1 A review of the effectiveness of blanket curtailment strategies in reducing bat fatalities at terrestrial

- 2 wind farms in North America
- 3 Evan M. Adams¹*, Julia Gulka¹, and Kathryn A. Williams¹
- 4 ¹Biodiversity Research Institute
- 5 *Corresponding author
- 6 E-mail: evan.adams@briwildlife.org
- 7

8 Abstract

9 Blanket curtailment of turbine operations during low wind conditions has become an accepted operational minimization tactic to reduce bat mortality at terrestrial wind facilities. Site-specific studies 10 11 have demonstrated that operational curtailment effectively reduces impacts, but the exact nature of the 12 relationship between increased cut-in speed and fatality reduction in bats remains unclear. To evaluate the 13 efficacy of differing blanket curtailment regimes in reducing bat fatality, we examined data from turbine 14 curtailment experiments in the United States and Canada in a meta-analysis framework. We tested 15 multiple statistical models to explore possible linear and non-linear relationships between turbine cut-in 16 speed and bat fatality reduction while controlling for control cut-in speed. Because the overall sample size for this meta-analysis was small (n = 36 control-treatment studies from 16 field sites from the American 17 Wind Wildlife Information Center and a recent review), we conducted a power analysis to assess the 18 19 number of control-impact curtailment studies that would be needed to understand the relationship 20 between fatality rate and change in cut-in speed under different fatality reduction scenarios. We also identified the characteristics of individual field studies that may influence their power to detect fatality 21 22 reduction due to curtailment. Using a response ratio approach, we found any curtailment strategy reduced fatality rates by 56% for studies included in this analysis (p < 0.001). However, we did not find strong 23 evidence for linear (p = 0.07) or non-linear (p > 0.11) associations between increasing cut-in speeds and 24 25 fatality reduction. The power analyses showed that the power to detect effects in the meta-analysis was low if fatality reductions were less than 50%. Synthesizing across all analyses, we need more well-26 27 designed curtailment studies to determine the effect of increasing curtailment speed and the effect size is 28 likely of a magnitude that we had limited power to detect.

29

30 Introduction

31 Wind energy development is increasing rapidly worldwide and hundreds of thousands of bat 32 fatalities are estimated to occur per year due to collisions with terrestrial wind energy facilities in North 33 America [1-3]. Turbine attraction is the leading explanation for high observed fatality rates, particularly 34 in migratory tree bats [4,5]. Between 70% and 80% of bats killed at wind energy facilities in the U.S. are 35 migratory tree bats, including hoary bat (Lasiurus cinereus), eastern red bat (L. borealis), and silver-36 haired bat (*Lasionycteris noctivangans*; [3,6–8]. While fatality rates are variable among sites, the 37 magnitude of mortality for some North American bat species is high enough to be considered a serious 38 conservation concern [9,10].

39 Curtailment of turbine operations during low wind conditions, particularly in late summer and fall 40 when fatality rates are highest, has become an accepted operational minimization tactic to reduce bat 41 fatality at terrestrial wind facilities [11]. By increasing the cut-in speed, or the wind speed at which a 42 turbine generator begins to produce electricity, curtailment reduces turbine blade spinning rates. Below the cut-in speed, turbine blades still spin with the wind but do so much more slowly, especially if blades 43 44 are "feathered" or pitched to catch as little wind as possible. Because bats tend to be more active at lower 45 wind speeds, increasing turbine cut-in speed can significantly reduce bat fatality [1,12]. However, a great 46 deal of variability has been reported in the level of fatality reduction achieved by curtailment, likely due to the effect of myriad factors (e.g., curtailment regime, time of year, weather, turbine dimensions, and 47 48 landscape characteristics; [8]) on fatality risk. While site-specific studies have demonstrated that 49 operational curtailment is effective at reducing impacts, the exact nature of the relationship between 50 increases in cut-in speed and fatality reduction in bats remains unclear.

For this study, "blanket" curtailment, in which wind speed and time of day/year are used to
determine when to curtail, has both operational and financial implications for wind facility operators [13].
At present, the exact nature of the trade-off between turbine energy production and bat fatality
minimization is poorly understood. Larger increases in cut-in speeds will further reduce power

generation. Still, the implications for fatality reduction are less clear, in part because this type of assessment requires intensive monitoring and is subject to errors introduced by imperfect detection and small sample sizes. Despite limited evidence that raising blanket cut-in speeds above 4.5 m/s will further reduce bat fatalities [14], regulators now have required operational minimization for some new wind projects in the United States and Canada at wind speeds up to 6.9 m/s [15]. A synthesis of the available data from designed curtailment studies will allow us to quantify better the relative benefits of increasing turbine cut-in speed for reducing bat collision fatality.

62 A meta-analysis framework is used to synthesize data across studies to determine the effect of 63 curtailment on bat fatality reduction. Meta-analysis provides a method to account for multiple types of uncertainty and use predictor variables to explain patterns between studies [16]. Random effects meta-64 analyses are needed to account for the uncertainty in effects from each study and the uncertainty in the 65 66 true effect size to which all studies contribute. Using such an approach, we aim to evaluate the current 67 knowledge of the effectiveness of blanket curtailment regimes in reducing bat fatalities at terrestrial wind projects in North America. We identified three objectives: 1) evaluate existing control-treatment 68 69 curtailment study data for bats in a meta-analysis framework to examine the relative benefit of increased 70 curtailment cut-in speeds and examine the importance of geography and turbine dimensions on fatality 71 reduction; 2) assess the power of the meta-analysis approach to quantify fatality reduction using a data 72 simulation approach; and 3) understand how different site or survey characteristics (e.g., fatality rates, 73 study length, and carcass persistence) influence the power of individual curtailment field studies to detect 74 a difference in bat fatality rates between control and treatment groups. These analyses are combined to 75 identify the most likely effect of blanket curtailment on bat fatality reduction, how much additional 76 information is needed to be certain of these effects, and how to design curtailment experiments to maximize the value of their results. 77

78 Methods

79 The study's overall goal was to understand the relationship between blanket curtailment cut-in 80 speed and bat fatality reduction at wind facilities in the United States and Canada. To achieve this goal, 81 we used a response ratio approach that focused on the differences in fatality rates between control and 82 curtailment treatments in available studies. We used a meta-analysis approach (hereafter referred to as the "meta-analysis") to control for variability among studies. As we did not have a predetermined assumption 83 84 about the nature of the relationship between fatality rate and the change in cut-in speed between control 85 and treatment, we tested multiple statistical models that allowed for both linear and non-linear 86 relationships between cut-in speed and the response to determine which best described the observed 87 pattern. Both the absolute cut-in speed and change in cut-in speed were allowed to influence the predicted fatality rate. Once the best models were selected, we used them to understand how covariates like study 88 location and turbine dimensions could influence the relationship between fatality rate and change in cut-in 89 90 speed.

91 Because the sample size for this analysis was small (n=36 control-treatment pairs), we also assessed the likelihood that the above *meta-analysis* would provide statistically significant results and 92 93 determined the number of control-treatment pairs needed in this meta-analytical framework to be confident in our understanding of the relationship between fatality rate and change in cut-in speed. Thus, 94 95 we conducted two types of power analyses. The first power analysis (the "meta-analysis power analysis") was designed to quantify the power of the meta-analysis under different hypothetical scenarios about the 96 97 relationship between fatality rate and change in cut-in speed. The first of these scenarios was an a 98 *posteriori* scenario based on the results of the best meta-analysis model using existing data, and four 99 additional *a priori* scenarios with different relationships between fatality reduction and cut-in speed were 100 also examined. The second power analysis (the "fatality estimation power analysis") was designed to inform future curtailment studies and fatality monitoring efforts at operating wind energy facilities. This 101 102 analysis assessed the relative quality of different fatality studies at the project scale and identified site and survey characteristics (e.g., fatality rate, study length, and carcass persistence) that influenced the power 103

of individual curtailment field studies to detect a difference in bat fatality rates between control and
 treatment groups. All analyses were conducted in R [17], and all analysis scripts were documented in
 Supplementary Information.

107 *Data Inclusion*

Data for this analysis were collected in part from the American Wind Wildlife Information Center 108 (AWWIC, Accessed in August 2019), which compiles private and public data from post-construction 109 fatality monitoring studies at individual wind energy projects in the United States (n=43) [7]. Data from 110 111 several additional studies in the U.S. and Canada were harvested from publicly available reports (n = 22) 112 with overlap to the AWWIC studies) [14]. Paired control-treatment curtailment studies (hereafter 113 'studies') with blanket curtailment treatments were of primary interest for this analysis. Studies were included in the analysis if there was both a treatment and control group of turbines with fatality estimates 114 115 at different cut-in speeds at the same project site (Fig. 1). Data were excluded from analysis if there was 116 no change in cut-in speed between treatments (e.g., testing other fatality reduction methods) or no measurement of treatment effect (e.g., no control treatment). The remaining studies in the database (n=36;117 Table 1) were conducted at 17 wind energy project sites in the U.S. and Canada from 2005-2016. There 118 119 were instances where multiple experimental cut-in speeds were tested simultaneously at the same project, 120 resulting in multiple studies from the same project and year that shared a control. Studies without precision estimates for their fatality ratios were included in the analysis by applying the global average 121 122 standard error.

Figure 1. PRISMA meta-analysis data flow diagram for the study. After receiving a list of all projects in AWWIC
 and the CanWea syntheses that reported turbine curtailment we removed duplicates between the two sources.
 Individual studies within those projects were determined to be suitable for analysis if they used a blanket curtailment
 treatment and there were multiple cut-in speeds that could be compared.

127

Fatality estimates in the AWWIC database, which were reported from the original studies, had already been adjusted for detection probability (observer ability to detect carcasses that are present) and carcass persistence (rate of removal of carcasses by scavengers) using searcher efficiency trials and

131 carcass persistence trials, respectively [18]. There are multiple approaches for correcting fatality estimates

that differ in their assumptions regarding how to account for detection error resulting from carcass

removal and searcher efficiency [18,19].

135 year, geographic region, and rotor diameter in meters (RD); control cut-in speed (Cont.), experimental cut-in speed

136 (Exp.), and change in cut-in speed (Δ), all in m/s; and treatment effect information, including the mean fatality ratio

 \pm SE (Fatal. Ratio) and percent decrease in fatality between treatments (%). Studies from the same project and year

138 were tested simultaneously and share a control. Some studies lacked information on fatality uncertainty; for these,

the global average standard error was applied to the fatality ratio.

