Human and wildlife use of mountain glacier habitat in western North America

Scott Hotaling[1], Jordan Boersma[1], Neil A. Paprocki[2], Alissa Anderson[3], Logan Whiles[3], Lucy

Ogburn[4], Sophia Kasper[4], Catharine White[5], Daniel H. Thornton[3], and Peter Wimberger[4,6]

Affiliations:

[1] School of Biological Sciences, Washington State University, Pullman, WA, USA; ORCID

#0000-0002-5965-0986

[2] Idaho Cooperative Fish and Wildlife Research Unit, College of Natural Resources, University

of Idaho, Moscow, ID, USA

[3] School of the Environment, Washington State University, Pullman, WA, USA

[4] Biology Department, University of Puget Sound, Tacoma, WA, USA

[5] Coast Mountain College, Terrace, BC, Canada

[6] Slater Museum of Natural History, Tacoma, WA, USA

Correspondence: Scott Hotaling, School of Biological Sciences, Washington State University,

Pullman, WA, USA; Email: scott.hotaling1@gmail.com; Phone: (828) 507-9950

Running head: Human and wildlife usage of a glacier

ABSTRACT

The global recession of glaciers and perennial snowfields is reshaping mountain ecosystems.

However, beyond physical changes to the landscape and altered downstream hydrology, the

implications of glacier decline are poorly known. Before predictions can be made about how

climate change may affect wildlife in glacier-associated ecosystems, a more thorough

accounting of the role that glaciers play in species' life histories is needed. In this study, we

deployed an elevational transect of wildlife cameras along the western margin of the Paradise

Glacier, a rapidly receding mountain glacier on the south side of Mount Rainier, WA, USA. From

June to September 2021, we detected at least 16 vertebrate species (seven birds, nine

mammals) using glacier-associated habitats. While humans, and primarily skiers, were the most

common species detected, we recorded 99 observations of wildlife (birds and mammals). These

included three species of conservation concern in Washington: wolverine (*Gulo gulo*), Cascade

red fox (*Vulpes vulpes cascadensis*), and White-tailed ptarmigan (*Lagopus leucura*).

Collectively, our results reveal a rich diversity of wildlife using a single mountain glacier and

adjacent habitat in the Pacific Northwest, emphasizing a largely overlooked risk of climate

change to mountain biodiversity. We highlight the global need for similar studies to better

understand the true scale of biodiversity that will be impacted by glacier recession in mountain

ecosystems.

INTRODUCTION

As climate change proceeds and mountain glaciers are lost, there is a pressing need to

understand how the loss of glacier ice will impact habitats and ecosystems (Hotaling et al. 2017,

Milner et al. 2017). Well-known impacts of glacier recession include the loss of key meltwater

resources which transport nutrients to downstream habitats (Ren et al. 2019), underpin

hydropower infrastructure (Finger et al. 2012), and provide key habitats for specialized aquatic

communities (Giersch et al. 2017, Muhlfeld et al. 2020). Often overlooked in this discussion,

however, is the impact of glacier loss on the biodiversity living on and adjacent to glaciers

(Stibal et al. 2020, Schano et al. 2021).

The biodiversity of glacier ecosystems has gained appreciation in recent decades. Long

considered devoid of life, a growing body of research has shown that glaciers host a thriving,

multi-kingdom community (Anesio et al. 2017, Shain et al. 2021), particularly in the uppermost

layer of snow and ice where intense solar radiation, atmospheric deposition, and available water

during melt make make for conditions that are more hospitable to life (Anesio & Laybourn-Parry

2012, Hotaling et al. 2021). Still, however, most studies have focused on the most dominant life

form on glaciers--microbes--leaving larger-bodied residents of glaciers and surrounding habitats

overlooked.

