36-year study reveals the effects of microhabitats on a wild wheat population

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Abstract

Long-term genetic studies of wild populations are very scarce but are essential for connecting ecological and population genetics models, and are also necessary for understanding how climate change can affect biodiversity. We present a study of the wild ancestor of modern wheat, carried out over a 36-year period during which atmospheric CO_2 increased by 70 ppm and temperature by 1.5°C. We genotyped 832 individuals from regular sampling along transects during the course of the experiment. Genotypes were clustered into ecological microhabitats over scales of tens of meters, and this clustering was remarkably stable over the 36 generations of the study. Simulations indicate that neutral processes alone are insufficient to explain this spatial and temporal stability, suggesting a role for fine-scale local adaptation to ecological parameters. Using a common-garden experiment, we showed that the genotypes found in distinct microhabitats differ phenotypically, further supporting the hypothesis of local adaptation. Our results provide a rare insight into the dynamics of a natural population at the scale of individuals over a period of environmental change.

Introduction

Climate change is a looming threat to plant biodiversity and food security (Hoffmann and Sgró, 2011; Pecl et al., 2017; Urban, 2015). This increases the importance of wild progenitors of crops because their genetic diversity is an immense reservoir of biotic and abiotic stress-resistance genes (Khoury et al., 2021). Sustained studies of progenitor species in their natural ecological niches can provide a better understanding of the likely impacts of climate change, and inform strategies for conservation of such resources *in-situ* as well as in gene banks (Wintle et al., 2019). We examined the genetic diversity of a population of wild emmer wheat (*Triticum turgidum ssp. dicoccoides*) in Ammiad, Israel, which has been monitored for 36 years (Figure 1A). The original motivation for studying this particular population long-term was its high genetic diversity compared to other wild wheat populations (Felsenburg et al., 1991; Levy and Feldman, 1988), together with variable topography (Figure 1A, Movie S1) and a diversity of ecological niches. These parameters were ideal for investigating adaptation to microhabitats and to analyze the dynamics of the population over time in order to assess strategies for *in-situ* conservation. When the experiment started in 1984, climate change was not yet an established fact. In retrospect, the increase in local ambient temperatures by ~1.5 degrees Celsius (Figure 1B) and in atmospheric CO_2 concentration by 70 ppm (Figure 1C) (Keeling et al., 2001) during the course of the experiment, enabled us also to assess the effect of climate change on the genetic makeup of the population. For example, we could test if the overall genetic variation of the population or the spatial distribution of genotypes had changed. The study was also motivated by the fact that wild emmer wheat biodiversity is an invaluable resource for food security. Wild emmer wheat, a self-pollinating,

allotetraploid annual cereal (genome BBAA), is the progenitor of durum wheat (*T.*

turgidum ssp. durum) and the maternal donor of the A and B sub-genomes of bread

wheat (*T. aestivum*, genome BBAADD)(Levy and Feldman, 2022), and any gene from

wild emmer wheat can be easily transferred to the background of domesticated wheat.

Dense population transects were regularly sampled nine times between 1984 and

2020, and each plant was genotyped by sequencing. Earlier spatio-temporal studies

of wild plants involved shorter periods, as well as less comprehensive or lower

resolution sampling (Franks et al., 2007; Fu et al., 2019; Horovitz and Feldman, 1991;

Kolis et al., 2022; Ozbek et al., 2007; Todesco et al., 2020; Volis et al., 2004). Indeed,

long-term studies of species in their natural habitat are exceedingly rare. In animals,

Soay sheep and Florida scrub-jays have been studied for periods ranging between 23

to 36 years (Aguillon et al., 2017; Ashraf et al., 2021; Chen et al., 2019; Stoffel et al.,

2021), comparable to our study in years (but not generations). In plants, detailed

studies of population structure did not cover more than eight years, and most did not

involve consecutive samplings (Frachon et al., 2017; Hendrick et al., 2016; Kolis et al.,

2022). Without such studies, it is difficult to connect abstract population genetics

models of allele-frequency change with real-world temporal and spatial scales. Our

dataset provides an almost unique insight into genetic change in a natural population

over time and space.

Results

In order to analyze the structure of the Ammiad population over time, 100 locations were marked by pegs and by GPS coordinates, along four transects traversing seven ecologically distinct microhabitats (Anikster and Noy-Meir, 1991) (Figure 1A, Dataset S1, Movie S1). Seeds were sampled within 1 meter of each location at each of nine time points, and sown in the greenhouse for genotyping. In total, we genotyped 832 individuals using a genotyping-by-sequencing (GBS) protocol (Poland et al., 2012), resulting in 5196 single copy, single-nucleotide-polymorphism (SNP) markers that were unambiguously positioned on the wild emmer wheat reference genome (Avni et al., 2017).

We found that the population is almost entirely composed of genetically distinct groups of nearly identical, highly homozygous individuals, clustered in space. Pairwise genetic identity follows a multimodal distribution that is characterized by pairs of plants that are either effectively identical, or share at 70-90% of SNP alleles (Figure 1D). We unambiguously assigned all individuals to distinct genotype groups (DGGs) using a cut-off of >98% allele-sharing. This resulted in 118 DGGs over all nine time points, including 61 singletons (DGGs with a single member). The 10 largest DGGs encompass 65% of the samples, while the numbers of DGGs per year varied between 32 (2014) to 49 (1988; Dataset S1). Rarefaction curves indicate that the sampling scheme was sufficient to detect almost all of the DGGs present close to the transects (Figure S1). These results provide a detailed picture of the genetic structure of the population through time and space.

Figure 1. Ammiad population study design and spatial structure. (A) Location of Ammiad near the lake of Galilee (star, top left and geographic coordinates, bottom right). Sampling points of transects A, B, C, D collected in the years 1984, 1988, 1992, 1996, 2002, 2014, 2016, 2018 and 2020. Sampling points are shown on-site for transects A, B, D, but transect C, shown in the dashed rectangle, is located ~500 meters north-west of A, B, D. Sampling points colors correspond to the different microhabitats labeled in the legend. The map on the right is from gov.map.co.il, and the map on the left from Google Maps. (B) Average yearly minimum (top) and maximum (bottom) temperature measured in Ayelet Haschahar weather station from 1984 to 2020. Collection years are marked in red on the linear model regression line. (C) CO_2 concentration 1983 to 2021. Data obtained from the Mauna Loa record, Scripps Institution of Oceanography, UC San Diego. (D) Distribution of pairwise identity values

between the plants in the collection. (E) Population structure of Ammiad in the first year of

collection (1984). Heatmap colors represent identity value for each pairwise comparison,

annotation colors (on the left) correspond to the list of microhabitats in Figure 1A.

Though outcrossing is thought to be relatively low in wheat (0.5%, Golenberg, 1988)

we also identified 15 (1.8%) highly heterozygous plants in Ammiad collection. Using

SNPMatch (Pisupati et al., 2017), we were able to identify their parental DGGs (Figure

S2B), which suggests that outcrossing happen between plants that are nearby. Based

on F_{IS}, we estimated an average outcrossing rate of 3.7% across the years in Ammiad,

with detectable variation between the different habitats (1.8-8.8%, Table S1). Such

ranges are compatible with the estimated values (2-9%) in the literature of bread wheat

(*Triticum aestivum*) using the same approach (Enjalbert and David, 2000).

