

1 **Parallel evolution of urban-rural clines in melanism in a widespread mammal**

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## 12 **Abstract**

13

14 Urbanization is the dominant trend of global land use change. The replicated nature of  
15 environmental change associated with urbanization should drive parallel evolution, yet insight  
16 into the repeatability of evolutionary processes in urban areas has been limited by a lack of  
17 multi-city studies. Here we leverage community science data on coat color in >60,000 eastern  
18 gray squirrels (*Sciurus carolinensis*) across 43 North American cities to test for parallel clines in  
19 melanism, a genetically based trait associated with thermoregulation and crypsis. We show the  
20 prevalence of melanism was positively associated with urbanization as measured by impervious  
21 cover. Urban-rural clines in melanism were strongest in the largest cities with extensive forest  
22 cover and weakest or absent in cities with warmer winter temperatures, where thermal selection  
23 likely limits the prevalence of melanism. Our results suggest that novel traits can evolve in a  
24 highly repeatable manner among urban areas, modified by factors intrinsic to individual cities,  
25 including their size, land cover, and climate.

26

## 27 **Introduction**

28

29 Urban land use is expanding at a rate faster than urban population growth<sup>1</sup>, transforming land  
30 cover, climate, hydrology, and biogeochemical cycling<sup>2</sup>. The dramatic environmental change  
31 associated with urbanization renders cities similar to one another in many abiotic and biotic  
32 characteristics<sup>3,4</sup>. For example, cities have extensive impervious cover (e.g., buildings,  
33 transportation networks), increased air temperature, and increased air, water, light, and noise  
34 pollution relative to their adjacent rural landscapes<sup>2</sup>.

35           The convergence of environmental conditions in cities makes urban ecosystems an ideal  
36 focus for addressing longstanding questions in evolutionary biology about the degree to which  
37 populations exhibit parallel evolution, that is, when multiple populations occupying similar  
38 environments evolve comparable traits<sup>5</sup>. Cities are thought to drive parallel evolution as urban  
39 populations adapt to similar selection pressures<sup>6,7,8,9</sup>. Cities are not homogenous, however, and  
40 variation in their characteristics likely mediates processes that facilitate or inhibit parallel  
41 evolution, generating variation in the shape of genotypic or phenotypic clines along urbanization  
42 gradients. Clines may be strongest in large cities where the environmental gradients driving  
43 divergent selection between urban and rural areas are steepest. Conversely, clines may be weak  
44 or nonexistent in small cities or cities with extensive greenspace connecting urban and rural  
45 populations via gene flow. Environmental conditions at large spatial scales – such as regional  
46 climate – may also modulate the evolution of urban-rural clines at a smaller scale. For example,  
47 directional selection for traits that reduce thermal stress at climate extremes may overwhelm  
48 selective pressures associated with local land use, weakening urban-rural clines<sup>10</sup>. Advances in  
49 our understanding of how landscape-scale differences among cities facilitate or constrain parallel  
50 evolution have been limited by a lack of studies conducted across multiple, replicate cities<sup>7,8,9</sup>.  
51 Addressing this knowledge gap will help clarify how urbanization, the megatrend of global land  
52 use change, will shape biodiversity.

53           Here we test for urban-rural clines as a signal of parallel evolution and examine how  
54 landscape features mediate the shape of clines. Our focus is on melanism in eastern gray  
55 squirrels (*Sciurus carolinensis*) in 43 cities in North America (Fig. 1, Supplementary Fig. 1).  
56 Gray squirrels are among the most ubiquitous of fauna in cities and rural woodlands in eastern  
57 North America<sup>11</sup>. Coat color in gray squirrels is typically gray or black (melanic) and inherited in

58 a simple Mendelian fashion at the melanocortin 1 receptor gene (*Mc1R*)<sup>12</sup>. Evidence suggests  
59 divergent selection associated with crypsis, predation risk, and road mortality favors the melanic  
60 morph in cities and the gray morph in rural woodlands<sup>13</sup>, which should cause a clinal decrease in  
61 melanism from urban to rural areas. However, urban-rural divergence in melanism has been  
62 documented in only two cities (Syracuse, NY<sup>13</sup> and Wooster, OH<sup>14</sup>).

63 We used a dataset on coat color from squirrel observations collected via the community  
64 science project *SquirrelMapper*<sup>13</sup> to test for parallel clines in melanism along gradients of  
65 impervious cover. We included city size, forest cover across each urbanization gradient, and  
66 winter temperature in a linear mixed model to test how these landscape-scale characteristics  
67 affect the prevalence of melanism in each city (i.e., main effects of each covariate) and drive  
68 variation in cline shape among cities (i.e., interaction effects with impervious cover). Gray  
69 squirrels can disperse long distances (>10 km)<sup>15</sup>, but movements are limited by forest  
70 fragmentation<sup>16</sup>, so we expected clines to be weakest in small cities and cities with extensive  
71 forest cover where gene flow between urban and rural areas should be greatest. We included  
72 winter temperature in the model because melanic squirrels are known to have greater  
73 thermogenic capacity than the gray morph at cold temperature<sup>17</sup>. Clines may be weak in cities  
74 with cold winter temperature if thermal selection overwhelms the selection pressures on  
75 melanism associated with urbanization.