Globally, 36 species of birds and mammals have been observed using glacier and perennial

snow habitat for some portion of their life history (Rosvold 2016). Mammals typically use these

habitats for relief from lower elevation conditions during warmer parts of the year, whereas birds

commonly forage on arthropods and seeds that have been atmospherically deposited on the ice

(Antor 1995, Rosvold 2016, Hotaling et al. 2021). However, closer links between wildlife and

glacier habitats have also been described. In western North America, the high abundance of

glacier ice worms (*Mesenchytraeus solifugus*)—which are commonly present at densities >200

per m^2 on glaciers during summer—may be a key resource subsidy for nesting birds (e.g., Gray-

crowned Rosy Finches, Hotaling et al. 2020). On glaciers of the high Andes in South America,

the White-winged Diuca Finch nests directly within crevasses (Hardy et al. 2018). Throughout

higher latitudes of the Northern Hemisphere, wolverines cache foods in snow and ice, and the

persistence of spring snow cover and cache longevity has been hypothesized as a primary

reason underpinning their cold, snowy distributions (the "refrigeration-zone" hypothesis, Inman

et al. 2012).

While links between larger-bodied vertebrates and icy habitats have been made for a range of

species, these observations typically stem from species-specific studies and likely

underestimate wildlife use of the cryosphere (the global collection of Earth's frozen waters). In

this study, we sought a more holistic understanding of human and wildlife usage of an alpine

glacier and adjacent habitat. We deployed a camera-trapping array along an elevational

transect of a mountain glacier on Mount Rainier, Washington, USA. In line with escalating

environmental harshness at higher elevations, we expected to observe an inverse relationship

between vertebrate use and elevation along our transect. However, we also expected to

observe species not previously associated with glacier or snow ecosystems, highlighting the

likely underestimation of glacier-linked vertebrate biodiversity globally and the pressing need to

better understand how climate change may impact enigmatic mountain species.

84   **MATERIALS AND METHODS**

85   Study area

Mount Rainier is a large stratovolcano in south-central Washington, USA (Figure 1a), and the

centerpiece of Mount Rainier National Park (MORA). At 4,392 m (14,411 feet) tall, Mount

Rainier is the highest peak in Washington state and is flanked by 29 named glaciers covering

~78 km^2 (Beason 2017). The Paradise Glacier lies on the southern side of the peak—with lower

and upper extents at 2,100 and 2,585 m, respectively—and is one of the smaller glaciers in

MORA at ~0.6 km^2 (Beason 2017). Like most glaciers around the world, Mount Rainier's

glaciers are in rapid retreat. From 1896 to 2015, glaciated area on Mount Rainier has decreased

by ~40%, a loss of 50.6 km^2 (Beason 2017). The Paradise Glacier, which was part of a larger

Paradise-Stevens Glaciers complex in 1896, has retreated by 87% over the same period

(Beason 2017). Still, Mount Rainier and its glaciers remain some of the snowiest places on

Earth. Annual snowfall at 1,646 m—as measured near the park visitor center at Paradise

(Figure 1)—has averaged 1,620 cm (639 inches) from 1920 - 2020 (National Park Service

2021a). On Mount Rainier, treeline occurs from roughly 1,829 - 1,981 m (6,000 - 6,500 feet).

100   Experimental design

From 17 June 2021 to 31 October 2021, we deployed wildlife cameras (Browning Strike Force

850 Extreme, Browning, USA) along the western margin of the Paradise Glacier (Figures 1b,c).

However, we only analyzed data through 25 September 2021 because variable but often

persistent snowpack either buried or obscured cameras during the final five weeks of the field

season (25 September - 31 October 2021). We placed six cameras along an elevational

transect from 2,164 to 2,469 m (7,100 - 8,100 feet). We re-visited the cameras every 3-5 weeks.

We focused our cameras to target the margins of the glacier where talus and boulders give way

to ice. Because the Paradise Glacier is such a dynamic environment during summer, with a

dramatically shifting physical surface (both in area and elevation), we moved the cameras as

needed to maintain close proximity (generally 0-10 m) to the ice surface. Due to the variable

terrain and goals of the study, we made no attempts to control for the amount of area each

camera could "see" which likely affected the comparability of our results across cameras.