Interestingly, we also observed a significant two-fold decrease (p-value 0.016) in the

outcrossing rates over time, from about 4.5% in 1984 to about 2% after 2016 (Figure

S2A).

Returning to the inbred individuals, the spatio-temporal distribution of DGGs shows

that plants of the same DGG tend to cluster together in both space and time (Figures

1E and 2). This clustering was also found in a principal component analysis (Figures

3A and 3B). Although there is little evidence for isolation by distance over the years of

collection and along the transects (Figure 3C and Figure S3), within years, the

probability of belonging to the same DGG decays rapidly with distance (Figure 4A).

Overall, we find a striking concordance between DGGs and microhabitats, which were

defined based on topographic and floristic data without knowledge of the genetic

structure of the population (Noy-Meir et al., 1991b) (Figure 2). For example, in transect

A there is a sharp transition between DGGs of the upper and middle north-facing slope

(Figure 2A). Another example is the genotype at collection point A59 in a unique

location under the shade of a *Ziziphus spina-christi* tree (Figure 2A). In contrast, in

transect D, where there is negligible microhabitat differentiation, plants of different

DGGs were much more mixed (Figure 2D). The spatial pattern was highly stable over

the course of the experiment. In contrast to the rapid spatial decay, the probability of

belonging to the same DGG decays only slowly over time (Figure 4B). For example,

33% of plants sampled in 1984 belonged to the same DGG as plants sampled at the

same position in 2020. In conclusion, our results show a strong spatial structure,

associated with microhabitat, which has remained stable for at least 36 years despite

changes in climatic conditions.

Figure 2. Spatio-temporal distribution of genotypes. (A-D) Distribution of samples colored

by distinct genotype group (DGG) in transects A, B, C and D. Y-axes indicate year of collection

(left) and altitude for 1984 (right). Singleton samples are marked in small black dots, while

doubleton groups are marked with "2" in the middle with the color of their closest DGG.

Doubleton groups that are closest to the same DGG were marked with colored asterisks. (E)

Dendrogram of DGGs with ≥ 3 samples, hierarchically clustered by UPGMA.

Figure 3. Principal Component Analysis (PCA) and isolation by distance of Ammiad collection. (A) PCA colored by microhabitat. (B) PCA colored by DGGs, black dots represent singleton DGGs. (B) Scatter plots of pairwise genetic identity and Euclidean distance (meters). Subplots show years of collection, and colors represent the different transects (A=blue, B=red, C=yellow, D=green).

Figure 4. Spatial clustering and temporal stability at Ammiad. Points show the probability

of observing genotypes of the same DGG at (A) pairs of sampling points in the same year at

increasing distances along each transect and (B) the same sampling point between pairs of

years. Grey regions indicate 95% confidence intervals over populations simulated under

conditions closest to empirical estimates (3 plants/m^2, mean seed dispersal of 1m, 0.5%

outcrossing per generation, 30% seeds germinating in the second year after dispersal). In A,

points show the average for sampling points in bins of 5 meters. (C) Sorting of plants of the

same DGG by microhabitat comparing observed values (red) to values obtained by permuting

habitat labels (violins).

The observed patterns are far from random. Across transects, pairs of the same DGG

were significantly more likely to be found in the same microhabitat in the observed

dataset than in 1000 permuted datasets where the order of microhabitats was

randomly shuffled (Figure 4C), or even when we control for distance and spatial

autocorrelation using partial Mantel tests and Moran-spectral-randomization Mantel

statistics (MSR-Mantel) (Figures 5A and 5B). Similarly, population differentiation

measured by F_{st} was consistently higher between microhabitats than within, even

when taking distance into account (Figures 5C and 5D), and genetic distances are

strongly correlated with microhabitat. Furthermore, an analysis of molecular variance

(AMOVA) showed that 28% of the genetic variation between samples is explained by

differences between microhabitats (Table S2).

Motivated by these results, we decided to test whether it was possible to explain the

patterns using explicit demographic models that incorporate both self-fertilization and

limited dispersal. The question is whether these neutral demographic processes - on

their own - could generate the strong spatial clustering and temporal stability observed

at Ammiad.

Figure 5. Genetic differentiation between and within habitats, accounting for spatial distance. (A-B) Histograms of 1000 spatially constrained randomizations of Mantel statistics (MSR-Mantel), testing for the role of habitats in determining the spatial autocorrelation in Ammiad. Observed values (black dots) versus randomizations for (A) pairwise genetic distances (p-value < 0.001) and (B) DGGs (p-value < 0.001). (C-D) Plants within each habitat were grouped in 30m demes. Pairwise F_{st} was calculated among all demes, within and between habitats. (C) Pairwise F_{st} plotted as a function of distance, measured as the average distance of plants in the two demes. Credibility intervals of a loess fit with degree 1 are shown in gray. (D) Pairwise F_{st} between demes with average distance <50m. Average observed values are shown as a black dot. Uncertainty was estimated with 1000 bootstraps of the parcels considered for the comparisons. Average within-habitat Fst was higher than between-habitat F_{st} in 3% of the bootstraps. For both C and D, Nm values (i.e. the theoretical number of migrants between parcels in an infinite island model) calculated are shown on the right axis.

We performed individual-based simulations of populations of plants in a homogeneous environment, evolving by seed dispersal and outcrossing only. We ran simulations using different values of outcrossing rate, plant density, seed dormancy and mean seed dispersal distance that approximate a range of biologically realistic values. Empirical values of outcrossing rates in wild wheat have been estimated to be approximately 0.5% (Golenberg, 1988), and we conservatively use this value here although our estimates above point to an outcrossing rate of around 3.7%, with up to 30% of germinating seedlings coming from the seed bank (Horovitz, 1998). Average plant density in the Ammiad population was measured in 1984 as 4 plants/m^2 (Noy-Meir et al., 1991a) and again in 2020 for transect A as 3.8 plants/m^2 (Dataset S1). We lack direct estimates of seed dispersal, but average gene flow, a composite of seed and pollen dispersal, was estimated to be 1.25 m per year (Golenberg, 1987). Seed dispersal occurs through a combination of gravity, ants, birds, grazing mammals and rodents, enabling both short and long-distance dispersal. For example, the appearance of a DDG prominent in transect C (Figure 2C, marked by yellow dot), in transect D in 2020 (Figure 2D, locations D198, D217), more than 700 meters away, demonstrates that long distance movement occasionally occurs.

In simulations with parameters supported by observations (1 m mean seed dispersal, 0.5% outcrossing, density of 3 plants/m2 and 30% seed dormancy) we find that neutral demographic processes alone could not account for the observed spatial clustering and temporal stability (Figure 4), in line with other studies (Nevo, 2009). However, the simulation could produce results consistent with the data if we assumed extremely low plant densities (≤1 plant/m2) or rates of gene flow (outcrossing < 0.5% when seed dispersal ≤ 1m) (Figure S4). Without better estimates of these parameters or direct

fitness estimates from reciprocal transplant experiments, it is not possible to definitely

reject a purely neutral explanation for the observed patterns. That said, the

correspondence between DGGs and microhabitats is strong (Figure 2, Figure 4C,

Figures 5C and 5D), and certainly not compatible with the outcrossing rates measured

in this experiment. Thus, the simulations indicate that it would be difficult to generate

the observed patterns of stability over 36 years through neutral demographic

processes alone.