76

## 77 **Results and Discussion**

78

79 Our analyses revealed that a positive relationship between the prevalence of melanism and  
80 degree of urbanization was replicated across multiple cities in North America (main effect of

81 impervious cover = 0.17, SE = 0.04,  $P < 0.001$ ; Figs. 1-2; Supplementary Table 2). Repeated  
82 shifts in phenotype along environmental gradients – such as the general increase in melanism  
83 from rural to urban areas documented here – is a hallmark of parallel evolution. Melanism in  
84 gray squirrels is a simple Mendelian trait coded by a 24-bp deletion at *Mc1R*<sup>12</sup>. Individuals with  
85 at least one copy of the deletion allele are melanic, so clines in melanism very likely occur at  
86 both phenotypic and genotypic levels.

87         The presence of parallel clines in heritable traits is often considered evidence of adaptive  
88 evolution because gene flow should homogenize populations in the absence of natural  
89 selection<sup>18</sup>. Known drivers of parallel adaptive evolution in cities include the urban heat island<sup>19</sup>  
90 and pollution<sup>20</sup>. Evidence suggests multiple selective mechanisms involving crypsis help  
91 maintain clines in melanism in gray squirrels. Predation – including hunting of squirrels by  
92 humans – is often the main driver of squirrel mortality in rural woodlands<sup>21,22,23</sup>, and the gray  
93 morph is more concealed than the melanic morph in regrown, secondary forests that dominate  
94 rural woodlands in our study area<sup>13</sup>. In contrast, predation pressure on tree squirrels tends to be  
95 relaxed in cities<sup>21,22</sup>, and relaxed selection against melanic squirrels in cities could generate the  
96 observed urban-rural clines. On the other hand, novel selection pressures in cities may favor  
97 melanics. The primary driver of mortality of tree squirrels in cities tends to be vehicular  
98 collisions<sup>22</sup>. The melanic morph is more visible to vehicle drivers than the gray morph on asphalt  
99 roads, which may lead to the melanic morph having greater survival in areas with high car  
100 traffic<sup>13</sup>. Additional selective mechanisms may play a role in maintaining urban melanism (e.g.,  
101 parasite pressure, pollution)<sup>24</sup>, including mechanisms involving correlated traits<sup>25</sup>. For example,  
102 melanic squirrels have less phylogenetically variable microbiomes than the gray morph in  
103 cities<sup>26</sup>. A more complete picture of the selective mechanisms driving urban-rural clines will

104 require linking trait variation to fitness differences between squirrel color morphs in urban and  
105 rural areas.

106 An alternative explanation for the urban-rural clines documented here is that melanic  
107 squirrels have been introduced by humans to some cities<sup>23</sup>. Strong genetic drift via founder  
108 events, combined with restricted gene flow along the urbanization gradient, can generate single-  
109 locus clines<sup>27</sup>. Variation in the size of founding populations likely causes some variation in the  
110 prevalence of melanism among cities, and thus the strength of urban-rural clines. At an extreme,  
111 urban-rural clines are absent in cities where melanism was historically absent and melanic  
112 squirrels have never been introduced. However, two factors lead us to believe that urban founder  
113 events are not the driving force behind the observed urban-rural clines. First, single-locus clines  
114 generated by neutral evolutionary forces require strong isolation-by-distance (IBD)<sup>27</sup>. We  
115 suspect IBD between urban and rural areas exists but is limited in strength by high effective  
116 population size in cities (e.g., urban squirrel densities 2.5 times greater than rural squirrel  
117 densities<sup>28</sup>), high capacity for long-distance dispersal<sup>15</sup>, and frequent translocation of squirrels<sup>29</sup>.  
118 Second, at northern latitudes where we observe most urban-rural clines, melanism was common  
119 in rural old growth forests prior to the 1800s<sup>30,31,32</sup>, including regions where it is now rare except  
120 in cities (e.g., upstate New York<sup>13,31</sup>). Founder events in cities do not explain the paucity of  
121 melanics in rural areas relative to the past. We hypothesize that the dramatic shift from old  
122 growth to secondary forests following European settlement in North America<sup>33,34,35</sup> changed the  
123 degree to which gray and melanic morphs are concealed from visual predators and hunters,  
124 favoring the gray morph in regrown forests<sup>13</sup>. Genomic analysis and field experiments are  
125 needed to clarify the role of natural selection versus nonadaptive processes in generating the  
126 observed urban-rural clines<sup>36</sup>.

127           Although clines in melanism tended to take the same form among cities, the strength of  
128 urban-rural clines varied, supporting the idea that repeated evolution is often inconsistent<sup>5</sup>. First,  
129 there was a strong negative relationship between melanism and winter temperature (main effect  
130 of winter temperature = -2.14, SE = 0.33,  $P < 0.001$ ; Figs. 1-2; Supplementary Table 2) that  
131 likely constrains the development of urban-rural clines. Urban-rural clines in melanism were  
132 most common in northern cities with cold winter temperature (Fig. 1b), whereas clines were  
133 absent in cities with warm winters where melanism was rare or nonexistent. The melanic morph  
134 has greater capacity for non-shivering thermogenesis and retains more heat than the gray morph  
135 at cold temperature<sup>17,37,38</sup>. Our results suggest thermal selection at a regional scale plays an  
136 important role in maintaining melanism in the north, allowing other processes operating at a finer  
137 spatial scale to cause the evolution of urban-rural clines. Winter temperature has also been  
138 shown to constrain the evolution of urban-rural clines in cyanogenesis in white clover (*Trifolium*  
139 *repens*)<sup>39</sup>. Importantly, our results do not suggest urban-rural clines in melanism in gray squirrels  
140 are caused by the urban heat island effect, which should cause the opposite clinal pattern than  
141 that observed.

