

27 **Abstract**

28 Pesticides are critical for invasive species management, but often have negative effects on non-
29 target native biota. Tolerance to pesticides should have an evolutionary basis, but this is poorly
30 understood. Invasive sea lamprey (*Petromyzon marinus*) populations in North America have
31 been controlled with a pesticide lethal to them at lower concentrations than native fishes. We
32 addressed how interspecific variation in gene expression and detoxification gene diversity confer
33 differential pesticide sensitivity in two fish species. We exposed sea lamprey and bluegill
34 (*Lepomis macrochirus*), a tolerant native species, to TFM, a pesticide commonly used in sea
35 lamprey control, and used whole-transcriptome sequencing of gill and liver to characterize the
36 cellular response. Comparatively, bluegill exhibited a larger number of detoxification genes
37 expressed and a larger number of responsive transcripts overall, which likely contributes to
38 greater tolerance to TFM. Understanding the genetic and physiological basis for pesticide
39 tolerance is crucial for managing invasive species.

40

41 Invasive species represent a considerable threat to local biodiversity and ecosystem
42 functioning, having significant economic costs associated with lost ecosystem services and
43 control efforts¹. Invasive species control is multifaceted and serves primarily to reduce, but not
44 necessarily eradicate, populations of the target organism². Often, chemical pesticides are used to
45 kill the target species while attempting to minimize non-target impacts². In aquatic environments,
46 this can be difficult as pesticides are often applied to large portions of the waterbody, thus
47 exposing all community members³. However, species-specific pesticides are relatively rare and,
48 consequently, there will likely be impacts on native species in most pesticide application
49 programs⁴.

50 The degree to which different species are adversely affected by pesticide applications can
51 be highly species-specific. Interspecific variation in pesticide sensitivities occurs in amphibians⁵,
52 invertebrates⁶, and fishes⁷, which suggests that species-specific differences in pesticide uptake,
53 biotransformation, and elimination (i.e., the disposition) are likely driving differential
54 sensitivity⁸. Further, phylogenetic signatures associated with toxicity tolerances have been
55 identified in several clades of aquatic ectotherms⁹. This suggests that toxicity tolerances likely
56 have an evolutionary basis and are in part genetically determined¹⁰. Recent transcriptional
57 profiling of several teleost fishes has shown considerable interspecific variation amongst genes
58 involved in detoxification of pesticides and other xenobiotics¹¹, underscoring an important
59 genomic role in toxicity tolerances. Consequently, using molecular toolsets for understanding
60 pesticide responses are valuable in providing insight into phylogenetic variation in physiological
61 response and in predicting community impacts of pesticide applications^{9,10}. This is relevant in
62 contemporary invasion control efforts, where minimizing harm to native species is often a key
63 objective¹². To date, our understanding of the genetic and evolutionary underpinnings of
64 interspecific variation in toxicity tolerance is limited, hindering our insight into community-level
65 effects of pesticide applications.

66 The sea lamprey (*Petromyzon marinus*) control program in the Laurentian Great Lakes of
67 North America (hereafter referred to as the Great Lakes) is an excellent model system to explore
68 the evolutionary and genetic foundations of differential toxicity tolerance to a pesticide in target
69 (i.e., invasive sea lamprey) and non-target (native) species. Sea lamprey, although native to the
70 Atlantic Ocean, likely gained access to Lake Ontario through canals in the mid-1800s; the

71 Welland Canal subsequently allowed access to the remaining Great Lakes by the early 1900s¹³.
72 By the 1950s, sea lamprey populations in the Great Lakes had exploded, decimating native
73 fisheries, notably lake trout (*Salvelinus namaycush*), having severe impacts on the environment
74 and economy of the region¹³. An international Great Lakes Fishery Commission (GLFC)
75 involving Canadian and American institutions was formed in 1955 to co-manage fisheries and to
76 control sea lamprey populations. Since the 1950s, the GLFC's primary method of sea lamprey
77 control has involved treating streams and rivers with pesticides, commonly referred to as
78 lampricides, that target burrowing sea lamprey larvae in Great Lakes tributaries¹³. Because the
79 filter-feeding larval stage lasts for ~4–5 years before sea lamprey metamorphose into parasitic
80 juveniles that disperse widely in the lakes, lampricide treatments in streams containing sea
81 lamprey larvae kill multiple year classes at once.

82 Two pesticides are used to control sea lamprey, 3-trifluoromethyl-4-nitrophenol (TFM)
83 and niclosamide (2',5-dichloro-4'-nitrosalicylanilide), each of which contains a halogenated
84 phenol ring that contributes to their toxicity³. In most treatments, TFM is applied at
85 concentrations reaching the sea lamprey 9-h LC_{99.9} of TFM¹⁴, the concentration needed to kill
86 99.9% of the exposed population, and is co-applied with 1-2% niclosamide because these
87 compounds interact in a greater than additive fashion¹⁵. While niclosamide is broadly toxic to
88 fishes³, TFM is highly toxic to lampreys relative to jawed fishes. The mode of action of TFM is
89 interference with oxidative phosphorylation in the mitochondria¹⁶, which lowers aerobic ATP
90 production forcing the animal to increasingly rely on unsustainable anaerobic glycolysis^{17–19}.
91 However, TFM sensitivity varies greatly among jawed fishes, which makes understanding the
92 broader non-target impacts of TFM difficult to resolve³. For example, lake sturgeon (*Acipenser*
93 *fulvescens*), a non-teleost jawed fish, are moderately sensitive to TFM³, while centrarchids,
94 which includes basses and bluegill (*Lepomis macrochirus*), have some of the highest reported
95 tolerances to TFM amongst teleost fishes in the Great Lakes, with reported LC values ~10-fold
96 higher than sea lamprey¹⁹.

97 Little is understood about the physiological mechanisms driving varying tolerances to
98 TFM among fishes³. Detoxification of TFM is believed to occur through a combination of Phase
99 I and Phase II biotransformation processes, which collectively act to enhance the water solubility
100 of the compound to improve elimination³. It is generally believed that UDP-

101 glucuronosyltransferase (Ugt) is the main enzyme responsible for TFM detoxification, via
102 conjugation of TFM through the addition of a UDP-glucuronic acid²⁰⁻²². The greater sensitivity
103 of lampreys to TFM is believed to be related to lower Ugt activities compared to other fishes²⁰⁻
104 ²², but the other cellular aspects of TFM detoxification are poorly characterized in fishes.
105 Furthermore, interspecific variation in detoxification gene diversity is believed to be important in
106 driving differences in toxicity tolerances among fishes¹¹, which could suggest that available *ugt*
107 genes impart differential TFM tolerances between sea lamprey and non-target fishes. Indeed,
108 compared to zebrafish (*Danio rerio*; N =40), sea lamprey have far fewer *ugt* genes (N= 2) in
109 their genome²³. A few comparative studies have been conducted looking at sea lamprey and
110 rainbow trout (*Oncorhynchus mykiss*), however little attention has been paid to the physiological
111 effects of TFM exposure in other fishes¹⁶⁻¹⁹. As TFM detoxification primarily occurs in the
112 liver, and the gills are one of the primary uptake sites, TFM physiological studies often focus on
113 these two tissue types in fishes³. However, a broader characterization of TFM toxicity,
114 particularly the expression patterns of Phase I and Phase II enzyme genes, in fishes is necessary.

115 We used a comparative transcriptomics approach to evaluate the effects of TFM on sea
116 lamprey and bluegill, species with the lowest and highest reported tolerances to TFM,
117 respectively¹⁹. The purpose of this study was to: 1) characterize interspecific variation in
118 detoxification gene families expressed in the transcriptome following TFM exposure, 2) assess
119 the potential mechanisms of TFM detoxification common to both species, and 3) identify key
120 taxon-specific differences in the physiological response to TFM between sea lamprey and a
121 TFM-tolerant teleost fish. We predicted that bluegill would have a greater diversity of
122 detoxification mechanisms alongside a large robust transcriptomic detoxification response that
123 contributes to their greater tolerance to TFM relative to sea lamprey. We also predicted that the
124 molecular signatures of TFM toxicity would become more evident with increasing exposure
125 duration as detoxification systems become exhausted; effects that would be more pronounced in
126 the sea lamprey. These results will help to identify the evolutionary and genetic underpinnings of
127 pesticide tolerance, which may be used to inform improvements in invasive species control
128 efforts across a diversity of contexts.

129 **Results**

130 *Detoxification gene expression patterns*

131 Sea lamprey and bluegill were exposed to TFM for 6, 12, and 24 h at their species-
132 specific 24 h-LC₁₀ (sea lamprey = 2.21 mg L⁻¹, bluegill = 22.06 mg L⁻¹) with gill and liver tissue
133 being sampled at 6, 12, and 24 h of exposure¹⁹. The mRNA from the gills and liver were
134 extracted and sequenced using a whole transcriptome approach for the identification of gene
135 targets associated with TFM detoxification and gene ontology (GO) term enrichment. As
136 variation in TFM tolerance is believed to stem from differences in biotransformation capacities
137 in fishes^{3,21}, we filtered for Phase I-III biotransformation transcripts from each species'
138 transcriptome. Proteins listed as being involved with detoxification, xenobiotic removal, and/or
139 organic compound breakdown were retained as these functions likely made them important in
140 TFM detoxification. Briefly, Phase I biotransformation generally includes the hydrolysis of
141 organic compounds while Phase II biotransformation involves conjugation of organics with both
142 processes enhancing xenobiotic solubility for easier elimination. Phase III biotransformation
143 typically involves facilitating the excretion of the xenobiotic²⁴.

144 Several Phase I detoxification transcripts were affected by TFM treatment in both
145 species. In sea lamprey, transcripts associated with Phase I detoxification processes were only
146 differentially expressed in the gills and consisted mainly of cytochrome P450s (six unique *cyp*'s;
147 Table 1). Notably, this included both *cyp1a1* (6 h) and *cyp1b1* (24 h) with a paraoxonase (*pon1*)
148 being the only non-*cyp* transcript differentially expressed (Table 1). Most *cyp* transcripts were
149 upregulated in response to TFM exposure at 6 or 24 h, but *pon1*, *cyp2j2*, and *cyp2d3* were
150 downregulated at 24 h. Contrastingly, bluegill demonstrated a larger induction of Phase I
151 detoxification transcripts (nine unique transcripts, 28 differentially expressed transcripts) in both
152 tissues across all exposure durations (Table 1). In bluegill gill, *cyp1a1* and *cyp1b1* were
153 upregulated across all timepoints (Table 1). In the bluegill liver, several *cyp* transcripts
154 responded positively to TFM treatment across exposure durations (e.g., *cyp1a1*, *cyp1b1*,
155 *cyp2c31*), but three transcripts (*cyp2b2*, *cyp2k1*, *cyp2k6*) were downregulated at 24 h (Table 1).

156 In the transcriptomes of both species, we identified several *ugt*'s likely involved in the
157 detoxification of TFM (Table 2). In sea lamprey, six clusters from the superTranscriptome
158 mapped to *ugt*'s. In contrast, the bluegill transcriptome had 17 clusters which mapped to *ugt*'s
159 (Table 2). There were also noticeable differences between sea lamprey and bluegill in expression
160 of the *ugt* transcripts in response to TFM exposure. Treatment with TFM did not result in any *ugt*

161 being differentially expressed in sea lamprey. Conversely, in bluegill, *ugt3* was upregulated in
162 the gill across all three exposure durations (Table 3) and in the liver at 12 and 24 h of exposure.
163 Interestingly, *ugt3* was not present in the lamprey transcriptome (Table 2). Hepatic *ugt2b9*
164 transcript levels were also elevated under TFM exposure at 12 and 24 h in bluegill liver, and
165 TFM exposure led to several *ugt* transcripts (i.e., *ugt1a2*, *ugt1a5*, *ugt2a1*) being downregulated
166 in the liver of bluegill at 12 and 24 h of TFM exposure (Table 3).

167 There was also a noticeable difference between bluegill and sea lamprey in the
168 expression patterns for other transcripts involved in Phase II detoxification (Table 4). In sea
169 lamprey, there were only three differentially expressed transcripts for Phase II genes, including
170 glutathione S-transferase (*gstt3*), sulfotransferase (*sult6b1*), and nicotinamide N-
171 methyltransferase (*nmmt*), in both liver and gill, all at 24 h of exposure (Table 4). In bluegill,
172 several sulfotransferases and a single N-acetyltransferase (*nat8*; Table 4) transcripts were
173 upregulated in the gill. In the bluegill liver, Phase II detoxification transcripts were only
174 differentially regulated at 24 h of TFM exposure with N-acetyltransferases (*nat2* and *nat8*) being
175 the only upregulated transcripts and two glutathione S-transferases (*gstt3* and *gstt1*), a thiopurine
176 S-methyltransferase (*tpmt*), and a sulfotransferase (*sult1c1*) being downregulated (Table 4).

177 Differential expression of Phase III detoxification transcripts was limited to 24 h of
178 exposure in sea lamprey. In the gills, five ATP-binding cassettes (*abc*; *abcc3*, *abccc5*, *abcb1*,
179 *abcb11*, and *abcg2*) were upregulated in response to TFM, whereas the two solute carrier family
180 transcripts (*slc*; *slc35d2* and *slc22a15b*) were downregulated (Table 5). In the lamprey liver, a
181 single *abc* and *slc* were the only transcripts differentially expressed in response to TFM (Table
182 5). In contrast, the gills of bluegill exhibited differential expression of Phase III transcripts across
183 all three timepoints consisting of four unique *abc*'s (*abca3*, *abcb1*, *abcb6*, and *abcc4*) and nine
184 unique *slc*'s (Table 5). In most instances, these branchial *abc*'s and *slc*'s were upregulated in
185 bluegill. In the bluegill liver, a single *abc* (*abcb6*) was found to be upregulated at 12 and 24 h of
186 TFM exposure (Table 5). Aside from *slc47a1*, all of the five-remaining hepatic *slc*'s were found
187 be upregulated under TFM treatment and with differential expression occurring at 12 and 24 h of
188 exposure (Table 5).

189 *Overall transcriptome patterns*

190 The overall number of differentially expressed transcripts and their timing were different
191 in the two species. In the sea lamprey, there was a delayed transcriptional response in the gills
192 following TFM exposure (Fig. 1a). At 6 and 12 h, we observed only 139 and 9 differentially
193 expressed superTranscriptome clusters, respectively (hereafter referred to as ‘clusters’).
194 However, by 24 h, there were 7,055 differentially expressed clusters. In contrast, bluegill gill
195 expression patterns were marked by a stepwise increase in the number of differentially expressed
196 superTranscriptome clusters. A total of 595 and 2,662 differentially expressed clusters were
197 already observed at 6 and 12 h of TFM exposure, respectively, although the total at 24 h (4,370)
198 was less than the 24-h total in sea lamprey gills (Fig. 1b).

199 The livers of sea lamprey showed a muted pattern of transcriptional activity under TFM
200 exposure with no changes in transcription within the first 12 h of exposure, and only 1,940
201 differentially expressed clusters by 24 h (Fig. 1c). In bluegill, there were moderate levels of
202 differential expression in the liver within the first 12 h of TFM exposure (388 and 1,464 at 6 and
203 12 h, respectively), and the number had increased to 5,685 differentially expressed clusters by 24
204 h of exposure (Fig. 1d).

205 *Lamprey GO term enrichment*

206 In sea lamprey gills, TFM treatment resulted in an enrichment of GO terms associated
207 with cellular growth and proliferation, immune function, and metabolism by 24 h of exposure.
208 Specifically, in transcripts that were upregulated, GO terms associated with cellular proliferation
209 and growth included regulation of cell proliferation, Ras protein signaling, regulation of I- κ B
210 kinase/signaling, and regulation of the cell cycle (Fig. 2a). GO terms related to immune function
211 including viral life cycle, antigen processing, and negative regulation of type I interferon were
212 also enriched for upregulated transcripts in sea lamprey (Fig. 2a). An upregulation in GO terms
213 associated with energy metabolism, including “cellular response to insulin stimulus”, and
214 “regulation of gluconeogenesis”, also featured prominently in the sea lamprey gill (Fig. 2a). Of
215 the gluconeogenic transcripts that were enriched, we specifically observed an upregulation in
216 mitochondrial pyruvate dehydrogenase kinase 2 (*pdk2*) and glycerol-3-phosphate phosphatase
217 (*pgp*).

218 In sea lamprey liver, there were no differentially expressed transcripts at 6 h and 12 h of
219 TFM exposure (Fig. 1c). At 24 h, there was an enrichment of upregulated transcripts associated

220 with the GO term “positive regulation of transcription” (Fig. 2b). Within this parent term, cell
221 growth and survival GO terms were enriched, which included regulation of cell proliferation, fat
222 cell differentiation, autophagy, and ubiquitination. Additionally, processes related to energy
223 metabolism appeared to be upregulated as there was enrichment of transcripts associated with
224 “cellular response to insulin stimulus” (e.g., *igf1r*, *insr*, *pck2*, *pdk2*) and “negative regulation of
225 lipid storage” (e.g., *abca1*, *nfkbia*, *osbpl8*; Fig. 2b).

226 *Bluegill GO term enrichment*

227 In bluegill gills, there was enrichment of upregulated transcripts associated with cellular
228 growth, proliferation, and death. For example, at 6 h of TFM exposure, there was enrichment of
229 apoptotic processes regulation (Fig. 3a). By 12 h, this included a “positive regulation of cell
230 proliferation”, “transforming growth factor β (TGF- β) responsiveness”, “regulation of
231 autophagy”, “regulation of apoptotic process”, and “ubiquitin dependent processes” (Fig. 3a). By
232 24 h, enrichment was largely restricted to cellular death including GO terms related to apoptosis
233 and macroautophagy (Fig. 3a). Several transcripts associated with apoptosis, including *akt2*,
234 several *bmp* and *map3k* transcripts, and *casp3*, were upregulated under TFM. For downregulated
235 transcripts in the gill, there was also further enrichment of processes related to cellular growth
236 and death (Fig. 3b). At 6 h, we observed a negative regulation of several growth-related
237 processes including a “positive regulation of cell proliferation” and “Wnt signaling” (Fig. 3b).
238 By 12 h, there was an enrichment of downregulated transcripts associated with “positive
239 regulation of intracellular signal transduction” in the gill which included several growth-related
240 processes (Fig. 3b). By 24 h, the parent term “cytokine-mediated signaling pathway” was the
241 largest enriched parent term and included the downregulation of transcripts associated with
242 several GO terms related to cell growth and survival, including cell proliferation, apoptosis, and
243 G2/M mitotic transition (Fig. 3b).

244 The bluegill gill’s response to TFM treatment involved expression changes of transcripts
245 associated with immune function. For upregulated transcripts, “negative regulation of
246 macrophage derived foam cell differentiation” (12 h) and “cellular response to transforming
247 growth factor beta stimulus” (12 and 24 h) were the main responses (Fig. 3a). However, there
248 was a large enrichment of GO terms associated with the immune response in downregulated
249 transcripts in the gill (Fig. 3b). At 12 h, cytokine signaling, and T cell migration were enriched.

250 By 24 h, the parent term “cytokine-mediated signaling pathway” included GO terms associated
251 with immune function including “inflammatory response” and NF- κ B signaling (Fig. 3b). There
252 was also an enrichment of downregulated transcripts associated with the parent terms “antigen
253 presentation via MHC class I” and “viral life cycle” (Fig. 3b).

