Simultaneous passive acoustic monitoring uncovers evidence of potentially overlooked temporal variation in an Amazonian bird community

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Abstract

The vocal activity and detectability of tropical birds are subject to high levels of temporal

heterogeneity, but quantifying patterns of diel and day-to-day variation in complex systems is

challenging with traditional point count methods. As a result, research concerning stochastic

temporal effects on tropical avian assemblages is limited, typically offering only broad

conclusions, e.g., overall activity is highest in the first few hours of the morning and some

species are active at different times of the day. Passive acoustic monitoring introduces several

advantages for studying temporal variation, particularly by enabling simultaneous and

continuous data collection across adjacent sites. Here, we employed autonomous recording units

to quantify temporal variation in avian vocal activity and observed species richness at an

Amazonian reserve in Madre de Dios, Peru—a region featuring some of Earth's richest, most

complex avian assemblages. We manually annotated 18 dawn hour recordings, collected

simultaneously from three separate days at the same six sites, which represent various

microhabitats and avian community compositions. We documented significant and consistent

temporal variation in avian vocal activity levels and observed species richness within the dawn

hour and across days. We found that temporal effects were stronger for vocal activity than for

observed species richness and that vocal activity patterns over the course of the dawn hour varied

on a species-to-species basis. Our results indicate that overlooked temporal variation in

Amazonian soundscapes may obfuscate the results of surveys that fail to sufficiently account for

temporal variables with simultaneous monitoring. While manual analysis of large volumes of

soundscape data remains challenging, such data should be collected as a supplement to

traditional surveys whenever possible. Rapidly advancing research concerning the automated

processing of acoustic data could lead to more efficient methods for reducing temporal bias and

improving the calibration and accuracy of ornithological surveys in the tropics.

46
47   **Keywords:** avian biodiversity, bioacoustics, passive acoustic monitoring, Amazon rainforest,
48   temporal variation.
49

50 **Resumen**

La actividad vocal y la detectabilidad de las aves tropicales están sujetas a altos niveles de

heterogeneidad temporal, pero cuantificar con precisión los patrones de variación diurna y diaria

en sistemas complejos es un reto con los métodos tradicionales de conteo por puntos. En

consecuencia, la investigación sobre los efectos temporales estocásticos en los conjuntos de aves

tropicales es limitada y normalmente sólo ofrece conclusiones generales, como que la actividad

general es mayor en las primeras horas de la mañana y que algunas especies están activas en

distintos momentos del día. El monitoreo acústico pasivo presenta varias ventajas sobre los

métodos tradicionales para estudiar la variación temporal, en particular al permitir el recojo

simultáneo de datos en lugares adyacentes. En este trabajo empleamos unidades de grabación

autónomas para cuantificar la variación temporal en los niveles de actividad vocal de las aves y

la riqueza de especies observada en una reserva amazónica en Madre de Dios, Perú, una región

que cuenta con algunos de los conjuntos de aves más ricos y complejos del planeta. Anotamos

manualmente 18 grabaciones de la hora del amanecer, recogidas simultáneamente en tres días

distintos en los mismos seis lugares, que representan una variedad de microhábitats y

composiciones de comunidades de aves. Se documentó una variación temporal significativa y

consistente en los niveles de actividad vocal de las aves y en la riqueza de especies observada

dentro de la hora del amanecer y a lo largo de los días. Descubrimos que los efectos temporales

eran mayores en la actividad vocal que en la riqueza de especies observada, y que los patrones de

actividad diurna variaban de especie a especie a escala fina. Nuestros resultados indican que la

variación temporal que se pasa por alto en los paisajes sonoros amazónicos puede ofuscar los

resultados de los estudios que no tienen suficientemente en cuenta las variables temporales con

un seguimiento simultáneo. Nuevas investigaciones sobre el tratamiento automatizado de los

datos acústicos podrían conducir a métodos más eficaces para reducir el sesgo temporal y

mejorar la calibración y precisión de los estudios ornitológicos en los trópicos.

1 Introduction

The tropics account for an overwhelming share of Earth's avian diversity, featuring more than

75% of all species and over 90% of terrestrial birds (Barlow *et al.* 2018). Still, tropical regions

are underrepresented in ornithological literature relative to temperate regions, even though

tropical birds have fundamentally different life history strategies, behavioral ecology, and vocal

patterns than temperate birds, requiring separate study (Stutchbury & Morton 2008). The largest

intact tropical system on the planet is the Amazon basin (Allan *et al.* 2017), which is the

epicenter of global biodiversity (Antonelli *et al.* 2018). Like most of the tropics, the Amazon is

undergoing rapid ecological change, primarily the result of agricultural expansion (Lapola *et al.*

2023), exacerbated by the increasing impacts of climate change (Xu *et al.* 2020). These

anthropogenic impacts are significantly outpacing natural processes in Amazonia (Albert *et al.*

2023), threatening the resilience and stability of the region (Lovejoy & Nobre 2019, Boulton *et*

al. 2022). Despite its importance, however, the Amazon's avifauna remains poorly documented

and is subject to flawed baseline species occurrence data and significant knowledge gaps (Lees *et*

al. 2014), which are exacerbated by systemic barriers to researchers from the Global South

(Soares *et al.* 2022). Efforts to increase ornithological survey coverage in the region are

complicated by the difficulty of tropical field surveys. Overwhelming avian species richness,

including many rare and similar species, challenging logistics, and poor visibility conditions

contribute to the unreliability of avian surveys in Amazonia, where upwards of 95% of birds are

heard but never seen by a field observer (Robinson *et al.* 2018).

The vocal activity and detectability of Neotropical birds are subject to high levels of

temporal heterogeneity. Possible drivers of this variation include seasonality (Pérez-Granados &

Schuchmann 2022), weather and feeding opportunities (Metcalf *et al.* 2021), foraging strata

(Berg *et al.* 2006), temporal partitioning to avoid signal masking from other birds (Planqué &

Slabbekoorn 2008, Luther 2008, Luther 2009, Hart *et al.* 2021) and vocalizing insects (Hart *et al.*

2015, Alvarez-Berríos *et al.* 2016, Aide *et al.* 2017, Metcalf *et al.* 2020), along with a wide range

of poorly-understood floristic, geographic, and spatial variables (Menger *et al.* 2017). Some

broad patterns of diel variation in the detectability and vocal activity of tropical birds are well

established; for example, birds are most vocally active during the first 2–3 hours after dawn

(Lynch 1995, Woltmann 2005). However, the temporal variation of tropical soundscapes is more

complex than hour-long intervals can capture (Verner & Ritter 1986, Rodriguez *et al.* 2014,

Metcalf *et al.* 2021). Many Amazonian species are only aurally detectable within strict temporal

niches (Gil & Llusia 2020). These species may only vocalize during a single 5-minute period in

the morning, perhaps accompanied by an even shorter bout in the evening (Parker 1991). Some

Amazonian species have been shown to sing as infrequently as twice in 50 days (Jirinec *et al.*

2018). In addition to affecting estimates of total species richness, this fine-scale temporal

variation in vocal activity can also influence patterns of observed community composition.

Temporal effects on detectability vary by functional or taxonomic group. For example, canopy

species are known to reach their activity peak later in the morning than understory species (Blake

1992), and species with high sensitivity to habitat fragmentation may have proportionately

higher detection rates in pre-dawn surveys than less sensitive species (Woltmann 2005).

Quantifying the fine-scale patterns of temporal variation in tropical bird communities,

both within and across days, is difficult. Non-simultaneous survey methods are subject to a litany

of spatial, temporal, and observer biases, and therefore, capturing such granular effects can

require infeasibly large sample sizes (Lynch 1995). For example, Esquivel & Peris (2008) found

that four visits per point are necessary to account for the temporal variation in bird activity in the

Atlantic Forest of Paraguay. This sampling design does not account for possible variation in

vocal activity levels between different days, between different observers (Robinson *et al.* 2018),

nor travel time between sites. Additionally, Rodriguez *et al.* (2014) found that temporal effects

on acoustic activity patterns were spatially heterogeneous at adjacent recording sites in French

Guiana. Therefore, precise assessments of temporal variation may be required on a per-site basis.

More costly and labor-intensive approaches, for example, using a team of point count

technicians to visit multiple points simultaneously, may also introduce observer effects.

Amazonian soundscapes include thousands of vocalization types, inducing high rates of false-

negative and false-positive identification errors (Remsen 1994), and it can take months for

skilled observers to reach a level of competence attainable in just one to two weeks of field

experience in temperate systems (Parker 1991). Observer error rates have been experimentally

quantified in relation to aural identification (Simons *et al.* 2007), visual identification (Hull *et al.*

2010), and the subjectivity of abundance estimates (Cerqueira *et al.* 2013). These dynamics are

not limited to inexperienced surveyors; established ornithologists can also be subject to non-

trivial identification error rates in environments as complex as Amazonia (Lees *et al.* 2014).

