Interactions between evolved pesticide resistance and pesticide exposure on immunity against pathogens

Running Title: Pesticide and pathogen interactions

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Abstract

Chemical and microbial pesticides are both important tools in combating agricultural

pests and disease vectors. However, the interactive effects of chemical and microbial agents on

insect physiology could result in either antagonism or facilitation among control strategies, as

both can activate stress, detoxification, and immune responses. In natural populations, moreover,

the evolution of resistance to chemical pesticides is common and often involves differential

regulation of mechanisms involved in the primary responses to both pathogens and pesticides,

adding an additional degree of complexity for predicting net control outcomes. To investigate the

interactive effects of chemical pesticide resistance, exposure, and bacterial infection on host

outcomes, we experimentally evolved resistance to two different classes of pesticides

(organophosphates and pyrethroids) in the red flour beetle, *Tribolium castaneum*. We exposed

pesticide susceptible and resistant lines to pesticides, the entomopathogen and biocontrol agent

Bacillus thuringiensis (Bt), or both. Pesticide resistance and Bt exposure were individually

associated with slower development, indicating sub-lethal fitness costs to resistance and

infection, respectively. After organophosphate exposure, however, beetles developed more

quickly and were more likely to survive if also exposed to Bt. We used RNAseq to examine the

interactive effects of pesticide resistance, pesticide exposure, and Bt exposure on gene

expression. Pyrethroid-resistant insects exhibited dampened immune responses to Bt infection

relative to susceptible ones. In a similar vein, simultaneous exposure to organophosphates and Bt

resulted in muted stress-associated transcriptional responses compared to exposure with only one

factor. Our data suggest that evolved resistance, pesticide exposure, and biopesticide infection all

exert contrasting effects on life history parameters and host physiology that could ultimately

influence the demography and control of pest populations in complex ways.

Introduction

Control of insect agricultural pests and disease vectors remains a central problem worldwide, as widespread use of chemical pesticides has led to the rise of pesticide resistance amongst important insect species (Bass & Jones, 2018). Insect resistance management (IRM) and integrative pest management (IPM) strategies strive to overcome problems of evolved resistance by alternating or implementing the use of both chemical pesticides and biological control measures (Sappington & Miller, 2017; Zhu et al., 2016). Entomopathogens, including bacteria, fungi, and viruses, can complement or even replace the use of chemical pesticides by increasing the mortality rate and slowing development and reproduction in pest and vector populations (Cory, 2017; Liu et al., 2017). The physiological effects of chemical and microbial control agents are not completely independent, however, as both can activate stress, detoxification, and immune responses (James & Xu, 2012). Since the evolution of pesticide resistance also often arises through these mechanisms (Minetti, Ingham, & Ranson, 2020), host responses to chemical and microbial control agents could mediate facilitation or antagonism among control strategies on both proximate and evolutionary time scales (Booton et al., 2019; Hite & Cressler, 2018; Tompkins et al., 2001; Wolz et al., 2021). A better understanding of the molecular basis and phenotypic outcomes of these interactions would improve the design and predictive power of pest and vector control strategies (Bass & Jones, 2018; Sappington & Miller, 2017) while also providing fundamental insight into the integration and evolution of organismal stress and immune responses.

Pesticides target a diverse set of physiological functions in insects, ranging from neurotoxic activity to the regulation of growth and development. The mechanisms of pesticide resistance show a similarly diverse set of solutions within and among pesticide classes, including

target-site modifications, increased metabolic detoxification, and cuticular modifications (Boyer

et al., 2012; Rösner et al., 2020). Organophosphates (OP), which inhibit acetylcholinesterase

(AChE) to overexcite cholinergic synapses (Casida & Durkin, 2013), and pyrethroids (Pyr),

which disrupt voltage-gated sodium channel (vgsc) function (Davies et al., 2007), are two

pesticide classes widely used in agricultural systems and against disease vectors. Target-site

mutations and AChE gene duplications have been described for several OP-resistant insects

(Siegfried & Scharf, 2001), while Pyr resistance has been associated with mutations in the

voltage-gated sodium channel (Soderlund, 2011). Target site mutations are not the only path to

resistance, however. OP and Pyr pesticides are mainly detoxified through oxidation and

hydrolysis, and resistance associated with differential expression of diverse canonical

detoxification genes, including cytochrome P450s, esterases, and glutathione S-transferases

(GSTs), has also been described for both pesticide types (Boyer et al., 2012; Meinke et al., 2021;

Oakeshott et al., 2013; Pavlidi et al., 2018; Sogorb & Vilanova, 2002). Moreover, recent studies

(David et al., 2014; Lilly et al., 2016; Tandonnet et al., 2020; Zimmer et al., 2017) have

implicated changes in cuticle and serine endopeptidase gene expression in increased resistance to

penetration by OP and Pyr pesticides.

Recent evidence suggests that the mechanisms of pesticide resistance could also impact

immune and physiological responses against parasites (Minetti et al., 2020). For example,

esterase-mediated pesticide resistance in *Culex pipiens* is associated with changes in immune

gene expression including constitutive upregulation of antimicrobial peptide (AMP) and nitric

oxide synthase (NOS) genes (Vézilier et al., 2013). Moreover, resistance-associated constitutive

changes in metabolic detoxification mechanisms like cytochrome P450s or GSTs can alter

concentrations of damaging reactive oxygen species (ROS) that a pathogen would encounter

within the host (Minetti et al., 2020; Pang et al., 2021) and influence the success of pathogen

colonization, growth, and transmission (Minetti et al., 2020; Ndo et al., 2019; Wang et al., 2016).

For example, changes in cap'n'collar transcription factor expression have been shown to alter

detoxification gene expression and increase ROS levels, thereby conferring pesticide resistance

in several species (Kalsi & Palli, 2017; Lu et al., 2020; Misra et al., 2011) while also modifying

vector competence in *Aedes aegypti* (Bottino-Rojas et al., 2018). Mutations within *ace-1* (AChE)

and *kdr* (vgsc) genes in mosquitoes (*Anopheles gambiae*) confer target-site resistance to OP and

Pyr pesticides, respectively. Populations of mosquito strains in possession of these mutant alleles

support a higher prevalence of *Plasmodium falciparum* parasites, but *kdr* mutations were also

associated with reduced midgut oocyst burden in infected individuals (Alout et al., 2013).

Even in the absence of evolved resistance, host exposure to pesticides may impact

pathogen growth directly through contact with toxins or indirectly through the induction of insect

detoxification enzymes (Duke, 2018; Minetti et al., 2020). Exposure to pesticides can also have

complex effects on components of the cellular, humoral, and oxidative stress responses of host

immunity (James & Xu, 2012). For example, exposure to OP pesticides has been associated with

increased hemocyte numbers and phenoloxidase (PO) and encapsulation activity in wax moth

(*Galleria mellonella*) and Colorado potato beetle (*Leptinotarsa decemlineata*) larvae (Dubovskiy

et al., 2014). However, dual exposure to OP and a pathogenic virus in silkworm larvae (*Bombyx*

mori) resulted in increased mortality and differential expression of oxidative stress and AMP

genes (Gu et al., 2016). Exposure to Pyr pesticides, meanwhile, has been associated with

increased melanization responses and decreased replication of *Escherichia coli* bacteria (Hauser

& Koella, 2020) and decreased *P. falciparum* infection prevalence and intensity in *A. gambiae*

(Kristan et al., 2018). Pyr exposure is also hypothesized to impact the production of other

immune responses including serine proteases, lytic enzymes including esterases and lysozymes,

and reactive oxidative stress responses (James & Xu, 2012).

Clearly, both pesticide resistance and exposure independently have important effects on

insect immunity against pathogens. However, it remains an open question whether resistance and

exposure influence host-microbe interactions in the same direction, and whether this influence

arises through the same or different mechanisms. To address this gap in knowledge, we

experimentally evolved resistance to two different classes of pesticides (OP and Pyr) in the red

flour beetle (*Tribolium castaneum*), an agricultural pest and model for studies on insect

genomics, immunity and resistance (Herndon et al., 2020; Rösner et al., 2020). To investigate the

interactive effects of pesticide resistance, exposure, and infection, we exposed our evolved lines

to *Bacillus thuringiensis* (Bt), an entomopathogenic gram-positive bacterium that has been

developed into a widely used biopesticide (Sansinenea, 2012).

We first investigated the main and interactive effects of selection regime (pesticide

resistant and susceptible) and pesticide exposure on host fitness-associated phenotypes and

bacterial load after Bt infection. Having established these phenotypes, we turned to

transcriptional data to identify potential mechanisms that could explain the observed phenotypes.

We first asked whether, in the absence of pesticides, evolution regime alone influences

constitutive gene expression and the transcriptional response to infection. We next included

pesticide exposure into the regime-by-infection interaction to investigate whether pesticides

facilitate or antagonize the host response to infection, and whether differential gene expression

depends on the experimental evolution regime. Finally, we compared the results of OP and Pyr

treatments to determine whether evolved resistance or exposure to pesticides with different

physiological targets exert different effects on host-pathogen interactions. Our study provides a

comprehensive window into the physiological and evolutionary processes that shape interactions

among two important ecological stressors.

Methods and Materials

T. castaneum Control and Pesticide-selected Colony Maintenance

We used six *T. castaneum* populations (Adairville, Coffee, Dorris, RR, Snavely, WF

Ware) collected from the southeast USA (Jent et al., 2019) that had been maintained under

laboratory control conditions (standard diet of whole wheat flour + 5% yeast; 30°C; 70%

humidity; in the dark) for at least two years prior to the start of the experiment.

For each of the six ancestral *T. castaneum* populations, we initially exposed

approximately 200 larvae to control conditions (no pesticides) or to pesticides (organophosphate

[OP] or pyrethroid [Pyr]) (Suppl. Fig. 1). Initial exposure doses were calculated as one tenth the

manufacturer's recommended dose (OP [Malathion 50% EC, Southern Ag]- 0.103 mg/ml; Pyr

[Demon WP, 40% cypermethrin, Syngenta]- 0.0251 mg/ml). We approximated LC50 pesticide

concentrations for ancestral populations using dose response curves (Birnbaum et al., 2021), and

adjusted these after the second generation (LC50: 5.14 mg/ml OP and 0.188 mg/ml Pyr).

