Stability of the genetic structure of a wild wheat population across

microhabitats over 36 years

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Abstract

Long-term genetic studies of wild populations are very scarce but are essential for connecting ecological and population genetics models, and are also necessary for understanding how climate change can affect biodiversity. We present a study of the wild ancestor of modern wheat, carried out over a 36-year period during which atmospheric CO_2 increased by 70 ppm and temperature by $1.5^o C$. We genotyped 832 individuals that had been collected along the same transects during the course of the experiment. To our surprise, very little changed. Genotypes were clustered into ecological microhabitats over scales of tens of metres, and this clustering was stable over the 36 generations of the study. Simulations indicate that neutral processes alone are unlikely to fully explain this stability, suggesting a role for fine-scale local adaptation. Our results provide a rare insight into the dynamics of a wild population at the scale of individuals.

Introduction

Climate change is a looming threat to plant biodiversity and food security (Hoffmann and Sgró, 2011; Pecl et al., 2017; Urban, 2015). This increases the importance of wild progenitors of crops because their genetic diversity is an immense reservoir of biotic and abiotic stress-resistance genes (Khoury et al., 2021). Long-term studies of progenitor species in their natural ecological niches can provide a better understanding of the likely impacts of climate change, and inform strategies for conservation of such resources *in-situ* as well as in gene banks (Wintle et al., 2019).

We examined the genetic diversity of a population of wild emmer wheat (*Triticum turgidum ssp. dicoccoides*) in Ammiad, Israel, which has been monitored for 36 years (Fig. 1A). During this period, local ambient temperatures increased by ~1.5 degrees Celsius (Fig. S1) and atmospheric CO_2 concentration increased by more than 20% (Fig. S2) (Keeling et al., 2001). This population was previously shown to be highly diverse in comparison with other populations, suggesting that it is located at the centre of distribution of the species (Felsenburg et al., 1991; Levy and Feldman, 1988). Wild emmer wheat is a self-pollinating, allotetraploid annual cereal (genome BBAA) that was domesticated about 10,000 years ago, during the Neolithic period (Avni et al., 2017; Nave et al., 2019; Salamini et al., 2002). It is the direct progenitor of durum wheat (*T. turgidum ssp. durum*) and the maternal donor of the A and B sub-genomes of bread wheat (*T. aestivum*, genome BBAADD). The population has been sampled at nine time points between 1984 and 2020. At each time point, seeds were sampled from plants at 100 marked locations along four transects traversing seven ecologically distinct microhabitats (Anikster and Noy-Meir, 1991) (Fig. 1A, Table S1, Movie S1). We genotyped 832 individuals using a genotyping-by-sequencing (GBS) protocol (Poland et al., 2012),

resulting in 5196 single-nucleotide-polymorphism (SNP) markers spread throughout the

genome (Fig. S3).

Earlier spatio-temporal studies of wild plants involved shorter periods, as well as less

comprehensive and less precise sampling (Franks et al., 2007; Fu et al., 2019; Horovitz and

Feldman, 1991; Ozbek et al., 2007; Volis et al., 2004). Our dataset thus provides a unique

resource for assessing genetic changes in a natural population in its natural habitat over many

generations, and connects population genetic data with ecological sampling.

Results

As expected for a highly selfing species (The 1001 Genomes Consortium, 2016), the

population is composed of genetically distinct groups of nearly identical, highly homozygous

individuals clustered in space. Pairwise genetic identity follows a multimodal distribution that

is characterised by pairs of plants that are either identical at almost all SNPs, or identical at

70% to 90% of SNPs (Fig. 1B). We unambiguously assigned all plants to distinct genotype

groups (DGGs) of individuals that were identical at >98% of markers. This resulted in 118

DGGs over all nine time points, including 61 singletons (DGGs with a single member). The

10 largest DGGs encompass 65% of the samples, while the numbers of DGGs per year varied

between 32 (2014) to 49 (1988; Table S2). Rarefaction curves indicate that the sampling

scheme was sufficient to detect almost all of the DGGs present close to the transects (Fig. S4).

These results provide a detailed picture of the genetic structure of the population through time

and space.

