Simultaneous passive acoustic monitoring uncovers evidence of potentially overlooked temporal variation in an Amazonian bird community

W. Alexander Hopping[1], Christopher J. Sayers II[2], Noe Roger Huaraca-Charca[3], Holger Klinck[1]

[1] K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Cornell University, Ithaca, NY, USA
[2] Department of Ecology and Evolutionary Biology, University of California, Los Angeles, CA 90095, USA
[3] Inkaterra Asociación, Victor Larco Herrera 130, Miraflores, Lima, Peru

Corresponding Author:
W. Alexander Hopping[1]

K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Cornell University, 159 Sapsucker Woods Road, Ithaca, NY 14850, USA

Email address: wah63@cornell.edu

Abstract

The vocal activity and detectability of tropical birds are subject to high levels of temporal

heterogeneity, but quantifying fine-grain patterns of diel and day-to-day variation in complex

systems is challenging with traditional point-count methods. As a result, research concerning

stochastic temporal effects on tropical avian assemblages is limited, typically offering only broad

conclusions, e.g., that overall activity is highest in the first few hours of the morning and that

some species are active at different times of day. Passive acoustic monitoring introduces several

advantages for studying temporal variation, particularly by enabling simultaneous and

continuous data collection across adjacent sites. Here, we employed autonomous recording units

to quantify temporal variation in avian vocal activity levels and observed species richness at an

Amazonian reserve in Madre de Dios, Peru—a region featuring some of Earth's richest, most

complex avian assemblages. We manually annotated 18 dawn hour recordings, collected

simultaneously from three separate days at the same six sites, which represent a variety of

microhabitats and avian community compositions. We documented significant and consistent

temporal variation in avian vocal activity levels and observed species richness within the dawn

hour and across days. We found that temporal effects were stronger for vocal activity than for

observed species richness and that diel activity patterns varied on a species-to-species basis at a

fine scale. Our results indicate that overlooked temporal variation in Amazonian soundscapes

may obfuscate the results of surveys that fail to sufficiently account for temporal variables with

simultaneous monitoring. Further research concerning the automated processing of acoustic data

could lead to more efficient methods for reducing temporal bias and improving the calibration

and accuracy of ornithological surveys in the tropics.

Keywords: avian biodiversity, bioacoustics, passive acoustic monitoring, Amazon rainforest, temporal variation.

1 Introduction

The tropics account for an overwhelming share of Earth's avian diversity, featuring more than

75% of all species and over 90% of terrestrial birds (Barlow *et al.* 2018). Still, tropical regions

are underrepresented in ornithological literature relative to temperate regions, even though

tropical birds have fundamentally different life history strategies, behavioral ecology, and vocal

patterns than temperate birds, necessitating separate studies (Stutchbury & Morton 2008). The

largest intact tropical system on the planet is the Amazon basin (Allan *et al.* 2017), which is

often considered the epicenter of global biodiversity (Jenkins *et al.* 2013). Like most of the

tropics, the Amazon is undergoing major ecological change (Edelman 2014), primarily from the

region's rapidly expanding anthropogenic land use footprint (Laurence *et al.* 2014), as well as the increasing impacts of climate change (Lovejoy 2013). Despite its importance, however, the Amazon's avifauna remains poorly documented and is subject to flawed baseline species occurrence data and significant knowledge gaps (Lees *et al.* 2014), which are exacerbated by systemic barriers to researchers from the Global South (Soares *et al.* 2022). Efforts to increase ornithological survey coverage in the region are complicated by the difficulty of tropical field surveys: overwhelming avian species richness, including many rare and similar species, challenging logistics, and poor visibility conditions contribute to the unreliability of avian surveys in Amazonia, where upwards of 95% of birds are heard but never seen by a field observer (Robinson *et al.* 2018).

The vocal activity and detectability of Neotropical birds are subject to high levels of spatiotemporal heterogeneity, including seasonality (Pérez-Granados & Schuchmann 2022), biotic interactions like cicada choruses (Hart *et al.* 2015), and a wide range of poorly-understood floristic, geographic, and spatial variables (Menger *et al.* 2017). Some broad patterns of diel variation in the detectability and vocal activity of tropical birds are well established; for example, that birds are most vocally active during the first 2–3 hours after dawn (Lynch 1995, Woltmann 2005). However, the temporal variation of tropical soundscapes is more complex than hour-long intervals can capture (Verner & Ritter 1986, Rodriguez *et al.* 2014, Metcalf *et al.* 2021). Many Amazonian species are only aurally detectable within strict temporal niches (Gil & Llusia 2020). These species may only vocalize during a single 5 min period in the morning, perhaps accompanied by an even shorter bout in the evening (Parker 1991). Some Amazonian species have been shown to sing as infrequently as twice in 50 days (Jirinec *et al.* 2018). In addition to

affecting estimates of total species richness, this fine-scale temporal variation in vocal activity

can also influence patterns of observed community composition. Temporal effects on

detectability vary by functional or taxonomic group; for example, canopy species are known to

reach their activity peak later in the morning than understory species (Blake 1992), and species

with high sensitivity to habitat fragmentation may have proportionately higher detection rates in

pre-dawn surveys than less sensitive species (Woltmann 2005).

Quantifying the fine-scale patterns of temporal variation in tropical bird communities,

both within and across days, is difficult. Non-simultaneous survey methods are subject to a litany

of spatial, temporal, and observer biases, and therefore capturing such granular effects can

require infeasibly large sample sizes (Lynch 1995). For example, Esquivel & Peris (2008) found

that four visits per point are necessary to account for the temporal variation in bird activity in the

Atlantic Forest of Paraguay. This sampling design does not account for possible variation in

vocal activity levels between different days, between different observers (Robinson *et al.* 2018),

nor travel time between sites. Additionally, Rodriguez *et al.* (2014) found that temporal effects

on acoustic activity patterns were spatially heterogeneous at adjacent recording sites in French

Guiana. Precise assessments of temporal variation, therefore, may need to be conducted on a per-

site basis.

More costly and labor-intensive approaches, for example, using a team of point count

technicians to visit multiple points at the same time, may also introduce confounding variables.

Amazonian soundscapes include thousands of vocalization types, inducing high rates of false-

negative and false-positive identification errors (Remsen 1994), and it can take months for

skilled observers to reach a level of competence attainable in just one to two weeks of field

experience in temperate systems (Parker 1991). Observer error rates have been experimentally

quantified in relation to aural identification (Simons *et al.* 2007), visual identification (Hull *et al.*

2010), and the subjectivity of abundance estimates (Cerqueira *et al.* 2013). These dynamics are

not limited to inexperienced surveyors; established ornithologists can also be subject to non-

trivial identification error rates in environments as complex as Amazonia (Lees *et al.* 2014).

Overlooked species tend to reflect non-random subsets of assemblages, and the specific nature of

these biases may differ between individual observers, even amongst experts (Blackburn &

Gaston 1998). Therefore, standardized surveys that do not produce archivable raw data and seek

to uncover subtle spatiotemporal variation in avian vocal activity, observed species richness, and

community structure should ideally be conducted by a single observer (Blake & Loiselle 2015,

Robinson *et al.* 2018).

Simultaneous passive acoustic monitoring (PAM), obtained with the use of autonomous

recording units (ARUs), has the capacity to eliminate much of the temporal bias in surveys by

collecting continuous data from adjacent locations concurrently when temporal variables are

uniform (Venier *et al.* 2012, Tegeler *et al.* 2012). PAM can address observer bias since data

collection is independent of technician skill level, and recordings collected simultaneously can

be annotated by a single observer, replayed & reanalyzed, archived publicly, and distributed to

experts for an independent review of identifications (Rempel *et al.* 2005, Robinson *et al.* 2018).

PAM has been shown to present a suitable supplement or alternative to point counts, both in the

specific region where this study was conducted (Haselmayer & Quinn 2000) and in general

(Shonfield & Bayne 2017, Darras *et al.* 2019, Blake 2021) while providing numerous other

advantages (Acevedo & Villanueva-Rivera 2006, Newson *et al.* 2017, Darras *et al.* 2018, Jorge *et al.* 2018, Pillay *et al.* 2019, Sugai & Llusia 2019).