				Cut-in	Speed		Effect		
Project Name	Year	Region	RD	Cont.	Exp.	Δ	Fatal. Ratio	%	Source
Anonymous East	2014	East	77	3.5	4.5	1.0	0.50 ± 0.16	50	AWWIC ²
Anonymous East	2015	East	77	3.5	5.5	2.0	1.00 ± 0.42	0	AWWIC
Anonymous East	2015	East	77	3.0	4.0	1.0	0.91 ± 0.36	9	AWWIC
Anonymous East	2016	East	77	3.0	5.0	2.0	0.69 ± 0.33	31	AWWIC
Anonymous East	2016	East	77	3.0	6.0	3.0	0.53 ± 0.34	47	AWWIC
Casselman Wind	2008	East	77	3.5	5.0	1.5	0.13 ± 0.14	87	AWWIC
Casselman Wind	2008	East	77	3.5	6.5	3.0	0.26 ± 0.18	74	AWWIC
Casselman Wind	2009	East	77	3.5	5.0	1.5	0.32 ± 0.17	68	AWWIC
Casselman Wind	2009	East	77	3.5	6.5	3.0	0.24 ± 0.15	76	AWWIC
Criterion	2012	East	93	4.0	5.0	1.0	0.38 ± 0.14	62	AWWIC
Laurel Mountain	2011	East	82	3.5	4.5	1.0	0.42 ± 0.15	58	AWWIC
Pinnacle Wind Force	2012	East	95	3.0	5.0	2.0	0.53 ± 0.15	47	AWWIC
Pinnacle Wind Force	2013	East	95	3.0	5.0	2.0	0.42 ± 0.21	58	AWWIC
Pinnacle Wind Force	2013	East	95	3.0	6.5	3.5	0.25 ± 0.14	75	AWWIC
Anonymous Midwest	2010	Midwest/West	82	3.5	4.8	1.3	0.53 ± 0.25	47	CanWEA ³
Anonymous Midwest	2010	Midwest/West	82	3.5	4.0	0.5	0.28 ± 0.13	72	CanWEA
Anonymous Pac. SW	2012	Midwest/West	101	3.5	4.8	1.3	0.80 ± 0.37	20	CanWEA
Anonymous Pac. SW	2012	Midwest/West	101	3.5	4.0	0.5	0.65 ± 0.30	35	CanWEA
Anonymous Pac. SW	2012	Midwest/West	101	3.5	4.8	1.3	0.62 ± 0.29	38	CanWEA
Fowler Ridge 1	2010	Midwest/West	89	3.5	5.0	1.5	0.50 ± 0.11	50	AWWIC
Fowler Ridge 1	2010	Midwest/West	89	3.5	6.5	3.0	0.21 ± 0.07	79	AWWIC
Fowler Ridge 1	2011	Midwest/West	89	3.5	4.5	1.0	0.64 ± 0.29	36	AWWIC
Fowler Ridge 1	2011	Midwest/West	89	3.5	5.5	2.0	0.38 ± 0.18	62	AWWIC
Fowler Ridge 1	2012	Midwest/West	89	3.5	5.0	1.5	0.16 ± 0.06	84	AWWIC
Lakefield	2016	Midwest/West	77	3.5	5.0	1.5	0.56 ± 0.34	44	AWWIC
Summerview	2005	Midwest/West	80	4.0	7.0	3.0	0.61 ± 0.28	39	CanWEA
Summerview	2007	Midwest/West	80	4.0	5.5	1.5	0.94 ± 0.26	6	CanWEA
Wild Cat 1	2013-15	Midwest/West	100	5.0	7.0	2.0	0.20 ± 0.07	80	AWWIC
Wild Cat 1	2016	Midwest/West	100	5.0	6.9	1.9	0.41 ± 0.33	59	AWWIC
Bull Hill	2013	Northeast	100	3.0	5.0	2.0	0.70 ± 0.23	30	AWWIC
Enbridge Wind	2012	Northeast	82	3.5	5.5	2.0	0.38 ± 0.18	62	CanWEA
Raleigh Wind	2014	Northeast	77	3.5	4.5	1.0	0.23 ± 0.05	77	CanWEA
Sheffield 2012		Northeast	94	4.0	6.0	2.0	0.37 ± 0.13	63	AWWIC
Talbot Wind ¹	2013	Northeast	101	3.5	5.5	2.0	0.04 ± 0.18	96	CanWEA ¹
Wolfe Island	2011	Northeast	93	3.2	4.5	1.3	0.52 ± 0.24	48	CanWEA
Wolfe Island	2011	Northeast	93	3.2	5.5	2.3	0.40 ± 0.18	60	CanWEA

140 ¹ This study was a statistical outlier that did not meet the assumptions of the meta-analysis and thus was excluded

² Data obtained from the American Wind Wildlife Information Center (AWWIC) database, which includes both public and

¹³⁴ Table 1. Bat fatality curtailment study data from the AWWIC and CanWEA databases, including project name,

142 private data.

- **143** 3 Data obtained from a review by the Canadian Wind Energy Associate [CanWEA 14] and public reports cited therein.
- 144

145 Studies in this analysis primarily used the Huso and Shoenfield estimators [20,21]. However, some 146 studies used the Erickson estimator [22], MNRF estimator [23], or custom calculations to adjust fatality estimates for carcass persistence and searcher efficiency. Adjusted bat fatality estimates per turbine were 147 presented in the database with upper and lower 90-95% confidence intervals (CIs). As the experimental 148 149 period and hours per night when experiments were implemented varied among studies, fatality estimates 150 were converted to bat fatality per turbine-hour by dividing fatalities per turbine by the total number of hours of curtailment (number of nights*hours per night). Study periods varied somewhat between 151 152 individual curtailment studies, with some studies examining specific time periods throughout the night or 153 focusing on different windows of time during fall migration; our approach controls for study-specific 154 variability by pairing control-treatment groups for analysis, but does assume that the relationship between turbine-hours and fatalities is robust to potential variation in the effect of curtailment through time. Few 155 156 studies reported species-specific fatality rates, so fatality estimates were for all bat species combined. 157 Meta-analysis 158 The effect size of each study was calculated as a log-transformed ratio between the estimated

159 fatality of the treatment and the control, both in the unit of bat fatalities per turbine hour (i.e., the log-160 transformed response ratio, hereafter 'RR'). In instances where only a percent decrease was reported, this 161 was used to calculate the RR $(\log(1-(\% \text{ decrease}/100)) = RR)$. This effect size approach controls for differences in study design ranging from site-specific effects to the choice of fatality estimator. 162 Manufacturer cut-in speed can vary among turbine makes and models. In most studies, the control 163 group's cut-in speed was 3.5 m/s (a common cut-in speed set by turbine manufacturers), though values 164 165 ranged from 3.0 to 5.0 m/s. Experimental cut-in speeds varied from 4.0 to 7.0 m/s. Due to this variation 166 and the small sample size of available studies, the change in cut-in speed between treatment and control

167 (Δ cut-in speed) was used as the estimate of treatment magnitude. Thus, the analysis focused on the effect 168 of relative rather than absolute change in curtailment cut-in speed.

We used a random effects meta-analysis that accounts for heterogeneity in the true effect 169 170 (between-study variance) and sampling error (within-study variance; [24]). The inclusion of between-171 study variance (i.e., the random effect τ) allows for the incorporation of additional uncertainty in the 172 analysis by assuming the true effect is a random variable that is realized at different magnitudes in 173 different studies. Confidence intervals (90% or 95% depending on the estimator used) for control and 174 treatment fatality estimates were converted into standard error (SE) estimates assuming an approximately 175 normal distribution. While confidence intervals were slightly asymmetrical, a normal approximation was the best available strategy for conversion given the variation in fatality estimators used across studies. For 176 independent studies (i.e., those with no shared control), standard error estimates of the RR were 177 178 calculated using the delta method [25,26]. In instances where multiple studies shared a common control 179 (i.e., were conducted simultaneously at the same project site; n=23), the correlation among the studies was calculated by decoupling the associated dependence into a single estimate of uncertainty for each 180 181 study [26,27]. In instances where no estimate of uncertainty was provided in the original study, the mean SE of all studies after decoupling was applied to the estimate. 182

To conduct the meta-analysis and explore the possible relationships between Δ cut-in speed and 183 184 bat fatality rates, we ran two types of models with the RR as the dependent variable and Δ cut-in as the 185 primary explanatory variable, with control cut-in speed included as an additional covariate. Using the 186 'metagen' and 'metareg' functions from the *meta* R package [28] we tested: 1) non-linear categorical 187 model specifications where studies were binned into three discrete categories to simplify model fitting 188 $(1 = \Delta \text{ cut-in values} \ge 0.5 \text{ and} < 1.4 \text{ m/s}, 2 = \Delta \text{ values} \ge 1.4 \text{ and} < 2.6 \text{ m/s}, \text{and} 3 = \Delta \text{ values} \ge 2.6 \text{ m/s}); \text{ and}$ 189 2) *linear continuous* model specifications that treated Δ cut-in as a continuous variable. As the categorical 190 model ignored the ordinal relationships among treatment groups, the continuous model was implemented to help determine the degree of bias in this approach. We also explored the influence of bin choice on the 191

fit of the categorical model (Appendix A) and other types of models that test for non-linear relationships in fatality ratio and Δ cut-in (e.g., continuous quadratic relationships) before determining the best approach for this question.

195 For all model types, variation in individual study precision (within-study variance) was accounted 196 for using a weighted regression approach, so that studies with higher precision influenced the model 197 parameter estimates more than studies with lower precision. Study weight was determined as a function of 198 the inverse square of the study standard error plus the overall between-study variance (τ^2). τ^2 was 199 estimated using a restricted maximum likelihood approach [29]. One study (Talbot Wind) was determined 200 to be an outlier and was removed due to disproportionately high leverage compared other studies. 201 Additional covariates included rotor diameter (RD) and geographic region. Geographic region (Northeast, East, Midwest/West) was based upon EPA ecoregions (https://www.epa.gov/eco-research/ecoregions) but 202 203 consolidated to ensure enough studies per category for inclusion in the model (Table 1). Hub height was 204 considered for inclusion as a covariate but had little variation across studies (n=30 studies with hub height 205 of 80 m). There was not enough data to consider interactions among these covariates. We also lacked data 206 to consider controlling site dependencies among studies using a random effect. We examined between-207 study heterogeneity and model goodness of fit using Cochran's Q (QE), τ^2 , and I² model statistics [30]. 208 Model selection was performed based on AIC_c values and model weights were calculated based on these 209 values for each model type separately.