Overlooked species tend to reflect non-random subsets of assemblages, and the specific nature of

these biases may differ between individual observers, even amongst experts (Blackburn &

Gaston 1998). Therefore, standardized surveys that do not produce archivable raw data and seek

to uncover subtle spatiotemporal variation in avian vocal activity, observed species richness, and

community structure should ideally be conducted by a single observer (Blake & Loiselle 2015,

Robinson *et al.* 2018).

Simultaneous passive acoustic monitoring (PAM) conducted with autonomous recording

units (ARUs) can eliminate much of the temporal bias in surveys by collecting continuous data

from adjacent locations simultaneously (Venier *et al.* 2012, Tegeler *et al.* 2012). Such effects are

difficult to quantify with traditional surveys, and establishing their presence and magnitude may

be useful for improving the calibration of point count studies, even with relatively small volumes

of ARU data. PAM can also address observer bias since data collection is independent of

technician skill level, and recordings collected simultaneously can be annotated by a single

observer, replayed and reanalyzed, archived publicly, and distributed to experts for an

independent review of identifications (Rempel *et al.* 2005, Robinson *et al.* 2018). PAM has been

shown to present a suitable supplement or alternative to point counts, both in Amazonia

(Haselmayer & Quinn 2000), and in general (Shonfield & Bayne 2017, Darras *et al.* 2019, Blake

2021), while providing numerous other advantages (Acevedo & Villanueva-Rivera 2006,

Newson *et al.* 2017, Darras *et al.* 2018, Jorge *et al.* 2018, Pillay *et al.* 2019, Sugai & Llusia

2019).

This study aimed to investigate the hypothesis that Amazonian soundscapes may be

subject to overlooked temporal variation in avian vocal activity, both between days and

throughout a single dawn hour, which in the absence of simultaneous survey methods, could

mask important differences in vocal activity levels, community composition, and observed

species richness between sites. We tested this hypothesis by deploying an array of ARUs at a

single Amazonian reserve in Madre de Dios, Peru, and manually annotated a subsample of the

resulting bioacoustic data. We predicted that stochastic effects of diel and day-to-day variation

on vocal activity and observed species richness would be consistent across sites and, therefore,

would be possible to control for with simultaneous monitoring.

2 Materials & Methods

2.1 Study area

We conducted this study at Inkaterra Reserva Amazónica (ITRA) in the Madre de Dios

Department of southeastern Peru (12°32'07.8" S, 69°02'58.2" W). This 191-ha private reserve is

located at an elevation of approximately 200 m along the Madre de Dios River, directly across

from the Reserva Nacional de Tambopata—one of the most biodiverse regions on the planet

(Foster *et al.* 1994) (Fig. 1). The habitat at ITRA primarily consists of *várzea* floodplain forest,

the most endangered forest type in the southwestern Amazon (Phillips *et al.* 1994), interspersed

with seasonally flooded *Mauritia* palm swamp forest.

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184 **Figure 1: Swift recorder locations and typical forest habitat at Inkaterra Reserva**
185 **Amazónica, Madre de Dios, Peru.** Recorder locations covered a range of habitats, including
186 edge habitat adjacent to clearings for small-scale agriculture (Site A), forest degraded by
187 selective logging outside of the reserve's boundaries (Site F), mature *várzea* in the reserve
188 interior (Site C and Site E), and mixed palm swamp (*aguajal*)-*várzea* habitat (Site B and Site D).
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2.2 Acoustic monitoring

We collected acoustic data from 14 January to 2 February 2019, using six ARUs (Swift recorder,

K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Cornell

University, Ithaca, NY, USA). We deployed the ARUs across the entire reserve, at a minimum

distance of 350 m from each other and 450 m from the river, to limit spatially overlapping

detections (pseudoreplication) and background noise (Ralph *et al.* 1995, Yip *et al.* 2017, Darras

et al. 2018, Haupert *et al.* 2022), respectively. Deployment locations were chosen to represent a

gradient of intactness and forest habitat type, including edge habitat adjacent to clearings for

small-scale agriculture (Site A), forest degraded by selective logging outside of the reserve's

boundaries (Site F), mature *várzea* in the reserve's interior (Site C and Site E), and mixed palm

swamp (*aguajal*)-*várzea* habitat (Site B and Site D) (Fig. 1). Placed at the height of 1.5 m above

the ground to maximize the sound detection space and minimize sound shadows (Darras *et al.*

2018), the 6 ARUs recorded continuously (mono, WAV format) throughout the deployment

period using a sampling rate of 48 kHz (16-bit resolution) and a gain setting of 35 dB. The

microphone sensitivity of the Swift was –44 dBV/Pa (+/– 3 dB) and featured a flat frequency

response (+/– 3 dB) in the frequency range of the vocalizations of interest. The clipping level of

the analog-to-digital (ADC) converter was +/– 0.9 V.

2.3 Annotation process

We manually annotated eighteen total dawn hours, from 05:00–06:00 PET (10:00–11:00 UTC),

representing six sites on three days, using the Raven Pro Sound Analysis Software (version 1.6;

K. Lisa Yang Center for Conservation Bioacoustics, Cornell Lab of Ornithology, Cornell

University, Ithaca, NY, USA). Annotations were performed by a single observer, the lead author,

to control for observer bias. A local expert, Noe Huaraca-Charca, also annotated a subset of these

recordings (n = 4) as a quality control measure. The three days, 16 January (Day 1), 20 January

(Day 2), and 31 January (Day 3), 2019, were randomly chosen from a subset of days that

featured less than 5 min of heavy rain during the dawn hour. The dawn hour was selected

because it typically contains the highest level of vocal activity and diversity among Neotropical

birds (Berg *et al.* 2006), including many species that only sing in narrow time windows around

first light (Parker 1991, Blake 1992). Additionally, the dawn soundscape in Amazonia features

far less anthropogenic activity and insect noise than in other periods, reducing interference with

avian species detection (Metcalf *et al.* 2020). While annotating short, randomly-selected samples

spread out over a longer survey period can improve estimates of alpha and gamma diversity

(Wimmer *et al.* 2013, Metcalf *et al.* 2021), this study was designed to assess temporal variation

over the course of a contiguous hour, and annotations were therefore conducted in this format.

Spectrograms were viewed in 60 s increments split over two rows (Hann window, 3 dB

bandwidth = 100 Hz, 690 FFT window size, 1024-sample DFT, 95% temporal overlap) (Fig. 2).

To reduce bias introduced by improved sound identification knowledge developed throughout

the project, the order in which recordings were annotated was determined using a random

number generator. Vocalizations were labeled for both species and a handful of broader groups

with confusing or indistinguishable calls (e.g., "TRSP" as "trogon species" [Trogonidae sp.]).

Background vocalizations, audible but not clearly visible on a spectrogram, were denoted with a

'1' and excluded from analysis. To account for inter-specific variation in call volume and

frequency, consistency for applying 'background' designation was approached on a species-

specific basis. In cases where vocalizations from the same species were separated by fewer than

5 s, they were included as part of the same annotation, while annotations were split if the gap

between vocalizations exceeded 5 s. An assortment of unlabeled sound clips featuring identified

and unidentified species were sent to regional experts for secondary verification. Annotations

were reviewed repeatedly until unidentified vocalizations made up less than 10% of the total

avian vocal activity on a given recording. We made considerable effort to accurately label as

many foreground and background vocalizations as possible to improve the value of this dataset

for training soundscape-based automated identification algorithms.

Figure 2: The annotation process. Spectrograms showing one minute of sound from Site B,
Day 2, before and after manual annotation in Raven Pro 1.6.
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2.4 Statistical analyses

To assess fine-scale temporal variation in dawn chorus vocal activity within recordings and

across sampling sites and days, we subsampled each 60 min recording at 1 min intervals, totaling

1,080 samples across all sites and days. For each 1 min interval, we calculated species richness

(SR), vocal prevalence (VP), and total vocal prevalence (TVP). Species richness refers to the

number of unique species detected, while vocal prevalence refers to the number of 10 s intervals

featuring a vocalization of a given species. Therefore, a single species could attain a maximum

VP of 6 per 1 min interval or 360 per hour recording. Total vocal prevalence (TVP)—a proxy for

the total avian vocal activity in a soundscape—refers to the sum of VP for all detected taxa in a

given subsample. We assessed the relationship between TVP and SR using a Pearson's *R*

correlation analysis.

259   Certain taxa were excluded from species richness calculations for data quality reasons:

parrots (Psittacidae sp.), because they occurred almost exclusively as calling flyovers, as well as

Columbiformes (pigeons and doves), and three congeneric tinamou species, Cinereous Tinamou

(*Crypturellus cinereus*), Little Tinamou (*C. soui*), and Bartlett's Tinamou (*C. bartletti*), which

had overlapping vocalizations and were inconsistently identified with confidence. However,

these species were included in calculations of TVP, as were vocalizations that were identified to

a broader group but not at the species level (e.g., Trogon sp. [Trogonidae sp.]) (For a full list of

species and exclusions for SR and TVP calculations, see Supporting Information Appendix S1).

