

1 **Ecosystem size-induced environmental fluctuations affect the temporal dynamics of**
2 **community assembly mechanisms**

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23 **Abstract**

24 Understanding processes that determine community membership and abundance is important for
25 many fields from theoretical community ecology to conservation. However, spatial community
26 studies are often conducted only at a single timepoint despite the known influence of temporal
27 variability on community assembly processes. Here we used a spatiotemporal study to determine
28 how environmental fluctuation differences induced by mesocosm volumes (larger volumes were
29 more stable) influence assembly processes of aquatic bacterial metacommunities along a press
30 disturbance gradient. By combining path analysis and network approaches, we found mesocosm
31 size categories had distinct relative influences of assembly process and environmental factors
32 that determined spatiotemporal bacterial community composition, including dispersal and
33 species sorting by conductivity. These processes depended on, but were not affected
34 proportionately by, mesocosm size. Low fluctuation, large mesocosms primarily developed
35 through the interplay of species sorting that became more important over time and transient
36 priority effects as evidenced by more time-delayed associations. High fluctuation, small
37 mesocosms had regular disruptions to species sorting and greater importance of ecological drift
38 and dispersal limitation indicated by lower richness and higher taxa replacement. Together, these
39 results emphasize that environmental fluctuations influence ecosystems over time and its impacts
40 are modified by biotic properties intrinsic to ecosystem size.

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46 **Introduction**

47 The community composition of both micro- and macro-organisms at a given point in
48 space and time results from the interaction of multiple assembly processes, including ecological
49 drift, species sorting (environmental filtering), dispersal, and speciation (1-5). Most
50 observational metacommunity studies, however, focus only on spatial snapshots without
51 considering temporal dynamics of community assembly and association networks, or historical
52 contingencies (2, 6). Hence, we still lack knowledge about the underlying mechanisms and
53 regulating factors that temporal dynamics encompass.

54 When species sorting assembles communities, their composition tracks changes in
55 environmental conditions that occur in time and space (2, 7). However, environmental tracking
56 can be hindered or disrupted (8). Such asynchrony can lead to historical contingencies by priority
57 effects (e.g., (8, 9), which can occur during early community formation or when communities re-
58 assemble following perturbation. An important consequence of priority effects is that they
59 impede or delay environmental tracking enacted by species sorting.

60 Environmental changes may influence temporal community assembly processes and the
61 strength of this can be regulated by ecosystem size (e.g., 6). Studies have shown that microbial
62 communities exposed to disturbances are initially, and often to a strong degree, stochastically
63 assembled, but that the importance of species sorting increases later during community re-
64 assembly as more species from the regional species pool arrive (11-14). Rapidly fluctuating
65 environmental conditions, however, may continuously disrupt environmental tracking by
66 reducing opportunities for species sorting to select and shape local communities before the
67 environmental conditions change again. This might promote coexistence of species with different
68 niche optima (20-22) and, thus, reduce beta diversity (16), or could cause extinctions that bolster

69 dispersal limitation and priority effects (23). Nevertheless, many studies happen in controlled
70 settings; thus, we lack knowledge on the temporal dynamics of these processes within larger,
71 more complex habitats which track environmental changes (2, 17). Disturbance strength may
72 uniquely affect microbial communities in ecosystems of different sizes as ecosystem size may
73 influence assembly processes by increasing habitat heterogeneity, community abundance (6, 18,
74 19), and the pace at which communities track environmental changes. For instance, communities
75 may experience different environmental variability including press disturbances (e.g., climate
76 warming, eutrophication, or saltwater incursion), periodic and stochastic environmental
77 fluctuations, where the latter may influence community assembly in response to the former over
78 time and space.

79 Here, we implemented an experiment with freshwater bacterial metacommunities to test
80 how different ecosystem size-induced environmental fluctuations influence the temporal
81 dynamics of community assembly mechanisms. We collected a 64-day time series from
82 mesocosms that allowed bacterial communities sufficient time to experience natural
83 environmental fluctuations. Specifically, we set-up a natural experimental landscape with
84 mesocosms containing identical lake water that differed in volume, which induced differences in
85 environmental fluctuation intensity among the mesocosms. We created a press disturbance by
86 applying a salinity gradient to each set of mesocosm volumes as it has been shown that salinity
87 affects bacterial communities in many ecosystems (e.g., (10-13)). We hypothesized that the
88 importance of species sorting would increase over time in local communities of larger
89 mesocosms that experience relatively minor environmental fluctuations because their
90 communities will have sufficient time for species selection in response to the initial salinity.
91 Second, other environmental changes occurring in mesocosms would be slow in large

92 mesocosms and this would allow time for taxa to be recruited from internal and external
93 dispersal sources and to become active. We expected that species sorting related to salinity
94 differences across communities, i.e., at the metacommunity scale, promotes recruitment of taxa
95 best suited to the salinity. Last, we hypothesized that stochastic and/or dispersal-related assembly
96 processes should be more important in small mesocosms where communities experience strong
97 environmental fluctuations that continually disrupt environmental tracking. We combined
98 quantitative path analysis methods that aim to estimate metacommunity processes with a network
99 approach that identifies environmental tracking patterns through local and time-delayed co-
100 occurrences to provide insights into temporal dynamics of microbial ecosystems (14).

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102 **Methods**

103 *Experimental set-up*

104 Three different sizes (24.5, 70, or 200 L) of hard-shell polyethylene mesocosms were
105 arranged in a field beside Lake Erken (16 per size category) and filled with 0.1 mm filtered lake
106 water from Lake Erken in Sweden (59°51'N 18°35'E) (water properties in Supplementary
107 Information). Mesocosms were seeded with 1 L of sieved and mixed surface sediments collected
108 from Lake Erken at ~0.5 m water depth.

109 To induce species sorting with a press disturbance, a salinity gradient was created within
110 each mesocosm size category using nitrate- and phosphate-free sea salt (Red Sea Aquatics Ltd,
111 Verneuil-sur-Avre, France) and ranged from freshwater (0 ‰) to 6 ‰ with 0.4 ‰ increments
112 (rationale for range in Supplementary Information). Mesocosm water surface area and volume
113 were proportional such that air or rain dispersal was proportional across size classes. Mesocosm

114 sediment was also a recruitment source (15-17). Equal mesocosm bottom surface areas allowed
115 for equal recruitment independent of fluctuation category.

116 *Monitoring and sampling*

117 Mesocosms were monitored on days 1, 2, and 4, and then every fourth day for 64 days
118 from July to September 2016. Monitoring included depth profiles of conductivity (to measure
119 salinity changes) and temperature, and depth integrated pH, chlorophyll-*a*, and colored dissolved
120 organic matter (CDOM) fluorescence (see Supplementary Information for details). Weather data
121 from Svanberga, Sweden (0.87 km southwest of the site) included daily precipitation and hourly
122 air temperature (Swedish Meteorological and Hydrological Institute). Every eighth day, water
123 was collected for total organic carbon (TOC), total nitrogen (TN), and total phosphorus (TP) and
124 analyzed using established methods (18).

125 Water samples for enumerating microorganism cells were collected simultaneously with
126 bacterial community composition (below) and preserved with sterile formaldehyde to 2.5 % (19).
127 Samples were stained with SYTO™ 13 Green Fluorescent Nucleic Acid Stain (ThermoFisher
128 Scientific), counted (CyFlow Space flow cytometer, Partec, Münster, Germany) and analyzed
129 using FlowingSoft software (Perttu Terho). Total community size was calculated as cell
130 abundance (mL^{-1}) multiplied by mesocosm volume.

131 *Bacterial community composition*

132 Mesocosm water was collected on days 1, 2, 4, 8 and every 8th day thereafter for 64 days
133 to assess community composition through 16S rRNA amplicon sequencing to specifically detect
134 active members (20). Depth integrated 0.5 L water samples, and air and rain immigration
135 samples (see Supplementary Information), were collected and filtered onto 0.2 μm pore-size
136 filters (47 mm Supor-200 filters, Pall Corporation, Hampshire, UK) until 5 minutes or 0.5 L

137 volume was reached. Filters were flash-frozen in liquid nitrogen and stored at -80°C . DNA from
138 initial lake water and sediment used in the experiment were sampled to learn initial communities
139 and seed banks.

140 Nucleic acids were extracted using a modified protocol from Easy-DNATM kit
141 (Invitrogen, Carlsbad, CA, USA). See Supplementary Information and DOI for a detailed
142 protocol (dx.doi.org/10.17504/protocols.io.xekfjcw). Samples were submitted to SNP&SEQ
143 Technology platform at SciLife in Uppsala, Sweden for two Illumina MiSeq PE300bp
144 sequencing runs with v3 chemistry.

145 *Data processing*

146 Sequencing resulted in 35.6 million paired reads from 609 demultiplexed samples
147 including 12 extraction and PCR negatives. Primers were removed from sequences using
148 cutadapt v 2.7 ref. (21). The DADA2 pipeline (22) was used for sequence processing and
149 taxonomy assignment of Amplicon Sequence Variants (ASVs) using the SILVA v. 138.1
150 reference database (23) (Supplementary Information, Table S1).

151 For beta diversity analyses, ASVs with counts less than 10 were removed and samples
152 were subsampled to a minimum of 5 028 reads, retaining 7 983 unique ASVs. Samples not
153 meeting the 5 028 reads requirement were excluded (Table S2). Both alpha and beta diversity
154 datasets represented 99 % coverage. Raw sequences are available in the European Nucleotide
155 Archive (study accession number PRJEB26595).

156 *Statistical analyses*

157 Statistical analyses were conducted in R (v3.4.3 and v4.0.2) ref. (24) with package
158 “vegan” (25) unless otherwise specified.

159 Fluctuation magnitudes among mesocosm sizes

160 For each environmental variable, fluctuations data were analyzed using the mean of
161 absolute differences of mesocosms in a size category between one date and the previous
162 sampling date. For variables with depth profiles (conductivity and temperature), the absolute
163 difference at each depth was used to calculate the mean change per mesocosm. The
164 environmental variables dataset is in the DiVA repository (26). To determine if the magnitude of
165 changes differed between mesocosm sizes over time, nonparametric tests for repeated measures
166 with an ANOVA-type statistic (ATS) were used (R package and function *nparLD*, ref. (27)).
167 Mesocosms were assessed using principal components analysis (PCA) of original and absolute
168 changes of environmental variables (both log-transformed) and fit with environmental vectors
169 (Fig. S2).

