

1 Multiple parasitoid species enhance top-down control, but parasitoid  
2 performance is context-dependent

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12 **Running Head:** Trophic interactions depend on context

13 **Data availability:** All raw data used for this study are available from the Zenodo database:

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15 **Abstract.** Ecological communities are composed of a multitude of interacting species, forming  
16 complex networks of interactions. Current global changes are altering community composition  
17 and, as such, we need to understand if the mechanisms structuring species interactions are  
18 consistent across different species compositions. However, it is a challenge to identify which  
19 aspects of species interactions are primarily driven by community structure and which by species  
20 identity. Here we compared the outcome of host-parasitoid interactions across four community  
21 modules that are common in host-parasitoid communities (i.e., host-parasitoid, exploitative  
22 competition, alternative host, and combined exploitative competition and alternative host modules)  
23 to measure the influence of structure and identity on the outcome of host-parasitoid interactions.  
24 In a laboratory experiment, we produced nine different species assemblage combinations per  
25 community module using a pool of three *Drosophila* host and three larval parasitoid species. We  
26 found that multiple parasitoid species enhanced host suppression due to sampling effect and  
27 weaker interspecific than intraspecific competition between parasitoids. Presence of an alternative  
28 host species had no general effects on host suppression nor on parasitoid performance, therefore  
29 showing no evidence of indirect interactions between host species nor any host switching behavior.  
30 However, effects of community structure on parasitoid performance were species-specific and  
31 dependent on the identity of co-occurring species. Consequently, multiple parasitoid species  
32 generally strengthen top down-control, but performance of the parasitoids depends on the identity  
33 of either the co-occurring parasitoid species, the alternative host species, both, or none. This  
34 implies that effects of co-occurring species should be taken into account for understanding  
35 ecosystem functioning and services. Understanding the effects of the biotic context on species  
36 interactions is essential when developing biological control strategies, especially in open  
37 agricultural systems.

38 **Key words:** community modules, multiple predator effects, community composition, interaction  
39 modification, *Drosophila*

## 40 **Introduction**

41 In nature, species interact in a variety of ways, forming complex ecological networks (Fontaine et  
42 al. 2011, Kéfi et al. 2012, 2015, Piloosof et al. 2017, García-Callejas et al. 2018, Miele et al. 2019).  
43 How species interact depends on the structure of the community, but also on identity of species in  
44 the assemblage (Bográn et al. 2002). Effects of the biotic context on species interactions is  
45 particularly relevant to predicting natural pest control. Moreover, in response to environmental  
46 changes, such as climate warming, species ranges and phenologies are shifting (Parmesan and  
47 Yohe 2003). Differences in the sensitivity and responsiveness to these changes among species are  
48 disrupting historical patterns of interactions and co-occurrences, creating communities with novel  
49 species compositions (Alexander et al. 2015). To accurately predict top-down control and forecast  
50 the ecological consequences of changes in the biotic environment induced by global changes, we  
51 need to determine which aspects of species interactions are primarily driven by community  
52 structure, and which are driven by species identity.

53 Trophic (i.e., direct predation) and non-trophic (e.g., competition, facilitation) interactions act in  
54 combinations to shape communities (Thierry et al. 2019), and their dynamics (Kawatsu et al.  
55 2021). For example, a predator might switch prey species with the presence of a competing  
56 predator, or with change in relative prey abundances (Siddon and Witman 2004, Randa et al. 2009).  
57 Top-down control might be driven by a single influential predator species independently of  
58 predator diversity (Letourneau et al. 2009), be enhanced with multiple predator species (i.e., risk  
59 enhancement for the prey) if predators show some degree of niche differentiation (Greenop et al.  
60 2018), or be weakened when multiple predator species are present (i.e., risk reduction for the prey)

61 due to exploitative competition, interference or intraguild predation (Sih et al. 1998). Most studies  
62 looking at complex networks are observational (e.g., Tylianakis et al. 2007, Jeffs et al. 2021), and  
63 are typically unable to disentangle the potential mechanisms driving species interactions described  
64 above. Experimental systems are thus needed to separate the mechanisms structuring networks of  
65 interacting species. For this purpose, community modules (i.e., a small number of species  
66 interacting in a specified pattern; Holt 1997, also referred to as “motifs” in the literature; Milo et  
67 al. 2002) represent a powerful tool to isolate certain key interactions that structure complex  
68 networks. They are the building blocks of natural communities (Gilman et al. 2010), and thus allow  
69 us to disentangle the mechanisms structuring them. Common community modules in food webs  
70 include pairwise predator-prey (direct interactions), two prey species sharing a common natural  
71 enemy (indirect interactions; i.e., apparent competition or mutualism; hereafter alternative host  
72 module), or two predator species attacking the same prey species (either direct or indirect  
73 interactions; i.e., exploitative competition, interference, or facilitation; hereafter exploitative  
74 competition module). But experimental studies investigating the mechanisms structuring  
75 interactions with community modules rarely consider potential variations due to species-specific  
76 effects (but see Bográn et al. 2002, Snyder et al. 2006). Thus, it is unclear whether the mechanisms  
77 structuring interactions are consistent when looking at community modules of different species  
78 compositions (Cusumano et al. 2016).

79 Experiments manipulating interactions in different community contexts with different species  
80 assemblages usually manipulate species assemblage of one trophic level at a time. For instance,  
81 (Bográn et al. 2002) revealed competitive interactions among predator species in some, but not all  
82 the predator assemblages studied. However, the study used only one prey species. Snyder et al.  
83 (2006) found varying strength in the effect of predator species diversity on aphid suppression

84 depending on the aphid species considered, but did not vary species composition in the multiple  
85 predator treatment. Woodcock and Heard (2011) showed that risk of predation for planthoppers  
86 was reduced in systems where predators shared the same habitat domain, but did not manipulate  
87 the herbivore community. We need studies manipulating species assemblages at both trophic levels  
88 to understanding how the identity of co-occurring species affects the outcome of consumer-  
89 resource interactions.