2.4.1 Non-metric multidimensional scaling ordination (NMDS) and ANOVA

We used non-metric multidimensional scaling ordination (NMDS) and two-way repeated-

measures ANOVA to assess evidence for significant effects of site-specific differences in

community composition on SR and TVP. Substituting VP in place of abundance, we calculated

NMDS using the "bray" method, referring to Bray-Curtis dissimilarity, and metaMDS function

from the "vegan" package in R (Okansen *et al.* 2019). One advantage of using Bray-Curtis

dissimilarity is its ability to account for the relative abundances of different species or taxa

(Ricotta & Podani 2017). We plotted NMDS ordinations on two axes and ensured that the output

met minimum stress requirements (Kruskal 1964). To compare SR and TVP between sites, days,

and recordings, two-way repeated-measures ANOVA followed by Tukey's HSD tests for

multiple comparisons were conducted in R, with site and day as interacting independent

variables and the 1-minute time interval as a blocking variable.

2.4.2 Generalized Additive Models (GAMs)

We then constructed a set of Generalized Additive Models (GAMs) fit with Poisson distributions

and restricted maximum likelihood estimations to account for the non-linearity of within-

recording response curves for SR, TVP, and species-specific VP within the dawn hour. For both

the SR and TVP models, we included a smoothing term for the interaction between time of day

and recording to produce individual response curves for each recording, and a nested site-day

random effects term to account for pseudoreplication. Similarly, to produce species-specific VP

response curves across days and sites, we replicated this random effects structure, but removed

the interaction for the time interval smoothing term. As the amount of species present at time

step t directly influences the number of species that are likely to be detected at time step $t + 1$, we

accounted for this inherent temporal autocorrelation by fitting both models with a first-order

autoregressive covariance structure (AR1; Yang et al. 2012). This decision successfully removed

evidence of autocorrelation within model residuals. Because our model structure was

biologically informed, we chose to forgo model selection and base inferences and predictions

from this structure. We completed all statistical analyses in R (v. 4.2.0; R Core Team 2022)

using functions from the "tidyverse" (Wickham et al. 2019) and employing the "mgcv" package

to construct GAMs (Wood et al. 2016; Wood 2004, 2011, 2017).

3 Results

We identified 127 species over the 18 analyzed dawn hours, seven of which were recorded only

as background vocalizations; 17,036 individual bounding-box annotations were drawn in Raven

Pro 1.6, of which 94.8% were identified to the species or taxon level. Site A had the lowest

average species richness per 1 min interval (4.0; 43 total), followed by Site F (4.4; 42 total), Site

D (4.8; 49 total), Site C (4.9; 46 total), Site E (5.3; 55 total), and Site B (6.1; 50 total), which was

also the most active site by TVP. Aggregating detections across sampling sites, Day 1 featured

57 species, Day 2 featured 78 species, and Day 3 featured 56 species, with average SR values per

1 min interval of 4.79, 6.18, and 3.81, respectively. At individual sites, the percentage of total

vocal prevalence (TVP) that was not identified to a species or taxa group ranged from 4.1% (Site

E) to 7.1% (Site B) between sites, and from 5.1% (Day 3) to 5.4% (Day 1) between days.

311   Of the 120 foreground species detected, only two, Black-faced Antthrush (*Formicarius*

analis) and Buff-throated Woodcreeper (*Xiphorhynchus guttatus*), were detected on all 18

recordings, and six more, Hauxwell's Thrush (*Turdus hauxwelli*), Little Tinamou (*Crypturellus*

soui), Thrush-like Wren (*Camplyothynchs turdinus*), Amazonian Motmot (*Momotus momota*),

Plumbeous Pigeon (*Patagioenas plumbea*), and Screaming Piha (*Lipaugus cociferans*) were

detected on at least 15 recordings. Most species were rare; more than half ($n = 66$, 55%) were

detected on 4 or fewer 1-hr recordings, and roughly a quarter ($n = 27$, 23%) on only a single

recording (Supporting Information Appendix S2). Regarding total vocal prevalence, the most

abundant species were Hauxwell's Thrush, Black-faced Antthrush, and Thrush-like Wren. TVP

was highest at all six sites on 20 January (Day 2) and lowest on 31 January (Day 3) at every site

except Site B.

3.1 Non-metric multidimensional scaling ordination (NMDS) and ANOVA

Individual sites consistently clustered separately from each other in terms of their species

composition, as demonstrated by non-metric multidimensional scaling (NMDS) ordination (Fig.

3). Tukey's HSD tests indicated that a higher percentage of pairwise comparisons between sites

were significantly different ($p < 0.01$) for SR (9/15, 60%) than for TVP (6/15, 40%), and all

comparisons between days were significant for both TVP and SR. The rate that recordings from

different days exhibited significant pairwise differences ($p < 0.01$) remained essentially identical

whether they came from the same site or from different sites: 44.4% (8/18, same site) versus

42.2% (38/90, different sites) for SR and 55.6% (10/18) versus 56.7% (51/90) for TVP. In

general, pairwise comparisons of recordings from different days were more likely to differ

significantly ($p < 0.01$) than recordings from different sites. For SR, 42.6% of recordings from

different days (46/108) differed significantly, compared to 35.6% from different sites (48/135).

For TVP, 56.5% of recording pairings from different days (61/108) differed significantly,

compared to 45.2% from different sites (61/135). TVP and SR were least likely to differ between

recordings that came from different sites on the same day; just 22.2% of these pairwise

comparisons (10/45) differed significantly ($p < 0.01$) for each metric. For all Tukey's HSD

pairwise comparisons between sites, days, and individual recordings, see Supporting Information

Appendix S3.

342
343 **Figure 3: Non-metric multidimensional scaling (NMDS) ordination** by site using Bray-Curtis
344 dissimilarity, with vocal prevalence in place of abundance. Significantly representative species
345 along the ordination axes ($p < 0.02$) are shown with vectors. For site information, see Fig. 1.
346

3.2 Generalized Additive Models (GAMs)

SR and TVP models accounted for a substantial amount of variation in dawn chorus vocal

activity ($R^2 = 0.64$ and $R^2 = 0.83$, respectively, Fig. 4), in which smoothing terms for the

interaction between time interval and recording, as well as site-day random intercepts, were often

significant ($p < 0.05$; Supporting Information Appendix S4). The effect of time interval and day

was generally stronger than the effect of site in explaining variation in observed SR and TVP,

and TVP was strongly correlated with SR ($r = 0.85$, $p < 0.001$). We observed that diel variation

in avian vocal activity varied by species and taxonomic group, with the VP of different species

peaking at different time intervals within the dawn hour (Fig. 5).

Figure 4: Comparisons of (A) species richness and total vocal prevalence detected by time of day across days and sites overlaid with Generalized Additive Model (GAM) prediction curves ± 95% confidence intervals, along with (B) site-specific GAM prediction curves overlaid with day-specific mean GAM prediction curves ± 95% confidence intervals of species richness and vocal prevalence across days.

4 Discussion

We used simultaneous passive acoustic monitoring to quantify temporal variation in the vocal activity level and observed species richness of an Amazonian bird community. In line with our primary hypothesis, the magnitude of this temporal variation is large enough to potentially mask, or even supersede, observed differences in species richness and vocal activity levels between sites and habitat types—complicating the interpretation of point count studies and other non-simultaneous survey methods that fail to control for temporal variables. While temporal variation in avian activity over the course of the morning is well-documented in the tropics, the limitations of traditional non-simultaneous survey methods mean that point count studies on temporal effects in these systems offer only broad conclusions, typically relating to maximum temporal resolutions of one hour or more (Blake 1992, Lynch 1995, Woltmann 2005, Esquivel & Peris 2008). Fine-scale diel variation in the vocal activity and detectability of Amazonian birds has been referenced anecdotally for decades (Parker 1991). However, thanks to the development of autonomous recording units, which allow for considerably easier repeat sampling than traditional methods, we are beginning to understand more about these effects at higher temporal resolutions (Gil & Llusia 2020, Blake 2021, Metcalf *et al.* 2021, Pérez-Granados & Schuchmann 2023). While prior temporal effects studies have focused primarily on diel and seasonal temporal variation, to our knowledge, this is the first study to provide evidence that day-to-day temporal variation in the vocal activity and detectability of Amazonian bird communities may be strong enough to obscure statistically significant differences between sites, in the absence of simultaneous data collection.