142           Second, the positive relationship between urbanization and melanism was strongest in  
143 large cities with extensive forest cover (impervious\*city size interaction effect = 0.13, SE =  
144 0.03,  $P < 0.001$ ; impervious\*forest cover interaction effect = 0.15, SE = 0.03,  $P < 0.001$ ; Fig. 2;  
145 Supplementary Table 2), where melanism increased by nearly 20% from rural to urban areas  
146 (Fig. 2c). In contrast, clines were weak to nonexistent in smaller cities and those with less  
147 extensive forest cover (Fig. 2c). Although gray squirrels have substantial capacity for long-  
148 distance dispersal, gene flow between urban and rural areas should be most limited in large cities  
149 due to isolation-by-distance and mortality caused by extensive road networks during dispersal.

150 Gene flow may be strongest between urban and rural populations in smaller cities, maintaining  
151 maladapted morphs across the urbanization gradient<sup>9</sup>. Selection associated with road mortality in  
152 urban areas may also be weak in small cities with limited traffic, which could explain the lower  
153 prevalence of melanism in smaller cities (main effect of city size = 0.58, SE = 0.28,  $P = 0.037$ ;  
154 Fig. 2). Mechanistic studies are needed to provide insight into how road networks, traffic  
155 volume, and road-crossing structures (e.g., aerial utility lines that squirrels use to cross roads<sup>23</sup>)  
156 contribute to variation in gene flow and selection along urbanization gradients within and among  
157 cities.

158 The finding of strong clines in cities with extensive forest cover (Fig. 2) did not support  
159 our prediction that forest cover facilitates gene flow and weakens clines along the urbanization  
160 gradient. Gray squirrels require trees, and squirrel abundance and distribution along urbanization  
161 gradients is limited by tree cover<sup>28</sup>, particularly in large cities<sup>40</sup>. It is possible that cities with  
162 more extensive forest cover support a greater abundance of gray squirrels, increasing effective  
163 population size and thereby making selection more efficient in overcoming the effects of genetic  
164 drift on morph frequencies.

165 Community science datasets in ecology and evolution are powerful for conducting  
166 studies at large spatial scales and with levels of replication that would otherwise be infeasible.  
167 However, one challenge of using community science datasets is the potential for sampling  
168 biases<sup>41</sup>. For example, *SquirrelMapper* observers may be more likely to record observations of  
169 rare color morphs. Melanics were less common than the gray morph in most cities, so detection  
170 bias for rare morphs would inflate the estimated prevalence of melanism. However, we are  
171 unaware of a mechanism that would cause sampling bias to vary between urban and rural areas  
172 in a way that artificially generates the urban-rural clines documented here (e.g., bias toward



173 recording the melanic morph in urban areas and gray morph in rural areas), so we have no reason  
174 to believe the relative change in melanism along urbanization gradients is biased.

175 Our work reinforces the idea that ecological homogenization among cities likely drives  
176 parallel evolution<sup>19,20</sup>, and it suggests cities can function as refuges for unique phenotypic  
177 traits<sup>42</sup>. Our study also provides insight into how landscape-scale differences among cities alter  
178 phenotypic clines along urbanization gradients. We show the degree of parallelism in urban-rural  
179 clines among cities is imperfect, with environmental heterogeneity among cities contributing to  
180 variation in the shape of clines. Greenspace availability in particular has been a focus in the  
181 ecological literature for understanding among-city variation in the response of species  
182 distributions to urbanization<sup>40</sup> – our results show that greenspace availability in the form of forest  
183 cover likely mediates evolutionary responses to urbanization as well. Additional multi-city  
184 studies will be essential for understanding how the degree of parallel evolution depends on  
185 factors intrinsic to individual cities, including those not examined here (e.g., forest community  
186 composition). Finally, our work underscores the value of leveraging community science data to  
187 conduct studies of urban evolution across multiple cities.

188

## 189 **Methods**

190

191 **Squirrel datasets.** Squirrel records were sourced from the community science program  
192 *SquirrelMapper* ([squirrelmapper.org](http://squirrelmapper.org)). Starting in 2010, participants used a web platform to  
193 submit sightings of *S. carolinensis*, including the date of observation and color morph of each  
194 squirrel (“gray”, “melanic”, or “other”). Spatial coordinates for each observation were secured  
195 via Google Maps JavaScript API. A total of 5,899 records were submitted to the original

196 *SquirrelMapper* program from the native range of *S. carolinensis* in North America. Beginning  
197 in 2019, *SquirrelMapper* was integrated with *iNaturalist*, enabling users to submit georeferenced  
198 photos of each squirrel observation. All images of *S. carolinensis* submitted to *iNaturalist*  
199 through 21 December 2020 and classified as “research grade” were included for analysis (n =  
200 75,130). Data from the original *SquirrelMapper* program were merged with the *iNaturalist*  
201 dataset for analyses.