To account for the need to move the cameras throughout the melt season and to minimize their

visibility in the Mount Rainier wilderness, we employed a unique camera trap design (Figures

2a,b). Cameras were mounted to 5-gallon (18.9 liter) cylindrical Igloo coolers that were covered

in camouflage duct tape with small, camouflage ratchet straps. The coolers were filled with

snow and loose rocks from the trapping area and sealed with ⅛" coated steel cable and a small

U-clamp. Because we expected visitors to notice our cameras despite our efforts to blend them

into the landscape, we added a small label to the top of each trap noting its purpose with

contact information. When triggered, cameras took bursts of five photos in quick succession

with a 30-second rest period between triggers. While our camera traps were largely reliable and

stayed in working order throughout the study, we did experience a memory card failure at C5

from 20 August - 25 September.

126   Animal identification and data analysis

Images were visually inspected for any evidence of vertebrate detection (see example images in

Figure 3). When humans or wildlife (birds and mammals) were detected, they were identified to

the lowest taxonomic level possible by consulting guides for Mount Rainier National Park and

the Pacific Northwest as well as our own research team's expertise in mammals (A.A. and L.W.,

specifically) and birds (J.B., N.A.P., and P.W., specifically).Given the proximity of our cameras

to glacier ice or perennial snow, we considered any species recorded in this study to have a link

to glacierized habitats. This definition is slightly expanded relative to Rosvold (2016), the most

similar study to ours, which focused exclusively on birds and mammals that were directly on

snow and ice.

When multiple images from a camera burst included the same organism(s), it was only counted

once. The same is true for multiple bursts within minutes of each other. However, because our

camera array was arranged along an elevational transect, it is possible, and perhaps likely, that

some humans or wildlife were detected multiple times across different cameras. We did not

attempt to control for multiple identifications in space nor time beyond the methods described

above. Photos of the research team were also excluded from all analyses. Observations were

totaled and analyzed for each of the 14 weeks of the study. We note that the final "week" (week

14) contained two extra days due to where the study endpoint fell (25 September) relative to the

end of that particular week (23 September).

To test for a relationship between human and wildlife occurrences in our data set, we performed

a series of statistical analyses in R (R Core Team 2021). First, we grouped all human and

wildlife observations separately for each camera on a week-by-week basis. Then, we tested for

normality of the human and wildlife data sets using a Shapiro-Wilk normality test ("shapiro.test").

Because both data sets were significantly different from normal (P, Shapiro-Wilk < 0.001), we

performed correlation analyses with a Spearman's rank correlation. We tested for a correlation

between the number of human versus wildlife detections across all cameras and weeks in the

study.

RESULTS

Across 307 detections in the Mount Rainier alpine, we identified at least 16 vertebrate species

(seven birds, nine mammals) using glacier-associated habitats (Table 1). Humans were the

most frequently observed taxon ($N = 208$), followed by American pipits ($N = 18$), mountain goats

($N = 13$), marmot ($N = 13$), and wolverines ($N = 8$; Table 1, Figure 4). Just under one-fifth of

birds could not be identified to species but for those that were identified, American pipits and

White-tailed ptarmigan accounted for more than half of the observations. However, a number of

nondescript brown and gray "blurs" were also recorded, of which many were likely small birds,

and thus our results for birds are likely underestimations.

We observed the most humans at the third-highest camera, C3 ($N = 59$), and the most wildlife at

the lowest camera, C6 ($N = 36$; Figure 4). Interestingly, C6 also recorded the fewest humans (N

= 8) and the most taxonomically diverse wildlife: five mammals (Golden-mantled ground

squirrel, Hoary marmot, human, mountain goat, wolverine), three identifiable birds (American

pipit, Gray-crowned rosy finch, White-tailed ptarmigan), and several unidentifiable passerines

and other birds. Because we did not control for the amount of area each camera could "see,"

comparisons across cameras (and detection periods since cameras were adjusted throughout

the study), should be interpreted with caution.