The variable diel patterns in avian vocal activity that we observed between species and

taxonomic groups (Fig. 5) mean that visiting two sites with the same overall SR or TVP at

different times in the morning could result in considerably different observations of community

composition, even when both visits occur within the same hour. This species-to-species diel

variation in vocal activity and detectability may partially or entirely explain why the diel

variation patterns in SR and TVP that we documented within the dawn hour were spatially

heterogeneous. Consistent with the findings of Rodriguez *et al.* (2014), these patterns varied per-

site, even between sites in the same general habitat type and superficial condition. Because sites

with differing communities should be expected to exhibit differing patterns of diel variation in

overall vocal activity, the differing patterns of diel variation in TVP and SR that we documented

between sites are plausibly the result of the spatial heterogeneity of avian community

composition (Fig. 3).

A)

Species	0-5min	5-10min	10-15min	15-20min	20-25min	25-30min	30-35min	35-40min	40-45min	45-50min	50-55min	55-60min
Cinereous Tinamou	5	15	69	59	29	46	61	17	22	3	8	1
Little Tinamou	7	34	43	39	25	18	10	17	8	3	0	1
Undulated Tinamou	9	33	39	38	51	44	50	47	58	47	53	43
Plumbeous Pigeon	0	1	3	1	31	15	26	26	18	42	40	67
Ruddy Pigeon	0	1	1	0	8	2	2	22	26	14	0	0
Grey-fronted Dove	0	1	1	13	15	40	79	88	39	13	12	20
Amazonian Pygmy-Owl	51	48	21	4	1	2	0	0	0	0	0	0
Black-tailed Trogon	0	0	0	1	3	2	17	7	11	23	13	18
Collared Trogon	0	0	0	21	20	3	15	20	48	53	47	21
Amazonian Motmot	38	31	52	90	69	54	42	14	15	10	15	0
Mealy Parrot	0	0	0	0	7	9	21	29	23	10	5	11
Fasciated Antshrike	4	5	1	2	8	31	21	49	35	31	3	6
Plain-winged Antshrike	0	0	19	46	39	39	20	15	3	1	0	8
Dusky-throated Antshrike	0	1	0	6	14	10	17	24	30	33	19	13
Spot-winged Antshrike	0	0	20	26	30	10	13	23	13	8	5	2
Pygmy Antwren	0	0	0	0	8	11	10	12	19	17	11	10
Long-winged Antwren	4	19	17	39	27	3	6	3	3	12	0	8
Grey Antbird	0	0	10	37	40	43	60	100	93	69	80	57
Plumbeous Antbird	0	0	0	0	2	2	36	76	75	59	60	43
Goeldi's Antbird	0	2	13	18	20	27	19	23	22	18	17	7
Black-throated Antbird	0	0	1	0	4	4	2	14	28	18	8	0
Black-faced Antthrush	7	29	55	114	118	148	164	194	163	184	190	148
Rufous-fronted Antthrush	0	0	0	0	1	4	17	24	19	15	19	21
Cinnamon-throated Woodcreeper	7	25	39	24	21	4	20	8	13	6	0	1
Long-billed Woodcreeper	0	6	10	25	14	9	12	6	3	6	12	9
Amazonian Barred-Woodcreeper	6	46	62	50	17	14	1	0	1	0	0	2
Striped Woodcreeper	0	6	17	41	36	47	27	21	15	11	0	1
Buff-throated Woodcreeper	25	58	86	104	67	60	52	38	34	16	10	8
Purple-throated Fruitcrow	0	0	0	0	9	8	22	17	18	9	16	9
Screaming Piha	2	3	2	10	24	31	27	30	32	54	57	54
White-winged Becard	0	6	31	54	65	54	33	48	52	44	31	26
Wing-barred Piprites	0	0	1	1	4	2	9	2	7	19	35	29
Yellow-margined Flycatcher	0	1	0	1	0	1	1	3	24	24	20	19
Dusky-capped Flycatcher	0	0	0	0	0	3	13	33	15	17	4	0
Piratic Flycatcher	2	2	7	14	24	13	8	12	15	13	12	8
Dusky-capped Greenlet	0	0	0	0	11	32	28	38	74	99	99	75
Thrush-like Wren	1	5	20	44	73	100	90	122	101	102	114	119
Buff-breasted Wren	0	1	5	13	14	30	65	76	36	38	40	54
Hauxwell's Thrush	18	40	46	46	67	63	79	68	74	82	88	88
Russet-backed Oropendola	0	0	0	1	0	0	8	22	25	48	41	14
Amazonian Grosbeak	8	24	55	58	57	46	52	52	48	30	25	0

399
400 **Figure 5: Species-level diel variation in vocal prevalence within the dawn hour.** (A) Diel
401 variation in vocal prevalence by 5 min interval for species with total vocal prevalence > 75 from
402 all sites and days, and (B) model-generated prediction curves ± 95% confidence intervals of
403 species-specific vocal prevalence within the dawn hour for six species (B).
404

The spatial heterogeneity that we observed in contiguous lowland forest with

superficially similar conditions may have been the result of cryptic habitat diversity. Though

sometimes misrepresented as homogenous, Amazonian habitats are highly heterogeneous (Pires

& Prance 1985, Tuomisto *et al.* 1995), and important elements of their structural diversity

remain poorly understood (Macía 2011, Borges 2013). Interactions between habitat features and

avian community structure in the Amazon are extremely complex and are influenced by

variables including soil type (Borges 2013), riverine sediment concentration (Laranjeiras *et al.*

2021), the condition of surrounding habitat (Latta *et al.* 2011, Laurence *et al.* 2011, Woltmann

2012, Laurence *et al.* 2012, Wolfe *et al.* 2015, Barlow 2016, Menger *et al.* 2017, Hernández-

Palma & Stouffer 2018), the presence of large animals (Redford 1992, Estes *et al.* 2011), and

climate change (Blake & Loiselle 2015, Stouffer *et al.* 2020). As a result, Amazonian habitat

assessments, which generally rely on features discernable by field observers, can be overly

simplistic (Milliken *et al.* 2010), and we avoided incorporating them in our analysis for this

reason. The complexity of Amazonian systems increases the importance of controlling for

temporal variables, which may obscure these effects.

While this study was not designed to serve as a comprehensive census of the avian

community at ITRA, our observations were generally consistent with those documented by

previous census efforts in Madre de Dios's *várzea* forests (Terborgh *et al.* 1990, Martínez *et al.*

2023), namely that most species were rare, with a high proportion of TVP contributed by only a

few species, and that sites were compositionally heterogeneous. The observation that the site in

edge habitat (Site A) had the lowest average species richness per 1-min interval (4.03; 43 total),

followed by Site F (4.41; 42 total), in degraded forest outside of the reserve boundaries, is

consistent with the results of research focused on edge effects in Amazonia, which are known to

pervasively impact Amazonian bird communities (Terborgh *et al.* 1990, Laurence 2004, Barlow

et al. 2006, Broadbent *et al.* 2008, Laurence *et al.* 2011, Moura *et al.* 2016, Luther *et al.* 2020,

Stouffer 2020). This habitat usage pattern differs from observed dynamics in temperate systems,

where avian diversity tends to be higher in edge habitats than forest interiors (Baldi 1996, Lindell

et al. 2007). Most annotations occurred in the frequency range of 0.5–5 kHz, suggesting possible

signal masking from insects, which primarily occur between 4–12 kHz (Hart *et al.* 2015, Metcalf

et al. 2020).

The temporal variation in avian vocal activity levels and observed species richness that

we documented between days was greater and more consistent than we expected, generally

influencing SR and TVP more than differences between sites. Supporting the idea that temporal

effects may overwhelm statistically significant differences between sites, sampling the same site

on different days was more than twice as likely to result in significantly differing measures of SR

and TVP than sampling different sites on the same day, even though most site-to-site

comparisons indicated significant differences in SR or TVP. That temporal variation was

strongest for TVP is relevant for point counts because aural detection probability for a field

observer is a direct function of VP, which reflects the probability that a species vocalizes during

a given 10 s window. Standard duration point counts should have higher rates of false absences

on days with lower TVP, even with static species availability, because of the reduction in time

windows where species vocalize.

We found that TVP was a useful metric for quantifying avian vocal activity, and believe

that it is worthy of further study and use. TVP is a more stable indicator of overall activity levels

than raw call count, total number of annotations, or total annotation length because it is robust to

natural differences in vocalization patterns across species. When annotation boxes are split, if the

gap between vocalizations exceeds a set time interval (5 s in this study), alternative activity

metrics are dependent on the innate vocalization rate of a given species. For example, Hauxwell's

Thrush (HATH) and Black-faced Antthrush (BFAT) were two of the most common species in

this dataset but feature different vocalization styles. While HATHs often take only short breaks

of < 5 s between song bouts in the morning, BFATs sing with gaps between song bouts that are

generally greater than 5 s. A HATH singing throughout the morning may only result in a small

number of different annotation boxes, but a large total annotation length that incorporates the

time in-between song bouts, while a BFAT singing throughout the morning can feature a very

high number of total annotations, but will not include the gaps between song phrases because

they exceed 5 s in length. In our data, BFAT had a total annotation length of 4,456 seconds

across 1,500 individual bounding boxes, while HATH had a total annotation length of 25,308

seconds that constituted just 675 total bounding boxes. Using TVP eliminates this type of

artificial species-to-species variability in vocal activity estimates. As a result, TVP can improve

comparisons of annotated bioacoustic datasets that employ slightly different annotation protocols

and can reduce manual analysis time by eliminating the need to measure the distance between

vocalizations. While metrics that require species-level ID are highly contingent on observer skill

level, driving inconsistencies in Neotropical bird surveys (Robinson *et al.* 2018), TVP supports

using broader taxonomic groups for challenging identifications, potentially lowering time and

experience barriers for researchers. We found that TVP correlated strongly with SR ($r = 0.85$, p

< 0.001), meaning that it may be a viable proxy for species richness, even without requiring

species IDs, and should be easier to generate with automated approaches than metrics that

require species-level IDs. Because TVP essentially represents soundscape abundance, it can also

enable the calculation of abundance-based indices like Bray-Curtis dissimilarity and Shannon

diversity with acoustic data.