This study aimed to investigate the hypothesis that Amazonian soundscapes may be subject to overlooked temporal variation in avian vocal activity, both between days and throughout a single dawn hour, that, in the absence of simultaneous survey methods, could mask important differences in vocal activity levels, community composition, and observed species richness between sites. Such effects are difficult to quantify with traditional surveys, and establishing their presence and magnitude may be useful for improving the calibration of point count studies, even with relatively small volumes of PAM data. We predicted that stochastic effects of diel and day-to-day variation on vocal activity levels and observed species richness would be consistent across sites and, therefore, possible to control for with simultaneous monitoring. We addressed this question by deploying an array of ARUs at a single Amazonian reserve in Madre de Dios, Peru, and manually annotated a sample of the resulting bioacoustic data.

2 Materials & Methods

2.1 Study area

We conducted this study at Inkaterra Reserva Amazónica (ITRA) in the Madre de Dios Department of southeastern Peru (12°32'07.8" S, 69°02'58.2" W). This 191-ha private reserve is located at an elevation of approximately 200 m along the Madre de Dios River, directly across from the Reserva Nacional de Tambopata—one of the most biodiverse regions on the planet (Foster *et al.* 1994) (Fig. 1). The habitat at ITRA primarily consists of *várzea* floodplain forest,

the most endangered forest type in the southwestern Amazon (Phillips *et al.* 1994), interspersed with seasonally flooded *Mauritia* palm swamp forest.

Figure 1: Swift recorder locations and typical forest habitat at the study site. Recorder locations covered a range of habitats, including edge (Site A), degraded forest (Site F), mature *várzea* in the reserve interior (Site C and Site E), and mixed palm swamp (*aguajal*)-*várzea* habitat (Site B and Site D).

2.2 Acoustic monitoring

We collected acoustic data from 14 January to 2 February, 2019, using six ARUs (Swift recorder, K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Ithaca, NY, USA). We deployed the ARUs across the entire reserve, at a minimum distance of 350 m from each other and 450 m from the river, to limit spatially overlapping detections (pseudoreplication) and background noise (Ralph *et al.* 1995), respectively. Deployment locations were chosen to represent a gradient of intactness and forest habitat type, including edge (Site A), degraded (Site F), mature *várzea* in the reserve's interior (Site C and Site E), and mixed palm swamp (*aguajal*)-*várzea* habitat (Site B and Site D) (Fig. 1). Placed at the height of 1.5 m above the ground, the 6 ARUs recorded continuously (mono, WAV format) throughout the deployment using a sampling rate of 48 kHz (16-bit resolution) and a gain setting of 35 dB. The microphone sensitivity of the Swift was –44 dBV/Pa (+/– 3 dB) and featured a flat frequency response (+/– 3 dB) in the frequency range of the vocalizations of interest. The clipping level of the analog-to-digital (ADC) converter was +/– 0.9 V.

2.3 Annotation process

We manually annotated eighteen total dawn hours, from 05:00–06:00 PET (10:00–11:00 UTC), representing six sites on three days, using the Raven Pro Sound Analysis Software (version 1.6; K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Ithaca, NY, USA). Annotations were performed by a single observer, the lead author, to control for observer bias. A local expert, Noe Huaraca-Charca, also annotated a subset of these recordings (n = 4) as a quality control measure. The three days, 16 January (Day 1), 20 January (Day 2), and 31 January (Day 3), 2019, were randomly chosen from a subset of days that featured less than 5 min of

heavy rain during the dawn hour. The dawn hour was selected because it typically contains the highest level of vocal activity and diversity among Neotropical birds (Berg *et al.* 2006), including many species that only sing in narrow time periods around first light (Parker 1991, Blake 1992). Additionally, the dawn soundscape in Amazonia features far less anthropogenic activity and insect noise than in other periods, reducing interference with avian species detection (Metcalf *et al.* 2020). While annotating short, randomly-selected samples spread out over a longer survey period can improve estimates of alpha and gamma diversity (Wimmer *et al.* 2013, Metcalf *et al.* 2021), this study was designed to assess temporal variation over the course of a contiguous hour, and annotations were therefore conducted in this format.

Spectrograms were viewed in 60 s increments split over two rows (Hann window, 3 dB bandwidth = 100 Hz, 690 FFT window size, 1024-sample DFT, 95% temporal overlap) (Fig. 2). To reduce bias introduced by improved sound identification knowledge developed throughout the project, the order in which recordings were annotated was determined using a random number generator. Vocalizations were labeled for both species and a handful of broader groups with confusing or indistinguishable calls, e.g., "TRSP" as "trogon species" (*Trogonidae sp.*). Background vocalizations, audible but not clearly visible on a spectrogram, were denoted with a '1' and excluded from analysis. To account for inter-specific variation in call volume and frequency, consistency for applying 'background' designation was approached on a species-specific basis. In cases where vocalizations from the same species were separated by fewer than 5 s, they were included as part of the same annotation, while annotations were split if the gap between vocalizations exceeded 5 s. An assortment of unlabeled sound clips, featuring both identified and unidentified species, were sent to regional experts for secondary verification.

Annotations were reviewed repeatedly until unidentified vocalizations made up less than 10% of the total avian vocal activity on a given recording. Considerable effort was made to accurately label as many foreground and background vocalizations as possible, with the goal of improving the value of this dataset for training soundscape-based automated identification algorithms.

Figure 2: The annotation process. Spectrograms showing one minute of sound from Site B, Day 2, before and after manual annotation in Raven Pro 1.6.

2.4 Statistical analyses

To assess fine-scale temporal variation in dawn chorus vocal activity within recordings and across sampling sites and days, we subsampled each 60 min recording at 1 min intervals,

henceforth totaling 1,080 samples across all sites and days. For each 1 min interval, we calculated species richness (SR), vocal prevalence (VP), and total vocal prevalence (TVP). Species richness refers to the number of unique species detected, while vocal prevalence refers to the number of 10 s intervals featuring a vocalization of a given species. Therefore, a single species could attain a maximum VP of 6 per 1 min interval or 360 per hour recording. Total vocal prevalence (TVP)—a proxy for the total avian vocal activity in a soundscape—refers to the sum of VP for all detected taxa in a given subsample. We assessed the relationship between TVP and SR using a Pearson's *R* correlation analysis.

Certain taxa were excluded from species richness calculations for data quality reasons: parrots (*Psittacidae sp.*), because they occurred almost exclusively as calling flyovers, as well as *Columbiformes* (pigeons and doves) and three congeneric tinamou species, Cinereous Tinamou (*Crypturellus cinereus*), Little Tinamou (*C. soui*), and Bartlett's Tinamou (*C. bartletti*), which had overlapping vocalizations and were inconsistently identified with confidence. However, these species were included in calculations of TVP, as were vocalizations that were identified to a broader group but not at the species level (e.g., Trogon sp. [*Trogonidae sp.*]) (For a full list of species and exclusions for SR and TVP calculations, see Supporting Information Appendix S1).

We then constructed a set of Generalized Additive Models (GAMs) fit with Poisson distributions and restricted maximum likelihood estimations to account for the non-linearity of within-recording response curves for SR, TVP, and species-specific VP within the dawn hour. For both the SR and TVP models, we included a smoothing term for the interaction between time of day and recording to produce individual response curves for each recording, as well as day

and site as random effects intercepts to account for pseudoreplication. Similarly, to produce average species-specific VP response curves across days and sites, we replicated this random effects structure, but removed the interaction for the time interval smoothing term. As the amount of species present at time step t directly influences the amount of species that are likely to be detected at time step $t + 1$, we accounted for this inherent temporal autocorrelation by fitting both models with a first-order autoregressive covariance structure (AR1; Yang et al. 2012). This decision successfully removed evidence of autocorrelation within model residuals. Because our model structure was biologically-informed, we chose to forgo model selection and base inferences and predictions from this structure. We completed all statistical analyses in R (v. 4.2.0; R Core Team 2022) using functions from the "tidyverse" (Wickham et al. 2019) and employing the "mgcv" package to construct GAMs (Wood et al. 2016; Wood 2004, 2011, 2017).