In order to further examine the correspondence between DGGs and microhabitats, we

phenotyped 95 DGGs in a six-randomized-blocks common-garden experiment (Figure

6A). In order to reduce the genotype x environment effects as much as possible, plants

were sown in early November in their native soil brought from Ammiad site with no use

of fertilizers. We focused on measuring phenotypes related to yield and fecundity

(Figure S5).

Figure 6. Phenotypes measured in the common-garden experiment. (A) Panoramic picture of the experiment site and close up pictures inside the net house in Weizmann Institute of Science. (B) Phenotypic value (y-axis) in all growing conditions (blocks) for each genotype, sorted along the mean phenotypic value of each genotype (x-axis).

We estimated the heritability of each phenotype leveraging the correlation of each genotype's phenotype across blocks, calculating the proportion of total variance explained by the genotypes vs. the residual and that explained by blocks (Figure 6B and Figure S6A). Almost all phenotypes exhibited moderate to high heritability (23-78%), with seed length the most heritable (78.0+/-3.7%). The heritability of

germination time was indistinguishable from 0, and it was therefore excluded from

further analyses (after being used to calculate anthesis time, which instead showed a

heritability of 50.0+/-14.5%; Table S3).

We then aimed at testing whether habitat could explain the distribution of phenotypes,

while taking the distribution of genotypes into account. We assigned genotypes to

different habitats, considering a genotype representative of a habitat when at least

30% of the samples were found in that specific habitat (Table S4). For each trait, we

then tested whether habitat was making a significant contribution in explaining the

variation between phenotypes belonging to different genotypes (see Methods). After

a Bonferroni- correction for multiple testing, we observed that habitat was significantly

explaining differences in coleoptile color (p-value 0.020) and plant base pigmentation

during tillering (p-value 0.034). Other traits were non-significant, though a Stouffer's

test, corrected for the non-independence between phenotypes (Figure S6B), supports

the hypothesis that habitats explain differences between phenotypes in Ammiad

(Stouffer's method with Strube's adjustment p-value=0.0118). A linear discriminant

analysis confirms that coleoptile color is the most relevant trait for distinguishing

between habitats, followed by main tiller length and number of tillers (Figure S7). In

particular, the first dimension of the linear discriminant analysis, which explains 42.1%

of the variance explained by the analysis, separates habitats with more pigmented

plants and longer seeds and plants (valley and south-facing slope) from others (Figure

S6C, Table S5).

Discussion

Studies of natural populations rarely have fine-scale geographic resolution, and almost never have temporal resolution beyond the duration of a PhD (Siepielski et al., 2009). Our study, spanning over 36 years of dense sampling, thus provides a unique dataset that could improve our understanding of gene flow and local adaptation - often modeled quite abstractly in population genetics - on an ecological scale.

Our analysis revealed that the Ammiad population varies more in space than in time. We found unexpectedly strong long-term genetic stability, demonstrating that gene flow can effectively be highly limited in a wild cereal despite its wide distribution throughout the fertile crescent. Remarkably, the stable genetic clusters observed in Ammiad matched ecological microhabitats previously identified at the site, suggesting that natural selection, in addition to limited dispersal, contributed to shaping population structure. Previous studies of wild populations have demonstrated local adaptation (Abebe et al., 2015; Frachon et al., 2017; Gardiner et al., 2018; Hendrick et al., 2016; Schemske, 1984; Sedlacek et al., 2016; Volis et al., 2016, 2015) over large distances, but none at this spatial scale nor using long-term datasets. We also showed that the differences in genotypic compositions between habitats correspond to heritable differences in phenotypes, and that some of these phenotypes consistently differ among habitats. This further supports the hypothesis of local adaptation at the micro-environment level. Interestingly, such phenotypic differences indicate which traits are important for local adaptation, and could be useful for crop improvement. Specifically, we observed differences in anthocyanin content and coleoptile color, which have been associated to resistance to wheat fungal diseases (Khlestkina et al., 2011), as well as drought resistance (Shoeva (Tereshchenko) et al., 2012), suggesting that anthocyanin

biosynthesis regulation in wheat can be useful for adaptation to both biotic and abiotic

stress. The possibility that local adaptation occurs at such a fine spatial scale as in our

study is intriguing, and invites experimental confirmation on further populations.

In light of the apparent adaptation to microhabitats, it is notable that only minor

changes were seen on the temporal scale when atmospheric CO_2 increased by 70

ppm and temperature by 1.5^{o}C. We might have expected migration of genotypes from

the southern to the northern slope, for example. This genetic temporal stability could

be interpreted as the population not having enough time to adapt due to the rapid pace

of climatic changes, in which case we would expect a reduction in fitness of the

population. However, plant density was similar in 1984 and in 2020, suggesting that,

so far, this has not occurred. Alternatively, it might be that wild emmer wheat has

evolved to be resilient to temperature fluctuations that occurred since its speciation

0.8 MYA ago (Levy and Feldman, 2022) and that recent climate changes have not yet

reached critical levels - while local adaptation is very local indeed.

Archaeological excavations of a hunter-gatherer's site on the shore of Lake Galilee

uncovered an abundant wild emmer wheat, indicating that it has been present in the

region of Ammiad for at least 23,000 years (Snir et al., 2015). Since then, the Levant

has undergone major climatic changes, including several cycles of glaciation, warming

and cooling (Zeder, 2011), pointing to the robustness of wild emmer wheat.

Nevertheless, temperatures and atmospheric CO_2 concentrations have never been as

high as they are now during that period, therefore, it is unclear to what extent additional

changes in environmental conditions could occur before negative consequences for

population fitness became apparent.

Our results are relevant to conservation strategies as well as to the assessment of

biodiversity resilience under climate change. First, our study shows that *in-situ*

conservation can be quite effective in preserving diversity, without the need to

frequently sample seeds and store them in seed banks (*ex-situ*). Second, our results

highlight that effective *in-situ* and *ex-situ* conservation strategies of wild plants

biodiversity, particularly for wild grasses and crop progenitors such as wild emmer

wheat, should consider habitat diversity as a potential reservoir of adaptive diversity.

In fact, our long-term study would not have been possible if the site had not been

declared a natural reserve for wild-wheat conservation by the Israel nature and parks

authority. Indeed, this population is fenced and undergoes moderate grazing.

Unrestricted grazing, or urbanization, as seen in other parts of the Levant can reduce

and sometimes eliminate wild grasses. The Ammiad reserve is relatively small (~10

hectares), but contains much diversity. Establishing natural reserves of crop

progenitors, partially protected from anthropogenic activities, does not necessarily

require large areas and is a feasible goal that is critical for biodiversity conservation.

In conclusion, we provide a unique example of strong temporary stability in a wild

grass population, most likely due to a combination of local adaptation and limited

dispersal. This stability is remarkable given that, during the period of this study,

temperatures and CO_2 levels increased significantly (Figures 1B and 1C). Yet only

rather minor changes occurred in the population genetic structure, reflecting the ability

of wild plants to deal with changing environments and pointing to the robustness of

wild emmer wheat. Such robustness further demonstrates the importance of

conservation of wild relatives of crops in nature, as gene pools for breeding (Zsögön

et al., 2021).