210 Meta-analysis Power Analysis

To determine the number of studies required in a random effects meta-analysis to detect relative changes in RR with changing cut-in speed reliably, we implemented a power analysis at the meta-analysis scale using a simulation approach [31]. We conducted meta-analysis power analyses for the *non-linear categorical* and *linear continuous* descriptions of the relationship between Δ cut-in speed and RR. For the categorical relationship power analysis, simulations were designed using the Δ cut-in speed categories defined above to replicate the meta-analysis under multiple scenarios. The number of studies per Δ cut-in

category, fatality reduction for the first Δ cut-in speed category (β_0), and the subsequent reduction in the second and third categories (β_1 , β_2), were varied across simulations. The following linear regression equation was used for the categorical model:

220
$$RR = \beta_0 + \beta_1 X_1 + \beta_2 X_2 + \varepsilon + \tau$$

221 where X_1 and X_2 are dummy covariates that represent Δ cut-in Categories 2 and 3, respectively. The 222 uncertainty from the Gaussian error term (ε) and inter-study differences (τ) were added by using a normal 223 distribution with a mean of 0 and a standard deviation equal to that observed in the fatality ratio of 224 control-treatment data (Table 1; SD = 0.24). We used a uniform distribution to randomly assign a SE to each simulated study, which ranged between the minimum and maximum of the observed study standard 225 226 errors (Table 1; range: 0.05-0.34). Once the model was simulated, we used the methods described above 227 to estimate parameters. Each scenario was simulated 10,000 times with 5, 10, 20, and 30 studies per Δ cut-in category to achieve precise estimates of power. The statistical power of each parameter (β_0 , β_1 , and 228 β_2) and sign error (the probability that the estimate was the same sign as the given parameter; [32]) were 229 230 calculated to determine the effectiveness of the model in estimating the scenario parameters. Power was 231 determined by examining whether the results were significantly different from the value of no effect (1 for β_0 , and 0 for β_1 and β_2 ; $\alpha = 0.05$), and the sign error was computed by comparing the signs of the true 232 233 parameter value and the estimated value.

For the *linear continuous* models, Δ cut-in speed was randomly assigned to each study. To do this, we used the same category framework (where 5, 10, 20, or 30 studies were assigned to each Δ cut-in category), and studies in this category were randomly assigned a Δ cut-in speed from that category that was observed in the studies included in the meta-analysis. These values were then scaled (centered on zero) and used to build a linear model:

$$RR = \beta_0 + \beta_1 X_1 + \varepsilon + \tau$$

240 where X_1 is the scaled continuous Δ cut-in speed value for each study.

241	Five scenarios were simulated for the power analysis for the two model types (Table 2). Each
242	scenario was replicated at four different sample sizes (5, 10, 20, and 30 studies per Δ cut-in category).
243	Four of these scenarios were selected a priori to explore our power to detect different types of
244	relationships between fatality ratio and Δ cut-in. We included three scenarios thought to represent
245	plausible hypotheses based on observed results to date: 1) a 25% linear decrease in fatality per 1 m/s
246	increase in cut-in speed; 2) a 50% initial decrease in fatality with Category 1 Δ cut-in speed and
247	subsequently stable fatality rates; and 3) an initial 50% decrease in fatality with Category 1 Δ cut-in speed
248	and then 10% subsequent declines in fatality for Categories 2-3. The fourth scenario, a more extreme 50%
249	exponential decrease per 1 m/s increase in cut-in speed, was intended to provide context for interpreting
250	the results of other scenarios. We also included an a posteriori 'current knowledge' scenario that used
251	parameter estimates obtained from the top model in our meta-analysis (above).

252

Table 2. Parameters used in simulation scenarios for the meta-analysis scale power analysis of bat fatality with changes in cut-in speed. A total of 20 scenarios for each model type were run with different $\beta_0 \beta_1$ and β_2 values and

		p 0, p 1,
255	differing numbers of studies per category (<i>n</i> Studies). Parameters are log-transformed.	

Model	Scenario	β ₀	β1	β2	n Studies
Non-Linear Categorical	Current Knowledge	-0.67	-0.19	-0.46	5, 10, 20, 30
-	25% Linear Decrease	-0.29	-0.41	-1.1	5, 10, 20, 30
	50% Decrease then Stable	-0.69	0.00	0.00	5, 10, 20, 30
	50% Decrease then 10% Decline	-0.69	-0.22	-0.51	5, 10, 20, 30
	50% Exponential Decrease	-0.69	-0.69	-0.99	5, 10, 20, 30
Linear Continuous	Current Knowledge	-0.84	-0.17		5, 10, 20, 30
	25% Linear Decrease	-0.59	-0.59		5, 10, 20, 30
	50% Decrease then Stable	-0.52	-0.27		5, 10, 20, 30
	50% Decrease then 10% Decline	-0.70	-0.5		5, 10, 20, 30
	50% Exponential Decrease	-1.04	-0.89		5, 10, 20, 30

256

257 Thus, this power analysis represented a combination of *a priori* and *a posteriori* approaches designed to

understand the efficacy of the current study, estimate the number of studies needed to reduce uncertainty

in the meta-analysis, and inform the likelihood that the observed data could be generated by the *a priori*

scenarios.

261 Fatality Estimation Power Analysis

262 To understand the traits of effective curtailment experiments and provide guidelines for future studies, we used a data simulation approach to determine the effectiveness of project-scale curtailment 263 264 studies at detecting differences in bat fatality rates using a hierarchical simulation approach [33]. Three 265 different data sets were simulated to replicate the process by which true fatalities rates are estimated in 266 curtailment field studies. First, the true number of bat mortalities was simulated using a Poisson process. 267 New fatalities were generated each night for each turbine using a fatality rate per turbine-night as a 268 Poisson mean. Second, carcass persistence rate was estimated using a carcass persistence trial format. 269 Here, we used the exponential distribution to simulate the survival rates of 50 carcasses at the site based on the predefined median number of days of carcass persistence. Carcass searches were assumed to occur 270 271 every three days for the duration of the study, and the survival probability of the carcasses was used to estimate daily carcass persistence probabilities for the survey. Third, searcher efficiency data were 272 273 simulated based on 100 detection trials using a binomial distribution. These data sets were combined to 274 determine the number of carcasses detected by the surveyors at each survey interval. Detection probability for each carcass was a function of carcass persistence, which changed in a time-dependent 275 276 manner following the fatality event, and searcher efficiency, which was constant across time. The number 277 of observed mortalities was determined using a binomial draw from the combined probability of 278 persistence and detection for each fatality.

Forty-eight scenarios were used in this power analysis to explore the effects of effect size, study period, and carcass persistence on study design, and were based upon information from an early version of the subset AWWIC database with curtailment studies (June 2019). Simulation parameters were selected based on averages and ranges from the interim database, and are useful approximations of values in typical curtailment studies. The curtailment treatment effect was defined as either a 25% or 50% reduction in fatality rates (n=24 scenarios for each effect). These values were selected based on the 50% reduction approximated the average fatality reduction. The number of turbines (10 or 15), number of

experiment nights (45 or 90), fatality rate (0.1 or 0.3 mortalities/turbine-night), and carcass persistence
rate (3, 6, or 9 mean days of persistence) were varied to determine the effect of these variables on
statistical power. The number of turbines and experiment nights are combined as turbine-nights to
describe study effort. Chosen carcass persistence values tended to be on the lower end of the range of
observed values to test the power of these studies in more challenging environmental conditions.
Detection probability was fixed at 50% for all studies, the approximate median of the described studies.

292 Data were simulated for each scenario using base functions in R v. 3.6 [17] and package simsurv 293 [34]. Package *GenEst* [35] was then used to estimate the true number of fatalities for each treatment group 294 with the simulated data sets. This process was repeated 50,000 times to obtain consistent estimates of statistical power. This generalized fatality estimator ('GenEst') differs from those used by studies in the 295 296 AWWIC database but is considered the current best practice for estimating fatality from wind turbines 297 when the sample size is sufficiently large to estimate known biases [35]. The Bayesian posterior 298 distribution of the number of fatalities for each treatment group was estimated using the function 'estM' 299 in package *GenEst*. Simulated carcass observations, carcass persistence trial data, and searcher efficiency 300 data were used as inputs along with assumed static values for the proportion of area searched (50% for all turbines) and the search schedule (once every three days for all turbines). The mean number of mortalities 301 302 in the 25% and 50% reduction treatment groups (along with 95% credible intervals) were estimated using 303 a parametric bootstrapping approach (n = 1000). The 95% credible interval of the difference of the 304 GenEst-derived fatality estimates between these two groups was calculated to determine overlap with 305 zero and used to estimate statistical power for each scenario, and was determined by subtracting the 306 bootstrapped simulations for each treatment group. If a simulation study group did not detect any 307 carcasses, we did not include it in the power analysis calculation.

308

309 Results

310	Fatality ratios in the database representing fatality reduction due to curtailment ranged from 0.13
311	(87% decrease in fatalities) to 1.00 (0% decrease in fatalities) with an arithmetic mean of 0.46 (53%
312	decrease; n=35 studies). Lower fatality ratio values represented a greater reduction in bat fatality per
313	turbine-hour, while a value of one indicates no difference between curtailment treatment and control (0%
314	decrease). When examining fatality ratios by Δ cut-in category, the mean fatality ratio for Category 1 was
315	0.60 (n = 12), Category 2 was $0.41 (n = 18)$, and Category 3 was $0.37 (n = 6)$, suggesting a possible non-
316	linear relationship with Δ cut-in speed (Fig. 1).

Figure 2. The relationship between bat fatality and curtailment difference (Δ cut-in; calculated as a change in m/s
between the treatment and control groups) for 16 wind farms in North America. Some wind farms have multiple
data points as there were multiple years of experiments or multiple treatments tested within a year (n=36 studies).
Error bars represent the standard error of the fatality ratio. Talbot Wind was excluded from the meta-analysis as an
outlier.