202 We used the participatory science portal *Zooniverse* to crowdsource the classification of  
203 coat color of squirrels in images submitted to *iNaturalist*. Participants were asked to classify the  
204 number of squirrels observed in each image (“zero”, “one”, “two or more”) and the coat color of  
205 squirrels present in each image (“gray”, “melanic”, “other”, or “unclear”). Each image was  
206 classified by a minimum of 10 participants (range: 10-291; median = 10). We retained images  
207 with a minimum agreement threshold of 80% for the number of squirrels observed. Of the  
208 72,239 images (96%) that met the agreement threshold, 865 were classified as having no squirrel  
209 present (e.g., track or other sign), 70,258 had a single squirrel, and 1,116 had two or more  
210 squirrels. Of the images classified as having at least one squirrel, we retained 63,506 images that  
211 had a minimum agreement threshold of 80% for coat color of at least one squirrel in the image.  
212 We removed 5,182 images that did not have location data available and 4,035 images taken  
213 outside of North America where *S. carolinensis* is introduced. The final *iNaturalist* dataset  
214 consisted of 54,289 images, of which 54,215 were classified as having a single coat color present  
215 (gray: 46,630, melanic: 6,728, other: 857) and 74 had two color morphs present (gray and  
216 melanic: 68, gray and other: 5, melanic and other: 1). A random sample of 500 images classified  
217 by the authors showed that the classification scheme we used with 80% agreement thresholds  
218 was 100% accurate for the number of squirrels observed in the image and 99.8% accurate for

219 coat color, the only discrepancy being a single image classified as melanic that the authors  
220 classified as unclear. The final *iNaturalist* dataset was merged with squirrel records from the  
221 original *SquirrelMapper* project, yielding 60,262 squirrel records in total. Multiple coat colors  
222 observed at the same location were treated as separate records in this final dataset. Observations  
223 in the final dataset were made between 1980 and 2020 (median = 2019), with 97% of  
224 observations being reported since 2010. The 3% of observations prior to 2010 consisted of  
225 *iNaturalist* records and records by the authors and colleagues prior to *SquirrelMapper*.

226

227 **City selection.** We selected cities in the native range of *S. carolinensis* that had at least 100  
228 squirrel observations in the city and surrounding rural area. We defined the spatial extent of each  
229 city's footprint using the Urban Area National Shapefile Record Layout<sup>43</sup> for the United States  
230 and the Population Center Boundary File<sup>44</sup> for Canada. We calculated the distance from the  
231 centroid of each city footprint to its maximum extent in each of the four cardinal directions. We  
232 then buffered each urban area by 25% of the average distance from the city centroid to its extent  
233 (Supplementary Fig. 2). This buffer distance was large enough to include a substantial area of  
234 rural land surrounding each city but not so large as to produce excessive overlap between  
235 polygons of adjacent cities. We determined the number of squirrels in each city polygon  
236 (encompassing the city footprint and hinterland) after spatially thinning the squirrel records to a  
237 minimum distance of 10 m using the *spThin* package (v. 0.2.0)<sup>45</sup> in R<sup>46</sup>, which helps minimize  
238 spatial sampling bias. If two cities with  $\geq 100$  squirrels had polygons that overlapped, we  
239 retained the city with a greater sample size or spatial spread of squirrel observations. The final  
240 dataset included 26,924 squirrels among 43 cities (Fig. 1), and the median number of squirrels in  
241 each city was 254 (range: 101 – 4,731; Supplementary Table 1; maps of squirrel observations in

242 each city available for download at <https://github.com/bcosentino/squirrelMapper-urban-rural->  
243 [clines](#)).

244

245 **Environmental covariates.** We measured percent impervious cover within 1 km around each  
246 squirrel observation using the Global Man-made Impervious Surface Dataset (30-m resolution)<sup>47</sup>.  
247 A 1-km buffer was chosen to encompass the typical home range size of *S. carolinensis* (<5 ha)  
248 and long-distance movements typical of males during winter and summer mating seasons<sup>48</sup>. For  
249 each city and its surrounding rural area we measured forest cover and winter temperature. Forest  
250 cover was measured using the Global 2010 Tree Cover database (30-m resolution)<sup>49</sup>, and we  
251 used WorldClim (version 2.1) to measure winter temperature as mean temperature of the coldest  
252 quarter (BIO11, 30s / 1 km resolution)<sup>50</sup>. We also measured the area of each city footprint as a  
253 metric of city size. All spatial analyses were conducted in R using the *raster* (v. 3.3-13)<sup>51</sup> and  
254 *exactextractr* packages (v. 0.5.1)<sup>52</sup>.

255

256 **Statistical analyses.** We fit a generalized linear mixed model with a binomial error distribution  
257 to test for effects of impervious cover, city size, winter temperature, and forest cover on the  
258 distribution of melanism. Melanism was the response variable and represented by a binary  
259 indicator, with 1 corresponding to melanic and 0 to other coat colors for each squirrel  
260 observation. Impervious cover within 1 km of each observation was included as a fixed effect to  
261 represent each city's urbanization gradient. Because the range of impervious cover varied among  
262 cities, we rescaled impervious cover within each city to range from 0 (lowest impervious cover)  
263 to 1 (greatest impervious cover)<sup>39</sup>. This ensured the slope term for impervious cover has a  
264 consistent interpretation among cities, allowing us to test for a general urban-to-rural cline in