Materials and Methods

Sample collection and DNA extraction

Wild wheat spikes were collected in four different linear transects every 3 to 5 meters

from sampling points marked with pegs and GPS coordinates (Dataset S1). Plants

were assigned to a sampling point when collected in successive years in a radius of ~

1m from the peg.   A single seed was selected for each plant and DNA was extracted

from young seedlings. Tissue was collected and ground using mortar and pestle with

the addition of Merck sea sand in liquid nitrogen. DNA extraction was done using

Macherey-Nagel NucleoSpin® Plant II extraction kit, according to the manufacturer's

protocol, eluting in 50 ul EB total. Concentrations and DNA quality were estimated

using a NanoDrop® ND-1000 Spectrophotometer.

Genotyping by sequencing (GBS) and variant calling

Plant genotyping was done according to the Genotyping-By-Sequencing method

(Elshire et al., 2011; Poland et al., 2012). Samples were sequenced using Illumina

NextSeq 550 mid-output 150 base-pairs single-end kits. A reference sample, Zavitan

accession (Avni et al., 2017), was added to each run, undergoing the same plate GBS

library preparation and sequencing. Samples were divided by barcodes using a python

2.7 code. Mapping to the WEW_v2.0 genome assembly (Zhu et al., 2019) was done

using *bwa-mem* (Li, 2013). Conversion to binary alignment map (BAM) format, filtering

for uniquely mapped reads with minimum QUAL score of 30, sorting and indexing was

done using SAMtools (Li et al., 2009). Variant calling was done in parallel running on

all 845 (832 Ammiad samples and 13 Zavitan controls) samples using GATK

HaplotypeCaller 3.8 (Poplin et al., 2018). Filtering for quality (>30), read depth (>6),

genotype quality (>15), maximum missing sites (10%) and no-indels was done using

VCFtools (Danecek et al., 2011). Filtering for up to 10% heterozygous plants per

polymorphic position was done using TASSEL5 (Bradbury et al., 2007). Filtered VCF

was used for all further analyses.

Pairwise genetic identity

Identity was measured using R code which calculates the pairwise identity values

between all samples, divided by the numbers of comparisons. The pipeline also

includes the division of the samples to distinct genotype groups (DGGs), done by

setting a threshold of 98.1% identity, which is the lowest threshold where no sample

is assigned to two or more DGGs, and the construction of the dendrogram describing

the relatedness between the DGGs. Hierarchical clustering was done in R by UPGMA.

The 13 Zavitan reference samples added to each run were grouped together with

99.4% identity. Heatmap of samples from 1984 was done using *pheatmap*

(https://cran.r-project.org/web/packages/pheatmap/index.html)   with   default

hierarchical clustering.

Temperature and CO$_2$ concentration data

Data of daily minimum and maximum temperature from 1984 to 2020 were obtained

from Israel Meteorological Service at the Israel government portal

(https://ims.data.gov.il/he/ims/2). Data of weekly atmospheric CO2 concentrations

were obtained from The Scripps Institution of Oceanography UC San Diego (Keeling et al., 2001).

Rarefaction, principal component and identity by distance analyses

Rarefaction analysis was done using R, sampling backwards by the number of samples and different genotypes, per year, 100 times and adjusting a polynomial to each sampling.

Principal component analysis was done on the .vcf file of the entire Ammiad collection using *SNPRelate* R Package (Zheng et al., 2012). Identity by distance was correlated for each transect separately. Pairwise Euclidean distances was calculated using the *distHaversine* function of *Geosphere* R package (https://cran.rproject.org/web/packages/geosphere/index.html). Mantel test with 1000 permutations was used to calculate correlation between the genetic identity and the Euclidean distance matrices, using *mantel.test* in *Ape* R package (Paradis et al., 2004).

Outcrossing estimation and parental lines detection

To estimate outcrossing, we calculated F_{IS} using the R package hierfstat. Confidence intervals were calculated using 100 bootstraps and quantiles 2.5% and 97.5% using the function boot.ppfis. F_{IS} estimates were used to estimate outcrossing rates *f* using the relationship $F_{IS} = (1-f)/(1+f)$ (Wright, 1969). In order to detect the parental DGGs of heterozygous plants, we have used SNPMatch (Pisupati et al., 2017), running one plant against all DGGs.

Individual-based simulations

To investigate whether neutral demographic processes could account for the observed spatial clustering and temporal stability we performed individual-based simulations of populations of plants evolving via seed dispersal and outcrossing only. Briefly, we allow populations to evolve in a 2-dimensional continuous space, then draw a transect through the population and sample plants close to 30 evenly spaced sampling points in the last 36 generations to replicate the sampling procedure in the empirical dataset. An R package, *simmiad*, is provided to perform these simulations (DOI:10.5281/zenodo.4762083).

Populations begin with N plants of a single genotype distributed randomly through a homogeneous habitat with 1, 2, 3 or 5 plants/m^2. At each generation, N plants are sampled with replication to produce N seeds. Wild emmer wheat has a two-year seed bank, so from generation 3 onwards, seeds are drawn from the previous generation with probability *p*, and from two generations previous with probability 1-*p,* for values of *p* of 0.5, 0.7, 0.9 or 1 .Since we are only concerned with tracking unique genotypes, and information about seed dispersal is in any case lacking model outcrossing as analogous to mutation; seeds are randomly assigned as outcrossed with probabilities 0.25%, 0.5%, 1%, or 2%, in which case it is assigned a new unique genotype. Offspring disperse from the mother in a random direction at a distance drawn from an exponential distribution with a mean of 0.5, 0.75, 1 or 2m. To eliminate edge effects, the habitat is modelled as a torus, such that if dispersal takes a seed over the edge of the habitat it is reflected back to the opposite side. We allowed populations to evolve for 2000 generations to ensure populations had reached equilibrium. We then drew a transect of 30 sampling points 5m apart through the middle of the population, and

sampled the plant closest to each sampling point in each of the final 36 generations of

the simulations, replicating the sampling procedure in the field. For each combination

of plant densities, outcrossing rates, seed bank parameter p, and seed dispersal

distances we performed 100 replicate simulations.

Microhabitat sorting and analysis

We aimed to test the hypothesis that the population structure of Ammiad wild wheat

reflects that of microhabitats. We tested this using two different statistics and

permutation schemes. First, we tested whether plants in the same DGG tend to be

confined to a single microhabitat more than would be expected by chance using a

procedure analogous to a permutation test. We concatenated microhabitat and DGG

labels across transects, and calculated how often pairs of plants of the same DGG

were located in the same microhabitat. We then shuffled the order of microhabitats,

whilst keeping labels of the same microhabitat together, and recalculated how many

pairs of plants of the same DGG were located in the same microhabitat after shuffling.

We repeated this shuffling procedure 1000 times, and compared the observed number

of matching pairs within the same microhabitat to the distribution of shuffled datasets.

This procedure randomizes the relationship between the locations of DGGs and

microhabitat boundaries, but maintains the overall spatial structure of DGGs and

microhabitats.