We observed the highest number of detections during the first week of the study (18-24 June

2021) but this pattern was overwhelmingly driven by humans (Figure 4). Indeed, ~54% of all

human observations were within this period. For wildlife (mammals and birds), no clear temporal

pattern was present. At cameras where humans and wildlife were regularly observed (e.g., C4,

C5), we observed a temporal separation with humans being seen early in the season from mid-

June to mid-July and other taxa appearing later (after mid-July; Figure 4). However, we did not

observe a correlation between the presence of humans and wildlife across all cameras and

weeks of the study (P, Spearman's rho = 0.106).

During the study, we observed three taxa of conservation concern in Washington state:

wolverines (including two kits playing on the glacier), a Cascade red fox, and White-tailed

ptarmigan (Figures 3e-g). We observed wolverines at the three lowest cameras (C4-C6) in late

July/early August, a Cascade red fox at our C4 camera in mid-August, and White-tailed

ptarmigans at the two highest cameras (C1-C2) and the lowest camera (C6), including a pair of

ptarmigans together in mid-July at C2.

191   **DISCUSSION**

Mountain glaciers are in rapid decline around the world (Hugonnet et al. 2021). Concurrent with

this recession of glacier ice is the loss of critical habitat for glacier-associated biodiversity (Stibal

et al. 2021). To date, most of the research emphasis in glacier biology has centered on

understanding the distribution and functional roles of microbial life (e.g., Hotaling et al. 2021).

This is not surprising since microbes are the dominant life form in glacier environments and play

critical roles in the cryosphere, even altering the albedo of ice and impacting its rate of melt

(Ganey et al. 2017, Hotaling et al. 2021). However, overlooked in this emphasis have been

larger vertebrates that use glaciers and adjacent habitats for travel, nesting, foraging, and other

key aspects of their life histories (Rosvold 2016).

In this study, we used a camera trap array deployed along the margin of a rapidly receding

mountain glacier to gain a better understanding of human and wildlife use of these habitats. We

identified an array of birds and mammals, from songbirds to mesocarnivores, highlighting the

ecological complexity and potential importance of glaciers and alpine ecosystems to mountain

biodiversity. Indeed, many species observed in this study (e.g., American kestrel) have not been

previously linked to glacier habitats, raising questions about the degree to which components of

their life history have been overlooked. We also found limited, albeit non-significant, evidence

for interactions between human and wildlife usage of glacier ecosystems with little to no overlap

between the two groups in space and time.

From a conservation and management perspective, we observed three wildlife species that are

particularly notable: wolverine (*Gulo gulo*), Cascade red fox (*Vulpes vulpes cascadensis*), and

White-tailed ptarmigan (*Lagopus leucura*). All three are listed as "Species of Greatest

Conservation Need" under the Washington State Wildlife Action Plan (Washington Department

of Fish and Wildlife, 2021), with a need for monitoring and distribution information specifically

highlighted for wolverine and Cascade red fox. The successful identification of all three species

during the course of this study shows the value of wildlife camera trapping above treeline.

Furthermore, wolverines are exceptionally rare in the South Cascades with the family (adult

female, two kits) observed in this study likely representing just the third family in the range in a

century (National Park Service 2021b). Given their rarity in MORA and the South Cascades, we

assume that all of our observations were of the same wolverines, which were seen on three

separate cameras during the week of 30 July 2021. This burst of activity might indicate

successful foraging in the area as wolverines are known to use rendezvous sites near food

caches, particularly when kits are young (Inman et al. 2012). Notably, we only observed

mesocarnivores (e.g., wolverines) where we also observed common prey species (e.g.,

marmots; Table 1).

While our camera trap design and array in the Mount Rainier alpine was successful, it was not

without difficulties. Indeed, the challenging nature of camera trapping in a high-alpine landscape

should not be overlooked. In addition to highly variable, and often difficult, weather conditions,

the landscape does not lend itself well to traditional camera trapping approaches for three

reasons. (1) The physical landscape dramatically changes during summer due to snowmelt.

During 2021, seasonal snow depth at a substantially lower elevation than our study area peaked

in early summer at 5.33 m (210 inches). This seasonal snow completely melted by mid-July,

effectively lowering the habitat surface by several meters during the first month of our study

(mid-June to mid-July). Thus, camera trap locations were limited to areas that were snow-free

and had to be moved small distances throughout the season to track shifting ice margins.