265 melanism. City size (log-transformed), forest cover, and winter temperature were included as  
266 fixed effects to test whether these variables predict variation in the prevalence of melanism  
267 among cities. Two-way interaction terms were included between each predictor and impervious  
268 cover to test how city characteristics affect the shape of urban-to-rural clines in melanism. We  
269 standardized each predictor so that each variable had mean = 0 and SD = 1, which aided in  
270 model convergence and allowed us to compare the effect size of each slope term<sup>53</sup>. We also  
271 included city as a random effect to account for the nested observations within each city and to  
272 estimate unexplained variance in the prevalence of melanism among cities (Supplementary Table  
273 3). The model was fit with the *lme4* package in R (v. 1.1-26)<sup>54</sup>. Two-sided Wald tests were  
274 conducted for each parameter in the model. Binned residual plots were generated using the *arm*  
275 package (v. 1.11-2)<sup>55</sup> as recommended for models with a binary response variable<sup>56</sup>. Plots of  
276 binned residuals on fitted values and each predictor suggested adequate model fit and no  
277 substantial deviation from model assumptions.

278 We used the residual autocovariate approach<sup>57</sup> to account for spatial autocorrelation in  
279 the residuals of the model. We extracted residuals from the global model and created a spatial  
280 autocovariate for the residuals using the *spdep* package (v. 1.1-7)<sup>58</sup>. We set the radius  
281 neighborhood to 31 km (the greatest nearest neighbor distance in the thinned dataset) and used  
282 an inverse distance weighting scheme with style “B” for symmetric weights<sup>59</sup>. Our global model  
283 was then refit with the spatial autocovariate term as an additional predictor, which reduced  
284 spatial autocorrelation in the residuals from a maximum of  $r = 0.11$  within 1 km to  $r < 0.05$   
285 (Moran’s  $I$ ; Supplementary Fig. 3).

286 We refit our model using different spatial thinning distances (10 m, 50 m, 100 m) and  
287 buffer distances to measure impervious cover (500 m, 1 km, 10 km). These analyses showed the

288 model results were not sensitive to spatial thinning distances (Supplementary Table 4) or buffer  
289 distances to quantify impervious cover around squirrel observations (Supplementary Table 5).

290

## 291 **References**

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- 293 1. Angel, S. et al. The dimensions of global urban expansion: Estimates and projections for all  
294 countries, 2000–2050. *Prog. Plann.* **75**, 53-107 (2011).
- 295 2. Grimm, N. B. *et al.* Global change and the ecology of cities. *Science* **319**, 756-760 (2008).
- 296 3. McKinney, M. L. Urbanization as a major cause of biotic homogenization. *Biol. Conserv.*  
297 **127**, 247-260 (2006).
- 298 4. Groffman, P. M. *et al.* Ecological homogenization of urban USA. *Front. Ecol. Environ.* **12**,  
299 74-81 (2014).
- 300 5. Bolnick, D. I. *et al.* (Non)Parallel evolution. *Annu. Rev. Ecol. Evol. Syst.* **49**, 303-330 (2018).
- 301 6. Donihue, C. M. & Lambert, M. R. Adaptive evolution in urban ecosystems. *Ambio* **44**, 194-  
302 203 (2015).
- 303 7. Johnson, M. T. J. & Munshi-South, J. Evolution of life in urban environments. *Science* **358**,  
304 eaam8327 (2017).
- 305 8. Rivkin, L. R. *et al.* A roadmap for urban evolutionary ecology. *Evol. Appl.* **12**, 384-398  
306 (2019).
- 307 9. Santangelo, J. S. *et al.* Urban environments as a framework to study parallel evolution in  
308 *Urban Evolutionary Biology*, Szulkin, M., Munshi-South, J. & Charmantier, A. (Eds).  
309 (Oxford University Press, 2020).
- 310 10. Cosentino, B. J., Moore, J-D., Karraker, N. E., Ouellet, M. & Gibbs, J. P. Evolutionary

- 311 response to global change: climate and land use interact to shape color polymorphism in a  
312 woodland salamander. *Ecol. Evol.* **7**, 5426-5434 (2017).
- 313 11. Koprowski, J. L., Munroe, K. E. & Edelman, A. J. Gray not grey: ecology of *Sciurus*  
314 *carolinensis* in their native range in North America in *Grey Squirrels: Ecology and*  
315 *Management of an Invasive Species in Europe*, Shuttleworth, C. M., Lurz P. W. W. &  
316 Gurnell, J (Eds). (European Squirrel Initiative, 2016).
- 317 12. McRobie, H., Thomas, A. & Kelly, J. The genetic basis of melanism in the gray squirrel  
318 (*Sciurus carolinensis*). *J. Hered.* **100**, 709-714 (2009).
- 319 13. Gibbs, J. P., Buff, M. F. & Cosentino, B. J. The biological system – Urban wildlife,  
320 adaptation and evolution: urbanization as a driver of contemporary evolution in gray squirrels  
321 (*Sciurus carolinensis*) in *Understanding Urban Ecology*, Hall, M.A. & Balogh, S. (Eds).  
322 (Springer, 2019).
- 323 14. Lehtinen, R. M. *et al.* Dispatches from the neighborhood watch: Using citizen science and  
324 field survey data to document color morph frequency in space and time. *Ecol. Evol.* **10**,  
325 1526-1538.
- 326 15. Perlut, N. G. Long-distance dispersal by eastern gray squirrels in suburban habitats.  
327 *Northeast. Nat.* **27**, 195-200 (2020).
- 328 16. Goheen, J. R., Swihart, R. K., Gehring, T. M. & Miller, M. S. Forces structuring tree squirrel  
329 communities in landscapes fragmented by agriculture: species differences in perceptions of  
330 forest connectivity and carrying capacity. *Oikos* **102**, 95-103 (2003).
- 331 17. Ducharme, M. B., Laroche, J. & Richard, D. Thermogenic capacity in gray and black  
332 morphs of the gray squirrel, *Sciurus carolinensis*. *Physiol. Zool.* **62**, 1273-1292 (1989).
- 333 18. Linnen, C. R. & Hoekstra, H. E. Measuring natural selection on genotypes and phenotypes in