322

323 Meta-analysis

Thirty-five individual studies (from 16 projects) were included in the meta-analysis modeling. The estimated fatality ratio across all studies (i.e., the estimate before controlling for Δ cut-in speed) was 0.44 (95% CI: 0.36-0.49, z = -9.18, p < 0.001; Fig. 2). For both categorical and continuous model types, the models with Δ cut-in and control cut-in as covariates represented the best model fit, as indicated by AIC_c (Appendix A). Comparing across model types, the linear model represented the best fit (AIC_c = 55.67), likely due to model simplicity; the categorical model showed slightly worse fit (AIC_c=58.48). A forest plot was used to show that there was not evidence of publication bias.

331	The best fit lir	iear model had a la	arge and significant	amount of residual	heterogeneity between
-----	------------------	---------------------	----------------------	--------------------	-----------------------

studies (QE₃₂= 50.50, p = 0.02), with an among-study variance estimate (τ^2) of 0.10 (CI: 0.00, 0.19), while

- the percentage of overall variation across studies due to heterogeneity (I^2) was 39.2% (CI: 1.1-53.9%).
- Based on the linear model, the RR tended to decrease with increasing Δ cut-in (slope parameter β = -0.17,
- CI: -0.36-0.02; Fig 3); this relationship nearly met the requirement for statistical significance (z = -1.78, p
- = 0.07). Control cut-in speed was not a significant covariate (β = -0.14, CI: -0.32-0.05, z = -1.47, p =

- 337 0.14). There was no significant effect of rotor diameter (95% CI: -0.16- 0.24, z = 0.35, p = 0.72) or
- geographic region (Midwest vs. East, CI: -0.35- 0.49, z = 0.32, p = 0.75; Northeast vs. East, CI: -0.65,
- 0.33, z = -0.63, p = 0.53) on bat fatality ratios. The addition of these covariates did result in a slight
- decrease in model heterogeneity, however (rotor diameter: $QE_{31} = 50.0$, p = 0.02, $\tau^2 = 0.11$, $\tau^2 CI = 0.00$ -
- 341 0.20, $I^2 = 40.1\%$, $I^2 CI = 1.9-55.2\%$; geographic region: $QE_{30} = 46.3$, p = 0.03, $\tau^2 = 0.11$, $\tau^2 CI = 0.00-0.21$,
- 342 $I^2 = 38.7\%$, $I^2 CI = 0.0-55.3\%$).

Figure 3. The effect of curtailment regime on bat fatalities at terrestrial wind farms in North America from a metaanalysis incorporating within- and among-study variance. The plot shows the fatality ratio (black square) and 95% CI (error bars) of individual studies along with the mean effect size for each Δ cut-in category (grey diamonds). The 95% confidence interval of the overall effect is shown at the bottom (black diamond). Individual studies were weighted (out of 100%) based on study uncertainty (CI, in brackets) and distance from category mean effect, with square size indicating relative weighting. A fatality ratio of 1 indicates no difference in fatality rate between the control and experimental curtailment treatments.

- 350
- The best fit model with a categorical response to Δ cut-in speed also had a large and significant amount of residual heterogeneity between studies (QE₃₁= 50.76, p = 0.01), with an among-study variance estimate (τ^2) of 0.11 (CI: 0.01-0.21), while the percentage of overall variation across studies due to heterogeneity (I²) was 41.2% (CI: 2.9-56.0%). When examining fatality reduction by Δ cut-in speed, the model predicted a fatality ratio estimate for Category 1 of 0.52 and represented a significant reduction in fatality rates (β =-0.67, CI: -0.97 to -0.37, z = -4.38, p < 0.0001).

Figure 4. Meta-analysis estimated linear (black line) and categorical (pink) effect of Δ cut-in speed on bat fatality ratio at North American wind energy projects. Black dots represent fatality ratios for individual studies; note that uncertainty in individual study estimates, which influenced the meta-analysis parameter estimates, are not shown here (see Table 2 for these values). Categorical model points are based on mean Δ cut-in speed for the category. Error bars are 95% confidence intervals of estimates.

- 362
- 363 The model estimates for fatality ratios for Categories 2 and 3 were 0.42 and 0.34 respectively, but the
- marginal change of increasing Δ cut-in from Category 1 to Category 2 ($\beta_1 = -0.19$, CI: -0.59-0.20, z = -
- 365 0.96, p = 0.33) and from Category 1 to Category 3 (β_2 = -0.46, CI: -1.02-0.11, z = -1.58, p = 0.11) were
- small, with high amounts of uncertainty in the estimates (Fig. 3).Control cut-in speed was not a
- significant covariate (β = -0.13, CI: -0.32-0.06, z = -1.31, p =0.19). Analysis of study-scale covariates

revealed no significant effect of rotor diameter (95% CI: -0.19- 0.23, z = 0.16, p= 0.87) or geographic region (Midwest vs. East, CI: -0.35- 0.55, z = 0.45, p = 0.65; Northeast vs. East, CI: -0.71, 0.32, z = -0.73, p = 0.46) on bat fatality ratios. The addition of these covariates did not result in decreases in model residual heterogeneity.

372

373 Meta-analysis Power Analysis

Power analysis of the categorical model revealed that for most scenarios, five studies were 374 required to have adequate statistical power (>0.8) to determine an effect of curtailment on the fatality 375 ratios for Category 1 (β_0 ; 0.5-1.3 m/s Δ cut-in; Fig. 4). The exception was the 25% linear decrease 376 377 scenario, which required over 30 studies to achieve adequate power due to smaller changes at lower Δ 378 cut-in speeds. The statistical power of β_1 and β_2 (Δ cut-in Categories 2-3) were more variable across scenarios (Fig. 4). For β_1 (1.5-2.3 m/s Δ cut-in speed), the 50% exponential decrease scenario had 379 sufficient power at 20 or more studies, and the 25% linear decrease scenario had sufficient power at 30 380 studies per group, but no other scenario met the criteria for sufficient power. For β_2 (3-3.5 m/s Δ cut-in 381 382 speed), two scenarios achieved sufficient power with less than 10 studies per group (50 % exponential decrease, 25% linear decrease), while another two achieved sufficient power with 20-30 studies per group 383 384 (50% decrease followed by 10% decreases, and current knowledge scenario). Sign error decreased with 385 increasing sample size for all parameters except those that were set at zero (β_1 and β_2 in the 50% decrease then stable scenario) and decreased below 10% at 10 studies per category for most other parameter 386 387 estimates.

Figure 5. The relationship of statistical power and sign error with sample size in the categorical meta-analysis-scale power analysis of curtailment studies to reduce bat fatality rates. We examined the relationship between the number of studies per category of Δ cut-in speed (Category $1 = \beta_0 = 0.5 - 1.3$ m/s Δ cut-in speed, Category $2 = \beta_1 = 1.5 - 2.3 \Delta$ cut-in speed, Category $3 = \beta_2 = 3 - 3.5$ m/s Δ cut-in speed) and 1) the statistical power to detect change between categories (at top), and 2) the rate at which models would be expected to incorrectly predict the sign of parameter estimates (at bottom). Colors represent different curtailment regime scenarios. The horizontal dashed lines represent the 0.8 power threshold and 50% sign error threshold, respectively.

396	In comparison, the continuous model often had higher power, particularly for constant or
397	increasing relationships between RR and Δ cut-in (Fig. 5). Power to detect linear trends (β_1), particularly
398	for the scenarios with decreases at Δ cut-in speeds greater than 1.3 m/s, was greater than 0.8 even with
399	only 5 studies per group. Only the 50% then stable and current knowledge scenarios showed poor power,
400	likely needing 20-30 studies per group to measure the decrease accurately. The average value (or
401	intercept, β_0) was more difficult to precisely estimate, though this parameter is less important to the
402	present study as it does not estimate the change in effect with Δ cut-in. As the current knowledge scenario
403	had the smallest slope out of all the scenarios, it required the largest sample size to have sufficient
404	statistical power-around 30 studies per category. Sign error was low across all scenarios; it was lower
405	than 10% whenever the number of studies per group was greater than 10.
406 407 408	Figure 6. The relationship of statistical power and sign error with sample size in the linear continuous meta-analysis- scale power analysis of curtailment studies to reduce bat fatality rates. We examined the relationship between the number of studies per Δ cut-in speed category (Category 1 = β_0 = 0.5-1.3 m/s Δ cut-in speed, Category 2 = β_1 = 1.5-
409	2.3 Δ cut-in speed, Category 3 = β_2 = 3-3.5 m/s Δ cut-in speed) and 1) the statistical power to detect change in

fatality ratio (at top), and 2) the rate at which models would be expected to incorrectly predict the sign of parameter
 estimates (at bottom). Colors represent different curtailment regime scenarios. The horizontal dashed lines represent
 the 0.8 power threshold and 50% sign error threshold, respectively.

413

414 Fatality Estimation Study Power Analysis

415 At the scale of individual curtailment experiments at wind facilities, many factors influenced 416 these studies' statistical power and sign error. More turbine-nights increased the power of studies in all scenarios (Fig. 6). However, the importance of turbine-nights varied with several variables outside of 417 researcher control, such as effect size and carcass persistence. With a 25% fatality reduction between 418 419 experimental and control treatments, no tested scenario achieved statistical power of 0.8 when the control fatality rate was low (0.1 mortalities/turbine-night). For scenarios with a 25% reduction in fatality, 420 421 statistical power was high only when fatality rate, carcass persistence, and turbine-nights were also high (Fig. 6A). The statistical power of studies in the 50% fatality reduction scenarios was more resilient to 422 423 changes in sampling period and carcasses persistence than the lower-reduction scenarios. Statistical 424 power was above the 0.8 threshold across almost all scenarios with high fatality rates (0.3 fatalities per

425	turbine-night), and a large number of turbine-nights yielded strong statistical power even when the fatality
426	rate was lower (Fig. 6B). Sign error followed a similar pattern, errors occurred more often when fatality
427	rates and fatality reduction from curtailment were low (Fig. 6C). When fatality reduction was 50%, sign
428	error was almost always less than 10% (Fig. 6D). In summary, these simulation results suggest that many
429	curtailment study designs could be effective at detecting differences between treatments in situations with
430	high fatality rates and high carcass persistence. None of the tested study designs were effective in
431	detecting change when fatality reduction and carcass persistence were low. Based on the studies in our
432	database (which had a median number of 14 turbines and 75 experimental nights, 1050 turbine-nights,
433	and 16 of 36 studies with percent fatality reductions <50%), many studies could have low power and high
434	sign error if fatality rate and carcass persistence is low.

