Interactive effects of herbivory and the level and fluctuations of nutrient availability on

dominance of alien plants in synthetic native communities

Yanjun Li[1,2], Yingzhi Gao[2], Mark van Kleunen[3,4], Yanjie Liu[1,3*]

[1]Key Laboratory of Wetland Ecology and Environment, Northeast Institute of Geography and

Agroecology, Chinese Academy of Sciences, Changchun, 130102, China

[2]Key Laboratory of Vegetation Ecology, Northeast Normal University, Changchun, 130024,

China

[3]Ecology, Department of Biology, University of Konstanz, 78464 Konstanz, Germany

[4]Zhejiang Provincial Key Laboratory of Plant Evolutionary Ecology and Conservation, Taizhou

University, Taizhou 318000, China

*Email of corresponding author: liuyanjie@iga.ac.cn +86 431 82536096

Running title: Herbivory mediates nutrient-change effects on plant invasion

Abstract

Numerous studies have tested the fluctuating resource hypothesis and highlighted the role of

nutrient availability and fluctuations therein for invasion success of alien plants. Others have

tested the enemy-release hypothesis and highlighted the role of herbivores in invasion success.

However, how herbivory and the level and fluctuations in nutrient availability interact in driving

alien plant invasion into resident communities remains largely unexplored.

We grew eight invasive alien species as target species in pot⬜mesocosms with five

different synthetic native communities in a three-factorial design with two levels of nutrient

availability (low *vs* high), two levels of nutrient fluctuations (constant *vs* pulsed) and two levels

of herbivory (with *vs* without).

Increases in nutrient availability, irrespective of whether it was constant or pulsed,

significantly increased the absolute biomass production of the alien and native plants. The

presence of herbivores, on the other hand, significantly decreased the biomass production of

plants, in particular under high-nutrient availability. The relative biomass production of the alien

target plants, on the other hand, decreased in response to an increase in nutrient availability, and

increased in response to the presence of herbivores. We also found tentative evidence that

herbivory could interact with changes in nutrient availability and nutrient fluctuations to affect

the dominance of the alien target species (a marginally significant interaction; 90% CIs: [0.133,

2.745]).

Our multispecies experiment indicates that herbivory could mediate the interactive effect of

nutrient enrichment and variability in nutrient supply on invasion of alien plants into native

communities. Therefore, we recommend that studies testing the fluctuating resources hypothesis

should also consider interactive effect of other trophic levels.

Keywords: biological invasions, exotic plants, global change, native community, plant-herbivore

interactions, resource variability, trophic level

INTRODUCTION

Invasion by alien plants could reduce native biodiversity, influence ecosystem functions,

and degrade ecosystem services (Vilà et al. 2011). Due to the rapid globalization, the increase in

the number of naturalized alien plant species does not show any sign of saturation, and it was

recently predicted that their numbers may increase on average by 18% from 2005 to 2050

(Seebens et al. 2017 and 2021, Gippet and Bertelsmeier 2021). Therefore, a major research

objective in the field of ecology is to identify the mechanisms that underlie alien plant invasion

(Catford et al. 2009, Schultheis et al. 2015, Liu and van Kleunen 2019, Enders et al. 2020,

Speißer et al. 2021).

It has frequently been suggested that an increase in soil-nutrient availability is one of the

most important drivers of alien plant invasion (Davis et al. 2000, Bradley et al. 2010, Enders et al.

2020). As successful alien plant species are often introduced from anthropogenic and more

nutrient-rich environments, they might be more likely to be adapted to high-nutrient

environments (Dostál et al. 2013, Kalusová et al. 2017). Indeed, both a recent meta-analysis (Liu

et al. 2017) and a globally replicated study of 64 grasslands (Seabloom et al. 2015) showed that

successful alien plants respond more strongly to nitrogen enrichment than native plants do. So

far, empirical studies that tested how soil-nutrient changes affect alien plant invasion mainly

focused on changes in the mean nutrient level (Pyšek et al. 2012, Manea and Leishman 2015, Jia

et al. 2016, Heckman et al. 2017, Liu and van Kleunen 2017, Sardans et al. 2017a, Kelso et al.

2020). However, due to increasing occurrences of extreme events (e.g. droughts, floodings, heat

waves, fires), soil-nutrient changes also entail changes in their variability, and this may affect

plant invasions (Parepa et al. 2013, Liu and van Kleunen 2017). Therefore, it is important to test

how changes in soil-nutrient levels, as well as fluctuations in nutrient availability over time,

drive alien plant invasion in resident communities.

The fluctuating resource hypothesis proposes that temporal fluctuations in nutrient supply

could promote alien plant invasion in resident communities (Davis et al. 2000). However,

empirical studies testing the hypothesis found mixed results. For example, Parepa et al. (2013)

found that a pulsed nutrient supply, compared to a constant nutrient supply, increased the

dominance of the invasive plants *Fallopia japonica* and *F. × bohemica* in experimental plant

communities. In contrast, Liu et al. (2018) showed that a pulsed nutrient supply decreased the

dominance of invasive alien plants. Thus, more studies are needed to test the hypothesis, and

why the results might vary. The studies that tested the effect of temporal changes in nutrient

availability did so under overall nutrient-rich conditions. Under more nutrient-limiting conditions,

however, the effect of temporal fluctuations may be even stronger. In other words, mean nutrient

availability may interact with temporal fluctuations in nutrient availability to affect alien plant

invasion into resident communities. However, very few studies have tested whether this

expectation holds (but see Gao et al, 2021).

Although the fluctuating resource hypothesis has become a key theory in invasion ecology,

previous tests only used study systems consisting of a single trophic level (i.e., only considered

plant-plant interactions). Plant growth, however, can be strongly regulated by other trophic levels,

such as herbivores. This might be relevant for the fluctuating resource hypothesis as alien species

are likely to be released from most of their native enemies, and thus should suffer less herbivory

than native species in their introduced regions (Keane and Crawley 2002, Blumenthal 2006,

Zhang et al. 2018). Following this logic, the presence of herbivores, just like increases in

resource availability and fluctuations therein (Dawson et al. 2012, Parepa et al. 2013, Liu and

van Kleunen 2017, Pearson et al. 2018), could promote alien plant invasion in resident

communities. Moreover, as herbivore effects on plants are often regulated by soil-nutrient

availability, because plants growing in relatively high-nutrient conditions might be better able to

compensate or tolerate herbivory (Meyer 2000, Hawkes and Sullivan 2001, Gianoli and

Salgado-Luarte 2017, Hu and Dong 2019), the effects of herbivores might interact with the

effects of nutrients. Therefore, it is reasonable to assume that the presence of herbivores might

amplify the positive effect of increases in resource availability and fluctuations on alien plant

invasion.