254 TFM exposure also affected metabolic processes in bluegill gills. For upregulated
255 transcripts, this included enrichment of lipid metabolism (e.g., *abca1*, *cyp1a1*, *angptl4*; 12 h),
256 gluconeogenesis (e.g., *pdk2*, *pgp*, *ptpn2*; 12 h), reactive oxygen species (ROS; e.g., *cdkn1a*,
257 *slc25a33*, *thbs1*; 12 h), and glucose import (*insr*, *irs2*, *c1qtnf12*; 24 h). In downregulated
258 transcripts, enriched metabolic processes included a regulation of primary metabolic processes
259 and cellular amino acid metabolic processes (Fig. 3b). Tangentially, there was also an
260 enrichment of GO terms in upregulated transcripts that were associated with a response to an
261 organic compound (6h; Fig. 3a), which included the upregulation of several transcript targets,
262 notably cytochrome P450s (*cyp1a1* and *cyp1b1*), caspases (*casp3* and *casp7*), and an aryl
263 hydrocarbon receptor (*ahr*).

264 In the bluegill liver, TFM effects were generally restricted to transcripts pertaining to
265 cellular growth, proliferation, and cellular death. Specifically, in upregulated transcripts, there
266 was an enrichment of the parent GO terms “regulation of autophagy” (6 h), “positive regulation
267 of apoptotic process” (12 h), “macroautophagy” (24 h), and “ubiquitin-dependent protein
268 catabolism” (24 h; Fig. 4a). At 24 h, there was also enrichment of cell growth/death GO terms
269 for downregulated transcripts including regulation of programmed cell death and cell cycle
270 G2/M transitions (Fig. 4b). TFM also led to an enrichment of metabolic functions in the bluegill
271 liver. Specifically, glycogen metabolic processes were enriched in downregulated transcripts
272 which corresponded with lower differential expression of a glycogen debrancher enzyme (*agl*),
273 α -glucosidase (*gaa*), glycogen synthase (*gys2*), and glycogen phosphorylase (*pygl*; Fig. 4b).

274 **Discussion**

275 Variation in the availability and functional status of different detoxification pathways
276 plays a critical role in determining the interspecific and intraspecific sensitivity of animals to
277 different pesticides¹⁰. Accordingly, we predicted that interspecific variation in the expression and
278 diversity of transcripts coding for biotransformation enzymes would partially explain the
279 differences in TFM tolerances between sea lamprey and bluegill. As predicted, we identified

280 considerable interspecific variation in the transcriptomic responses to TFM exposure where all
281 three suites of biotransformation transcripts (i.e., Phase I-III) were more diverse and responsive
282 in bluegill relative to sea lamprey, suggesting a greater inherent ability to detoxify TFM. The
283 differences in the transcriptome responses between the species supports an evolutionary basis of
284 TFM tolerances in fishes. These results likely explain why >15-fold higher concentrations of
285 TFM are required to produce a toxic effect in bluegill versus sea lamprey³, as well as the 10-fold
286 higher 24 h LC₁₀ observed in bluegill compared to sea lamprey¹⁹. A functional transcriptomics
287 approach has therefore provided us with greater insight into the molecular basis for the different
288 sensitivities of these two species to TFM, and how transcriptomic responses associated with
289 detoxification genes explains these observations (Fig. 5).

290 *Evolutionary considerations in lampricide tolerances*

291 We found that differences in the expression and diversity of biotransformation transcripts
292 could help to explain interspecific variation in TFM sensitivity among target sea lamprey and
293 non-target teleost fishes. Specifically, differences in *ugt* transcript diversity and expression were
294 likely a major factor dictating TFM tolerances in fishes. Variation in detoxification capacities
295 have been exploited to improve the specificity of pesticides used across a broad range of
296 agricultural, commercial, and domestic settings, while minimizing impacts on non-target species.
297 For example, current agricultural practices make use of a combination of bioengineered crop
298 plants that are tolerant to pesticides (e.g., Roundup Ready™²⁵) as well as using taxon-specific
299 pesticides (e.g., 2,4-dichlorophenoxyacetic acid in cereal crops²⁶) to selectively control
300 dicotyledons (broadleaf plants) without harming valuable monocotyledon crops. In contrast,
301 many chemical insecticides (e.g., organophosphates and neonicotinoids) are broad-spectrum
302 rather than selective pesticides, with detrimental non-target effects generally being reduced by
303 applying them in a selective manner (e.g., via spot treatments or in baits)²⁷. Thus, furthering our
304 understanding of the genetic, physiological, and evolutionary mechanisms driving variation in
305 pesticide sensitivities can allow for the development of more effective and targeted pesticide
306 applications that minimize impacts on non-target species². As pesticides used as lampricides
307 cannot be applied in this manner, understanding genetic variation in detoxification mechanisms
308 is key in protecting local biodiversity.

309 The sea lamprey control program in the Laurentian Great Lakes is perhaps the most
310 notable and successful example of pesticides use in an invasive vertebrate³. High specificity was
311 key in ensuring minimal impacts on non-target species as the pesticide was being administered
312 directly into the natural environment. During the 1950's, researchers conducted a large-scale
313 series of toxicity tests, screening over 5,000 different compounds for an agent that could kill sea
314 lamprey with higher selectivity than native species. Eventually, compounds containing a
315 nitrophenol group were identified as the most selective agents, with TFM being the best
316 substance²⁸. Stream treatment with TFM has resulted in an estimated 90% reduction in sea
317 lamprey populations in the Great Lakes³.

318 The exact mechanism underpinning the variation in TFM sensitivity among fishes has
319 remained illusive. Phylogenetic variation in TFM sensitivity was apparent as closely related
320 species (i.e., within order) demonstrated similarities in TFM toxicity values^{3,19}. Notably, native
321 lampreys had comparable toxicity values (i.e., minimum lethal concentration 99% [MLC99]) to
322 sea lamprey and were also adversely affected by lampricide treatments²⁹, suggesting a taxonomic
323 basis for the sea lamprey's high sensitivity to TFM. Indeed, it is believed that Ugt's are the
324 principal enzyme in TFM detoxification as part of Phase II biotransformation^{3,20,21,30}. In sea
325 lamprey, Ugt activities and numbers of *ugt* gene isoforms appear to be low relative to teleost
326 fishes²¹⁻²³, suggesting an underlying genetic basis in dictating TFM tolerances. The sea
327 lamprey's phylogenetic position as a basal vertebrate could help explain the disparity in TFM
328 tolerances. Lampreys diverged from the lineage leading to jawed vertebrates ~500 million years
329 ago, most likely after one round of whole genome duplication (WGD), with the second WGD
330 occurring only in jawed vertebrates^{31,32} and a third WGD in teleosts³¹; these WGD events are
331 important in allowing the generation of new traits that can facilitate adaptation³¹. Consequently,
332 this provided teleosts with a more diverse set of detoxification genes relative to lampreys, and
333 perhaps even to non-teleost fishes like lake sturgeon, which are moderately sensitive to TFM
334 (although less than lampreys³).

335 *Patterns of ugt expression*

336 We show for the first time that there is a genetic basis underlying differences in TFM
337 tolerances among target and non-target freshwater vertebrates. Specifically, bluegill had a larger
338 diversity of *ugt*'s, compared to sea lamprey, and only bluegill showed differential expression of

339 *ugt*'s in their gill and liver transcriptomes due to TFM exposure. In line with our predictions, we
340 showed that differences in *ugt* expression patterns between bluegill and sea lamprey are likely
341 one of the main factors driving interspecific variation in TFM tolerance in fishes.

342 Bluegill exhibited differential expression in the *ugt* gene family including *ugt1*, *ugt2*, and
343 *ugt3*. In bluegill, *ugt3* appeared to be the most important *ugt* transcript in both tissues as it was
344 expressed at the highest levels and underwent the greatest fold-increase (upwards of 28x),
345 compared to *ugt1* and *ugt2*. *ugt3* genes have not been identified in fishes previously and was
346 only recently discovered in humans with their functional role remaining uncertain³³. However,
347 the Ugt3 enzyme appears to be involved in detoxifying polycyclic aromatic hydrocarbons, and 4-
348 nitrophenol in mammals³³, suggesting a role in TFM detoxification in bluegill as well. We also
349 noted an upregulation of *ugt2b9* in the bluegill liver. Thus, hepatic biotransformation using *ugt3*
350 and, to a lesser extent, *ugt2b9* may be a key factor in the high tolerance of bluegill to TFM
351 compared to sea lamprey and perhaps other teleost fishes³.

352 *Insight into detoxification pathways using transcriptomics*

353 Variation in the detoxification transcriptome between bluegill and sea lamprey was
354 evident in genes associated with Phase I biotransformation, predominantly genes coding for
355 *cyp*'s, which can detoxify organic compounds²⁴. These enzymes may have a role in TFM
356 detoxification as evident by the presence of Cyp-conjugated TFM metabolites in fishes²¹, but
357 their relative importance compared to Phase II processes is uncertain. In both bluegill and sea
358 lamprey, *cyp1a1* and *cyp1b1* appeared to be the dominant *cyp* transcripts that responded to TFM,
359 which is unsurprising given that *cyp1* genes are involved in the detoxification of a diversity of
360 organic compounds²⁴. Interestingly, bluegill had a far greater magnitude of response, as evident
361 by the high fold-changes in expression (upwards of ~187x), and a greater diversity of
362 differentially expressed *cyp* transcripts, compared to sea lamprey. However, it is unclear how
363 important Cyp's are for detoxifying TFM, as no studies have detected the corresponding Phase I
364 metabolites *in vivo*, as compared to the Phase II sulfate and glucuronide conjugates^{21,22,34}.
365 Further studies, both *in vitro* and *in vivo*, are required to better understand the functional role of
366 Cyp's in TFM biotransformation in both species.

367 Phase II biotransformation transcripts also showed a clear taxonomic divergence that may
368 help explain interspecific variation in TFM tolerances. Bluegill exhibited an upregulation of
369 sulfotransferases (N=3) and N-acetyltransferases (N =2) across both tissue types, whereas only a
370 single sulfotransferase and glutathione S-transferase responded in sea lamprey gill. As
371 sulfotransferases are important in TFM detoxification^{3,21}, the greater capacity to conjugate via
372 sulfation as well as through glucuronidation in bluegill likely supports high TFM metabolism
373 when compared to sea lamprey. The presence of N-acetyltransferases in bluegill also suggests
374 that TFM is undergoing acetylation²⁴, which has been characterized in bluegill previously²¹. Our
375 results suggest a taxonomic disparity in Phase II biotransformation pathways involved in TFM
376 detoxification, with bluegill able to upregulate several Phase II biotransformation processes to
377 effectively eliminate TFM compared with sea lamprey.

378 Interspecific variation was also evident in Phase III biotransformation processes, which
379 are mainly involved in the transport and elimination of toxicants²⁴. We detected a divergent
380 effect of TFM exposure on Phase III transcripts where numbers of responding *abc* and *slc*
381 transcripts were greater in sea lamprey and bluegill, respectively. This might suggest a disparity
382 in TFM elimination routes given that *Abc*'s are generally active transporters while *Slc*'s are
383 bidirectional passive transporters²⁴. In the lamprey, the high *abc* response, as well as the general
384 lack of transporter expression in the liver, may suggest a predominantly branchial-mediated
385 elimination³ as any TFM excretion would be against a concentration gradient (i.e. high
386 environmental TFM concentration). In bluegill, the high environmental concentrations of TFM
387 (~21 mg L⁻¹ [101 µM] in water, 23 nmol g⁻¹ ww in liver¹⁹) coupled with the lower branchial *abc*
388 expression, relative to lamprey, suggests that TFM elimination can be through biliary/renal
389 excretion as previously reported in teleosts^{3,30}. Together, this appears to suggest that bluegill
390 have a more responsive and effective system for TFM elimination.

391 *GO term enrichment and tissue responses to TFM*

392 In fishes, studies of TFM toxicity have largely been restricted to characterizations of
393 energy metabolism and mortality^{17,19}. However, we found several biomarkers associated with
394 inhibited cell cycle progression and growth, and higher levels of ubiquitination and apoptosis, in
395 both bluegill and sea lamprey, suggesting that TFM's effects extend beyond simply uncoupling
396 oxidative phosphorylation^{16,35}, indicating a new mode of TFM toxicity in fishes. Increases in

397 ROS or oxidative damage are likely the primary mechanism by which mitochondrial uncouplers
398 exert tissue-level repression of growth³⁶. While TFM effects on ROS generation have not been
399 quantified, TFM does increase mitochondrial respiration rates¹⁶, which may serve to elevate
400 cellular ROS levels³⁷. Coupled with the GO term enrichment patterns, our results suggest that
401 TFM-induced ROS is likely mediating cellular arrest as seen in niclosamide^{38,39} and 2,4-
402 dinitrophenol (DNP, a TFM analogue)^{40,41} exposures. While fish-specific examples are limited,
403 zebrafish (*Danio rerio*) exposed to 4-nitrophenol experienced reduced cellular growth and
404 proliferation, and heightened cell death³⁶, aligning with our results. At a broader scale, increases
405 in oxidative damage by ROS can impair cellular growth and cell cycle progression as well as
406 inducing apoptosis in fishes^{42,43}. Together, our results suggest that TFM has a potential role in
407 generating oxidative damage in the cell.

408 Despite TFM toxicity primarily affecting energy metabolism³, enrichment of these
409 metabolic GO terms was limited. In sea lamprey, there was a negative regulation of GO terms
410 associated with lipid storage, gluconeogenesis, and insulin responsiveness. Likely, these changes
411 were needed to increase energy mobilization for supporting heightened metabolic demands under
412 TFM exposure¹⁷⁻¹⁹. In bluegill, metabolic changes included positive regulation on
413 gluconeogenesis, glucose import, and lipid metabolism and negative regulation of glycogen
414 metabolism suggesting that increased energetic expenditure during TFM exposure is likely
415 occurring, probably due to increased rates of mitochondrial respiration and detoxification costs¹⁹.
416 Also, as predicted, the effects of TFM on both species' transcriptomes was more apparent with
417 longer exposure durations (i.e., more differentially expressed transcripts) suggesting an
418 exhaustion of detoxification systems.

419 *Conclusions*

420 Here, we showed how transcriptome responses demonstrate interspecific variation in
421 pesticide sensitivities, which are critical in developing more targeted invasive species control.
422 We identified differences in *ugt* diversity and responsiveness to TFM, which appeared to be
423 linked to differences in TFM sensitivities between bluegill and sea lamprey highlighting the
424 importance of evolutionary history in mediating toxicity resilience. In the context of sea lamprey
425 control, this knowledge will be useful in developing a predictive framework for assessing
426 community-scale impacts of lampricides. More broadly, similar principles are likely to apply to a

427 wide range of systems where invasive species control efforts use pesticides extensively. By
428 identifying key genes involved in detoxification in native fauna, we can develop methods for
429 addressing the relative impacts that pesticide treatments have on a system. For TFM specifically,
430 this would likely include developing screenings of Ugt's, Cyp's, and Sult's as potential
431 biomarkers of exposure. This approach could help shed light on more subtle interspecific
432 differences in sensitivity to TFM even among teleosts. As invasive species control efforts seek to
433 provide more effective and targeted pesticide applications¹³, furthering our understanding of the
434 toxicological mode of action and evolutionary variation in tolerances among species is likely to
435 greatly improve the use of pesticides.

436 **Methods**

437 *Animal care and holding*

438 Fish collection and holding conditions are presented in Lawrence et al.¹⁹. Briefly,
439 juvenile bluegill ($n = 200$; total length [TL] = 97.3 ± 0.88 mm; mass = 25.51 ± 0.75 g) were
440 sourced from Kinmount Fish Farm (Kinmount, ON, Canada) in September 2018. Bluegill were
441 transported to the animal holding facility at Wilfrid Laurier University (Waterloo, ON, Canada;
442 ~1000 L holding tank; ~ 0.5 L min⁻¹; T = 12–14°C; pH 8.1–8.2, alkalinity ~255 mg L⁻¹ as
443 CaCO₃), where they were provided with a combination of commercial fish feed (EWOS #1,
444 Cargill, ON, Canada) and bloodworms daily.

445 Larval sea lamprey ($n = 568$; TL = 104.94 ± 0.69 mm; mass = 1.47 ± 0.03 g) were
446 electrofished from a tributary of Lake Huron by the United States Fish and Wildlife Service in
447 April 2018 and were temporarily stored at the Hammond Bay Biological Station (Millersburg,
448 MI, USA). Sea lamprey were then transported to Wilfrid Laurier University and held under
449 similar water conditions to the bluegill, although the lamprey tank was smaller (~100 L) and
450 contained a 8–10 cm deep layer of sand to facilitate natural burrowing⁴⁴. Lamprey were fed on a
451 diet of baker's yeast weekly (see Lawrence et al.¹⁹ for specific details). All experimental series
452 were conducted under approval from the Wilfrid Laurier University Animal Care Committee
453 (Animal Use Protocol No. R18001) under the guidelines established by the Canadian Council of
454 Animal Care.

455 *TFM exposures*

456 Larval sea lamprey and bluegill were exposed to either control conditions (i.e., no
457 toxicant) or field grade TFM (35% active ingredient dissolved in isopropanol; Clariant,
458 Griesheim, Germany) for up to 24 h. The TFM exposure concentrations were equivalent to the
459 species-specific concentration of TFM that was lethal to 10% of each species over 24 h (24-h
460 LC₁₀), previously determined for the cohorts of sea lamprey (24-h LC₁₀ = 2.21 mg L⁻¹) and
461 bluegill (24-h LC₁₀ = 22.06 mg L⁻¹)¹⁹. Exposures were performed in triplicate with animals being
462 held in glass aquaria (10 L for sea lamprey larvae, 14 L for bluegill) in groups of 2–3 or 6
463 individuals per tank for bluegill and sea lamprey larvae, respectively. The TFM exposures were
464 static with temperature being maintained at the fish's acclimation temperature (~14–15°C).
465 Tissue sampling at 6 h, 12 h, and 24 h of exposure. At these discrete intervals, fish were netted
466 and placed one at a time into buffered tricaine methanesulfonate (MS-222; Syndel, Nanaimo,
467 BC, Canada; 1.5 g L⁻¹ MS-222 with 3.0 g L⁻¹ NaHCO₃) to euthanize the animals. The livers and
468 gill of both species were excised, and placed into *RNAlater* (Invitrogen, ThermoFisher
469 Scientific, Mississauga, ON, Canada), held at 4°C for at least 24 h, and then stored at -80°C for
470 later RNA extraction.

471 *RNA extraction and transcriptome sequencing*

472 Extractions of total RNA were performed using a commercially available kit (RNeasy
473 Plus Mini Kit; Qiagen, Toronto, ON, Canada) according to the manufacturer's specifications.
474 Following extraction, an initial quality control check was performed using the 260/280
475 absorbance ratio and the 260/230 absorbance ratio with a NanoDrop One Microvolume UV-Vis
476 Spectrophotometer (ThermoFisher, Mississauga, ON, Canada). Quality was further assessed
477 using a Qubit RNA IQ assay (ThermoFisher, Mississauga, ON, Canada), read on a Qubit 4
478 fluorometer (ThermoFisher, Mississauga, ON, Canada), as a final check of RNA integrity.
479 Samples were then diluted to a standardized volume (50 ng μL⁻¹) and stored at -80°C until
480 shipment to the sequencing facility.