Fig. 1. Spatial structure of the Ammiad population. (A) Location of Ammiad near the lake

of Galilee (star, top left and geographic coordinates, bottom right). Sampling points of

transects A, B, C, D collected in the years 1984, 1988, 1992, 1996, 2002, 2014, 2016, 2018

and 2020. Sampling points are shown on-site for transects A, B, D, but transect C, shown in

the dashed rectangle, is located ~500 meters north-west of A, B, D. Sampling points colours

correspond to the different microhabitats labelled in the legend. Image of map on the right is

from gov.map.co.il, map on the left from Google Maps. (B) Distribution of pairwise identity

values between the plants in the collection.

The spatio-temporal distribution of DGGs shows that plants of the same DGG tend to cluster

together in both space and time (Fig. 2 and Fig. S5). This clustering was also found in a

principal component analysis (Fig. S6). Within years, the probability of belonging to the same

DGG decays rapidly with distance (Fig. 3A), however, there is little evidence for isolation by

distance beyond this (Fig. S7, S8). Overall, we find a striking concordance between DGGs and

microhabitats, which were defined based on topographic and floristic data without knowledge

of the genetic structure of the population (Noy-Meir et al., 1991b) (Fig. 2). For example, in

transect A there is a sharp transition between DGGs of the upper and middle north-facing slope

(Fig. 2A). Another example is the genotype at collection point A59 in a unique location under

the shade of a *Ziziphus spina-christi* tree (Fig. 2A). In contrast, in transect D, where there is

negligible microhabitat differentiation, plants of different DGGs were much more mixed (Fig.

2D). The spatial pattern was highly stable over the course of the experiment. In contrast to the

rapid spatial decay, the probability of belonging to the same DGG decays only slowly over

time (Fig. 3). For example, 33% of plants sampled in 1984 belonged to the same DGG as plants

sampled at the same position in 2020. In conclusion, our results show a strong spatial structure,

associated with microhabitat, which has remained stable for at least 36 years despite changes

in climatic conditions.

Fig. 2. Spatio-temporal distribution of genotypes. (A-D) Distribution of samples coloured

by distinct genotype group (DGG) in transects A, B, C and D. Y-axes indicate year of

collection (left) and altitude for 1984 (right). Singleton samples are marked in small black

dots, while doubleton groups are marked with "2" in the middle with the colour of their

closest DGG. Doubleton groups that are closest to the same DGG were marked with coloured

asterisks. (E) Dendrogram of DGGs with ≥3 samples, hierarchically clustered by UPGMA.

Fig. 3. Spatial clustering and temporal stability at Ammiad. Points show the probability

of observing genotypes of the same DGG at (A) pairs of sampling points in the same year at

increasing distances along each transect and (B) the same sampling point between pairs of

years. Grey regions indicate 95% confidence intervals over populations simulated under

conditions closest to empirical estimates (3 plants/m^2, mean seed dispersal of 1m, 0.5%

outcrossing per generation, 30% seeds germinating in the second year after dispersal). In A,

points show the average for sampling points in bins of 5m.

The observed patterns cannot be explained by simple chance. Across transects, pairs of the

same DGG were significantly more likely to be found in the same microhabitat in the observed

dataset than in 1000 permuted datasets where the order of microhabitats was randomly shuffled

(Fig. S9), or even when we control for distance and spatial autocorrelation, which could be

caused by limited dispersal, as shown via partial Mantel tests and Moran-spectral-

randomization Mantel statistics (MSR-Mantel) (Fig. S10). Similarly, population differentiation

measured by F_{st} was consistently higher between microhabitats than within, even when taking

distance into account (Fig. S11), and genetic distances are strongly correlated with

microhabitat. Furthermore, an analysis of molecular variance (AMOVA) showed that 28% of

the genetic variation between samples is explained by differences between microhabitats

(Table S3). In order to further test the hypothesis that the correspondence between the

population structure of Ammiad and microhabitats is shaped by processes other than limited

dispersal, we implemented an explicit demographic model that also incorporates self-

fertilization and limited seed dispersal. The question is whether these neutral demographic

processes - on their own - could generate the strong spatial clustering and temporal stability

observed at Ammiad.