To test the individual effects of nutrient availability, nutrient fluctuations, herbivory, and

their interactions on alien plant invasion into resident communities, we grew eight invasive alien

species as target species in pot☐mesocosms with five different synthetic native communities,

each consisting of three grassland species. Then, we exposed the plants to eight combinations of

two nutrient availability (low *vs* high), two nutrient-fluctuation (constant *vs* pulsed) and two

herbivory (with *vs* without) treatments. By comparing the absolute aboveground biomass

production of the alien target species as well as their biomass production relative to the biomass

production of the native competitors, we addressed the following specific questions: 1) Do

nutrient availability, nutrient fluctuations and the presence of herbivores promote the absolute

and relative biomass of alien plants? 2) Does the effect of nutrient fluctuations on absolute and

relative biomass of alien plants depend on the overall nutrient availability level? 3) Does the

presence of herbivores interact with increases in nutrient availability and fluctuations therein to

affect the absolute and relative biomass of alien plants?

MATERIAL AND METHODS

Study species

To investigate the individual and interactive effects of nutrient availability, nutrient fluctuations and the presence of herbivores on alien plant invasion into resident communities, we chose eight invasive alien species as targets, and 15 native species as native community members from the herbaceous flora of China (Table S1). We classified these species as invasive alien or native to China based on information in the book "The Checklist of the Alien Invasive Plants in China" (Ma et al. 2018) and the Flora of China database (www.efloras.org). To cover a wide taxonomic breadth, we selected the eight alien target species from seven genera of three families. As plants with different life histories (i.e., annuals or perennials) may respond differently to nutrient availability and fluctuations (Roumet et al. 2006, Liu and van Kleunen 2017), we assured that both the alien targets and native community members included annuals and perennials (Table S1). Seeds of these species were collected in natural populations in grasslands in China or ordered from a commercial seed company (see Table S1).

To impose a herbivory treatment, we selected two aboveground insect herbivores. As natural systems usually include both generalist and specialist insect herbivores, we chose the generalist grasshopper *Stenocatantops splendens* and the specialist grasshopper *Locusta migratoria* (i.e., grass-feeder) as the shoot herbivores. The grasshoppers were acquired from a commercial insect company (Cangzhou Grasshoppers Breeding Center, China). As both grasshoppers and all plant species occur mainly in grasslands, and according to the GBIF database (www.gbif.org), overlap in their distributions, they are very likely to co-occur in nature.

Pre-cultivation and experimental setup

We conducted the experiment at the Northeast Institute of Geography and Agroecology,

Chinese Academy of Sciences (43°5'49"N, 125°24'40"E). From 7 May to 26 June 2020, we

sowed each of the invasive alien and native species separately into plastic circular trays

(diameter = 25.5 cm, height = 4 cm) filled with potting soil (Pindstrup Plus, Pindstrup Mosebrug

A/S, 103 Denmark; pH 6; 120 mg/L N; 12 mg/L P; 400 mg/L K; 28 mg/L Mg; 0.4 mg/L B; 2

mg/L Mo; 1042 mg/L Cu; 3 mg/L Mn; 0.9 mg/L Zn; 8 mg/L Fe). Because the time required for

germination varies among the species, we sowed them at different times (Table S1) to ensure that

at transplanting the seedlings were in a similar developmental stage. All trays with seeds were

kept in a greenhouse (temperature: 18-27°C; natural lighting with an intensity of ~75% of the

light outdoors; and ~68% relative humidity).

On 10 July 2020, we selected similar-sized seedlings from each of the eight invasive alien

and 15 native species, and transplanted them into 2.5☐L circular plastic pots (top diameter: 18.5

cm, bottom diameter: 12.5 cm, height: 15 cm) filled with a 1:1 mixture of sand and fine

vermiculite. We transplanted one seedling of an alien target species in the center of each pot. For

each of the eight alien target species, we transplanted a total of 40 seedlings into 40 pots (i.e.,

one individual per pot), resulting in total of 320 pots. Immediately after transplanting the alien

target species, we equally distributed the 40 pots of each alien target over five different native

communities (i.e., eight pots per native community). To create the five different native grassland

communities, we randomly assigned the 15 native species into five groups of three species

(Table S1). We planted two seedlings of each native community member, so that each pot

included six individuals of native species equally spaced in a circle (diameter = 10 cm) around

the alien target seedling. The two individuals of the same species were planted at opposite

positions of the circle (Fig. 1a). After transplanting, we randomly assigned all pots to two cages

(3.5 × 4.5 × 2.5 m) located outside of the pre-cultivation greenhouse. The cages were covered

with transparent plastic roofs and white nylon net (mesh size: 0.25 × 0.25 mm) all around. Half

of the pots of each combination of an alien target species and a specific native community were

assigned to one cage, and the remaining ones to the other cage (Fig. 1a). In other words, each

cage included 160 pots in total (8 alien target species × 5 native communities × 4 nutrient-supply

treatments [2 nutrient availability × 2 nutrient fluctuation treatments]). To avoid the loss of water

and nutrient solution, we put a plastic tray under each pot. We re-randomized the positions of the

pots in each cage after five weeks.

On 14 July 2020 (i.e., one week after transplanting), we started to apply the nutrient

treatments at weekly intervals for a total of 10 weeks (Fig. 1b). We applied two nutrient

availability treatments (low *vs* high) crossed with two nutrient-fluctuation treatments (constant *vs*

pulsed), using a 400% □strength Hoagland solution (Methods S1). During the experiment, we

added a total of 50 and 400 ml of the Hoagland solution to the low and high-nutrient availability

pots, respectively. Although it is difficult to compare the absolute nutrient levels in a pot

experiment to those found in a natural system, Liu et al. (2018) previously showed that both the

low- and high-nutrient availabilities are limiting plant growth, as is usually the case in nature.