Pesticide solutions diluted in DI water were combined with 0.15 g/ml standard diet, and

approximately 10 ml of the pesticide- or water-diet slurry (for controls) was added to 1 L plastic

colony containers and allowed to dry overnight at 55°C. We allowed approximately 200 pupae

and larvae to develop on the pesticide diet for one week before supplementing colonies with

100ml of fresh standard diet. We transferred individuals to new pesticide exposure containers

every four weeks. After six generations of selection, we tested for increased pesticide resistance

in F2 larvae (to avoid parental effects) from the pesticide-regime populations and found these

populations had significantly increased survival against pesticides compared to control-regime

populations (Birnbaum et al., 2021). Meanwhile, we continued selection in all populations for

another two generations (to generation 8).

Bt Spore Culture Preparation

To orally expose beetles to Bt, we created a spore culture supernatant according to

previously described protocols (Milutinović et al., 2013). Briefly, we inoculated 5 ml LB liquid

media with *Bacillus thuringiensis subsp. morrisoni bv. tenebrionis* (*Btt;* obtained from the

Bacillus Genetic Stock Center, Columbus, OH) from glycerol stock and incubated overnight at

30°C. We streaked liquid *Btt* from the overnight culture onto LB agar plates and incubated the

plates overnight at 30°C. The next day, 18 individual colonies were picked and each inoculated

into 5 ml of Bt medium (w/V–0.75% bacto peptone (Sigma), 0.1% glucose, 0.34% KH_2PO_4,

0.435% K_2HPO_4) with the addition of 25 µL of salt solution (w/V–2.46% $MgSO_4$, 0.04%

$MnSO_4$, 0.28% $ZnSO_4$, and 0.40% $FeSO_4$) and 6.25 µL of 1M $CaCl_2 \cdot 2H_2O$; these cultures and a

negative control (Bt medium only) were allowed to grow overnight on a bacterial shaker at 30°C,

200 rpm. The following day, 1 ml of the liquid overnight culture was inoculated into 200 ml Bt

medium with the addition of 1 ml salt solution and 50 µl calcium chloride. These cultures and a

negative control were incubated for seven days in darkness on a bacterial shaker at 30°C, 200

rpm; additional 1 ml salts solution and 50 µl calcium chloride were added after four days of

incubation. Spore cultures were centrifuged at 4,000 rpm for 15 minutes, washed once with

insect saline, and resuspended in Millipore H_2O. We calculated the final spore concentration of

$8.62*10^9$ cells/ml by counting spores using a Hausser Scientific counting chamber.

Pesticide and Bt Spore Oral Diet Preparation

We prepared control diet by mixing 0.15 mg/ml standard diet (described above) in DI

H_2O. We prepared pesticide diets by combining approximate LC50 pesticide dilutions in H_2O

with 0.15 mg/ml standard diet (9.75 µl/ml OP and 0.188 mg/ml Pyr). *Btt* diet was prepared by

mixing the previously described $8.62*10^9$ cells/ml spore culture suspension with 0.15 mg/ml

standard diet. Pesticide + *Btt* diet was prepared by diluting 9.75 µl/ml OP or 0.188 mg/ml Pyr in

the spore culture suspension with 0.15 mg/ml standard diet. Oral diet plates were prepared by

pipetting 50 µl of the appropriate treatment (Suppl. Fig. 1; control, control + *Btt*, OP, OP + *Btt*,

Pyr, Pyr + *Btt*) into 96-well cell culture plates. Plates were covered and allowed to dry overnight

at 55°C.

We investigated the effects of diet disk preparation and pesticide treatments on Bt spore

germination by preparing control, *Btt,* and *Btt* + pesticide diet disks as described above and

counting the number of formed colonies from each treatment (*Btt*, *Btt* + OP or *Btt* + Pyr n=10;

control n=6). We stored dried diet disks for four days at 30°C in the dark and then dissolved

individual disks in 1 mL ultrapure water. To initiate germination, we heated the samples to 50°C

for ten minutes. We conducted the spore germination on two separate days. On the first day, we

cultured bacteria from a $1:10^7$ dilution in ultrapure water. On the second day, we used a $1:10^6$

dilution, to ensure that sufficient colonies would grow. For the bacteria culture plates, we spread

100µl of the final dilution onto LB agar, included a negative control with only water, incubated

plates over night at 30°C and counted all *Btt* colonies after 15 hours. We analyzed the effect of

treatment on the number of colony forming units (CFUs) using a generalized linear mixed model

with day as a random factor and a negative binomial distribution.

Btt in vitro Growth Curve Assays

To measure *in vitro* effects of pesticides on Bt growth, we performed growth curve

analyses for OP and Pyr exposure separately. For each analysis, *Btt* from a glycerol stock was

freshly streaked onto Nutrient Broth (NB) agar plates and allowed to grow overnight at 30°C.

We grew overnight cultures from single colony inoculates (n = 9) in 5 ml NB liquid media at

30°C, 200 rpm; cultures were incubated for only 16 hours to avoid entrance into sporulation

phase. The following day, we grew log phase bacterial cultures by inoculating 100 µl of the

overnight culture into 2.9 ml NB liquid media and incubated for 3.5 hours at 30°C, 200 rpm.

To replicate *in vivo* conditions, each growth curve analysis included control NB, 1:10

pesticide NB dilution, and 1:100 NB pesticide dilution treatments based on larval exposure OP

and Pyr pesticide concentrations (described above). We added 200 µl of the appropriate

treatment solution to a 96-well CYTO-One spectrophotometer plate. An aliquot of log phase

bacterial culture (n = 9) was added to each well to achieve an initial $OD_{600} = 0.4$ (3 technical

replicates/ culture). OD_{600} bacterial growth was measured every 15 minutes for 17 hours using a

BioTek Epoch 2 spectrophotometer (30°C). Three or four independent growth curve replicates

were performed for each pesticide.

We obtained growth curve metrics using the R package growthcurver (Sprouffske &

Wagner, 2016). For OP and Pyr separately, we analyzed the individual effects of replicate and

treatment on growth rate values (r) obtained from growthcurver using non-parametric Kruskal-

Wallis tests. We measured the effect of pesticide treatment separately for each replicate if the

effect of replicate was significant.

Exposure of Control and Pesticide-selected Populations to Pesticide and Bt Oral Diets

After eight generations of selection, we generated same-age F2 larvae from pesticide-

selected and susceptible colonies. Approximately 200 adults from Gen. 8 pesticide-selected and

susceptible control colonies were allowed to reproduce on control diet for three days. We

removed these F0 adults and allowed offspring to develop to F1 adults. To generate F2

experimental larvae, we placed 4-5 replicates of 80 adults per selection line in 100 mm petri

dishes with control diet and allowed them to reproduce for 24 hours. F2 larvae were allowed to

develop for 12 days.

We placed twelve day old larvae (~4 mm long) on oral diet plates prepared as described

in the previous section (Suppl. Fig. 1; n = 96/ population/ treatment). Half of the plate (n = 48)

was monitored daily for three weeks for survival and development. Individuals in the remaining

half were sampled after two days of exposure prior to the onset of mortality, and used for

RNAseq analyses (described below). Pupae were transferred to clean 96-well plates and newly

eclosed adults were collected and stored at -80°C for use qPCR analyses of Bt infection

persistence across metamorphosis (described below).

Statistical Methods for Fitness Experiments

The effects of selection regime (Regime), pesticide exposure (PTx), Bt exposure (BtTx),

and interactions between the three factors on beetle survival and development were investigated

using population as a random factor in cox mixed effects survival models with the package

coxme (Therneau, 2015) in R. For both time to death and time to develop into adulthood, we

evaluated the individual effects of pesticide exposure (PTx), selection regime (Regime), and Bt

exposure (BtTx), and the interactions between selection regime and pesticide exposure

(Reg:PTx), selection regime and Bt exposure (Reg:BtTx), and co-exposure to pesticides and Bt

(PTx:BtTx). The full model for both time to death and adult development time is as follows:

Surv(Day, Status) ~ (PTx + BtTx + Regime + Regime:PTx + Regime:BtTx + PTx:BtTx +

(1|Pop)). Hazard ratios were calculated from coxme survival models and were plotted using the

sjPlot package in R (Lüdecke , 2021).

RNA Extractions and Sequencing

Samples for 3' TagSeq (Lohman, Weber, & Bolnick, 2016) were collected after two days

of exposure to oral diets for three populations (Coffee, Dorris, Snavely), prior to the main onset

of mortality. For each population, regime, and treatment, nine surviving larvae were collected

and immediately stored at -80°C until processing, and subsequently pooled larvae into three

replicates of three larvae each. We extracted total RNA using a Qiagen RNeasy Mini kit

according to manufacturer protocols, and confirmed nucleic acid concentration and purity using

a Nanodrop. We depleted remaining DNA using the Invitrogen RNaqueous Micro DNase

treatment according to manufacturer protocols. We confirmed the concentration and quality of

these DNA-depleted RNA samples using a Bioanlayzer (RIN > 9). We generated single indexed

3' TagSeq libraries from 1 µg total RNA using the Lexogen QuantSeq 3' mRNA-Seq Library

Prep Kit according to manufacturer's protocols; libraries were amplified using 15 PCR cycles.

Library concentration and quality were confirmed prior to PE100 sequencing on four runs on the

NextSeq platform (VANTAGE, Vanderbilt University).

RNAseq Analyses

Adapters were trimmed from the raw mRNAseq reads and low quality reads were

removed using fastqc (Andrews 2010); an average of 10,869,280 reads per sample were retained

(Suppl. Table 1). Sequence reads are deposited in the NCBI SRA under the BioProject

Accessions: PRJNA753142, PRJNA756317. We took advantage of the well-annotated, published

genome for *T. castaneum* (Herndon et al., 2020) to align trimmed reads to the most recent *T.*

castaneum assembly (Tcas5.2) using BWA-MEM (avg. reads mapped = 84.89%, s.d. = 0.023;

Suppl. Table 1) (Heng Li, 2013). Gene count files were generated using samtools (H Li et al.,

2009). We used DESeq2 (Love et al., 2014) to perform differential expression analyses.