Figure 7. The relationship of statistical power (A, B) and sign error (C, D) with sample size for curtailment
treatment groups using a simulation approach (n=50,000) using the Generalized Mortality Estimator (*GenEst*).
Variation in power across turbine-nights of study (T-N), fatality rates, and carcass persistence (in mean days of
persistence) is shown when curtailment is simulated to reduce fatality by 25% (A, C) and 50% (B, D). Simulations
assume a three-day search interval for fatality searches and 50% searcher efficiency.

440

441 Discussion

442 Like past studies, we found evidence that turbine blanket curtailment reduces fatality rates of bats 443 at wind farms at sites that have implemented the technique (as reviewed by Arnett et al. [36]). However, 444 the marginal effect of increasing turbine cut-in speed on fatality rates is more difficult to assess. Using a 445 *meta-analysis* approach, we estimated that the effect of a 0.5-1.3 m/s increase in cut-in speed resulted in a fatality ratio of 0.52, or a 48% reduction in bat fatalities. Estimated reductions in bat fatalities at higher Δ 446 cut-in speeds were not found to be significantly different than this value and had high modeled 447 448 uncertainty. The sample size was small, particularly at higher Δ cut-in speeds. Within the context of the *meta-analysis power analysis*, we only had the statistical power to consistently detect reductions of $\sim 50\%$ 449 450 per 1 m/s Δ cut-in speed. Combined with our meta-analysis results, it appears unlikely that larger 451 increases in Δ cut-in speeds beyond Category 1 result in >50% additional fatality reduction (e.g., the 50%) 452 exponential decrease scenario). Given that we lacked statistical power to detect changes in fatality ratio

453 less than 50%, it is possible the true effect could still be large enough to be ecologically relevant to bat 454 conservation and management. Other concurrent efforts to address a blanket curtailment found that 455 increased curtailment speeds did significantly affect bat mortality [37]. While the methods and data set 456 differ from the present study, Whitby et al. [37] corroborates our estimate of effect size and also show 457 how volatile these results can be when sample sizes are low.

458 Uncertainty in fatality reduction was variable across studies. While this imprecision was 459 accounted for in the meta-analysis framework and propagated into parameter estimates, uncertainty 460 should be minimized through careful study design to maximize the value of each study. Through our 461 *fatality estimation power analysis*, we found that with high fatality rates (≥ 0.3 fatalities per turbine-night) and carcass persistence ($\geq 6-9$ days), experimental studies were consistently successful in detecting 25-462 50% fatality reductions. However, studies with low fatality rates (0.1 fatalities per turbine-night) and 463 464 carcass persistence (3 days) were not adequate to detect 25% differences in fatality rates between treatment and controls groups even with high numbers (>2500) of turbine-nights. These results suggest 465 466 that effective monitoring studies can be conducted when assumptions are met (e.g., detection probability 467 is at least 50%), some number of studies could have low statistical power when using the *GenEst* modeling framework. As the additive effect of further increases in cut-in speed is uncertain, continuing to 468 469 conduct high quality curtailment experiments with a high number of experimental turbine-nights, 470 particularly if fatality rates are expected to be low, would provide data to better estimate the effect of 471 blanket curtailment and inform conservation and management activities for bats [38].

472 *Assessing the likelihood of curtailment effects on bat fatalities*

While the overall effect of blanket curtailment on bat fatality was clear, the relative effect of incrementally larger increases in curtailment cut-in speed was not. This result was likely due to both small effect size and sample size. The results of the current knowledge scenario in the *meta-analysis power analysis* suggested a low likelihood of detecting an effect of higher cut-in speeds with either the nonlinear categorical or linear continuous models, likely requiring around 25 additional studies at higher cut-

in speeds to precisely measure the effect. While *a posteriori* power analyses, such as this scenario, are
redundant with the statistical test on which they are based, these tests can still be useful for evaluating
study success and determining how much additional sampling effort is required [32].

481 However, as the current knowledge scenario parameter estimates are not precisely measured, the 482 *a priori* scenarios provide additional guidance on what effects are observable consistently within the 483 current analytical framework. At the current sample size, we have the power to detect differences in the 484 categorical framework at $\Delta 2$ m/s for the scenarios with the highest magnitude decrease (25% linear and 485 50% exponential) and detect linear trends for these same scenarios as well as the 50% initial decline/10% 486 long-term decline scenario. Given the lack of effects detected in the *meta-analysis* using the categorical 487 model (at higher Δ cut-in speeds), and the marginal effects detected in the linear model, it is unlikely that 488 the 25% linear or 50% exponential decrease scenarios represent the true effect. Thus, a smaller decrease is 489 more likely, though more data are needed to measure the effect precisely.

490 It is unlikely that fatality reduction and absolute cut-in speed are linearly related, as there is a high 491 potential for varying effects across sites [5,39], but control cut-in speed was not an important predictor in 492 our models, suggesting that the RR approach was effective in standardizing effect sizes across studies. 493 Neither rotor diameter nor geographic region explained much variation in RR, which may relate to the 494 scale of the variable; in ecoregion, just three broad geographic areas were used due to sample size 495 limitations. Previous research has also indicated that bat fatalities increased exponentially with tower 496 height [8,40], suggesting that more research is needed on the importance of turbine dimensions. Bat 497 mortality risk has also previously been related to habitat characteristics such as forested areas, slope, 498 temperature, and humidity [41,42], and mountain ridges have been recognized as important during 499 migration [43]. If more studies are completed across a wider range of study conditions, then detecting 500 sources of fatality reduction variation would be more effective. Testing curtailment efficacy at locations 501 with lower overall fatality rates could also be instructive and curtailment studies are suggested for sites 502 that typically have high enough fatality rates to elicit conservation concern.

503 Differentiation of fatality rates by species or species group could also help reduce our uncertainty. 504 Species-level traits such as migratory strategy, dispersal distance, and habitat association likely play an 505 important role in fatality risk [44]. For instance, long-distance migrants such as hoary bats, silver-haired 506 bats, and eastern red bats comprise a majority of fatalities at terrestrial wind energy facilities in North 507 America [3,8]. The project-specific risk is then correlated to species distributions, migratory routes, and 508 flight heights, among other characteristics [45]. Incorporating species-level information could improve 509 our understanding of bat fatality reduction, but this would require that species-level fatality estimates, or 510 at least species-group fatality estimates (i.e., migratory tree bats vs. *Myotis* spp.), be reported from 511 curtailment studies to allow for comparisons. Such estimates were not consistently reported by the studies included in our analysis, often due to insufficient sample size. 512

The precision of *meta-analysis* parameters is likely to be overestimated in this study. While the 513 514 random effects meta-analysis framework adds uncertainty to model estimates based on among-study 515 variance [24], we did not account for site dependence as modeling approaches yielded unstable results. Additionally, turbine operation, mortality estimator selection, and blanket curtailment implementation 516 517 varies substantially between sites (including time of year, time of night, species composition affected, choice of cut-in speed, and turbine feathering), and these differences could affect the results in ways that 518 519 are difficult to incorporate into meta-analyses due to incomplete documentation of these protocols. While 520 we controlled for some of these potential biases by including variables like control cut-in speed and multi-521 treatment controls, the remaining uncertainties will likely be reduced best with increased sample size or 522 protocol documentation.

523 *Recommendations for future studies*

If blanket curtailment greater than 1.5 m/s above manufacturer specifications continues to be implemented at wind facilities, additional experiments should be conducted to understand the relative benefit of these increased cut-in speeds for reducing bat fatalities. The number of studies that tested Δ cutin speeds greater than 1.5 m/s were relatively few, and more studies that target these larger changes are

528 needed. Estimates from the *meta-analysis power analysis* suggest that as many as 25 additional studies at $\Delta 2$ m/s cut-in speed would be needed to effectively exclude the possibility of a 20% reduction in fatality 529 530 (and even more are needed to detect an additional 10% reduction). Conducting studies that compare 531 multiple treatment groups against a control during the same time period at the same location would 532 provide greater inferential power to answer such questions; though the costs of each individual study 533 would increase compared to single treatment studies, more could be gained in terms of understanding the 534 benefits of higher Δ cut-in speeds. At the individual study level, statistical power is dependent on many 535 factors outside of the control of study designers (e.g., fatality rates and carcass persistence). Prior 536 knowledge of these parameters is valuable for designing effective studies, particularly if carcass persistence rates are expected to be lower than average (e.g., due to high scavenging activity at the site). 537 To facilitate inclusion of studies in future meta-analyses, curtailment experiments should report 538 539 fatality estimates for both control and treatment groups, carcass persistence rates, searcher efficiency,

whether feathering occurred), and turbine makes/models, with associated uncertainty values when
relevant. When sample size allows, fatality estimates should be reported by species or species group (e.g., *Myotis*) rather than for all bat species combined to facilitate taxon-specific assessments of curtailment
efficacy.

search frequency, search area coverage, number of turbine-nights of study, curtailment regime (including

540

545 Newer operational minimization strategies have been developed to achieve similar fatality 546 reductions as blanket curtailment but with lower energy loss at higher cut-in speeds [46,47]. "Smart" 547 curtailment strategies, for example, which use additional environmental data besides wind speed to 548 inform the assessment of mortality risk and vary curtailment implementation, show promise to reduce the 549 economic impact of curtailment on wind energy projects [48–50]. Several deterrent systems that 550 discourage bats from approaching turbines are also in development and show some promise for reducing fatalities while minimizing power loss [11,50–52], and could be particularly beneficial if used in 551 combination with curtailment at lower wind speeds. While such approaches are still being evaluated, they 552

may eventually represent a more cost-effective alternative to blanket curtailment, particularly blanketcurtailment at higher wind speeds.

555 *Conclusions*

556 The results of our meta-analysis suggest that blanket curtailment is effective at reducing bat 557 fatalities at terrestrial wind energy facilities, with the meta-analysis describing a mean fatality ratio of 558 0.44, or a 68% reduction in bat fatalities. Given our small sample size, particularly at higher Δ cut-in 559 speeds, our statistical power was limited to test the benefit of increasing cut-in speeds more than 1-1.3 560 m/s above the control cut-in speed. The power analysis suggests that differences in fatality ratio of 50% or 561 greater were often detectable even with small sample sizes (> 80% chance of significance), so it is likely 562 that the true value of incremental increases in Δ cut-in speed is below this 50% threshold. Whitby et al. 563 [37] suggest this is the case and that result combined with our marginally important effect in this study 564 provides more evidence that higher cut-in speeds can yield fewer mortalities. Though the small sample 565 sizes, low power in the present study, and variation in our respective results should engender caution when interpreting these findings. Given the scope of bat fatalities at terrestrial wind farms in North 566 America [3,53], we must learn more about the management effectiveness of curtailment, particularly at 567 larger Δ cut-in speeds. Further development of "smart" curtailment strategies may also reduce fatalities 568 569 while moderating impacts to project finances [49].