Within each nutrient-availability treatment, we created two different nutrient-supply patterns, a

constant nutrient supply and a pulsed nutrient supply (Fig. 1b). We supplied 5 ml of the nutrient

solution for the constant low-nutrient supply each week, and 40 ml of the same nutrient solution

for the constant high-nutrient supply. The pulsed treatment at low-nutrient availability consisted

of three weeks of 2 ml per week, followed by four weeks of 9.5 ml per week, and again three

weeks of 2 ml per week (Fig. 1b). The pulsed treatment at high-nutrient availability consisted of

three weeks of 10 ml a week, followed by four weeks of 85 ml, and again three weeks of 10 ml

(Fig. 1b). To avoid differences in water supply among the four treatments, we added extra water

to the amount of nutrient solution in each treatment to ensure that each pot received a total of 85

ml of water per nutrient application. In each cage, there were five replicates per alien target

species for each of the four nutrient-supply treatments (i.e., one replicate for each of the five

native communities). We watered all plants regularly by filling the dish under each pot to avoid

water limitation.

The herbivory treatment started on 14 August 2020, and ended six weeks later on 20

September 2020. We added the grasshoppers in one of the two cages, and treated the other cage

as control (Fig. 1a). As the commercial company hatched these two species of grasshoppers at

different times, we added *Locusta migratoria* from 14 August to 6 September 2020, and

Stenocatantops splendens from 31 August to 6 September 2020. We checked the herbivory

pressure every day to determine whether we would add more grasshoppers. During the

experiment, we added *Locusta migratoria* seven times (three times for 3–4th instars and four

times for adults), and *Stenocatantops splendens* two times (one time for 3–4th instars and the

other time for adults). In total, we added 354 individuals of *Locusta migratoria*, and 450

individuals of *Stenocatantops splendens* in the herbivory-treatment cage.

Measurements

On 25 September 2020 (i.e., eleven weeks after transplanting), we harvested the

aboveground biomass of all pots. For each pot, we first harvested the alien target species and

then the three native competitor species. As seven alien target plants died and three pots had

accidentally been treated with the wrong nutrient solution, we only harvested 310 instead of 320

pots (Table S1). We did not harvest the roots of the alien and native species, because the roots of

the species were intertwined, and it was impossible to separate them. All aboveground biomass

samples were dried at 65°C for 72 hours and then weighed. We calculated total aboveground

biomass per pot by summing the biomass of the alien target species and the three native

competitors. We also calculated the biomass proportion of the alien target species in each pot as

the ratio between the biomass of the alien target species and the total aboveground biomass per

pot.

Statistical analysis

To test the effects of nutrient availability (low *vs* high), nutrient fluctuations (constant *vs*

pulsed), herbivory treatments (with *vs* without) and their interactions on aboveground biomass

production of the alien target species, biomass production of the native communities and biomass

proportion of the alien target species in each pot, we fitted Bayesian multilevel models using the

function brm of the R package brms (Burkner, 2017) in R 4.0.2 (R Core Team, 2020). In all

models, we included the following explanatory variables nutrient availability (low *vs* high),

nutrient fluctuations (constant *vs* pulsed) and herbivory treatment (with *vs* without) as fixed

factors. To account for phylogenetic non-independence of species belonging to the same family

and for non-independence of replicates of the same species, we included identity of the alien

target species nested in their family as random factors in all models. To account for variation

among the five different native communities, we also included identity of the native community

as random factor in all models. To relax the homogeneity of variance assumption in all models,

we allowed the residual standard deviation sigma to vary by the identity of alien target species

(Zuur et al. 2009).

For all models, we used the default priors set by the brms package, and ran four independent

chains. The number of total iterations per chain was 5000, and the number of warm-up samples

was 2500. To directly test hypotheses about the main effects and interactive effects based on

each coefficient's posterior distribution, we used the sum coding, which effectively 'centers' the

effects to the grand mean (i.e. the mean value of all data observations) (Schad et al., 2020). To

implement this in brms, we used the functions contrasts and contr.sum of the stats package in R.

We considered the fixed effects nutrient availability, nutrient fluctuation and herbivory

treatments, and their interactions as significant when their 95% credible interval of the posterior

distribution did not overlap zero, and as marginally significant when the 90% credible interval

did not overlap zero. As we had only two cages available, one with herbivores and one without

herbivores, our herbivory treatment is obviously pseudoreplicated (Hurlbert 1984). This means

that the main effect of herbivory should be interpreted with care (Colegrave and Ruxton 2018).

However, it still allows us to test whether the effects of the nutrient availability and nutrient

fluctuation treatments differ between the cage with and without herbivores.

RESULTS

Averaged across the nutrient fluctuation and herbivory treatments, an increase in nutrient availability significantly increased the biomass production of alien target species (+591.6%; Table 1; Fig. 2 and Fig. S1) and the biomass production of native communities (+609.1%; Table 1; Fig. 2 and Fig. S1). The presence of herbivores, however, significantly decreased the biomass production of alien target species (-44.0%; Table 1; Fig. 2 and Fig. S2) and the biomass production of native communities (-40.2%; Table 1; Fig. 2 and Fig. S2). The negative effect of herbivory on the biomass production of both alien target species and native communities was stronger under high-nutrient availability (alien: -46.2%; native: -40.3%) than under low-nutrient availability (alien: -26.4%; native: -39.0%; significant NA × H interactions in Table 1; Fig. 2 and Fig. S3).

We found that an increase in nutrient availability significantly decreased (-6.4%; Table 1; Fig. 2 and Fig. S1), whereas the presence of herbivores significantly increased the relative biomass production of the alien target species in the native communities (+4.0%; Table 1; Fig. 2 and Fig. S2). Additionally, under low-nutrient availability, the pulsed nutrient supply tended to increase the relative biomass production of the alien target plants in the absence of herbivores (+8.0%), whereas the reverse was true for plants in the presence of herbivores (-18.8%; Fig. 2). Under low-nutrient availability, however, pulsed nutrient supply and the presence of herbivores did not affect the relative biomass production of the alien target species (Fig. 2). The dependency of the effect of the nutrient fluctuations treatment on the levels of the other treatments was indicated by a marginally significant interaction between herbivory, nutrient availability and fluctuation therein (90% CIs: [0.133, 2.745]) in Table 1.

DISCUSSION

Our multispecies experiment showed that an increase of nutrient availability promoted the biomass production for both the alien target species and native communities. However, herbivory suppressed the biomass production of plants, in particular under high-nutrient availability. We found that increases in nutrient availability suppressed, whereas herbivory promoted the dominance of the alien target species in native resident communities. Interestingly, we also found tentative evidence that herbivory could interact with changes in nutrient availability and nutrient fluctuations to affect the dominance of the alien target species. In other words, herbivory could mediate the interactive effect of nutrient enrichment and fluctuations in nutrient supply on alien plant invasion into native communities.