90 Here, we investigated the mechanisms structuring consumer-resource interactions using a host-  
91 parasitoid system. Parasitoids are a diverse group of insects that use other arthropods as a nursery  
92 for their offspring, ultimately killing their host to complete development (Godfray 2004).  
93 Parasitoids are important for top-down control in agricultural and natural ecosystems, and widely  
94 used as biological control agents. Interactions between hosts and parasitoids are easily observed,  
95 and represent a good model system to study how the structure and composition of communities  
96 influence species interactions. We used a set of three *Drosophila* species and three of their larval  
97 parasitoids from a natural tropical community in Australia (Jefferies et al. 2021). Under laboratory  
98 conditions, we reproduced isolated subsets of this community to look at direct and indirect  
99 interactions within host-parasitoid communities. We aimed to uncover general effects of  
100 community modules in our *Drosophila*-parasitoid system across different species assemblages,  
101 and detect any species-specific effects depending on the co-occurring species identity (using 9  
102 species assemblages for each of the four common community modules in host-parasitoid networks:  
103 host-parasitoid pair, exploitative competition module, alternative host module, and combined  
104 exploitative competition and alternative host module). We also explored the mechanisms behind  
105 the effects of community structure and composition on host-parasitoid interactions and top-down  
106 control on the hosts. Specifically, we tested the following hypotheses: (i) host suppression will be

107 higher with the presence of multiple parasitoid species (i.e., exploitative competition module), due  
108 to an increased chance that a more efficient parasitoid species will be present (i.e., sampling effect)  
109 (Pedersen and Mills 2004, Letourneau et al. 2009). However, (ii) parasitoid performance might  
110 decrease with the presence of multiple parasitoid species due to increased larval competition in  
111 multiparasitism events (Harvey et al. 2013). (iii) Interactions between a focal host-parasitoid pair  
112 will weaken with the presence of an alternative host because of trait- and density- mediated effects  
113 (Werner and Peacor 2003). (iv) Combined effects of exploitative competition among parasitoids  
114 and alternative host species in the four-species module will differ from three-species modules  
115 depending on the identity of the co-occurring species (Bográn et al. 2002, Sentis et al. 2017).

## 116 **Methods**

### 117 *Study system*

118 The experiment used cultures of *Drosophila* species and their associated parasitoids collected from  
119 two tropical rainforest locations in North Queensland Australia: Paluma (S18° 59.031' E146°  
120 14.096') and Kirrama Range (S18° 12.134' E145° 53.102') (< 100 m above sea level) (Jefferies et al.  
121 2021). *Drosophila* and parasitoid cultures were established between 2017 and 2018, identified  
122 using both morphology and DNA barcoding, and shipped to the Czech Republic under permit no.  
123 PWS2016-AU-002018 from Australian Government, Department of the Environment. All cultures  
124 were maintained at 23°C on a 12:12 hour light and dark cycle at Biology Centre, Czech Academy  
125 of Sciences. Three host species (*Drosophila birchii*, *D. simulans* and *D. pallidifrons*), and three  
126 larval parasitoid species *Asobara* sp. (Braconidae: Alysiinae; strain KHB, reference voucher no.  
127 USNMENT01557097, reference sequence BOLD process ID:DROP043-21), *Leptopilina* sp.  
128 (Figitidae: Eucolinae; strain 111F, reference voucher no. USNMENT01557117, reference  
129 sequence BOLD process ID:DROP053-21), and *Ganaspis* sp. (Figitidae: Eucolinae; strain 84BC,

130 reference voucher no. USNMENT01557102 and USNMENT01557297, reference sequence  
131 BOLD process ID:DROP164-21) were used (for more details on the parasitoid strains used see  
132 (Lue et al. 2021). *Drosophila* isofemale lines were kept on standard *Drosophila* medium (corn  
133 flour, yeast, sugar, agar and methyl-4-hydroxybenzoate) for approximately 45 to 70 non-  
134 overlapping generations. To revive genetic variation, four to seven lines from each host species  
135 were combined to establish two population cages per species of mass-bred lines prior the start of  
136 the experiment. Single parasitoid isofemale lines were used, and maintained for approximately 25  
137 to 40 non-overlapping generations prior to the start of the experiment by providing them every  
138 week with two-days-old larvae of *Drosophila melanogaster*. This host species was not used in the  
139 experiment, thus avoiding potential bias due to maternal effects.

#### 140 *Experimental design*

141 To investigate the effects of community structure and species composition on host-parasitoid  
142 interactions, we used four community modules, and nine different species assemblages each (six  
143 host and six parasitoid assemblages, from the pool of three host species and three parasitoid  
144 species) (Figure 1). Each unique treatment was replicated six times. Each replicate was represented  
145 by a set of two vials in one box, for a total of 216 boxes. Either conspecific (Figure 1a and c) or  
146 heterospecific (Figure 1b and d) parasitoids were used. The two vials contained *Drosophila* larvae  
147 from either the same host species (Figure 1a and b) or different host species (Figure 1c and d).  
148 Host and parasitoid densities were kept constant across treatments (i.e., substitutive design; e.g.,  
149 Snyder et al. 2008). We also included control treatments for each host species to acquire baseline  
150 levels of survival in the absence of parasitoids (replicated eight times each).

151 To initiate the experiment, twenty-five eggs of each host species were placed in a single 90 mm  
152 high and 28 mm diameter glass vial with 10mL of food media. To collect *Drosophila* eggs, an egg-

153 wash protocol was developed based on Nouhaud et al. 2018). One day before the egg-washed  
154 protocol was conducted, two egg-laying mediums (petri dishes with agar gel topped with yeast  
155 paste) were introduced in each population cage for flies to laying eggs overnight. We used a #3  
156 round paintbrush and distilled water to rinse out the yeast paste and collect the eggs on a fine black  
157 mesh that allowed only yeast and water to filter through, leaving the eggs on the surface. Eggs  
158 were transferred into petri dishes containing PBS (1 mM Calcium chloride dihydrate, 0.5 mM  
159 Magnesium Chloride Hexahydrate, pH = 7.4), and collected under microscope using a pipette to  
160 transfer eggs from the petri dish with PBS to the experimental vials.