The effects of Bt on *T. castaneum* immunity have been extensively studied (Behrens et

al., 2014; Ferro et al., 2019; Greenwood et al., 2017; reviewed in Pinos et al., 2021; Tate &

Graham, 2017). However, as highlighted by a recent review (Siegwart et al., 2015), interactions

between Bt and chemical pesticides in *T. castaneum* are largely unexplored (Malik, 2012). First,

to investigate the effects of Bt exposure and pesticide selection regime on gene expression in the

absence of pesticide exposure, we subset only pesticide-free and pesticide-free + Bt libraries and

examined the effects of selection regime, Bt exposure, and their interaction on gene expression

(no pesticide model form: count matrix ~ Reg + BtTx + Reg:BtTx). Next, we analyzed the main

effects of pesticide exposure, pesticide selection regime, and Bt exposure and the interactive

effects between selection regime and pesticide exposure, selection regime and Bt exposure, and

pesticide and Bt exposure for OP and Pyr separately (OP and Pyr model form: count matrix ~

PTx + Reg + BtTx + Reg:PTx + Reg:BtTx + PTx:BtTx).

Differentially expressed transcripts were annotated using the Ensembl *T. castaneum*

database and the R package 'biomaRt' (Durinck et al., 2009). We manually curated the annotated

gene lists to categorize differentially expressed transcripts into groups potentially impacted by

pesticide resistance and exposure and Bt treatment such as those involved in detoxification,

immunity, development, and cuticle associated transcripts (Suppl. Table 2). Gene ontology (GO)

enrichment analyses for the subsets of significantly differentially expressed gene sets were

performed using the online Gene Ontology Resource (Ashburner et al., 2000; Gene Ontology

Consortium, 2021).

To analyze global patterns of gene expression, we used a weighted gene co-expression

network analysis with the R package 'wgcna' (Langfelder & Horvath, 2008). This approach

allows us to group genes with correlated expression patterns into modules and statistically

associate gene module expression with experimental factors. We constructed the initial network

using all replicate samples for all treatments (n = 126) with a variance stabilizing transformation.

A soft thresholding power of 6 was used to construct the topological overlap matrix, and we then

identified modules with highly correlated module expression patterns (p < 0.05).

Quantification of Bt in Adults

While larval exposure to pesticides and pathogens can impose costs to larval traits, recent

studies suggest that larval infections can persist across metamorphosis into adulthood (Johnston

& Rolff, 2015), thereby also imposing longer-term consequences for organismal fitness. To

investigate if larvae orally exposed to Bt harbored chronic infections as adults, we quantified *Btt*

in whole body adults using qPCR. We extracted RNA from newly eclosed individual adults

stored at -80°C using a Qiagen RNeasy Mini kit (Qiagen) according to the manufacturer's

protocols. We reverse transcribed 1µl of RNA to cDNA using the SuperScript VILO cDNA

Synthesis Kit (Invitrogen) per manufacturer's instructions. For the RTqPCR, we used the same

protocol as described in Jent et al. (2019) with the PowerUp Sybr Green Master Mix (Life

Technologies) on the QuantStudio 6 machine (Applied Biosystems). We quantified bacterial load

relative to host tissue by comparing Ct values from a Bt specific 16S rRNA amplicon (Tate &

Graham, 2017) to the beetle housekeeping gene RPS18 (Lord et al., 2010; Tate et al., 2017).

Sample sizes included 39 individuals for the control group (exposed to neither *Btt* nor

pesticides), 44 individuals for the group exposed to *Btt* but not to pesticides, 30 individuals for

the group exposed to OP and *Btt* and 13 individuals for the group exposed to Pyr and *Btt*.

Samples sizes varied between groups because we were limited to surviving adults.

For the statistical analysis, we compared the deltaCt values (housekeeping gene – target

gene) of the three Bt exposed groups to the control group in an GLMM with beetle population

and qPCR plate as random factors and a Gaussian distribution (Lord et al., 2010; Venables &

Ripley, 2002). The linearized values span orders of magnitude, so log transformation normalizes

them for statistical analysis.

Septic Infection Experiment

While we used the natural oral infection route for our experiments in this study, many

studies of host-pathogen interactions in model and non-model insects use the septic infection

route to analyze immune responses and infection outcomes. To provide context for these and

future studies, we wanted to clarify whether the interaction between pesticides and infection

outcomes might be infection route-dependent. Therefore, we exposed 14-day old larvae from

control- and OP-regimes to control and OP oral diets (as described above) for three days, and

then septically infected individuals with Bt and recorded their survival.

We produced a mixture of *Btt* cultures of vegetative cells from both the logistic growth

and early stationary phase for septic infections as previously described (Jent et al., 2019), which

caused around 50% mortality in preliminary experiments. We pricked larvae laterally between

the head and second segment with an ultra-fine borosilicate needle dipped in the bacterial

suspension and recorded survival after 24 hours. We repeated the experiment three times with 20

individuals per population (Coffee or Dorris), selection regime (OP selected and control) and

pesticide treatment (OP exposure and control) each (n = 480).

Differences in survival between the control group (control selection regime, no pesticide)

and the other groups were analyzed in a GLMM with experiment (block) and population as

random factors and a binomial error distribution (Bates et al., 2015).

Results

To investigate how evolved pesticide resistance and pesticide exposure impact host

interactions with pathogens, we exposed control-regime and pesticide-selected populations to

control and pesticide oral treatments, with and without Bt. We measured the main and interactive

effects of pesticide selection regime (Reg), pesticide treatment (PTx), and Bt treatment (BtTx) on

survival, development, and gene expression for OP and Pyr separately.

Effects of Pesticide Selection and Exposure to Pesticides and Bt on Survival and Development

Overall, exposure to both pesticide types, OP (OP PTx) and Pyr (Pyr PTx), decreased

survival likelihood, while there was no overall effect of Bt treatment (BtTx) on survival (Figs.

1A-B; Suppl. Table 3). Pesticide selection significantly improved survival after pesticide

exposure compared to control-regime populations, reinforcing previous observations of evolved

pesticide resistance (Suppl. Figs. 2A-B, compare OP Reg:PTx and Pyr Reg:PTx; Suppl. Table 3).

Interestingly, there was a positive significant interaction between OP and Bt treatment indicating

that larvae dually exposed to OP and *Btt* had improved survival compared to larvae exposed to

OP only (Fig. 1A, compare OP PTx and BtTx; Suppl. Fig. 2A; Suppl. Table 3). There was no

significant interaction between Pyr exposure and Bt treatment (Pyr PTx:BtTx) or between either

pesticide selection regime and Bt treatment (OP Reg:BtTx, Pyr Reg:BtTx) on survival (Fig. 1B;

Suppl. Table 3).

Pesticide selection regime (OP Regime, Pyr Regime), pesticide exposure (OP PTx, Pyr

PTx), and Bt treatment (BtTx) individually delayed development, indicating a cost from each

(Figs. 1C-D; Suppl. Table 3). There was no significant interaction between pesticide regime and

Bt treatment on development (OP Reg:BtTx, Pyr Reg:BtTx; Suppl. Table 3). However, resistant

individuals developed faster than susceptible individuals in the presence of pesticides, as

indicated by a significant interaction between pesticide regime and treatment (OP Reg:PTx, Pyr

Reg:PTx; Fig. 1C, D; Suppl Fig. 2C. D; Suppl Table 3). There was also a significant positive

interaction effect between both pesticide treatments and *Btt* exposure (OP PTx:BtTx, Pyr

PTx:BtTx), indicating that larvae exposed to both Bt and pesticide treatments developed faster

than those exposed to only pesticide or only *Btt* (Suppl. Figs. 2C-D; Suppl. Table 3).

Direct Effects of Pesticides on Bt

To investigate direct interactions between pesticides and Bt, we used *in vitro* growth

curves and viable spore counts of oral diet disks. *In vitro Btt* growth rate significantly increased

upon OP exposure, but there was no effect of Pyr exposure on *in vitro* growth rate (Suppl. Fig. 3;

Suppl. Table 4). The germination rates of *Btt* spores on the diet disks exhibited no significant

differences among pesticide treatments (Suppl. Fig. 4; Suppl. Table 5).

The influence of pesticide exposure on route-specific infection outcomes

While oral infection with *Btt* did not result in immediate excess mortality compared to

unexposed individuals, we considered the possibility that infection outcomes could be route-

specific or lead to chronic rather than acute infections. To investigate if larvae exposed to *Btt*

from their diet disks remained infected as adults, we quantified Bt in newly emerged adults using

qPCR. We found that adults exposed as larvae to *Btt* but not pesticides still harbored Bt

infections, while Bt loads were not detectable in individuals raised on either control or pesticide

+ *Btt* diet disks (Fig. 2A; Suppl. Table 6). Thus, dual exposure with either pesticide prevented

chronic Bt infection in adults.

To determine whether the interaction between pesticides and infection outcomes might be

route-dependent, we septically infected larvae from each OP regime and pesticide treatment.

Individuals from the susceptible regime that were simultaneously exposed to OP had lower

survival against septic *Btt* infection compared to the other groups (Fig. 2B; Suppl. Table 7)

indicating an outsized cost of combined OP exposure and septic Bt infection in control regime

larvae. This result contrasts with the improved survival observed in the Bt x OP treatment

interaction term after oral exposure (Fig. 1A, OP PTx:BtTx).

Effects of Pesticide Selection and Btt Exposure on Gene Expression in the Absence of Pesticide

Exposure

We investigated the interaction between pesticide selection regime and Bt exposure on

gene expression in the absence of pesticides, as this provides information on the main host

response to oral Bt infection as well as whether the evolution of pesticide resistance modifies the

host response to Bt. Using only the data from individuals never exposed to pesticides, we

identified 104 genes that were differentially expressed in *Btt*-exposed individuals (Fig. 3A,

BtTx). This gene set largely concurred with previously published results on Bt infection in *T.*

castaneum (Behrens et al., 2014; Greenwood et al., 2017; Tate & Graham, 2017), and included

the upregulation of AMPs (two attacins and two defensins), a pathogenesis-related protein, and a

cytochrome P450 (Fig. 4, no pesticide model BtTx). GO enrichment analyses revealed the

importance of immune system processes and defense against bacteria, among other immune

related terms, in the genes differentially expressed with *Btt* treatment in the absence of pesticide

exposure (Suppl. Fig. 5A, no pesticide model BtTx).