Based on our collective results, we suggest that bioacoustic data should be collected to

complement traditional avian surveys whenever possible. In addition to the short-term benefits of

quantifying diel and day-to-day temporal variation, even small bioacoustic datasets serve as

ecological time capsules (Sugai & Llusia 2019), which are particularly important in an era of

rapid global change. ARUs can generate greater data volume than traditional methods without

meaningfully increasing field time or cost (Hobson *et al.* 2002, Acevedo & Villanueva-Rivera

2006, Tegeler *et al.* 2012). While manually analyzing acoustic data can be time-intensive,

limiting its utility, acoustic indices (Jorge *et al.* 2018, Metcalf *et al.* 2020) and automated

identification programs like BirdNET (Kahl *et al.* 2021) are continuously improving and could

have a transformative effect on global ornithological research (Pérez-Granados 2023). Due to

their relative rarity, fully-annotated tropical soundscapes are critical for developing these

automated programs, and this dataset has already been used for this purpose (Kahl *et al.* 2020).

Studies focused exclusively on birds can come at the expense of developing knowledge of other

groups (Gardner *et al.* 2008), but bioacoustic data collected for ornithological research often

contains vocalizations of non-target taxa and other soundscape elements, enabling more holistic

biodiversity research (Newson *et al.* 2017). To help advance these efforts, we encourage

researchers collecting bioacoustic data to make use of open-access repositories like Zenodo

(European Organization For Nuclear Research & OpenAIRE 2013) to host their datasets, as we

have (Hopping *et al.* 2022).

To conclude, we found that simultaneous passive acoustic monitoring coupled with

manual annotation revealed significant diel and day-to-day variation in the vocal activity and

observed species richness of an Amazonian bird community, an ecological pattern that would be

difficult to ascertain using traditional field methods. The magnitude of this temporal variation

was large enough that it could mask meaningful differences between sites if temporal bias is not

sufficiently accounted for in study designs. This research provides a case study for using passive

acoustic monitoring to quantify temporal variation in tropical ornithological surveys, even with a

relatively small sample of sites and days. Future studies could help explain the mechanisms for

this variation and improve methods for processing and interpreting large volumes of acoustic

data in complex systems.

Authors' Contributions

W.A.H. and H.K. designed the methodology and secured the funding to conduct the study;

W.A.H. and C.J.S. analyzed the data; W.A.H. and N.H.C. collected and annotated the recordings,

and W.A.H. led the writing of the manuscript. All authors contributed critically to the drafts and

gave final approval for publication. The Spanish abstract was translated by N.H.C.

Conflict of Interest

The authors declare that there is no conflict of interest.

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Intellectual Contributions

Our study involved collaboration with multiple Peruvian NGOs, particularly the Inkaterra Association (ITA), in both Lima and Madre de Dios. While in the field, we taught local researchers how to set up and maintain Swift recorders, how to annotate recordings in Raven Pro, and then left 9 recorders with them for their own use. We shared a copy of all collected data and our results with ITA, and one of 'ITA's local researchers (Noe Huaraca-Charca) is a coauthor on this paper.

Data Availability Statement

All software curated for this research are archived and available at

https://github.com/csayers2/Inkaterra-ARU. The raw acoustic data and annotations referenced in

this study can be found at https://zenodo.org/record/7079124 (DOI: 10.5281/zenodo.7079124).

References

Acevedo, M. A., & Villanueva-Rivera, L. J. (2006). Using Automated Digital Recording Systems as Effective Tools for the Monitoring of Birds and Amphibians. *Wildlife Society Bulletin*, *34*(1), 211–214. https://doi.org/10.2193/0091-7648(2006)34[211:uadrsa]2.0.co;2

Aide, T. M., Hernández-Serna, A., Campos-Cerqueira, M., Acevedo-Charry, O., & Deichmann, J. L. (2017). Species richness (of insects) drives the use of acoustic space in the tropics. *Remote Sensing*, *9*(11), 1–12. https://doi.org/10.3390/rs9111096

Albert, J.S.; Carnaval, A.C.; Flantua, S.G.A.; Lohmann, L.G.; Ribas, C.C.; Riff, D.; Carrillo, J.D.; Fan, Y.; Figueiredo, J.J.P.; Guayasamin, J.M.; *et al.* (2023). Human impacts outpace natural processes in the Amazon. *Science, 379,* eabo5003. https://doi.org/10.1126/science.abo5003

Alvarez-Berríos, N., Campos-Cerqueira, M., Hernández-Serna, A., Delgado, C. J. A., Román-Dañobeytia, F., & Aide, T. M. (2016). Impacts of small-scale gold mining on birds and anurans near the Tambopata Natural Reserve, Peru, assessed using passive acoustic monitoring. *Tropical Conservation Science*, *9*(2), 832–851. https://doi.org/10.1177/194008291600900216

Allan, J. R., Venter, O., & Watson, J. E. M. (2017). Temporally inter-comparable maps of terrestrial wilderness and the Last of the Wild. *Scientific Data*, *4*, 1–8. https://doi.org/10.1038/sdata.2017.187

Antonelli, A., Zizka, A., Carvalho, F. A., Scharn, R., Bacon, C. D., Silvestro, D., & Condamine, F. L. (2018). Amazonia is the primary source of Neotropical biodiversity. *Proceedings of the National Academy of Sciences of the United States of America*, *115*(23), 6034–6039. https://doi.org/10.1073/pnas.1713819115

Baldi, A. (1996). Edge effects in tropical versus temperate forest bird communities: Three alternative hypotheses for the explanation of differences. *Acta Zoologica Academiae Scientiarum Hungaricae*, *42*(3), 163–172.

Barlow, J., Peres, C. A., Henriques, L. M. P., Stouffer, P. C., & Wunderle, J. M. (2006). The responses of understorey birds to forest fragmentation, logging and wildfires: An Amazonian synthesis. *Biological Conservation*, *128*(2), 182–192. https://doi.org/10.1016/j.biocon.2005.09.028

Barlow, J., França, F., Gardner, T. A., Hicks, C. C., Lennox, G. D., Berenguer, E., … Graham, N. A. J. (2018). The future of hyperdiverse tropical ecosystems. *Nature*, *559*(7715), 517–526. https://doi.org/10.1038/s41586-018-0301-1

Berg, K. S., Brumfield, R. T., & Apanius, V. (2006). Phylogenetic and ecological determinants of the neotropical dawn chorus. *Proceedings of the Royal Society B: Biological Sciences*, *273*(1589), 999–1005. https://doi.org/10.1098/rspb.2005.3410

Blackburn, T. M., & Gaston, K. J. (1998). Some methodological issues in macroecology. *American Naturalist*, *151*(1), 68–83. https://doi.org/10.1086/286103

Blake, J. G. (1992). Temporal Variation in Point Counts of Birds in a Lowland Wet Forest in Costa Rica. *The Condor*, *94*(1), 265–275. https://doi.org/10.2307/1368816

Blake, J. G. (2021). Acoustic monitors and direct observations provide similar but distinct perspectives on bird assemblages in a lowland forest of eastern Ecuador. *PeerJ*, (January). https://doi.org/10.7717/peerj.10565

Blake, J. G., & Loiselle, B. A. (2015). Enigmatic declines in bird numbers in lowland forest of eastern Ecuador may be a consequence of climate change. *PeerJ*, *2015*(8), 1–20. https://doi.org/10.7717/peerj.1177

Borges, S. H. (2013). Bird species distribution in a complex Amazonian landscape: Species diversity, compositional variability and biotic-environmental relationships. *Studies on Neotropical Fauna and Environment*, *48*(2), 106–118. https://doi.org/10.1080/01650521.2013.841627

Boulton, C. A., Lenton, T. M., & Boers, N. (2022). Pronounced loss of Amazon rainforest resilience since the early 2000s. *Nature Climate Change*, *12*(3), 271–278. https://doi.org/10.1038/s41558-022-01287-8