- 334 the wild. *Cold Spring Harb. Symp. Quant. Biol.* **74**, 155-168 (2010).
- 335 19. Campbell-Staton, S. C. *et al.* Parallel selection on thermal physiology facilitates repeated  
336 adaptation of city lizards to urban heat islands. *Nat. Ecol. Evol.* **4**, 652-658 (2020).
- 337 20. Reid, N. M. *et al.* The genomic landscape of rapid repeated evolutionary adaptation to toxic  
338 pollution in wild fish. *Science* **354**, 1305-1308 (2016).
- 339 21. Bowers, M. A. & Breland, B. Foraging of gray squirrels on an urban-rural gradient: use of  
340 the GUD to assess anthropogenic impact. *Ecol. Appl.* **6**, 1135-1142 (1996).
- 341 22. McCleery, R. A., Lopez, R. R., Silvy, N.J. & Gallant, D. L. Fox squirrel survival in urban  
342 and rural environments. *J. Wildl. Manage.* **72**, 133-137 (2008).
- 343 23. Benson, E. The urbanization of the eastern gray squirrel in the United States. *J. Am. Hist.*  
344 **100**, 691-710 (2013).
- 345 24. Leveau, L. United colours of the city: A review about urbanization impact on animal colours.  
346 *Austral Ecol.* **46**, 670-679 (2021).
- 347 25. Ducrest, A.-L., Keller, L. & Roulin, A. Pleiotropy in the melanocortin system, coloration,  
348 and behavioural syndromes. *Trends Ecol. Evol.* **23**, 502-510 (2008).
- 349 26. Stothart, M. R. & Newman, A. E. M. Shades of grey: host phenotype dependent effect of  
350 urbanization on the bacterial microbiome of a wild mammal. *Anim. Microbiome.* **3**, 46  
351 (2021).
- 352 27. Vasemägi, A. The adaptive hypothesis of clinal variation revisited: Single-locus clines as a  
353 result of spatially restricted gene flow. *Genetics* **173**, 2411-2414 (2006).
- 354 28. Merrick, M. J., Evans, K. L. & Bertolino, S. Urban grey squirrel ecology, associated impacts,  
355 and management challenges in *Grey Squirrels: Ecology and Management of an Invasive*  
356 *Species in Europe*, Shuttleworth, C. M., Lurz P. W. W. & Gurnell, J (Eds). (European



- 357 Squirrel Initiative, 2016).
- 358 29. Chipman, R., Slate, D., Rupprecht, C. & Mendoza, M. Downside risk of wildlife  
359 translocation in *Towards the Elimination of Rabies in Eurasia*, Dodet, B., Fooks, A. R.,  
360 Müller, T. & Tordo, N. (Eds). (Dev. Biol. Basel, Karger, 2008).
- 361 30. Allen, D. L. *Michigan Fox Squirrel Management*. (Michigan Department of Conservation,  
362 1943).
- 363 31. Schorger, A.W. Squirrels in early Wisconsin. *Trans. Wis. Acad. Sci. Arts Lett.* **39**, 195-247  
364 (1949).
- 365 32. Robertson, G. I. *Distribution of Color Morphs of Sciurus carolinensis in Eastern North*  
366 *America*. (University of Western Ontario, 1973).
- 367 33. MacCleery, D. W. *American Forests: A History of Resiliency and Recovery*. (Forest History  
368 Society, 2011).
- 369 34. Foster, D. R. *et al. Wildlands and Woodlands: A Vision for the New England Landscape*.  
370 (Harvard University Press, 2010).
- 371 35. Thompson, R. T., Carpenter, D. N., Cogbill, C. V., & Foster, D. R. Four centuries of change  
372 in northeastern United States forests. *PLoS ONE* **8(9)**, e72540 (2013).
- 373 36. Lambert, M. R. *et al.* Adaptive evolution in cities: progress and misconceptions. *Trends Ecol.*  
374 *Evol.* **36**, 239-257 (2021).
- 375 37. Farquhar, D. N. *Some Aspects of Thermoregulation as Related to the Geographic*  
376 *Distribution of the Northern Melanic Phase of the Grey Squirrel*. (York University, 1974).
- 377 38. Innes, S. & D. M. Lavigne. Comparative energetics of coat colour polymorphs in the eastern  
378 gray squirrel *Sciurus carolinensis*. *Can. J. Zool.* **57**, 585-592 (1979).
- 379 39. Santangelo, J. S. *et al.* Predicting the strength of urban-rural clines in a Mendelian