Only two transcripts (an elastin and one unannotated transcript) were upregulated upon

Btt infection in OP-selection regime individuals relative to susceptible ones (Fig. 3A, no

pesticide model OP Reg:BtTx interaction), indicating that evolved resistance to OP minimally

affects the baseline response to infection in the absence of pesticides. Fourteen transcripts were

differentially expressed in Pyr-selected individuals relative to susceptible-regime individuals

after Btt infection (Fig. 3A; no pesticide model Pyr Reg:BtTx interaction), including the

downregulation of two defensins, a pathogenesis-related protein, and a histidine-rich

glycoprotein (Fig. 4B, no pesticide model Pyr Reg:BtTx). Notably, these genes were

differentially expressed in the opposite direction in Pyr regime-Btt infected larvae compared to

gene expression induced by the main effect of Bt exposure, indicative of dampened expression of

these genes (Fig. 4B, no pesticide model Pyr Reg:BtTx vs. BtTx). GO enrichment analysis

revealed significant enrichment of humoral immune response and defense response among the

differentially expressed genes from this interaction term (Suppl. Fig. 5A, Pyr Reg:BtTx).

Effects of Pesticide Selection and Dual Exposure to Pesticides and Pathogens on Gene

Expression

Next, we introduced the pesticide-exposed samples into the analysis, still treating OP and

Pyr-associated treatments separately, to investigate interactions between pesticide selection

regime, pesticide exposure, and Bt exposure on gene expression. For both pesticide class

analyses, pesticide exposure (OP PTx, Pyr PTx) had the greatest effect on gene expression and

induced differential expression of many canonical detoxification genes, including several

cytochrome P450s, esterases, GSTs, UDP-glucuronosyltransferases (UGTs), and ABC

transporters (Figs. 3B-C; Suppl. Fig. 6). Genes differentially expressed after both OP and Pyr

exposure were commonly enriched for GO processes involved in xenobiotic and detoxification

responses including oxidoreductase and serine hydrolase activity (Suppl. Fig. 5B).

Genes differentially expressed with respect to pesticide selection regime and with respect

to the interaction between selection regime and pesticide treatment can identify genes associated

with pesticide resistance. Overall, Pyr-regime had a greater effect on differential gene expression

compared to OP-regime (Fig. 3). Similarly, a greater number of genes were differentially

expressed with the interaction between Pyr regime and Pyr exposure compared to OP regime and

OP exposure (Figs. 3B, C). Both OP-regime and Pyr-regime induced differential expression of

canonical detoxification (e.g., P450s, esterases, GSTs, and ABC transporters) and cuticle-

associated genes (Suppl. Figs. 6, 7). GO enrichment analyses revealed the importance of

structural constituent of cuticle for genes differentially expressed with OP regime and the

interaction between OP regime and OP exposure (Suppl. Fig. 5B; OP Reg and OP Reg:PTx).

Pyr-regime individuals differentially expressed genes enriched for molecular functions involved

in oxidoreductase activity, heme binding, and structural constituent of cuticle, while genes

differentially expressed with the interaction between Pyr regime and Pyr exposure were enriched

for these functions and also serine hydrolase activity (Suppl. Fig. 5B; Pyr Reg and Pyr Reg:PTx).

In the models containing pesticide exposure samples, no genes were differentially

expressed with the interaction between OP regime and Btt treatment (Fig. 3B; OP Reg:BtTx). In

contrast, 61 genes were differentially expressed with the interaction between Pyr selection

regime and Btt treatment (Fig. 3C; Pyr Reg:BtTx), although there was no significant GO term

enrichment. Pyr regime larvae had lower induced expression of a defensin-like transcript

compared to control-regime larvae when exposed to Btt (Fig. 5B).

When considering the interaction effect between pesticide exposure and Bt infection, 175

genes were differentially expressed with dual OP and Btt exposure (Fig. 3B; OP PTx:BtTx), and

only one gene (an upregulated pupal cuticle protein) was differentially expressed with dual

exposure to Pyr and Btt (Fig. 3C; Pyr PTx:BtTx). OP pesticide exposure interacted with Btt

exposure to induce differential gene expression of immune, detoxification, cuticle, and

development genes compared to expression induced by the main effect of Btt exposure (Fig. 4A).

Further analysis revealed that these interaction effects were mainly driven by significantly

dampened upregulation upon combined OP and Btt exposure compared to upregulation induced

with Btt alone; transcripts downregulated with Btt alone were not differentially expressed or

were upregulated with combined OP and Btt exposure (Fig. 4A). For example, while Btt

exposure alone induced upregulation of two attacins, three cytochrome P450s, and several

cuticular proteins, larvae exposed to both OP and Btt had more modest differential expression of

these same transcripts (e.g. *attacin-1* in Fig. 5A). Genes differentially expressed with the

interaction between OP treatment and Btt exposure were enriched for GO terms involved in the

molecular functions of oxidoreductase activity and structural constituent of cuticle (Suppl. Fig.

5B; OP PTx:BtTx).

Correlated patterns of gene expression using WGCNA

Weighted gene correlation network analysis (WGCNA) allows for the identification of

genes with expression correlated and significantly associated with experimental factors. We

identified 25 modules containing genes with highly correlated expression, and several of these

modules significantly correlated with one or more experimental factors (OP Regime, Pyr

Regime, OP PTx, Pyr PTx, BtTx; Suppl. Fig. 8). We mapped differentially expressed transcripts

to each gene module and identified one that was significantly associated with all five

experimental factors (Suppl. Fig. 8; magenta module) and was also associated with a high

number of significantly differentially expressed transcripts (Suppl. Fig. 8; 201 DE transcripts/

258 total). GO enrichment analysis of the magenta module transcripts revealed the importance of

genes involved in hydrolase activity (GO:0016787; FDR = $6.07*10^{-7}$) and monooxygenase

activity (GO:0004497; FDR = $1.81*10^{-6}$).

Discussion

Our study experimentally evolved pesticide resistance in flour beetles (*T. castaneum*) to

investigate the interactive effects of pesticide resistance and pesticide exposure on immunity and

host-pathogen interactions. Pesticide-resistant insects and insects exposed to Btt both developed

more slowly than their counterparts, indicating sub-lethal fitness costs to resistance and infection,

respectively. We found that Pyr-resistant insects exhibited dampened immune responses to *Btt*

infection relative to their control-regime counterparts, although overall survival and development

did not differ among OP- and Pyr-resistant and susceptible insects after *Btt* exposure. While the

evolution of resistance did not substantially influence host infection phenotypes, exposure to

pesticides did. Specifically, we observed that larvae were more likely to survive OP exposure if

they were also orally exposed to *Btt*, with gene expression patterns suggesting dampened stress

responses in those beetles. Taken together, our data suggest that evolved resistance, pesticide

exposure, and biopesticide infection all exert contrasting effects on life history parameters that

could ultimately influence the demography and control of pest populations in complex ways. The

factorial nature of our study design, moreover, provides a new window into the physiological

basis and fitness consequences of host responses to overlapping environmental stressors.

Insects, including *T. castaneum*, can rapidly evolve resistance to pesticides through

mechanisms that enhance metabolic detoxification and cuticular modifications (Rösner et al.,

2020). Here, populations selected for pesticide resistance evolved it within 6-8 generations

(Birnbaum et al., 2021), demonstrating increased survival and faster development after pesticide

exposure compared to control-selected populations. Differential gene expression between

pesticide- and control- selection regimes suggested that OP resistance is associated with

constitutive changes in cuticle gene expression while Pyr resistance is associated with

oxidoreductase activity, heme binding, and cuticular modifications, consistent with previous

studies in OP and Pyr resistant insects (Carvalho et al., 2013; Tandonnet et al., 2020; Zimmer et

al., 2017). Thus, while both OP and Pyr resistance involved changes in cuticle gene expression,

Pyr resistance was also associated with changes in metabolic detoxification.

Oral Bt infection did not decrease survival in our study but was associated with slower

development in *T. castaneum* larvae, indicating a cost of infection as previously observed in the

same system (Tate & Graham, 2015) and in other insects challenged with Bt toxins (Hernández-

Martínez et al., 2017; Rahman et al., 2011). Previously documented variation in resistance in *T.*

castaneum against Cry3 proteins effective against other Tenebrionid beetles may explain why *Btt*

had negative effects on development and could colonize our insects but not instigate high

mortality (Milutinović et al., 2013; Oppert et al., 2011). Exposure to *Btt* induced the upregulation

of AMPs (attacins and defensins) and one pathogenesis-related protein as well as differential

expression of detoxification genes, including P450s and a CEH, GST, an ABC transporter, and

several cuticle and development related genes. Previous transcriptomic studies of oral Bt

infection in *T. castaneum* have identified similar transcripts, including attacin and defensin

AMPs, a pathogenesis-related protein, a P450, and a GST (Behrens et al., 2014; Greenwood et

al., 2017), supporting the assumption that *Btt* oral exposure in our experiments presented an

immune challenge despite the lack of infection-induced mortality.

Previous work has suggested that exposure to OP and Pyr pesticides can reduce defenses

against pathogens and stimulate an array of changes in humoral, cellular, and oxidative stress

components of immunity (Dubovskiy et al., 2014; Gu et al., 2016; Hauser & Koella, 2020). Our

results reveal a strong interaction effect between Bt and OP exposure, in particular, that

manifests at both phenotype and gene expression levels. For example, dual oral exposure to *Btt*

and OP improved larval survival and partially rescued development time relative to OP exposure

alone, without exacerbating *Btt*-induced mortality. At the same time, *Btt* co-exposure modified

the differential expression of 175 genes relative to OP exposure alone (Fig. 3B), contrasting with

the negligible effect observed in the comparable Pyr + *Btt* scenario. It is possible that OP and *Btt*

antagonize each other indirectly, e.g. through induction of cross-protective host defenses. For

example, while *Btt* exposure induced upregulation of attacins, cytochrome P450s, and cuticle-

related transcripts, combined OP and *Btt* exposure dampened the expression of these transcripts.