170 Community composition, diversity, and recruitment

171 Non-metric multidimensional scaling (NMDS) with Bray-Curtis dissimilarities was used
172 to visualize bacterial community composition and environmental variables. Shannon's index and
173 Pielou's evenness were calculated and richness was estimated using the package "breakaway"
174 (28). Temporal beta diversity differences in each mesocosm were evaluated by comparing each
175 community with the previous using Jaccard pairwise dissimilarity values. Dissimilarity was
176 partitioned between taxa turnover (taxa replacement) and community nestedness (chronological
177 subsets of taxa) using package "betapart" v1.5.2 ref. (29). Variation from each partition captured
178 by mesocosm size was compared using PERMANOVA tests (30) with function *adonis* and 999
179 permutations.

180 Recruitment was evaluated by pooling each mesocosm's active ASVs across days; ASVs
181 present on day one were removed from the pool leaving those recruited during the experiment.
182 Recruited ASVs were matched with their source seed bank(s) based on DNA from sediment,

183 initial lake water, air, and rain. Sources for unmatched ASVs were considered unknown. For
184 each mesocosm, the percent of recruited ASVs was calculated, split into each source, and
185 examined across the salinity gradient using Pearson's correlations.

186 Path analysis

187 To detect drivers of metacommunity dynamics, a spatiotemporal path analysis was used
188 (31). This method calculates dissimilarity for all community pairs sampled over time and space
189 and estimates, as individual paths on this beta-diversity measure, the influences of spatial
190 distance (Δx), temporal distance (Δt), environmental distance (ΔE), mean community size ($\langle J \rangle$,
191 cell abundance multiplied by mesocosm volume), and absolute differences in community size
192 (ΔJ) and taxa richness (ΔS). Nestedness between sites should explain a positive link between
193 differences in community size and richness thereby increasing community dissimilarity (31).
194 Bray-Curtis dissimilarity was used for the community dissimilarity matrix (β_{bc}). A permutation-
195 based approach adjusted with Benjamini-Hochberg procedure indicated path significance. Model
196 fit was assessed with the standardized root mean square residual (SRMR). The analysis was run
197 separately for each mesocosm size using the days required for network analysis (Supplementary
198 Information), with the *sem* function in R package "lavaan" (32).

199 Network analysis

200 To uncover local and time-delayed microbial associations and the extrinsic effects of
201 environmental variables on bacteria, extended local similarity analysis (eLSA) was applied (14).
202 Given our temporal data, this approach detects undirected associations (e.g., without time
203 delays), and associations where the change of one factor (a taxon or environmental variable)
204 chronologically leads or follows another factor. For a link between taxa and environmental
205 variables, the association type (delayed or non-delayed) can indicate tracking that is time-lagged

206 due to transient priority effects, or simultaneous through species sorting, respectively.
207 Associations were determined for each mesocosm size using eLSA wherein mesocosms from
208 each size category were used as fluctuation level replicates ($n = 16$). Because of the within- and
209 across-size variability of bacterial communities (e.g., significant differences in taxa richness), we
210 selected and analyzed only the core bacterial groups for each mesocosm size to make it
211 comparable. Hence, networks used the 50 most abundant ASVs from each size category. eLSA
212 (v1.0.2) was run over eight sampling points, allowing for local similarity (LS) correlations
213 between samples taken eight days apart ($d = 1$). LS correlations (LS value ≥ 0.05 ; $Q \leq 0.01$) were
214 visualized in Cytoscape v3.8.2 (33). Network characteristics were calculated using the Cytoscape
215 plugin NetworkAnalyzer (34). See Supplementary Information for details on sample selection,
216 dominant ASV abundances, and statistics.

217 **Results**

218 *Environmental fluctuations in mesocosms*

219 Environmental variable fluctuations corresponded with mesocosm size and reflected
220 rainfall and air temperature (Fig. S1, Table S3, Fig. S2). Size categories experienced
221 significantly different conductivity and temperature fluctuations. After four days small and
222 medium mesocosm conductivity fluctuated more than large mesocosms (Fig. S1, Table S3).
223 Mean temperature fluctuation increased inversely with mesocosm size (Table S3). Mesocosm
224 depth profiles showed stable conductivity, but temperature decreased with depth in medium and
225 large mesocosms (Fig. S3).

226 Mesocosm sizes differed in nutrient concentrations and the absolute change of other
227 environmental variables (chl-*a*, CDOM, pH, TN, TOC, TP and cell abundance, Table S3) and
228 most pairwise comparisons showed that the degree of change differed significantly between sizes

229 with the greatest changes in small mesocosms. Absolute changes between sampling dates and
230 individual timepoints grouped according to size (Fig. S2). Measured nutrients and conductivity
231 positively correlated with decreasing mesocosm sizes (environmental vector correlations, $p <$
232 0.05). For water temperature, sampling date was more influential than mesocosm size. Cell
233 abundances (cells mL^{-1}) increased over time and was highest in small and medium mesocosms
234 (ATS, $p < 0.001$, Fig. S4A). From day 24, total community abundance per mesocosm was lower
235 in small than medium and large mesocosms (ATS, all $p < 0.002$, Fig. S4B).

236 *Community composition, diversity, and recruitment*

237 Bacterial community composition shifted with time and initial conductivity in all
238 mesocosm sizes (Fig. S5). Diversity indices (estimated richness, Pielou's evenness, Shannon's
239 index) did not differ on the first day (two-way ANOVA, all $p > 0.05$), but over time all three
240 indices differed by mesocosm size (Fig. 1, repeated measures ANOVA, all overall $p \leq 0.001$;
241 pairwise Bonferroni adjusted). All sizes differed significantly in bacterial richness which was
242 lowest in small and highest in large mesocosms (all $p \leq 0.001$). The more evenly distributed
243 ASV abundances in large mesocosms widened the separation in Shannon's index diversity
244 between large and small or medium mesocosms, indicating a greater presence of dominant
245 and/or rare taxa in smaller mesocosms (Fig. 1). Mesocosm size explained some variability in
246 beta diversity from turnover (F model = 13, $R^2 = 0.04$, $p \leq 0.001$) with communities in small
247 mesocosms experiencing higher turnover by taxa replacement than those in large mesocosms
248 (Wilcoxon Test, $W = 4276$, Bonferroni $p.adj. = 0.02$, Fig. S6A). Statistically, mesocosm size did
249 not explain variability in nestedness, although communities in large mesocosms trended towards
250 greater nested species loss (Fig. S6B).

251 Recruited ASVs as a percentage of total unique ASVs, had weak negative correlations
252 with the salinity press disturbance in small and large mesocosms ($r = -0.35$ and -0.39 , $p < 0.005$,
253 respectively, Fig. S7). Less than 15 % of recruited ASVs in each mesocosm were attributed to a
254 known source. In all mesocosm sizes, recruitment from water declined significantly with initial
255 salinity (small: $r = -0.90$, medium: $r = -0.85$, large: $r = -0.89$, all $p < 0.001$). Recruitment from
256 sediment showed different patterns across salinity levels in small and large mesocosms: it
257 decreased in small mesocosms and was unchanged in large mesocosms ($r = -0.70$, $p = 0.002$; $r =$
258 0.45 , $p = 0.08$, respectively). Sediment was typically the largest recruitment source in the most
259 saline mesocosms. Air and rain recruitment was related to salinity level only in the medium
260 mesocosms where it was weakly positively correlated (air: $r = 0.52$, $p = 0.04$, rain: $r = 0.58$, $p =$
261 0.02).

262 *Path analysis*

263 Bacterial metacommunities in mesocosms of different sizes experienced disparate
264 relative influences from species sorting by environmental variation, demographic stochasticity,
265 and dispersal limitation (Fig. 2). The model fit for small mesocosms was roughly twice that of
266 medium and large mesocosms (Fig. 2).

267 Species sorting (ΔE) had the most influential direct effect on community dissimilarity
268 (β_{bc}) (Fig. 2). This effect was strongest in small mesocosms and similar in medium and large
269 mesocosms (sum of absolute standardized estimates 0.925, 0.773, and 0.766, respectively), but
270 all sizes had significant environmental distance and community dissimilarity relationships
271 (Tables S4-S6). Small mesocosms had five significant relationships between community
272 dissimilarity and environmental variables (conductivity, temperature, chlorophyll-*a*, TOC, and
273 TN); large mesocosms had three (conductivity, temperature, and chlorophyll-*a*), and medium

274 mesocosms had only conductivity. Conductivity correlated most strongly with community
275 dissimilarity of medium, followed by large and small mesocosms. Significant correlations
276 between temporal (Δt) or spatial (Δx) distance and environmental (ΔE) distance were positive
277 and increased with mesocosm size. The indirect effect of time on community dissimilarity
278 through species sorting was apparent with all measured variables except conductivity.

279 Demographic stochasticity was indicated by significant negative relationships between
280 mean community size ($\langle J \rangle$) and community dissimilarity in all mesocosm sizes (Fig. 2, Tables
281 S4-S6). Small mesocosms had the strongest influence by demographic stochasticity. All
282 mesocosm sizes had positive correlations between temporal distance and community
283 dissimilarity indicating additional demographic stochasticity. Relationship strengths differed
284 with size: temporal changes had the greatest influence in large, then small, then medium
285 mesocosms.

286 Dispersal limitation shown as a positive correlation between geographic distance and
287 community dissimilarity appeared only for small mesocosms (Fig. 2). Large mesocosms had a
288 significant negative correlation between geographic distance and community dissimilarity but
289 this was considered an artefact of the linear modelling framework (31) and negligible compared
290 with the relationship between space and community dissimilarity via the environmental variation
291 pathway.

292 The path analysis for medium and large mesocosms also suggested an effect of taxa
293 nestedness whereby communities form as subsets of original communities over time or space
294 (Tables S4-S6). First, differences in community richness (ΔS) positively correlated with
295 community dissimilarity. This relationship was strongest in large mesocosms. Second, a

296 significant positive relationship occurred between differences in community size (ΔJ) and
297 richness in medium and large mesocosms.