To investigate this, we performed individual-based simulations of populations of plants in a

homogeneous environment, evolving by seed dispersal and outcrossing only. We ran

simulations using different values of outcrossing rate, plant density, seed dormancy and mean

seed dispersal distance that approximate a range of biologically realistic values. Empirical

values of outcrossing rates in wild wheat have been estimated to be approximately 0.5%

(Golenberg, 1988) with up to 30% of germinating seedlings coming from the seed bank

(Horovitz, 1998). Average plant density in the Ammiad population was measured in 1984 as

4 plants/m^2 (Noy-Meir et al., 1991a) and again in 2020 for transect A as 3.8 plants/m^2 (Fig.

S12, Table S4). We lack direct estimates of seed dispersal, but average gene flow, a composite

of seed and pollen dispersal, was estimated to be 1.25 m per year (Golenberg, 1987). Seed

dispersal occurs through a combination of gravity, ants, birds, grazing mammals and rodents,

enabling both short and long-distance dispersal. For example, the appearance of a DDG

prominent in transect C (Fig. 2C, marked by yellow dot), in transect D in 2020 (Fig. 2D,

locations D198, D217), more than 700 meters away, demonstrates that long distance movement

occasionally occurs.

In simulations with the parameters we consider most likely (1 m mean seed dispersal, 0.5%

outcrossing, density of 3 plants/m^2 and 30% seed dormancy) we find that neutral demographic

processes alone could not account for the observed spatial clustering and temporal stability

(Fig. 3), in line with other studies (Nevo, 2009). However, the simulation could produce results

consistent with the data if we assumed extremely low plant densities ($\leq$1 plant/m^2) or rates of

gene flow (outcrossing <0.5% when seed dispersal $\leq$ 1m) (Fig. S13-S16). Without better

estimates of these parameters or direct fitness estimates from reciprocal transplant experiments,

it is not possible to definitely reject a purely neutral explanation for the observed patterns. That

said, the correspondence between DGGs and microhabitats is strong (Fig. 2 and Fig. S9-S11).

Thus, the simulations indicate that it would be difficult to generate the observed patterns of

stability over 36 years through neutral demographic processes alone.

Discussion

Studies of natural populations rarely have fine-scale geographic resolution, and almost never

have temporal resolution beyond the duration of a PhD (Siepielski et al., 2009). Our study,

spanning over 36 years of dense sampling, thus provides a unique dataset that could improve

our understanding of gene flow and local adaptation - often modelled quite abstractly in

population genetics - on an ecological scale.

Our analysis revealed that the collection of Ammiad population varies more in space than in

time. We found unexpectedly strong long-term genetic stability, demonstrating that gene flow

can be highly limited in a wild cereal despite its relatively wide distribution throughout the

fertile crescent. Remarkably, the stable genetic clusters observed in Ammiad matched

ecological microhabitats previously identified at the site, suggesting that natural selection, in

addition to limited dispersal, contributed to shaping population structure. Previous studies of

wild populations have demonstrated local adaptation (Abebe et al., 2015; Frachon et al., 2017;

Gardiner et al., 2018; Schemske, 1984; Sedlacek et al., 2016; Volis et al., 2016, 2015) over

large distances, but none at this spatial scale nor using long-term datasets. The possibility that

local adaptation occurs at such a fine spatial scale as in our study is particularly intriguing, and

invites experimental confirmation.

These results are relevant to conservation strategies as well as to the assessment of biodiversity

resilience under climate change. First, our study shows that *in-situ* conservation can be quite

effective in preserving diversity, without the need to frequently sample seeds and store them

in seed banks (*ex-situ*). Second, our results highlight that effective *in-situ* and *ex-*

situ conservation strategies of wild plants biodiversity, particularly for wild grasses and crop

progenitors such as wild emmer wheat, should consider habitat diversity as a potential reservoir

of adaptive diversity. In fact, our long-term study would not have been possible if the site had

not been declared a natural reserve for wild-wheat conservation by the Israel nature and parks

authority. Indeed, this population is fenced and undergoes moderate grazing. Unrestricted

grazing, or urbanization, as seen in other parts of the Levant can reduce and sometimes

eliminate wild grasses. The Ammiad reserve is relatively small (~10 hectares), but contains

much diversity. Establishing natural reserves of crop progenitors, partially protected from

anthropogenic activities, does not necessarily require large areas and is a feasible goal that is

critical for biodiversity conservation.