We assessed patterns in community composition across sites using non-metric multidimensional scaling ordination (NMDS). Substituting VP in place of abundance, we calculated NMDS using the "bray" method and metaMDS function from the "vegan" package in R (Okansen *et al.* 2019). We plotted NMDS ordinations on two axes and ensured that the output met minimum stress requirements (Kruskal 1964). To compare TVP and SR between sites, days, and recordings, two-way repeated-measures ANOVA followed by Tukey's HSD tests for multiple comparisons were conducted in R, with site and day as interacting independent variables and the 1min time interval as a blocking variable.

3 Results

We identified 127 species over the 18 analyzed dawn hours, seven of which were recorded only as background vocalizations; 17,036 individual bounding-box annotations were drawn in Raven

Pro 1.6, of which 94.8% were identified to the species or taxon level. Site A had the lowest average species richness per 1 min interval (4.0; 43 total), followed by Site F (4.4; 42 total), Site D (4.8; 49 total), Site C (4.9; 46 total), Site E (5.3; 55 total), and Site B (6.1; 50 total), which was also the most active site by TVP. Aggregating detections across sampling sites, Day 1 featured 57 species, Day 2 featured 78 species, and Day 3 featured 56 species, with average SR values per 1 min interval of 4.79, 6.18, and 3.81, respectively. At individual sites, the percent of total vocal prevalence (TVP) that was not identified to a species or taxa group ranged from 4.1% (Site E) to 7.1% (Site B) between sites, and from 5.1% (Day 3) to 5.4% (Day 1) between days. Most annotations occurred in the frequency range 0.5–5 kHz, which features the highest diversity of Amazonian bird vocalizations at dawn, and the least interference from insects, which contribute to significant signal masking between 4–12 kHz (Hart *et al.* 2015, Alvarez-Berríos *et al.* 2016, Aide *et al.* 2017, Metcalf *et al.* 2020). Of the 120 foreground species detected, only two, Black-faced Antthrush (*Formicarius analis*) and Buff-throated Woodcreeper (*Xiphorhynchus guttatus*), were detected on all 18 recordings, and six more, Hauxwell's Thrush (*Turdus hauxwelli*), Little Tinamou (*Crypturellus soui*), Thrush-like Wren (*Camplyothynchs turdinus*), Amazonian Motmot (*Momotus momota*), Plumbeous Pigeon (*Patagioenas plumbea*), and Screaming Piha (*Lipaugus cociferans*) were detected on at least 15 recordings. Most species were rare; more than half (*n* = 66, 55%) were detected on 4 or fewer 1 hour recordings, and roughly a quarter (*n* = 27, 23%) on only a single recording (Supporting Information Appendix S2). Regarding total vocal prevalence, the most abundant species were Hauxwell's Thrush, Black-faced Antthrush, and Thrush-like Wren. At all six sites, TVP was highest on 20 January (Day 2), and was lowest on 31 January (Day 3) at every site but Site B.

SR and TVP models accounted for a substantial amount of variation in dawn chorus vocal activity ($R^2 = 0.62$ and $R^2 = 0.78$, respectively), in which smoothing terms for the interaction between time interval and recording, as well as day and site intercepts, were often highly significant ($p < 0.01$; Supporting Information Appendix S3). TVP was strongly correlated with SR ($R^2 = 0.85$, $p < 0.001$). Individual sites consistently clustered separately from each other in terms of their species composition, as demonstrated by non-metric multidimensional scaling (NMDS) ordination (Fig. 4). Tukey's HSD tests indicated that 8 of the 15 pairwise comparisons (53.3%) between sites featured significant differences in TVP ($p < 0.05$), 6 (40%) of which were highly significant ($p < 0.01$), while 9/15 (60%) were highly significant for SR ($p < 0.01$). All comparisons between days were highly significant ($p < 0.01$) for both TVP and SR. Between individual recordings from different sites on the same day, 10/45 (22.2%) were highly significant ($p < 0.01$) for both TVP and SR. Concerning recordings from the same sites but different days, 8/18 (44.4%) featured highly significant ($p < 0.01$) differences in SR, compared to 10/18 (55.6%) for TVP. Between recordings where both day and site were different, 38/90 (42.2%) had highly significant differences in SR, compared to 51/90 (56.7%) for TVP. Tukey's HSD pairwise comparisons between sites, days, and recordings are featured in Supporting Information Appendix S4.

Figure 3: Comparisons of (A) species richness and total vocal prevalence detected by time of day across days and sites overlaid with model-generated prediction curves ± 95% confidence intervals, along with (B) site-specific mean prediction curves overlaid with day-specific mean prediction curves ± 95% confidence intervals of species richness and vocal prevalence across days.

Figure 4: Non-metric multidimensional scaling (NMDS) ordination using Bray-Curtis dissimilarity, with vocal prevalence in place of abundance, displayed by site. Significantly representative species ($p < 0.02$) are shown with vectors.

4 Discussion

We used simultaneous passive acoustic monitoring to quantify temporal variation in the vocal activity level and observed species richness of an Amazonian bird community. We documented significant temporal variation both across days and within the dawn hour at the same sites and time of year. The effect of time interval and day were generally stronger than the effect of site in explaining variation in observed species richness (SR) and total vocal prevalence (TVP). In line with our major hypothesis, the magnitude of this temporal variation is large enough to potentially mask, or even supersede, observed differences in species richness and vocal activity levels between sites and habitat types—complicating the interpretation of point-count studies and other non-simultaneous survey methods that fail to control for temporal variables. Additionally, we

observed that diel variation in avian vocal activity varies by species and taxonomic group (Fig. 5); therefore, even when summary numbers such as SR and TVP are stable at different points in the morning, these numbers can reflect considerably different community composition. While temporal variation in avian activity over the course of the morning is well-documented in the tropics, the limitations of traditional non-simultaneous survey methods mean that point-count studies on temporal effects in these systems offer only broad conclusions, typically relating to intervals of one hour or more (Blake 1992, Lynch 1995, Woltmann 2005, Esquivel & Peris 2008). Fine-scale diel variation in the vocal activity and detectability of Amazonian birds has been referenced anecdotally for decades (Parker 1991). However, thanks to the development of autonomous recording units, which allow for considerably easier repeat sampling than traditional methods, we are beginning to understand more about these effects at higher temporal resolutions (Gil & Llusia 2020, Blake 2021, Metcalf *et al.* 2021, Pérez-Granados & Schuchmann 2023). While prior temporal effects studies have focused primarily on diel and seasonal temporal variation, to our knowledge this is the first study to provide evidence that day-to-day temporal variation in the vocal activity and detectability of Amazonian bird communities may be strong enough to obscure statistically significant differences between sites, in the absence of simultaneous data collection.

The differing patterns of diel variation in TVP and SR that we documented between sites are plausibly explained partially or entirely by the spatial heterogeneity of avian community composition. Though sometimes misrepresented as homogenous, Amazonian habitats are highly heterogeneous (Pires & Prance 1985, Tuomisto *et al.* 1995), and important elements of their structural diversity remain poorly understood (Macía 2011, Borges 2013). Interactions between

habitat features and avian community structure in the Amazon are extremely complex, and are influenced by variables, including soil type (Borges 2013), riverine sediment concentration (Laranjeiras *et al.* 2021), the condition of surrounding habitat (Latta *et al.* 2011, Laurence *et al.* 2011, Woltmann 2012, Laurence *et al.* 2012, Wolfe *et al.* 2015, Barlow 2016, Menger *et al.* 2017, Hernández-Palma & Stouffer 2018), the presence of large animals (Redford 1992, Estes *et al.* 2011), and climate change (Blake & Loiselle 2015, Stouffer *et al.* 2020). As a result, Amazonian habitat assessments, which are generally reliant on features discernable by field observers, can be overly simplistic (Milliken *et al.* 2010), and we avoided incorporating them in our analysis for this reason. However, all sampled sites were unique, featuring differing community compositions (Fig. 4), even in contiguous forest of comparable superficial condition. Likewise, we found that patterns of diel variation within the dawn hour were spatially heterogeneous and varied on a per-site basis, consistent with the findings of Rodriguez *et al.* (2014). Given that diel variation in vocal activity and detectability varies on a species-to-species basis (Fig. 5), sites with differing communities should be expected to exhibit differing patterns of diel variation in overall vocal activity. This diel variation means that visiting two sites with the same overall species richness at different times of the morning could result in considerably different observations of community composition and species richness, even when both visits occur within the same hour. Determining the mechanisms driving temporal variation in avian vocal activity is outside of the scope of this research, but possible contributors include weather and feeding opportunities (Metcalf *et al.* 2021), along with temporal partitioning to avoid signal masking from vocalizing insects (Hart *et al.* 2015, Aide *et al.* 2017).