Although each pot had one invasive alien and six native plants, the invasive plant accounted for about one quarter of the aboveground biomass in each pot (Fig. 2c). This suggests that the invasive alien species were more dominant than the native species. This dominance decreased with an increase of the average nutrient availability, as nutrient availability promoted growth more strongly for the native communities (+609.1%) than for the alien target species (+591.6%). This finding does not support the idea that increased nutrient availability could promote alien plant invasion in resident communities, although many theoretical (Sher and Hyatt 1999, Davis et al. 2000) and empirical studies (Parepa et al. 2013, Seabloom et al. 2015, Liu et al. 2017, Liu and van Kleunen 2017, Gao et al. 2021) found evidence for this. A potenial reason could be that in our study the native species themselves are also quite common, and it are usually the rare species that take less advantage of increased nutrient availability (Dawson et al. 2012).

Previous empirical studies testing the nutrient-fluctuation effect on alien plant invasion found mixed results (Parepa et al. 2013, Dener et al. 2016, Liu and van Kleunen 2017, Liu et al.

2018, Zhao et al. 2020, Gao et al. 2021). We hypothesized that this might be because the effect

of temporal fluctuations may be even stronger under more nutrient-limiting conditions than

under less nutrient-limiting conditions. However, we found no significant interactive effect of

nutrient availability and fluctuations therein on the aboveground biomass and dominance of alien

target plants in native communities. This is in line with the results of a recent case study by Gao

et al. (2021). It is worth noting though that in the absence of herbivory, the absolute and relative

aboveground biomass of the invasive alien plants tended to be higher in the low, pulsed nutrient

treatment than in the low, constant treatment, while the reverse was true in the presence of

herbivory.

Not surprisingly, herbivory decreased the biomass production of plants (also see the total

biomass production per pot in Fig. S4a and Fig. S4b). In particular, the biomass suppression of

herbivory was stronger under high-nutrient availability than under low-nutrient availability (Fig.

S4c). This seems inconsistent with previous findings that plants compensate or tolerate herbivory

more when growing in high nutrient conditions (Meyer 2000, Hawkes and Sullivan 2001,

Alvarenga et al. 2019, Garcia and Eubanks 2019, Getman-Pickering et al. 2021, Zhong et al.

2021). However, it should be noted that in our experiment, the herbivores could choose between

the plants. As the plants grown at high-nutrient availabilities might be more nutritious (Deng et

al. 2017, Sardans et al. 2017b, Scalon et al. 2017, Oldroyd and Leyser 2020), and have decreased

plant secondary metabolite concentrations such as tannins (Li et al. 2016, Jamieson et al. 2017),

the herbivores might have preferably fed on the plants with high nutrient availabilities (Kudo

2003, Borgström et al. 2017, Sardans et al. 2017b, Anderson et al. 2018, Hernán et al. 2019). On

the other hand, we found that herbivory increased the dominance of the alien target species in the

native resident communities. This is also not surprisin, because according to the enemy release

hypothesis, alien plants often have escaped from many of their herbivorous enemies, and

therefore could in their introduced range outcompete resident plants (Keane and Crawley 2002).

When growing under low-nutrient availability and also in the absence of herbivores, the

nutrient pulse promoted the dominance of alien plants, which supports the fluctuating resource

hypothesis (Davis et al. 2000). However, when growing under low-nutrient availability but in the

presence of herbivores, the nutrient pulse suppressed the dominance of alien plants in pots. This

may because herbivores often reduce the abundance (biomass, cover) of dominant species (Olff

and Ritchie 1998, Koerner et al. 2018). On the other hand, under high-nutrient availablity, the

nutrient pulse and the herbivory treatment hardly affected the dominance of alien plants. One

plausibility explanation for this finding is that the overall high-nutrient availability reduces, or

even cancels, the nutrient-limitation shifts over time caused by nutrient fluctuations (Allington et

al. 2013, Liu and van Kleunen 2017, Liu et al. 2018), resulting in very weak effects of the

nutrient pulse on dominance changes in plant communities (Sakavara et al. 2018). In addition to

the weak evidence that herbivory mediated the effects of changes in nutrient availability and

fluctuations on alien plant invasion, a recent case study also found that the parasitic plant

Cuscuta australis could also regulate the effects of nutrient availablity and fluctuations on the

invasion success of the alien plant *Bidens pilosa* (Gao et al. 2021). Therefore, if organisms from

other trophic levels frequently mediate the effects of nutrient fluctuations on alien plant invasion,

we recommend that studies testing the fluctuating resources hypothesis should more frequently

consider the interactive effect of other trophic levels.

CONCLUSIONS

The fluctuating resource hypothesis has become a key theory for explaining invasion

success of alien plants. However, our study is, to the best of our knowledge, the first

multi-species experiment that tested how another trophic level influences the effects of nutrient

fluctuations on alien plant invasion. Partly in line with the fluctuating resource hypothesis, we

found tentative evidence that nutrient variability promotes alien plant invasion only under overall

low-nutrient conditions, and only in the absence of herbivores. Therefore, other trophic levels,

such as herbivores in our study, might mediate the interactive effect of nutrient enrichment on

alien plant invasion into resident communities.

ACKNOWLEDGEMENTS

We thank Xue Zhang, Mingxin Pan, Chunling Chang, Huifei Jin, Lichao Wang, Lingxi

Wang, Liping Shan, Han Yu for set-up of the experiment and plant harvest, and we thank

Mingxin Pan for weighing the biomass. This work was supported by funding from the Chinese

Academy of Sciences (Y9B7041001).

AUTHOR CONTRIBUTIONS

Y Liu conceived the idea and designed the experiment. Y Li and Y Gao performed the

experiment. Y Li and Y Liu analyzed the data. Y Li and Y Liu wrote the first draft of the

manuscript, with further inputs from Y Gao and M van Kleunen.

DATA AVAILABILITY

Should the manuscript be accepted, the data supporting the results will be archived in Dryad

and the data DOI will be included at the end of the article.

LITERATURE CITED

Alvarenga, R., A. M. Auad, J. C. Moraes, S. E. B. da Silva, and B. S. Rodrigues. 2019. Tolerance to nymphs and adults of *Mahanarva spectabilis* (Hemiptera: Cercopidae) by forage plants in fertilized soils. Pest Management Science 75:2242-2250.