Second, we tested whether population differentiation, measured by F_{st}, is higher when

measured between plants occurring in different microhabitats than in the same

microhabitat. To this goal we subdivided plants into demes with a maximum distance

between plants of 30m. We then computed the pairwise Fst between demes using the

R package *pegas* (Paradis, 2010). To correct for the effects of isolation by distance and spatial structure we stratified comparisons by distance, measured as the distance between the average locations of plants in the two demes.

We computed an analysis of molecular variance using the R package *pegas* (Paradis, 2010), using plants sampled across all years and restricting the analysis only to genotypes present in Ammiad. We included two factors, microhabitat and year of collection, with the latter nested within the former.

To account for spatial autocorrelation we first performed a partial Mantel test using the function *multi.mantel* from the R package *phytools* (Revell, 2012). Specifically, we considered as a response variable the pairwise genetic distances used to define DGGs. We tested for the effect of habitat, represented as a distance matrix with value "0" for plants found in the same microhabitat and "1" for plants found in two different microhabitats. In addition, we included in the test a matrix of time distances between the sampling of two plants (distance measured in years) and spatial distance, measured in meters and obtained from latitude and longitude using the function *earth.dist* from the R package *fossil* (https://palaeo-electronica.org/2011_1/238/index.html). We detected a significant effect of space and microhabitat (p-values <0.005 for both matrices) and a non-significant effect of time (p-value 0.863).

Note that Mantel and partial Mantel tests have severely been criticized for their inaccuracy in presence of high spatial autocorrelation (Crabot et al., 2019; Legendre et al., 2015). To account for these known biases, we used the approach proposed by (Crabot et al., 2019), which leverages a Moran spectral randomization procedure to generate random samples which maintain the same autocorrelation properties of the

original dataset. To this goal, we used the R packages *sdpep* to process the spatial

matrices and the *msr* function of the package *adespatial* (Guénard and Legendre,

2022). The distances above were transformed into Euclidean distances with function

cailliez from the R package *adegenet* (Jombart and Ahmed, 2011). Using this

approach, we tested the effect of habitat on both DGGs and pairwise genetic

distances, using 1000 randomizations. In both cases the effect of microhabitat was

significant (p-value<0.05).

Common garden experiment and analysis

We established a six-randomized-blocks common garden experiment in a net-house

in the Weizmann Institute of Science. In order to keep conditions as similar as possible

to the conditions in Ammiad site, seeds were sown in mid-November, in Terra Rosa

soil that was brought from Ammiad location and the experiment was subjected to

winter rainfall. We chose 95 samples representing 95 DGGs, not including DDGs of

heterozygous singletons. Phenotypes measured were: germination date, coleoptile

color (anthocyanin pigment/no pigment), erect growth habit (scored from 1 to 5), plant

base pigment during tillering (scored from 1 to 3), anthesis date, phenotypes during

anthesis - main tiller length (cm) from the soil to the basis of the spike, main tiller flag

leaf width (cm), main tiller number of spikelets, main tiller spike angle (scored from 1

to 3) and number of tillers. After seed collection we measured dry seed phenotypes of

10 seeds per plant: seed area (mm^2), seed length (mm), seed width (mm) and 10

seeds' weight (gr).

Phenotypic   analyses   were   done   in   R

(https://github.com/TalDM/Ammiad/tree/main/data_analysis/phenotypes.R).

Heritability estimates were performed using the R package *brms* (Bürkner, 2017),

implementing a linear mixed model with both genotypes and block as random factors.

For coleoptile color we implemented a logistic linear mixed model. Heritability was

calculated as var_g/(var_g+var_b+var_r), where var_g indicates the variance

explained by genotypes, var_b that explained by blocks, and var_r the residual

variance. Tests of the role of habitat were performed using the R package *lme4* (Bates

et al., 2015), testing an analogous model to that used to calculate heritability versus a

model including habitat as an additional fixed factor, using the R function anova.

Specifically, we tested models with the structure

"phenotype~habitats+(1|block)+(1|new_dgg)" vs. a set of null models not including the

factor habitats.

Stouffer's test was calculated using the R package *poolr* (Cinar and Viechtbauer,

2022). The linear discriminant analysis was computed and evaluated using the R

package *caret* (Kuhn, 2008). Linear discriminant analyses were computed using the

R function caret, after z-transformation of all the phenotypic data.

Data and code availability

546 Sequencing data generated in this study are available in the National Center for

547 Biotechnology Information (NCBI) Sequence Read Archive (SRA) under the accession

548 number PRJNA779576. All data processing and data analysis are publicly available

549 on GitHub (https://github.com/TalDM/Ammiad). An R package, *simmiad*, is provided

550 to perform the individual-based simulations (DOI:10.5281/zenodo.4762083).

Author Contributions

T.D-M., M.F., Y.A. and A.A.L, conceived and designed the study. T.D-M. carried out the molecular work, SNP calling, genetic and phenotypic analyses with input from T.J.E., F.M., M.N. and A.A.L. T.J.E. performed the simulations and microhabitat sorting analyses with input from T.D-M., M.N. and A.A.L. F.M. carried out the population differentiation analysis, the MSR-Mantel tests and the phenotypic analyses with input from T.D-M. and A.A.L. T.D-M., O.R. and A.A.L, carried out the common-garden experiment. T.D-M., H.S., J.M., N.A-R., A.R., M.F., Y.A, and A.A.L carried out the collection. T.D-M., T.J.E., F.M, M.N. and A.A.L wrote the manuscript. M.N. and A.A.L supervised the project.

Competing Interest Statement

The authors declare no competing interests.

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References

Abebe TD, Naz AA, Léon J. 2015. Landscape genomics reveal signatures of local adaptation in barley (Hordeum vulgare L.). *Front Plant Sci* **6**:813.

Aguillon SM, Fitzpatrick JW, Bowman R, Schoech SJ, Clark AG, Coop G, Chen N. 2017. Deconstructing isolation-by-distance: The genomic consequences of limited dispersal. *PLOS Genet* **13**:e1006911.

Anikster Y, Noy-Meir I. 1991. The wild-wheat field laboratory at Ammiad. *Isr J Bot* **40**:351–362.

Ashraf B, Hunter DC, Bérénos C, Ellis PA, Johnston SE, Pilkington JG, Pemberton JM, Slate J. 2021. Genomic prediction in the wild: A case study in Soay sheep. *Mol Ecol.*

Avni R, Nave M, Barad O, Baruch K, Twardziok SO, Gundlach H, Hale I, Mascher M, Spannagl M, Wiebe K, Jordan KW, Golan G, Deek J, Ben-Zvi B, Ben-Zvi G, Himmelbach A, MacLachlan RP, Sharpe AG, Fritz A, Ben-David R, Budak H, Fahima T, Korol A, Faris JD, Hernandez A, Mikel MA, Levy AA, Steffenson B, Maccaferri M, Tuberosa R, Cattivelli L, Faccioli P, Ceriotti A, Kashkush K, Pourkheirandish M, Komatsuda T, Eilam T, Sela H, Sharon A, Ohad N, Chamovitz DA, Mayer KFX, Stein N, Ronen G, Peleg Z, Pozniak CJ, Akhunov ED, Distelfeld A. 2017. Wild emmer genome architecture and diversity elucidate wheat evolution and domestication. *Science* **357**:93–97.