161 After 48 hours, two vials with *Drosophila* second instar larvae (initially eggs) were placed in a  
162 hermetically sealed plastic box (15x11x19 cm) with four 3-to-5-days-old parasitoids (1:1 sex  
163 ratio). Twenty-four hours later, parasitoids were removed, and vials were removed from the boxes  
164 and plugged for rearing (Figure S1). Every vial was checked daily for emergences until the last  
165 individual emerged (up to 41 days for the species with the longest developmental time). We  
166 stopped collecting host emerges after five consecutive days without emerges to avoid collecting  
167 the second generation. All emerges were collected, identified, sexed, and stored in 95% ethanol.  
168 A total of 11,400 host eggs were collected across 456 experimental vials, of which 7,494 (65.7%)  
169 successfully emerged as adults (3,702 hosts and 3,792 parasitoids).

170 To explore parasitism strength depending on host density, we estimated parasitoid functional  
171 responses for each of the three parasitoid species on each of the three host species (nine host-  
172 parasitoid combinations). An additional 14,760 host eggs were collected for this experiment across  
173 432 experimental vials. The method used to set the experiment was the same as described above,  
174 and both experiments were run in parallel. The detailed method is presented in Appendix S2, and  
175 the parasitoid functional responses are presented in Figure S2. The empirical results for the host



176 density at 25 eggs per 10mL of food media also served to measure degree of infestation and  
177 successful parasitism rate for each of the nine host-parasitoid pairs (Appendix S3 Figure S3).

### 178 *Data analysis*

179 We characterized the outcome of host-parasitoid interactions by a combination of degree of  
180 infestation (DI) for each host species (i.e., the probability of a larvae to be parasitized and die),  
181 and successful parasitism rate (SP) for each host-parasitoid pair representing parasitoid  
182 performance (i.e., the probability of a parasitized host to give rise to an adult parasitoid) (Carton  
183 and Kitano 1981, Boulétreau and Wajnberg 1986). Degree of infestation (DI) was calculated as  
184 the proportion of host attacked (the difference between adult hosts emerging from the controls  
185 without parasitoids and from the experiment) among the total of hosts (set to 0 if the number of  
186 hosts emerging from the experiment was greater than the controls). Successful parasitism rate (SP)  
187 was calculated as the proportion of parasitoid adults emerging among the number of hosts attacked  
188 (Carton and Kitano 1981, Boulétreau and Wajnberg 1986). If no parasitoid emerged or if the  
189 number of hosts attacked was estimated to be zero, SP was set to 0. If the number of parasitoids  
190 that emerged was greater than the estimated number of hosts attacked, SP was set to 1. For  
191 treatments with one parasitoid species, we assumed that each of the two parasitoid individuals  
192 were attacking hosts equally, therefore the number of parasitoid adults emerging was divided by  
193 two to allow comparison of parasitism rates between single and multiple parasitoid species.

194 Data were analyzed using generalized linear mixed-effects models (GLMMs). Model assumptions  
195 were verified with the *DHARMA* package (Hartig 2019). To correct for overdispersion of the  
196 residuals, data were modeled using a beta binomial error distribution and a logit link function using  
197 the *glmmTMB* function from the *TMB* package (Lüdecke et al. 2019). We compared the AIC of  
198 models with or without zero inflation. We chose the models with the smaller AIC, which were the

199 zero-inflated models. Three model types were used to investigate general effects of community  
200 modules, species-specific responses, and effects of community composition for each focal species.  
201 (i) “Community module models” used two explanatory variables and their two-way interaction to  
202 account for the fully-factorial design of the experiment that resulted in four community modules  
203 (exploitative competition treatment with two levels: presence or not of a parasitoid heterospecific,  
204 and alternative host treatment with two levels: presence or not of an alternative host species). Box  
205 ID (214 levels) was included as a random factor to remove the variation between the different  
206 species assemblages and thus extract general effects of community modules. Host species (three  
207 levels) for DI, and host-parasitoid pairs for SP were also included as random factors to remove the  
208 variation between different species. (ii) “Species-specific community module models” used the  
209 same explanatory variables and Box ID as a random factor as previously described, but host species  
210 and host-parasitoid pairs were included as fixed factors to test if effects varied depending on the  
211 focal species. All three and two-way interactions between treatments (exploitative competition and  
212 alternative host), host species, and host-parasitoid pairs were tested and kept in our models if found  
213 to be significant based on backward model selection using Likelihood-ratio tests. Models for SP  
214 were also run for each host-parasitoid pair separately to quantify differences in the sign and  
215 magnitude of the effects of community structure on pairwise interaction depending on the focal  
216 species. (iii) “Community composition models” used species assemblages rather than community  
217 modules as explanatory variables (host species assemblage: 6 levels, and parasitoid species  
218 assemblage: 6 levels). The two-way interaction between host and parasitoid assemblages was  
219 always kept in the models to account for the fully-factorial design of the experiment. Models for  
220 DI were run for each host species, and models for SP were run for each host-parasitoid pair  
221 separately. Blocks (6 levels) were included in all models as a random effect. Significance of the

222 effects was tested using Wald type III analysis of deviance with Likelihood-ratio tests. Factor  
223 levels of community modules and species assemblages were compared to the reference module  
224 and species assemblages of the host-parasitoid pair in isolation by Tukey's HSD *post hoc*  
225 comparisons of all means, using the *emmeans* package (Lenth 2018).

226 To further investigate emergent effects of parasitoid diversity on host suppression, we compared  
227 the observed and expected effect magnitude of multiple parasitoid species on host suppression for  
228 each host species ( $n = 3$ ), each heterospecific parasitoid combination ( $n = 3$ ), and each replicate ( $n$   
229  $= 6$ ), following the framework proposed by Schmitz (2007). Observed values smaller than  
230 estimated translate to risk enhancement for the host with multiple parasitoid species, while  
231 observed values bigger than estimated reflect risk reduction. Detailed methods and results are  
232 presented in Appendix S4. All analyses were performed using R 4.0.2 (Team 2017).