- 380 polymorphism along a latitudinal gradient. *Evol. Lett.* **4**, 212-225 (2020).
- 381 40. Fidino, M. *et al.* Landscape-scale differences among cities alter common species' responses  
382 to urbanization. *Ecol. Appl.* **31**, e02253 (2021).
- 383 41. Dickinson, J. L., Zuckerberg, B. & Bonter, D. N. Citizen science as an ecological research  
384 tool: challenges and benefits. *Annu. Rev. Ecol. Evol. Syst.* **41**, 149-172 (2010).
- 385 42. Alberti, M. Global urban signatures of phenotypic change in animal and plant populations.  
386 *Proc. Natl. Acad. Sci. U.S.A.* **114**, 8951-8956 (2017).
- 387 43. United States Census Bureau. 2019 TIGER/Line Shapefiles (machine-readable data files)  
388 <https://www2.census.gov/geo/tiger/TIGER2019/UAC/> (2019).
- 389 44. XX. Statistics Canada. Population Centre Boundary File, Census year 2016  
390 <https://www150.statcan.gc.ca/n1/en/catalogue/92-166-X> (2017).
- 391 45. Aiello-Lammens, M. E. *et al.* spThin: an R package for spatial thinning of species occurrence  
392 records for use in ecological niche models. *Ecography* **38**, 541-545 (2015).
- 393 46. R Core Team. R: A language and environment for statistical computing. R Foundation for  
394 Statistical Computing, Vienna, Austria. (2020).
- 395 47. Brown de Colstoun, E. C. *et al.* *Documentation for the Global Man-made Impervious*  
396 *Surface (GMIS) Dataset from Landsat*. (NASA Socioeconomic Data and Applications  
397 Center, 2017).
- 398 48. Steele, M. A. & Koprowski, J. L. *North American Tree Squirrels*. (Smithsonian Books,  
399 2001).
- 400 49. Hansen, M. C. *et al.* High-resolution global maps of 21<sup>st</sup>-century forest cover change. *Science*  
401 **342**, 850-853 (2013).
- 402 50. Fick, S. E. & Hijmans R. J. WorldClim 2: new 1km spatial resolution climate surfaces for

- 403 global land areas. *Int. J. Climatol.* **37**, 4302-4315 (2017).
- 404 51. Hijmans, R. L. raster: Geographic data analysis and modeling. R package version 3.3-13.  
405 (2020). <https://CRAN.R-project.org/package=raster>.
- 406 52. Baston, D. exactextractr: Fast extraction from raster datasets using polygons. R package  
407 version 0.5.1. (2020). <https://CRAN.R-project.org/package=exactextractr>.
- 408 53. Harrison, X. A. *et al.* A brief introduction to mixed effects modelling and multi-model  
409 inference in ecology. *PeerJ* **6**, e4794 (2018).
- 410 54. Bates, D. , Maechler, M., Bolker, B. & Walker, S. Fitting linear mixed-effects models using  
411 lme4. *J. Stat. Softw.* **67**, 1-48 (2015).
- 412 55. Gelman, A. & Su, Y. arm: Data analysis using regression and multilevel/hierarchical models.  
413 R package version 1.11-2. (2020). <https://CRAN.R-project.org/package=arm>.
- 414 56. Gelman, A. & Hill, J. *Data Analysis Using Regression and Multilevel/Hierarchical Models*.  
415 (Cambridge University Press, 2007).
- 416 57. Crase, B., Liedloff, A. C., & Wintle, B. A. A new method for dealing with residual spatial  
417 autocorrelation in species distribution models. *Ecography* **35**, 879-888 (2012).
- 418 58. Bivand, R. S. & Wong, D. W. S. Comparing implementations of global and local indicators  
419 of spatial association. *TEST* **27**, 716-748 (2018).
- 420 59. Bardos, D. C., Guillera-Aroita, G. & Wintle, B. A. Valid auto-models for spatially  
421 autocorrelated occupancy and abundance data. *Methods Ecol. Evol.* **6**, 1137-1149 (2015).

422

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424

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427

428 **Author Contributions**

429

430 B.J.C. and J.P.G designed the study. B.J.C. led the data analysis and wrote the first draft of the  
431 manuscript. J.P.G. contributed ideas for data analysis and edited the manuscript.