Broadbent, E. N., Asner, G. P., Keller, M., Knapp, D. E., Oliveira, P. J. C., & Silva, J. N. (2008). Forest fragmentation and edge effects from deforestation and selective logging in the Brazilian Amazon. *Biological Conservation*, *141*(7), 1745–1757. https://doi.org/10.1016/j.biocon.2008.04.024

Cerqueira, M. C., Cohn-Haft, M., Vargas, C. F., Nader, C. E., Andretti, C. B., Costa, T. V. V., … Ferraz, G. (2013). Rare or elusive? A test of expert knowledge about rarity of Amazon forest birds. *Diversity and Distributions*, *19*(7), 710–721. https://doi.org/10.1111/ddi.12033

Darras, K., Batáry, P., Furnas, B., Celis-Murillo, A., Van Wilgenburg, S. L., Mulyani, Y. A., &
Tscharntke, T. (2018). Comparing the sampling performance of sound recorders versus
point counts in bird surveys: A meta-analysis. *Journal of Applied Ecology*, *55*(6), 2575–
2586. https://doi.org/10.1111/1365-2664.13229

Darras, K., Batáry, P., Furnas, B. J., Grass, I., Mulyani, Y. A., & Tscharntke, T. (2019).
Autonomous sound recording outperforms human observation for sampling birds: a
systematic map and user guide. *Ecological Applications*, *29*(6).
https://doi.org/10.1002/eap.1954
629

630 Estes, J. A., Terborgh, J., Brashares, J. S., Power, M. E., Berger, J., Bond, W. J., … Wardle, D.
631   A. (2011). Trophic downgrading of planet earth. *Science*, *333*(6040), 301–306.
632   https://doi.org/10.1126/science.1205106
633

634 European Organization For Nuclear Research & OpenAIRE. (2013). *Zenodo*.
635   doi:10.25495/7GXK-RD71
636

637 Gardner, T. A., Barlow, J., Araujo, I. S., Ávila-Pires, T. C., Bonaldo, A. B., Costa, J. E., …
638   Peres, C. A. (2008). The cost-effectiveness of biodiversity surveys in tropical forests.
639   *Ecology Letters*, *11*(2), 139–150. https://doi.org/10.1111/j.1461-0248.2007.01133.x
640

Gil, D., & Llusia, D. (2020). *The Bird Dawn Chorus Revisited*. https://doi.org/10.1007/978-3-
030-39200-0_3

Haddad, N. M., Brudvig, L. A., Clobert, J., Davies, K. F., Gonzalez, A., Holt, R. D., …
Townshend, J. R. (2015). Habitat fragmentation and its lasting impact on 'Earth's
ecosystems. *Science Advances*, *1*(2), 1–10. https://doi.org/10.1126/sciadv.1500052

Hart, P. J., Hall, R., Ray, W., Beck, A., & Zook, J. (2015). Cicadas impact bird communication
in a noisy tropical rainforest. *Behavioral Ecology*, *26*(3), 839–842.
https://doi.org/10.1093/beheco/arv018

Hart, P. J., Ibanez, T., Paxton, K., Tredinnick, G., Sebastián-González, E., & Tanimoto-Johnson,
A. (2021). Timing Is Everything: Acoustic Niche Partitioning in Two Tropical Wet Forest
Bird Communities. *Frontiers in Ecology and Evolution*, *9*(October), 1–10.
https://doi.org/10.3389/fevo.2021.753363

Haselmayer, J., & Quinn, J. S. (2000). A comparison of point counts and sound recording as bird
survey methods in amazonian southeast Peru. *Condor*, *102*(4), 887–893.
https://doi.org/10.2307/1370317
656

657 Haupert, S., Sèbe, F., & Sueur, J. (2022). Physics-based model to predict the acoustic detection
658   distance of terrestrial autonomous recording units over the diel cycle and across seasons:
659   Insights from an Alpine and a Neotropical forest. *Methods in Ecology and Evolution*, *14*(2),
660   614–630. https://doi.org/10.1111/2041-210X.14020
661

662 Hernández-Palma, A., & Stouffer, P. C. (2018). Matrix and area effects on the nutritional
663   condition of understory birds in Amazonian rainforest fragments. *Perspectives in Ecology*
664   *and Conservation*, *16*(3), 139–145. https://doi.org/10.1016/j.pecon.2018.06.003
665

666 Hobson, K. A., Rempel, R. S., Greenwood, H., Turnbull, B., & Van Wilgenburg, S. L. (2002).
667   Acoustic surveys of birds using electronic recordings: New potential from an
668   omnidirectional microphone system. *Wildlife Society Bulletin*, *30*(3), 709–720.
669

670 Hopping, W. A., Kahl, S. & Klinck, H. (2022). A collection of fully-annotated soundscape
671   recordings from the Southwestern Amazon Basin (Version 1) [Data set]. Zenodo.
672   https://doi.org/10.5281/zenodo.7079124
673

674 Hull, J. M., Fish, A. M., Keane, J. J., Mori, S. R., Sacks, B. N., & Hull, A. C. (2010). Estimation
675   of Species Identification Error: Implications for Raptor Migration Counts and Trend
676   Estimation. *Journal of Wildlife Management*, *74*(6), 1326–1334.
677   https://doi.org/10.2193/2009-255
678

Jirinec, V., Elizondo, E. C., Rutt, C. L., & Stouffer, P. C. (2018). Space use, diurnal movement,
and roosting of a Variegated Antpitta (Grallaria varia) in central Amazonia. *Ornitologia*
Neotropical, *29*(February), 13–20. https://doi.org/10.58843/ornneo.v29i1.325

Jorge, F. C., Machado, C. G., da Cunha Nogueira, S. S., & Nogueira-Filho, S. L. G. (2018). The
effectiveness of acoustic indices for forest monitoring in Atlantic rainforest fragments.
Ecological Indicators, *91*(August 2017), 71–76.
https://doi.org/10.1016/j.ecolind.2018.04.001
686

Kahl, S., Clapp, M., Hopping, W. A., Goëau, H., Glotin, H., Planqué, R., Vellinga, W. P., Joly,
A. (2020). Overview of BirdCLEF 2020: Bird sound recognition in complex acoustic
environments. In: CLEF task overview 2020, CLEF: Conference and Labs of the Evaluation
Forum, Sep. 2020, Thessaloniki, Greece.

Kahl, S., Wood, C. M., Eibl, M., & Klinck, H. (2021). BirdNET: A deep learning solution for
avian diversity monitoring. *Ecological Informatics*, *61*(December 2020), 101236.
https://doi.org/10.1016/j.ecoinf.2021.101236

Kruskal, J.B. (1964). Non-metric Multidimensional Scaling: A Numerical Method.
Psychometrika, 29, 115-129. http://dx.doi.org/10.1007/BF02289694
696

697 Lapola D. M., Pinho P., Barlow J., Aragão L. E. O. C., Berenguer E., *et al.*, (2023). The drivers
698   and impacts of Amazon forest degradation. *Science*, 379(6630), eabp8622.
699   http://dx.doi.org/10.1126/science.abp8622
700

701 Laranjeiras, T. O., Naka, L. N., Leite, G. A., Cohn-Haft, M. (2021). Effects of a major
702   Amazonian river confluence on the distribution of floodplain forest avifauna, *Journal of*
703   *Biogeography*, 48, 4, (847-860). https://doi.org/10.1111/jbi.14042
704

705   Latta, S. C., Tinoco, B. A., Astudillo, P. X., & Graham, C. H. (2011). Patterns and magnitude of
706   temporal change in avian communities in the ecuadorian andes. *Condor*, *113*(1), 24–40.
707   https://doi.org/10.1525/cond.2011.090252
708

709   Laurance, S. G. W. (2004). Responses of understory rain forest birds to road edges in central
710   Amazonia. *Ecological Applications*, *14*(5), 1344–1357. https://doi.org/10.1890/03-5194
711

712   Laurance, W. F., Camargo, J. L. C., Luizão, R. C. C., Laurance, S. G., Pimm, S. L., Bruna, E.
713   M., … Lovejoy, T. E. (2011). The fate of Amazonian forest fragments: A 32-year
714   investigation. *Biological Conservation*, *144*(1), 56–67.
715   https://doi.org/10.1016/j.biocon.2010.09.021
716

717   Laurance, W. F., Useche, D. C., Rendeiro, J., Kalka, M., Bradshaw, C. J. A., Sloan, S. P., …
718   Campbell, M. (2012). Averting biodiversity collapse in tropical forests. *Nature*, *489*, 290–
719   294.
720

721   Lees, A. C., Naka, L. N., Aleixo, A., Cohn-Haft, M., Piacentini, V. de Q., Santos, M. P. D., &
722   Silveira, L. F. (2014). Conducting rigorous avian inventories: Amazonian case studies and a
723   roadmap for improvement. *Revista Brasileira de Ornitologia*, *22*(2), 107–120.
724