481 RNA-seq library preparation and sequencing were performed by the Centre d'Expertise
482 et de Services Génome Québec (Montreal, QC, Canada). Before sequencing, samples were again
483 verified for integrity using both spectrophotometry (NanoDrop ND 1000; ThermoFisher,
484 Mississauga, ON, Canada) and electrophoresis (Bioanalyzer 2100; Agilent, Santa Clara, CA,
485 USA; RIN ranges: sea lamprey, 8.4–10.0; bluegill 8.5–9.8). Sequencing of cDNA libraries was

486 achieved using the NovaSeq 6000 sequencing system (Illumina, Vancouver, BC, Canada). Total
487 sequencing read counts can be found in Table S1.

488 *Transcriptome assembly, alignment, and annotation*

489 RNA-seq QC

490 Raw reads were downloaded from Genome Quebec for processing and downstream
491 analysis on a University of Manitoba personal Linux Server, as well as supercomputers
492 belonging to Westgrid (Grex) and Compute Canada (Beluga, Cedar, and Graham). Fastqc (v
493 0.11.9) was run on the raw sequencing data to assess quality, and MultiQC
494 (<https://github.com/ewels/MultiQC>) was used to generate reports from the Fastqc for easier
495 viewing across samples^{45,46}. It is important to note here that RNA-seq data files contained
496 sequencing data from fish exposed to TFM as well as those exposed to niclosamide or a mixture
497 of TFM and niclosamide. The latter two exposures are part of a complementary but ongoing
498 study and are not included in the final results here. Raw sequencing quality and duplication rates
499 met our expectations across all 148 sea lamprey samples and all 163 bluegill samples (combined
500 numbers of gill and liver tissue for each species that were obtained from the exposure series).
501 Adapters were trimmed from the raw sequences using Trimmomatic (v0.36), Fastqc and
502 MultiQC were run again to ensure successful trimming⁴⁷.

503 *De novo* transcriptome assembly

504 Since bluegill does not have a reference genome to map reads back to, *de novo*
505 transcriptome assembly was carried out for both bluegill and sea lamprey to keep methods
506 consistent⁴⁸. These *de novo* assemblies were each constructed from a list containing a single
507 library for each tissue and timepoint. These input libraries were selected based on the highest
508 read count, accounting for duplication rates for each condition. For bluegill ($n = 26$), we used
509 one paired-end library from liver and gill for each of the following treatments and timepoints: 6,
510 12, and 24 h for niclosamide ($n = 6$), TFM ($n = 6$) and TFM:niclosamide mixture ($n = 6$), as well
511 as 0, 6, 12, and 24 h for the control ($n = 8$). For lamprey ($n = 22$), paired-end libraries from liver
512 and gill for 6, 12, and 24 h for niclosamide ($n = 6$), TFM ($n = 6$), and control ($n = 6$) were used,
513 but the TFM:niclosamide mixture only had libraries at 6 and 12 h ($n = 4$), since the combination
514 killed 93% of all lamprey by 12 h of exposure (see Lawrence et al.¹⁹).

515 Each assembly was generated using Trinity (v2.8.5), run with the following command:
516 Trinity --max_memory 250G --seqType fq --samples_file samples.txt --KMER_SIZE 25 --
517 SS_lib_type RF --CPU 24 --bflyCalculateCPU --output full_trinity_assembly &>log_RF_2.txt
518 &⁴⁹. Following Trinity, the completeness of the assemblies was assessed using BUSCO (v4.1.4)
519 to test for the presence of Benchmarking Universal Single Copy Orthologs that are conserved
520 within all eukaryotes using the eukaryota_odb10 database⁵⁰. The bluegill assembly was 98.4%
521 complete with one fragmented and three missing BUSCOs out of the database of 255. The sea
522 lamprey was 96.1% complete with six fragmented and four missing BUSCOs (Table 1). These
523 results are comparable to a recently published *de novo* larval sea lamprey transcriptome
524 assembled from muscle, liver, and brain. The bluegill assembly had 559,416,397 assembled
525 bases in 498,997 transcripts with a median contig length of 413, contig N50 of 2,992, and %GC
526 of 44.38. The sea lamprey assembly had 377,365,173 assembled bases in 474,000 transcripts
527 with a median contig length of 358, contig N50 of 1,648, and %GC of 53.10.

528 De novo transcriptome annotation

529 Trinity assemblies were processed with the Trinotate (v3.2) annotation protocol
530 (<http://trinotate.github.io>), which incorporates many bioinformatics tools, namely TransDecoder
531 (v5.5.0), BLASTP (v 2.10.0), BLASTX (v2.10.0), HMMER (packaged in Trinotate v3.2),
532 TMHMM (v2.0c), SignalP (v4.1), and RNAmmer (packaged in Trinotate v3.2) to identify open
533 reading frames (ORFs) within the assembly and test these ORFs against databases of known
534 proteins, mRNA transcripts, transmembrane helices, signal peptides, and ribosomal RNA
535 sequences in order to provide candidate annotations and gene ontology (GO) terms for
536 transcripts within the assembly^{49,51–55}. Predicted annotations were used in downstream
537 differential expression analyses.

538 Super transcriptome assembly

539 Although *de novo* assemblies are often essential when reference genomes do not exist,
540 carrying out differential expression analyses on raw assemblies is complicated by the fact that
541 multiple transcripts are often generated for each gene. To combat this issue, we used the Corset-
542 Lace pipeline (<https://github.com/Oshlack/Lace>) to generate SuperTranscripts from our *de novo*
543 assemblies. First, we mapped reads back to the Trinity assembly using Bowtie2, then carried out
544 hierarchical clustering of transcripts based on shared reads and expression across transcripts,

545 finally creating an assembly of SuperTranscripts based on clustering of gene groups^{56–58}. For
546 bluegill, Corset-Lace reduced 498,997 transcripts to 109,702 SuperTranscripts, and 474,000
547 transcripts were reduced to 129,219 SuperTranscripts in sea lamprey. These
548 SuperTranscriptomes were run through BUSCO, as above, and bluegill SuperTranscriptome
549 displayed 84.7% completeness, whereas the lamprey reported 91% completeness. Although both
550 of these measures are slight decreases from the original 98.4% and 96.1% for bluegill and
551 lamprey, respectively, the biggest change in the BUSCO report was the number of complete and
552 single-copy BUSCOs compared to the number of complete and duplicated BUSCOs (Table S2).

553 Given that the genes in the BUSCO list represent Benchmarking Universal Single-Copy
554 Orthologs across eukaryotes, the high percentage of “complete duplicated BUSCOs (D)”, 58.8%
555 and 60.4% in bluegill and lamprey, respectively, indicated that our *de novo* Trinity assemblies
556 contained numerous transcripts for the same gene, which would be problematic when assessing
557 differential expression. Our SuperTranscriptomes decreased the overall transcript number in our
558 assemblies by ~3.4–4.5 times, slightly decreased the overall percentage of complete BUSCOs
559 (the sum of single-copy and duplicated BUSCOs) by ~1.05–1.15 times, but most importantly,
560 decreased the number of duplicated BUSCOs by ~17–21 times. The Corset-Lace pipeline greatly
561 increased the number of complete single-copy BUSCOs (S) for both species. Although the
562 Corset-Lace pipeline resulted in slightly fewer genes overall, we believe that those remaining
563 were better assembled, with far fewer genes present across multiple transcripts in the assembly,
564 which increased our power for differential expression analyses.

565 *Differential expression analyses*

566 Reads from bluegill and lamprey TFM experimental and control conditions were mapped to
567 the corresponding SuperTranscriptome using STAR (v2.6.1a)⁵⁹. The resulting BAM files from
568 STAR were used to generate transcript counts using featureCounts (part of subread v2.0.0)⁶⁰ for
569 both exon and transcript-based analyses.

570 Transcriptomic analysis was performed in R studio (v1.3.1093) using the R programming
571 language (v3.5.1)⁶¹. Principal analyses of differential expression patterns were performed using
572 the package ‘edgeR’ (v3.24.3)^{62,63} using a quasi-likelihood pipeline⁶⁴. Count data were first
573 filtered to remove lowly expressed genes with the “filterByExpr” function which used the
574 experimental design matrix to determine minimum gene count thresholds. Filtered count data

575 were then normalized using the trimmed mean of M-values⁶⁵. At this point, normalized data were
576 visually inspected using principal component analysis (PCA; packages ‘FactoMineR’⁶⁶ and
577 ‘factoextra’⁶⁷; see Fig. S1) and a multi-dimensional scaling (MDS) plot (package ‘Glimma’⁶⁸).
578 Individual samples were also visually inspected using a mean difference (MD) plot.

579 To determine patterns of differential expression, count data were first modeled against a
580 negative binomial distribution⁶³. From this, we obtained dispersal estimates using a Cox-Reid
581 profile adjusted likelihood method⁶³, which were visually inspected by use of a biological
582 coefficient of variation (BCV) plot. Quasi-likelihood (QL) dispersion estimates were determined
583 and were subject to a QL F-test to determine transcripts that had statistically significant
584 differential gene expression⁶⁹. From the resulting QL F-tests, we pulled out pairwise contrasts of
585 interest which include control versus TFM exposed fish for each timepoint of exposure (i.e., 6,
586 12, 24 h). Statistical significance of pairwise comparison was accepted at $\alpha = 0.05$ where all p -
587 values were adjusted for false discovery rates (FDR) via a Benjamini-Hochberg correction⁷⁰.

588 *Enrichment analyses*

589 One of our primary goals was to ascertain specific pathways and processes that were
590 underlying the transcriptome response to TFM exposure. To do so, we conducted analyses in
591 pathway enrichment using the R package ‘enrichR’ (v2.1)⁷¹, which allows R access to the
592 Enrichr databases (<https://maayanlab.cloud/Enrichr/>)⁷². This platform compares differentially
593 expressed genes in a dataset and determines a list of GO terms that are enriched under the
594 experimental treatment⁷². Here, we used three main databases for comparison:
595 GO_Biological_Process_2018, GO_Molecular_Function_2018, and
596 GO_Cellular_Component_2018. Before making database comparisons, unannotated
597 superTranscript clusters were removed from the analysis. For annotated transcripts, UniProt gene
598 names were retrieved using the UniProt’s Retrieve/ID mapping tool
599 (<https://www.uniprot.org/uploadlists/>). In each transcripts list enrichR was run and GO term lists
600 were subsequently filtered for those that were statistically significant (adjusted p -value < 0.05)
601 and had four or more genes associated with a particular GO term⁷³.

602 To identify general patterns and biological features of interest, we simplified the resulting
603 enrichR GO term lists using a web-based platform, REVIGO (<http://revigo.irb.hr/>)⁷⁴, which
604 summarizes GO term patterns by grouping terms based on their similarity and then creates a

605 hierarchical structure of the terms⁷⁴. Our REVIGO analyses were at a similarity level of 0.5 (i.e.,
606 small) and used the adjusted *p*-values generated from enrichR. Treemap files from the REVIGO
607 output were used to generate GO term enrichment plots in ggplot2⁷⁵. To reduce term
608 redundancy, we opted to only use biological processes and molecular functions from the
609 REVIGO analysis. In cases where the biological processes and molecular functions provided
610 functionally similar results, only the biological processes were displayed in the Results section
611 with the molecular functions being available in the Supplementary Materials. Similarly,
612 summary plots that offered no insight on specific processes (e.g., DNA replication, protein
613 kinase activity, GTPase activities, etc.) were placed into Supplementary Materials.

614 *Identification of differentially expressed detoxification transcripts*

615 We next sought to identify specific mechanisms underlying TFM detoxification and
616 understand species-specific differences in detoxification ability (see Lawrence et al.¹⁹ for a
617 review). To achieve these goals, we also incorporated a targeted approach, which required us to
618 identify known and suspected candidates for TFM detoxification, including *ugt*'s, one of the
619 principal enzymes believed to be involved in Phase II TFM detoxification (see Wilkie et al.³), as
620 well as other enzymes/proteins involved with Phases I, II, and III of detoxification. In the case of
621 *ugt* transcripts, *ugt*-associated transcripts were first identified in each species' annotated
622 transcriptome. These transcripts were then used to filter each list of species- and tissue-specific
623 differentially expressed transcripts to examine how these enzymes responded to TFM exposure.

624 In the case of the Phase I–III proteins, we opted to filter a broad list of common protein
625 families involved in detoxification from the differentially expressed transcripts lists and then
626 manually check their functional role based on their UniProt IDs. As discussed in the Results, we
627 filtered transcripts in the transcriptome associated with detoxification, xenobiotic removal,
628 and/or organic compound breakdown. In the case of Phase III transcripts, we also included
629 transporters that were involved in organic ion transport and those transporting UDP/glucuronide
630 substrates. Phase I search terms included cytochromes P450 (CYPs), alcohol dehydrogenase
631 (ADH), aldehyde dehydrogenase (ALDH), monoamine oxidases (MAO), and paraoxonases
632 (PON). Phase II protein search terms included sulfotransferases (SULT), glutathione transferases
633 (GST), glycine N-methyltransferase (GLYAT), N-terminal acetyltransferases (NAT), and

634 methyltransferases (MT). Phase III search terms included solute carrier's (SLC) and ATP-
635 binding cassette (ABC) transporters.

636 Code and Data Availability

637 All commands and scripts used for quality control (QC), assembly, annotation, read
638 mapping, transcript counting and RNAseq analyses are available at [https://github.com/phil-
639 grayson/transcriptome_lawrence_jeffries/](https://github.com/phil-grayson/transcriptome_lawrence_jeffries/). The raw sequence reads are available at the National
640 Center for Biotechnology Information Sequence Read Archive (accession number
641 SUB8714632). Gene lists and GO term list can be found in the Supplementary Materials.

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649 collecting the larval lamprey used in this study.

650 **Contributions**

651 KMJ, MPW, RGM, JMW, CJG, and MFD were responsible for the experimental design. The
652 labs of KMJ and MPW carried out the exposures and sample collection. Data and sample
653 analyses were conducted by MJL, PG, and JDJ. The first draft of the manuscript was written by
654 MJL and PG, with all authors contributing to the refinement and production of the final
655 manuscript.

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Table 1: Differentially expressed transcripts associated with phase I detoxification for sea lamprey larvae (*Petromyzon marinus*) and bluegill (*Lepomis macrochirus*) following 6, 12, and 24 h TFM (3-trifluoromethyl-4'-nitrophenol) exposure.

Species	Tissue	Exposure Duration (h)	Cluster ID	UniProt Gene Name	FDR	Fold Change
<i>Sea Lamprey</i>	Gill	6	Cluster_41283.36015	<i>cyp1a1</i>	0.041	5.06
	Gill	6	Cluster_41283.41047	<i>cyp1a1</i>	0.049	1.65
	Gill	24	Cluster_41283.1293	<i>cyp1b1</i>	0.002	2.57
	Gill	24	Cluster_41283.2130	<i>cyp1a5</i>	<0.001	6.95
	Gill	24	Cluster_41283.54788	<i>cyp3a21</i>	0.003	1.46
	Gill	24	Cluster_41283.54890	<i>pon1</i>	0.033	-1.19
	Gill	24	Cluster_41283.8821	<i>cyp2j2</i>	<0.001	-2.75
<i>Bluegill</i>	Gill	24	Cluster_41283.60437	<i>cyp2d3</i>	0.001	-2.01
	Gill	6	Cluster_20758.1	<i>cyp1b1</i>	<0.001	11.57
	Gill	6	Cluster_49386.0	<i>cyp1b1</i>	<0.001	16.23
	Gill	6	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	108.31
	Gill	12	Cluster_20758.1	<i>cyp1b1</i>	<0.001	12.17
	Gill	12	Cluster_49386.0	<i>cyp1b1</i>	<0.001	18.85
	Gill	12	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	117.16
	Gill	24	Cluster_11231.2	<i>mao</i>	0.011	1.347
	Gill	24	Cluster_20758.1	<i>cyp1b1</i>	<0.001	18.37
	Gill	24	Cluster_49386.0	<i>cyp1b1</i>	<0.001	40.30
	Gill	24	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	186.76
	Liver	6	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	6.42
	Liver	12	Cluster_49386.0	<i>cyp1b1</i>	<0.001	12.93
Liver	12	Cluster_51425.10202	<i>cyp2c31</i>	0.012	1.75	
Liver	12	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	12.80	

Liver	24	Cluster_20758.1	<i>cyp1b1</i>	0.044	2.14
Liver	24	Cluster_49386.0	<i>cyp1b1</i>	<0.001	12.36
Liver	24	Cluster_51425.10202	<i>cyp2c31</i>	0.003	1.78
Liver	24	Cluster_51425.1737	<i>cyp1a1</i>	<0.001	11.63
Liver	24	Cluster_51425.22684	<i>cyp2k1</i>	<0.001	2.46
Liver	24	Cluster_51425.24180	<i>cyp2k1</i>	0.027	2.56
Liver	24	Cluster_51425.30529	<i>cyp3a40</i>	<0.001	2.19
Liver	24	Cluster_51425.31208	<i>cyp2k1</i>	0.005	1.71
Liver	24	Cluster_51425.36982	<i>cyp2k1</i>	<0.001	2.18
Liver	24	Cluster_51425.6113	<i>cyp2k6</i>	0.020	1.68
Liver	24	Cluster_51425.29803	<i>cyp2b2</i>	0.013	-3.08
Liver	24	Cluster_51425.38557	<i>cyp2k1</i>	0.008	-1.66
Liver	24	Cluster_51425.7362	<i>cyp2k6</i>	0.010	-1.70
Liver	24	Cluster_51425.38638	<i>cyp2b4</i>	0.002	2.20

Transcripts were selected using UniProt descriptions where the enzyme was noted to be involved in xenobiotic and/or drug detoxification processes. FDR represents the false discovery rate of the pairwise contrasts and the fold change is against respective controls for a specific timepoint. A positive and negative fold change represents an upregulation or downregulation of the transcript relative to the control. Abbreviations for gene names are as follows: cytochrome P450 (*cyp*), monoamine oxidase (*mao*), and paraoxonase (*pon*).

Table 2: Identification of all UDP-glucuronosyltransferase (*ugt*) transcripts in the annotated transcriptomes of sea lamprey larvae (*Petromyzon marinus*) and bluegill (*Lepomis macrochirus*).