Bates D, Mächler M, Bolker BM, Walker SC. 2015. Fitting Linear Mixed-Effects Models Using lme4. *J Stat Softw* **67**:1–48. doi:10.18637/JSS.V067.I01

Bradbury PJ, Zhang Z, Kroon DE, Casstevens TM, Ramdoss Y, Buckler ES. 2007. TASSEL: software for association mapping of complex traits in diverse samples. *Bioinformatics* **23**:2633–2635.

Bürkner PC. 2017. brms: An R Package for Bayesian Multilevel Models Using Stan.

604 *J Stat Softw* **80**:1–28.

Chen N, Juric I, Cosgrove EJ, Bowman R, Fitzpatrick JW, Schoech SJ, Clark
605
606 **AG, Coop G.** 2019. Allele frequency dynamics in a pedigreed natural
607 population. *Proc Natl Acad Sci U S A* **116**:2158–2164.

Cinar O, Viechtbauer W. 2022. The poolr Package for Combining Independent and
608
609 Dependent p Values. *J Stat Softw* **101**:1–42.

Crabot J, Clappe S, Dray S, Datry T. 2019. Testing the Mantel statistic with a
610
611 spatially-constrained permutation procedure. *Methods Ecol Evol* **10**:532–540.

Danecek P, Auton A, Abecasis G, Albers CA, Banks E, DePristo MA, Handsaker
612
613 **RE, Lunter G, Marth GT, Sherry ST, McVean G, Durbin R, Group 1000**
614 **Genomes Project Analysis.** 2011. The variant call format and VCFtools.
615 *Bioinformatics* **27**:2156–2158.

Elshire RJ, Glaubitz JC, Sun Q, Poland JA, Kawamoto K, Buckler ES, Mitchell
616
617 **SE.** 2011. A Robust, Simple Genotyping-by-Sequencing (GBS) Approach for
618 High Diversity Species. *PLoS One* **6**:e19379.

Enjalbert J, David JL. 2000. Inferring recent outcrossing rates using multilocus
619
620 individual heterozygosity: application to evolving wheat populations. *Genetics*
621 **156**:1973–1982.

Felsenburg T, Levy AA, Galili G, Feldman M. 1991. Polymorphism of high-
622
623 molecular-weight glutenins in wild tetraploid wheat: spatial and temporal
624 variation in a native site. *Isr J Bot* **40**:451–479.

Frachon L, Libourel C, Villoutreix R, Carrère S, Glorieux C, Huard-Chauveau C,
625
626 **Navascués M, Gay L, Vitalis R, Baron E, Amsellem L, Bouchez O, Vidal M,**
627 **Le Corre V, Roby D, Bergelson J, Roux F.** 2017. Intermediate degrees of
628 synergistic pleiotropy drive adaptive evolution in ecological time. *Nat Ecol Evol*
629 *2017 110* **1**:1551–1561.

Franks SJ, Sim S, Weis AE. 2007. Rapid evolution of flowering time by an annual
630
631 plant in response to a climate fluctuation. *Proc Natl Acad Sci* **104**:1278–1282.

Fu Y-B, Peterson GW, Horbach C, Konkin DJ, Beiles A, Nevo E. 2019. Elevated
632
633 mutation and selection in wild emmer wheat in response to 28 years of global
634 warming. *Proc Natl Acad Sci* **116**:20002–20008.

Gardiner LJ, Joynson R, Omony J, Rusholme-Pilcher R, Olohan L, Lang D, Bai
635
636 **C, Hawkesford M, Salt D, Spannagl M, Mayer KFX, Kenny J, Bevan M, Hall**
637 **N, Hall A.** 2018. Hidden variation in polyploid wheat drives local adaptation.
638 *Genome Res* **28**:1319–1332.

Golenberg EM. 1988. Outcrossing rates and their relationship to phenology in
Triticum dicoccoides. *Theor Appl Genet* **75**:937–944.

Golenberg EM. 1987. Estimation of gene flow and genetic neighborhood size by
indirect methods in a selfing annual, Triticum dicoccoides. *Evolution (N Y)*
41:1326–1334.

Guénard G, Legendre P. 2022. Hierarchical Clustering with Contiguity Constraint in
R. *J Stat Softw* **103**:1–26.

Hendrick MF, Finseth FR, Mathiasson ME, Palmer KA, Broder EM, Breigenzer
P, Fishman L. 2016. The genetics of extreme microgeographic adaptation: an
integrated approach identifies a major gene underlying leaf trichome divergence
in Yellowstone Mimulus guttatus. *Mol Ecol* **25**:5647–5662.

Hoffmann AA, Sgró CM. 2011. Climate change and evolutionary adaptation. *Nat*
2011 4707335 **470**:479–485.

Horovitz A. 1998. The soil seed bank in wild emmerThe Proceedings of
International Symposium of In Situ Conservation of Plant Genetic Diversity. pp.
185–188.

Horovitz A, Feldman M. 1991. Evaluation of the wild-wheat study at Ammiad. *Isr J*
Bot **40**:501–508.

Jombart T, Ahmed I. 2011. adegenet 1.3-1: new tools for the analysis of genome-
wide SNP data. *Bioinformatics* **27**:3070–3071.

Keeling CD, Piper SC, Bacastow RB, Wahlen M, Whorf TP, Heimann M, Meijer
HA. 2001. Exchanges of Atmospheric CO2 and 13CO2 with the Terrestrial
Biosphere and Oceans from 1978 to 2000. I. Global Aspects. *I Glob Asp SIO*
Ref Ser **Scripps In**:88.

Khlestkina E, Antonova E, Pershina L, Soloviev A, Badaeva E, Börner A, Salina
E. 2011. Variability of Rc (red coleoptile) Alleles in Wheat and Wheat-alien
Genetic Stock Collections. *Cereal Res Commun 2011 394* **39**:465–474.

Khoury CK, Carver D, Greene SL, Williams KA, Achicanoy HA, Schori M, León
B, Wiersema JH, Frances A. 2021. Crop wild relatives of the United States
require urgent conservation action. *Proc Natl Acad Sci U S A* **117**:33351–33357.

Kolis KM, Berg CS, Nelson TC, Fishman L. 2022. Population genomic
consequences of life-history and mating system adaptation to a geothermal soil
mosaic in yellow monkeyflowers. *Evolution (N Y)* **76**:765–781.

Kuhn M. 2008. Building Predictive Models in R Using the caret Package. *J Stat*
Softw **28**:1–26.

Legendre P, Fortin MJ, Borcard D. 2015. Should the Mantel test be used in spatial
analysis? *Methods Ecol Evol* **6**:1239–1247. doi:10.1111/2041-210X.12425

Levy AA, Feldman M. 2022. Evolution and origin of bread wheat. *Plant Cell.*

Levy AA, Feldman M. 1988. Ecogeographical distribution of HMW glutenin alleles in
populations of the wild tetraploid wheat Triticum turgidum var. dicoccoides.
Theor Appl Genet **75**:651–658.

Li H. 2013. Aligning sequence reads, clone sequences and assembly contigs with
BWA-MEM.

Li H, Handsaker B, Wysoker A, Fennell T, Ruan J, Homer N, Marth G, Abecasis
G, Durbin R, Subgroup 1000 Genome Project Data Processing. 2009. The
Sequence Alignment/Map format and SAMtools. *Bioinformatics* **25**:2078–2079.