Interestingly, single exposure to OP or *Btt* induced downregulation of an ABC transporter but

this transcript was upregulated with co-exposure to OP and *Btt*. ABC transporters have been

identified as receptors for Bt Cry proteins and aid in membrane pore formation (Güney et al.,

2021; Heckel, 2021; Pauchet et al., 2016), and the additional function of ABC transporters as

xenobiotic pumps may help to mediate Bt toxicity (Dermauw & Leeuwen, 2014; Wu et al.,

2019). It is also possible that OP and *Btt* interact directly, as would be the case if *Btt* actively

degrades OP for energy (Kumar et al., 2018) or if OP facilitates *Btt* germination rates. Our *in*

vitro data suggest that *Btt* vegetative cells in liquid medium replicate faster in the presence of

OP, lending credence to a direct interaction. On the other hand, there was no difference in

germinated spore counts between control-diet and OP-diet disks, suggesting that direct

interactions, if they matter *in vivo*, are not relevant until after the host ingests the diet.

Notably, we observed that larvae reared on diets containing both pesticides and *Btt* did

not have detectable Bt loads as adults, in contrast to the chronically infected adults reared on *Btt*

alone. The detection of Bt in adults is perhaps surprising given the shedding of the meconium

during metamorphosis (Castagnola & Jurat-Fuentes, 2016; Hakim et al., 2010; Moll et al., 2001).

However, previous studies have suggested that gut bacteria can persist through development into

adulthood (Hammer et al., 2014; Johnston & Rolff, 2015). It is possible that pesticide exposure

induced an excessive cost that selectively killed individuals with the highest *Btt* loads, but the

survival data contradict this explanation because pesticide-exposed individuals actually survived

as well or better with Btt than without. It is also possible that pesticide-induced mechanisms

prevented Bt from establishing an infection in the host or helped hosts clear infection.

Supporting the idea of localized cross-protective mechanisms, a previous study of oral vs. septic

Btt infection in *T. castaneum* demonstrated that oral infection induced greater changes in cuticle-

related gene expression compared to septic infection (Behrens et al., 2014). In our study, OP-

exposed individuals were not protected from greater mortality upon septic Bt infection,

suggesting that infection route has important implications for pesticide-pathogen interactions.

Future work dissecting the relative importance of these direct and indirect interactions will be

crucial for refining integrative pest management (IPM) strategies employing both chemical and

biopesticides.

The impact of evolved resistance on immunity is still poorly understood (Minetti et al.,

2020), and in the absence of pesticides we did not observe any significant interactions between

pesticide resistance and Bt exposure on survival or development. Few genes were differentially

expressed between OP resistant and susceptible beetles after *Btt* exposure, but *Btt* exposure in

Pyr-resistant populations was associated with dampened upregulation of several genes, including

two defensins, a histidine-rich glycoprotein, and a pathogenesis-related protein, that were more

strongly upregulated in Pyr-susceptible populations after Btt exposure. It is also possible that

Pyr-resistant populations had constitutively higher immune gene expression resulting in higher

immune activation upon Bt infection. Moreover, an apolipophorin-III transcript, previously

shown to be important in immunity against coleopteran-specific Bt toxins (Contreras et al.,

2013), was upregulated in Pyr-resistant populations. In contrast, cuticle-related protein and

neuropeptide transcripts were downregulated in Pyr-regime larvae exposed to *Btt*.

While we did not observe any phenotypic differences with pesticide selection regime and

Btt infection here, we also did not observe mortality from *Btt* alone and it is possible that the

immune challenge was not strong enough to elicit an interaction with resistance mechanisms.

However, the differential expression of immune and cuticle functions in Pyr-regime larvae

exposed to *Btt* could have important implications for insect interactions with other pathogens or

xenobiotic pressures. Furthermore, changes in oxidoreductase activity observed after Pyr

selection are likely to alter the status of reactive oxygen species (ROS) within the host and may

impact immunity against different pathogens (Minetti et al., 2020; Wang et al., 2016). Finally, it

is possible that adaptation to pesticides could constrain host evolutionary responses to pathogens,

or vice versa. Future experiments could simultaneously select for dual resistance to both

pesticides and pathogens to see if evolutionary trajectories are similar or constrained relative to

selection under a single stressor.

Conclusions

A better understanding of the complex interactions between pesticide exposure and

resistance for host-pathogen outcomes is critical for effective pest and vector control strategies

(James & Xu, 2012; Minetti et al., 2020; Orr et al., 2020). Here, we identified antagonistic

interactions between Bt and OP that mitigated the effects of pesticides on fitness-related host

traits. Our results suggest that both pesticides and Bt individually influence important

demographic parameters like development time and mortality rate, but that co-exposed hosts

may actually perform better than those exposed to pesticides alone. We also observed dampened

expression of important candidate genes upon dual exposure to OP and Btt and in Pyr-resistant

individuals exposed to Btt compared to the main effect of Bt exposure. Are these within-host

interaction effects sufficient to alter the dynamics of disease and economic damage in pest

populations? Mathematical models would be useful to address this question, as the answer will

likely depend on feedbacks from ecological variables like intraspecific competition, relative

transmission among single- and co-exposed hosts, and the relative costs of pesticides and

infection to host reproductive rates (Booton et al., 2019; Booton et al., 2018; Kuniyoshi &

Santos, 2017). Moreover, our data suggest that the rise of pesticide resistance within populations

could reduce the magnitude of the interaction effects, likely altering population outcomes.

Overall, our results shine light on the physiological and evolutionary processes that shape

interactions among two important ecological stressors, and should prompt further study of their

interactions at multiple levels of biological organization on ecological and evolutionary time

scales.

Data Accessibility: Sequence reads are deposited in the NCBI SRA under the BioProject

Accessions: PRJNA753142, PRJNA756317.

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Literature Cited

Alout, H., Ndam, N. T., Sandeu, M. M., Djégbe, I., Chandre, F., Dabiré, R. K., … Cohuet, A. (2013). Insecticide Resistance Alleles Affect Vector Competence of *Anopheles gambiae* s.s. for *Plasmodium falciparum* Field Isolates. *PLoS ONE*, *8*(5), e63849. https://doi.org/10.1371/journal.pone.0063849

Andrews, S. 2010. FastQC: A Quality Control Tool for High Throughput Sequence Data [Online]. Available online at: http://www.bioinformatics.babraham.ac.uk/projects/fastqc/

Ashburner, M., Ball, C. A., Blake, J. A., Botstein, D., Butler, H., Cherry, J. M., … Sherlock, G. (2000). Gene ontology: tool for the unification of biology. The Gene Ontology Consortium. *Nature Genetics*, *25*(1), 25–29. https://doi.org/10.1038/75556

Bass, C., & Jones, C. M. (2018). Editorial overview: Pests and resistance: Resistance to pesticides in arthropod crop pests and disease vectors: mechanisms, models and tools. *Current Opinion in Insect Science*, *27*, iv–vii. https://doi.org/10.1016/j.cois.2018.04.009

Bates, D., Mächler, M., Bolker, B., & Walker, S. (2015). Fitting Linear Mixed-Effects Models Using lme4. *Journal of Statistical Software*, *67*(1). https://doi.org/10.18637/jss.v067.i01

Behrens, S., Peuß, R., Milutinović, B., Eggert, H., Esser, D., Rosenstiel, P., … Kurtz, J. (2014). Infection routes matter in population-specific responses of the red flour beetle to the entomopathogen *Bacillus thuringiensis*. *BMC Genomics*, *15*(1), 445–17. https://doi.org/10.1186/1471-2164-15-445

Birnbaum, S. S. L., Schulz, N. K. E., Garrett, D. S., & Tate, A. T. (2021). Experimental evolution of insect resistance to two pesticide classes reveals mechanistic diversity and context-dependent fitness costs. *BioRxiv*, 2021.09.03.458899. https://doi.org/10.1101/2021.09.03.458899

Booton, R. D., Iwasa, Y., & Childs, D. Z. (2019). How do toxicants affect epidemiological dynamics? *Oikos*, *128*(5), 729–740. https://doi.org/10.1111/oik.05654

Booton, R. D., Yamaguchi, R., Marshall, J. A. R., Childs, D. Z., & Iwasa, Y. (2018). Interactions between immunotoxicants and parasite stress: Implications for host health. *Journal of Theoretical Biology*, *445*, 120–127. https://doi.org/10.1016/j.jtbi.2018.02.018

Bottino-Rojas, V., Talyuli, O. A. C., Carrara, L., Martins, A. J., James, A. A., Oliveira, P. L., & Paiva-Silva, G. O. (2018). The redox-sensing gene Nrf2 affects intestinal homeostasis, insecticide resistance, and Zika virus susceptibility in the mosquito *Aedes aegypti*. *Journal of Biological Chemistry*, *293*(23), 9053–9063. https://doi.org/10.1074/jbc.ra117.001589

Boyer, S., Zhang, H., & Lempérière, G. (2012). A review of control methods and resistance mechanisms in stored-product insects. *Bulletin of Entomological Research*, *102*(2), 213–229. https://doi.org/10.1017/s0007485311000654

Carvalho, R. A., Omoto, C., Field, L. M., Williamson, M. S., & Bass, C. (2013). Investigating
the Molecular Mechanisms of Organophosphate and Pyrethroid Resistance in the Fall
Armyworm *Spodoptera frugiperda*. *PLoS ONE*, *8*(4), e62268-11.
https://doi.org/10.1371/journal.pone.0062268

Casida, J. E., & Durkin, K. A. (2013). Neuroactive insecticides: targets, selectivity, resistance,
and secondary effects. *Annual Review of Entomology*, *58*, 99–117.
https://doi.org/10.1146/annurev-ento-120811-153645

Castagnola, A., & Jurat-Fuentes, J. L. (2016). Intestinal regeneration as an insect resistance
mechanism to entomopathogenic bacteria. *Current Opinion in Insect Science*, *15*, 104–110.
https://doi.org/10.1016/j.cois.2016.04.008

Consortium, G. O. (2021). The Gene Ontology resource: enriching a GOld mine. *Nucleic Acids
Research*, *49*(D1), D325–D334. https://doi.org/10.1093/nar/gkaa1113