## 233 **Results**

### 234 *Effects of community structure on host suppression*

235 The presence of multiple parasitoid species in the module significantly increased the probability  
236 of hosts being parasitized (DI) by 48% (95% CI: 26-70%) (community module model:  $\chi^2_{(1)} = 7.08$ ,  
237  $P = 0.008$ ; Post Hoc Odds Ratio (OR) exploitative competition module/pairwise interaction = 1.58,  
238  $P = 0.076$ ; OR exploitative competition and alternative host module/pairwise interaction = 1.32,  $P$   
239  $= 0.376$ ). Moreover, effects of multiple parasitoid species on host suppression were greater than  
240 expected in four host-parasitoid cases (Figure S4). However, DI did not significantly change with  
241 the presence of an alternative host species ( $\chi^2_{(1)} = 0.56$ ,  $P = 0.452$ ; OR alternative host  
242 module/pairwise interaction = 0.96,  $P = 0.984$ ), and the two-way interaction between alternative  
243 hosts and exploitative competition treatments had no significant effect (community module model:  
244  $\chi^2_{(1)} = 0.22$ ,  $P = 0.638$ ) (Figure 2a).

245 *Effects of host species and community composition on host suppression*

246 Host DI did not differ significantly across host species (species-specific community module  
247 model:  $\chi^2_{(2)} = 0.07$ ,  $P = 0.965$ ). The directionality of the effect of parasitoid diversity did not vary  
248 depending on species assemblages (community composition models; Figure S5 and Appendix S5).

249 *Effects of community structure on parasitoid performance*

250 Community modules had no general effects of successful parasitism rates (SP) (community  
251 module model; Figure 2b), but the effects significantly varied across host-parasitoid pairs (species-  
252 specific community module model; three-way interaction:  $\chi^2_{(8)} = 36.81$ ,  $P < 0.0001$ ; Table 1).  
253 The interaction between exploitative competition and alternative host treatments had a significant  
254 effect on SP for one out of the nine host-parasitoid pairs (*Ganaspis sp.* on *D. simulans*). Successful  
255 parasitism rate of two other host-parasitoid pairs significantly decreased with exploitative  
256 competition between parasitoid species (*Ganaspis sp.* on *D. birchii* and on *D. pallidifrons*), while  
257 SP significantly increased with the presence of an alternative host species in *Asobara sp.* on *D.*  
258 *simulans*. Successful parasitism rate for the rest of the host-parasitoid pairs did not significantly  
259 change between community modules when compared to the host-parasitoid pair in isolation (Table  
260 1).

261 *Effects of community composition on parasitoid performance*

262 Effects of an alternative host and a parasitoid competitor on parasitoid performance varied  
263 depending on co-occurring species identity, both in terms of magnitude and direction of their  
264 response (community composition models). The interaction between host and parasitoid species  
265 assemblages had a significant effect on SP for four out of the nine host-parasitoid pairs: *Asobara*  
266 *sp.* on *D. simulans*, *Leptopilina sp.* on *D. birchii*, and *Ganaspis sp.* on *D. birchii* and on *D.*

267 *simulans*. Effects of species assemblages on SP for each host-parasitoid pair are summarized in  
268 Table 2 and Figure 3, and presented in Appendix S6.

## 269 **Discussion**

270 Our results confirm some general effects of community structure on consumer-resource  
271 interactions over a number of species combinations, but also reveal important species-specific  
272 effects linked to the identity of species composing the community. Specifically: (i) the presence  
273 of multiple parasitoid species consistently increased host suppression due to a combination of  
274 sampling effect and resource partitioning between parasitoid species. On the contrary, (ii) the  
275 presence of an alternative host had no general effect on host suppression, but increased or  
276 decreased successful parasitism rate depending on host-parasitoid pairs and co-occurring species  
277 identity. These species-specific effects on parasitoid performance have important implications for  
278 biological control programs that often overlook effects of co-occurring species on biological  
279 control agent efficiency to suppress pests.

### 280 *Positive effects of consumer diversity on top-down control*

281 The presence of multiple parasitoid species generally increased host suppression (community  
282 module model). We found a greater top-down control than expected with multiple parasitoid  
283 species for four out of the nine species combinations, suggesting weaker interference between  
284 parasitoids heterospecific than between conspecific parasitoids (Snyder et al. 2008, Snyder and  
285 Tylianakis 2012). Weak predator interference can be obtained if predators show some degree of  
286 resource partitioning (Pedersen and Mills 2004), which have been reported in several experimental  
287 studies (reviewed in Letourneau et al. 2009). Resource partitioning can be more important than  
288 diversity *per se* to explain an increase in top-down control with multiple predator species (Finke  
289 and Snyder 2008). The positive effects of parasitoid diversity on host suppression were also driven

290 by the presence of the most efficient parasitoid species for the focal host (e.g., presence of  
291 *Ganaspis sp.* for *D. simulans*: Figure S3a and Figure S5c), which could explain why there were no  
292 significant differences between observed and expected effect magnitude of multiple parasitoid  
293 species in some cases. These results match the sampling effect model suggesting that an increase  
294 in top-down control with an increase in consumer diversity is explained because of an increasing  
295 probability that a superior natural enemy species will be present in the community (Myers et al.  
296 1989). Here, both the sampling effect and resource partitioning seem to be in play to explain the  
297 increase of host suppression with increase in parasitoid diversity. Our results demonstrating the  
298 enhancement of top-down control with multiple parasitoid species across several species  
299 assemblages, in addition to the results from previous studies (reviewed in Letourneau et al. 2009),  
300 highlight the importance of preserving predator diversity for ecosystem functioning.