Nevo E. 2009. Evolution in action across life at "Evolution Canyons", Israel. *Trends*
Evol Biol **1**:e3–e3.

Noy-Meir I, Agami M, Anikster Y. 1991a. Changes in the population density of wild
emmer wheat (Triticum turgidum var. dicoccoides) in a mediterranean
grassland. *Isr J Bot* **40**:385–395.

Noy-Meir I, Agami M, Cohen E, Anikster Y. 1991b. Floristic and ecological
differentiation of habitats within a wild wheat population at Ammiad. *Isr J Bot*
40:363–384.

Ozbek O, Millet E, Anikster Y, Arslan O, Feldman M. 2007. Spatio-temporal
genetic variation in populations of wild emmer wheat, Triticum turgidum ssp.
dicoccoides, as revealed by AFLP analysis. *Theor Appl Genet 2007 1151*
115:19–26.

Paradis E. 2010. pegas: an R package for population genetics with an integrated-
modular approach. *Bioinformatics* **26**:419–420.

Paradis E, Claude J, Strimmer K. 2004. APE: Analyses of phylogenetics and
evolution in R language. *Bioinformatics* **20**:289–290.

Pecl GT, Araújo MB, Bell JD, Blanchard J, Bonebrake TC, Chen IC, Clark TD,
Colwell RK, Danielsen F, Evengård B, Falconi L, Ferrier S, Frusher S,
Garcia RA, Griffis RB, Hobday AJ, Janion-Scheepers C, Jarzyna MA,
Jennings S, Lenoir J, Linnetved HI, Martin VY, McCormack PC, McDonald
J, Mitchell NJ, Mustonen T, Pandolfi JM, Pettorelli N, Popova E, Robinson
SA, Scheffers BR, Shaw JD, Sorte CJB, Strugnell JM, Sunday JM, Tuanmu
MN, Vergés A, Villanueva C, Wernberg T, Wapstra E, Williams SE. 2017.
Biodiversity redistribution under climate change: Impacts on ecosystems and
human well-being. *Science* **355**.

Pisupati R, Reichardt I, Seren Ü, Korte P, Nizhynska V, Kerdaffrec E, Uzunova
K, Rabanal FA, Filiault DL, Nordborg M. 2017. Verification of Arabidopsis
stock collections using SNPmatch, a tool for genotyping high-plexed samples.
Sci Data **4**:170184.

Poland JA, Brown PJ, Sorrells ME, Jannink J-L. 2012. Development of High-
Density Genetic Maps for Barley and Wheat Using a Novel Two-Enzyme
Genotyping-by-Sequencing Approach. *PLoS One* **7**:e32253.

Poplin R, Ruano-Rubio V, DePristo MA, Fennell TJ, Carneiro MO, Auwera GA
Van der, Kling DE, Gauthier LD, Levy-Moonshine A, Roazen D, Shakir K,
Thibault J, Chandran S, Whelan C, Lek M, Gabriel S, Daly MJ, Neale B,
MacArthur DG, Banks E. 2018. Scaling accurate genetic variant discovery to
tens of thousands of samples. *bioRxiv* 201178.

Revell LJ. 2012. phytools: an R package for phylogenetic comparative biology (and
other things). *Methods Ecol Evol* **3**:217–223.

Schemske DW. 1984. Population structure and local selection in Impatiens pallida
(balsaminaceae), a selfing annual. *Evolution (N Y)* **38**:817–832.

Sedlacek J, Cortés AJ, Wheeler J, Bossdorf O, Hoch G, Klápště J, Lexer C,
Rixen C, Wipf S, Karrenberg S, van Kleunen M. 2016. Evolutionary potential
in the Alpine: trait heritabilities and performance variation of the dwarf willow
Salix herbacea from different elevations and microhabitats. *Ecol Evol* **6**:3940–
3952.

Shoeva (Tereshchenko) O, Khlestkina E, Gordeeva E, Arbuzova VS, Salina E.
2012. Relationship between anthocyanin biosynthesis and abiotic stress in
wheat. *Proc15th EWAC Conf* **2011**:72–75.

Siepielski AM, Dibattista JD, Carlson SM. 2009. It's about time: The temporal
dynamics of phenotypic selection in the wild. *Ecol Lett* **12**:1261–1276.

Snir A, Nadel D, Groman-Yaroslavski I, Melamed Y, Sternberg M, Bar-Yosef O,
Weiss E. 2015. The Origin of Cultivation and Proto-Weeds, Long Before
Neolithic Farming. *PLoS One* **10**:e0131422.

Stoffel MA, Johnston SE, Pilkington JG, Pemberton JM. 2021. Genetic
architecture and lifetime dynamics of inbreeding depression in a wild mammal.
Nat Commun 2021 121 **12**:1–10.

Todesco M, Owens GL, Bercovich N, Légaré JS, Soudi S, Burge DO, Huang K,
Ostevik KL, Drummond EBM, Imerovski I, Lande K, Pascual-Robles MA,
Nanavati M, Jahani M, Cheung W, Staton SE, Muños S, Nielsen R, Donovan
LA, Burke JM, Yeaman S, Rieseberg LH. 2020. Massive haplotypes underlie

ecotypic differentiation in sunflowers. *Nat 2020 5847822* **584**:602–607.

Urban MC. 2015. Accelerating extinction risk from climate change. *Science*
348:571–573.

Volis S, Anikster Y, Olsvig-Whittaker L, Mendlinger S. 2004. The influence of
space in genetic-environmental relationships when environmental heterogeneity
and seed dispersal occur at similar scale. *Am Nat* **163**:312–327.

Volis S, Ormanbekova D, Shulgina I. 2016. Fine-scale spatial genetic structure in
predominantly selfing plants with limited seed dispersal: A rule or exception?
Plant Divers **38**:59–64.

Volis S, Ormanbekova D, Yermekbayev K, Song M, Shulgina I. 2015. Multi-
Approaches Analysis Reveals Local Adaptation in the Emmer Wheat (Triticum
dicoccoides) at Macro- but not Micro-Geographical Scale. *PLoS One*
10:e0121153.

Wintle BA, Kujala H, Whitehead A, Cameron A, Veloz S, Kukkala A, Moilanen A,
Gordon A, Lentini PE, Cadenhead NCR, Bekessy SA. 2019. Global synthesis
of conservation studies reveals the importance of small habitat patches for
biodiversity. *Proc Natl Acad Sci U S A* **116**:909–914.

Wright S. 1969. Evolution and the Genetics of Populations. Vol. 2: The Theory of
Gene Frequencies. Chicago, University of Chicago Press.

Zeder MA. 2011. The origins of agriculture in the Near East. *Curr Anthropol* **52**.

Zheng X, Levine D, Shen J, Gogarten SM, Laurie C, Weir BS. 2012. A high-
performance computing toolset for relatedness and principal component
analysis of SNP data. *Bioinformatics* **28**:3326.

Zhu T, Wang L, Rodriguez JC, Deal KR, Avni R, Distelfeld A, McGuire PE,
Dvorak J, Luo M-C. 2019. Improved Genome Sequence of Wild Emmer Wheat
Zavitan with the Aid of Optical Maps. *G3 (Bethesda)* **9**:619–624.