570 The number of available studies in the current analysis limited our analytical options and findings in several ways. If blanket curtailment continues to be a common strategy at wind speeds at ~ 5 m/s or 571 572 above (i.e., Δ cut-in speed of >1.5 with a standard factory cut-in speed of 3.5 m/s), we would recommend conducting additional experimental curtailment studies with blanket curtailment treatments at these higher 573 574 cut-in speeds to strengthen our understanding of the relationship between increasing cut-in speeds and bat 575 fatality rates. Such studies must be carefully designed, ideally using an adaptive management framework [54], to consider such variables as the expected fatality rate and carcass persistence rate when selecting a 576 577 search interval and defining the number of turbine-nights to monitor. Studies at sites with expected low

fatality rates and low carcass persistence, in particular, must be carefully designed, and power analyses are an important tool to ensure adequate statistical power to detect changes across treatment and control groups. While such studies would improve our understanding of the relationship between fatality and cutin speed, given the results of our power analysis, a large number of these studies may be required to develop reliable estimates across sites for larger Δ cut-in speeds. While this could be costly, the potential effect of increasing cut-in speed on bat mortality could be ecologically important for species of conservation concern.

585

586 Acknowledgments

The authors are grateful to the Wind Wildlife Research Fund for providing financial support to make this work possible. The American Wind Wildlife Institute provided valuable feedback on earlier drafts of this manuscript along with four anonymous reviewers. We want to thank the American Wind Wildlife Information Center (AWWIC) database manager, Ryan Butryn, for data collation and management, and the AWWI project managers for logistical support and input on the draft report. We would also like to thank all the data contributors who conducted curtailment studies and made the data available for this meta-analysis.

595 Literature Cited

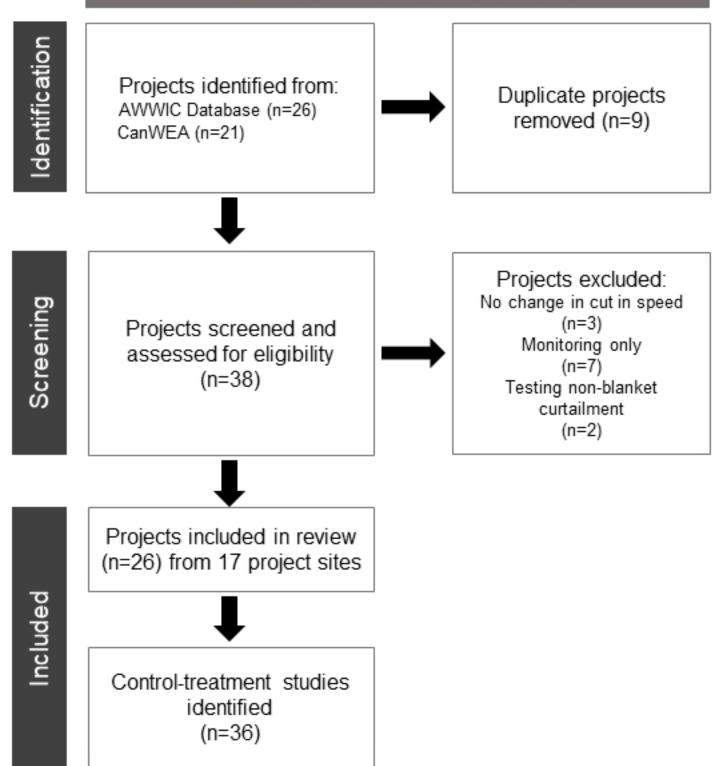
- Arnett EB, Huso MMP, Schirmacher MR, Hayes JP. Altering turbine speed reduces bat mortality at wind-energy facilities. Front Ecol Environ. 2011;9: 209–214. doi:10.1890/100103
- Arnett EB, Baerwald EF. Impacts of wind energy development on bats: Implications for
 conservation. In: Adams RA, Pedersen SC, editors. Bat Evolution, Ecology, and Conservation.
 New York, New York: Springer Science+Business Media; 2013. pp. 435–456. doi:10.1007/978-1 4614-7397-8
- 3. Zimmerling JR, Francis CM. Bat mortality due to wind turbines in Canada. J Wildl Manage.
 2016;80: 1360–1369. doi:10.1002/jwmg.21128
- 604 4. Cryan PM, Barclay RM. Causes of bat fatalities at wind turbines: hypotheses and predictions. J
 605 Mammal. 2009;90: 1330–1340.
- 5. Cryan PM, Gorresen PM, Hein CD, Schirmacher MR, Diehl RH, Huso MM, et al. Behavior of
 bats at wind turbines. Proc Natl Acad Sci. 2014;111: 15126–15131. doi:10.1073/pnas.1406672111
- 6. Baerwald EF, Edworthy J, Holder M, Barclay RMR. A Large-Scale Mitigation Experiment to
 609 Reduce Bat Fatalities at Wind Energy Facilities. J Wildl Manage. 2009;73: 1077–1081.
 610 doi:10.2193/2008-233
- Arnett EB, Baerwald EF. Impacts of Wind Energy Development on Bats: Implications for
 Conservation. In: Adams RA, Pedersen SC, editors. Bat Evolution, Ecology, and Conservation.
 New York, NY: Springer Science+Business Media; 2013. pp. 435–456. doi:10.1007/978-1-46147397-8
- 8. American Wind Wildlife Institute. A summary of bat fatality data in a nationwide database.
 Washington D.C. 41 pp.; 2019. Available: www.awwi.org
- 617 9. Kunz TH, Arnett EB, Cooper BM, Erickson WP, Larkin RP, Mabee T, et al. Assessing impacts of
 618 wind-energy development on nocturnally active birds and bats: A guidance document. J Wildl
 619 Manage. 2007;71: 2449–2486. doi:10.2193/2007-270
- Frick WF, Baerwald EF, Pollock JF, Barclay RM., Szymanski JA, Weller TJ, et al. Fatalities at
 wind turbines may threaten population viability of a migratory bat. Biol Conserv. 2017;209: 172–
 177.
- Arnett EB, Johnson GD, Erickson WP, Hein CD. A synthesis of operational mitigation studies to reduce bat fatalities at wind energy facilities in North America. National Renewable Energy
 Laboratory. Bat Conservation Interational, Austin, Texas. 37 pp.; 2013.
- Baerwald EF, Edworthy J, Holder M, Barclay RMR. A large-scale mitigation experiment to reduce bat fatalities at wind energy facilities. J Wildl Manage. 2009;73: 1077–1081.
- Arnett EB, Huso MMP, Hayes JP, Schirmacher M. Effectiveness of changing wind turbine cut-in speed to reduce bat fatalities at wind facilities. Bats and Wind Energy Cooperative, Bat Conservation International, Austin, Texas. 57 pp.; 2010.
- 631 14. Canadian Wind Energy Association. Wind energy and bat conservation A review by the
 632 Canadian Wind Energy Association. DNVGL. Hovik, Norway. 253 pp.; 2018. Available:
 633 https://canwea.ca/wp-content/uploads/2018/08/windenergyandbatconservationreview.pdf
- 634 15. American Wind Wildlife Institute. Bats and wind energy: Impacts, mitigation, and tradeoffs.
 635 Washington D.C. 42 pp.; 2018.

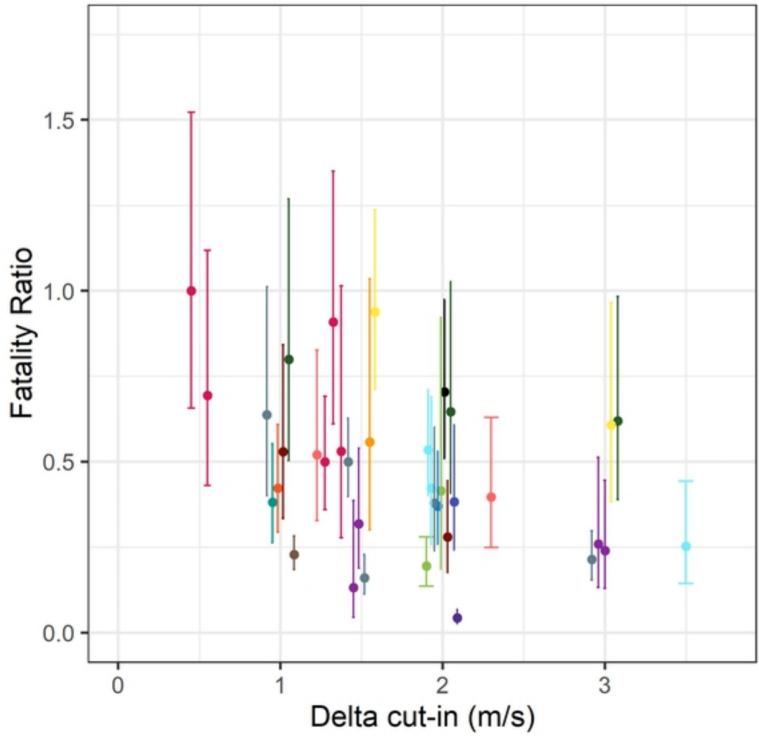
636	16.	Gurevitch J, Koricheva J, Nakagawa S, Stewart G. Meta-analysis and the science of research
637	17	synthesis. Nature. 2018;555: 175–182. doi:10.1038/nature25753
638 639 640	17.	R Core Team. R: a language and environment for statistical computing. Vienna, Austria. http://www.r-project.org: R Foundation for Statistical Computing; 2020. Available: http://www.r-project.org
641 642	18.	Huso M, Dalthorp D, Miller TJ, Bruns D. Wind energy development: methods to assess bird and bat fatality rates post- construction. Human-Wildlife Interact. 2016;10: 62–70.
643 644 645	19.	Bernardino J, Bispo R, Costa H, Mascarenhas M. Estimating bird and bat fatality at wind farms: a practical overview of estimators, their assumptions and limitations. New Zeal J Zool. 2013;40: 63–74. doi:10.1080/03014223.2012.758155
646 647	20.	Huso MMP. An estimator of wildlife fatality from observed carcasses. Environmetrics. 2011;22: 318–329. doi:10.1002/env.1052
648 649 650	21.	Shoenfeld PS. Suggestions regarding avian mortality extrapolation. West Virgina Highl Conserv. Davis, West Virginia. 6 pp.; 2004. Available: http://www.wvhighlands.org/Birds/SuggestionsRegardingAvianMortalityExtrapolation.pdf
651 652 653	22.	Erickson WP, Strickland MD, Johnson GD, Kern JW. Examples of statistical methods to assess risk of impacts to birds from wind plants. Proceedings of National Avian-Wind Power Planning Meeting III. King City, Ontario.; 1998. pp. 172–182.
654 655 656 657	23.	Ontario Ministry of Natural Resources. Bats and bat habitats: guidelines for wind power projects. Ontario, Canada. 24 pp.; 2011. Available: https://tethys.pnnl.gov/sites/default/files/publications/Ontario_Ministry_of_Natural_Resources_20 11.pdf
658 659	24.	Viechtbauer W. Conducting meta-analyses in R with the metafor. J Stat Softw. 2010;36: 1–48. doi:10.18637/jss.v036.i03
660	25.	Cox C. Delta Method. Encycl Biostat. 2005; 2.
661 662	26.	Lajeunesse MJ. On the meta-analysis of response ratios for studies with correlated and multi- group design. Ecology. 2011;92: 2049–2055.
663 664 665	27.	Han B, Duong D, Sul J, de Bakker P, Raychaudhuri S. A general framework for meta-analyzing dependent studies with overlapping subjects in association mapping. Hum Mol Genet. 2016;25: 1857–1866.
666	28.	Schwarzer G. meta: An R package for meta-analysis. R News. 2007;7: 40–45.
667 668	29.	Veroniki A, Jackson D, Viechtbauer W, Bender R, J B, G K, et al. Methods to estimate the between-study variance and its uncertainty in meta-analysis. Res Synth Methods. 2016;7: 55–79.
669 670	30.	Higgins JPT, Thompson SG, Deeks JJ, Altman DG. Measuring inconsistency in meta-analyses. Br Med J. 2003;327: 557–560. doi:10.1136/bmj.327.7414.557
671 672	31.	Arnold BF, Hogan DR, Colford Jr. JM, Hubbard AE. Simulation methods to estimate design power: an overview for applied research. BMC Med Res Methodol. 2011;11: 94.
673 674	32.	Gelman A, Carlin J. Beyond power calculations: assessing Type S (Sign) and Type M (Magnitude) errors. Perpectives Psychol Sci. 2014;9: 641–651.
675	33.	Kery MA, Royle JA. Applied hierarchical modeling in ecology: Analysis of distribution,