#### 301 *No indirect interactions among prey species*

302 Contrary to the presence of an additional parasitoid species, we did not detect any effect of an  
303 additional host species on host suppression. This suggests that host switching behaviors and any  
304 indirect interactions between host species were absent over this one generation experiment.  
305 Another empirical study failed to uncover any evidence of indirect interactions in natural host-  
306 parasitoid food webs (Kaartinen and Roslin 2013). Indirect interactions between prey, mediated  
307 by a shared natural enemy, is supposedly common in nature (e.g., Morris et al. 2001), and an  
308 important mechanism for prey species coexistence via density- and trait-mediated effects (Holt  
309 1977, Fleury et al. 2004, van Veen et al. 2005, McPeck 2019). But direct exploitative competition  
310 between host species, which was not allowed in our experiment, might have a stronger effect on  
311 host population and community structure than effects of an alternative host species (Jones et al.  
312 2009). Furthermore, frequency-dependent attack rates, in which a predator switches between two

313 prey species depending on their relative densities (Bonsall and Hassell 1999) is another important  
314 mechanism determining the strength of predator-prey interactions (van Veen et al. 2005), but was  
315 not tested in the present study. Varying host and non-host densities could also change parasitoid  
316 foraging behavior (Kehoe et al. 2016). In our study, potential effects of an alternative host (i.e.,  
317 alternative host module and both exploitative competition and alternative host module) on host-  
318 parasitoid interactions were tested over a single generation and with constant host density.  
319 Experimental studies manipulating alternative prey density over several generations are needed to  
320 detect indirect interactions between host species. However, we could expect that an experiment  
321 performed on our study system over multiple generations would have shown effects of an  
322 alternative host species on host suppression due to changes in host relative abundances. Indeed,  
323 we found that the strength of parasitism for our nine host-parasitoid pairs increased logistically  
324 with host density (i.e., type II parasitoid functional responses; Appendix S2 and Figure S2).  
325 Beyond a host density threshold, parasitoids cannot attack more hosts in a given time period. This  
326 threshold changes depending on both host and parasitoid identity (Figure S2), meaning that the  
327 number of hosts attacked for each species will change depending on host relative abundances and  
328 the identity of co-occurring species.

### 329 *Importance of community composition for consumer-resource interactions*

330 Our study was based on a particular set of interacting species, yet even the relatively small number  
331 of species used in our experiment allowed us to uncover species-specific response within a given  
332 community module. Community modules have been extensively used as a tool to study the  
333 mechanisms structuring and stabilizing complex natural communities (Bascompte and Melián  
334 2005, Rip et al. 2010), yet the effects of species identity in such studies is often ignored. Our results  
335 highlight that variation in directionality and magnitude of the effects of a particular community

336 module on host-parasitoid interactions depend on the species assemblage considered. This is  
337 particularly relevant to predicting natural top-down control, and the consequences of climate  
338 change driven shifts in community composition for ecosystem functioning. Potential effects of co-  
339 occurring species (alternative prey species and/or other predators) should be taken into  
340 consideration when developing biological control strategies, especially when choosing the  
341 biological control agent in open agricultural system. For example, presence of non-target species  
342 might weaken the biological control agent-pest interaction (Wajnberg et al. 2001). Even with no  
343 direct effects of presence of a non-target species, presence of a natural enemy to the alternative  
344 prey species could decrease the biological control agent performance, and prevent its establishment  
345 in the community. However, the differences between the number of hosts suppressed and the  
346 outcome for the parasitoid populations depending on the biotic context might be a unique feature  
347 of host-parasitoid systems. It remains to be determined if the identity of co-occurring species is as  
348 important in prey-predator systems where the death of the prey generally results in energy flow to  
349 the predators.

350 Successful parasitism rate increased in modules with a parasitoid competitor compared to the  
351 pairwise interaction in six species combinations (out of the 54 species combinations representing  
352 host-parasitoid pairs in modules with exploitative competition between parasitoid species),  
353 suggesting that some parasitoid species benefit from presence of an heterospecific. According to  
354 a recent review on interspecific interactions among parasitoids (Cusumano et al. 2016), and to the  
355 best of our knowledge, only one study showed facilitation between two parasitoid species  
356 (Poelman et al. 2014). Our case seems to be different because successful parasitism rates did not  
357 increase for both parties, which implies competitive interactions between parasitoids. Parasitoids  
358 can compete both as adults for oviposition (i.e., extrinsic competition or interference) and as larvae



359 within a host (i.e., intrinsic competition) (Harvey et al. 2013). Extrinsic competition has negative  
360 effects on parasitoid attack rates, linked to search efficiency and handling time, leading to a  
361 potential decrease in host mortality (Xu et al. 2016). In our study, modules with the pairwise  
362 interaction in isolation had two parasitoid conspecifics (i.e., substitutive design). Our results  
363 suggest that in these six cases, where parasitoid performance increased with the presence of a  
364 competitor, interspecific extrinsic competition between parasitoids was weaker than intraspecific  
365 extrinsic competition. On another hand, intrinsic competition is the result of a super- or multi-  
366 parasitism event when two parasitoids (conspecifics or heterospecific, respectively) parasitize the  
367 same host individual. It is usually detrimental for the host survival, and would explain an increase  
368 in host suppression with multiple parasitoid species despite a decrease in parasitoid performance  
369 in some cases (i.e., unsuccessful parasitism but with host suppression; e.g., Thierry et al. 2021b).

370 Our contrasting results on successful parasitism rate depending on the host-parasitoid pair and the  
371 other parasitoid species present in the community are likely due to differences in traits (e.g.,  
372 immune response of the hosts, and oviposition behavior and virulence of the parasitoids; Carton  
373 et al. 2008). For example, *D. simulans* often encapsulates parasitoid eggs, while encapsulation  
374 rates of *D. birchii* are low in our system. *Asobara sp.* rarely lay more than one egg in a same host  
375 individual but develop fast to avoid encapsulation, while super-parasitism is common for *Ganaspis*  
376 *sp.* and could be a strategy of this parasitoid to overcome host immune response. The different trait  
377 combinations and trade-offs across host and parasitoid species are likely an important mechanism  
378 driving species interactions and co-occurrences in natural communities (Wong et al. 2019).