298 *Association networks*

299 Association networks of the 50 most abundant ASVs (members of Actinobacteriota,
300 Bacteroidota, Cyanobacteria, Planctomycetota and Proteobacteria) differed among the three
301 mesocosm sizes (Fig. 3, Table S7). The number of total edges and ASV nodes increased with
302 mesocosm size, and the proportion of delayed (time-shifted) associations were higher in larger
303 mesocosms (small: 25.9 %, medium: 49.8 %, large: 46 %) (Table S7). Small mesocosms had the
304 most ASVs ($n = 18$) that were unassociated with environmental variables and bacterial
305 abundance while medium and large mesocosms had only 4 and 6 ASVs, respectively (Table S7).
306 Network subsets showed no connection between conductivity and ASVs of small mesocosms,
307 while conductivity influenced many abundant ASVs from medium and large mesocosms (mainly
308 phylum Proteobacteria). In large mesocosms, conductivity had a direct (non-delayed) influence
309 on ASVs (except one Cyanobacterium), while in medium mesocosms it had both time-shifted
310 (e.g., mainly positive in Bacteroidota and negative in Proteobacteria) and non-delayed (e.g.,
311 Cyanobacteria) associations with taxa (Fig. S8, Table S8).

312 Association networks were quantitatively compared by mesocosm size with commonly
313 used topological characteristics. Negative associations, average number of neighbors, and
314 network density (the proportion of possible edges that are associated with nodes) increased with
315 mesocosm size (Table S7). Further, small and medium mesocosms networks were less
316 centralized (the concentration of centrality among the nodes) than those in large mesocosms.
317 When considering only taxa associations, small mesocosms had the least centralized network
318 with more taxa displaying similar numbers of links (Table S7).

319

320 **Discussion**

321 Here we show how differences in environmental fluctuation strengths due to differences
322 in ecosystem, i.e. mesocosm, size influenced the temporal dynamics of community assembly in
323 response to a salinity press disturbance (Fig. 4). First, species sorting was generally the most
324 influential process for all mesocosms but differences in how species sorting operated among
325 mesocosm sizes at the community (path analysis) and individual taxa levels (association network
326 analysis). These evaluations indicated that under low environmental fluctuations, dominant ASV
327 populations were effective trackers of environmental conditions. When ecosystem size-induced
328 environmental fluctuations were strong (i.e., small mesocosms), environmental tracking was
329 disrupted. Second, the salinity press disturbance initiated environmental tracking, especially
330 under stable conditions (i.e., larger mesocosms), through the recruitment of taxa from seed banks
331 (mainly sediment at high salinity). Third, stochasticity and dispersal-related assembly processes
332 (e.g., dispersal limitation) generally were more important for communities of small ecosystems.
333 Overall, our study aligns with previous findings that ecosystem size influences community
334 assembly processes (35-37), but we identified this effect to derive from environmental
335 fluctuations created by ecosystem size differences and corresponding differences in species
336 sorting effects.

337

338 *Salinity press disturbance enforces environmental tracking*

339 Differences in the magnitude of salinity press disturbances induced clear compositional
340 shifts within and across mesocosms over time. This was expected as we used salinity to induce
341 species sorting because it is an environmental factor that causes clear taxonomic differences in

342 aquatic bacterial communities (13, 38-41). However, there were disparities in how well
343 communities in each mesocosm size tracked temporal changes in salinity.

344 The path analysis and network analysis results indicated that species sorting patterns
345 differed across mesocosm sizes and were altered by time. The direct effects of significant
346 environmental variables with unidirectional influences (i.e., salinity and temperature) on species
347 sorting were most influential in medium and large, stabler mesocosms. However, when variables
348 prone to feedbacks (i.e., nutrients, see below) were included into the total environmental effect
349 on composition, species sorting was greatest in small, highly fluctuating mesocosms. In contrast,
350 the indirect effect of time on composition via species sorting increased with mesocosm size and
351 was driven primarily by changes in all environmental variables except for salinity (which
352 changed minimally within a mesocosm compared to the salinity gradient). This temporal pattern
353 generally agreed with the network results of the 50 most abundant bacteria which showed that
354 they best tracked multiple environmental variables over time in large and medium mesocosms.
355 Here, almost all ASVs directly linked to environmental variables and populations of core groups
356 of taxa oscillated correspondingly with temporal salinity changes. In addition, when we isolated
357 taxa and salinity associations, the network approach revealed many non-delayed associations in
358 large mesocosms, indicating that the most abundant bacterial populations rapidly
359 (simultaneously) tracked changes. Despite the path analysis results, no associations were found
360 in the small mesocosms which could otherwise indicate salinity tracking through time.

361 Several reasons may explain the contradiction between the two analytical approaches
362 regarding direct species sorting. First, in the network analysis, we calculated associations only
363 among the 50 most abundant taxa, thus, we likely overlook conditionally rare taxa that can be
364 temporarily abundant (42) as a consequence of the rapidly changing environment in small

365 mesocosms. This is supported by the trend of higher taxa turnover and direct demographic
366 stochasticity (discussed below) in small mesocosms. Another explanation may include the
367 phenomenon that bacterial communities can be an imprint of past environmental conditions (43)
368 and the correlations detected between community dissimilarity and environmental variables
369 might coincide with prior processes. Last, the analytical approaches generally agreed concerning
370 direct effects by salinity, but differed for variables with the potential for feedbacks (i.e.,
371 nutrients). Although the path analysis portrays nutrients as effect variables, they are also
372 modified by microorganisms. Likely due to the salinity and greater proportions of sediment,
373 small mesocosms had greater nutrient concentrations and higher cell densities including from
374 observed algal blooms. These conditions could increase competition which hinders synchrony
375 between abiotic variables and taxa abundances (44).

376 Taken together, our findings (conceptualized in Fig. 4) around the importance of species
377 sorting and the strong temporal influences highlight the distinct differences in the mechanisms
378 underlying species sorting in mesocosms of different sizes. These findings became apparent
379 through combining the path analysis, which captures both spatial and temporal patterns at the
380 whole community level, and the network analysis, which captures time-associated patterns of the
381 most abundant populations.

382

383 *Ecosystem size regulates community assembly and associations among bacterioplankton*

384 While the different environmental conditions from the press disturbance and ongoing
385 environmental changes throughout the experiment might explain why species sorting was the
386 main driver of metacommunity assembly, our study suggests that other factors related to

387 ecosystem size (e.g., spatial environmental heterogeneity) could further regulate
388 metacommunities.

389 The importance of species sorting can increase with environmental heterogeneity, i.e.,
390 the number of niches that are available for colonization across patches (45, 46). In our study,
391 large mesocosms contained greater spatial environmental heterogeneity evidenced by depth
392 associated changes in temperature and light. Hence, across mesocosms spatial environmental
393 heterogeneity could explain why species sorting was more apparent in large compared to small
394 mesocosms based on the network analysis. This might also explain the increased associations in
395 larger mesocosms, enhancing the probabilities for true biotic interactions. This increase may be
396 attributed to (i) the greater availability of niches (and consistency of nutrients) found in larger
397 mesocosms, or (ii) the synchronous establishment of bacteria which might have a better chance
398 in a stable environment. In a study of protists experiencing light-dark fluctuations in aquatic
399 microcosms and models, high fluctuations disrupted species synchrony between patches (47).
400 Spatial heterogeneity could also explain the greater bacterial richness as ecosystem size
401 increased.

402 Network topological features were partially influenced by mesocosm size: bacteria were
403 more connected in medium or large than small mesocosms, suggesting that abundance dynamics
404 were less similar across small mesocosms and indicating asynchrony among dominant bacteria.
405 In these less densely populated, large mesocosms, competition may have lessened which can
406 lead to greater synchrony between species that is driven by changes in abiotic conditions (44). In
407 our mesocosms, more connected, centralized communities with greater network density occurred
408 as size increased, indicating the presence of subnetworks or cliques and that, due to lower
409 fluctuation strength, the establishment of more connected, denser networks were common under

410 stabler environments (Fig. 4). Because size replicates in the network analysis spanned the
411 salinity gradient, this further suggests that the spatial environmental heterogeneity of salinity had
412 less importance for potential biotic associations among stabler mesocosms.

413 Taken together, we suggest that these patterns indicate mesocosm size-specific
414 mechanisms of species sorting: in small mesocosms, changes in community composition from
415 species sorting primarily occurred through taxa replacement in response to variation in multiple
416 environmental factors. In contrast, in larger mesocosms, environmental change was more gradual
417 and cascaded into compositional differences through abundant bacteria tracking environmental
418 changes over time by changing in relative population size, with lower replacement (Fig. 4).

419

420 *Recruitment of the members of bacterioplankton*

421 Initial community size due to differences in mesocosm volume might have affected the
422 resulting community composition through species sorting, but other factors related to the
423 experimental set-up were unlikely to have substantial influence. The experimental set-up ensured
424 no extensive differences in the recruitment of novel species from external sources due to
425 proportional mesocosm surface areas and equal initial sediment volumes. There were also no
426 differences in estimated richness of active bacteria between mesocosms on the first day of the
427 experiment. The dispersal sources (rain and air deposition as well as seed banks in sediments and
428 lake water) all harbored high diversity and in previous studies were shown to be important
429 recruitment sources for novel taxa following salinity disturbances (15-17) and other types of
430 environmental change (48). Although large mesocosms contained more total microorganisms
431 and thereby possibly a larger planktonic seed bank from which taxa could respond to the salinity
432 disturbance, recruitment from the water seed bank declined with salinity in all mesocosm sizes

433 (Fig. 4). Even with reduced dispersal, future studies that extend temporal sampling beyond the
434 64 days sampled here may eventually see eco-evolutionary processes such as increased tracking
435 of environmental conditions in small mesocosms due to bacterial diversification, which can be
436 intensified by a history of environmental adversity (49). The high percentage of ASVs for which
437 we did not identify a source (85%) could indicate dispersal from other sources such as the snails
438 we observed on most mesocosms. or the effect of sequencing depth which can miss the rarest
439 taxa.

440

441 *Roles of stochasticity and dispersal-related processes*

442 Demographic stochasticity (leading to ecological drift) was an important driver of
443 community assembly of all mesocosms via community size with the strongest direct effect in
444 small mesocosms (Fig. 4). This result is bolstered by previous studies showing that ecological
445 drift more often occurs in small communities (50, 51) especially when the importance of species
446 sorting is weak (52) or when the effective community size is small due to dispersal limitation
447 (53). This may be why we did not detect synchronous environmental tracking from the dominant
448 populations across the small mesocosms. Drift can also alter the outcome of niche selection (54).
449 Nevertheless, the effect of time on community composition indicated that large mesocosm
450 communities were most influenced by demographic stochasticity arising from temporal
451 influences. In this case, large mesocosms may more strongly reflect (i) random changes in births
452 and deaths from a community that grew in number over time, (ii) stochasticity based on priority
453 effects from slower time-delayed tracking, or (iii) may reflect sampling timepoints that
454 underrepresented the larger total community.