725   Lindell, C. A., Riffell, S. K., Kaiser, S. A., Battin, A. L., Smith, M. L., & Sisk, T. D. (2007).
726   Edge Responses of Tropical and Temperate Birds. *The Wilson Journal of Ornithology*,
727   *119*(2), 205–220. https://doi.org/10.1676/05-133.1
728

729   Lovejoy, T. E., & Nobre, C. (2019). Amazon tipping point: Last chance for action. *Science*
730   *Advances*, *5*(12), 4–6. https://doi.org/10.1126/sciadv.aba2949
731

732   Luther, D.. (2008). Signaller: Receiver coordination and the timing of communication in
733   Amazonian birds. *Biology Letters*, *4*(6), 651–654. https://doi.org/10.1098/rsbl.2008.0406
734

735   Luther, D. (2009). The influence of the acoustic community on songs of birds in a neotropical
736   rain forest. *Behavioral Ecology*, *20*(4), 864–871. https://doi.org/10.1093/beheco/arp074
737

738   Luther, D. A., Cooper, W. J., Wolfe, J. D., Bierregaard, R. O., Gonzalez, A., & Lovejoy, T. E.
739   (2020). Tropical forest fragmentation and isolation: Is community decay a random process?
740   *Global Ecology and Conservation*, *23*, e01168. https://doi.org/10.1016/j.gecco.2020.e01168
741

Lynch, J.F. 1995. Effects of Point Count Duration, Time-of-Day, and Aural Stimuli on
Detectability of Migratory and Resident Bird Species in Quintana Roo, Mexico. Ralph, C.
John; Sauer, John R.; Droege, Sam, Tech. Ed. 1995. *Monit. bird Popul. by point counts.*
Gen. Tech. Rep. PSW-GTR- 149. Albany, CA U.S. Dep. Agric. For. Serv. Pacific
Southwest Res. Stn. p. 1-6 149.

Macía, M. J. (2011). Spatial distribution and floristic composition of trees and lianas in different
forest types of an Amazonian rainforest. *Plant Ecology*, *212*(No. 7 (July 2011)), 1159–1177.
https://doi.org/10.1007/s11258-011-9895-1

Martínez, A.E., Ponciano, J.M., Gomez, J.P., Valqui, T., Novoa, J. & Antezana, M. et al. (2023).
The structure and organisation of an Amazonian bird community remains little changed
after nearly four decades in Manu National Park. *Ecology Letters*, 26, 335–346.
https://doi.org/10.1111/ele.14159

Menger, J., Magnusson, W. E., Anderson, M. J., Schlegel, M., Pe'er, G., & Henle, K. (2017).
Environmental characteristics drive variation in Amazonian understorey bird assemblages.
PLoS ONE, *12*(2), 1–20. https://doi.org/10.1371/journal.pone.0171540

Metcalf, O. C., Barlow, J., Devenish, C., Marsden, S., Berenguer, E., & Lees, A. C. (2020).
Acoustic indices perform better when applied at ecologically meaningful time and
frequency scales. *Methods in Ecology and Evolution*, *2020*(October), 1–11.
https://doi.org/10.1111/2041-210x.13521

Metcalf, O. C., Barlow, J., Marsden, S., Moura, N. G. De, Berenguer, E., Ferreira, J., & Lees, A.
C. (2021). Optimizing tropical forest bird surveys using passive acoustic monitoring and
high temporal resolution sampling. *Remote Sensing in Ecology and Conservation*, 1–12.
https://doi.org/10.1002/rse2.227

Milliken, W., Zappi, D., Sasaki, D., Hopkins, M., & Pennington, R. T. (2010). Amazon
vegetation: How much don't we know and how much does it matter? *Kew Bulletin*, *65*(4),
691–709. https://doi.org/10.1007/s12225-010-9236-x

Moura, N. G., Lees, A. C., Aleixo, A., Barlow, J., Berenguer, E., Ferreira, J., … Gardner, T. A.
(2016). Idiosyncratic responses of Amazonian birds to primary forest disturbance.
Oecologia, *180*(3), 903–916. https://doi.org/10.1007/s00442-015-3495-z

Newson, S. E., Bas, Y., Murray, A., & Gillings, S. (2017). Potential for coupling the monitoring
of bush-crickets with established large-scale acoustic monitoring of bats. *Methods in
Ecology and Evolution*, *8*(9), 1051–1062. https://doi.org/10.1111/2041-210X.12720

Oksanen, J., Guillaume Blanchet, F., Friendly, M., Kindt, R., Legendre, P., McGlinn, D.,
Minchin, P. R., O'hara, R. B., Simpson, G. L., Solymos, P., Stevens, M. H. H., Szoecs, E.,
& Wagner, H. (2019). vegan: Community ecology package. R package version 2.5-6.
Retrieved from https:// CRAN.R-project.org/package=vegan

Parker, T. (1991). On the Use of Tape Recorders in Avifaunal Surveys. *The Auk: Ornithological
Advances*, *108*(02), 443–444. https://doi.org/10.1093/auk/108.2.443

Pérez-Granados, C. (2023). BirdNET: applications, performance, pitfalls and future
opportunities. *Ibis*. https://doi.org/10.1111/ibi.13193

Pérez-Granados, C., & Schuchmann, K. L. (2022). Diel and seasonal variation of Striped Cuckoo
(Tapera naevia) vocalizations revealed using automated signal recognition. *Ibis*, *165*(1),
179–189. https://doi.org/10.1111/ibi.13129

Pillay, R., Fletcher, R. J., Sieving, K. E., Udell, B. J., & Bernard, H. (2019). Bioacoustic
monitoring reveals shifts in breeding songbird populations and singing behaviour with
selective logging in tropical forests. *Journal of Applied Ecology*, *56*(11), 2482–2492.
https://doi.org/10.1111/1365-2664.13492

Pires, J.M., Prance, G. 1985. The vegetation types of the Brazilian Amazon. In: Prance GT,
Lovejoy TE, editors. Key Environment – Amazonia. New York (NY): Pergamon Press. p.
109–145.

Planqué, R., & Slabbekoorn, H. (2008). Spectral Overlap in Songs and Temporal Avoidance in a
Peruvian Bird Assemblage. *Ethology*, *114*(3), 262–271. https://doi.org/10.1111/j.1439-
0310.2007.01461.x

R Core Team. 2022. *R: a language and environment for statistical computing*. R Foundation for
Statistical Computing, Vienna. https://www.R-project.org/

Ralph, C. J., Sauer, J. R., & Droege, S. (1995). Monitoring Bird Populations by Point Counts.
The Journal of Wildlife Management, *61*(4), 1453. https://doi.org/10.2307/3802161

Redford, K. H. (1992). The Empty Forest. *BioScience*, *42*(No. 6 (Jun., 1992)), 412–422.
https://doi.org/https://doi.org/10.2307/1311860

Rempel, R. S., Hobson, K. A., Holborn, G., Van Wilgenburg, S. L., & Elliott, J. (2005).
Bioacoustic monitoring of forest songbirds: interpreter variability and effects of
configuration and digital processing methods in the laboratory. *Journal of Field
Ornithology*, *76*(1), 1–11. https://doi.org/10.1648/0273-8570-76.1.1

Remsen, J. V. (1994). Use and Misuse of Bird Lists in Community Ecology and Conservation.
The Auk, *111*(1), 225–227. https://doi.org/10.2307/4088531

Ricotta, C., & Podani, J. (2017). On some properties of the Bray-Curtis dissimilarity and their
ecological meaning. *Ecological Complexity, 31,* 201-205.
https://doi.org/10.1016/j.ecocom.2017.07.003.

Robinson, W. D., Lees, A. C., & Blake, J. G. (2018). Surveying tropical birds is much harder
than you think: a primer of best practices. *Biotropica*, *50*(6), 846–849.
https://doi.org/10.1111/btp.12608

Rodriguez, A., Gasc, A., Pavoine, S., Grandcolas, P., Gaucher, P., & Sueur, J. (2014). Temporal
and spatial variability of animal sound within a neotropical forest. *Ecological Informatics*,
21, 133–143. https://doi.org/10.1016/j.ecoinf.2013.12.006

Shonfield, J., & Bayne, E. M. (2017). Autonomous recording units in avian ecological research:
current use and future applications. *Avian Conservation and Ecology*, *12*(1).
https://doi.org/10.5751/ace-00974-120114

Simons, T. R., Alldredge, M. W., Pollock, K. H., & Wettroth, J. M. (2007). Experimental
analysis of the auditory detection process on avian point counts. *Auk*, *124*(3), 986–999.
https://doi.org/10.1642/0004-8038(2007)124[986:EAOTAD]2.0.CO;2

Soares, L., Cockle, K. L., Inzunza, E. R., Ibarra, J. T., Miño, C. I., Zuluaga, S., … Echeverry-
galvis, M. A. (2023). Neotropical ornithology: Reckoning with historical assumptions ,
removing systemic barriers , and reimagining the future. *Ornithological Applications*,
125(1), 1–31. http://dx.doi.org/10.32942/osf.io/yu2fx