### 379 *Conclusion*

380 Our study is the first to our knowledge to investigate the effects of community module of different  
381 species assemblages on consumer-resource interactions at both trophic levels. Studies looking at

382 interactions between predator species on single prey (e.g., Valente et al. 2019, Ortiz-Martínez et  
383 al. 2019), or studies looking at the effect of an alternative prey with single predator species (e.g.,  
384 Ishii and Shimada 2012, De Rijk et al. 2016), overlook important mechanisms present in nature  
385 where a number of species co-occur. With current global changes such as climate warming, the  
386 structure and composition of communities is changing, either via direct effects on species  
387 performance (Thierry et al. 2021a), via effects on interactions (Hance et al. 2007, Thierry et al.  
388 2019), or due to shift in ranges and phenology, and an increase in invasive species. It is therefore  
389 imperative to take into account the context in which species interact, both abiotic and biotic, as all  
390 those factors are likely to act together in influencing the outcome of the interaction between focal  
391 species (Gilman et al. 2010).

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398 anonymous reviewer for comments that improved the manuscript.

## 399 **Author contributions**

400 MT conceived the project; NP and JH contributed to the experimental design; MT, NP, MG, and  
401 GP collected the data; MT analyzed the data. All authors contributed critically to the draft and  
402 gave final approval for publication.

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569

570 **Table 1.** Odds ratios of a successful parasitism in each community module (exploitative competition, alternative host, and both  
571 exploitative competition and alternative host) compared to the host-parasitoid pair in isolation for each pair (host abbreviations: b: *D.*  
572 *birchii*, p: *D. pallidifrons*, s: *D. simulans*, and parasitoid abbreviations: A: *Asobara sp.*, L: *Leptopilina sp.*, G: *Ganaspis sp.*). Odds Ratios  
573 superior or inferior to 1 translate an increased or a decreased probability of having successful parasitism, respectively. Results come  
574 from the species-specific community module models run for each host-parasitoid pair separately. Odds Ratios are presented with the  
575 significance of the effect: (\*\*\*) P < 0.001, (\*\*) P < 0.01, (\*) P < 0.05, (ns) P > 0.05.

576

Module	b-A		p-A		s-A		b-L		p-L		s-L		b-G		p-G		s-G	
<b>Expl. comp.</b>	1.49	(ns)	0.63	(ns)	1.56	(ns)	0.85	(ns)	1.36	(ns)	0.00	(ns)	0.30	*	0.19	*	0.67	(ns)
<b>App. comp.</b>	0.95	(ns)	1.03	(ns)	4.29	***	1.16	(ns)	1.65	(ns)	0.00	(ns)	0.94	(ns)	0.92	(ns)	0.83	(ns)
<b>Expl. + app. comp.</b>	0.65	(ns)	0.38	(ns)	0.76	(ns)	1.44	(ns)	1.82	(ns)	0.00	(ns)	0.73	(ns)	0.45	(ns)	4.12	***
<b>Df residuals</b>	65		65		65		65		65		65		64		64		65	

577

578

579

580 **Table 2.** Effects of community composition on the probability of successful parasitism for each host-parasitoid pair. Effects are shown  
581 by the summary of Likelihood-ratio chi-square tests on the community composition models with the effects of host and parasitoid species  
582 assemblages (3 levels each) (host abbreviations: b: *D. birchii*, p: *D. pallidifrons*, s: *D. simulans*, and parasitoid abbreviations: A: *Asobara*  
583 *sp.*, L: *Leptopilina sp.*, G: *Ganaspis sp.*). For p-A and s-L models contain only host species assemblage as a fixed effect due to  
584 convergence problem with the full model. Degrees of freedom (Df) are given for each factor and for the residuals.  $\chi^2$  values are presented  
585 with the significance of the effect: (\*\*\*)  $P < 0.001$ , (\*\*)  $P < 0.01$ , (\*)  $P < 0.05$ , (ns)  $P > 0.05$ .

586

Effects	Df	b-A	p-A	s-A	b-L	p-L	s-L	b-G	p-G	s-G
Host sp.	2	11.14 **	1.12 (ns)	15.56 ***	4.80 (ns)	4.83 (ns)	10.08 **	34.14 ***	2.23 (ns)	19.71 ***
Parasitoid sp.	2	6.63 *	-	0.11 (ns)	38.32 ***	4.36 (ns)	-	2.57 (ns)	1.73 (ns)	13.81 **
Host x para	4	7.80 (ns)	-	24.19 ***	40.57 ***	7.81 (ns)	-	25.22 ***	9.01 (ns)	20.51 ***
Df residuals	60		66	60	60	60	66	59	59	60