Changing snow depth also meant that framing the view, and controlling for the field of vision

across cameras, was nearly impossible. For instance, one view that was completely snow-

covered in early July was entirely rock and talus within a matter of days. The end of the season

offers another complication as seasonal snow can accumulate rapidly, particularly in "sheltered"

areas that are ideal for obscuring camera traps. Thus, we recommend that similar future studies

in the PNW or comparable areas conclude by the end of September or that researchers devise

a strategy for finding snow-covered cameras and/or keeping them free of snow. (2) Hiking

and/or game trails are minimal or non-existent above treeline. Thus, an exceptionally large and

consistent survey area exists with little to no mechanism for targeting areas where wildlife may

be more abundant. To this end, a denser camera array would have certainly yielded better

quantifications of vertebrate presence and interactions (e.g., links between carnivores and prey

or humans and wildlife). Finally, (3) we did not observe many songbirds on the glacier and in

surrounding habitats despite well-known connections between the two (e.g., Hotaling et al.

2020). This was likely due to a combination of our cameras not being sensitive enough to detect

them and images with potential birds in-flight being too blurry for identification. Indeed, a

number of photographs included unidentifiable brown or gray blurs that we could not confidently

categorize as a bird or something else (e.g., a flying insect). One way to overcome this, or at

least improve the detection rate of small birds, would be to pair cameras with acoustic recorders

at each site so song could be used in combination with (or in lieu of) imagery.

259   **CONCLUSION**

In this study, we showed that a rich community of birds and mammals use glacier habitat in

western North America. Given the relatively small number of camera traps placed in the vast

alpine habitat of Mount Rainier with no game-trails and few natural chokepoints to increase

success, it is likely that our results are a significant underestimation of the true scale of

biodiversity in these habitats. For instance, in a global meta-analysis, Rosvold (2016) identified

36 species of birds and mammals that are known to use snow and ice habitats for some portion

of their life history. Here, we observed 16 species on a small glacier during a single summer

with a limited camera trap array. Thus, our current understanding clearly underestimates the

true scale of wildlife associations with mountain glaciers. As glaciers continue to rapidly recede

worldwide, there is a growing need to document and understand these wildlife-ice connections

to properly account for the biodiversity impacts of cryosphere decline.

ACKNOWLEDGEMENTS

This project was funded by grants from the Northwest Ecological Research Institute and

Western North American Naturalist. S.H. was also supported by NSF award #OPP-1906015.

This research was conducted under Mount Rainier National Park permit #MORA-2021-SCI-

0016. We thank Ben Lee for assistance in the field and park staff for their support during the

permitting process. We also thank David Sousa and the co-owners of Tatoosh Meadows for

allowing us to camp on their property during this research. We gratefully acknowledge that our

research was conducted on the traditional lands of the Cowlitz, Muckleshoot, Nisqually,

Puyallup, Squaxin Island, and Yakama tribes.

LITERATURE CITED

Anesio, A. M. & J. Laybourn-Parry. 2012. Glaciers and ice sheets as a biome. Trends in

Ecology & Evolution 27:219-225.

Anesio, A. M., S. Lutz, N. A. Chrismas & L.G. Benning. 2017. The microbiome of glaciers and

ice sheets. npj Biofilms and Microbiomes 3:1-11.

Antor, R. J. 1995 The importance of arthropod fallout on snow patches for the foraging of high-

alpine birds. Journal of Avian Biology 26:81-85.

Beason, S. R. 2017. Change in glacial extent at Mount Rainier National Park from 1896–2015.

Natural Resource Report NPS/MORA/NRR-2017/1472, National Park Service.

Finger, D., G. Heinrich, A. Gobiet & A. Bauder. 2012. Projections of future water resources and

their uncertainty in a glacierized catchment in the Swiss Alps and the subsequent effects

on hydropower production during the 21st century. Water Resources Research 48.