Species	Cluster ID	UniProt Gene Name	Uniprot Entry ID	Expect Value (E)	Percent ID (%)	Protein Name
<i>Sea lamprey</i>						
	Cluster_41283.10908	<i>ugt2c1</i>	P36514	2.93x10 ⁻²⁴	62.50	UDP-glucuronosyltransferase 2c1
	Cluster_41283.12396	<i>ugt2a1</i>	P0DTE4	1.45x10 ⁻⁸⁶	36.96	UDP-glucuronosyltransferase 2a1
	Cluster_41283.4253	<i>ugt2a2</i>	Q6PDD0	2.35x10 ⁻⁵⁷	40.35	UDP-glucuronosyltransferase 2a2
	Cluster_41283.48705	<i>ugt2a2</i>	Q6PDD0	1.13x10 ⁻²⁴	56.99	UDP-glucuronosyltransferase 2a2
	Cluster_41283.5366	<i>ugt2b1</i>	P09875	3.84x10 ⁻¹⁴	63.27	UDP-glucuronosyltransferase 2b1
	Cluster_41283.69509	<i>ugt2a2</i>	Q6PDD0	0.00	51.15	UDP-glucuronosyltransferase 2a2
<i>Bluegill</i>						
	Cluster_21009.0	<i>ugt3</i>	Q91280	1.22x10 ⁻²⁸	41.61	UDP-glucuronosyltransferase
	Cluster_3692.0	<i>ugt3a1</i>	Q3UP75	3.36x10 ⁻⁸⁵	46.47	UDP-glucuronosyltransferase 3a1
	Cluster_51425.15463	<i>ugt2a1</i>	P0DTE4	1.56x10 ⁻⁶⁸	36.47	UDP-glucuronosyltransferase 2a1
	Cluster_51425.15468	<i>ugt2a1</i>	P0DTE4	1.67x10 ⁻⁶³	59.87	UDP-glucuronosyltransferase 2a1
	Cluster_51425.15474	<i>ugt2a1</i>	Q80X89	1.26x10 ⁻¹¹⁰	39.82	UDP-glucuronosyltransferase 2a1
	Cluster_51425.15477	<i>ugt2b20</i>	O77649	3.76x10 ⁻⁵³	53.57	UDP-glucuronosyltransferase 2b20
	Cluster_51425.15481	<i>ugt2a1</i>	P0DTE4	1.40x10 ⁻⁶²	59.63	UDP-glucuronosyltransferase 2a1
	Cluster_51425.17384	<i>ugt3</i>	Q91280	0.00	67.26	UDP-glucuronosyltransferase
	Cluster_51425.25248	<i>ugt2b9</i>	O02663	2.49x10 ⁻⁶	96.00	UDP-glucuronosyltransferase 2b9
	Cluster_51425.26320	<i>ugt2a1</i>	P0DTE4	1.15x10 ⁻¹²⁶	38.21	UDP-glucuronosyltransferase 2a1
	Cluster_51425.32583	<i>ugt1a5</i>	Q64638	1.09x10 ⁻⁷	50.00	UDP-glucuronosyltransferase 1a5
	Cluster_51425.33579	<i>ugt2a1</i>	P36510	0.00	58.26	UDP-glucuronosyltransferase 2a1

Cluster_51425.33985	<i>ugt2b9</i>	O02663	1.35×10^{-14}	39.13	UDP-glucuronosyltransferase 2b9
Cluster_51425.36243	<i>ugt1a2</i>	P20720	2.74×10^{-124}	42.06	UDP-glucuronosyltransferase 1-2
Cluster_51425.37011	<i>ugt2a1</i>	Q80X89	4.91×10^{-132}	42.56	UDP-glucuronosyltransferase 2a1
Cluster_51425.38640	<i>ugt2b20</i>	O77649	1.39×10^{-9}	58.49	UDP-glucuronosyltransferase 2b20
Cluster_51425.5745	<i>ugt2a1</i>	PODTE4	1.09×10^{-56}	64.44	UDP-glucuronosyltransferase 2a1

Table 3: Summary of the differentially expressed UDP-glucuronosyltransferase (UGT) transcripts for bluegill (*Lepomis macrochirus*) following 6, 12, and 24 h TFM (3-trifluoromethyl-4'-nitrophenol) exposure.

Tissue	Exposure Duration (h)	Cluster ID	UniProt Gene Name	FDR	Fold Change
Gill	6	Cluster_51425.17384	<i>ugt3</i>	<0.001	5.69
Gill	12	Cluster_51425.17384	<i>ugt3</i>	<0.001	11.22
Gill	24	Cluster_51425.17384	<i>ugt3</i>	<0.001	15.88
Liver	12	Cluster_51425.17384	<i>ugt3</i>	<0.001	13.06
Liver	12	Cluster_51425.33985	<i>ugt2b9</i>	0.003	1.79
Liver	12	Cluster_51425.36243	<i>ugt1a2</i>	0.004	-1.90
Liver	24	Cluster_21009.0	<i>ugt3</i>	0.024	1.56
Liver	24	Cluster_51425.17384	<i>ugt3</i>	<0.001	28.14
Liver	24	Cluster_51425.33985	<i>ugt2b9</i>	<0.001	2.67
Liver	24	Cluster_51425.15468	<i>ugt2a1</i>	<0.001	-7.19
Liver	24	Cluster_51425.15481	<i>ugt2a1</i>	0.001	-3.14
Liver	24	Cluster_51425.32583	<i>ugt1a5</i>	0.038	-1.53
Liver	24	Cluster_51425.36243	<i>ugt1a2</i>	<0.001	-2.55

FDR represents the false discovery rate of the pairwise contrasts and the fold change is against respective controls for a specific timepoint. A positive and negative fold change represents an upregulation or downregulation of the transcript relative to the control.

Table 4: Differentially expressed transcripts associated with phase II detoxification for sea lamprey larvae (*Petromyzon marinus*) and bluegill (*Lepomis macrochirus*) following 6, 12, and 24 h TFM (3-trifluoromethyl-4'-nitrophenol) exposure.

Species	Tissue	Exposure Duration (h)	Cluster ID	UniProt Gene Name	FDR	Fold Change
<i>Sea Lamprey</i>						
	Gill	24	Cluster_41283.24751	<i>gstt3</i>	<0.001	2.26
	Gill	24	Cluster_41283.69009	<i>sult6b1</i>	<0.001	2.29
	Liver	24	Cluster_41283.18986	<i>nnmt</i>	0.026	-1.76
<i>Bluegill</i>						
	Gill	6	Cluster_51425.33237	<i>sult6b1</i>	0.006	1.54
	Gill	12	Cluster_3656.0	<i>nat8</i>	0.005	2.07
	Gill	12	Cluster_51425.33237	<i>sult6b1</i>	<0.001	2.19
	Gill	12	Cluster_52314.5	<i>tpmt</i>	0.032	-1.30
	Gill	24	Cluster_51425.10275	<i>sult1st2</i>	0.025	1.29
	Gill	24	Cluster_51425.33237	<i>sult6b1</i>	<0.001	3.81
	Gill	24	Cluster_36463.1	<i>nat2</i>	0.026	-1.57
	Liver	24	Cluster_36463.1	<i>nat2</i>	0.022	1.70
	Liver	24	Cluster_3656.0	<i>nat8</i>	<0.001	2.34
	Liver	24	Cluster_51425.35293	<i>gstt3</i>	0.048	-1.37
	Liver	24	Cluster_51425.6836	<i>sult1c1</i>	<0.001	-1.63
	Liver	24	Cluster_52314.5	<i>tpmt</i>	<0.001	-1.81
	Liver	24	Cluster_7193.0	<i>gstt1</i>	0.019	-1.70

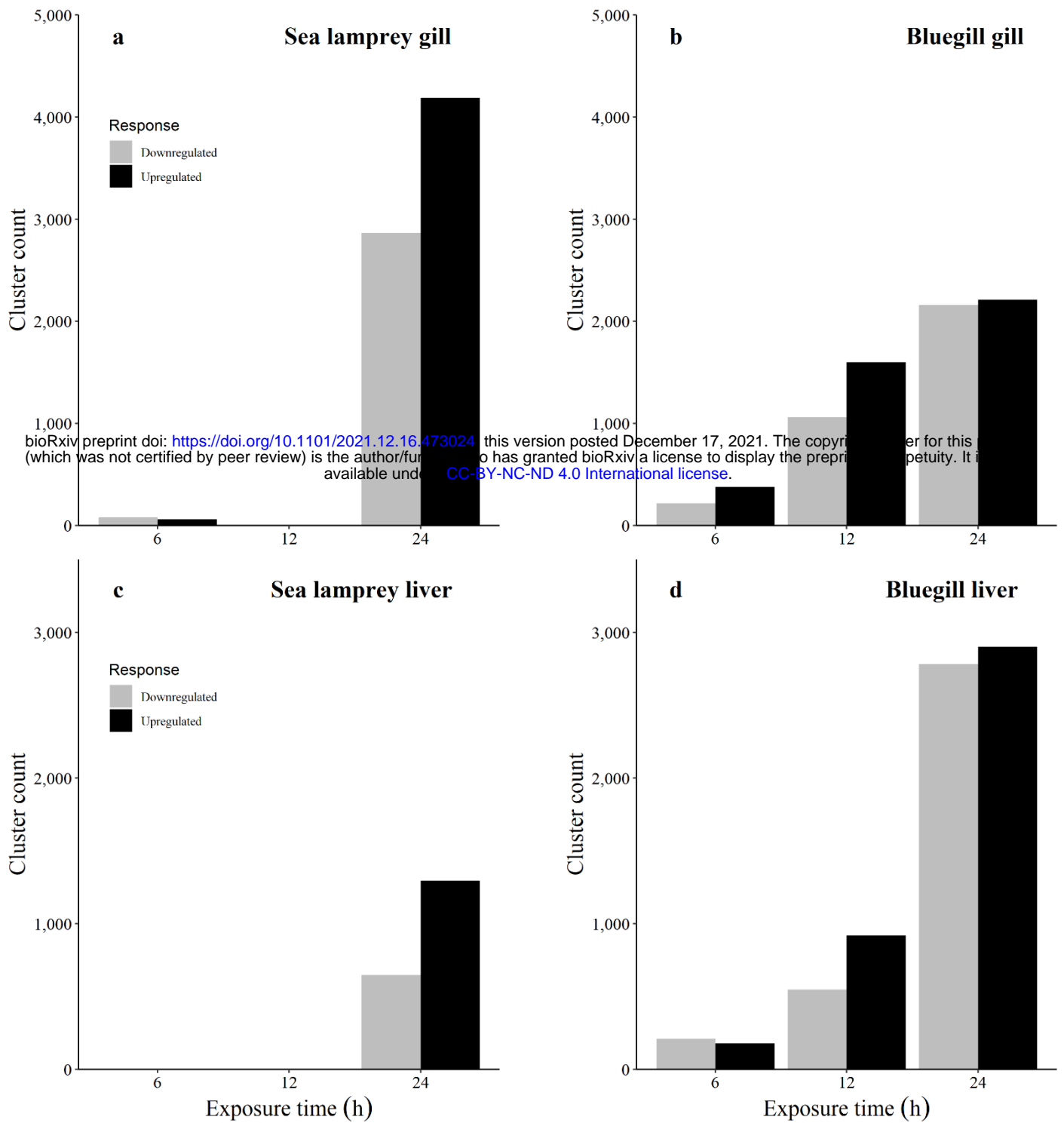
Transcripts were selected using UniProt descriptions where the enzyme was noted to be involved in xenobiotic and/or drug detoxification processes. Transcripts excluded UDP-glucuronosyltransferases (see Table 3). FDR represents the false discovery rate of the pairwise contrasts and the fold change is against respective controls for a specific timepoint. A positive and negative fold change represents an upregulation or downregulation of the transcript relative to the control. Abbreviations for gene names are as follows: Glutathione S-transferase (*gstt*), sulfotransferase (*sult*), N-acetyltransferase (*nat*), nicotinamide N-methyltransferase (*nnmt*), and thiopurine S-methyltransferase (*tpmt*).

Table 5: Differentially expressed transcripts associated with phase III detoxification for sea lamprey larvae (*Petromyzon marinus*) and bluegill (*Lepomis macrochirus*) following 6, 12, and 24 h TFM (3-trifluoromethyl-4'-nitrophenol) exposure.

Species	Tissue	Exposure Duration (h)	Cluster ID	UniProt Gene Name	FDR	Fold Change
<i>Sea Lamprey</i>						
	Gill	24	Cluster_34003.4	<i>abcc3</i>	0.008	1.55
	Gill	24	Cluster_41283.10772	<i>abcc5</i>	<0.001	1.95
	Gill	24	Cluster_41283.22177	<i>abcb1</i>	<0.001	5.56
	Gill	24	Cluster_41283.58441	<i>abcb11</i>	<0.001	4.31
	Gill	24	Cluster_41283.61313	<i>abcb1</i>	<0.001	4.95
	Gill	24	Cluster_41283.62778	<i>slc7a5</i>	<0.001	1.73
	Gill	24	Cluster_41283.9558	<i>abcg2</i>	<0.001	2.90
	Gill	24	Cluster_41283.1906	<i>slc35d2</i>	<0.001	-1.50
	Gill	24	Cluster_41283.24514	<i>slc22a15b</i>	0.014	-1.38
	Liver	24	Cluster_41283.10772	<i>abcc5</i>	0.017	2.34
	Liver	24	Cluster_41283.69161	<i>slc35a3</i>	0.016	-2.10
<i>Bluegill</i>						
	Gill	6	Cluster_49443.0	<i>slco3a1</i>	0.012	-1.30
	Gill	12	Cluster_50801.0	<i>slc29a3</i>	0.011	1.23
	Gill	12	Cluster_51425.13125	<i>slco2b1</i>	0.033	1.31
	Gill	12	Cluster_51425.24527	<i>slc35d1</i>	0.046	1.19
	Gill	12	Cluster_51425.35543	<i>slc7a8</i>	0.046	1.32
	Gill	12	Cluster_51425.4852	<i>slc35a3</i>	0.002	1.32
	Gill	12	Cluster_47743.0	<i>slc35d2</i>	0.034	-1.25
	Gill	12	Cluster_49443.0	<i>slco3a1</i>	0.004	-1.24
	Gill	24	Cluster_1028.4	<i>abca3</i>	<0.001	1.65
	Gill	24	Cluster_10382.9	<i>abcc4</i>	0.039	1.23
	Gill	24	Cluster_38275.4	<i>slc22a15b</i>	0.032	1.16

Gill	24	Cluster_50801.0	<i>slc29a3</i>	<0.001	1.49
Gill	24	Cluster_51425.13125	<i>slco2b1</i>	0.007	1.46
Gill	24	Cluster_51425.3133	<i>slc17a5</i>	0.006	1.21
Gill	24	Cluster_51425.4852	<i>slc35a3</i>	<0.001	1.68
Gill	24	Cluster_52136.2	<i>abcb6</i>	<0.001	1.44
Gill	24	Cluster_47743.0	<i>slc35d2</i>	0.001	-1.47
Gill	24	Cluster_49443.0	<i>slco3a1</i>	0.026	-1.22
Gill	24	Cluster_51425.1412	<i>abcb1</i>	0.022	-1.70
Liver	12	Cluster_43105.2	<i>slc17a5</i>	0.005	1.44
Liver	12	Cluster_51425.33972	<i>slc7a5</i>	0.027	1.58
Liver	12	Cluster_51425.35543	<i>slc7a8</i>	0.014	1.83
Liver	12	Cluster_52136.2	<i>abcb6</i>	0.001	2.66
Liver	24	Cluster_38275.4	<i>slc22a15b</i>	<0.001	1.56
Liver	24	Cluster_43105.2	<i>slc17a5</i>	<0.001	1.98
Liver	24	Cluster_51425.3133	<i>slc17a5</i>	0.022	1.21
Liver	24	Cluster_51425.33972	<i>slc7a5</i>	<0.001	1.88
Liver	24	Cluster_51425.4852	<i>slc35a3</i>	0.037	1.52
Liver	24	Cluster_52136.2	<i>abcb6</i>	<0.001	2.70
Liver	24	Cluster_17571.3	<i>slc47a1</i>	0.002	-1.70
Liver	24	Cluster_51425.13125	<i>slco2b1</i>	0.002	-1.87

Transcripts were selected using UniProt descriptions where the enzyme was noted to be involved in xenobiotic and/or drug detoxification processes as well as those capable of transporting conjugate molecules for glucuronidation. FDR represents the false discovery rate of the pairwise contrasts and the fold change is against respective controls for a specific timepoint. A positive and negative fold change represents an upregulation or downregulation of the transcript relative to the control. Abbreviations for gene names are as follows: ATP-binding cassette transporter (*abc*) and solute carrier (*slc*).



24

25 **Figure 1:** Total counts of differentially expressed superTranscriptome clusters in the gills (a,b)
 26 and liver (c,d) of sea lamprey larvae and bluegill following a 24 h TFM exposure. Grey bars
 27 represent differentially expressed clusters that were downregulated whereas black bars represent
 28 upregulated clusters. All clusters were deemed to be significant at a false discovery rate (FDR) of
 29 $\alpha = 0.05$.

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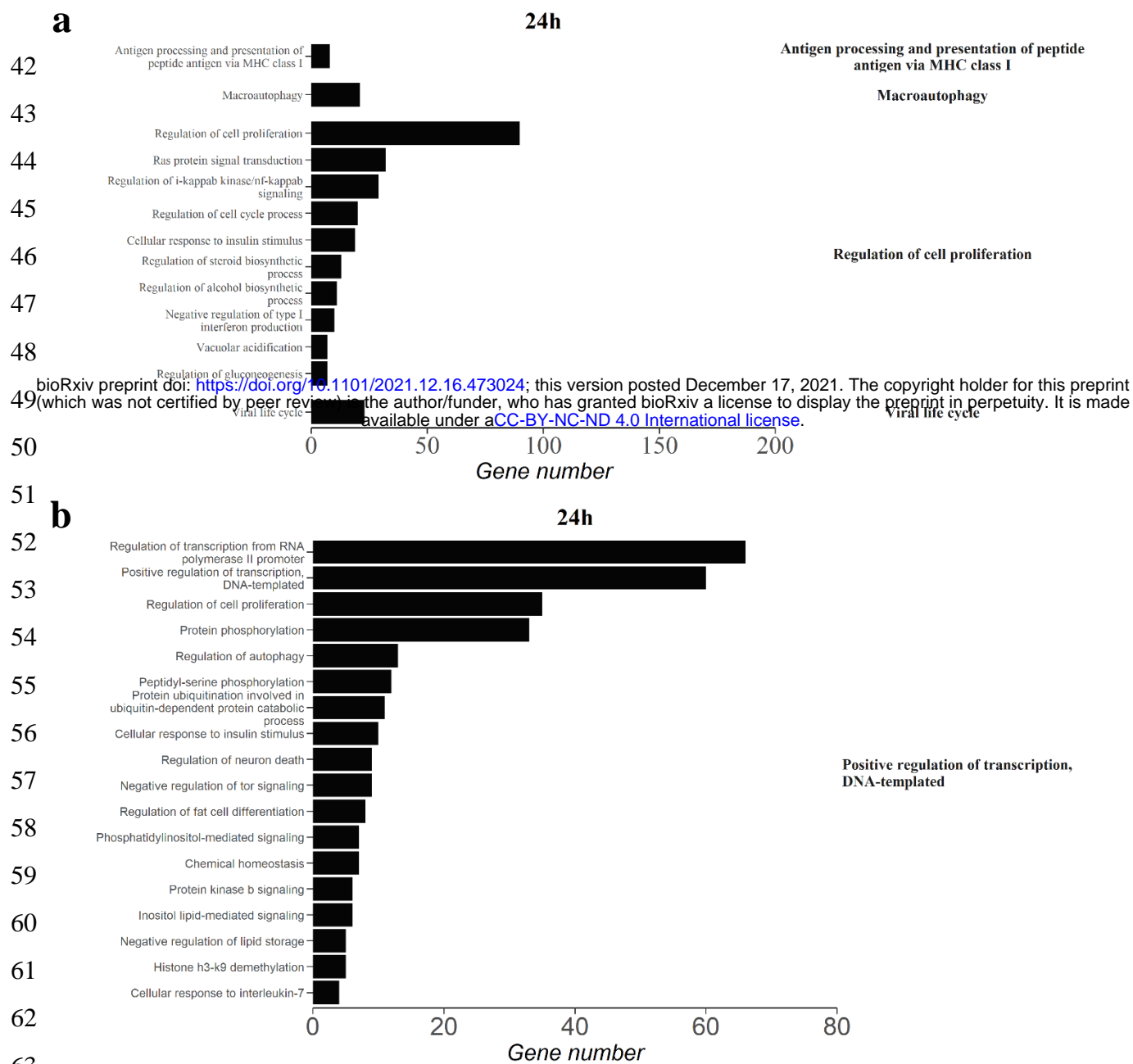
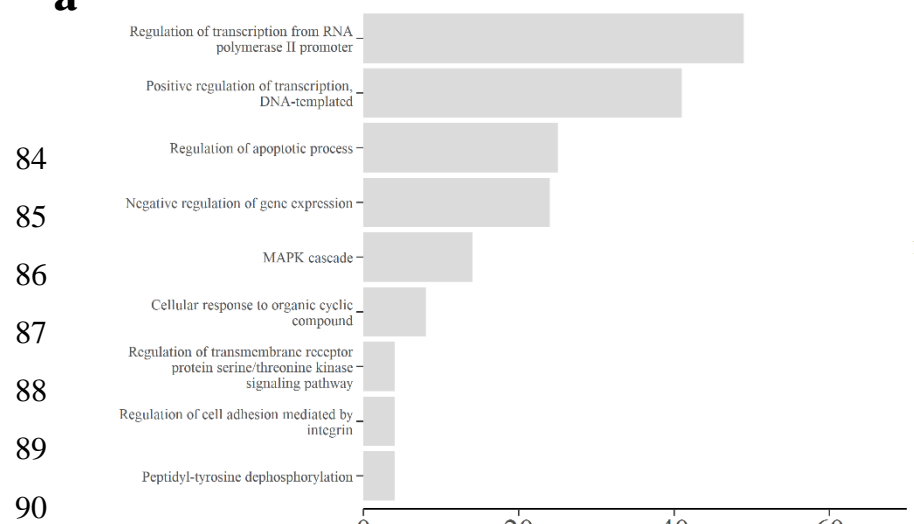


Figure 2: Summary of enriched gene ontology (GO) terms with transcripts that were upregulated following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 2.21 mg L⁻¹ nominal) exposure in the gills (a) and liver (b) of sea lamprey (*Petromyzon marinus*) larvae. Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).

