (98.8)		30.2 (83.4)
0022SH-0022W	30.9 (20.0)		26.6 (67.1)
Average (% of No. of shared reads)	66.1 (84.1)	-	36.1 (72.0)







Sample ID	Region	Acinetobacter	Brucella	Chryseobacterium	Clostridium	Corynebacterium	Haemophilus	Moraxella	Mycobacterium	Nocardia	Pseudomonas	Staphylococcus	Streptococcus	Vibrio
		Relative abundance [%]												
425SH	Kinki	-	-	0.03	0.06	0.30	-	1.09	2.13	0.03	-	-	-	-
425W		-	0.46	0.15	-	-	-	-	0.43	-	-	0.03	0.15	-
858SH	Kinki	-	-	-	-	-	-	0.12	-	-	-	-	-	-
858W		-	0.27	-	-	-	-	-	18.28	0.06	-	0.21	0.12	-
513OSH	Kinki	-	-	-	-	3.71	0.12	35.74	-	-	-	0.82	0.82	-
513OW		-	-	-	0.06	-	-	-	0.12	-	-	8.30	0.15	-
94OSH	Kinki	-	-	-	-	-	-	-	-	-	-	0.76	0.49	-
94OW		-	-	-	-	0.03	-	-	4.01	8.24	-	-	-	-
67OSH	Kinki	-	0.58	-	-	-	-	-	-	-	-	-	-	-
67OW		-	-	1.70	-	-	0.12	0.09	0.85	-	-	-	-	0.03
710OSH	Kinki	-	0.03	-	-	1.16	-	0.94	-	-	-	0.15	0.15	-
710OW		-	-	0.24	-	0.09	3.07	-	5.87	-	-	-	-	0.03
712KSH	Kinki	-	0.09	-	-	-	-	-	-	-	-	-	-	-
712KW		-	-	0.21	-	-	-	-	-	0.06	-	-	-	-
820tSH	Kinki	0.03	-	-	-	5.20	-	22.48	-	-	-	-	-	-
820tW		-	-	0.76	-	-	-	-	0.03	-	-	-	-	-
852SH	Kanto	-	0.40	-	-	-	-	0.06	-	-	-	0.03	-	-
852W		0.06	0.43	-	-	0.09	-	-	13.78	0.91	-	0.12	0.43	0.03
513SSH	Kanto	-	-	-	-	-	-	0.06	-	-	-	-	-	-
513SW		-	-	0.12	-	-	-	0.03	0.46	-	-	-	0.30	0.24
517ISH	Kanto	-	-	-	-	0.06	-	0.03	-	-	-	-	-	-
517IW		-	-	0.24	-	0.03	0.06	-	0.33	-	-	-	3.32	-
531HMSH	Kanto	-	-	-	-	-	-	0.03	-	-	-	-	-	-
531HMW		-	-	0.15	-	-	-	-	0.06	0.09	-	-	1.58	-
614KSH	Kanto	-	-	-	-	0.24	-	0.79	0.27	-	-	-	-	-
614KW		-	-	0.27	-	-	-	-	0.85	-	-	0.03	-	0.03
612YSH	Kanto	-	-	-	-	-	-	5.66	-	-	-	9.67	-	-
612YW		0.03	0.36	-	0.06	-	-	-	0.03	0.46	-	-	0.40	-
917MSH	Tohoku	0.52	-	0.21	-	10.10	-	1.52	-	-	-	19.95	4.84	0.67
917MW		0.06	0.12	-	0.12	0.79	-	-	6.75	0.13	-	0.09	32.03	0.03
0022SH	Tohoku	-	0.03	-	-	0.03	-	2.46	-	-	-	0.58	1.13	0.40
0022W		-	-	0.33	-	-	-	-	0.06	-	-	-	0.12	0.03
Mean [%]		0.02	0.21	0.01	0.01	0.78	0.01	3.82	0.40	0.00	0.01	2.46	0.26	0.04
Median [%]		0.00	0.11	0.00	0.00	0.00	0.00	0.40	0.00	0.00	0.00	0.03	0.00	0.00
Appearance frequency [%]		16	59	9.0	19	44	9.0	81	34	3.0	9.0	53	47	13

Shade key