In conclusion, we provide evidence for strong genetic stability in a wild grass, most likely due

to a combination of local adaptation and limited dispersal. This stability is remarkable given

that, during the period of this study, temperatures and CO_2 levels increased significantly (Fig.

S1, S2). Yet only rather minor changes occurred in the population genetic structure, reflecting

the ability of wild plants to deal with changing environments and pointing to the robustness of

wild emmer wheat. Such robustness further demonstrates the importance of conservation of

wild relatives of crops in nature, as gene pools for breeding (Zsögön et al., 2021).

Archaeological excavations of a hunter-gatherers' site on the shore of Lake Galilee uncovered

an abundant wild emmer wheat, indicating that it has been present in the region of Ammiad for

at least 23,000 years (Snir et al., 2015). Since then, the Levant has undergone major climatic

changes, including several cycles of glaciation, warming and cooling (Zeder, 2011), pointing

to the robustness of wild emmer wheat. Nevertheless, temperatures and atmospheric CO_2

concentrations have never been as high as they are now during that period, therefore, it is

unclear to what extent additional changes in environmental conditions could occur before

negative consequences for population fitness became apparent.

Materials and Methods

Sample collection and DNA extraction

Wild wheat spikes were collected in four different linear transects every 3 to 5 meters from

sampling points marked with pegs and GPS coordinates (Table S1). Plants were assigned to a

sampling point when collected in successive years in a radius of ~ 1m from the peg.   A single

seed was selected for each plant and DNA was extracted from young seedlings. Tissue was

collected and ground using mortar and pestle with the addition of Merck sea sand in liquid

nitrogen. DNA extraction was done using Macherey-Nagel NucleoSpin® Plant II extraction

kit, according to the manufacturer's protocol, eluting in 50 ul EB total. Concentrations and

DNA quality were estimated using a NanoDrop® ND-1000 Spectrophotometer.

Genotyping by sequencing (GBS) and variant calling

Plant genotyping was done according to the Genotyping-By-Sequencing method (Elshire et al.,

2011; Poland et al., 2012). Samples were sequenced using Illumina NextSeq 550 mid-output

150 base-pairs single-end kits. A reference sample, Zavitan accession (Avni et al., 2017), was

added to each run, undergoing the same plate GBS library preparation and sequencing. Samples

were divided by barcodes using a python 2.7 code. Mapping to the WEW_v2.0 genome

assembly (Zhu et al., 2019) was done using *bwa-mem* (Li, 2013). Conversion to binary

alignment map (BAM) format, filtering for uniquely mapped reads with minimum QUAL score

of 30, sorting and indexing was done using SAMtools (Li et al., 2009). Variant calling was

done in parallel running on all 845 (832 Ammiad samples and 13 Zavitan controls) samples

using GATK HaplotypeCaller 3.8(Poplin et al., 2018). Filtering for quality (>30), read depth

(>6), genotype quality (>15), maximum missing sites (10%) and no-indels was done using

VCFtools (Danecek et al., 2011). Filtering for up to 10% heterozygous plants per polymorphic

position was done using TASSEL5 (Bradbury et al., 2007). Filtered VCF was used for all

further analyses.

Pairwise genetic identity

Identity was measured using R code which calculates the pairwise identity values between all

samples, divided by the numbers of comparisons. The pipeline also includes the division of the

samples to distinct genotype groups (DGGs), done by setting a threshold of 98.1% identity,

which is the lowest threshold where no sample is assigned to two or more DGGs. and the

construction of the dendrogram describing the relatedness between the DGGs. Hierarchical

clustering was done in R by UPGMA. The 13 Zavitan reference samples added to each run

were grouped together with 99.4% identity. Heatmap of samples from 1984 was done using

pheatmap (https://cran.r-project.org/web/packages/pheatmap/index.html) with default

hierarchical clustering.

Temperature and CO2 concentration data

Data of daily minimum and maximum temperature from 1984 to 2020 were obtained from

Israel Meteorological Service at the Israel government portal (https://ims.data.gov.il/he/ims/2).