676		abundance and species richness in R and BUGS. London: Academic Press; 2016.
677 678	34.	Brilleman S. simsurv: Simulate Survival Data. https://cran.r-project.org/package=simsurv: R package version 0.2.3; 2019. Available: https://cran.r-project.org/package=simsurv
679 680 681	35.	Dalthorp D, Simonis J, Madsen L, Huso M, Rabie P, Minz J, et al. GenEst: Generalized Mortality Estimator. https://cran.r-project.org/package=GenEst: R package version 1.2.3; 2019. Available: https://cran.r-project.org/package=GenEst
682 683 684	36.	Arnett EB, Johnson GD, Erickson WP, Hein CD. A Synthesis of Operational Mitigation Studies to Reduce Bat Fatalities at Wind Energy Facilities in North America. Austin, TX; 2013. doi:10.1016/j.jallcom.2014.09.009
685 686 687	37.	Whitby MD, Schirmacher MR, Frick WF. The State of the Science on Operational Minimization to Reduce Bat Fatality at Wind Energy Facilities. A Rep Submitt to Natl Renew Energy Lab. Austin, Texas; 2021.
688 689 690	38.	Bulling L, Köppel J. Exploring the trade-offs between wind energy and biodiversity conservation. In: Geneletti D, editor. Handbook on Biodiversity and Ecosystem Services in Impact Assessment. 2016. p. 299.
691 692	39.	Horn JW, Arnett EB, Kunz TH. Behavioral responses of bats to operating wind turbines. J Wildl Manage. 2008;72: 123–132. doi:10.2193/2006-465
693 694 695	40.	Barclay RMR, Baerwald EF, Gruver JC. Variation in bat and bird fatalities at wind energy facilities: assessing the effects of rotor size and tower height. Can J Zool. 2007;85: 381–387. doi:10.1139/Z07-011
696 697	41.	Santos H, Rodrigues L, Jones G, Rebelo H. Using species distribution modelling to predict bat fatality risk at wind farms. Biol Conserv. 2013;157: 178–186. doi:10.1016/j.biocon.2012.06.017
698 699 700	42.	Rydell J, Bach L, Dubourg-Savage M-J, Green M, Rodrigues L, Hedenström A. Bat mortality at wind turbines in northwestern Europe. Acta Chiropterologica. 2010;12: 261–274. doi:10.3161/150811010x537846
701 702	43.	Baerwald EF, Barclay RMR. Geographic variation in activity and fatality of migratory bats at wind energy facilities. J Mammal. 2009;90: 1341–1349. doi:10.1644/09-mamm-s-104r.1
703 704 705	44.	Thaxter CB, Buchanan GM, Carr J, Butchart SHM, Newbold T, Green RE, et al. Bird and bat species' global vulnerability to collision mortality at wind farms revealed through a trait-based assessment. Proc R Soc B Biol Sci. 2017;284. doi:10.1098/rspb.2017.0829
706 707	45.	Roemer C, Disca T, Coulon A, Bas Y. Bat flight height monitored from wind masts predicts mortality risk at wind farms. Biol Conserv. 2017;215: 116–122. doi:10.1016/j.biocon.2017.09.002
708 709 710	46.	Arnett E, Hein C, Schirmacher M, Huso M, Szewczak J. Correction: evaluating the effectiveness of an ultrasonic acoustic deterrent for reducing bat fatalities at wind turbines. PLoS One. 2013;8: 10.1371/annotation/a81f59cb-0f82-4c84-a743-895acb4.
711 712 713	47.	Behr O, Brinkmann R, Hochradel K, Mages J, Korner-Nievergelt F, Niermann I, et al. Mitigating bat mortality with turbine-specific curtailment algorithms: A model based approach. Wind Energy and Wildlife Interactions. Springer; 2017. pp. 135–160.
714 715	48.	Martin CM, Arnett EB, Stevens RD, Wallace MC. Reducing bat fatalities at wind facilities while improving the economic efficiency of operational mitigation. J Mammal. 2017;98: 378–385.
716	49.	Hayes MA, Hooton LA, Gilland KL, Grandgent C, Smith RL, Lindsay SR, et al. A smart

717 curtailment approach for reducing bat fatalities and curtailment time at wind energy facilities. Ecol 718 Appl. 2019;29: 1-18. doi:10.1002/eap.1881 50. Weaver SP. Understanding Wind Energy Impacts to Bats and Testing Reduction Strategies in 719 720 South Texas. Texas State University. 2019. Available: https://digital.library.txstate.edu/handle/10877/8463 721 51. Johnson JB, Ford WM, Rodrigue JL, Edwards JW. Effects of acoustic deterrents on foraging bats. 722 Res Note NRS-129. Newtown Square, PA. 5 pp.: U.S. Departmenf of Agriculture, Forest Service, 723 Northern Research Station; 2012. Report No.: Res. Note NRS-129. 724 Romano WB, Skalski JR, Townsend RL, Kinzie KW, Coppinger KD, Miller MF. Evaluation of an 725 52. 726 acoustic deterrent to reduce bat mortalities at an Illinois wind farm. Wildl Soc Bull. 2019:43: 608-618. doi:10.1002/wsb.1025 727 Hayes MA. Bats killed in large numbers at United States wind energy facilities. Bioscience. 728 53. 729 2013;63: 975-979. doi:10.1093/biosci/biu056 730 54. Williams BK. Adaptive management of natural resources-framework and issues. J Environ Manage. 2011;92: 1346-1353. 731