Contreras, E. a, Rausell, C., & Real, M. D. (2013). *Tribolium castaneum* Apolipophorin-III acts
as an immune response protein against *Bacillus thuringiensis* Cry3Ba toxic activity. *Journal
of Invertebrate Pathology*, *113*(3), 209–213. https://doi.org/10.1016/j.jip.2013.04.002

Cory, J. S. (2017). Evolution of host resistance to insect pathogens. *Current Opinion in Insect
Science*, *21*, 54–59. https://doi.org/10.1016/j.cois.2017.04.008

David, J.-P., Faucon, F., Chandor-Proust, A., Poupardin, R., Riaz, M. A., Bonin, A., … Reynaud,
S. (2014). Comparative analysis of response to selection with three insecticides in the dengue
mosquito *Aedes aegypti* using mRNA sequencing. *BMC Genomics*, *15*(1), 174.
https://doi.org/10.1186/1471-2164-15-174

Davies, T. G. E., Field, L. M., Usherwood, P. N. R., & Williamson, M. S. (2007). DDT,
pyrethrins, pyrethroids and insect sodium channels. *IUBMB Life*, *59*(3), 151–162.
https://doi.org/10.1080/15216540701352042

Dermauw, W., & Leeuwen, T. V. (2014). The ABC gene family in arthropods: comparative
genomics and role in insecticide transport and resistance. *Insect Biochemistry and Molecular
Biology*, *45*, 89–110. https://doi.org/10.1016/j.ibmb.2013.11.001

Dubovskiy, I. M., Yaroslavtseva, O. N., Kryukov, V. Y., Benkovskaya, G. V., & Glupov, V. V.
(2014). An increase in the immune system activity of the wax moth *Galleria mellonella* and
of the Colorado potato beetle *Leptinotarsa decemlineata* under effect of organophosphorus
insecticide. *Journal of Evolutionary Biochemistry and Physiology*, *49*(6), 592–596.
https://doi.org/10.1134/s0022093013060066

Duke, S. O. (2018). Interaction of Chemical Pesticides and Their Formulation Ingredients with
Microbes Associated with Plants and Plant Pests. *Journal of Agricultural and Food
Chemistry*, *66*(29), 7553–7561. https://doi.org/10.1021/acs.jafc.8b02316

Durinck, S., Spellman, P. T., Birney, E., & Huber, W. (2009). Mapping identifiers for the
integration of genomic datasets with the R/Bioconductor package biomaRt. *Nature Protocols*,
4(8), 1184–1191. https://doi.org/10.1038/nprot.2009.97

Ferro, K., Peuß, R., Yang, W., Rosenstiel, P., Schulenburg, H., & Kurtz, J. (2019). Experimental
evolution of immunological specificity. *Proceedings of the National Academy of Sciences of*
the United States of America, *116*(41), 20598–20604.
https://doi.org/10.1073/pnas.1904828116

Greenwood, J. M., Milutinović, B., Peuß, R., Behrens, S., Esser, D., Rosenstiel, P., … Kurtz, J.
(2017). Oral immune priming with *Bacillus thuringiensis* induces a shift in the gene
expression of *Tribolium castaneum* larvae. *BMC Genomics*, *18*(1), 1–14.
https://doi.org/10.1186/s12864-017-3705-7

Gu, Z., Li, F., Hu, J., Ding, C., Wang, C., Tian, J., … Li, B. (2016). Sublethal dose of phoxim
and *Bombyx mori* nucleopolyhedrovirus interact to elevate silkworm mortality. *Pest*
Management Science, *73*(3), 554–561. https://doi.org/10.1002/ps.4326

Güney, G., Cedden, D., Hänniger, S., Heckel, D. G., Coutu, C., Hegedus, D. D., … Toprak, U.
(2021). Silencing of an ABC transporter, but not a cadherin, decreases the susceptibility of
Colorado potato beetle larvae to *Bacillus thuringiensis ssp. tenebrionis* Cry3Aa toxin.
Archives of Insect Biochemistry and Physiology, *108*(2), e21834.
https://doi.org/10.1002/arch.21834

Hakim, R. S., Baldwin, K., & Smagghe, G. (2010). Regulation of Midgut Growth, Development,
and Metamorphosis. *Entomology*, *55*(1), 593–608. https://doi.org/10.1146/annurev-ento-
112408-085450

Hammer, T. J., McMillan, W. O., & Fierer, N. (2014). Metamorphosis of a Butterfly-Associated
Bacterial Community. *PLoS ONE*, *9*(1), e86995.
https://doi.org/10.1371/journal.pone.0086995

Hauser, G., & Koella, J. C. (2020). Larval exposure to a pyrethroid insecticide and competition
for food modulate the melanisation and antibacterial responses of adult *Anopheles gambiae*.
Scientific Reports, *10*(1), 1364. https://doi.org/10.1038/s41598-020-58415-7

Heckel, D. G. (2021). The Essential and Enigmatic Role of ABC Transporters in Bt Resistance
of Noctuids and Other Insect Pests of Agriculture. *Insects*, *12*(5), 389.
https://doi.org/10.3390/insects12050389

Hernández-Martínez, P., Gomis-Cebolla, J., Ferré, J., & Escriche, B. (2017). Changes in gene
expression and apoptotic response in *Spodoptera exigua* larvae exposed to sublethal
concentrations of Vip3 insecticidal proteins. *Scientific Reports*, *7*(1), 16245.
https://doi.org/10.1038/s41598-017-16406-1

Herndon, N., Shelton, J., Gerischer, L., Ioannidis, P., Ninova, M., Dönitz, J., … Bucher, G.
718   (2020). Enhanced genome assembly and a new official gene set for *Tribolium castaneum*.
719   *BMC Genomics*, *21*, 47. https://doi.org/10.1186/s12864-019-6394-6

720   Hite, J. L., & Cressler, C. E. (2018). Resource-driven changes to host population stability alter
721   the evolution of virulence and transmission. *Philosophical Transactions of the Royal Society*
722   *B: Biological Sciences*, *373*(1745), 20170087–11. https://doi.org/10.1098/rstb.2017.0087

723   James, R. R., & Xu, J. (2012). Mechanisms by which pesticides affect insect immunity. *Journal*
724   *of Invertebrate Pathology*, *109*(2), 175–182. https://doi.org/10.1016/j.jip.2011.12.005

725   Jent, D., Perry, A., Critchlow, J., & Tate, A. T. (2019). Natural variation in the contribution of
726   microbial density to inducible immune dynamics. *Molecular Ecology*, *28*(24), 5360–5372.
727   https://doi.org/10.1111/mec.15293

728   Johnston, P. R., & Rolff, J. (2015). Host and Symbiont Jointly Control Gut Microbiota during
729   Complete Metamorphosis. *PLoS Pathogens*, *11*(11), e1005246.
730   https://doi.org/10.1371/journal.ppat.1005246

731   Kalsi, M., & Palli, S. R. (2017). Cap n collar transcription factor regulates multiple genes coding
732   for proteins involved in insecticide detoxification in the red flour beetle, *Tribolium*
733   *castaneum*. *Insect Biochemistry and Molecular Biology*, *90*, 43–52.
734   https://doi.org/10.1016/j.ibmb.2017.09.009

735   Kristan, M., Abeku, T. A., & Lines, J. (2018). Effect of environmental variables and kdr
736   resistance genotype on survival probability and infection rates in *Anopheles gambiae* (s.s.).
737   *Parasites & Vectors*, *11*(1), 560. https://doi.org/10.1186/s13071-018-3150-8

738   Kumar, S., Kaushik, G., Dar, M. A., Nimesh, S., López-Chuken, U. J., & Villarreal-Chiu, J. F.
739   (2018). Microbial Degradation of Organophosphate Pesticides: A Review. *Pedosphere*, *28*(2),
740   190–208. https://doi.org/10.1016/s1002-0160(18)60017-7

741   Kuniyoshi, M. L. G., & Santos, F. L. P. dos. (2017). Mathematical modelling of vector-borne
742   diseases and insecticide resistance evolution. *Journal of Venomous Animals and Toxins*
743   *Including Tropical Diseases*, *23*(1), 34. https://doi.org/10.1186/s40409-017-0123-x

744   Langfelder, P., & Horvath, S. (2008). WGCNA: an R package for weighted correlation network
745   analysis. *BMC Bioinformatics*, *9*(1), 559–559. https://doi.org/10.1186/1471-2105-9-559

746   Li, H, Handsaker, B., Wysoker, A., Fennell, T., Ruan, J., Homer, N., … Subgroup, 1000
747   Genome Project Data Processing. (2009). The Sequence Alignment/Map format and
748   SAMtools. *Bioinformatics*, *25*(16), 2078–2079. https://doi.org/10.1093/bioinformatics/btp352

749   Li, Heng. (2013). Aligning sequence reads, clone sequences and assembly contigs with BWA-
750   MEM. *ArXiv.v*, 1–3. Retrieved from https://arxiv.org/pdf/1303.3997.pdf

Lilly, D. G., Latham, S. L., Webb, C. E., & Doggett, S. L. (2016). Cuticle Thickening in a
Pyrethroid-Resistant Strain of the Common Bed Bug, *Cimex lectularius* L. (Hemiptera:
Cimicidae). *PLoS ONE*, *11*(4), e0153302. https://doi.org/10.1371/journal.pone.0153302

Liu, F., Huang, W., Wu, K., Qiu, Z., Huang, Y., & Ling, E. (2017). *Exploiting Innate Immunity*
for Biological Pest Control (1st ed.). Elsevier Ltd. https://doi.org/10.1016/bs.aiip.2017.02.001

Lohman, B. K., Weber, J. N., & Bolnick, D. I. (2016). Evaluation of TagSeq, a reliable low-cost
alternative for RNAseq. *Molecular Ecology Resources*, *16*(6), 1315–1321.
https://doi.org/10.1111/1755-0998.12529

Lord, J. C., Hartzer, K., Toutges, M., & Oppert, B. (2010). Evaluation of quantitative PCR
reference genes for gene expression studies in *Tribolium castaneum* after fungal challenge.
Journal of Microbiological Methods, *80*(2), 219–221.
https://doi.org/10.1016/j.mimet.2009.12.007

Love, M. I., Huber, W., & Anders, S. (2014). Moderated estimation of fold change and
dispersion for RNA-seq data with DESeq2. *Genome Biology*, *15*(12), 550.
https://doi.org/10.1186/s13059-014-0550-8

Lu, K., Cheng, Y., Li, W., Li, Y., Zeng, R., & Song, Y. (2020). Activation of CncC pathway by
ROS burst regulates cytochrome P450 CYP6AB12 responsible for λ-cyhalothrin tolerance in
Spodoptera litura. *Journal of Hazardous Materials*, *387*, 121698.
https://doi.org/10.1016/j.jhazmat.2019.121698

Lüdecke D 2021. sjPlot: Data Visualization for Statistics in Social Science. R package version
2.8. 9, https://CRAN.R-project.org/package=sjPlot.