Data of weekly atmospheric CO_2 concentrations were obtained from The Scripps Institution

of Oceanography UC San Diego (Keeling et al., 2001).

Rarefaction, principal component and identity by distance analyses

Rarefaction analysis was done using R, sampling backwards by the number of samples and

different genotypes, per year, 100 times and adjusting a polynomial to each sampling.

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Principal component analysis was done on the .vcf file of the entire Ammiad collection using

SNPRelate R Package (Zheng et al., 2012). Identity by distance was correlated for each transect

separately. Pairwise Euclidean distances was calculated using the *distHaversine* function of

Geosphere R package (https://cran.rproject.org/web/packages/geosphere/index.html). Mantel

test with 1000 permutations was used to calculate correlation between the genetic identity and

the Euclidean distance matrices, using *mantel.test* in *Ape* R package (Paradis et al., 2004).

Individual-based simulations

To investigate whether neutral demographic processes could account for the observed spatial

clustering and temporal stability we performed individual-based simulations of populations of

plants evolving via seed dispersal and outcrossing only. Briefly, we allow populations to evolve

in a 2-dimensional continuous space, then draw a transect through the population and sample

plants close to 30 evenly spaced sampling points in the last 36 generations to replicate the

sampling procedure in the empirical dataset. An R package, *simmiad*, is provided to perform

these simulations (DOI:10.5281/zenodo.4762083).

Populations begin with N plants of a single genotype distributed randomly through a

homogeneous habitat with 1, 2, 3 or 5 plants/m^2. At each generation, N plants are sampled with

replication to produce N seeds. Wild emmer wheat has a two-year seed bank, so from

generation 3 onwards, seeds are drawn from the previous generation with probability p, and

from two generations previous with probability 1-p, for values of p of 0.5, 0.7, 0.9 or 1 .Since

we are only concerned with tracking unique genotypes, and information about seed dispersal

is in any case lacking model outcrossing as analogous to mutation; seeds are randomly assigned

as outcrossed with probabilities 0.25%, 0.5%, 1%, or 2%, in which case it is assigned a new

unique genotype. Offspring disperse from the mother in a random direction at a distance drawn

from an exponential distribution with a mean of 0.5, 0.75, 1 or 2m. To eliminate edge effects,

the habitat is modelled as a torus, such that if dispersal takes a seed over the edge of the habitat

it is reflected back to the opposite side. We allowed populations to evolve for 2000 generations

to ensure populations had reached equilibrium. We then drew a transect of 30 sampling points

5m apart through the middle of the population, and sampled the plant closest to each sampling

point in each of the final 36 generations of the simulations, replicating the sampling procedure

in the field. For each combination of plant densities, outcrossing rates, seed bank parameter *p*,

and seed dispersal distances we performed 100 replicate simulations.

Microhabitat sorting and analysis

We aimed to test the hypothesis that the population structure of Ammiad wild wheat reflects

that of microhabitats. We tested this using two different statistics and permutation schemes.

First, we tested whether plants in the same DGG tend to be confined to a single microhabitat

more than would be expected by chance using a procedure analogous to a permutation test. We

concatenated microhabitat and DGG labels across transects, and calculated how often pairs of

plants of the same DGG were located in the same microhabitat. We then shuffled the order of

microhabitats, whilst keeping labels of the same microhabitat together, and recalculated how

many pairs of plants of the same DGG were located in the same microhabitat after shuffling.

We repeated this shuffling procedure 1000 times, and compared the observed number of

matching pairs within the same microhabitat to the distribution of shuffled datasets. This

procedure randomises the relationship between the locations of DGGs and microhabitat

boundaries, but maintains the overall spatial structure of DGGs and microhabitats.

Second, we tested whether population differentiation, measured by F_{st}, is higher when

measured between plants occurring in different microhabitats than in the same microhabitat.

To this goal we subdivided plants into demes with a maximum distance between plants of 30m.

We then computed the pairwise Fst between demes using the R package *pegas* (Paradis, 2010).

To correct for the effects of isolation by distance and spatial structure we stratified comparisons

by distance, measured as the distance between the average locations of plants in the two demes.