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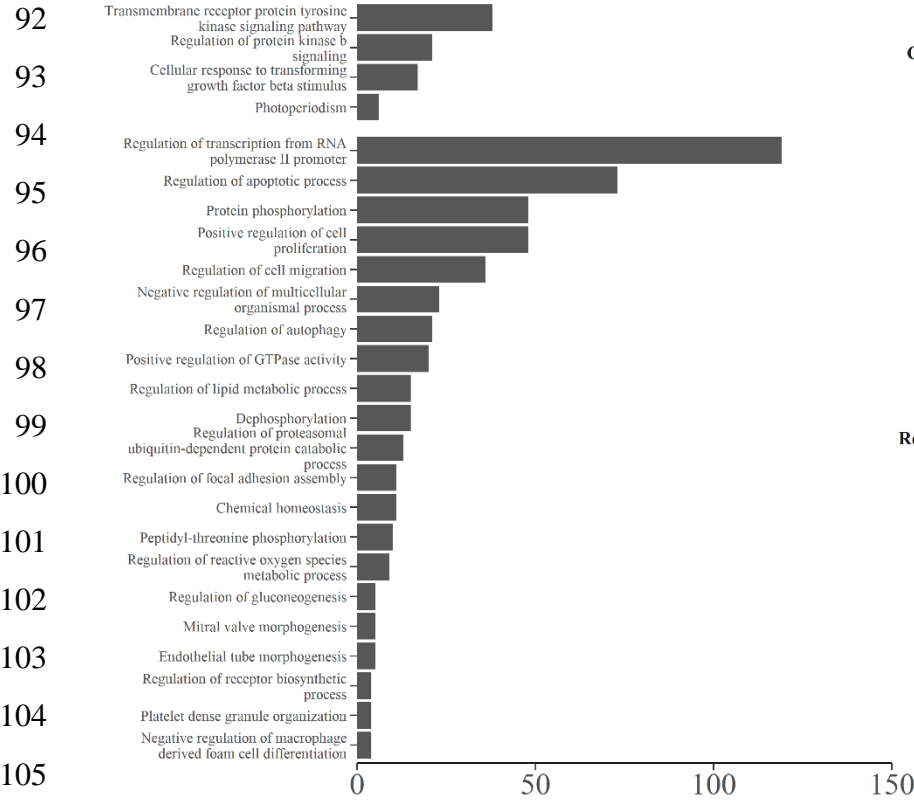
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Regulation of transcription from RNA polymerase II promoter

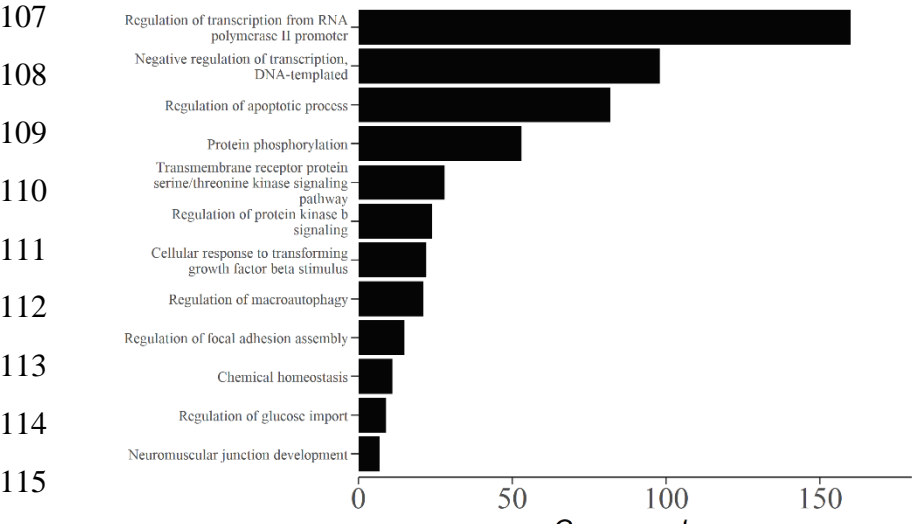
bioRxiv preprint doi: <https://doi.org/10.1101/2021.12.16.473024>; this version posted December 17, 2021. The copyright holder for this preprint (which was not certified by peer review) is the author/funder, who has granted bioRxiv a license to display the preprint in perpetuity. It is made available under aCC-BY-NC-ND 4.0 International license.



Cellular response to transforming growth factor beta stimulus

Regulation of transcription from RNA polymerase II promoter

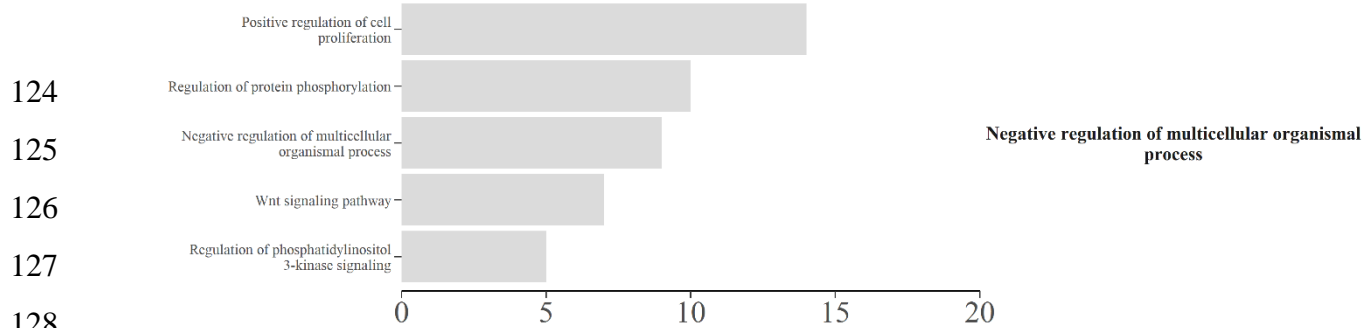
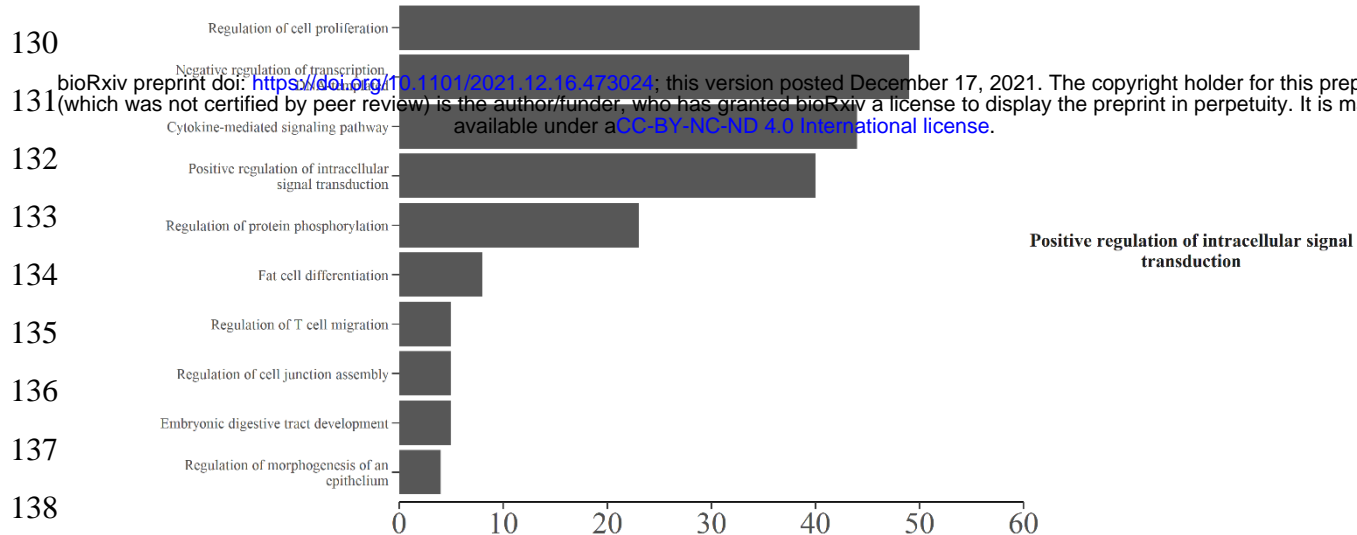
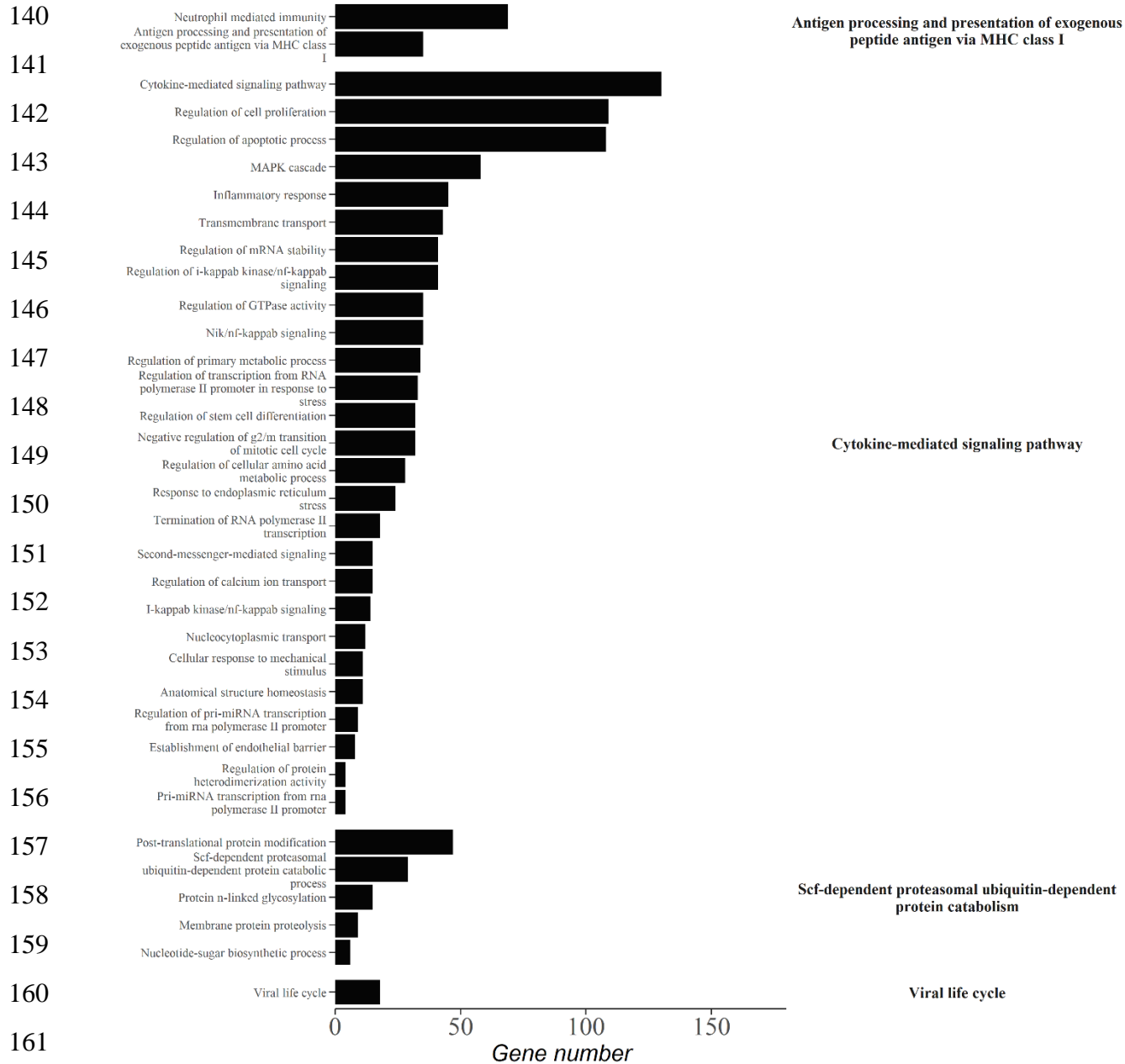
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Regulation of transcription from RNA polymerase II promoter

Gene number

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b**6h****12h****24h**

Gene number

164 **Figure 3:** Summary of enriched gene ontology (GO) terms with transcripts that were upregulated
165 (a) and downregulated (b) following 6 h (light grey), 12 h (dark grey), and 24 h (black) of TFM
166 (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of bluegill
167 (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery
168 rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least
169 four transcripts were considered significantly enriched. REVIGO was used to summarize GO
170 terms to reduce redundancy and group according to similarity (right labels).