Malik, K. (2012). Study on the combined insecticidal effect of pyrethroid, *Azadirachta indica*
and boric acid on the *Bacillus thuringiensis* efficacy in *Tribolium castaneum*. *African Journal*
of Microbiology Research, *6*(27). https://doi.org/10.5897/ajmr11.1030

Meinke, L. J., Souza, D., & Siegfried, B. D. (2021). The Use of Insecticides to Manage the
Western Corn Rootworm, *Diabrotica virgifera virgifera*, LeConte: History, Field-Evolved
Resistance, and Associated Mechanisms. *Insects*, *12*(2), 112.
https://doi.org/10.3390/insects12020112

Milutinović, B., Stolpe, C., Peuβ, R., Armitage, S. A. O., & Kurtz, J. (2013). The Red Flour
Beetle as a Model for Bacterial Oral Infections. *PLoS ONE*, *8*(5), e64638-11.
https://doi.org/10.1371/journal.pone.0064638

Minetti, C., Ingham, V. A., & Ranson, H. (2020). Effects of insecticide resistance and exposure
on *Plasmodium* development in *Anopheles* mosquitoes. *Current Opinion in Insect Science*,
39, 42–49. https://doi.org/10.1016/j.cois.2019.12.001

Misra, J. R., Horner, M. A., Lam, G., & Thummel, C. S. (2011). Transcriptional regulation of
xenobiotic detoxification in *Drosophila*. *Genes & Development*, *25*(17), 1796–1806.
https://doi.org/10.1101/gad.17280911

Moll, R. M., Romoser, W. S., Modrakowski, M. C., Moncayo, A. C., & Lerdthusnee, K. (2001).
Meconial Peritrophic Membranes and the Fate of Midgut Bacteria During Mosquito (Diptera:
Culicidae) Metamorphosis. *Journal of Medical Entomology*, *38*(1), 29–32.
https://doi.org/10.1603/0022-2585-38.1.29

Ndo, C., Kopya, E., Irving, H., & Wondji, C. (2019). Exploring the impact of glutathione S-
transferase (GST)-based metabolic resistance to insecticide on vector competence of
Anopheles funestus for *Plasmodium falciparum*. *Wellcome Open Research*, *4*, 52.
https://doi.org/10.12688/wellcomeopenres.15061.2

Oakeshott, J. G., Farnsworth, C. A., East, P. D., Scott, C., Han, Y., Wu, Y., & Russell, R. J.
(2013). How many genetic options for evolving insecticide resistance in heliothine and
spodopteran pests? *Pest Management Science*, *69*(8), 889–896.
https://doi.org/10.1002/ps.3542

Oppert, B., Morgan, T. D., & Kramer, K. J. (2011). Efficacy of *Bacillus thuringiensis* Cry3Aa
protoxin and protease inhibitors against coleopteran storage pests. *Pest Management Science*,
67(5), 568–573. https://doi.org/10.1002/ps.2099

Orr, J. A., Vinebrooke, R. D., Jackson, M. C., Kroeker, K. J., Kordas, R. L., Mantyka-Pringle,
C., … Piggott, J. J. (2020). Towards a unified study of multiple stressors: divisions and
common goals across research disciplines. *Proceedings of the Royal Society B*, *287*(1926),
20200421. https://doi.org/10.1098/rspb.2020.0421

Pang, R., Xing, K., Yuan, L., Liang, Z., Chen, M., Yue, X., … Zhang, W. (2021). Peroxiredoxin
alleviates the fitness costs of imidacloprid resistance in an insect pest of rice. *PLOS Biology*,
19(4), e3001190. https://doi.org/10.1371/journal.pbio.3001190

Pauchet, Y., Bretschneider, A., Augustin, S., & Heckel, D. (2016). A P-Glycoprotein Is Linked
to Resistance to the *Bacillus thuringiensis* Cry3Aa Toxin in a Leaf Beetle. *Toxins*, *8*(12),
362–12. https://doi.org/10.3390/toxins8120362

Pavlidi, N., Vontas, J., & Leeuwen, T. V. (2018). The role of glutathione S-transferases (GSTs)
in insecticide resistance in crop pests and disease vectors. *Current Opinion in Insect Science*,
27, 97–102. https://doi.org/10.1016/j.cois.2018.04.007

Pinos, D., Andrés-Garrido, A., Ferré, J., & Hernández-Martínez, P. (2021). Response
Mechanisms of Invertebrates to *Bacillus thuringiensis* and Its Pesticidal Proteins.
Microbiology and Molecular Biology Reviews, *85*(1), e00007-20.
https://doi.org/10.1128/mmbr.00007-20

Rahman, M., Glatz, R., Roush, R., & Schmidt, O. (2011). Developmental Penalties Associated with Inducible Tolerance in *Helicoverpa armigera* to Insecticidal Toxins from *Bacillus thuringiensis*. *Applied and Environmental Microbiology*, *77*(4), 1443–1448. https://doi.org/10.1128/aem.01467-10

Rösner, J., Wellmeyer, B., & Merzendorfer, H. (2020). *Tribolium castaneum*: A Model for Investigating the Mode of Action of Insecticides and Mechanisms of Resistance. *Current Pharmaceutical Design*, *26*(29), 3554–3568. https://doi.org/10.2174/1381612826666200513113140

Sansinenea, E. (2012). *Bacillus thuringiensis Biotechnology* (E. Sansinenea, Ed.). Springer, Dordrecht. https://doi.org/10.1007/978-94-007-3021-2

Sappington, T. W., & Miller, N. J. (2017). Editorial overview: Pests and resistance: Shedding the albatross of resistance starts by embracing the ecological complexities of its evolution. *Current Opinion in Insect Science*, *21*, v–viii. https://doi.org/10.1016/j.cois.2017.07.003

Siegfried, B. D., & Scharf, M. E. (2001). *Mechanisms of Organophosphate Resistance in Insects*. In *Biochemical Sites of Insecticide Action and Resistance*: *Vol. 48* (pp. 269–291). Springer Berlin Heidelberg. https://doi.org/10.1007/978-3-642-59549-3_13

Siegwart, M., Graillot, B., Lopez, C. B., Besse, S., Bardin, M., Nicot, P. C., & Lopez-Ferber, M. (2015). Resistance to bio-insecticides or how to enhance their sustainability: a review. *Frontiers in Plant Science*, *6*, 381. https://doi.org/10.3389/fpls.2015.00381

Soderlund, D. M. (2011). Molecular mechanisms of pyrethroid insecticide neurotoxicity: recent advances. *Archives of Toxicology*, *86*(2), 165–181. https://doi.org/10.1007/s00204-011-0726-x

Sogorb, M. A., & Vilanova, E. (2002). Enzymes involved in the detoxification of organophosphorus, carbamate and pyrethroid insecticides through hydrolysis. *Toxicology Letters*, *128*(1–3), 215–228. https://doi.org/10.1016/s0378-4274(01)00543-4

Sprouffske, K., & Wagner, A. (2016). Growthcurver: an R package for obtaining interpretable metrics from microbial growth curves. *BMC Bioinformatics*, *17*(1), 172. https://doi.org/10.1186/s12859-016-1016-7

Tandonnet, S., Cardoso, G. A., Mariano-Martins, P., Monfardini, R. D., Cunha, V. A. S., Carvalho, R. A. de, & Torres, T. T. (2020). Molecular basis of resistance to organophosphate insecticides in the New World screw-worm fly. *Parasites & Vectors*, *13*, 562. https://doi.org/10.1186/s13071-020-04433-3

Tate, A. T., Andolfatto, P., Demuth, J. P., & Graham, A. L. (2017). The within-host dynamics of infection in trans-generationally primed flour beetles. *Molecular Ecology*, *26*(14), 3794–3807. https://doi.org/10.1111/mec.14088

Tate, A. T., & Graham, A. L. (2015). Trans-generational priming of resistance in wild flour
beetles reflects the primed phenotypes of laboratory populations and is inhibited by co-
infection with a common parasite. *Functional Ecology*, *29*(8), 1059–1069.
https://doi.org/10.1111/1365-2435.12411

Tate, A. T., & Graham, A. L. (2017). Dissecting the contributions of time and microbe density to
variation in immune gene expression. *Proceedings of the Royal Society B: Biological*
Sciences, *284*(1859), 20170727–20170728. https://doi.org/10.1098/rspb.2017.0727

Therneau, T. 2015. Package 'coxme'. Mixed Effects Cox Models. R package version, 2.24.
URL http://r-forge.r-project.org.

Tompkins, D. M., Dobson, A. P., Arneberg, P., Begon, M. E., Cattadori, I. M., Greenman, J. V.,
… Wilson, K. (2001). *Parasites and host population dynamics* (P. J. Hudson, Annapaola
Rizzoli, B T Grenfell, Hans Heesterbeek, & A. P. Dobson, Eds.). In (pp. 45–62).

Venables WN, Ripley BD. 2002. *Modern Applied Statistics with S*, Fourth edition. Springer,
New York. ISBN 0-387-95457-0, https://www.stats.ox.ac.uk/pub/MASS4/.