Allington, G. R. H., D. N. Koons, S. K. Morgan Ernest, M. R. Schutzenhofer, and T. J. Valone. 2013. Niche opportunities and invasion dynamics in a desert annual community. Ecology Letters 16:158-166.

Anderson, T. M., et al. 2018. Herbivory and eutrophication mediate grassland plant nutrient responses across a global climatic gradient. Ecology 99:822-831.

Blumenthal, D. M. 2006. Interactions between resource availability and enemy release in plant invasion. Ecology Letters 9:887-895.

Borgström, P., J. Strengbom, L. Marini, M. Viketoft, and R. Bommarco. 2017. Above□ and belowground insect herbivory modifies the response of a grassland plant community to nitrogen eutrophication. Ecology 98:545-554.

Bradley, B. A., D. M. Blumenthal, D. S. Wilcove, and L. H. Ziska. 2010. Predicting plant invasions in an era of global change. Trends in Ecology & Evolution 25:310-318.

Burkner, P. C. 2017. brms: An R package for bayesian multilevel models using stan. Journal of Statistical Software 80:1-28.

Catford, J. A., R. Jansson, and C. Nilsson. 2009. Reducing redundancy in invasion ecology by integrating hypotheses into a single theoretical framework. Diversity and Distributions

15:22-40.

Colegrave, N., and G. D. Ruxton. 2018. Using biological insight and pragmatism when thinking

about pseudoreplication. Trends in Ecology & Evolution 33:28-35.

Davis, M. A., J. P. Grime, and K. Thompson. 2000. Fluctuating resources in plant communities: a

general theory of invasibility. Journal of Ecology 88:528-534.

Dawson, W., R. P. Rohr, M. van Kleunen, and M. Fischer. 2012. Alien plant species with a wider

global distribution are better able to capitalize on increased resource availability. New

Phytologist 194:859-867.

Dawson, W., M. Fischer, and M. van Kleunen. 2012. Common and rare plant species respond

differently to fertilisation and competition, whether they are alien or native. Ecology

Letters 15:873-880.

Dener, E., A. Kacelnik, and H. Shemesh. 2016. Pea plants show risk sensitivity. Current Biology

26:1763-1767.

Deng, Q., D. F. Hui, S. Dennis, and K. C. Reddy. 2017. Responses of terrestrial ecosystem

phosphorus cycling to nitrogen addition: A meta-analysis. Global Ecology and

Biogeography 26:713-728.

Dostál, P., W. Dawson, M. van Kleunen, L. H. Keser, and M. Fischer. 2013. Central European

plant species from more productive habitats are more invasive at a global scale. Global

Ecology and Biogeography 22:64-72.

Enders, M., et al. 2020. A conceptual map of invasion biology: Integrating hypotheses into a

consensus network. Global Ecology and Biogeography 29:978-991.

Garcia, L. C., and M. D. Eubanks. 2019. Overcompensation for insect herbivory: a review and

meta□analysis of the evidence. Ecology 100:e02585.

Gao, F. L., Q. S. He, R. Q. Xie, J. H. Hou, C. L. Shi, J. M. Li, and F. H. Yu. 2021. Interactive

effects of nutrient availability, fluctuating supply, and plant parasitism on the

post-invasion success of *Bidens pilosa*. Biological Invasions

https://doi.org/10.1007/s10530-021-02555-y.

Getman□Pickering, Z. L., G. M. Stack, and J. S. Thaler. 2021. Fertilizer quantity and type alter

mycorrhizae□conferred growth and resistance to herbivores. Journal of Applied Ecology

58:931-940.

Gippet, J. M. W., and C. Bertelsmeier. 2021. Invasiveness is linked to greater commercial

success in the global pet trade. Proceedings of the National Academy of Sciences

118:e2016337118.

Gianoli, E., and C. Salgado-Luarte. 2017. Tolerance to herbivory and the resource availability

hypothesis. Biology Letters 13:20170120.

Hawkes, C. V., and J. J. Sullivan. 2001. The impact of herbivory on plants in different resource

conditions: a meta□analysis. Ecology 82:2045-2058.

Heckman, R. W., F. W. Halliday, P. A. Wilfahrt, and C. E. Mitchell. 2017. Effects of native

diversity, soil nutrients, and natural enemies on exotic invasion in experimental plant

communities. Ecology 98:1409-1418.

Hernán, G., I. Castejón, J. Terrados, and F. Tomas. 2019. Herbivory and resource availability

shift plant defense and herbivore feeding choice in a seagrass system. Oecologia

189:719-732.

Hu, X. T., and B. C. Dong. 2019. Herbivory and nitrogen availability affect performance of an

invader *Alternanthera philoxeroides* and its native congener *A. sessilis*. Flora

257:151412.

Hurlbert, S.H. 1984. Pseudoreplication and the design of ecological field experiments.

Ecological Monographs 54:187-211.

Jamieson, M. A., L. A. Burkle, J. S. Manson, J. B. Runyon, A. M. Trowbridge, and J. Zientek.

2017. Global change effects on plant–insect interactions: the role of phytochemistry.

Current Opinion in Insect Science 23:70-80.

Jia, J. J., Z. C. Dai, F. Li, and Y. J. Liu. 2016. How will global environmental changes affect the

growth of alien plants? Frontiers in Plant Science 7:1623.

Kalusová, V., et al. 2017. Naturalization of European plants on other continents: the role of

donor habitats. Proceedings of the National Academy of Sciences 114:13756-13761.

Keane, R. M., and M. J. Crawley. 2002. Exotic plant invasions and the enemy release hypothesis.

Trends in Ecology & Evolution 17:164-170.

Kelso, M. A., R. D. Wigginton, and E. D. Grosholz. 2020. Nutrients mitigate the impacts of

extreme drought on plant invasions. Ecology 101:e02980.

Koerner, S. E., et al. 2018. Change in dominance determines herbivore effects on plant

biodiversity. Nature Ecology & Evolution 2:1925-1932.

Kudo, G. 2003. Variations in leaf traits and susceptibility to insect herbivory within a *Salix*

miyabeana population under field conditions. Plant Ecology 169:61-69.

Li, F., T. L. Dudley, B. Chen, X. Chang, L. Liang, and S. Peng. 2016. Responses of tree and

insect herbivores to elevated nitrogen inputs: A meta-analysis. Acta Oecologica

77:160-167.