455 Dispersal limitation as driver of metacommunity dynamics (considering all mesocosms at
456 one time point) was present only in small sized mesocosms and suggests that multiple
457 communities emerged from similar initial conditions in the small mesocosms. However, the
458 interpretation of the dispersal limitation is ambiguous (e.g.(55)). It could be true dispersal
459 limitation whereby niche spaces that are opened (i.e., when species become inactive in response
460 to the initial salinity changes which increase habitat specialists (56) and/or the strong
461 environmental changes) remain empty (57). However, it does not necessarily indicate true
462 dispersal limitation between patches (55) or reduced immigration from a regional pool. Instead,
463 it can be explained by the low richness of these mesocosm communities decreasing the
464 likelihood that they contain superb dispersers. When dispersal rates are low, local adaptations to
465 environmental fluctuations can lead to strengthened priority effects by preemptive taxa (58),
466 which might have occurred during the experiment. For example, the many time-delayed
467 associations between salinity and bacterial taxa in medium mesocosms could be a sign of
468 transient priority effects where taxa (i.e., Bacteroidota and Proteobacteria but not Cyanobacteria)
469 maintained abundances for a short period without environmental tracking. Nevertheless, with our
470 data and the applied approaches, it is not possible to clearly support or exclude assembly
471 processes and other factors that regulate them.

472

473 *Conclusions*

474 Overall, our results partially align with those from previous studies which show that after
475 disturbances, stochastic community assembly initially is important, but the dominant influence
476 shifts to deterministic processes in later successional stages (e.g. (59)), especially when
477 environmental conditions are stable. Dispersal limitation and ecological drift (demographic

478 stochasticity) were drivers of metacommunity dynamics after community establishment with
479 strong environmental fluctuations. Mesocosms with reduced environmental fluctuations may
480 facilitate considerable time-delayed species sorting and thus possibly, a transient influence of
481 priority effects. The novelty of our study is that we could show, by applying both path and
482 network approaches, that the trajectories of (meta)community development are influenced by
483 size-induced environmental fluctuations in concert with a salinity press disturbance. Our results
484 represent the advantage of joining a network analysis together with metacommunity models, and
485 stress that environmental fluctuations are important to consider in future community assembly
486 studies as they can modify community assembly under natural conditions.

487

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504

505 **Data accessibility statement:** The data supporting the results are archived in the public
506 repository European Nucleotide Archive with accession number PRJEB26595 and
507 environmental data are made available in the Swedish institutional repository, DiVA, (diva-
508 portal.org) with the following accession number: diva2:1210995.

509

510 **Competing Interests**

511 The authors declare no competing interests.

512

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659

660 **Figure Legends**

661 **Figure 1.** Temporal patterns of alpha diversity metrics for bacterial communities in dispersal
662 sources (air and rain), source media (sediment and water) (DNA) and mesocosm water (RNA).
663 Error bars are standard error. Diversity metrics for large mesocosms greater than both small and
664 medium mesocosms (repeated measures ANOVA, pairwise t-test with Bonferroni correction, $p <$
665 0.05). Source media $n = 3$, mesocosm sizes $n = 16$, air and rain $n = 1$. Note difference in y-axis
666 scales.

667
668 **Figure 2.** Path analysis diagrams of small, medium, and large mesocosm sizes. The influence of
669 spatial distances (Δx), temporal distances (Δt), environmental distances (ΔE), mean community
670 size ($\langle J \rangle$), absolute difference in community size (ΔJ) and species richness (ΔS) on community
671 dissimilarity (β_{bc}) was quantified following Jabot et al.'s framework (2020). Arrow width
672 represents standardized estimate strength with positive estimate arrows in solid lines and
673 negative estimates in dashed lines. For environmental variables, the absolute values of
674 standardized estimates were added. Effects shown have $p < 0.05$. SRMR = Standardized Root
675 Mean Square Residual. See Tables S4-S6 for standardized estimate values.