A)

Species	0-5min	5-10min	10-15min	15-20min	20-25min	25-30min	30-35min	35-40min	40-45min	45-50min	50-55min	55-60min
Cinereous Tinamou	5	15	69	59	29	46	61	17	22	3	8	1
Little Tinamou	7	34	43	39	25	18	10	17	8	3	0	1
Undulated Tinamou	9	33	39	38	51	44	50	47	58	47	53	43
Plumbeous Pigeon	0	1	3	1	31	15	26	26	18	42	40	67
Ruddy Pigeon	0	1	1	0	8	2	2	22	26	14	0	0
Grey-fronted Dove	0	1	1	13	15	40	79	88	39	13	12	20
Amazonian Pygmy-Owl	51	48	21	4	1	2	0	0	0	0	0	0
Black-tailed Trogon	0	0	0	1	3	2	17	7	11	23	13	18
Collared Trogon	0	0	0	21	20	3	15	20	48	53	47	21
Amazonian Motmot	38	31	52	90	69	54	42	14	15	10	15	0
Mealy Parrot	0	0	0	0	7	9	21	29	23	10	5	11
Fasciated Antshrike	4	5	1	2	8	31	21	49	35	31	3	6
Plain-winged Antshrike	0	0	19	46	39	39	20	15	3	1	0	8
Dusky-throated Antshrike	0	1	0	6	14	10	17	24	30	33	19	13
Spot-winged Antshrike	0	0	20	26	30	10	13	23	13	8	5	2
Pygmy Antwren	0	0	0	0	8	11	10	12	19	17	11	10
Long-winged Antwren	4	19	17	39	27	3	6	3	3	12	0	8
Grey Antbird	0	0	10	37	40	43	60	100	93	69	80	57
Plumbeous Antbird	0	0	0	0	2	2	36	76	75	59	60	43
Goeldi's Antbird	0	2	13	18	20	27	19	23	22	18	17	7
Black-throated Antbird	0	0	1	0	4	4	2	14	28	18	8	0
Black-faced Antthrush	7	29	55	114	118	148	164	194	163	184	190	148
Rufous-fronted Antthrush	0	0	0	0	1	4	17	24	19	15	19	21
Cinnamon-throated Woodcreeper	7	25	39	24	21	4	20	8	13	6	0	1
Long-billed Woodcreeper	0	6	10	25	14	9	12	6	3	6	12	9
Amazonian Barred-Woodcreeper	6	46	62	50	17	14	1	0	1	0	0	2
Striped Woodcreeper	0	6	17	41	36	47	27	21	15	11	0	1
Buff-throated Woodcreeper	25	58	86	104	67	60	52	38	34	16	10	8
Purple-throated Fruitcrow	0	0	0	0	9	8	22	17	18	9	16	9
Screaming Piha	2	3	2	10	24	31	27	30	32	54	57	54
White-winged Becard	0	6	31	54	65	54	33	48	52	44	31	26
Wing-barred Piprites	0	0	1	1	4	2	9	2	7	19	35	29
Yellow-margined Flycatcher	0	1	0	1	0	1	1	3	24	24	20	19
Dusky-capped Flycatcher	0	0	0	0	0	3	13	33	15	17	4	0
Piratic Flycatcher	2	2	7	14	24	13	8	12	15	13	12	8
Dusky-capped Greenlet	0	0	0	0	11	32	28	38	74	99	99	75
Thrush-like Wren	1	5	20	44	73	100	90	122	101	102	114	119
Buff-breasted Wren	0	1	5	13	14	30	65	76	36	38	40	54
Hauxwell's Thrush	18	40	46	46	67	63	79	68	74	82	88	88
Russet-backed Oropendola	0	0	0	1	0	0	8	22	25	48	41	14
Amazonian Grosbeak	8	24	55	58	57	46	52	52	48	30	25	0

Figure 5: Species-level diel variation in vocal prevalence within the dawn hour. Diel variation in VP by 5 min interval for species with TVP > 75 from all sites and days (A), and model-generated prediction curves ± 95% confidence intervals of species-specific VP within the dawn hour for six species (B).

While this study was not designed to serve as a comprehensive census of the avian community at ITRA, our observations were generally consistent with those documented by previous census efforts in Madre de Dios's *várzea* forests (Terborgh *et al.* 1990), namely that most species were rare, with a high proportion of TVP contributed by only a few species, and that sites were compositionally heterogeneous. That the site in edge habitat (Site A) had the lowest average species richness per 1 min interval (4.03; 43 total), followed by Site F (4.41; 42 total), in degraded forest outside of the reserve boundaries, are consistent with the results of research focus on edge effects in Amazonia, which are known to pervasively impact Amazonian bird communities (Terborgh *et al.* 1990, Laurence 2004, Barlow *et al.* 2006, Broadbent *et al.* 2008, Laurence *et al.* 2011, Whitworth *et al.* 2015, Wolfe *et al.* 2015, Haddad *et al.* 2015, Moura *et al.* 2016, Stouffer 2020). This habitat usage pattern differs from observed dynamics in temperate systems, where avian diversity tends to be higher in edge habitats than forest interiors (Baldi 1996, Lindell *et al.* 2007).

The temporal variation in avian vocal activity levels and observed species richness that we documented between days was greater and more consistent than we expected. Based on our GAM results, these temporal effects were strongest and most uniform for TVP ($R^2 = 0.78$), which was lowest on Day 3 at all six sites and highest on Day 2 at five of six sites. Likewise, pairwise comparisons between recordings from different days more often featured highly significant ($p < 0.01$) differences in TVP (61/108 comparisons, 56.5%) than SR (46/108, 42.6%). Supporting the idea that temporal effects may overwhelm statistically significant differences between sites, the rate that recordings from different days exhibited significant differences in TVP or SR remained essentially identical, whether or not recordings were from the same site or

from different sites (55.6% vs 56.7% for TVP, 44.4% vs 42.2% for SR). However, only 22.2% of comparisons between recordings from different sites on the same day featured highly significant ($p < 0.01$) differences in either TVP or SR. Therefore, sampling the same site on different days was more than twice as likely to result in differing measures of TVP and SR than sampling different sites on the same day, even though the majority of site-to-site comparisons indicated significant differences in TVP or SR. That temporal variation was strongest for TVP is relevant for point-counts because aural detection probability for a field observer is a direct function of VP, which reflects the probability that a species vocalizes during a given 10 s window. Standard duration point counts should have higher rates of false absences on days with lower TVP, even with static species availability, because of the reduction in time windows where species are vocalizing.

We found that TVP was a useful metric for quantifying avian vocal activity, and believe that it is worthy of further study and use. TVP is a more stable indicator of overall activity levels than raw call count, total number of annotations, or total annotation length, because it smooths out natural differences in vocalization patterns across species. When annotation boxes are split if the gap between vocalizations exceeds a set time interval (5 s in this study), alternative activity metrics are highly dependent on the innate vocalization rate of a given species. For example, Hauxwell's Thrush (HATH) and Black-faced Antthrush (BFAT) were two of the most common species in this dataset, but they feature different vocalization styles. While HATH often take only short breaks of < 5 s between song bouts in the morning, BFAT sing with gaps between song bouts that are generally greater than 5 s. A HATH singing throughout the morning may only result in a small number of different annotation boxes, but a large total annotation length that

incorporates the time in-between song bouts, while a BFAT singing throughout the morning can feature a very high number of total annotations, but will not include the gaps between song phrases, because they exceed 5 s in length. In our data, HATH had a total annotation length of 25,308 seconds over 675 total bounding boxes, while BFAT had a total annotation length of 4,456 seconds across 1,500 individual bounding boxes; using TVP eliminates this type of artificial species-to-species variability in vocal activity estimates. As a result, TVP can improve comparisons of annotated bioacoustic datasets that employ slightly different annotation protocols, and can reduce manual analysis time by eliminating the need to measure the distance between vocalizations. While metrics that require species-level ID are highly contingent on observer skill level, driving inconsistencies in Neotropical bird surveys (Robinson *et al.* 2018), TVP supports the use of broader taxonomic groups for challenging identifications, potentially lowering time and experience barriers for researchers. We found that TVP correlated strongly with SR ($R^2 = 0.85$, $p < 0.001$), meaning that it may be a viable proxy for species richness, even without requiring species IDs, and should be easier to generate with automated approaches than metrics that require species-level IDs. Because TVP essentially represents soundscape abundance, it can also enable the calculation of abundance-based indices like Bray-Curtis dissimilarity and Shannon diversity with acoustic data.