Liu, Y. J., A. M. O. Oduor, Z. Zhang, A. Manea, I. M. Tooth, M. R. Leishman, X. L. Xu, and M.

van Kleunen. 2017. Do invasive alien plants benefit more from global environmental

change than native plants? Global Change Biology 23:3363-3370.

Liu, Y. J., and M. van Kleunen. 2017. Responses of common and rare aliens and natives to

nutrient availability and fluctuations. Journal of Ecology 105:1111-1122.

Liu, Y. J., and M. van Kleunen. 2019. Nitrogen acquisition of Central European herbaceous

plants that differ in their global naturalization success. Functional Ecology 33:566-575.

Liu, Y. J., X. Q. Zhang, and M. van Kleunen. 2018. Increases and fluctuations in nutrient

availability do not promote dominance of alien plants in synthetic communities of

common natives. Functional Ecology 32:2594-2604.

Manea, A., and M. R. Leishman. 2015. Competitive interactions between established grasses and

woody plant seedlings under elevated CO_2 levels are mediated by soil water availability.

Oecologia 177:499-506.

MA, J. S, and H. R. Li. 2018. The checklist of alien invasive plants in china. Higher education

press, Beijing, China.

Meyer, G. A. 2000. Interactive effects of soil fertility and herbivory on *Brassica nigra*. Oikos

88:433-441.

Oldroyd, G. E., and O. Leyser. 2020. A plant's diet, surviving in a variable nutrient environment.

Science 368:eaba019.

Olff, H., and M. E. Ritchie. 1998. Effects of herbivores on grassland plant diversity. Trends in

Ecology & Evolution 13:261-265.

Parepa, M., M. Fischer, and O. Bossdorf. 2013. Environmental variability promotes plant

invasion. Nature Communications 4:1-4.

Pearson, D. E., Y. K. Ortega, D. Villarreal, Y. Lekberg, M. C. Cock, Ö. Eren, and J. L. Hierro.

2018. The fluctuating resource hypothesis explains invasibility, but not exotic advantage

following disturbance. Ecology 99:1296-1305.

Pyšek, P., V. Jarošík, P. E. Hulme, J. Pergl, M. Hejda, U. Schaffner, and M. Vilà. 2012. A global

assessment of invasive plant impacts on resident species, communities and ecosystems:

the interaction of impact measures, invading species' traits and environment. Global

Change Biology 18:1725-1737.

Roumet, C., C. Urcelay, and S. Díaz. 2006. Suites of root traits differ between annual and

perennial species growing in the field. New Phytologist 170:357-368.

R Core Team. 2020. R: A language and environment for statistical computing. R Foundation for

Statistical Computing, Vienna, Austria.

Sardans, J., M. Bartrons, O. Margalef, A. Gargallo☐Garriga, I. A. Janssens, P. Ciais, M.

Obersteiner, B. D. Sigurdsson, H. Y. H. Chen, and J. Peñuelas. 2017a. Plant invasion is

associated with higher plant-soil nutrient concentrations in nutrient-poor environments.

Global Change Biology 23:1282-1291.

Sardans, J., O. Grau, H. Y. H. Chen, I. A. Janssens, P. Ciais, S. Piao, and J. Peñuelas. 2017b.

Changes in nutrient concentrations of leaves and roots in response to global change

factors. Global Change Biology 23:3849-3856.

Sakavara, A., G. Tsirtsis, D. L. Roelke, R. Mancy, and S. Spatharis. 2018. Lumpy species

coexistence arises robustly in fluctuating resource environments. Proceedings of the

National Academy of Sciences 115:738-743.

Scalon, M. C., M. Haridasan, and A. C. Franco. 2017. Influence of long-term nutrient

manipulation on specific leaf area and leaf nutrient concentrations in savanna woody

species of contrasting leaf phenologies. Plant and Soil 421:233-244.

Schad, D. J., S. Vasishth, S. Hohenstein, and R. Kliegl. 2020. How to capitalize on a priori

contrasts in linear (mixed) models: A tutorial. Journal of Memory and Language 110:

104038.

Schultheis, E. H., A. E. Berardi, and J. A. Lau. 2015. No release for the wicked: enemy release is

dynamic and not associated with invasiveness. Ecology 96:2446-2457.

Seabloom, E. W., et al. 2015. Plant species' origin predicts dominance and response to nutrient

enrichment and herbivores in global grasslands. Nature Communications 6:1-8.

Seebens, H., et al. 2017. No saturation in the accumulation of alien species worldwide. Nature

Communications 8:1-9.

Seebens, H., et al. 2021. Projecting the continental accumulation of alien species through to 2050.

Global Change Biology: 27:970-982.

Sher, A. A., and L. A. Hyatt. 1999. The disturbed resource-flux invasion matrix: a new

framework for patterns of plant invasion. Biological Invasions 1:107-114.

Speißer, B., Y. J. Liu, and M. van Kleunen. 2021. Biomass responses of widely and less☐widely

naturalized alien plants to artificial light at night. Journal of Ecology 109:1819-1827.

Vilà, M., J. L. Espinar, M. Hejda, P. E. Hulme, V. Jarošík, J. L. Maron, J. Pergl, U. Schaffner, Y.

Sun, and P. Pyšek. 2011. Ecological impacts of invasive alien plants: a meta-analysis of

their effects on species, communities and ecosystems. Ecology Letters 14:702-708.

Zhang, Z. J., X. Y. Pan, D. Blumenthal, M. van Kleunen, M. Liu, and B. Li. 2018. Contrasting

effects of specialist and generalist herbivores on resistance evolution in invasive plants.

Ecology 99:866-875.

Zhao, C. Y., Y. Y. Liu, X. P. Shi, and Y. J. Wang. 2020. Effects of soil nutrient variability and

competitor identify on growth and co-existence among invasive alien and native clonal

plants. Environmental Pollution 261:113894.

Zhong, Z. W., X. F. Li, D. Sanders, Y. M. Liu, L. Wang, Y. K. Ortega, D. E. Pearson, and D. L.

Wang. 2021. Soil engineering by ants facilitates plant compensation for large herbivore

removal of aboveground biomass. Ecology:e03312.

Zuur, A., E. N. Ieno, N. Walker, A. A. Saveliev, and G. M. Smith. 2009. Mixed effects

models and extensions in ecology with R. Springer, New York, USA.