We computed an analysis of molecular variance using the R package *pegas* (Paradis, 2010),

using plants sampled across all years and restricting the analysis only to genotypes present in

Ammiad. We included two factors, microhabitat and year of collection, with the latter nested

within the former.

To account for spatial autocorrelation we first performed a partial Mantel test using the function

multi.mantel from the R package *phytools*. Specifically, we considered as a response variable

the pairwise genetic distances used to define DGGs. We tested for the effect of habitat,

represented as a distance matrix with value "0" for plants found in the same microhabitat and

"1" for plants found in two different microhabitats. In addition, we included in the test a matrix

of time distances between the sampling of two plants (distance measured in years) and spatial

distance, measured in metres and obtained from latitude and longitude using the function

earth.dist from the R package fossil. We detected a significant effect of space and microhabitat

(p-values <0.005 for both matrices) and a non-significant effect of time (p-value 0.863).

Note that Mantel and partial Mantel tests have severely been criticised for their inaccuracy in

presence of high spatial autocorrelation (Crabot et al., 2019; Legendre et al., 2015). To account

for these known biases, we used the approach proposed by (Crabot et al., 2019), which

leverages a Moran spectral randomization procedure to generate random samples which

maintain the same autocorrelation properties of the original dataset. To this goal, we used the

R packages *sdpep* to process the spatial matrices and the *msr* function of the package

adespatial. The distances above were transformed into Euclidean distances with function

cailliez from the R package *adegenet*. Using this approach, we tested the effect of habitat on

both DGGs and pairwise genetic distances, using 1000 randomizations. In both cases the effect

of microhabitat was significant (p-value<0.05).

Data and code availability

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Author Contributions

T.D-M., M.F., Y.A. and A.A.L, conceived and designed the study. T.D-M. carried out the molecular work, SNP calling and genetic analyses with input from T.J.E., F.M., M.N. and A.A.L. T.J.E. performed the simulations and microhabitat sorting analyses with input from T.D-M., M.N. and A.A.L. F.M. carried out the population differentiation analysis and the MSR-Mantel tests. T.D-M., H.S., J.M., N.A-R., A.R., M.F., Y.A, and A.A.L carried out the collection. T.D-M., T.J.E., F.M, M.N. and A.A.L wrote the manuscript. M.N. and A.A.L supervised the project.

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Ethics declaration

The authors declare no competing interests.

389 **Supporting Information**

Fig. S1. **Average yearly minimum (A) and maximum (B) temperature measured in Ayelet Haschahar weather station from 1984 to 2020**. Collection years are marked in red on the linear model regression line.

Fig. S2. **CO₂ concentration 1983 to 2021**. Data obtained from the Mauna Loa record, Scripps

Institution of Oceanography, UC San Diego.

404   Fig. S3. **SNP density of 5196 markers along the wild emmer wheat genome.**

Fig. S4. **Rarefaction curves.** Adjusted 2nd degree polynomials of 100x backwards sample of

411   unique genotype number and number of samples collected per year. Dashed horizontal line

412   represents the observed number of unique genotypes per year, dashed vertical lines represent

413   number of samples per year. Light blue line and red dot represent range and mean of inflection

414   points over replicate subsamples respectively.

420   Fig. S5. **Population structure of Ammiad in the first year of collection (1984).** Heatmap

421   colours represent identity value for each pairwise comparison, annotation colours (on the left)

422   correspond to the list of microhabitats.

Fig. S6. **Principal Component Analysis of Ammiad collection.** (A) Coloured by microhabitat. (B) Coloured by DGGs, black dots represent singleton DGGs, which are not related to each other.

431 Fig. S7. **No evidence for isolation by distance in Ammiad collection.** Scatter plots of

432 pairwise genetic identity and Euclidean distance (meters). Subplots show years of collection,

433 and colours represent the different transects (A=blue, B=red, C=yellow, D=green).

Pairwise identity

Distance (m)

436   Fig. S8. **Correlation and significance of isolation by distance in Ammiad collection.**

437   Pairwise genetic identity by Euclidean distance (meters). Subplots show years of collection

438   (rows) and transects (columns). The title of each plot indicates its number of samples and p-

439   value from a Mantel test with 1000 permutations.