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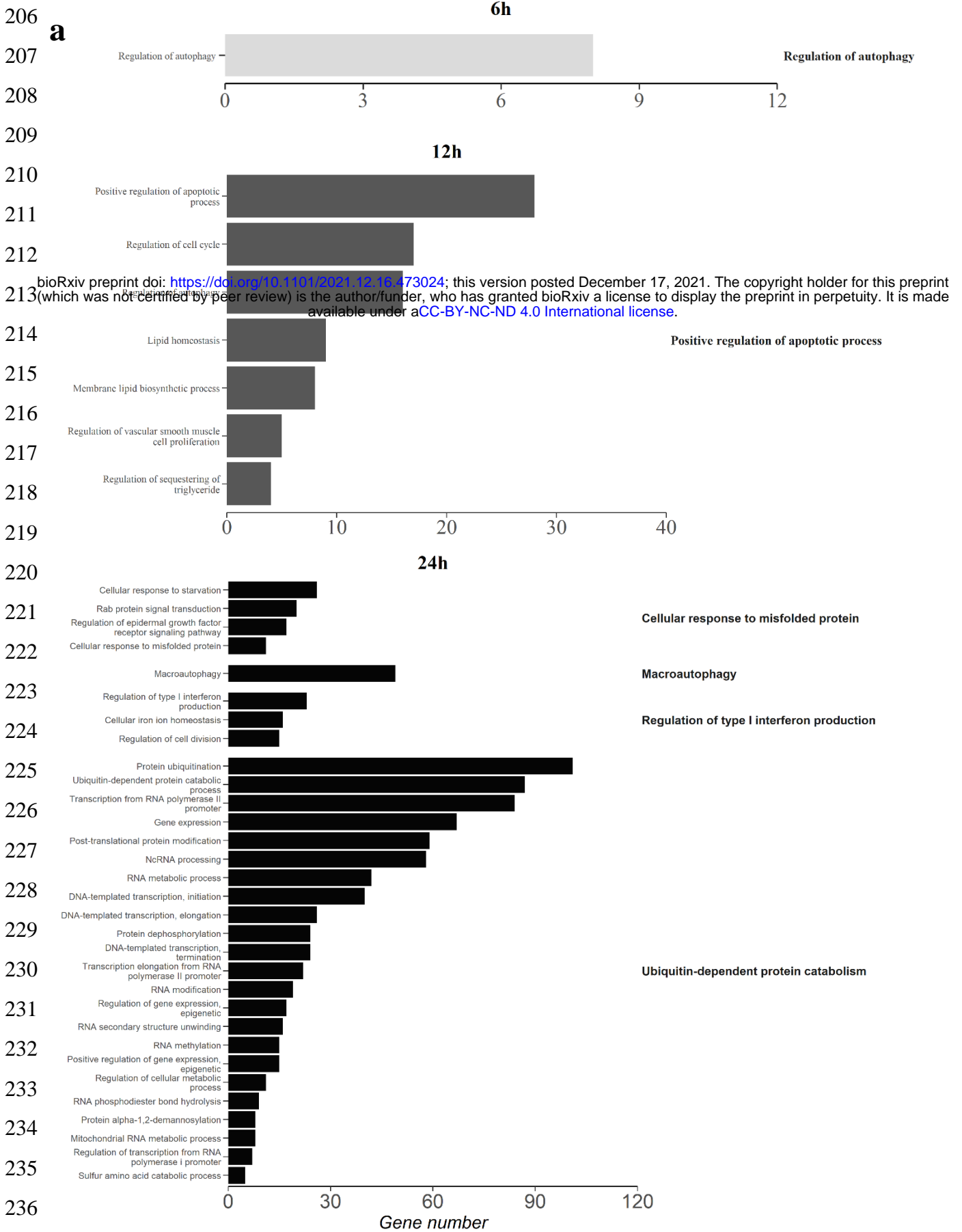
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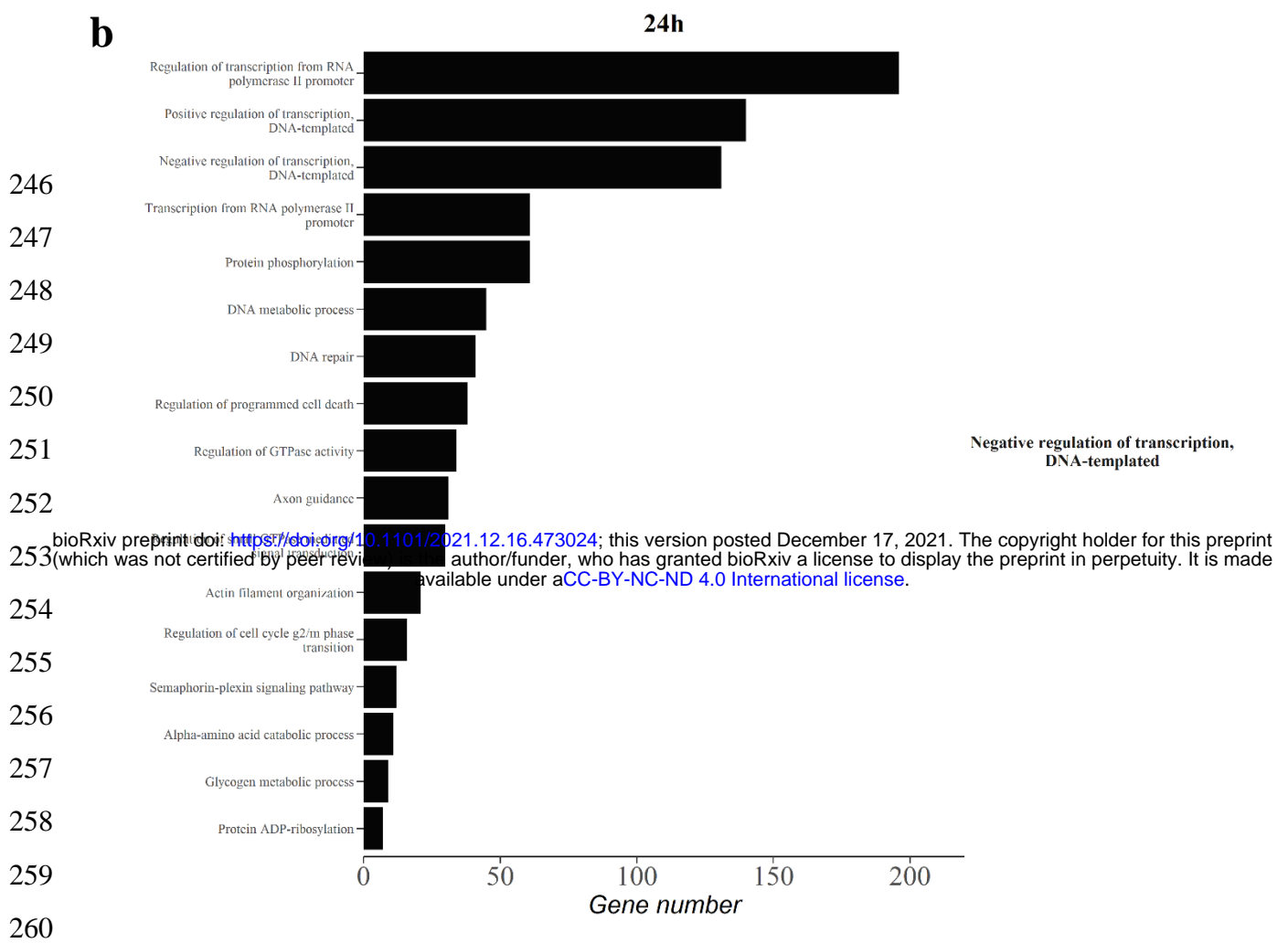
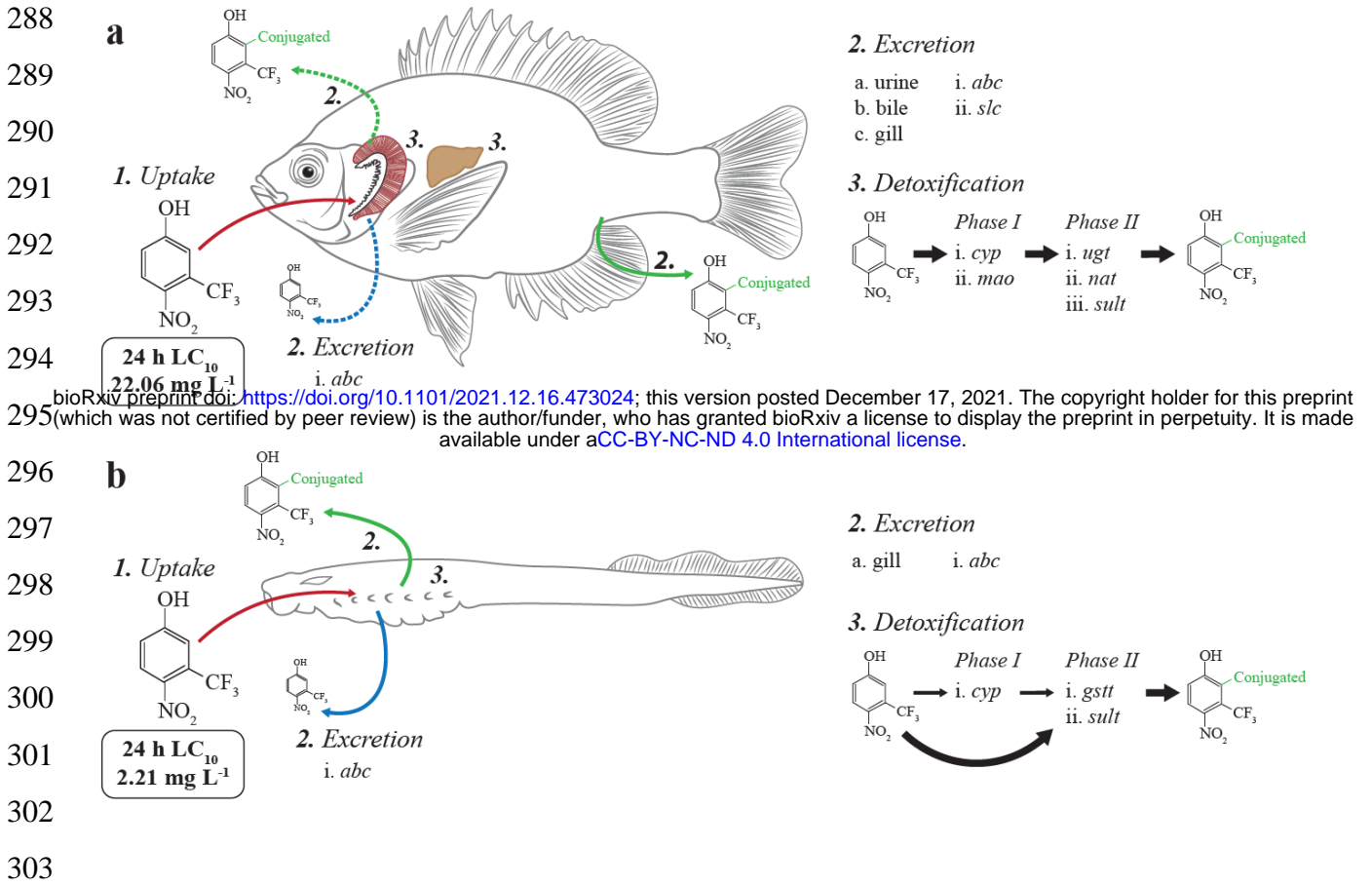


Figure 4: Summary of enriched gene ontology (GO) terms with transcripts that were upregulated (a) and downregulated (b) following 6 h (light grey), 12 h (dark grey) and 24 h (black) of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the liver of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).



304 **Figure 5:** Overview of the pathways of TFM (3-trifluoromethyl-4'-nitrophenol) uptake,
305 excretion, biotransformation, and elimination in bluegill (*Lepomis macrochirus*; **a**) and sea
306 lamprey larvae (*Petromyzon marinus*; **b**). Detoxification transcripts that are presented here were
307 found to have higher differential expression following exposure to TFM (bluegill = 22.06 mg L⁻¹
308 ¹, lamprey = 2.21 mg L⁻¹) over a 24 h period.

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2 **Table S1:** Total sequencing reads counts by species and tissue.

Species	Tissue	Total Reads
Sea lamprey	Gill	1.73x10 ⁹
	Liver	1.19x10 ⁹
Bluegill	Gill	1.59x10 ⁹
	Liver	1.81x10 ⁹

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25 **Table S2.** BUSCO reports comparing the initial Trinity assembly to the Corset-Lace assembly

Species	Assembly	Complete BUSCOs (C) = S+D	Complete Single-Copy BUSCOs (S)	Complete duplicated BUSCOs (D)	Fragmented BUSCOs (F)	Missing BUSCOs (M)
<i>Bluegill</i>	Trinity	98.4%	39.6%	58.8%	0.4%	1.2%
	Corset/Lace	84.7%	82.0%	2.7%	9.4%	5.9%
<i>Sea lamprey</i>	Trinity	96.1%	35.7%	60.4%	2.4%	1.5%
	Corset/Lace	91%	87.5%	3.5%	6.3%	2.7%

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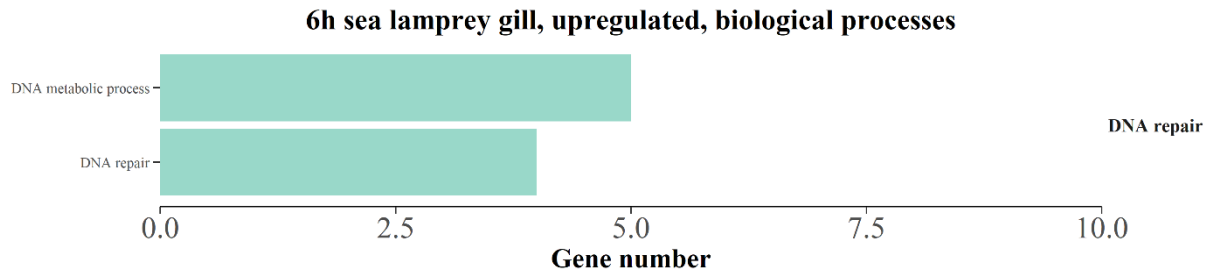
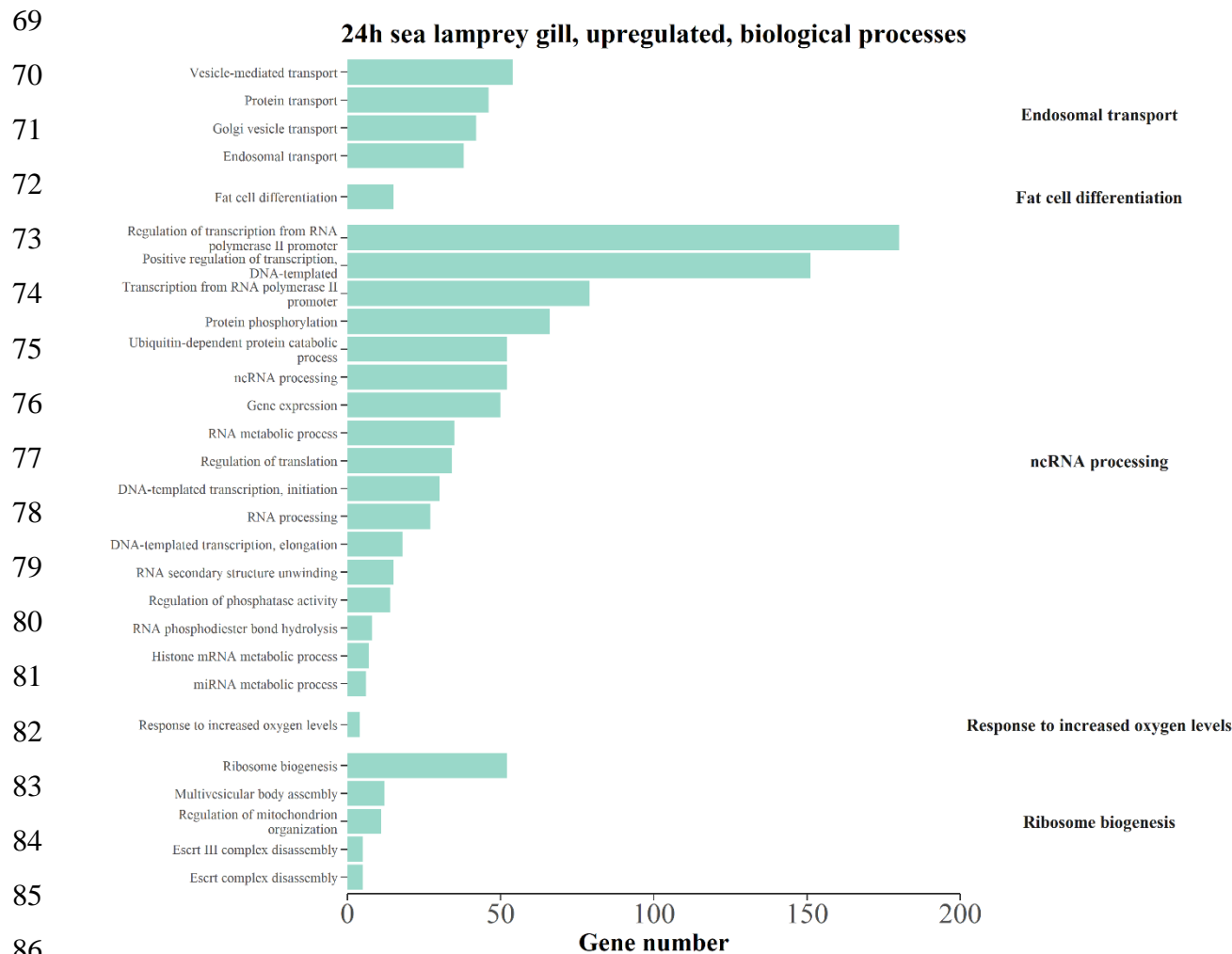


Figure S1: Summary of enriched gene ontology (GO) terms associated with biological processes in transcripts that were upregulated following 6 h of TFM (3-trifluoromethyl-4'-nitrophenol; 2.21 mg L⁻¹ nominal) exposure in the gills of sea lamprey (*Petromyzon marinus*) larvae. Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).



87 **Figure S2:** Summary of enriched gene ontology (GO) terms associated with biological processes
 88 in transcripts that were upregulated following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol;
 89 2.21 mg L⁻¹ nominal) exposure in the gills of sea lamprey (*Petromyzon marinus*) larvae.
 90 Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO
 91 terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were
 92 considered significantly enriched. REVIGO was used to summarize GO terms to reduce
 93 redundancy and group according to similarity (right labels).

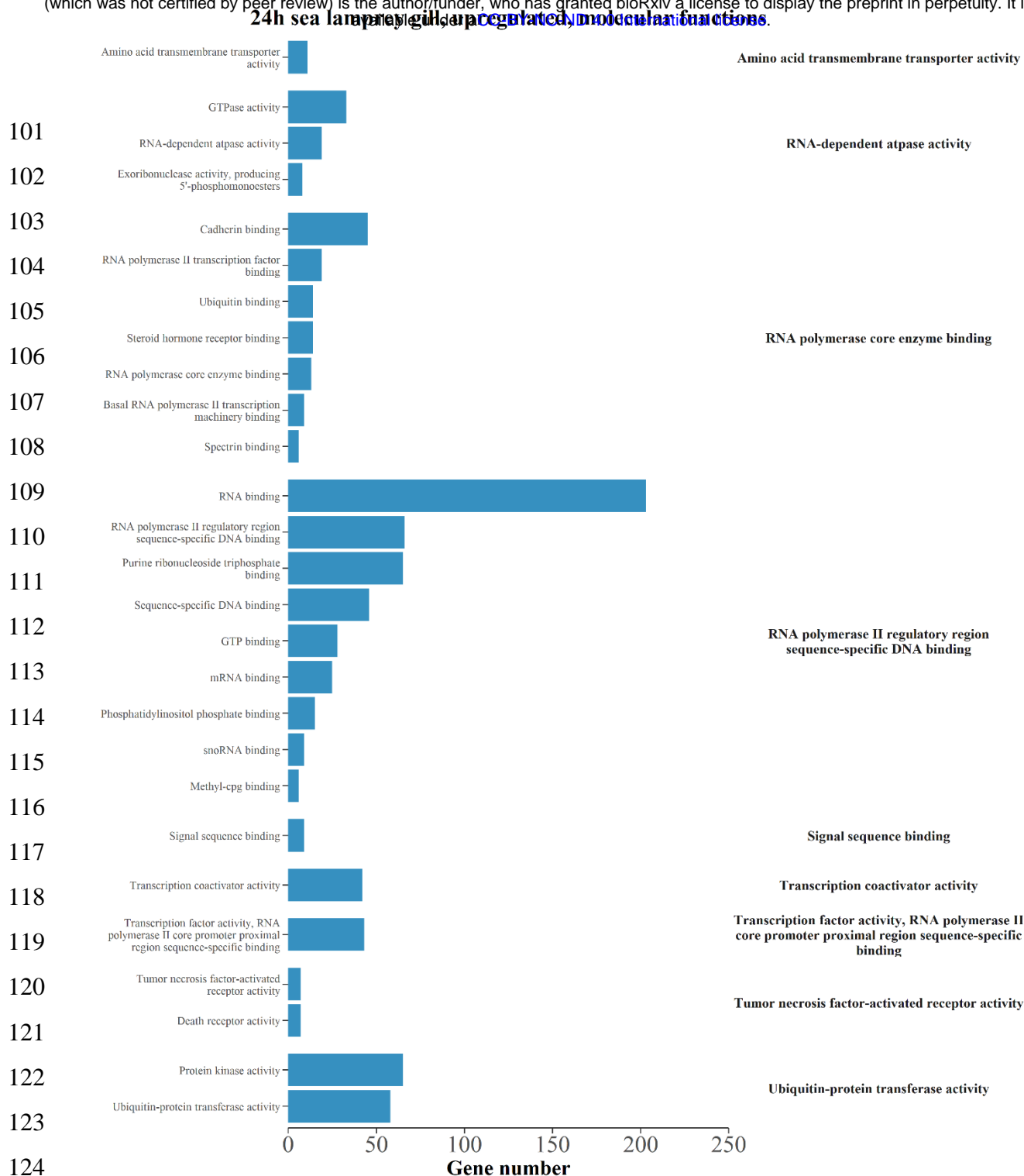
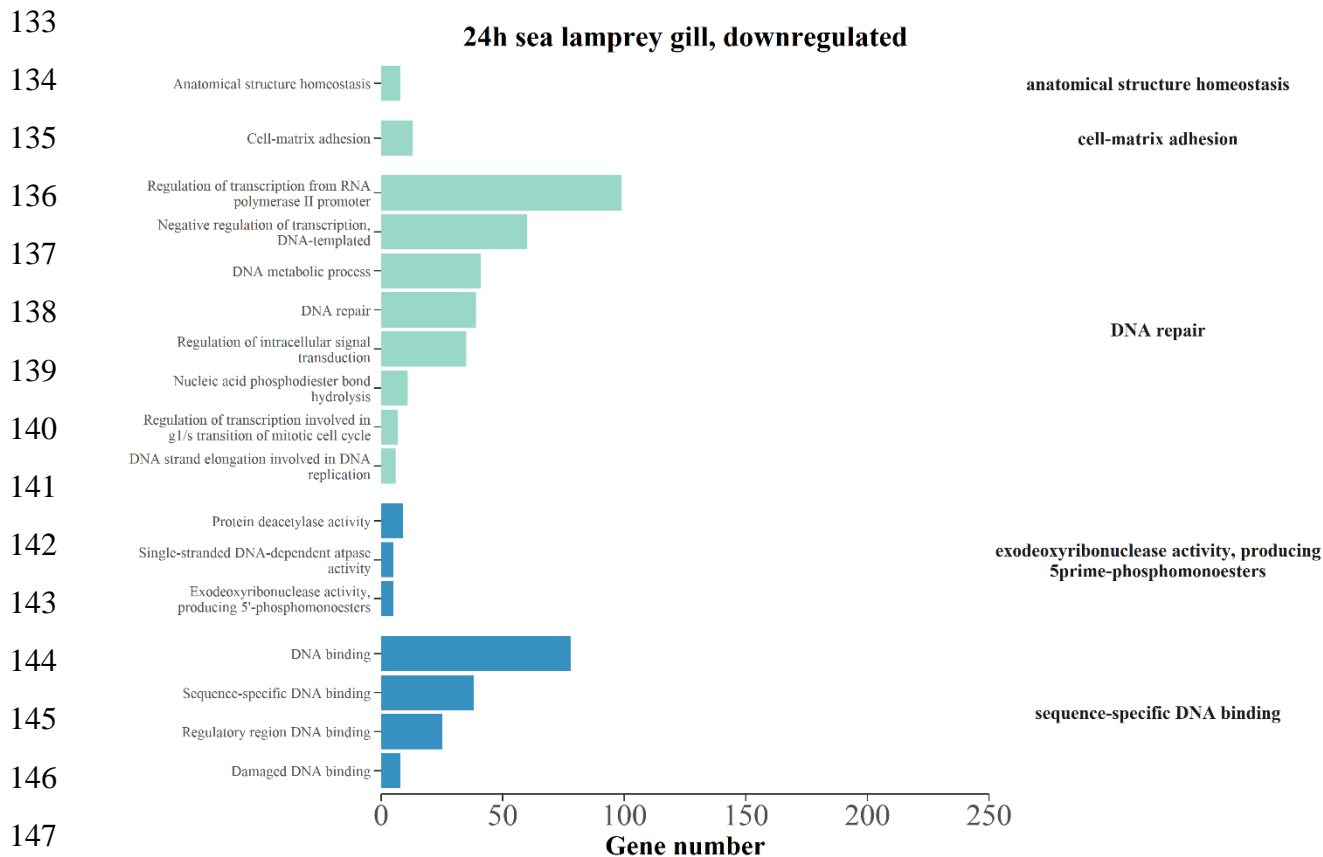


Figure S3: Summary of enriched gene ontology (GO) terms associated with molecular functions in transcripts that were upregulated following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 2.21 mg L⁻¹ nominal) exposure in the gills of sea lamprey (*Petromyzon marinus*) larvae. Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).