 Table 1 Output of the models testing effects of nutrient availability (low *vs* high), nutrient fluctuations (constant *vs* pulsed), herbivory

516   treatments (with *vs* without), and their interactions on aboveground biomass production of the alien target species, biomass production

517   of the native communities and biomass proportion of the alien target species in each pot. Shown are the model estimates and standard

518   errors as well as the lower (L) and upper (U) values of the 95% and 90% credible intervals (CI).

		Estimate	SE	L95% CI	U95% CI	L90% CI	U90% CI
Biomass production of alien target plants	Intercept	1.030	0.754	-0.565	2.526	-0.193	2.192
	Nutrient availability (NA)	0.438*	0.058	0.328	0.555	0.344	0.535
	Nutrient fluctuation (NF)	0.019	0.047	-0.075	0.111	-0.058	0.096
	Herbivory treatment (H)	-0.103*	0.048	-0.198	-0.008	-0.182	-0.024
	NA×H	-0.108*	0.047	-0.202	-0.017	-0.187	-0.032
	NF×H	-0.016	0.047	-0.107	0.075	-0.092	0.062
	NA×NF	0.056	0.047	-0.035	0.151	-0.020	0.133
	NA×NF×H	0.016	0.047	-0.073	0.107	-0.060	0.094
Biomass production of native communities	Intercept	3.556*	1.160	1.090	5.775	1.566	5.380
	Nutrient availability (NA)	3.231*	0.112	3.013	3.451	3.050	3.414
	Nutrient fluctuation (NF)	0.026	0.110	-0.191	0.241	-0.153	0.204
	Herbivory treatment (H)	-1.149*	0.108	-1.354	-0.938	-1.326	-0.970
	NA×H	-0.757*	0.109	-0.971	-0.545	-0.937	-0.577
	NF×H	-0.136	0.111	-0.351	0.082	-0.317	0.045
	NA×NF	-0.037	0.111	-0.252	0.180	-0.217	0.148
	NA×NF×H	-0.075	0.110	-0.291	0.142	-0.255	0.107
Biomass proportion of alien target plants	Intercept	23.611*	11.536	0.453	46.093	4.783	42.029
	Nutrient availability (NA)	-1.945*	0.778	-3.464	-0.411	-3.224	-0.684
	Nutrient fluctuation (NF)	0.621	0.781	-0.897	2.121	-0.656	1.898
	Herbivory treatment (H)	4.363*	0.859	2.627	6.035	2.916	5.762
	NA×H	-1.083	0.768	-2.608	0.413	-2.355	0.167
	NF×H	-0.591	0.763	-2.097	0.908	-1.845	0.663
	NA×NF	0.932	0.777	-0.579	2.444	-0.344	2.217
	NA×NF×H	1.436†	0.794	-0.128	3.030	0.133	2.745

519　Parameters whose 95% credible intervals do not overlap with zero are indicted with asterisks (*), and whose 90% credible intervals do

520　not overlap with zero are indicated with daggers (†).Residual standard deviations *sigma* for individual alien species are found in Table

521　S2.

 FIGURE LEGENDS

Figure 1 Graphical illustration of the experimental design. Overview of the herbivory treatment,

the herbivory-treatment cage during the experiment, and the pot design (a); the amount of

nutrient solution supplied each week during the ten weeks of the experiment (b). The constant

and pulsed nutrient supply under low- or high-nutrient availability received the same total

amount of nutrients during the ten weeks.

Figure 2 Mean values (±SE) of biomass production of alien target species (a) and native

communities (b) and biomass proportion of the alien target species (c) under each combination of

two nutrient availability (low *vs* high), two nutrient-fluctuation (constant *vs* pulsed) and two

herbivory (with *vs* without) treatments. Parameters whose 95% credible intervals do not overlap

with zero are indicted with asterisks (*), whose 90% credible intervals do not overlap with zero

are indicated with daggers (†), and whose 90% credible intervals do not overlap with zero are

indicated with "ns".

Figure 1 Graphical illustration of the experimental design. Overview of the herbivory treatment, the herbivory-treatment cage during the experiment, and the pot design (a); the amount of nutrient solution supplied each week during the ten weeks of the experiment (b). The constant and pulsed nutrient supply under low- or high-nutrient availability received the same total amount of nutrients during the ten weeks.

Figure 2 Mean values (±SE) of biomass production of alien target species (a) and native

communities (b) and biomass proportion of the alien target species (c) under each combination of

two nutrient availability (low *vs* high), two nutrient-fluctuation (constant *vs* pulsed) and two

herbivory (with *vs* without) treatments. Parameters whose 95% credible intervals do not overlap

with zero are indicted with asterisks (*), whose 90% credible intervals do not overlap with zero

are indicated with daggers (†), and whose 90% credible intervals do not overlap with zero are

indicated with "ns".

SUPPORTING INFORMATION

Methods S1

Recipe for 400% strength Hoagland's Complete Nutrient Solution

The preparation of Hoagland nutrient solution essentially followed recommendations by

Hoagland and Arnon (1950) and Liu et. al (2017), with the exception of the form in which iron

was added (see below). We prepared the stock solutions 1-6 and micronutrient-stock solution

given below, and used the amounts indicated to prepare 1 liter of nutrient solution:

(1) 4 mL of 1.00 M/L KH_2PO_4

(2) 20 mL of 1.00 M/L KNO_3

(3) 20 mL of 1.00 M/L $Ca(NO_3)_2$

(4) 8 mL of 1.00 M/L $MgSO_4$

(5) 4 ml of micronutrient stock solution (see recipe below)

(6) 10 ml of 1000 mg/liter iron from iron chelate (Fe-EDTA)

* Micronutrient-stock solution per liter:

2.86 g H_3BO_3

1.81 g $MnCl_2 \cdot 4H_2O$

0.22 g $ZnSO_4 \cdot 7H_2O$

0.08 g $CuSO_4 \cdot 5H_2O$

0.018 g H_2MoO_4 (Assaying 94% MoO_3)

* Hoagland's recipe called for 1 ml of 0.5% iron tartrate stock per liter of nutrient solution but we

used iron chelate instead.

Reference: Hoagland, D.R. & Arnon, D.I. 1950. The water-culture method of growing plants without soil. California

Agricultural Experiment Station, Circular 347.