Based on our collective results, we suggest that bioacoustic data should be collected as a complement to traditional avian surveys whenever possible. In addition to the short-term benefits for quantifying diel and day-to-day temporal variation that we documented, even small bioacoustic datasets serve as ecological time capsules (Sugai & Llusia 2019), which are particularly important in this time of rapid global change. ARUs can generate greater data

volume than traditional methods without meaningfully increasing field time or cost (Hobson *et al.* 2002, Acevedo & Villanueva-Rivera 2006, Tegeler *et al.* 2012). While manually analyzing acoustic data can be time-intensive, limiting its utility, acoustic indices (Jorge *et al.* 2018, Metcalf *et al.* 2020) and automated identification programs like BirdNET (Kahl *et al.* 2021) are continuously improving and could have a transformative effect on global ornithological research (Pérez-Granados 2023). Due to their relative rarity, fully-annotated tropical soundscapes are critical for developing these automated programs, and this dataset has already been used for this purpose (Kahl *et al.* 2020). Studies focused exclusively on birds can come at the expense of developing knowledge of other groups (Gardner *et al.* 2008), but bioacoustic data collected for ornithological research often contains vocalizations of non-target taxa and other soundscape elements, enabling more holistic biodiversity research (Newson *et al.* 2017). To help advance these efforts, we suggest that researchers collecting bioacoustic data make use of open-access repositories like Zenodo (European Organization For Nuclear Research & OpenAIRE 2013) to host their datasets.

We found that simultaneous passive acoustic monitoring coupled with manual annotation revealed significant diel and day-to-day variation in the vocal activity and observed species richness of an Amazonian bird community, an ecological pattern that would be very difficult to ascertain using traditional field methods. The magnitude of this temporal variation was large enough that it could mask meaningful differences between sites if temporal bias is not sufficiently accounted for in study designs. This research provides a case study for using passive acoustic monitoring to quantify temporal variation in tropical ornithological surveys, even with a relatively small sample of sites and days. Future studies could help explain the mechanisms for

this variation and improve methods for processing and interpreting large volumes of acoustic data in complex systems.

Authors' Contributions

W.A.H. and H.K. designed the methodology and secured the funding to conduct the study; W.A.H. and C.J.S. analyzed the data; W.A.H. and N.H.C. collected and annotated the recordings, and W.A.H. led the writing of the manuscript. All authors contributed critically to the drafts and gave final approval for publication. The Spanish abstract was translated by N.H.C.

Conflict of Interest

The authors declare that there is no conflict of interest.

Funding

Funding for equipment was provided by the K. Lisa Yang Center for Conservation Bioacoustics at the Cornell Lab of Ornithology (https://www.birds.cornell.edu/ccb/), with support from Innóvate Perú (https://www.proinnovate.gob.pe/), CORBIDI (http://www.corbidi.org/), and the Inkaterra Association (https://www.inkaterra.com/inkaterra-asociacion-org/en/). Travel expenses were funded by the Cornell Lab of Ornithology (https://www.birds.cornell.edu/home/).

Acknowledgements

We would like to thank the Inkaterra Association (ITA) staff for providing logistical support and excellent field station facilities, particularly Dennis Osorio and Kevin Jiménez Gonzales, who

helped set up recorders. John Fitzpatrick, Fernando Angulo, Will Sweet, Ken Rosenberg, and

Alex Wiebe helped identify unknown vocalizations, and Richard Fuller, Ken Rosenberg, Andrew

Farnsworth, Connor Wood, and several anonymous reviewers provided comments on the drafts.

Intellectual Contributions

Our study involved collaboration with multiple Peruvian NGOs, particularly the Inkaterra

Association (ITA), in both Lima and Madre de Dios. While in the field, we taught local

researchers how to set up and maintain Swift recorders, how to annotate recordings in Raven

Pro, and then left 9 recorders with them for their own use. We shared a copy of all collected data

and our results with ITA, and one of 'ITA's local researchers (Noe Huaraca-Charca) is a coauthor

on this paper.

Data Availability Statement

All software curated for this research are archived and available at

https://github.com/csayers2/Inkaterra-ARU. The raw acoustic data and annotations referenced in

this study can be found at https://zenodo.org/record/7079124 (DOI: 10.5281/zenodo.7079124).

References

Acevedo, M. A., & Villanueva-Rivera, L. J. (2006). Using Automated Digital Recording
    Systems as Effective Tools for the Monitoring of Birds and Amphibians. *Wildlife Society
    Bulletin*, *34*(1), 211–214. https://doi.org/10.2193/0091-7648(2006)34[211:uadrsa]2.0.co;2

Aide, T. M., Hernández-Serna, A., Campos-Cerqueira, M., Acevedo-Charry, O., & Deichmann,
    J. L. (2017). Species richness (of insects) drives the use of acoustic space in the tropics.
    Remote Sensing, *9*(11), 1–12. https://doi.org/10.3390/rs9111096

Alvarez-Berríos, N., Campos-Cerqueira, M., Hernández-Serna, A., Delgado, C. J. A., Román-Dañobeytia, F., & Aide, T. M. (2016). Impacts of small-scale gold mining on birds and anurans near the Tambopata Natural Reserve, Peru, assessed using passive acoustic monitoring. *Tropical Conservation Science*, *9*(2), 832–851. https://doi.org/10.1177/194008291600900216

Allan, J. R., Venter, O., & Watson, J. E. M. (2017). Temporally inter-comparable maps of terrestrial wilderness and the Last of the Wild. *Scientific Data*, *4*, 1–8. https://doi.org/10.1038/sdata.2017.187

Baldi, A. (1996). Edge effects in tropical versus temperate forest bird communities: Three alternative hypotheses for the explanation of differences. *Acta Zoologica Academiae Scientiarum Hungaricae*, *42*(3), 163–172.

Barlow, J., Peres, C. A., Henriques, L. M. P., Stouffer, P. C., & Wunderle, J. M. (2006). The responses of understorey birds to forest fragmentation, logging and wildfires: An Amazonian synthesis. *Biological Conservation*, *128*(2), 182–192. https://doi.org/10.1016/j.biocon.2005.09.028

Barlow, J., França, F., Gardner, T. A., Hicks, C. C., Lennox, G. D., Berenguer, E., … Graham, N. A. J. (2018). The future of hyperdiverse tropical ecosystems. *Nature*, *559*(7715), 517–526. https://doi.org/10.1038/s41586-018-0301-1

Berg, K. S., Brumfield, R. T., & Apanius, V. (2006). Phylogenetic and ecological determinants of the neotropical dawn chorus. *Proceedings of the Royal Society B: Biological Sciences*, *273*(1589), 999–1005. https://doi.org/10.1098/rspb.2005.3410

Blackburn, T. M., & Gaston, K. J. (1998). Some methodological issues in macroecology. *American Naturalist*, *151*(1), 68–83. https://doi.org/10.1086/286103

Blake, J. G. (1992). Temporal Variation in Point Counts of Birds in a Lowland Wet Forest in Costa Rica. *The Condor*, *94*(1), 265–275. https://doi.org/10.2307/1368816

Blake, J. G. (2021). Acoustic monitors and direct observations provide similar but distinct perspectives on bird assemblages in a lowland forest of eastern Ecuador. *PeerJ*, (January). https://doi.org/10.7717/peerj.10565