Zsögön A, Peres LEP, Xiao Y, Yan J, Fernie AR. 2021. Enhancing crop diversity
for food security in the face of climate uncertainty. *Plant J* **109**:402–414.

Supporting Information

Figure S1. Rarefaction curves. Adjusted 2nd degree polynomials of 100x backwards sample of unique genotype number and number of samples collected per year. Dashed horizontal line represents the observed number of unique genotypes per year, dashed vertical lines represent number of samples per year. Light blue line and red dot represent range and mean of inflection points over replicate subsamples respectively.

Figure S2. **Outcrossing in Ammiad.** (A) Outcrossing over time in Ammiad. The red continuous line represents the estimated average decrease of 0.06% every year, with confidence intervals shown as dashed lines. The highest rate of outcrossing is estimated for the middle north-facing slope, consistently with the identification of two highly heterozygous plants in this habitat from 1984 and 1988 (A17 and A20, respectively). Individual yearly estimates are shown as black dots. (B) Parental DGGs of 1984A17 (left panel) and 1988A47 (right panel) detected using SNPMatch. F1 plants are marked as a black dot.

Distance (m)

798 **Figure S3. Correlation and significance of isolation by distance in Ammiad collection.**

799 Pairwise genetic identity by Euclidean distance (meters). Subplots show years of collection

800 (rows) and transects (columns). The title of each plot indicates its number of samples and p-

801 value from a Mantel test with 1000 permutations.

Figure S4. Overlap in spatial clustering between observed data and populations simulated under neutral demography and the probabilities that observed temporal stability could be generated by neutral demography. (A) The proportion of replicate simulations where pairs of sampling points at different distances were more likely to be occupied by plants of the same DGG than in any observed transect. **(B)** The proportion of replicate simulations where a single sampling point sampled in different harvest years was more likely to be occupied by plants of the same DGG than in any observed transect. Panels indicate increasing mean seed dispersal distances (left to right) and increasing outcrossing rates (top to bottom). 'Dormancy' indicates the proportion of seeds drawn from the seed bank in each generation, and 'Density' the number of plants per square meter.

Figure S5. Distribution of phenotypes measured in common-garden experiment.

Figure S6. Trait heritability and phenotype-habitat linear discriminant analysis. (A) Posterior probability distribution of the heritability estimates across phenotypes. (B)

829   Correlation among phenotypes in the data used to estimate the role of habitat. Warmer colors

830   indicate stronger positive correlation. (C) Projection of the first two dimensions of the linear

831   discriminant analysis of the phenotypes of the genotypes sampled from the wild wheat

832   Ammiad population. Different colors indicate genotypes distinctive of different habitats as

833   defined in the text.

Figure S7. Phenotypes most relevant to distinguish different habitats according to the linear discriminant analysis, sorted top to bottom according to their overall importance across habitats.

 Supplementary Tables

865 **Table S1. Bootstrap-based ranges of F_{IS} and estimated outcrossing rates per habitat.**

	F_{IS}	F_{IS}	outcrossing	outcrossing
Habitat	2.5%	97.5%	2.5%	97.5%
Upper North facing slope	0.9074	0.9246	0.0392	0.0485
Middle north facing slope	0.8385	0.8654	0.0722	0.0878
Narrow valley	0.9548	0.9627	0.0190	0.0231
South facing slope	0.9274	0.9348	0.0337	0.0377
Plateau	0.9565	0.9640	0.0183	0.0222
East facing slope	0.9294	0.9405	0.0307	0.0366
Valley	0.9535	0.9614	0.0197	0.0238
Karst	0.9082	0.9317	0.0354	0.0481

Table S2. Analysis of molecular variance of genotypes sampled in Ammiad across all
years, subdivided by habitat, using only polymorphic sites within Ammiad. Sum of
squared deviations (SSD) and mean square deviations (MSD), variance components (σ^2) and
corresponding p-values were computed using the R package *pegas*. The Φ statistic for habitat
was used as an estimate of the proportion of molecular variance explained by the habitat
structure.

Component	SSD	MSD	df	σ^2	p-value	Φ
Habitat	0.9946	0.1421	7	0.0014	$<10^{-12}$	0.2833
Error	2.9429	0.0036	824	0.0036	/	/
Total	3.9374	0.0047	831	/	/	/

876 **Table S3. Heritability of the phenotypes, measured as the proportion of variance among**
877 **genotypes and blocks (growing plots) explained by genotype, using a Bayesian model**
878 **implemented in the R package *brms* (Bürkner, 2017).**

Phenotype	heritability	Standard Deviation
Seed length	0.78019	0.034308
Erect growth habit	0.663759	0.041745
Main tiller flag leaf width	0.657141	0.041966
Main tiller length	0.612437	0.073655

Plant base pigment during tillering	0.5053	0.05331
Coleoptile color	0.505163	0.090388
Anthesis time	0.500273	0.077041
Seed weight	0.485189	0.055428
Seed area	0.444486	0.052507
Main tiller number of spikelets	0.412969	0.097037
Seed width	0.384354	0.052098
Main tiller spike angle	0.358738	0.049909
Number of tillers	0.231275	0.062326
Germination date	0.029275	0.027938

881 **Table S4**. **Association between habitat and phenotypes.** Generalized linear mixed models

882 comprising the random effect habitats were tested against a set of null models for each

883 phenotype, not including habitat.

Phenotype	p.value	Bonferroni corrected p.value
Coleoptile color	0.001666	0.019988
Erect growth habit	0.092703	1
Plant base pigment during tillering	0.002859	0.034312
Main tiller length	0.249044	1
Main tiller flag leaf width	0.03783	0.453954
Main tiller number of spikelets	0.953601	1
Number of tillers	0.557119	1
Main tiller spike angle	0.34719	1
Seed weight	0.486148	1
Seed length	0.893096	1
Seed width	0.185839	1
Anthesis time	0.056984	0.683811

884 **Table S5. Linear discriminant analysis dimension values.**

	LD1	LD2	LD3	LD4	LD5
Coleoptile color	0.627	1.483	1.001	0.230	0.270
Erect growth habit	-0.330	0.115	0.093	-0.335	0.606

Plant base pigment during tillering	-0.651	0.036	0.508	-0.145	-0.114
Main tiller length	-0.288	0.128	-0.340	-0.332	-0.685
Main tiller flag leaf width	-0.384	0.373	-0.610	0.129	0.130
Main tiller number of spikelets	-0.055	-0.057	-0.175	0.315	0.157
Number of tillers	-0.233	-0.361	0.095	-0.320	0.010
Main tiller spike angle	0.304	-0.165	-0.094	-0.275	-0.462
Seed weight	0.015	0.219	-0.068	0.528	-0.965
Seed length	-0.533	-0.054	-0.126	0.017	0.413
Seed width	0.235	-0.549	-0.318	-0.819	0.518
Anthesis time	-0.249	-0.364	-0.620	0.339	-0.197

Movie S1 (separate file). Drone video of the different transects and microhabitats in Ammiad wild wheat reserve.

Dataset S1 (separate file). Dataset of Ammiad samples by genotype, coordinates and spike density measured at transect A in 2020.