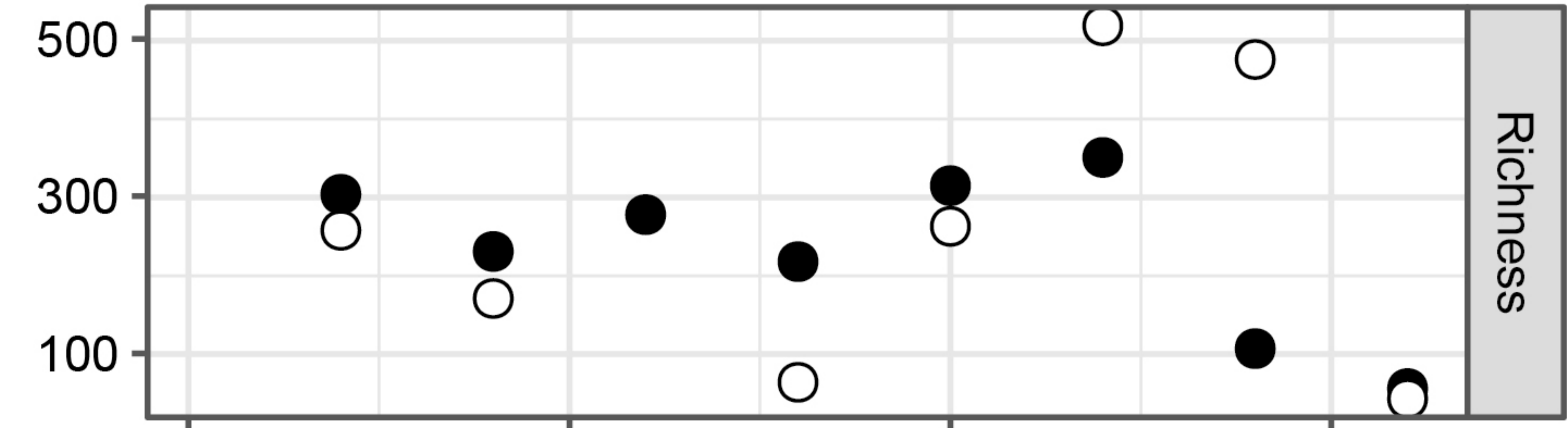
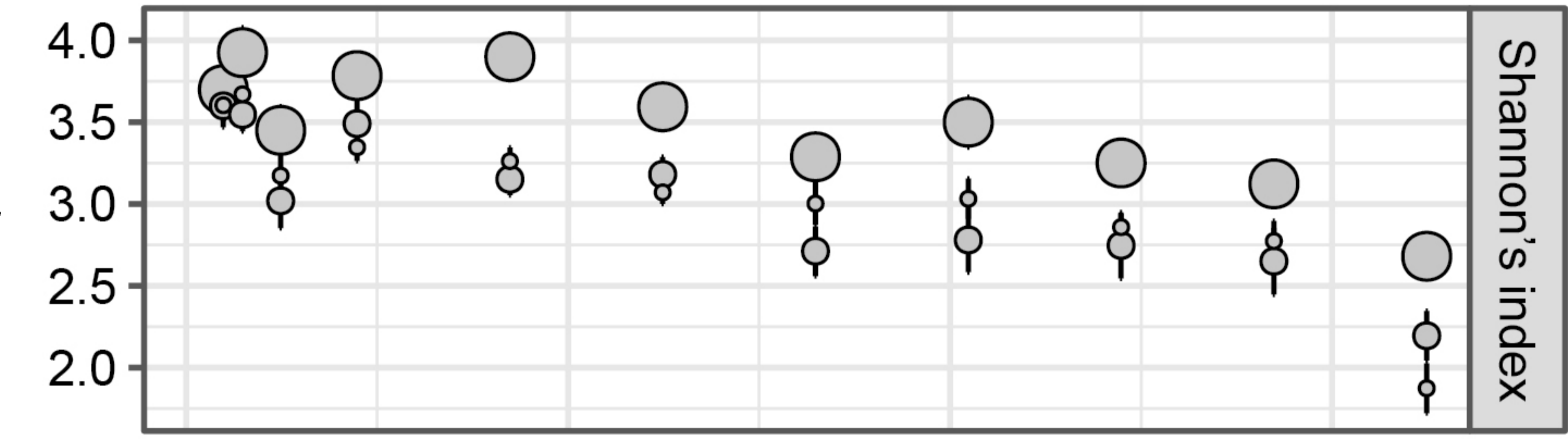
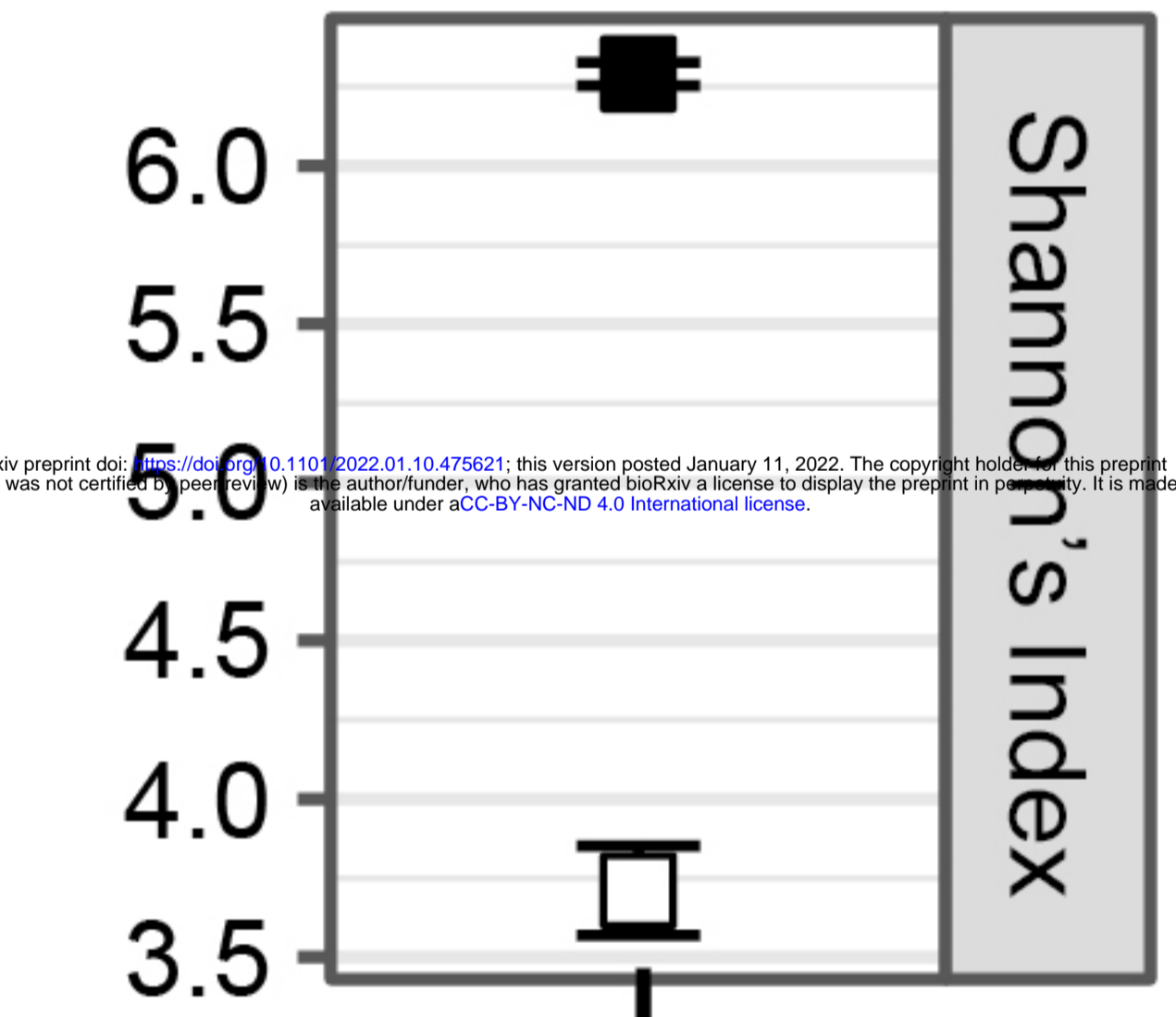
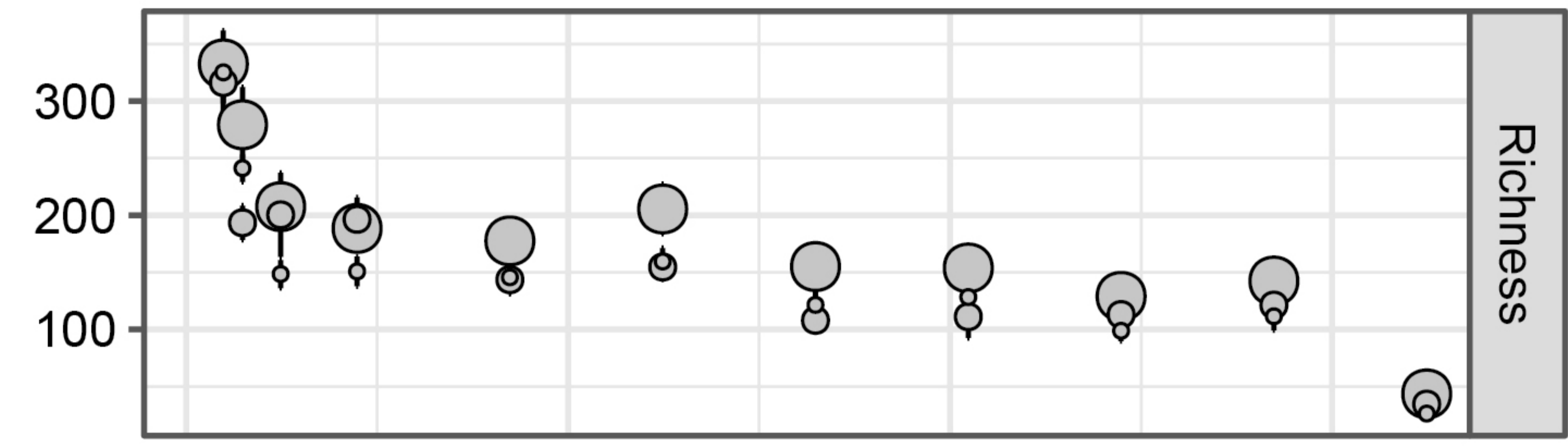
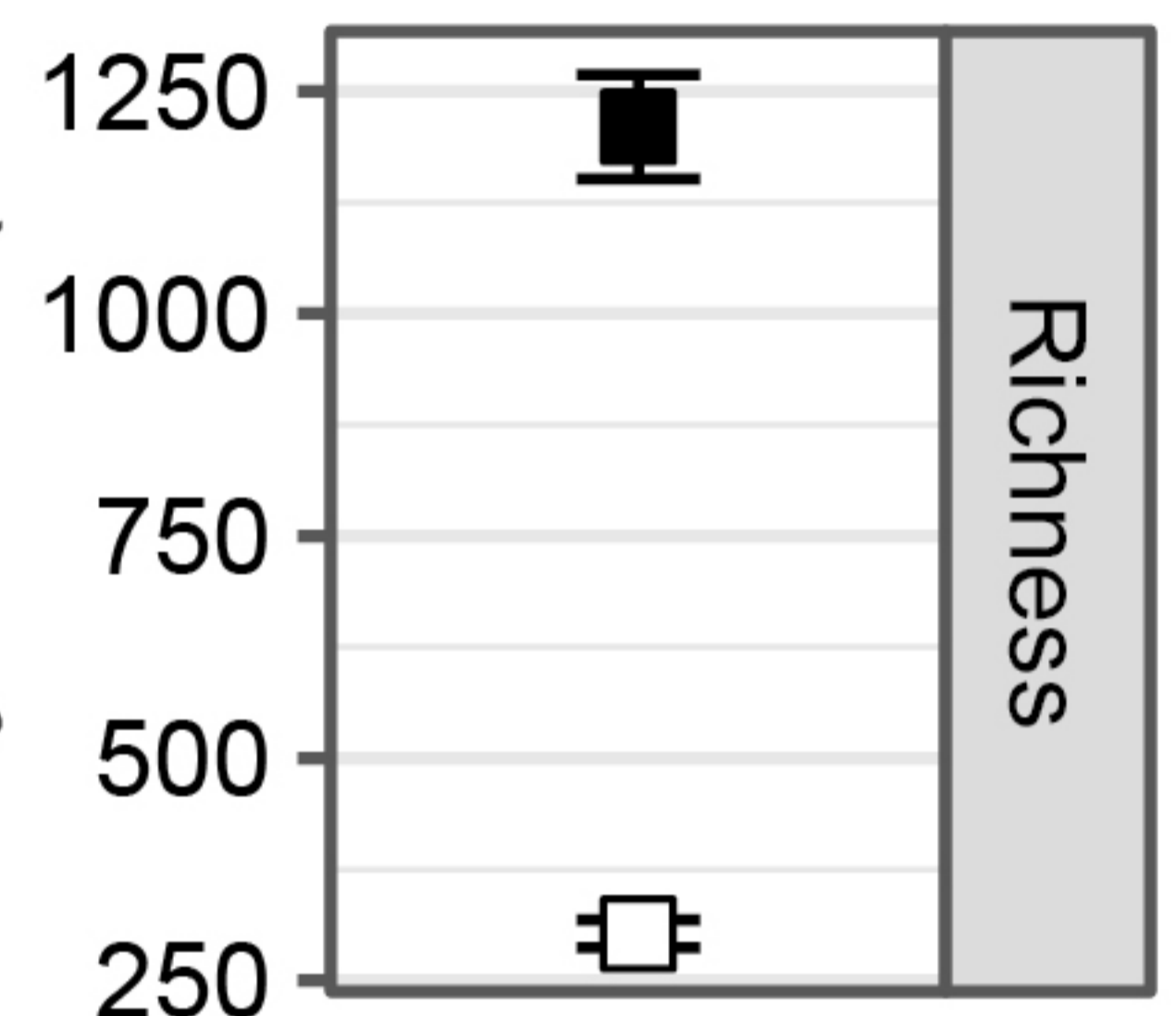
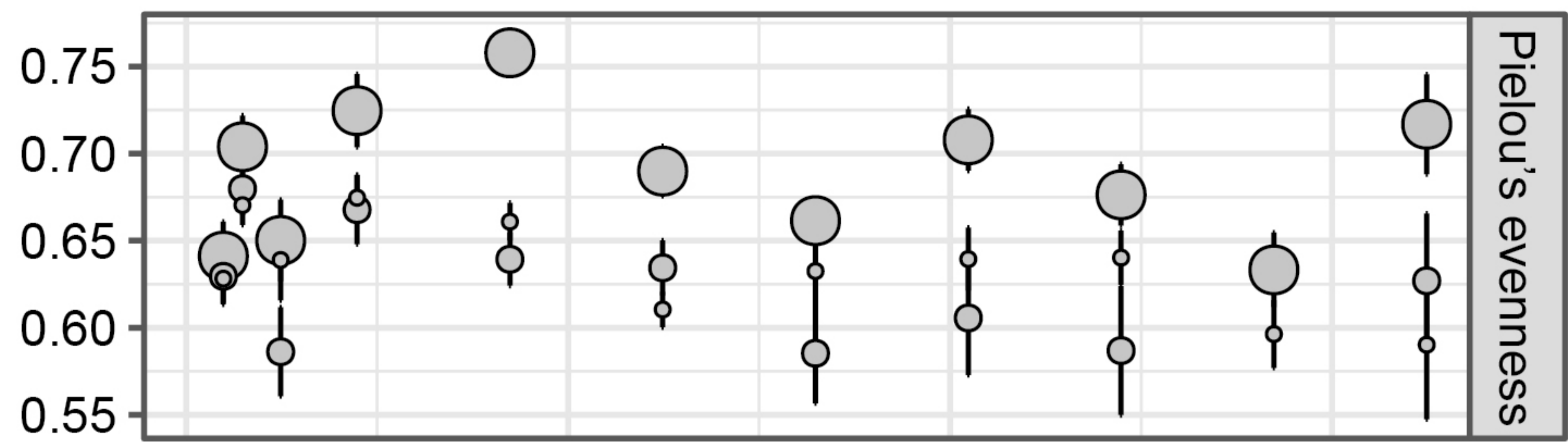
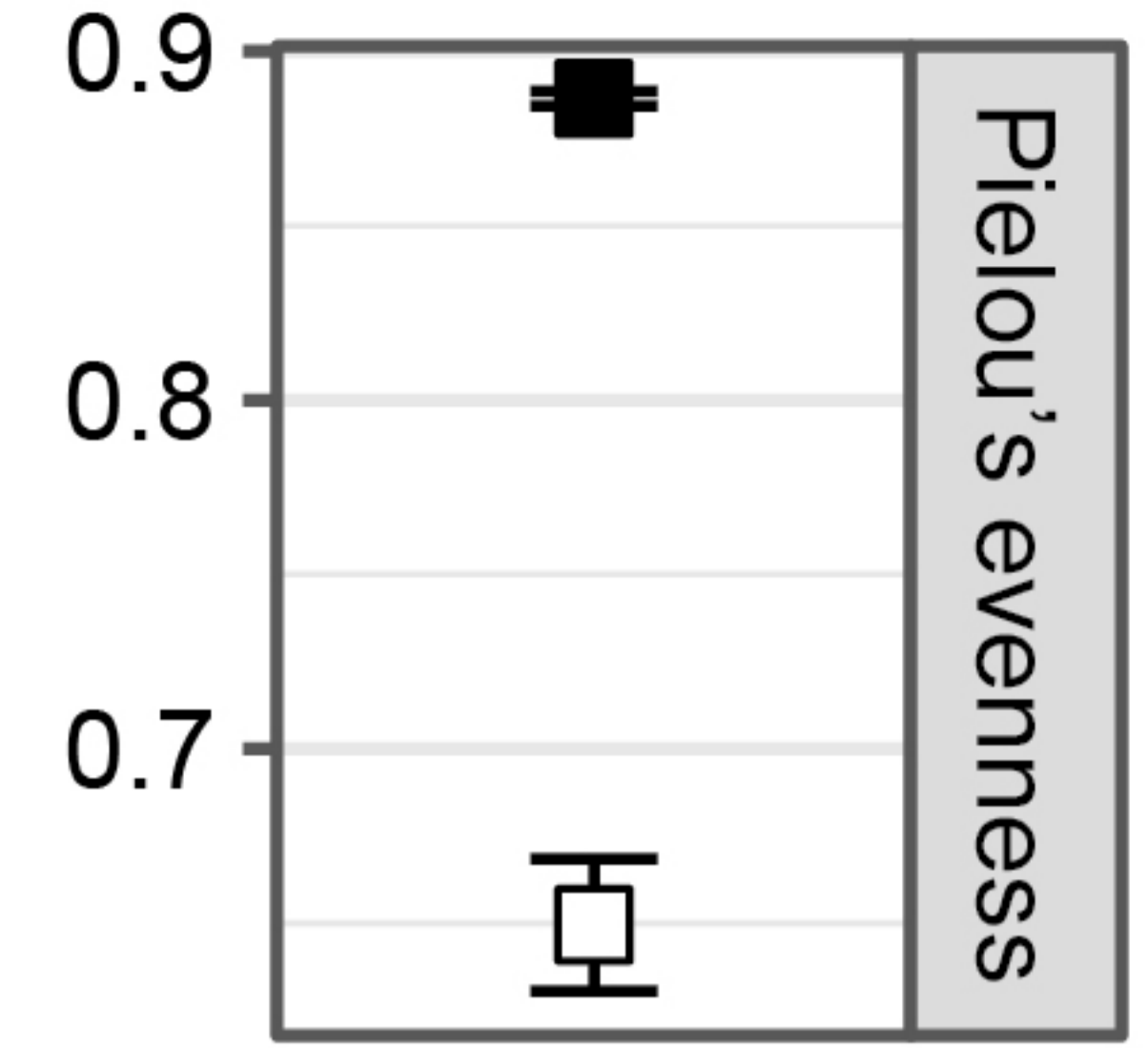
676
677 **Figure 3.** Association networks and the relative abundances of the 50 most abundant bacteria of
678 the three mesocosm size categories ($n = 16$). All significant ($(p \leq 0.01$ and $Q \leq 0.01)$) pairwise
679 local similarity (LS) correlations ≥ 0.05 are shown as edges in the networks. Each node
680 represents an ASV (ellipse) or an environmental factor (rectangle). Edge transparency is
681 proportional to the association strength (based on LS values). Solid lines refer to positive
682 associations while dashed lines to negative ones. Edge colors indicate delayed (blue) and non-