Blake, J. G., & Loiselle, B. A. (2015). Enigmatic declines in bird numbers in lowland forest of eastern Ecuador may be a consequence of climate change. *PeerJ*, *2015*(8), 1–20. https://doi.org/10.7717/peerj.1177

Borges, S. H. (2013). Bird species distribution in a complex Amazonian landscape: Species diversity, compositional variability and biotic-environmental relationships. *Studies on Neotropical Fauna and Environment*, *48*(2), 106–118. https://doi.org/10.1080/01650521.2013.841627

Broadbent, E. N., Asner, G. P., Keller, M., Knapp, D. E., Oliveira, P. J. C., & Silva, J. N. (2008). Forest fragmentation and edge effects from deforestation and selective logging in the Brazilian Amazon. *Biological Conservation*, *141*(7), 1745–1757. https://doi.org/10.1016/j.biocon.2008.04.024

Cerqueira, M. C., Cohn-Haft, M., Vargas, C. F., Nader, C. E., Andretti, C. B., Costa, T. V. V., … Ferraz, G. (2013). Rare or elusive? A test of expert knowledge about rarity of Amazon forest birds. *Diversity and Distributions*, *19*(7), 710–721. https://doi.org/10.1111/ddi.12033

Darras, K., Batáry, P., Furnas, B. J., Grass, I., Mulyani, Y. A., & Tscharntke, T. (2019). Autonomous sound recording outperforms human observation for sampling birds: a systematic map and user guide. *Ecological Applications*, *29*(6). https://doi.org/10.1002/eap.1954

Edelman, A. et al. State of the Tropics: 2014 Report (James Cook Univ., Cairns, 2014).Available at: https://researchonline.jcu.edu.au/35471/1/SOTT_ExcecutiveSummary_Print1.pdf

Estes, J. A., Terborgh, J., Brashares, J. S., Power, M. E., Berger, J., Bond, W. J., … Wardle, D. A. (2011). Trophic downgrading of planet earth. *Science*, *333*(6040), 301–306. https://doi.org/10.1126/science.1205106

European Organization For Nuclear Research & OpenAIRE. (2013). *Zenodo*. doi:10.25495/7GXK-RD71

Gardner, T. A., Barlow, J., Araujo, I. S., Ávila-Pires, T. C., Bonaldo, A. B., Costa, J. E., … Peres, C. A. (2008). The cost-effectiveness of biodiversity surveys in tropical forests. *Ecology Letters*, *11*(2), 139–150. https://doi.org/10.1111/j.1461-0248.2007.01133.x

Gil, D., & Llusia, D. (2020). *The Bird Dawn Chorus Revisited*. https://doi.org/10.1007/978-3-030-39200-0_3

Haddad, N. M., Brudvig, L. A., Clobert, J., Davies, K. F., Gonzalez, A., Holt, R. D., … Townshend, J. R. (2015). Habitat fragmentation and its lasting impact on 'Earth's ecosystems. *Science Advances*, *1*(2), 1–10. https://doi.org/10.1126/sciadv.1500052

Hart, P. J., Hall, R., Ray, W., Beck, A., & Zook, J. (2015). Cicadas impact bird communication in a noisy tropical rainforest. *Behavioral Ecology*, *26*(3), 839–842. https://doi.org/10.1093/beheco/arv018

Haselmayer, J., & Quinn, J. S. (2000). A comparison of point counts and sound recording as bird survey methods in amazonian southeast Peru. *Condor*, *102*(4), 887–893. https://doi.org/10.2307/1370317

Hernández-Palma, A., & Stouffer, P. C. (2018). Matrix and area effects on the nutritional condition of understory birds in Amazonian rainforest fragments. *Perspectives in Ecology and Conservation*, *16*(3), 139–145. https://doi.org/10.1016/j.pecon.2018.06.003

Hobson, K. A., Rempel, R. S., Greenwood, H., Turnbull, B., & Van Wilgenburg, S. L. (2002). Acoustic surveys of birds using electronic recordings: New potential from an omnidirectional microphone system. *Wildlife Society Bulletin*, *30*(3), 709–720.

Hull, J. M., Fish, A. M., Keane, J. J., Mori, S. R., Sacks, B. N., & Hull, A. C. (2010). Estimation of Species Identification Error: Implications for Raptor Migration Counts and Trend Estimation. *Journal of Wildlife Management*, *74*(6), 1326–1334. https://doi.org/10.2193/2009-255

Jenkins, C. N., Pimm, S. L., & Joppa, L. N. (2013). Global patterns of terrestrial vertebrate diversity and conservation. *Proceedings of the National Academy of Sciences of the United States of America*, *110*(28), E2603–E2610. https://doi.org/10.1073/pnas.1302251110

Jirinec, V., Elizondo, E. C., Rutt, C. L., & Stouffer, P. C. (2018). Space use, diurnal movement, and roosting of a Variegated Antpitta (Grallaria varia) in central Amazonia. *Ornitologia Neotropical*, *29*(February), 13–20. https://doi.org/10.58843/ornneo.v29i1.325

Jorge, F. C., Machado, C. G., da Cunha Nogueira, S. S., & Nogueira-Filho, S. L. G. (2018). The effectiveness of acoustic indices for forest monitoring in Atlantic rainforest fragments. *Ecological Indicators*, *91*(August 2017), 71–76. https://doi.org/10.1016/j.ecolind.2018.04.001

Kahl, S., Clapp, M., Hopping, W. A., Goëau, H., Glotin, H., Planqué, R., Vellinga,W. P., Joly, A. (2020). Overview of BirdCLEF 2020: Bird sound recognition in complex acoustic environments. In: CLEF task overview 2020, CLEF: Conference and Labs of the Evaluation Forum, Sep. 2020, Thessaloniki, Greece.

Kahl, S., Wood, C. M., Eibl, M., & Klinck, H. (2021). BirdNET: A deep learning solution for avian diversity monitoring. *Ecological Informatics*, *61*(December 2020), 101236. https://doi.org/10.1016/j.ecoinf.2021.101236

Kruskal, J.B. (1964) Non-metric Multidimensional Scaling: A Numerical Method. Psychometrika, 29, 115-129. http://dx.doi.org/10.1007/BF02289694

Laranjeiras, T. O., Naka, L. N., Leite, G. A., Cohn‐Haft, M. (2021). Effects of a major Amazonian river confluence on the distribution of floodplain forest avifauna, *Journal of Biogeography*, 48, 4, (847-860). https://doi.org/10.1111/jbi.14042

Latta, S. C., Tinoco, B. A., Astudillo, P. X., & Graham, C. H. (2011). Patterns and magnitude of temporal change in avian communities in the ecuadorian andes. *Condor*, *113*(1), 24–40. https://doi.org/10.1525/cond.2011.090252

Laurance, S. G. W. (2004). Responses of understory rain forest birds to road edges in central Amazonia. *Ecological Applications*, *14*(5), 1344–1357. https://doi.org/10.1890/03-5194

Laurance, W. F., Camargo, J. L. C., Luizão, R. C. C., Laurance, S. G., Pimm, S. L., Bruna, E. M., … Lovejoy, T. E. (2011). The fate of Amazonian forest fragments: A 32-year investigation. *Biological Conservation*, *144*(1), 56–67. https://doi.org/10.1016/j.biocon.2010.09.021

Laurance, W. F., Sayer, J., & Cassman, K. G. (2014). Agricultural expansion and its impacts on tropical nature. *Trends in Ecology and Evolution*, *29*(2), 107–116. https://doi.org/10.1016/j.tree.2013.12.001

Laurance, W. F., Useche, D. C., Rendeiro, J., Kalka, M., Bradshaw, C. J. A., Sloan, S. P., … Campbell, M. (2012). Averting biodiversity collapse in tropical forests. *Nature*, *489*, 290–294.

Lees, A. C., Naka, L. N., Aleixo, A., Cohn-Haft, M., Piacentini, V. de Q., Santos, M. P. D., & Silveira, L. F. (2014). Conducting rigorous avian inventories: Amazonian case studies and a roadmap for improvement. *Revista Brasileira de Ornitologia*, *22*(2), 107–120.