149 **Figure S4:** Summary of enriched gene ontology (GO) terms associated with biological processes
 150 (green) and molecular functions (blue) in transcripts that were downregulated following 24 h of
 151 TFM (3-trifluoromethyl-4'-nitrophenol; 2.21 mg L⁻¹ nominal) exposure in the gills of sea
 152 lamprey (*Petromyzon marinus*) larvae. Transcripts were considered differentially regulated at a
 153 false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p <
 154 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to
 155 summarize GO terms to reduce redundancy and group according to similarity (right labels).

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24h sea lamprey liver, upregulated, molecular functions

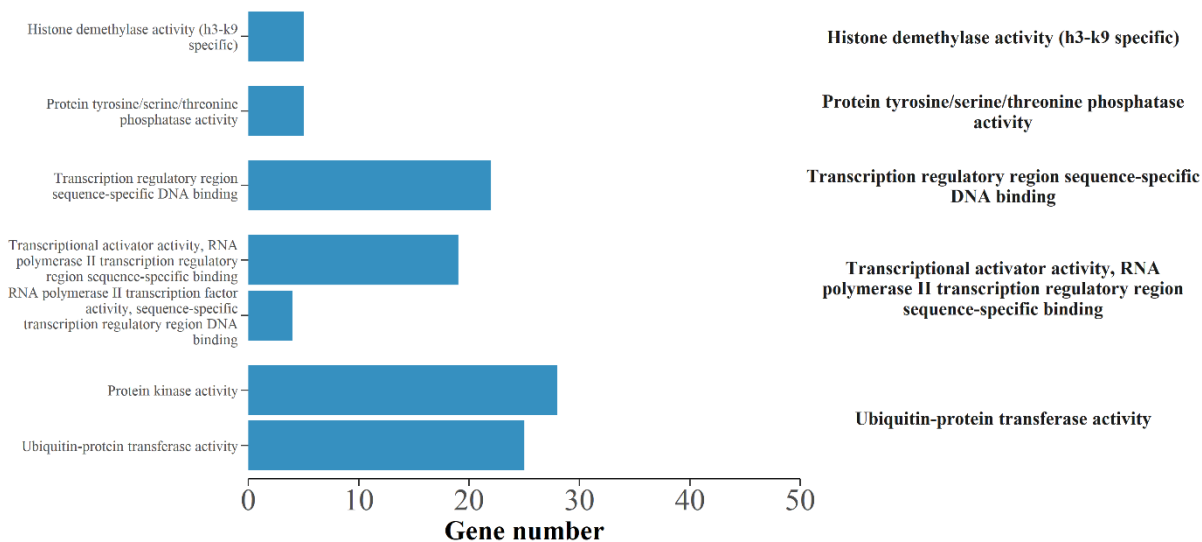
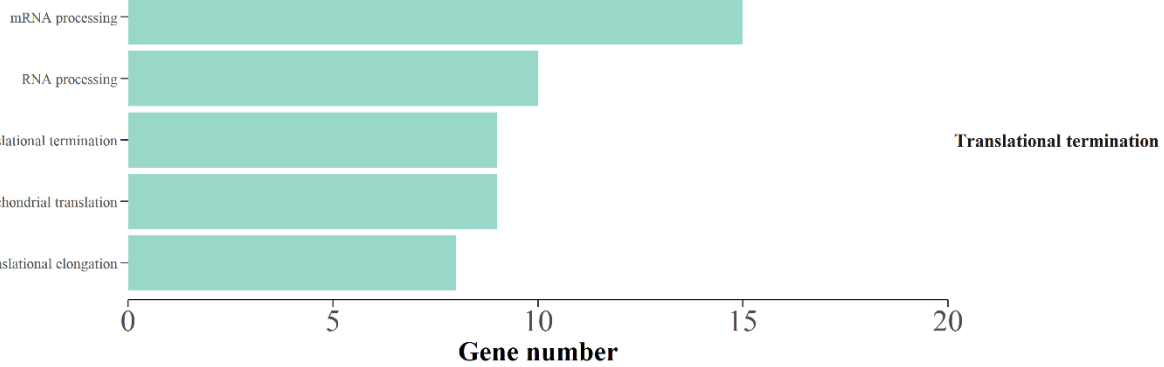


Figure S5: Summary of enriched gene ontology (GO) terms associated with molecular functions in transcripts that were upregulated following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 2.21 mg L⁻¹ nominal) exposure in the liver of sea lamprey (*Petromyzon marinus*) larvae. Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).

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24h sea lamprey liver, downregulated, biological processes

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206 **Figure S6:** Summary of enriched gene ontology (GO) terms associated with biological processes
207 in transcripts that were downregulated following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol;
208 2.21 mg L⁻¹ nominal) exposure in the liver of sea lamprey (*Petromyzon marinus*) larvae.
209 Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO
210 terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were
211 considered significantly enriched. REVIGO was used to summarize GO terms to reduce
212 redundancy and group according to similarity (right labels).

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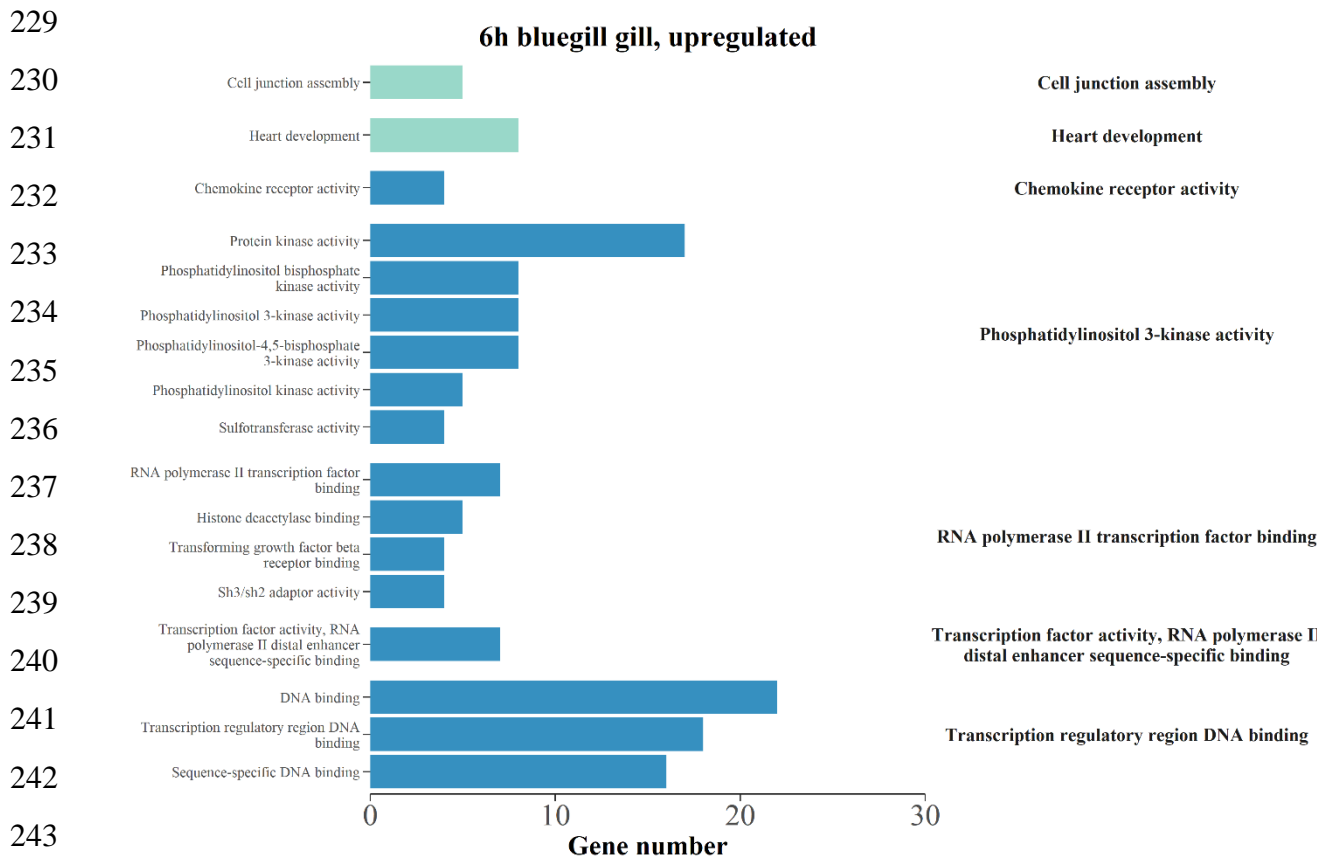
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245 **Figure S7:** Summary of enriched gene ontology (GO) terms associated with biological processes

246 (green) and molecular functions (blue) in transcripts that were upregulated following 6 h of TFM

247 (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of bluegill

248 (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery

249 rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least

250 four transcripts were considered significantly enriched. REVIGO was used to summarize GO

251 terms to reduce redundancy and group according to similarity (right labels).

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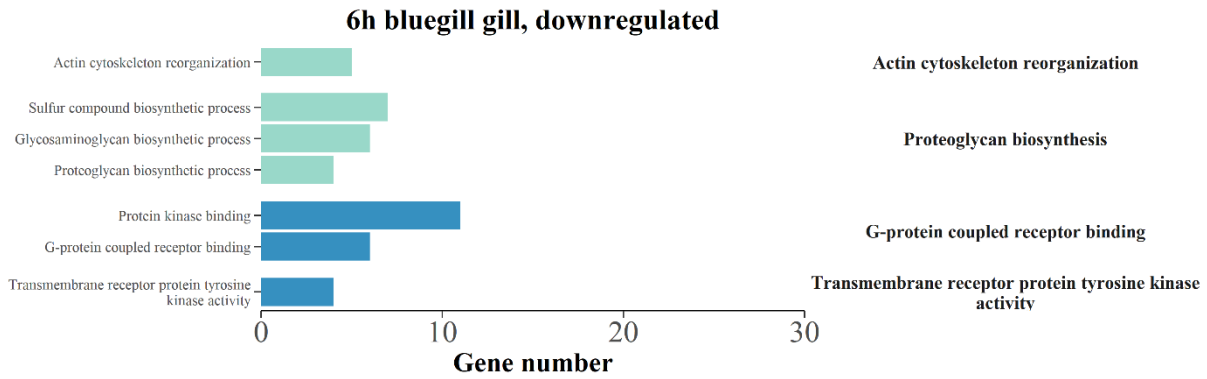
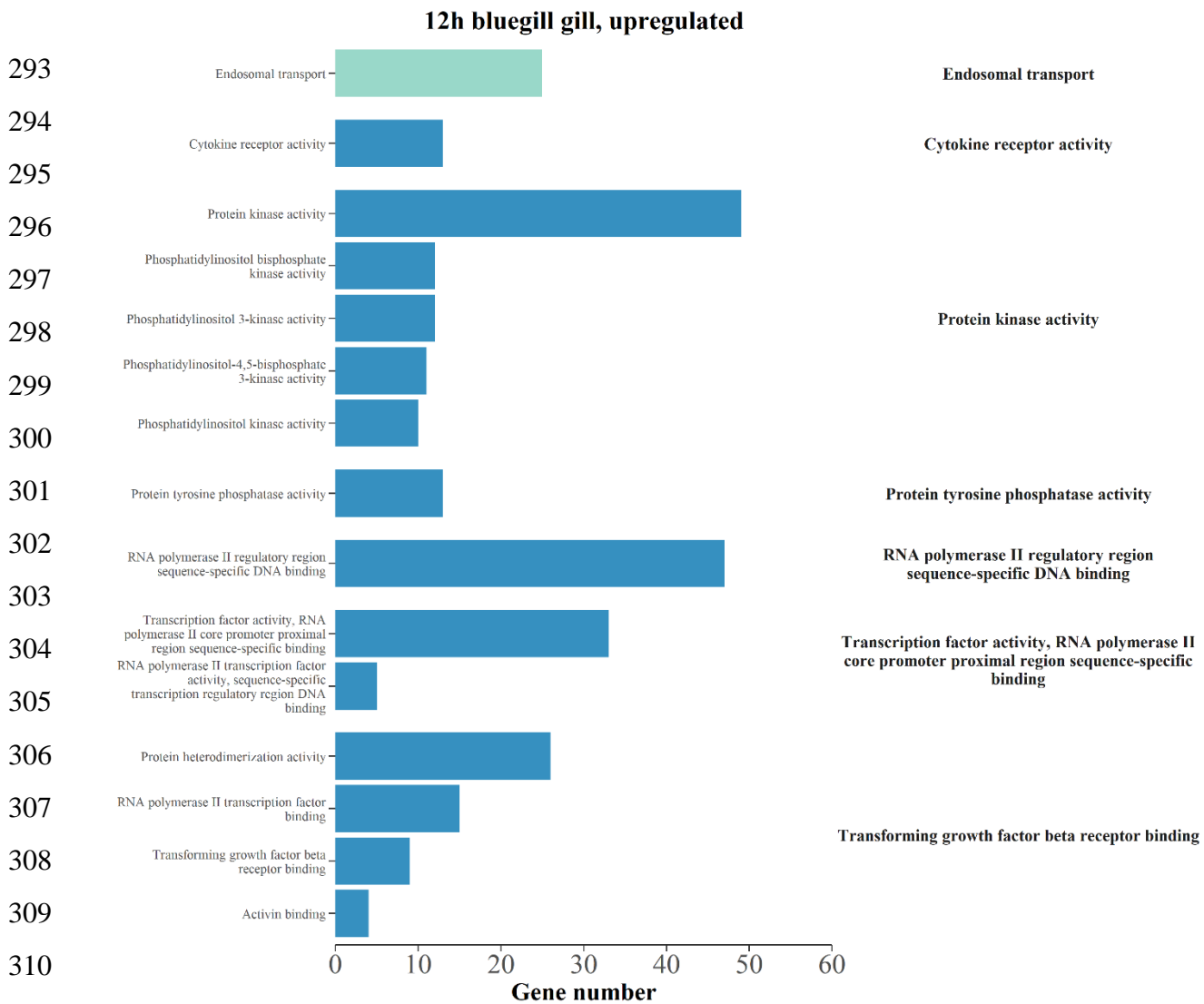


Figure S8: Summary of enriched gene ontology (GO) terms associated with biological processes (green) and molecular functions (blue) in transcripts that were downregulated following 6 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).



311
 312 **Figure S9:** Summary of enriched gene ontology (GO) terms associated with biological processes
 313 (green) and molecular functions (blue) in transcripts that were upregulated following 12 h of
 314 TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of bluegill
 315 (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery
 316 rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least
 317 four transcripts were considered significantly enriched. REVIGO was used to summarize GO
 318 terms to reduce redundancy and group according to similarity (right labels).

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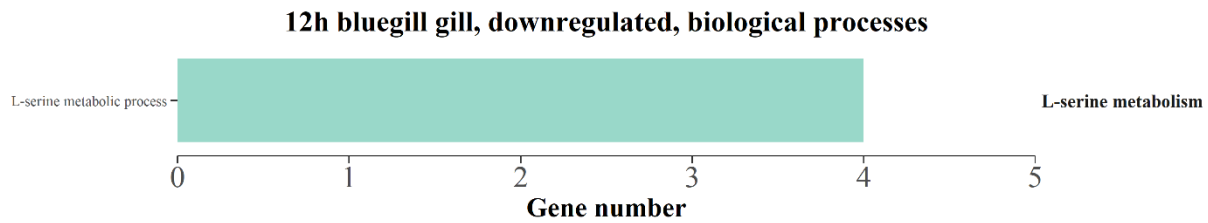
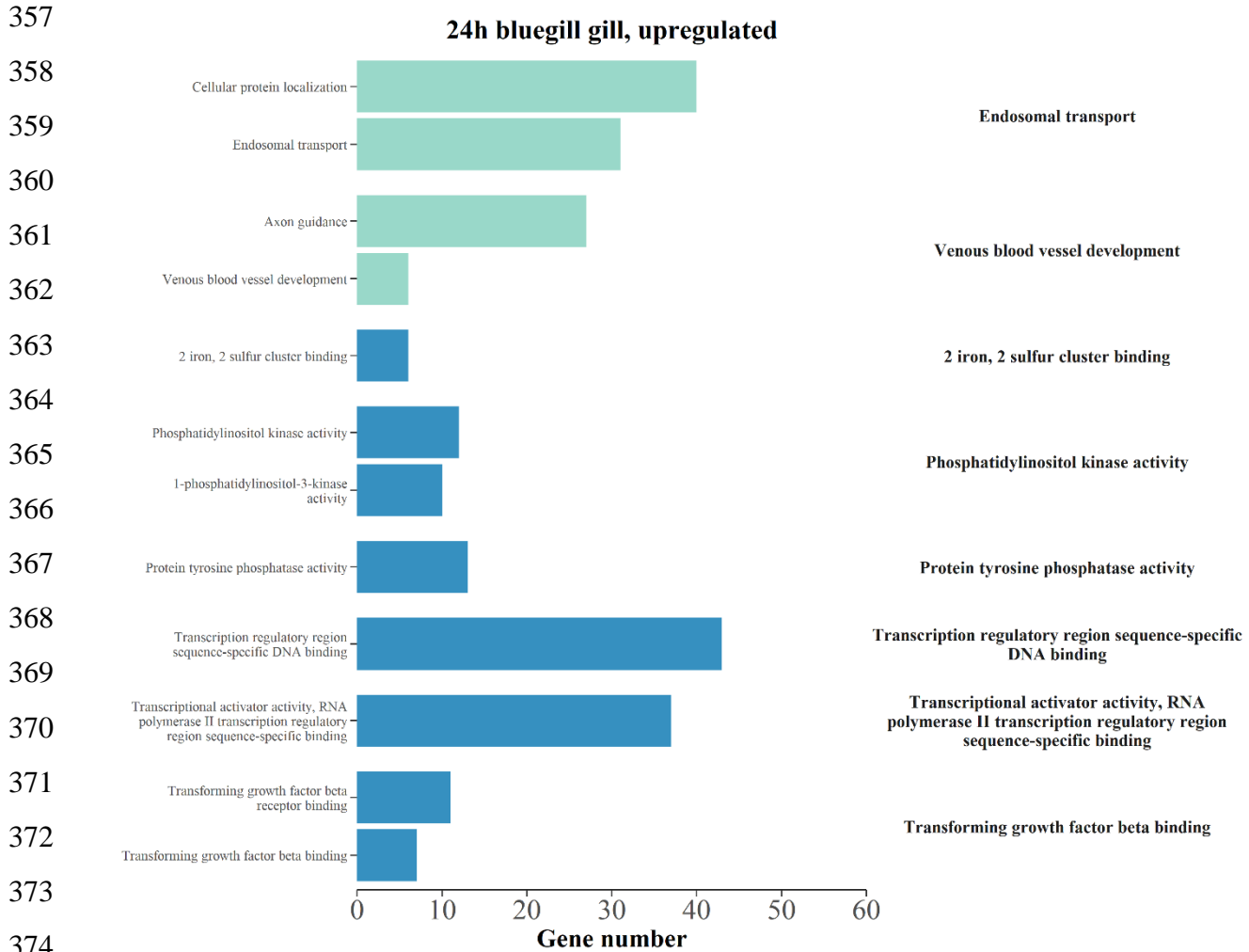


Figure S10: Summary of enriched gene ontology (GO) terms associated with biological processes in transcripts that were downregulated following 12 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).