683 delayed (black) associations between ASVs and/or environmental variables. Arrows point
684 toward the lagging node.

685

686 **Figure 4.** Conceptual figure for the interpretation of statistical results and patterns based on path
687 analysis, network analysis, and the partitioning of beta-diversity. In our study, the dominant
688 deterministic force was the applied salinity press disturbance. Ecosystem size was manipulated
689 by different volumes of mesocosms. Darker shading in bars indicates greater influence of the
690 process.

Diversity Value (mean)



Mesocosm Size

- Large
- Medium
- Small

Dispersal Source

- Air
- Rain

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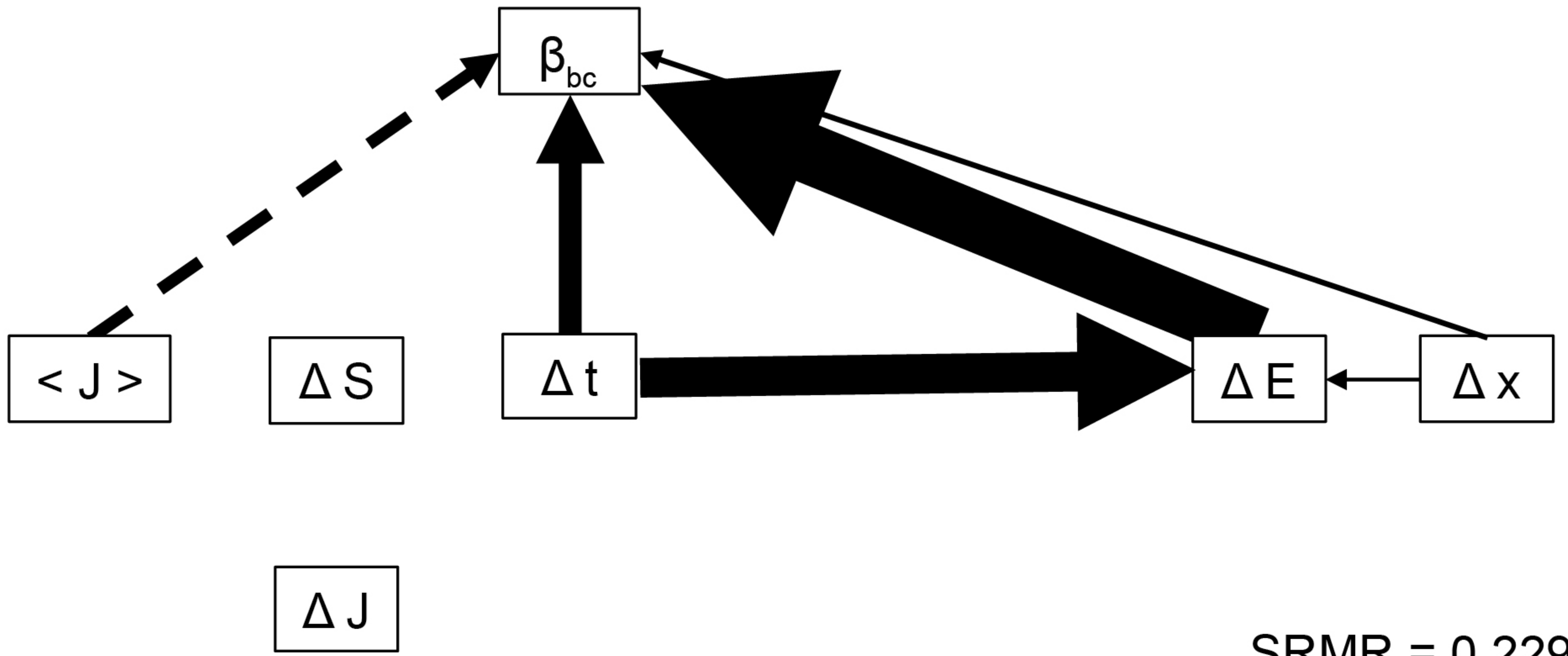
Time (days)

- Source sediment
- Source water

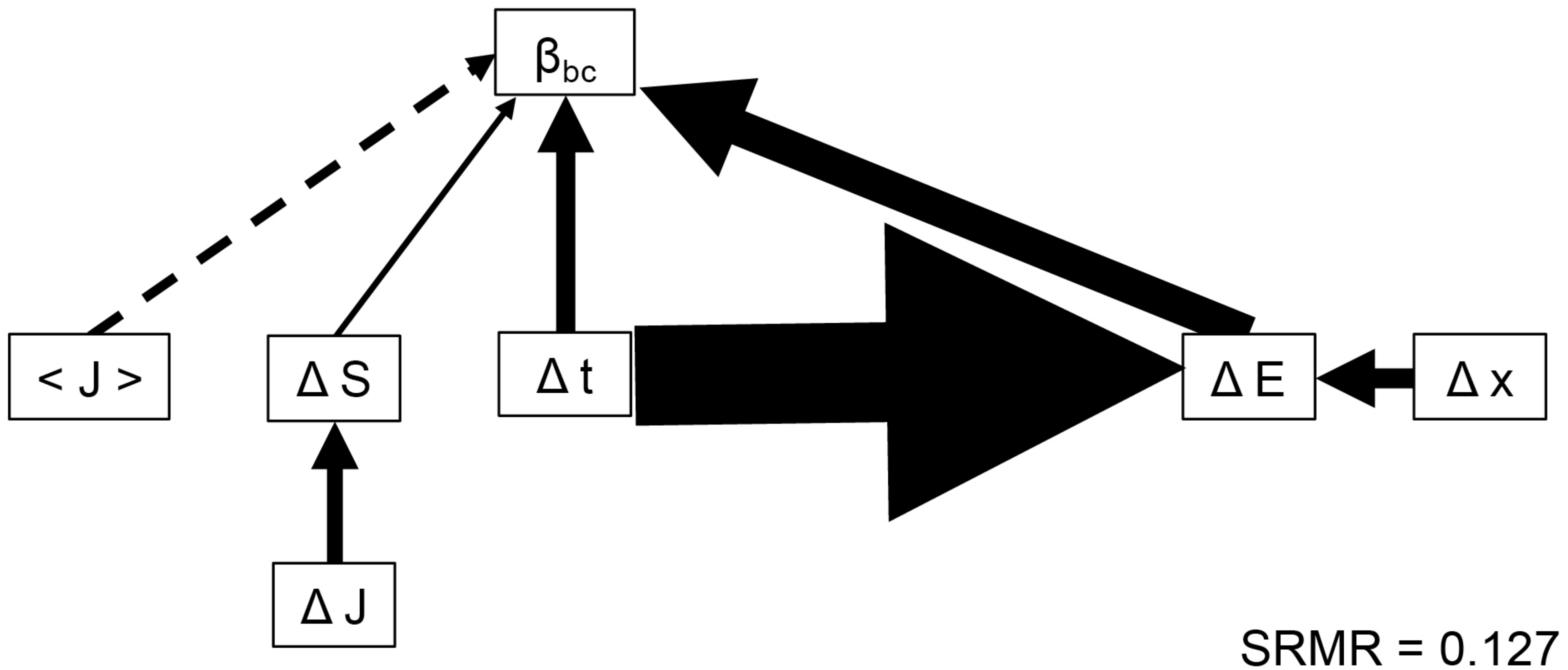
Time (days)