Ganey, G. Q., M. G. Loso, A. B. Burgess & R. J. Dial. 2017. The role of microbes in snowmelt

and radiative forcing on an Alaskan icefield. Nature Geoscience 10:754-759.

Giersch, J. J., S. Hotaling, R. P. Kovach, L. A. Jones & C. C. Muhlfeld. 2017. Climate-induced

glacier and snow loss imperils alpine stream insects. Global Change Biology 23:2577-

2589.

Hardy, S. P., D. R. Hardy & K. C. Gil. 2018. Avian nesting and roosting on glaciers at high

elevation, Cordillera Vilcanota, Peru. The Wilson Journal of Ornithology 130:940-957.

Hotaling, S., D. S. Finn, J. J. Giersch, D. W. Weisrock & D. Jacobsen. 2017. Climate change

and alpine stream biology: progress, challenges, and opportunities for the future.

Biological Reviews 92:2024-2045.

Hotaling, S., S. Lutz, R. J. Dial, A. M. Anesio, L. G. Benning, A. G. Fountain ... & T. L. Hamilton.

2021. Biological albedo reduction on ice sheets, glaciers, and snowfields. Earth-Science

Reviews 103728.

Hotaling, S., P. H. Wimberger, J. L. Kelley & H. E. Watts. 2020. Macroinvertebrates on glaciers:

a key resource for terrestrial food webs? Ecology 101:e02947.

Hugonnet, R., R. McNabb, E. Berthier, B. Menounos, C. Nuth, L. Girod ... & A. Kääb. 2021.

Accelerated global glacier mass loss in the early twenty-first century. Nature 592:726-

731.

Inman, R.M., A. J. Magoun, J. Persson & J. Mattisson. 2012. The wolverine's niche: linking

reproductive chronology, caching, competition, and climate. Journal of Mammalogy

93:634-644.

Lofroth, E. C. & J. Krebs. 2007. The abundance and distribution of wolverines in British

Columbia, Canada. The Journal of Wildlife Management 71:2159-2169.

Milner, A. M., K. Khamis, T. J. Battin, J. E. Brittain, N. E. Barrand, L. Füreder ... & L. E. Brown.

2017. Glacier shrinkage driving global changes in downstream systems. Proceedings of

the National Academy of Sciences 114:9770-9778.

Muhlfeld, C. C., T. J. Cline, J. J. Giersch, E. Peitzsch, C. Florentine, D. Jacobsen & S. Hotaling.

2020. Specialized meltwater biodiversity persists despite widespread deglaciation.

Proceedings of the National Academy of Sciences 117:12208-12214.

National Park Service. 2021a. Annual snowfall totals at Paradise, 1920 to 2020 [Online].

Retrieved from: https://www.nps.gov/mora/planyourvisit/annual-snowfall-totals.htm on 29

August 2021.

National Park Service. 2021b. Second family of wolverines documented at Mount Rainier

National Park [Online]. Retrieved from: https://www.nps.gov/mora/learn/news/second-

family-of-wolverines-documented-at-mount-rainier-national-park.htm on 5 September

2021.

R Core Team. 2021. R: A language and environment for statistical computing.

Ren, Z., N. Martyniuk, I. A. Oleksy, A. Swain & S. Hotaling. 2019. Ecological stoichiometry of

the mountain cryosphere. Frontiers in Ecology and Evolution 7:360.

Rosvold, J. 2016. Perennial ice and snow-covered land as important ecosystems for birds and

mammals. Journal of Biogeography 43:3-12.

Schano, C., C. Niffenegger, T. Jonas & F. Korner-Nievergelt. 2021. Hatching phenology is

lagging behind an advancing snowmelt pattern in a high-alpine bird. Scientific Reports

11:1-11.

Shain, D. H., P. M. Novis, A. G. Cridge, K. Zawierucha, A. J. Geneva & P. K. Dearden. 2021.

Five animal phyla in glacier ice reveal unprecedented biodiversity in New Zealand's

Southern Alps. Scientific Reports 11: 1-7.

Stibal, M., J. A. Bradley, A. Edwards, S. Hotaling, K. Zawierucha, J. Rosvold ... & A. M. Anesio.