376 **Figure S11:** Summary of enriched gene ontology (GO) terms associated with biological
 377 processes (green) and molecular functions (blue) in transcripts that were upregulated following
 378 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the gills of
 379 bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false
 380 discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and
 381 at least four transcripts were considered significantly enriched. REVIGO was used to summarize
 382 GO terms to reduce redundancy and group according to similarity (right labels).

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419 **Figure S12:** Summary of enriched gene ontology (GO) terms associated with biological
420 processes (green) and molecular functions (blue) in transcripts that were downregulated
421 following 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the
422 gills of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a
423 false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p <
424 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to
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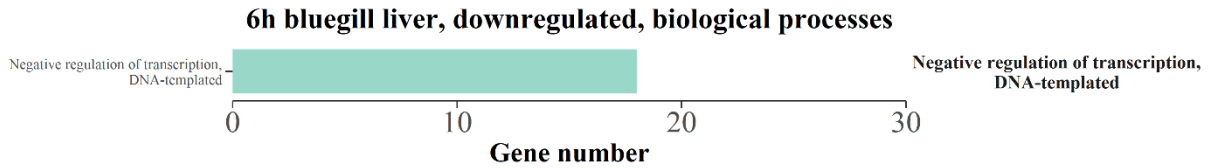


Figure S13: Summary of enriched gene ontology (GO) terms associated with biological processes in transcripts that were downregulated following 6 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the liver of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to summarize GO terms to reduce redundancy and group according to similarity (right labels).

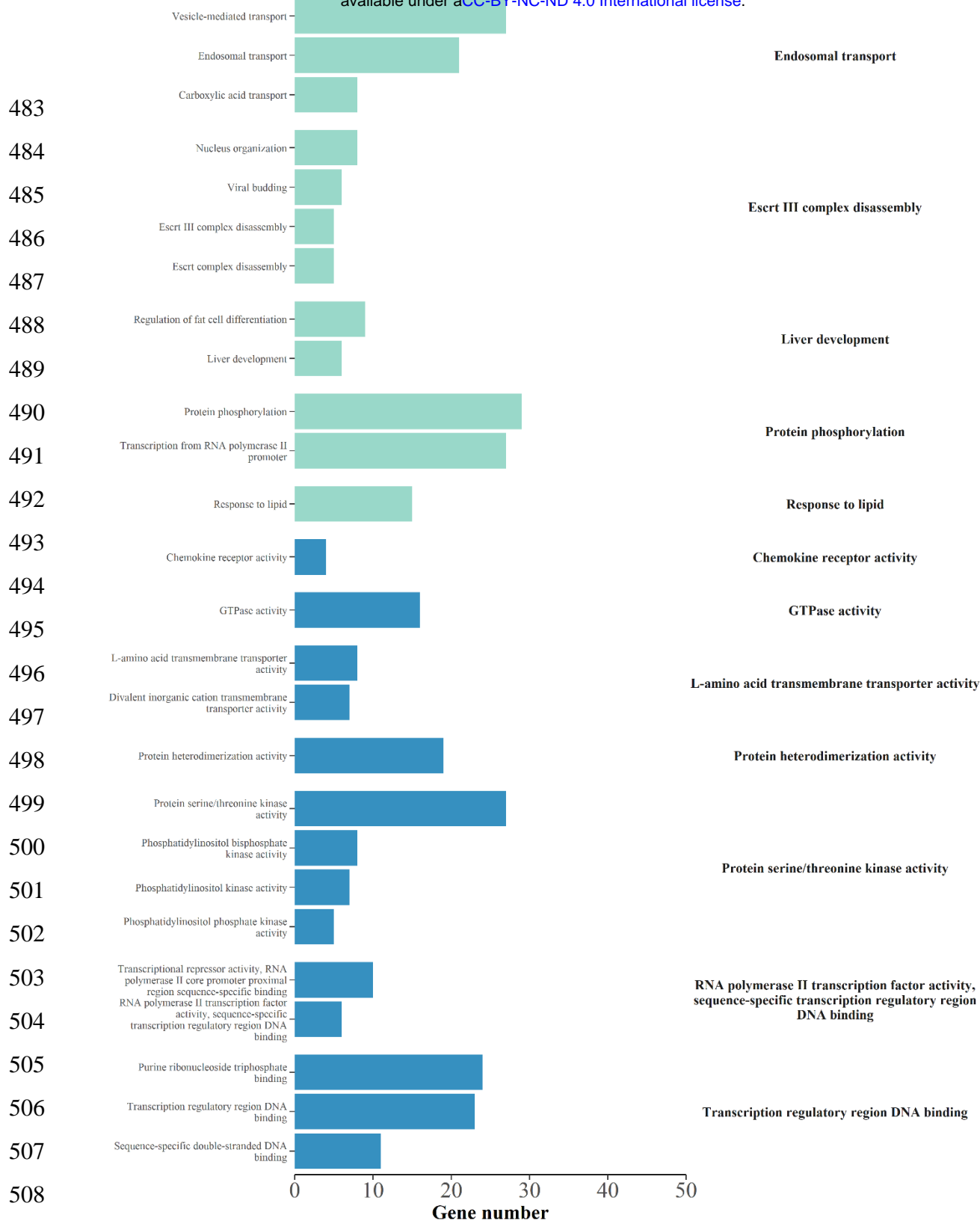


Figure S14: Summary of enriched gene ontology (GO) terms associated with biological processes (green) and molecular functions (blue) in transcripts that were upregulated following 12 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the liver of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false

514 discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted $p < 0.05$ and
515 at least four transcripts were considered significantly enriched. REVIGO was used to summarize
516 GO terms to reduce redundancy and group according to similarity (right labels).

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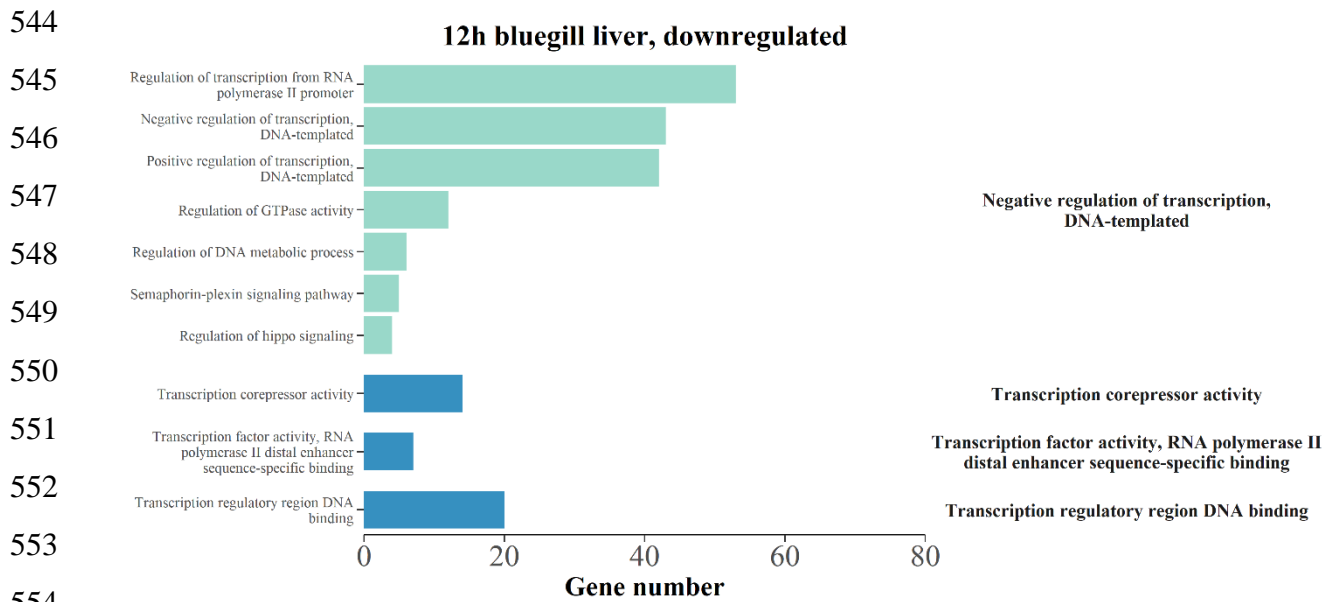
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555 **Figure S15:** Summary of enriched gene ontology (GO) terms associated with biological
556 processes (green) and molecular functions (blue) in transcripts that were downregulated
557 following 12 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the
558 liver of bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a
559 false discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p <
560 0.05 and at least four transcripts were considered significantly enriched. REVIGO was used to
561 summarize GO terms to reduce redundancy and group according to similarity (right labels).

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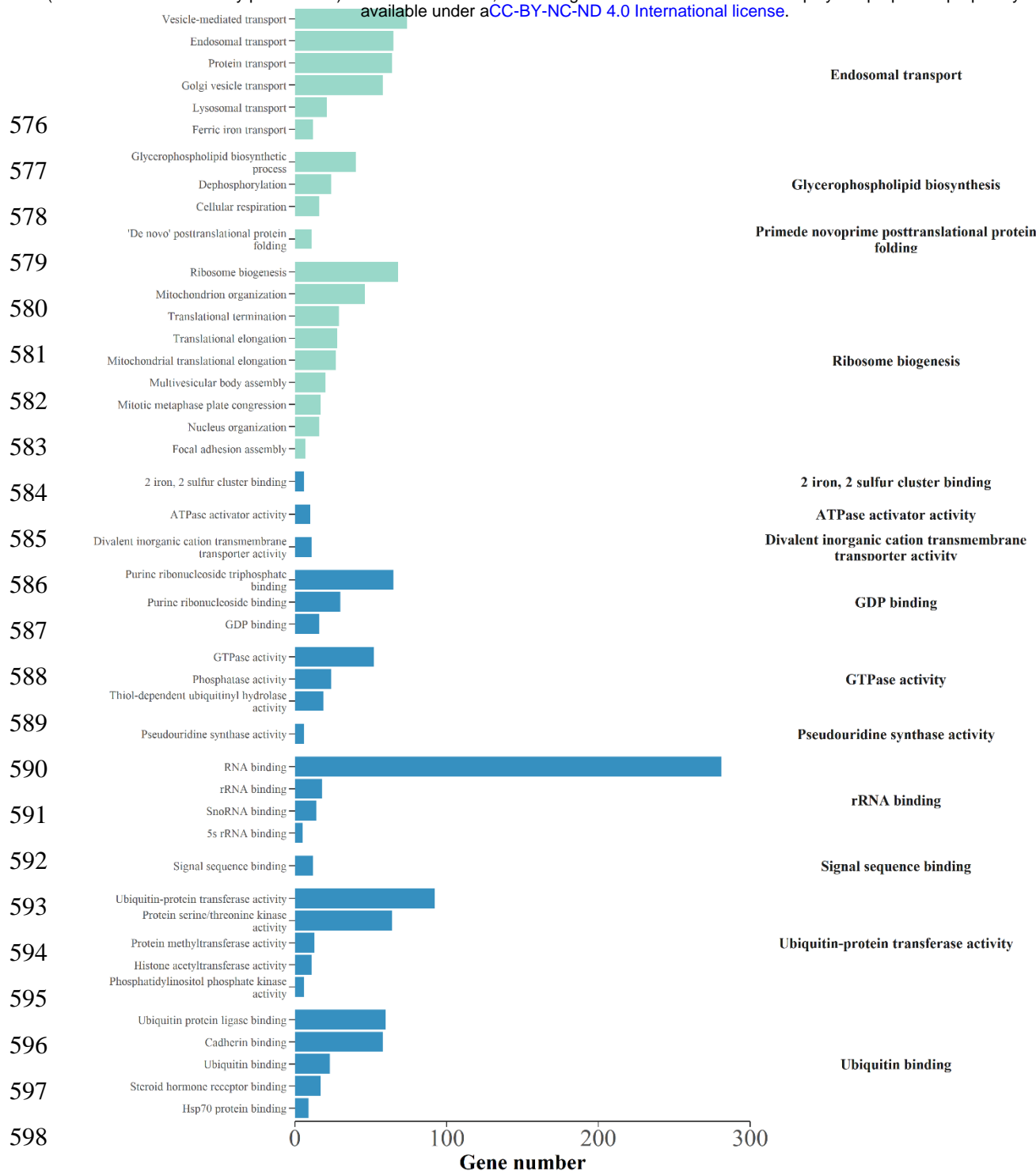
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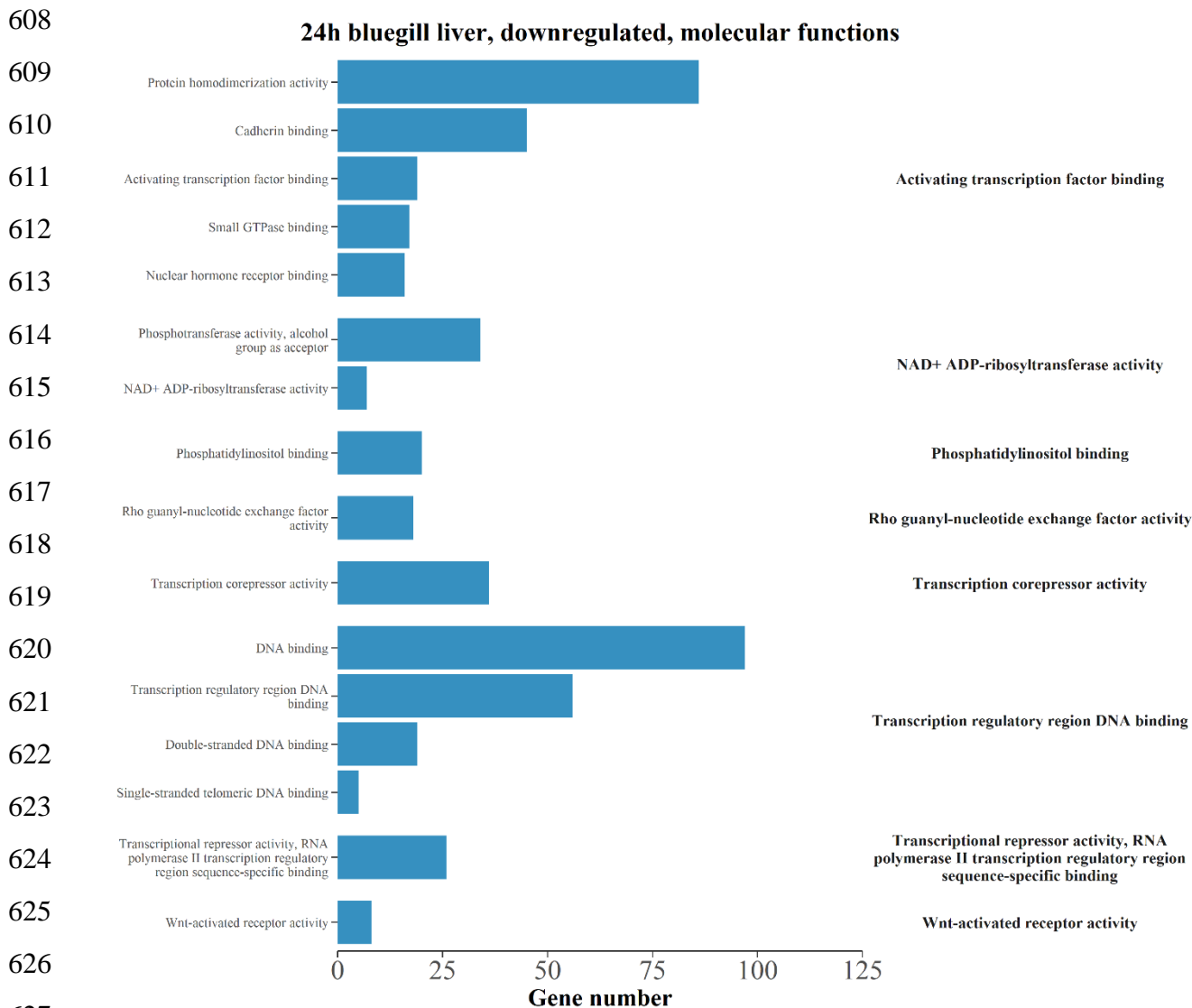
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 600 **Figure S16:** Summary of enriched gene ontology (GO) terms associated with biological
 601 processes (green) and molecular functions (blue) in transcripts that were upregulated following
 602 24 h of TFM (3-trifluoromethyl-4'-nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the liver of
 603 bluegill (*Lepomis macrochirus*). Transcripts were considered differentially regulated at a false
 604 discovery rate < 0.05. Only GO terms from the functional analysis with an adjusted p < 0.05 and
 605 at least four transcripts were considered significantly enriched. REVIGO was used to summarize
 606 GO terms to reduce redundancy and group according to similarity (right labels).

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629 **Figure S17:** Summary of enriched gene ontology (GO) terms associated with molecular
 630 functions in transcripts that were downregulated following 24 h of TFM (3-trifluoromethyl-4'-
 631 nitrophenol; 22.06 mg L⁻¹ nominal) exposure in the liver of bluegill (*Lepomis macrochirus*).
 632 Transcripts were considered differentially regulated at a false discovery rate < 0.05. Only GO
 633 terms from the functional analysis with an adjusted p < 0.05 and at least four transcripts were
 634 considered significantly enriched. REVIGO was used to summarize GO terms to reduce
 635 redundancy and group according to similarity (right labels).

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