432

433 **Data Availability**

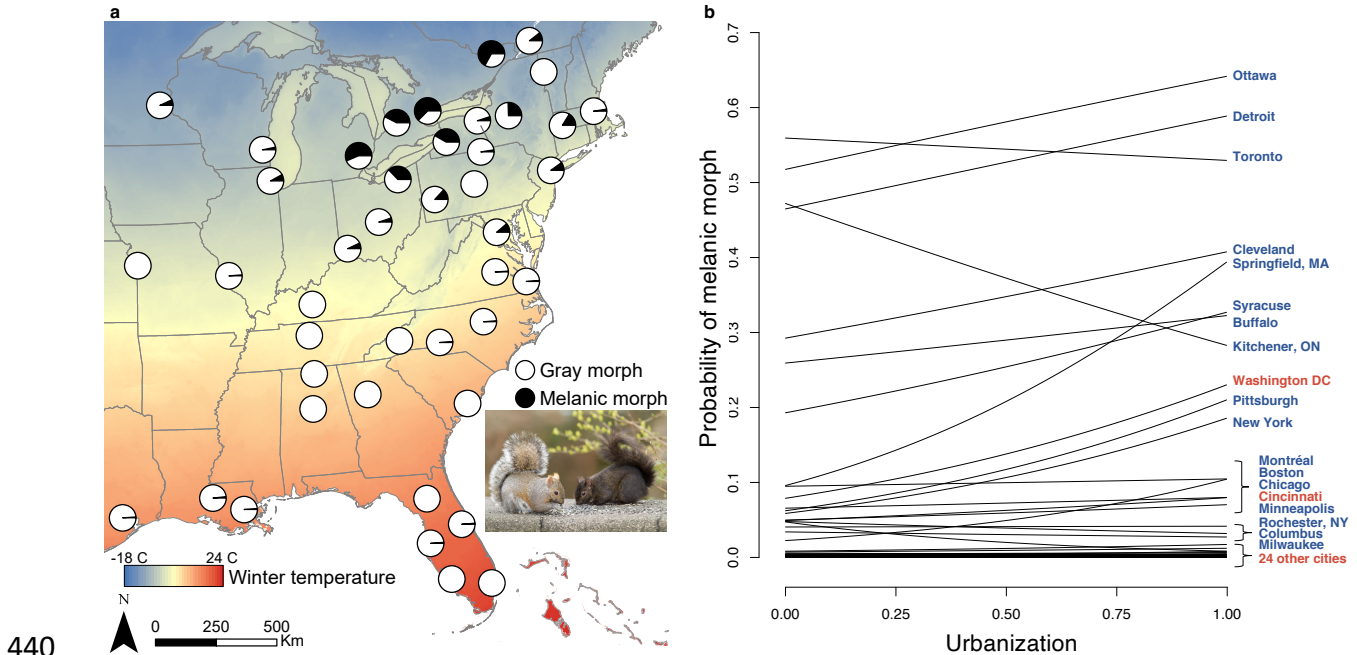
434 The observations of eastern gray squirrels (*Sciurus carolinensis*) generated and analyzed for this  
435 project can be found at <https://github.com/bcosentino/squirrelMapper-urban-rural-clines>.

436

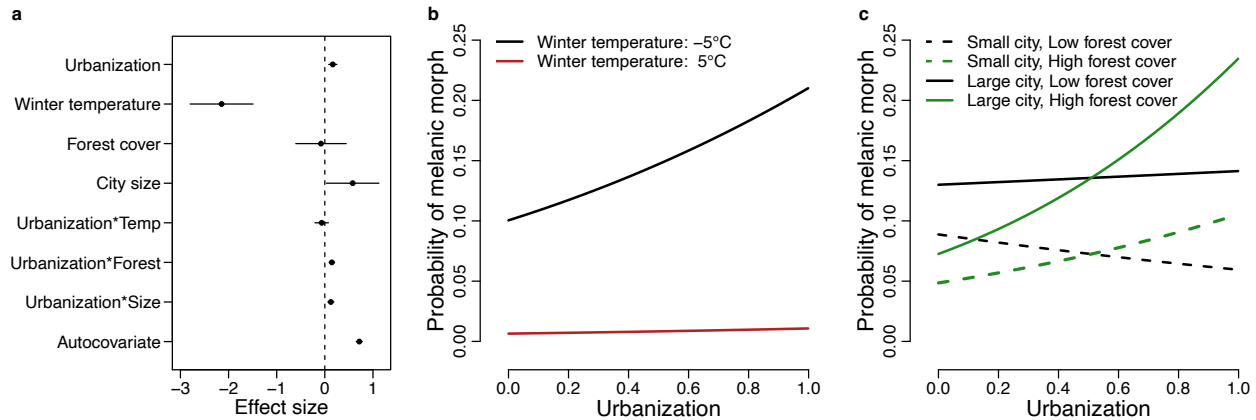
437 **Competing interests**

438

439 The authors declare no competing interests.



**Fig. 1. a,** Distribution of coat color morphs of 26,924 eastern gray squirrels (*Sciurus carolinensis*) among 43 cities in North America. Pie charts show the proportion of melanic and gray color morphs in each city. Gray morph included rare color morphs that accounted for 2% of observations (Supplementary Table 1). **b,** Clines in melanism along urbanization gradients in each city. Urbanization was measured as standardized impervious cover. Regression lines represent predicted effect of urbanization on melanism in each city based on a linear mixed model including effects of city size, forest cover, and winter temperature (Fig. 2, Supplementary Table 2). City names are color-coded to show winter temperatures above (red) and below (blue) the median winter temperature. Photo ‘grey squirrel, black squirrel’ (tinyurl.com/pauha835) by Eyesplash Photography is licensed under CC BY-NC-ND 2.0 (<https://creativecommons.org/licenses/by-nc-nd/2.0/>).



453

454 **Fig. 2. a**, Effect sizes for fixed effects of urbanization, winter temperature, forest cover, city size,

455 interaction terms with urbanization, and a spatial autocovariate from a linear mixed model of

456 melanism in 26,924 eastern gray squirrels (*Sciurus carolinensis*) among 43 cities in North

457 America. Effect sizes represent standardized regression coefficients and include 95% confidence

458 intervals. **b**, Relationship of the probability of melanism to urbanization at low (-5°C) and high

459 (5°C) winter temperature while holding city size and forest cover constant. **c**, Relationship of the

460 probability of melanism to urbanization at varying levels of city size and forest cover. Effects of

461 urbanization on melanism are shown for combinations of small cities (25,000 ha), large cities

462 (350,000 ha), low forest cover (15%), and high forest cover (45%) while holding winter

463 temperature constant at -5°C. Urbanization was measured as standardized impervious cover in all

464 panels.