2020. Glacial ecosystems are essential to understanding biodiversity responses to

glacier retreat. Nature Ecology & Evolution 4:686-687.

Washington Department of Fish and Wildlife. 2021. Species in Washington [Online]. Retrieved

from: https://wdfw.wa.gov/species-habitats/species on 15 November 2021.

Table 1. Taxa detected on the Paradise Glacier and directly adjacent habitat of Mount Rainier, WA,

USA. N_{DET}: number of detections. Taxa were classified to the lowest taxonomic possible.

Species	Order, Family	Camera	N_{DET}
Birds			
American kestrel (*Falco sparverius*)	Falconiformes, Falconidae	C5	1
American pipit (*Anthus rubescens*)	Passeriformes, Motacillidae	C1-4, C6	18
Common raven (*Corvus corvax*)	Passeriformes, Corvidae	C2	1
Gray-crowned rosy finch (*Leucosticte tephrocotis*)	Passeriformes, Fringillidae	C3-4, C6	5
Horned lark (*Eremophila alpestris*)	Passeriformes, Alaudidae	C1, C3	5
Mountain bluebird (*Sialia corrucoides*)	Passeriformes, Turdidae	C4	1
White-tailed ptarmigan (*Lagopus leucura*)	Galliformes, Phasianidae	C1-2, C6	6
Unidentified passerines	Passeriformes	C6	4
Unidentified birds	n/a	C2, C4-6	4
Mammals			
Bushy-tailed wood rat (*Neotoma cinereal*)	Rodentia, Cricetidae	C1-2	2
Cascade red fox (*Vulpes vulpes cascadensis*)	Carnivora, Canidae	C4	1
Cascade Golden-mantled ground squirrel (*Callospermophilus saturatus*)	Rodentia, Sciuridae	C1-2, C4, C6	2
Hoary marmot (*Marmota caligata*)	Rodentia, Sciuridae	C1, C4, C6	13
Human (*Homo sapiens*)	Primates, Hominidae	C1-6	208
Mountain goat (*Oreamnos americanus*)	Artiodactyla, Bovidae	C4-6	13
Pacific jumping mouse (*Zapus trinotatus*)	Rodentia, Cricetidae	C1	4
Wolverine (*Gulo gulo*)	Carnivora, Mustelidae	C4-6	8
Unknown *Mustela*	Carnivora, Mustelidae	C4	1

351
352
Figure 1. (a) Mount Rainier, Washington, USA, viewed from the south with our camera locations

labeled. Satellite imagery from Google Earth. (b, c) Sentinel-2 satellite imagery of our camera

locations when they were first deployed (b) and once the bulk of seasonal snow had melted (c).

The outline of the Paradise Glacier (thin green line) in (b) and (c) was approximated from

Beason (2017).

Figure 2. (a) A close-up view of an example camera trap on the margin of the Paradise Glacier

on Mount Rainier, Washington, USA. (b) A more distant view of the camera (within the red

square) showing how it blends into the surrounding talus. Photographs: © Scott Hotaling.

Figure 3. Examples of taxa detected with camera traps along the western margin of the

Paradise Glacier, Mount Rainier, WA, USA. (a) Common raven (*Corvus corvax*), (b) humans

(*Homo sapiens*), (c) Mountain bluebird (*Sialia corrucoides*), (d) mountain goat (*Oreamnos*

americanus), (e) wolverine (*Gulo gulo*), (f) Cascade red fox (*Vulpes vulpes cascadensis*), (g)

White-tailed ptarmigan (*Lagopus leucura*), (h) Hoary marmot (*Marmota caligata*).

Figure 4. Humans and wildlife observed in this study by elevation (top to bottom) and seasonal

timing (grouped on a per week basis; left to right for each taxonomic group). Observations refer

to unique sightings and the total observations for a given taxonomic group are given at the top

of the plot (N_{ALL}). The viewer should note the substantial scale difference of the y-axis for human

observations relative to wildlife. A short-term camera failure at C5 is denoted by the labeled light

pink square.