Table S1 Detailed information on the eight alien target species and 15 native competitor species used in this study

Species	Family	Status	Life history	Seed sources	Sowing date	Wrong/dead pots
Bidens frondosa L.	Compositae	Alien	Annual	Zhejiang	26.06.2020	1[#]
Bidens pilosa L.	Compositae	Alien	Annual	Zhejiang	26.06.2020	0
Solidago canadensis L.	Compositae	Alien	Perennial	Zhejiang	26.06.2020	0
Xanthium strumarium L.	Compositae	Alien	Annual	Nei Mongol	26.06.2020	1[#]
Abutilon theophrasti Medik.	Malvaceae	Alien	Annual	Jilin	26.06.2020	1[#]
Hibiscus trionum L.	Malvaceae	Alien	Annual	Jilin	05.07.2020	3*
Lolium perenne L.	Poaceae	Alien	Perennial	Greenwood flower seed company	26.06.2020	3*
Paspalum notatum Flüggé	Poaceae	Alien	Perennial	Zhejiang	15.05.2020	1*
Catolobus pendulus (L.) Al-Shehbaz[5]	Brassicaceae	Native	Biennial	Nei Mongol	26.06.2020	-
Cynanchum chinense R. Br.[2]	Apocynaceae	Native	Annual	Nei Mongol	26.06.2020	-
Lappula myosotis V. Wolf[1]	Boraginaceae	Native	Annual or Biennial	Nei Mongol	26.06.2020	-
Artemisia rubripes Nakai[4]	Compositae	Native	Perennial	Nei Mongol	26.06.2020	-
Achillea asiatica Serg.[2]	Compositae	Native	Perennial	Nei Mongol	26.06.2020	-
Convolvulus arvensis L.[3]	Convolvulaceae	Native	Perennial	Jilin	26.06.2020	-
Gypsophila licentiana Hand.-Mazz.[4]	Caryophyllaceae	Native	Perennial	Nei Mongol	26.06.2020	-
Euphorbia humifusa Willd.[5]	Euphorbiaceae	Native	Annual	Nei Mongol	26.06.2020	-
Chloris virgata Sw.[2]	Poaceae	Native	Annual	Nei Mongol	26.06.2020	-
Cleistogenes squarrosa (Trin.) Keng[5]	Poaceae	Native	Perennial	Jilin	26.06.2020	-
Digitaria sanguinalis (L.) Scop.[1]	Poaceae	Native	Annual	Jilin	26.06.2020	-
Echinochloa crus-galli (L.) P.Beauv.[4]	Poaceae	Native	Annual	Jilin	26.06.2020	-
Setaria pumila (Poir.) Roem. & Schult.[3]	Poaceae	Native	Annual	Jilin	27.06.2020	-
Portulaca oleracea L.[3]	Portulacaceae	Native	Annual	Nei Mongol	26.06.2020	-
Rumex crispus L.[1]	Polygonacea	Native	Perennial	Nei Mongol	26.06.2020	-

576   [1,2,3,4,5]Allocation of native species to the five native communities; each group included three different native species

577   [#] Number of pots treated with wrong nutrient solution during the experiment; * Number of pots in which the alien target plant died during the experiment

578 **Table S2** Output of the models residual standard deviations sigma for individual alien species. Shown are the model estimates and

579 standard errors as well as the lower (L) and upper (U) values of the 95% and 90% credible intervals (CI).

		Estimate	SE	L95% CI	U95% CI	L90% CI	U90% CI
	Abutilon theophrasti Medik.	-0.722*	0.129	-0.962	-0.460	-0.923	-0.502
	Bidens frondose L.	1.075*	0.125	0.842	1.331	0.877	1.288
	Bidens pilosa L.	0.338*	0.122	0.116	0.591	0.148	0.546
Biomass production of alien target plants	*Hibiscus trionum* L.	-0.783*	0.171	-1.113	-0.441	-1.062	-0.496
	Lolium perenne L.	0.705*	0.134	0.451	0.981	0.490	0.933
	Paspalum notatum Flüggé	0.452*	0.121	0.225	0.700	0.261	0.658
	Solidago canadensis L.	-0.476*	0.125	-0.710	-0.223	-0.678	-0.264
	Xanthium strumarium L.	0.684*	0.126	0.456	0.949	0.488	0.901
	Abutilon theophrasti Medik.	0.374*	0.130	0.132	0.638	0.169	0.594
	Bidens frondose L.	0.594*	0.127	0.361	0.857	0.394	0.812
	Bidens pilosa L.	0.514*	0.126	0.283	0.780	0.317	0.731
Biomass production of native communities	*Hibiscus trionum* L.	0.603*	0.127	0.372	0.869	0.403	0.821
	Lolium perenne L.	0.507*	0.154	0.223	0.827	0.265	0.770
	Paspalum notatum Flüggé	0.960*	0.121	0.735	1.208	0.767	1.164
	Solidago canadensis L.	0.722*	0.117	0.505	0.966	0.538	0.922
	Xanthium strumarium L.	0.556*	0.123	0.328	0.813	0.363	0.766
	Abutilon theophrasti Medik.	2.466*	0.133	2.221	2.740	2.255	2.692
	Bidens frondose L.	2.565*	0.135	2.314	2.838	2.350	2.795
	Bidens pilosa L.	2.207*	0.128	1.968	2.471	2.003	2.427
Biomass proportion of alien target plants	*Hibiscus trionum* L.	2.627*	0.140	2.364	2.912	2.404	2.863
	Lolium perenne L.	3.001*	0.149	2.723	3.304	2.764	3.257
	Paspalum notatum Flüggé	2.913*	0.124	2.684	3.170	2.718	3.126
	Solidago canadensis L.	2.411*	0.125	2.180	2.666	2.214	2.626
	Xanthium strumarium L.	2.591*	0.127	2.353	2.850	2.390	2.805

Figure S1 Mean values (±SE) of biomass production of the alien target species (a) and the native

communities (b) and biomass proportion of the alien target species (c) under different nutrient

availability conditions.

Figure S2 Mean values (±SE) of biomass production of the alien target species (a) and the native

communities (b) and biomass proportion of the alien target species (c) growing with and without

herbivory.

Figure S3 Mean values (±SE) of biomass production of native communities (a) and alien target

species (b) under each combination of two nutrient availability (low *vs* high) and two herbivory

(with *vs* without) treatments.

Figure S4 Mean values (±SE) of total biomass per pot under each combination of two nutrient

availability (low *vs* high), two nutrient fluctuation (constant *vs* pulsed) and two herbivory (with

vs without) treatments (a), with and without herbivory (b), and under each combination of two

nutrient availability (low *vs* high) and two herbivory (with *vs* without) treatments (c). Parameters

whose 95% credible intervals do not overlap with zero are indicted with asterisks (*).