Identification of Studies

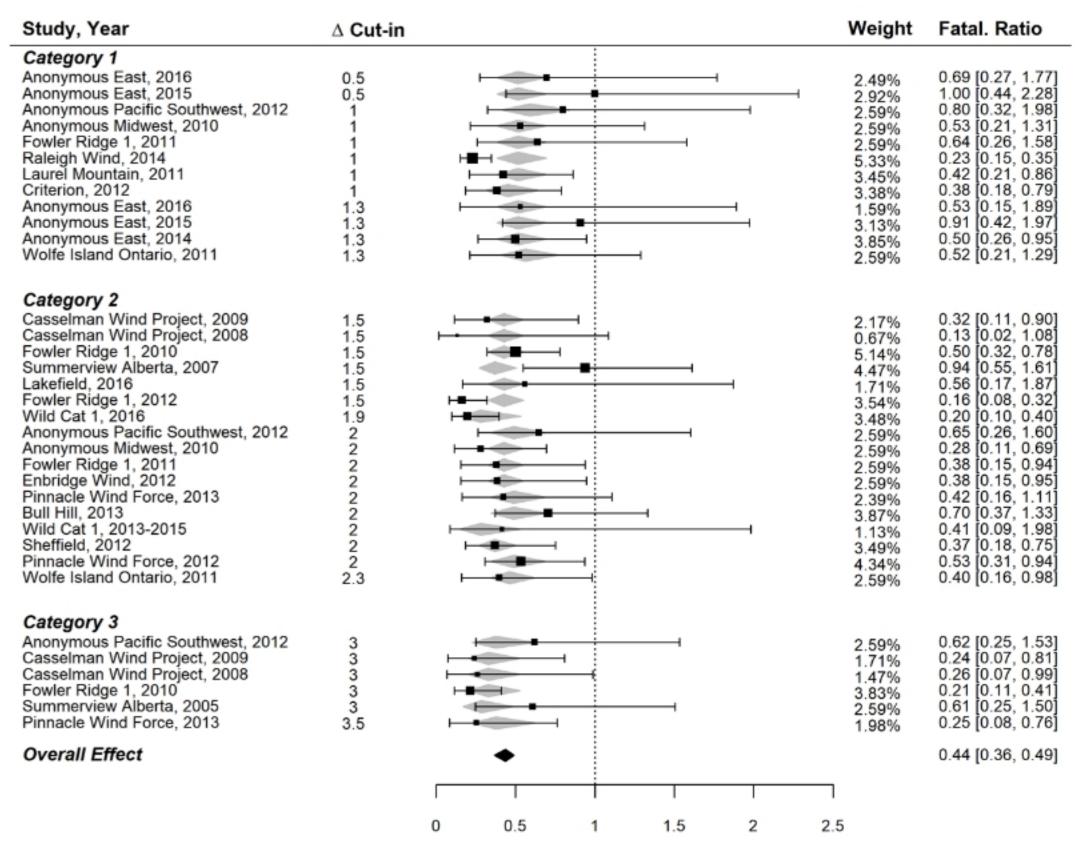




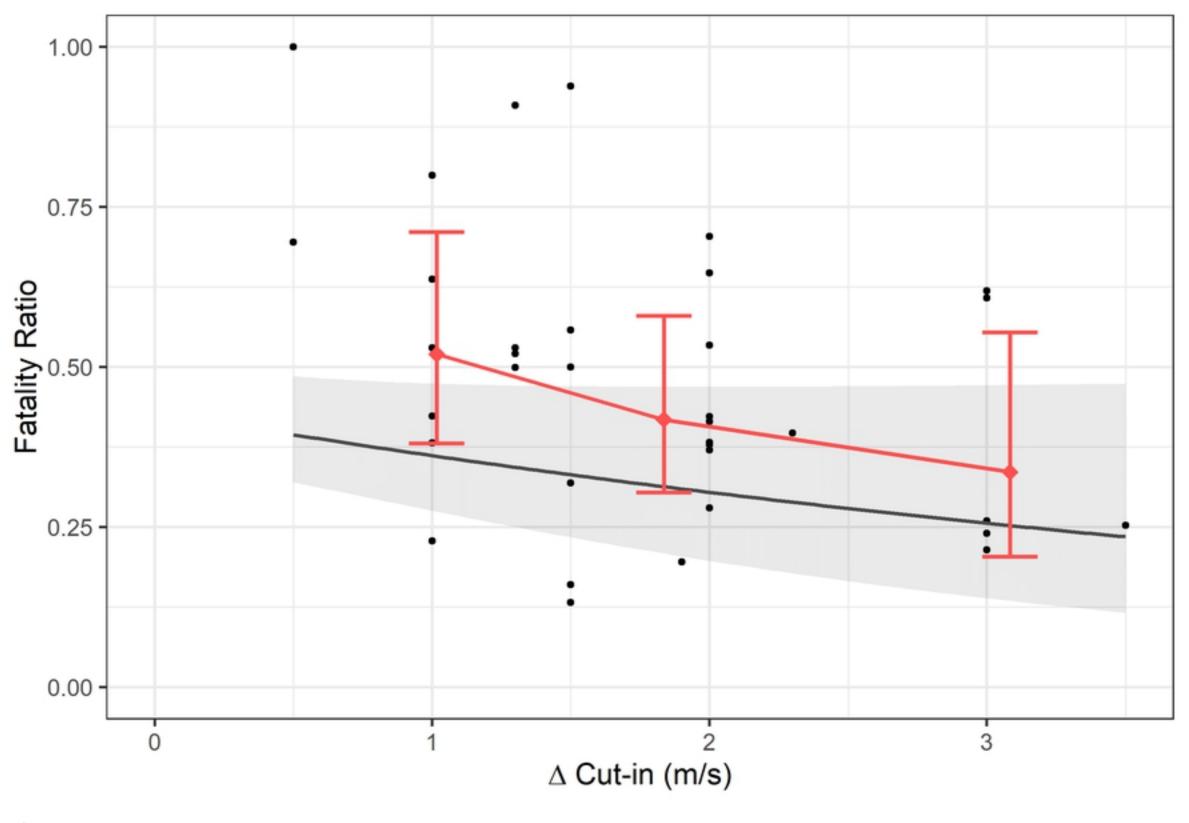
Project

- Anonymous Midwest
- Anonymous Pacific Southwest
- Beech Ridge
- ← Bull Hill
- Casselman Wind Project
- Criterion
- Enbridge Wind
- Fowler Ridge 1
- Lakefield
- Laurel Mountain
- Pinnacle Wind Force
- Raleigh Wind
- Sheffield
- Summerview Alberta
- Talbot Wind
- Wild Cat 1
- Wolfe Island Ontario

Figure 2



Fatality Ratio



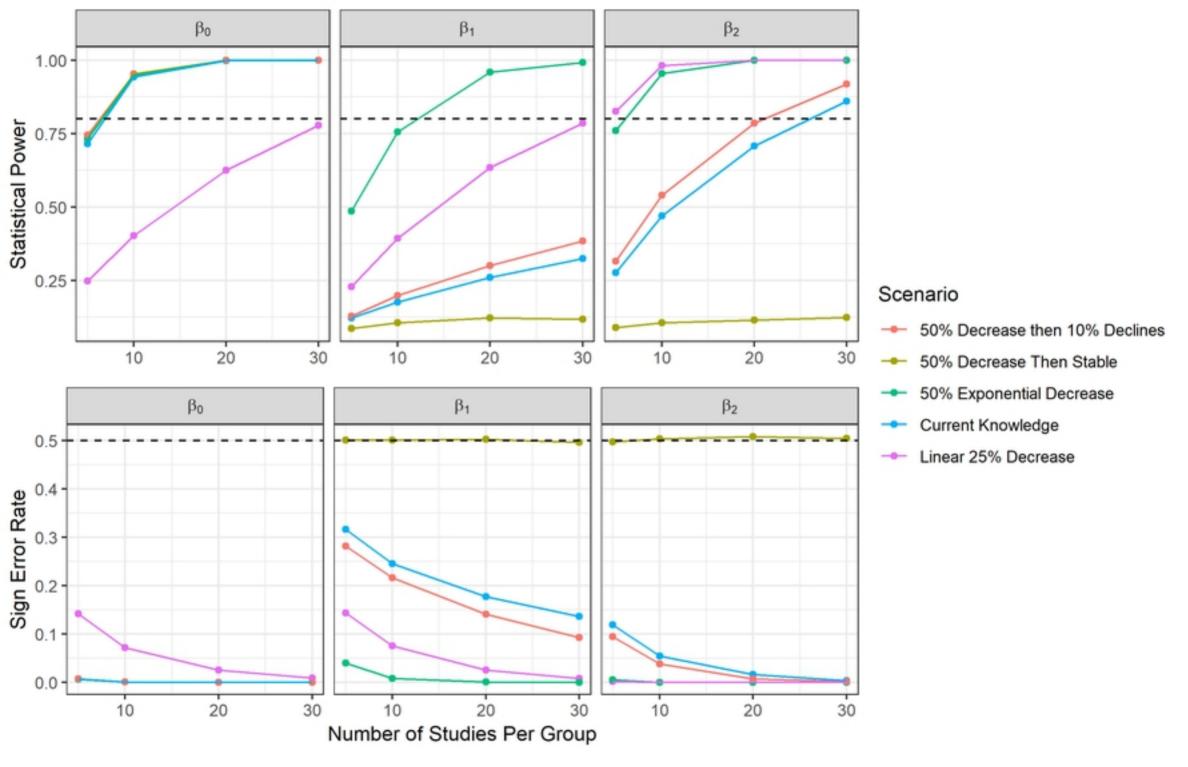


Figure 5

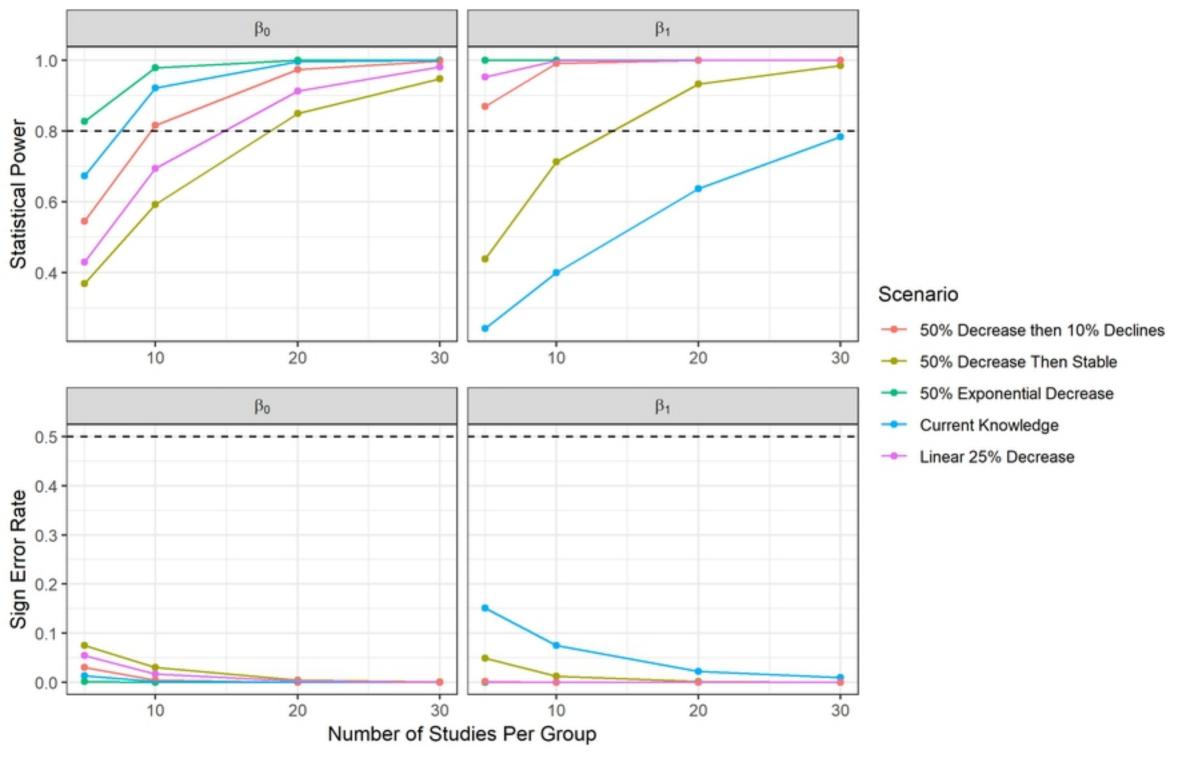


Figure 6