Lindell, C. A., Riffell, S. K., Kaiser, S. A., Battin, A. L., Smith, M. L., & Sisk, T. D. (2007). Edge Responses of Tropical and Temperate Birds. *The Wilson Journal of Ornithology*, *119*(2), 205–220. https://doi.org/10.1676/05-133.1

Lovejoy, T. (2013). Tropical forests under a changing climate and innovations in tropical forest management. *Tropical Conservation Science*, *6*(3), 311–314. https://doi.org/10.1177/194008291300600301

Lynch, J.F. 1995. Effects of Point Count Duration, Time-of- Day, and Aural Stimuli on Detectability of Migratory and Resident Bird Species in Quintana Roo, Mexico. Ralph, C. John; Sauer, John R.; Droege, Sam, Tech. Ed. 1995. *Monit. bird Popul. by point counts.* Gen. Tech. Rep. PSW-GTR- 149. Albany, CA U.S. Dep. Agric. For. Serv. Pacific Southwest Res. Stn. p. 1-6 149.

Macía, M. J. (2011). Spatial distribution and floristic composition of trees and lianas in different forest types of an Amazonian rainforest. *Plant Ecology*, *212*(No. 7 (July 2011)), 1159–1177. https://doi.org/10.1007/s11258-011-9895-1

Menger, J., Magnusson, W. E., Anderson, M. J., Schlegel, M., Pe'er, G., & Henle, K. (2017). Environmental characteristics drive variation in Amazonian understorey bird assemblages. *PLoS ONE*, *12*(2), 1–20. https://doi.org/10.1371/journal.pone.0171540

Metcalf, O. C., Barlow, J., Devenish, C., Marsden, S., Berenguer, E., & Lees, A. C. (2020). Acoustic indices perform better when applied at ecologically meaningful time and frequency scales. *Methods in Ecology and Evolution*, *2020*(October), 1–11. https://doi.org/10.1111/2041-210x.13521

Metcalf, O. C., Barlow, J., Marsden, S., Moura, N. G. De, Berenguer, E., Ferreira, J., & Lees, A. C. (2021). Optimizing tropical forest bird surveys using passive acoustic monitoring and

high temporal resolution sampling. *Remote Sensing in Ecology and Conservation*, 1–12. https://doi.org/10.1002/rse2.227

Milliken, W., Zappi, D., Sasaki, D., Hopkins, M., & Pennington, R. T. (2010). Amazon vegetation: How much don't we know and how much does it matter? *Kew Bulletin*, *65*(4), 691–709. https://doi.org/10.1007/s12225-010-9236-x

Moura, N. G., Lees, A. C., Aleixo, A., Barlow, J., Berenguer, E., Ferreira, J., … Gardner, T. A. (2016). Idiosyncratic responses of Amazonian birds to primary forest disturbance. *Oecologia*, *180*(3), 903–916. https://doi.org/10.1007/s00442-015-3495-z

Newson, S. E., Bas, Y., Murray, A., & Gillings, S. (2017). Potential for coupling the monitoring of bush-crickets with established large-scale acoustic monitoring of bats. *Methods in Ecology and Evolution*, *8*(9), 1051–1062. https://doi.org/10.1111/2041-210X.12720

Oksanen, J., Guillaume Blanchet, F., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin, P. R., O'hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E., & Wagner, H. (2019). vegan: Community ecology package. R package version 2.5-6. Retrieved from https:// CRAN.R-project.org/package=vegan

Parker, T. (1991). On the Use of Tape Recorders in Avifaunal Surveys. *The Auk: Ornithological Advances*, *108*(02), 443–444. https://doi.org/10.1093/auk/108.2.443

Pérez-Granados, C. (2023). BirdNET: applications, performance, pitfalls and future opportunities. *Ibis*. https://doi.org/10.1111/ibi.13193

Pérez-Granados, C., & Schuchmann, K. L. (2022). Diel and seasonal variation of Striped Cuckoo (Tapera naevia) vocalizations revealed using automated signal recognition. *Ibis*, *165*(1), 179–189. https://doi.org/10.1111/ibi.13129

Pillay, R., Fletcher, R. J., Sieving, K. E., Udell, B. J., & Bernard, H. (2019). Bioacoustic monitoring reveals shifts in breeding songbird populations and singing behaviour with selective logging in tropical forests. *Journal of Applied Ecology*, *56*(11), 2482–2492. https://doi.org/10.1111/1365-2664.13492

Pires, J.M., Prance, G. 1985. The vegetation types of the Brazilian Amazon. In: Prance GT, Lovejoy TE, editors. Key Environment – Amazonia. New York (NY): Pergamon Press. p. 109–145.

R Core Team. 2022. *R: a language and environment for statistical computing*. R Foundation for Statistical Computing, Vienna. https://www.R-project.org/

Ralph, C. J., Sauer, J. R., & Droege, S. (1995). Monitoring Bird Populations by Point Counts. *The Journal of Wildlife Management*, *61*(4), 1453. https://doi.org/10.2307/3802161

Redford, K. H. (1992). The Empty Forest. *BioScience*, *42*(No. 6 (Jun., 1992)), 412–422. https://doi.org/https://doi.org/10.2307/1311860

Rempel, R. S., Hobson, K. A., Holborn, G., Van Wilgenburg, S. L., & Elliott, J. (2005). Bioacoustic monitoring of forest songbirds: interpreter variability and effects of configuration and digital processing methods in the laboratory. *Journal of Field Ornithology*, *76*(1), 1–11. https://doi.org/10.1648/0273-8570-76.1.1

Remsen, J. V. (1994). Use and Misuse of Bird Lists in Community Ecology and Conservation. *The Auk*, *111*(1), 225–227. https://doi.org/10.2307/4088531

Robinson, W. D., Lees, A. C., & Blake, J. G. (2018). Surveying tropical birds is much harder than you think: a primer of best practices. *Biotropica*, *50*(6), 846–849. https://doi.org/10.1111/btp.12608

Rodriguez, A., Gasc, A., Pavoine, S., Grandcolas, P., Gaucher, P., & Sueur, J. (2014). Temporal and spatial variability of animal sound within a neotropical forest. *Ecological Informatics*, *21*, 133–143. https://doi.org/10.1016/j.ecoinf.2013.12.006

Shonfield, J., & Bayne, E. M. (2017). Autonomous recording units in avian ecological research: current use and future applications. *Avian Conservation and Ecology*, *12*(1). https://doi.org/10.5751/ace-00974-120114

Simons, T. R., Alldredge, M. W., Pollock, K. H., & Wettroth, J. M. (2007). Experimental analysis of the auditory detection process on avian point counts. *Auk*, *124*(3), 986–999. https://doi.org/10.1642/0004-8038(2007)124[986:EAOTAD]2.0.CO;2

Soares, L., Cockle, K. L., Inzunza, E. R., Ibarra, J. T., Miño, C. I., Zuluaga, S., … Echeverry-galvis, M. A. (2023). Neotropical ornithology: Reckoning with historical assumptions , removing systemic barriers , and reimagining the future. *Ornithological Applications*, *125*(1), 1–31. http://dx.doi.org/10.32942/osf.io/yu2fx

Simons, T. R., Alldredge, M. W., Pollock, K. H., & Wettroth, J. M. (2007). Experimental analysis of the auditory detection process on avian point counts. *Auk*, *124*(3), 986–999. https://doi.org/10.1642/0004-8038(2007)124[986:EAOTAD]2.0.CO;2

Stouffer, P. C., Jirinec, V., Rutt, C. L., Bierregaard, R. O., Hernández-Palma, A., Johnson, E. I., … Lovejoy, T. E. (2020). Long-term change in the avifauna of undisturbed Amazonian rainforest: ground-foraging birds disappear and the baseline shifts. *Ecology Letters*. https://doi.org/10.1111/ele.13628

Stutchbury, B. J. M., & Morton, E. S. (2008). Recent advances in the behavioral ecology of tropical birds. *Wilson Journal of Ornithology*, *120*(1), 26–37. https://doi.org/10.1676/07-018.1

Sugai, L. S. M., & Llusia, D. (2019). Bioacoustic time capsules: Using acoustic monitoring to document biodiversity. *Ecological Indicators*, *99*(December 2018), 149–152. https://doi.org/10.1016/j.ecolind.2018.12.021

Tegeler, A. K., Morrison, M. L., & Szewczak, J. M. (2012). Using extended-duration audio recordings to survey avian species. *Wildlife Society Bulletin*, *36*(1), 21–29. https://doi.org/10.1002/wsb.112

Terborgh, J., Robinson, S. K., Parker III, T. A., & Munn, C. A. (1990). Structure and Organization of an Amazonian Forest Bird Community. *Ecological Monographs*, *60*(2), 213–238.