587

588 **Figure legends**

589

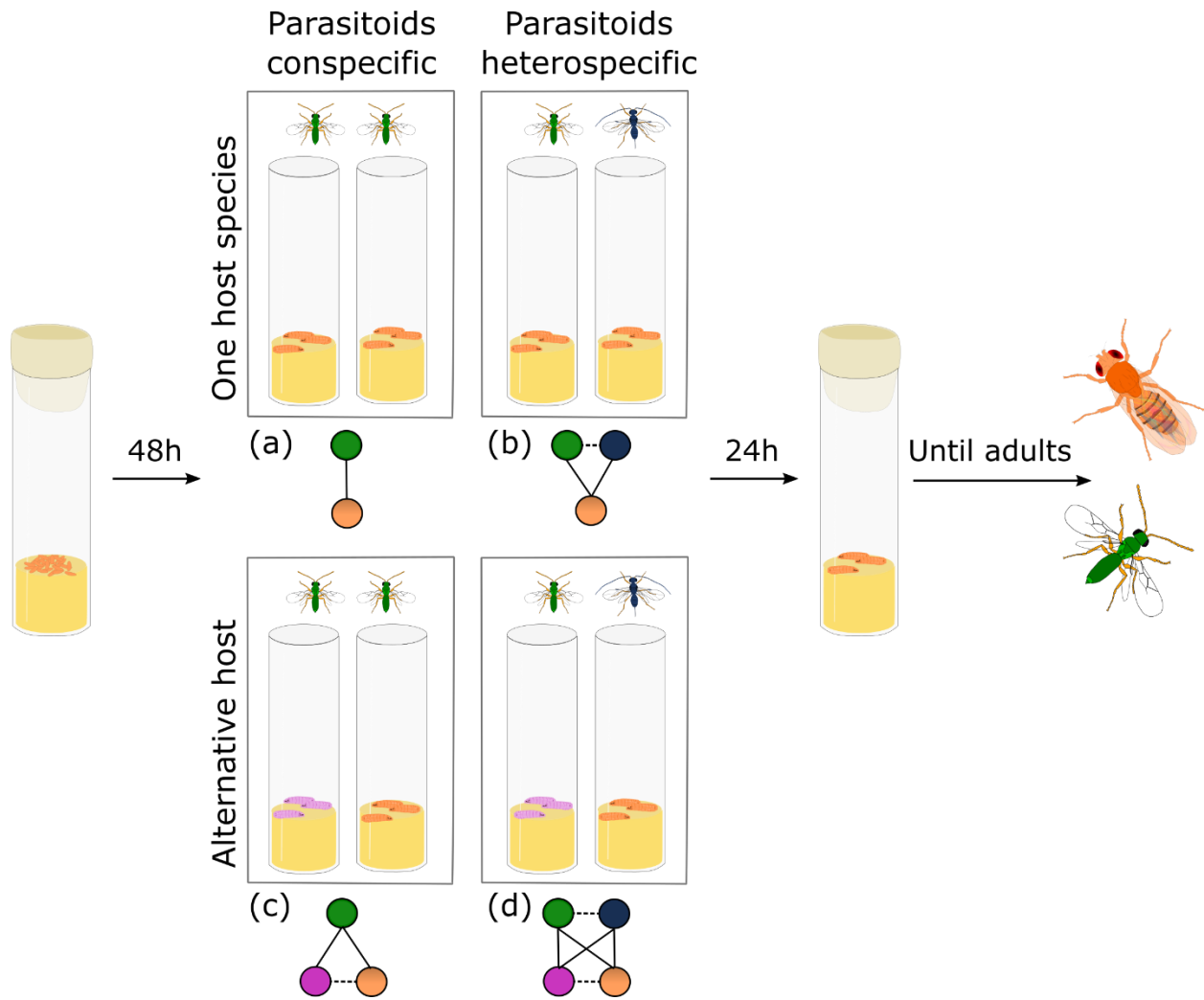
590 **Figure 1.** Schematic representation of the experimental treatments with the potential direct and  
591 indirect interactions in each community module. Orange and pink nodes and larvae represent  
592 different *Drosophila* host species, and green and blue nodes and wasps represent different  
593 parasitoid species, assembled in a fully factorial design in four different community modules  
594 represented schematically below their corresponding experimental box: a) host-parasitoid pair  
595 (one host and one parasitoid species), b) exploitative competition module (one host and two  
596 parasitoid species), c) alternative host module (two host and one parasitoid species) and, d) both  
597 exploitative competition and alternative host module (two host and two parasitoid species). In the  
598 community module schemas, solid lines represent trophic interactions, and dashed lines represent  
599 non-trophic interactions (in b) either exploitative competition, interference, or facilitation between  
600 parasitoids, c) indirect interactions between hosts, and d) potential for all the above). Direct  
601 interaction between host species were not allowed. See (Thierry et al. 2019)(Thierry et al.  
602 2019)(Thierry et al. 2019)(Thierry et al. 2019)(Thierry et al. 2019) for a  
603 detailed description of each interaction type

604

605 **Figure 2.** Effects of community structure (represented by the community module bellow each  
606 treatment) on (a) degree of infestation and on (b) successful parasitism rate. Different capital letters  
607 denote significant differences between community structure from the community module models.  
608 The small points represent the observed values, and the large points represent the predicted values  
609 with their 95% confidence intervals

610

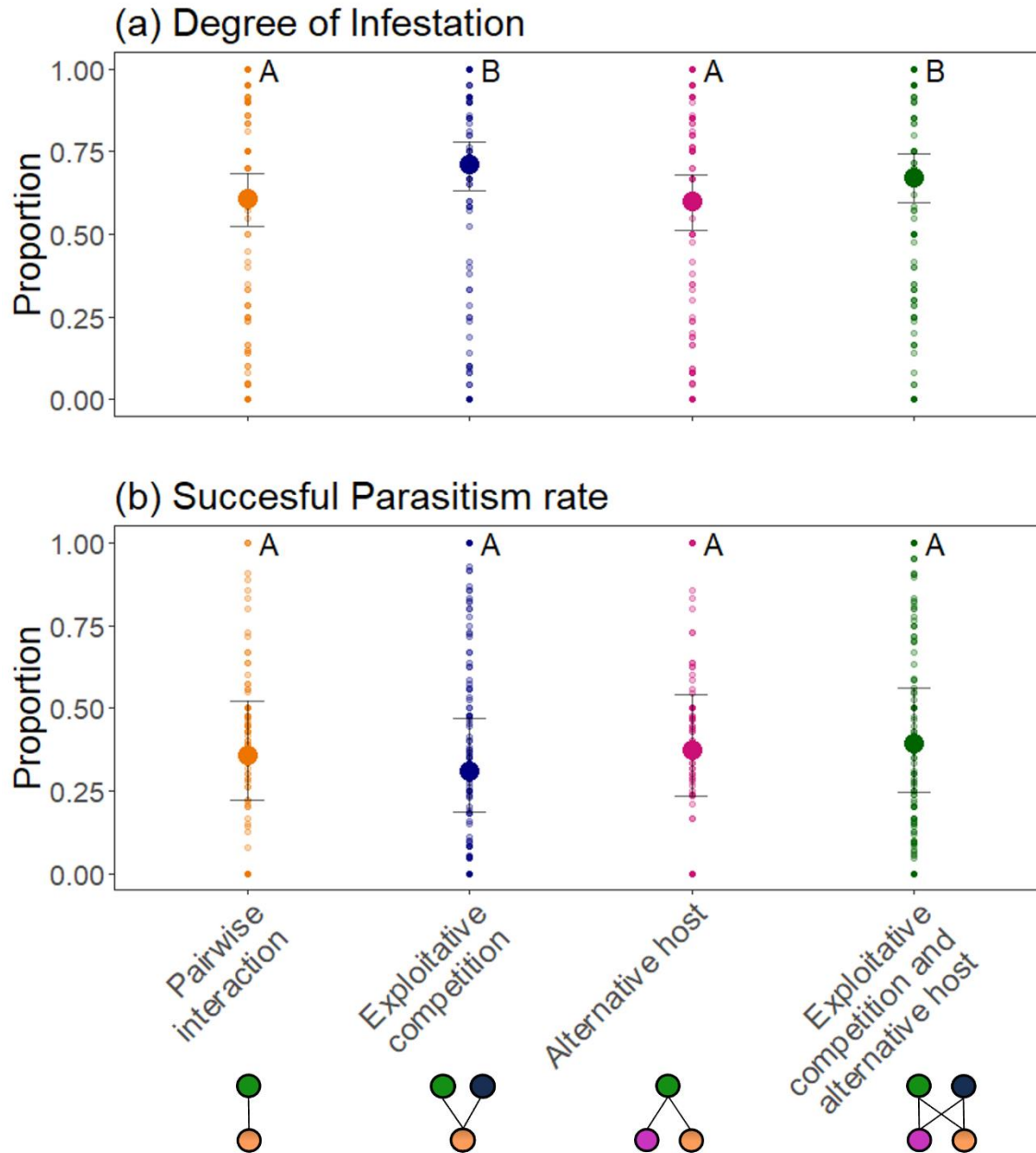
611 **Figure 3.** Effects of community composition (identity of the alternative host and the parasitoid  
612 heterospecific) on the successful parasitism rate of each parasitoid species on each host species  
613 [rows are host species (host abbreviations: b: *D. birchii*, p: *D. pallidifrons*, s: *D. simulans*), and  
614 columns are parasitoid species (parasitoid abbreviations: A: *Asobara sp.*, L: *Leptopilina sp.*, G:  
615 *Ganaspis sp.*)]. Host assemblages are represented by the different colors, and parasitoid  
616 assemblages are on the x axis. For SP p-A and SP s-L, only effect of host assemblages was analysed  
617 due to convergence problem with the full model, and are represented for all parasitoid assemblages  
618 combined. The small points represent the observed values, and the large points represent the  
619 predicted values with their 95% confidence intervals from the community composition models