451 Fig. S9. **Sorting of plants of the same DGG by microhabitat comparing observed values**

452 **(red) to values obtained by permuting habitat labels (violins).**

458 Fig. S10. **Histograms of 1000 spatially constrained randomizations of Mantel statistics**

459 **(MSR-Mantel).** Observed values (black dots) versus randomizations for (A) pairwise genetic

460 distances (p-value < 0.001) and (B) DGGs (p-value < 0.001).

Fig. S11. **Genetic differentiation between and within habitats, accounting for spatial**

distance. Plants within each habitat were grouped in 30m demes. Pairwise F_{st} was calculated

among all demes, within and between habitats. A) Pairwise F_{st} plotted along distance, measured

as the average distance of plants in the two demes. Credibility intervals of a loess fit with

degree 1 are shown in grey. B) Pairwise F_{st} between demes with average distance <50m.

Average observed values are shown as a black dot. Uncertainty was estimated with 1000

bootstraps of the parcels considered for the comparisons. Average within-habitat Fst was

higher than between-habitat F_{st} in 3% of the bootstraps. For both A and B, Nm values (i.e. the

theoretical number of migrants between parcels in an infinite island model) calculated are

shown on the right axis.

Fig. S12. **Spike density in transect A.** Spike density (corresponding to density of fertile plants)

was measured around and between sampling sites (Table S4). Points represent the number of

fertile spikes per 1 m^2 around the sampling point. Light blue bars represent normalized (to area)

number of spikes from one sampling point to the following.

Fig. S13. **Spatial clustering along a transect in simulated populations.** Subplots show the probability of observing plants of the same DGG at pairs of sampling points at increasing distances, averaged over 100 replicate simulations. Panels indicate increasing mean seed dispersal distances (left to right) and increasing outcrossing rates (top to bottom). 'Dormancy' indicates the proportion of seeds drawn from the seed bank in each generation, and 'Density' the number of plants per square meter.

Fig. S14. **Overlap in spatial clustering between observed data and populations simulated under neutral demography.** Plots show the proportion of replicate simulations where pairs sampling points at different distances were more likely to be occupied by plants of the same DGG than in any observed transect. Panels indicate increasing mean seed dispersal distances (left to right) and increasing outcrossing rates (top to bottom). 'Dormancy' indicates the proportion of seeds drawn from the seed bank in each generation, and 'Density' the number of plants per square meter.

Fig. S15. **Temporal stability in simulated populations.** Subplots show the probability of

observing plants of the same DGG at a single sampling in different years, averaged over 100

replicate simulations. Panels indicate increasing mean seed dispersal distances (left to right)

and increasing outcrossing rates (top to bottom). 'Dormancy' indicates the proportion of seeds

drawn from the seed bank in each generation, and 'density' the number of plants per square

meter.

Fig. S16. **Probabilities that observed temporal stability could be generated by neutral demography.** Plots show the proportion of replicate simulations where a single sampling point sampled in different harvest years was more likely to be occupied by plants of the same DGG than in any observed transect. Panels indicate increasing mean seed dispersal distances (left to right) and increasing outcrossing rates (top to bottom). 'Dormancy' indicates the proportion of seeds drawn from the seed bank in each generation, and 'Density' the number of plants per square meter.

Table S1. **Ammiad sampling points coordinates.**

Table S2. **Dataset of Ammiad samples by genotype.**

Table S3. **Analysis of molecular variance of genotypes sampled in Ammiad across all**
years, subdivided by habitat, using only polymorphic sites within Ammiad. Sum of
squared deviations (SSD) and mean square deviations (MSD), variance components (σ^2) and
corresponding p-values were computed using the R package *pegas*. The Φ statistic for habitat
was used as an estimate of the proportion of molecular variance explained by the habitat
structure.

Component	SSD	MSD	df	σ^2	p-value	Φ
Habitat	0.9946	0.1421	7	0.0014	$<10^{-12}$	0.2833
Error	2.9429	0.0036	824	0.0036	/	/
Total	3.9374	0.0047	831	/	/	/

553   Table S4. **Spike density measured at transect A in 2020.**

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