Tuomisto, H., Ruokolainen, K., Kalliola, R., Linna, A., Danjoy, W., & Rodriguez, Z. (1995). Dissecting amazonian biodiversity. *Science*, *269*(5220), 63–66. https://doi.org/10.1126/science.269.5220.63

Verner, J. & Ritter, L.V. 1986. Hourly variation in morning point counts of birds. *The Auk,* 103, 117–124. https://doi.org/10.1093/AUK/103.1.117.

Whitworth, A., Beirne, C., Rowe, J., Ross, F., Acton, C., Burdekin, O., & Brown, P. (2015). The response of faunal biodiversity to an unmarked road in the Western Amazon. *Biodiversity and Conservation*, *24*(7), 1657–1670. https://doi.org/10.1007/s10531-015-0883-y

Wickham, H., Averick, M., Bryan, J., et al. 2019. Welcome to the tidyverse. *J. Open Source Softw*. 4:1686. https://doi.org/10.21105/joss.01686

Wimmer, J., Towsey, M., Roe, P., & Williamson, I. (2013). Sampling environmental acoustic recordings to determine bird species richness. *Ecological Applications*, *23*(6), 1419–1428.

Wolfe, J. D., Stouffer, P. C., Mokross, K., Powell, L. L., & Anciaes, M. M. (2015). Island vs. countryside biogeography: An examination of how Amazonian birds respond to forest clearing and fragmentation. *Ecosphere*, *6*(12), 1–14. https://doi.org/10.1890/ES15-00322.1

Woltmann, S. (2005). Patterns of daily temporal variation in detectability of forest birds in Bolivia. *Ornitologia Neotropical*, *16*(3), 337–346.

Woltmann, S., Kreiser, B. R., & Sherry, T. W. (2012). Fine-scale genetic population structure of an understory rainforest bird in Costa Rica. *Conservation Genetics*, *13*(4), 925–935. https://doi.org/10.1007/s10592-012-0341-2

Wood, S.N. (2004). Stable and Efficient Multiple Smoothing Parameter Estimation for Generalized Additive Models. *J. Am. Stat. Assoc.* 99: 673–686.

Wood, S.N. (2011). Fast stable restricted maximum likelihood and marginal likelihood estimation of semiparametric generalized linear models. *J. R. Stat. Soc. Ser. B Stat. Methodol.* 73: 3–36.

Wood, S.N. (2017). *Generalized Additive Models: An Introduction with R (2nd edition)*. Chapman and Hall/CRC.

Wood, S.N., Pya, N. & Säfken, B. (2016). Smoothing Parameter and Model Selection for General Smooth Models. *J. Am. Stat. Assoc.* 111: 1548–1563.

Venier, L. A., Holmes, S. B., Holborn, G. W., McIlwrick, K. A., & Brown, G. (2012). Evaluation of an automated recording device for monitoring forest birds. *Wildlife Society Bulletin*, *36*(1), 30–39. https://doi.org/10.1002/wsb.88

Verner, J., & Ritter, L. V. (1986). Hourly Variation in Morning Point Counts of Birds. *The Auk*, (January), 117–124.

Yang, L., Qin, G., Zhao, N., Wang, C. & Song, G. (2012). Using a generalized additive model with autoregressive terms to study the effects of daily temperature on mortality. *BMC Med. Res. Methodol.* 12: 165.

Swift recorder placement in "várzea" forest habitat

Peru

Madre de Dios

Madre de Dios

Study Area

Puerto Maldonado

Recent land clearing

Site F

Site A

Site B

Site C

Site D

Site E

Field Station

Inkaterra Reserva Amazónica Ecolodge

Madre de Dios River

Legend

Swift Recorder Placement

Reserve Boundary

Image © 2023 CNES / Airbus

Google Earth

Imagery Date: 7/29/2019   12°31'24.59" S   69°02'08.45" W   elev   0 m   eye alt   4.31 km

A)

Species	0-5min	5-10min	10-15min	15-20min	20-25min	25-30min	30-35min	35-40min	40-45min	45-50min	50-55min	55-60min
Cinerous Tinamou	5	15	69	59	29	46	61	17	22	3	8	1
Little Tinamou	7	34	43	39	25	18	10	17	8	3	0	1
Undulated Tinamou	9	33	39	38	51	44	50	47	58	47	53	43
Plumbeous Pigeon	0	1	3	1	31	15	26	26	18	42	40	67
Ruddy Pigeon	0	1	1	0	8	2	2	22	26	14	0	0
Grey-fronted Dove	0	1	1	13	15	40	79	88	39	13	12	20
Amazonian Pygmy-Owl	51	48	21	4	1	2	0	0	0	0	0	0
Black-tailed Trogon	0	0	0	1	3	2	17	7	11	23	13	18
Collared Trogon	0	0	0	21	20	3	15	20	48	53	47	21
Amazonian Motmot	38	31	52	90	69	54	42	14	15	10	15	0
Mealy Parrot	0	0	0	0	7	9	21	29	23	10	5	11
Fasciated Antshrike	4	5	1	2	8	31	21	49	35	31	3	6
Plain-winged Antshrike	0	0	19	46	39	39	20	15	3	1	0	8
Dusky-throated Antshrike	0	1	0	6	14	10	17	24	30	33	19	13
Spot-winged Antshrike	0	0	20	26	30	10	13	23	13	8	5	2
Pygmy Antwren	0	0	0	0	8	11	10	12	19	17	11	10
Long-winged Antwren	4	19	17	39	27	3	6	3	3	12	0	8
Grey Antbird	0	0	10	37	40	43	60	100	93	69	80	57
Plumbeous Antbird	0	0	0	0	2	2	36	76	75	59	60	43
Goeldi's Antbird	0	2	13	18	20	27	19	23	22	18	17	7
Black-throated Antbird	0	0	1	0	4	4	2	14	28	18	8	0
Black-faced Antthrush	7	29	55	114	118	148	164	194	163	184	190	148
Rufous-fronted Antthrush	0	0	0	0	1	4	17	24	19	15	19	21
Cinnamon-throated Woodcreeper	7	25	39	24	21	4	20	8	13	6	0	1
Long-billed Woodcreeper	0	6	10	25	14	9	12	6	3	6	12	9
Amazonian Barred-Woodcreeper	6	46	62	50	17	14	1	0	1	0	0	2
Striped Woodcreeper	0	6	17	41	36	47	27	21	15	11	0	1
Buff-throated Woodcreeper	25	58	86	104	67	60	52	38	34	16	10	8
Purple-throated Fruitcrow	0	0	0	0	9	8	22	17	18	9	16	9
Screaming Piha	2	3	2	10	24	31	27	30	32	54	57	54
White-winged Becard	0	6	31	54	65	54	33	48	52	44	31	26
Wing-barred Piprites	0	0	1	1	4	2	9	2	7	19	35	29
Yellow-margined Flycatcher	0	1	0	1	0	1	1	3	24	24	20	19
Dusky-capped Flycatcher	0	0	0	0	0	3	13	33	15	17	4	0
Piratic Flycatcher	2	2	7	14	24	13	8	12	15	13	12	8
Dusky-capped Greenlet	0	0	0	0	11	32	28	38	74	99	99	75
Thrush-like Wren	1	5	20	44	73	100	90	122	101	102	114	119
Buff-breasted Wren	0	1	5	13	14	30	65	76	36	38	40	54
Hauxwell's Thrush	18	40	46	46	67	63	79	68	74	82	88	88
Russet-backed Oropendola	0	0	0	1	0	0	8	22	25	48	41	14
Amazonian Grosbeak	8	24	55	58	57	46	52	52	48	30	25	0

B)