Richness

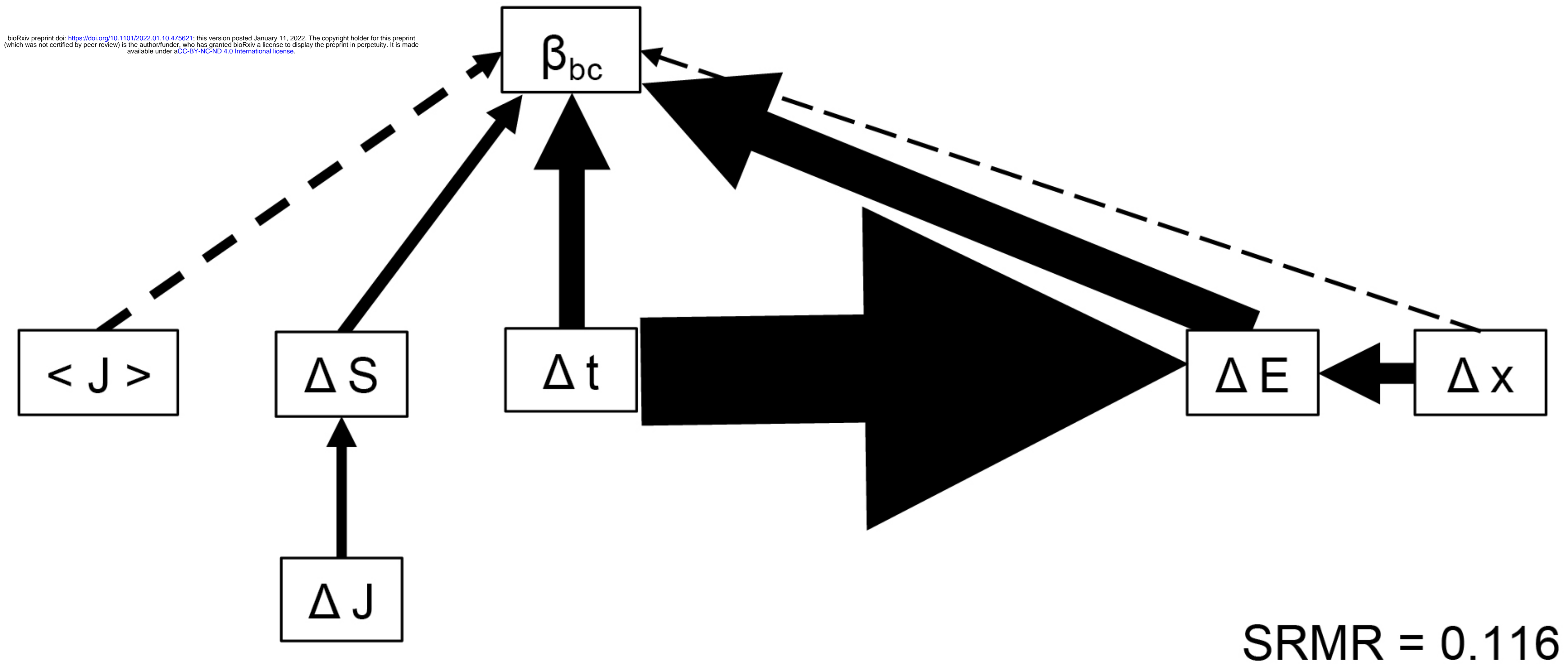
Small Mesocosms



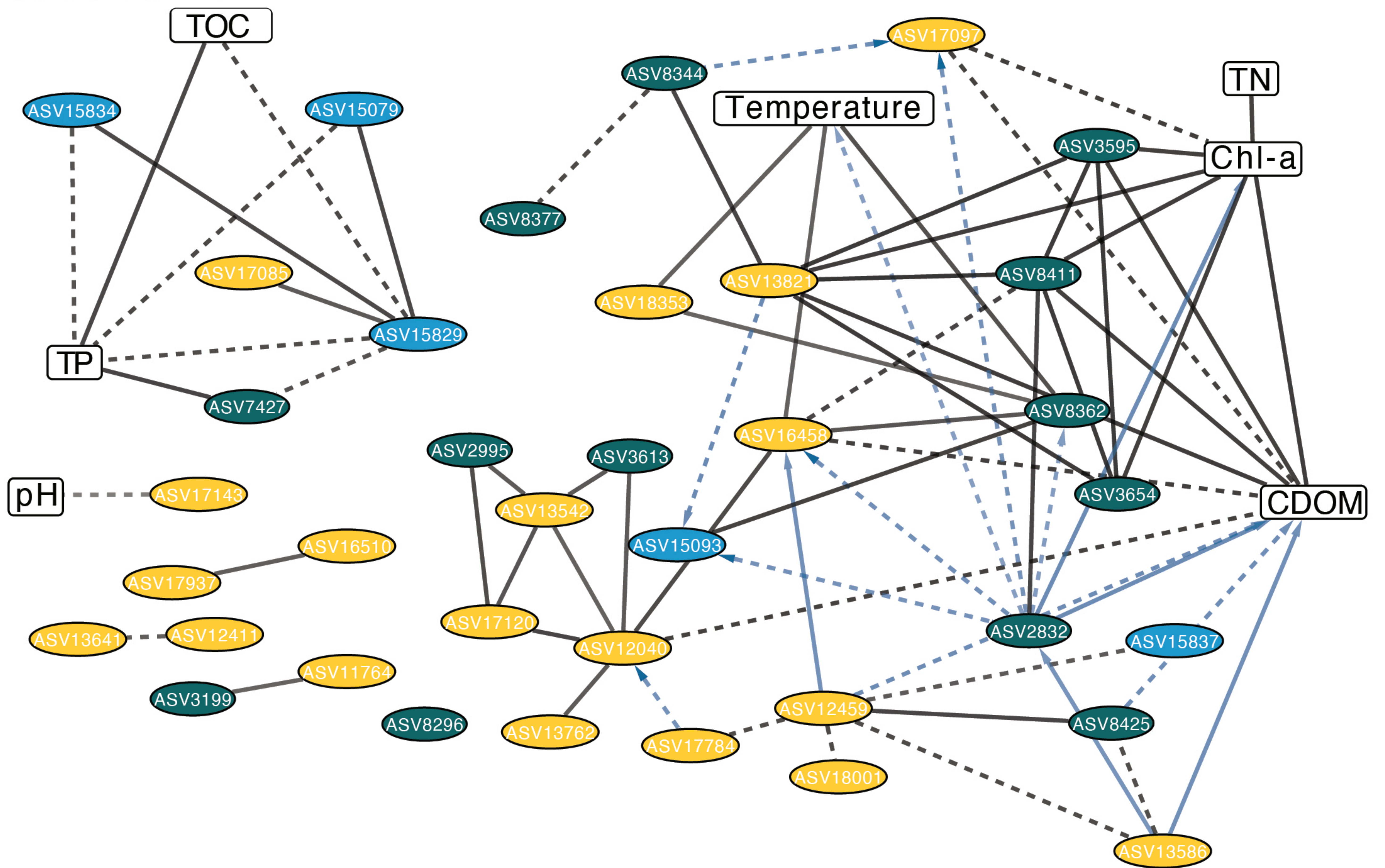
Medium Mesocosms



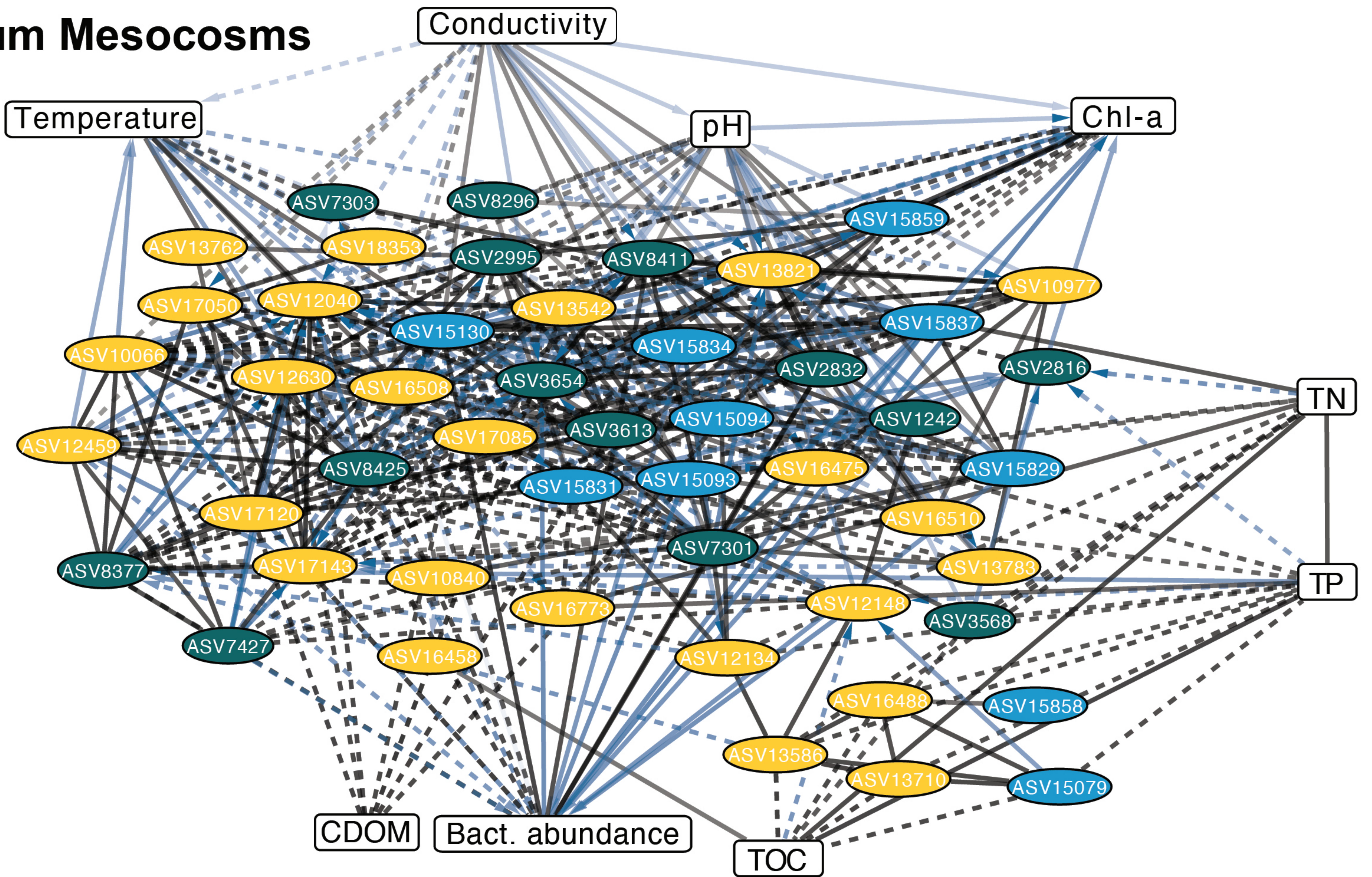
Large Mesocosms



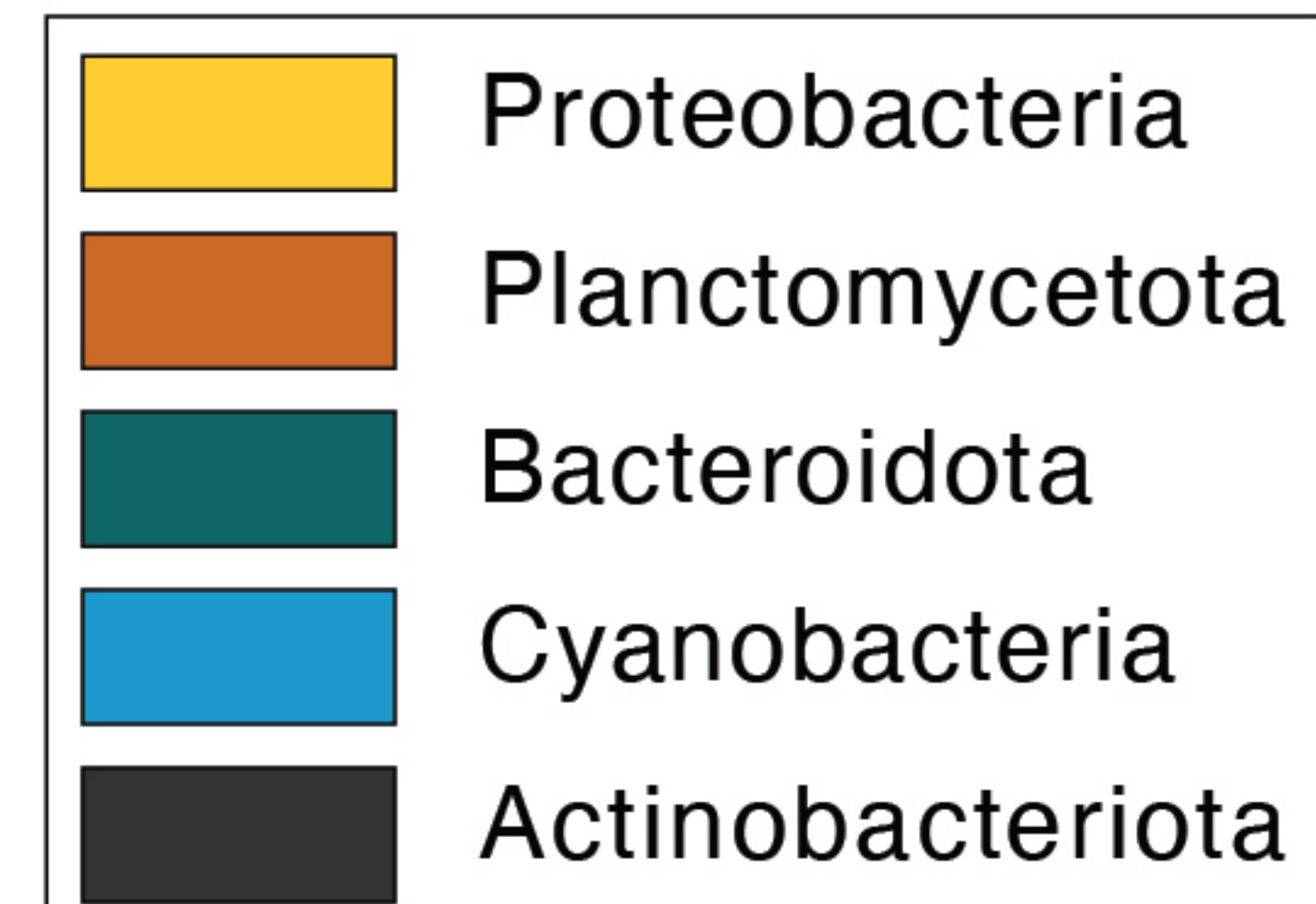
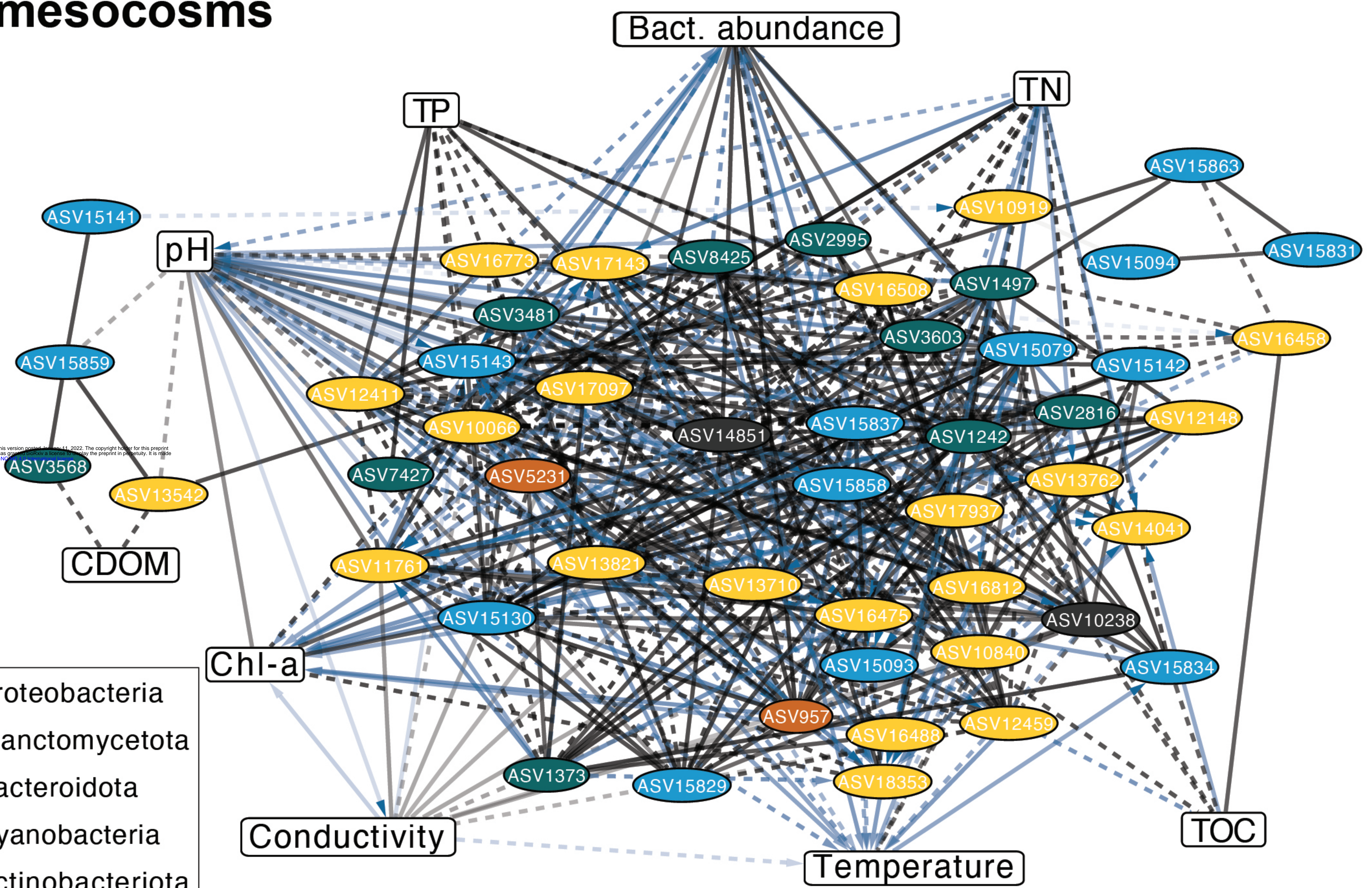
Small Mesocosms



Medium Mesocosms



Large mesocosms



← Environmental fluctuation →

Deterministic force

[salinity press disturbance]

Network properties

→ Ecosystem size →

bacterial richness

turnover ↔ nestedness

species sorting

recruitment from water

recruitment from sediment

demographic stochasticity

dispersal limitation

biotic interactions & transient priority effects

Network centralization

Time-shifted associations

Network connectivity, nr. of nodes & edges

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