Simons, T. R., Alldredge, M. W., Pollock, K. H., & Wettroth, J. M. (2007). Experimental
analysis of the auditory detection process on avian point counts. *Auk*, *124*(3), 986–999.
https://doi.org/10.1642/0004-8038(2007)124[986:EAOTAD]2.0.CO;2

Stouffer, P. C., Jirinec, V., Rutt, C. L., Bierregaard, R. O., Hernández-Palma, A., Johnson, E. I.,
… Lovejoy, T. E. (2020). Long-term change in the avifauna of undisturbed Amazonian
rainforest: ground-foraging birds disappear and the baseline shifts. *Ecology Letters*.
https://doi.org/10.1111/ele.13628
854
Stutchbury, B. J. M., & Morton, E. S. (2008). Recent advances in the behavioral ecology of
tropical birds. *Wilson Journal of Ornithology*, *120*(1), 26–37. https://doi.org/10.1676/07-
018.1

Sugai, L. S. M., & Llusia, D. (2019). Bioacoustic time capsules: Using acoustic monitoring to
document biodiversity. *Ecological Indicators*, *99*(December 2018), 149–152.
https://doi.org/10.1016/j.ecolind.2018.12.021
861
862 Tegeler, A. K., Morrison, M. L., & Szewczak, J. M. (2012). Using extended-duration audio
863   recordings to survey avian species. *Wildlife Society Bulletin*, *36*(1), 21–29.
864   https://doi.org/10.1002/wsb.112
865
866 Terborgh, J., Robinson, S. K., Parker III, T. A., & Munn, C. A. (1990). Structure and
867   Organization of an Amazonian Forest Bird Community. *Ecological Monographs*, *60*(2),
868   213–238.
869
870 Tuomisto, H., Ruokolainen, K., Kalliola, R., Linna, A., Danjoy, W., & Rodriguez, Z. (1995).
871   Dissecting amazonian biodiversity. *Science*, *269*(5220), 63–66.
872   https://doi.org/10.1126/science.269.5220.63
873
874 Verner, J. & Ritter, L.V. 1986. Hourly variation in morning point counts of birds. *The Auk,* 103,
875   117–124. https://doi.org/10.1093/AUK/103.1.117.
876
877 Wickham, H., Averick, M., Bryan, J., et al. 2019. Welcome to the tidyverse. *J. Open Source*
878   *Softw*. 4:1686. https://doi.org/10.21105/joss.01686
879
Wimmer, J., Towsey, M., Roe, P., & Williamson, I. (2013). Sampling environmental acoustic
recordings to determine bird species richness. *Ecological Applications*, *23*(6), 1419–1428.

Wolfe, J. D., Stouffer, P. C., Mokross, K., Powell, L. L., & Anciaes, M. M. (2015). Island vs.
countryside biogeography: An examination of how Amazonian birds respond to forest
clearing and fragmentation. *Ecosphere*, *6*(12), 1–14. https://doi.org/10.1890/ES15-00322.1

Woltmann, S. (2005). Patterns of daily temporal variation in detectability of forest birds in
Bolivia. *Ornitologia Neotropical*, *16*(3), 337–346.

Woltmann, S., Kreiser, B. R., & Sherry, T. W. (2012). Fine-scale genetic population structure of
an understory rainforest bird in Costa Rica. *Conservation Genetics*, *13*(4), 925–935.
https://doi.org/10.1007/s10592-012-0341-2

Wood, S.N. (2004). Stable and Efficient Multiple Smoothing Parameter Estimation for
Generalized Additive Models. *J. Am. Stat. Assoc.* 99: 673–686.

Wood, S.N. (2011). Fast stable restricted maximum likelihood and marginal likelihood
estimation of semiparametric generalized linear models. *J. R. Stat. Soc. Ser. B Stat.*
Methodol. 73: 3–36.

Wood, S.N. (2017). *Generalized Additive Models: An Introduction with R (2nd edition)*.
Chapman and Hall/CRC.

Wood, S.N., Pya, N. & Säfken, B. (2016). Smoothing Parameter and Model Selection for
General Smooth Models. *J. Am. Stat. Assoc.* 111: 1548–1563.

Venier, L. A., Holmes, S. B., Holborn, G. W., McIlwrick, K. A., & Brown, G. (2012).
Evaluation of an automated recording device for monitoring forest birds. *Wildlife Society*
Bulletin, *36*(1), 30–39. https://doi.org/10.1002/wsb.88

Verner, J., & Ritter, L. V. (1986). Hourly Variation in Morning Point Counts of Birds. *The Auk*,
(January), 117–124.

Xu, X., Jia, G., Zhang, X., Riley, W. J., & Xue, Y. (2020). Climate regime shift and forest loss
amplify fire in Amazonian forests. *Global Change Biology*, *26*(10), 5874–5885.
https://doi.org/10.1111/gcb.15279

Yang, L., Qin, G., Zhao, N., Wang, C. & Song, G. (2012). Using a generalized additive model
with autoregressive terms to study the effects of daily temperature on mortality. *BMC Med.*
Res. Methodol. 12: 165.

A)

Species	0-5min	5-10min	10-15min	15-20min	20-25min	25-30min	30-35min	35-40min	40-45min	45-50min	50-55min	55-60min
Cinerous Tinamou	5	15	69	59	29	46	61	17	22	3	8	1
Little Tinamou	7	34	43	39	25	18	10	17	8	3	0	1
Undulated Tinamou	9	33	39	38	51	44	50	47	58	47	53	43
Plumbeous Pigeon	0	1	3	1	31	15	26	26	18	42	40	67
Ruddy Pigeon	0	1	1	0	8	2	2	22	26	14	0	0
Grey-fronted Dove	0	1	1	13	15	40	79	88	39	13	12	20
Amazonian Pygmy-Owl	51	48	21	4	1	2	0	0	0	0	0	0
Black-tailed Trogon	0	0	0	1	3	2	17	7	11	23	13	18
Collared Trogon	0	0	0	21	20	3	15	20	48	53	47	21
Amazonian Motmot	38	31	52	90	69	54	42	14	15	10	15	0
Mealy Parrot	0	0	0	0	7	9	21	29	23	10	5	11
Fasciated Antshrike	4	5	1	2	8	31	21	49	35	31	3	6
Plain-winged Antshrike	0	0	19	46	39	39	20	15	3	1	0	8
Dusky-throated Antshrike	0	1	0	6	14	10	17	24	30	33	19	13
Spot-winged Antshrike	0	0	20	26	30	10	13	23	13	8	5	2
Pygmy Antwren	0	0	0	0	8	11	10	12	19	17	11	10
Long-winged Antwren	4	19	17	39	27	3	6	3	3	12	0	8
Grey Antbird	0	0	10	37	40	43	60	100	93	69	80	57
Plumbeous Antbird	0	0	0	0	2	2	36	76	75	59	60	43
Goeldi's Antbird	0	2	13	18	20	27	19	23	22	18	17	7
Black-throated Antbird	0	0	1	0	4	4	2	14	28	18	8	0
Black-faced Antthrush	7	29	55	114	118	148	164	194	163	184	190	148
Rufous-fronted Antthrush	0	0	0	0	1	4	17	24	19	15	19	21
Cinnamon-throated Woodcreeper	7	25	39	24	21	4	20	8	13	6	0	1
Long-billed Woodcreeper	0	6	10	25	14	9	12	6	3	6	12	9
Amazonian Barred-Woodcreeper	6	46	62	50	17	14	1	0	1	0	0	2
Striped Woodcreeper	0	6	17	41	36	47	27	21	15	11	0	1
Buff-throated Woodcreeper	25	58	86	104	67	60	52	38	34	16	10	8
Purple-throated Fruitcrow	0	0	0	0	9	8	22	17	18	9	16	9
Screaming Piha	2	3	2	10	24	31	27	30	32	54	57	54
White-winged Becard	0	6	31	54	65	54	33	48	52	44	31	26
Wing-barred Piprites	0	0	1	1	4	2	9	2	7	19	35	29
Yellow-margined Flycatcher	0	1	0	1	0	1	1	3	24	24	20	19
Dusky-capped Flycatcher	0	0	0	0	0	3	13	33	15	17	4	0
Piratic Flycatcher	2	2	7	14	24	13	8	12	15	13	12	8
Dusky-capped Greenlet	0	0	0	0	11	32	28	38	74	99	99	75
Thrush-like Wren	1	5	20	44	73	100	90	122	101	102	114	119
Buff-breasted Wren	0	1	5	13	14	30	65	76	36	38	40	54
Hauxwell's Thrush	18	40	46	46	67	63	79	68	74	82	88	88
Russet-backed Oropendola	0	0	0	1	0	0	8	22	25	48	41	14
Amazonian Grosbeak	8	24	55	58	57	46	52	52	48	30	25	0

B)