620

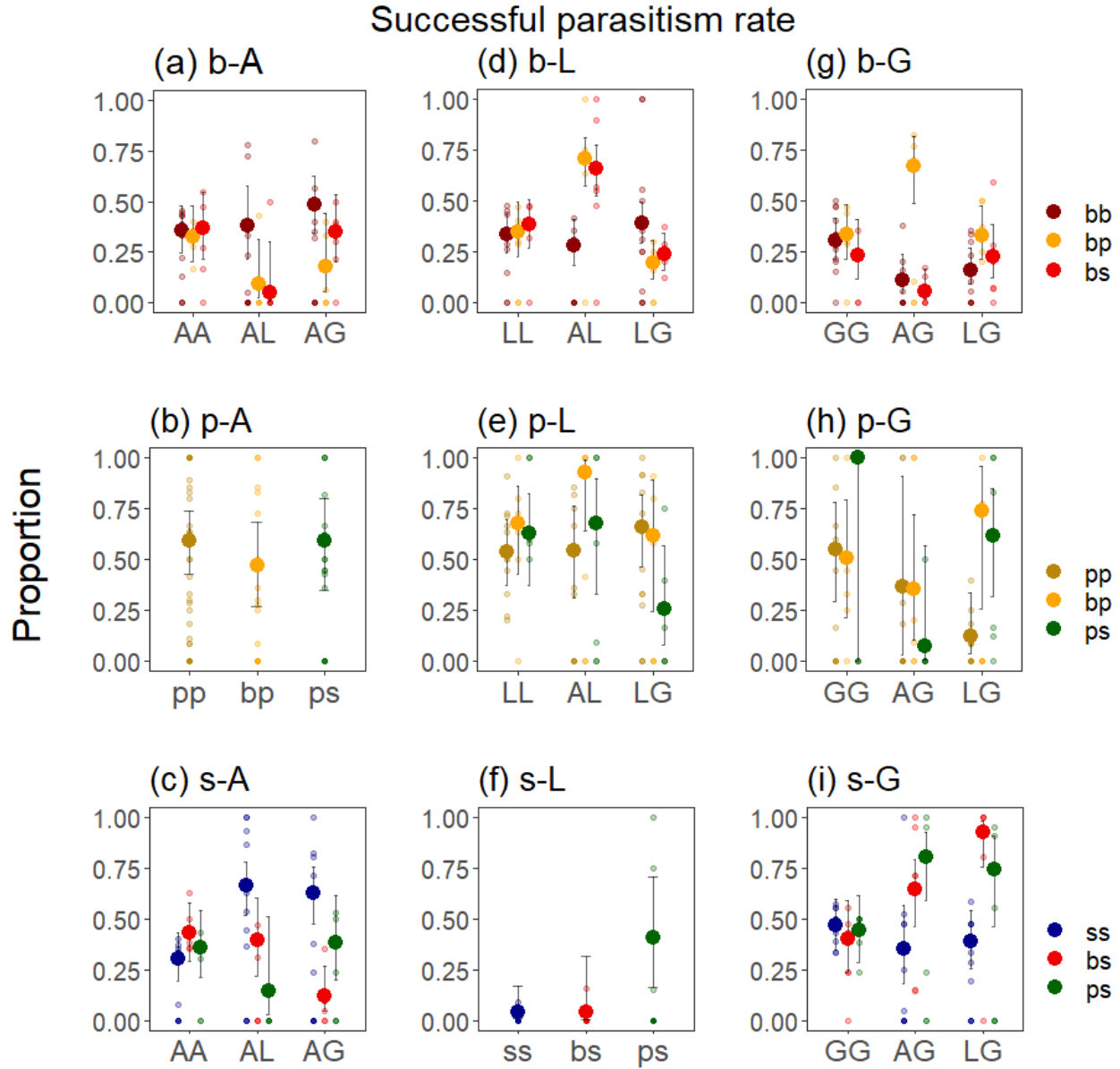
621 Figure 1.





622

623 Figure 2.



624

625 Figure 3.