Vézilier, J., Nicot, A., Lorgeril, J., Gandon, S., & Rivero, A. (2013). The impact of insecticide
resistance on *Culex pipiens* immunity. *Evolutionary Applications*, *6*(3), 497–509.
https://doi.org/10.1111/eva.12037

Wang, X., Martínez, M.-A., Dai, M., Chen, D., Ares, I., Romero, A., … Yuan, Z. (2016).
Permethrin-induced oxidative stress and toxicity and metabolism. A review. *Environmental*
Research, *149*, 86–104. https://doi.org/10.1016/j.envres.2016.05.003

Wolz, M., Schrader, A., Whitelaw, E., & Müller, C. (2021). Gregarines modulate insect
responses to sublethal insecticide residues. *Oecologia*, 1–11. https://doi.org/10.1007/s00442-
021-05086-4

Wu, C., Chakrabarty, S., Jin, M., Liu, K., & Xiao, Y. (2019). Insect ATP-Binding Cassette
(ABC) Transporters: Roles in Xenobiotic Detoxification and Bt Insecticidal Activity.
International Journal of Molecular Sciences, *20*(11), 2829.
https://doi.org/10.3390/ijms20112829

Zhu, F., Lavine, L., O'Neal, S., Lavine, M., Foss, C., & Walsh, D. (2016). Insecticide Resistance
and Management Strategies in Urban Ecosystems. *Insects*, *7*(1), 2.
https://doi.org/10.3390/insects7010002

Zimmer, C. T., Panini, M., Singh, K. S., Randall, E. L., FIELD, L. M., Roditakis, E., … Bass, C.
(2017). Use of the synergist piperonyl butoxide can slow the development of alpha-
cypermethrin resistance in the whitefly *Bemisia tabaci*. *Insect Molecular Biology*, *26*(2), 152–
163. https://doi.org/10.1111/imb.12276

Figure Legends

Figure 1. Overall effects of OP and Pyr factors on survival and development. **A.** Main and

interaction effects of OP selection regime (Regime), OP treatment (PTx), and Bt treatment

(BtTx) on survival. **B.** Main and interaction effects of Pyr selection regime (Regime), Pyr

treatment (PTx), and Bt treatment (BtTx) on survival. **C.** Main and interaction effects of OP

selection regime, OP treatment, and Bt treatment on adult development. **D.** Main and interaction

effects of Pyr selection regime, Pyr treatment, and Bt treatment on adult development. Positive

significant effects (higher survival or faster development) are colored blue, negative significant

effects (lower survival or slower development) are red, and nonsignificant effects are black

(Suppl. Fig. 2; Suppl. Table 2).

Figure 2. The influence of pesticide exposure on route-specific infection outcomes. **A.** Adult Bt

load after larval oral exposure**.** Adults exposed to Bt but not pesticides as larvae have detectable

levels of Bt, but adults also exposed to pesticides as larvae do not; significant differences

between groups are denoted with an asterisk ($p < 0.0005$; Suppl. Table 6). **B.** Survival 24h post

septic Bt infection. Circles indicate surviving proportions of individuals per block. Significant

differences between groups are denoted with asterisks ($p < 0.0005$; Suppl. Table 7).

Figure 3. Number of significantly differentially expressed genes (padj < 0.05) for factors within

each DESeq2 model (number of DE transcripts are denoted above bars; positive counts are

upregulated transcripts and negative counts are downregulated). **A.** no pesticide model: main and

interaction effects of pesticide selection regime and Bt treatment in the absence of pesticide

exposure. Hatched bars indicate no pesticide treatment in the model. **B.** OP model: main and

interaction effects of OP pesticide exposure, selection regime, and Bt exposure. **C.** Pyr model:

main and interaction effects of Pyr pesticide exposure, selection regime, and Bt exposure.

OP Regime, Pyr Regime = main effects of OP and Pyr selection regimes, respectively; BtTx =

main effect of Bt treatment, OP Reg:BtTx, Pyr Reg:BtTx = interactions between OP and Pyr

regime, respectively, and Bt treatment; OP PTx, Pyr PTx = main effects of OP and Pyr exposure,

respectively; OP Reg:PTx, Pyr Reg:PTx = interactions between regime and exposure for OP and

Pyr, respectively; OP Reg:BtTx, Pyr Reg:BtTx = interactions between OP and Pyr regime,

respectively, and Bt treatment, OP PTx:BtTx, Pyr PTx:BtTx = interactions between OP and Pyr

exposure, respectively, and Bt exposure.

Figure 4. Profiles of differentially expressed transcripts demonstrate opposing responses in Bt-

infected individuals that were or were not also exposed to or selected with pesticides. Transcripts

significantly differentially expressed (padj < 0.05) with Bt treatment and interactions with **A.** OP

exposure, and **B.** Pyr selection regime and exposure. The DE Factor/ Model legend refers to the

factor corresponding to the differentially expressed gene set and the DESeq2 model used (no

pesticide- hatched bars; OP or Pyr- solid bars). Colored circles on the left indicate transcripts

included in the magenta WGCNA module. Transcript abbreviations are as follows: Stc = shuttle

craft-like protein, Def = defensin, Att = attacin, PR = pathogenesis-related protein, HRG =

histidine-rich glycoprotein, SP = serine protease, P450 = cytochrome P450, CEH = carboxylic

ester hydrolase, GST = glutathione S-transferase, ABC = ABC transporter, Chit = chitinase

protein, CP = cuticle-related protein, PM = peritrophic matrix related protein, Vit = vitellogenin,

Ins = insulin-like peptide, LSP = larval serum protein, AchE = acetylcholinesterase, KLF =

Krüppel-like factor, ApoIII = apolipophorin-III, Neur = neuropeptide.

Figure 5. Normalized gene counts illustrate dampening effects of Pyr selection regime and OP

treatment on immune gene expression induced with Bt expsoure. **A.** OP (OP model- OP PTx)

and Bt treatment (OP model- BtTx) effects are shown for TC007737 (attacin1), and **B.** Pyr

Regime (Pyr model- Pyr Reg) and Bt treatment (Pyr model- BtTx) effects are shown for

TC010517 (defensin-like).

Supplemental Table Legends

Suppl. Table 1. Read counts and alignment statistics for RNAseq libraries.

Suppl. Table 2. Significantly differentially expressed (DE) transcripts from the no pesticide

model (Conly), OP model (allOP), and Pyr model (allPyr) ($p < 0.05$). Fold change differences

(log2FC) and adjusted p-values (padj) are listed for each factor within each model (BtTx: main

effect of Bt; OPReg, PyrReg: main effect of OP and Pyr selection regime, respectively; OPPTx,

PyrPTx: main effect of OP and Pyr pesticide treatment; OPReg:PTx, PyrReg:PTx: interaction

effect between pesticide selection regime and exposure; OPReg:BtTx, PyrReg:BtTx: interaction

effect between selection regime and Bt expsoure; OPPTx:BtTx, PyrPTx:BtTx: interaction effect

between pesticide and Bt exposure. DE transcripts belonging to WGCNA modules are listed

(moduleColor).

Suppl. Table 3. Main and interaction effects of pesticide exposure, selection regime, and Bt

exposure on survival and adult development (coxme(Surv(Day, Status) ~ PTx+ Regime + BtTx

+ Reg:PTx + Reg:BtTx + PTx:BtTx +(1|Pop)); compared to control regime and treatments).

Suppl. Table 4. Effect of pesticide exposure on in vitro Bt growth. The effects of OP and Pyr

treatment were analyzed separately. The effect of treatment was analyzed for each; if replicate

was non-significant, all replicates were combined.

Suppl. Table 5. Effect of pesticide treatment on germinated spores from oral diet disks.

Suppl. Table 6. Effect of pesticides on chronic Bt infection in adults.

Suppl. Table 7. Survival after septic Bt infection.

Supplemental Figure Legends

Suppl. Figure 1. Experimental design schematic. Larvae from six distinct populations were

exposed to control, organophosphate (OP), or pyrethroid (Pyr) diets for more than eight

generations. Adults from Gen. 8 pesticide-selected and control populations were allowed to

reproduce on control diets, and twelve-day old F2 larvae were exposed to control and pesticide

diets with and without *Btt*.

Suppl. Fig. 2. Average proportion for survival and development for OP and Pyr separately.

Average proportion of surviving individuals with **A.** OP selection and dual exposure to OP and

Btt, and **B.** Pyr selection and dual exposure to with and *Btt*. Average proportion of developed

adults with **C.** OP selection and dual exposure to OP and *Btt*, and **D.** Pyr selection and dual

exposure to Pyr and *Btt*.

Suppl. Figure 3. *In* vitro growth rate (r) of *Btt* under control and pesticide conditions. **A.** *In vitro*

growth rate (r) of *Btt* under control and OP conditions. Data were not normally distributed, so

non-parametric tests were used. There was a significant difference between replicates, so the

effect of treatment was analyzed for each replicate separately (Suppl. Table 4). K-W test

revealed significant positive effect of OP on *Btt* growth rate for three of the four replicates (reps

2-4; Suppl. Table 4). **B.** *In vitro* growth rate (r) of *Btt* under control and Pyr conditions. Data

were not normally distributed, so non-parametric tests were used. There was no significant

difference between replicates or with the effect of treatment (Suppl. Table 4).

Suppl. Fig. 4. Germination of spores from diet disks in the absence or presence of either

pesticide (Bt, OP + *Btt* or Pyr + *Btt*: n=10, control: n=6). There was no significant difference

between Bt control and Bt pesticide diet disks (Suppl. Table 5).

Suppl. Figure 5. Top significantly enriched GO terms for differentially expressed gene sets. **A.**

Biological process GO terms, **B.** Molecular function GO terms. Differentially expressed gene

sets are denoted on the x-axis and refer to the DESeq2 model (no P = no pesticide model, OP =

OP model, Pyr = Pyr model) and the factor corresponding to the differentially expressed gene set

(BtTx = main effect of Bt treatment; OP Reg, Pyr Reg = main effect of OP and Pyr selection

regimes, repsectively; OP PTx, Pyr PTx = main effect of OP and Pyr exposure, respectively;

OPReg:PTx, PyrReg:PTx = interaction between OP and Pyr regime and exposure, respectively;

PyrReg:BtTx = interaction between Pyr regime and Bt treatment; OPTx:BtTx = interaction

between OP exposure and Bt treatment).

Suppl. Fig. 6. Canonical detoxification genes significantly differentially expressed (padj < 0.05)

with **A.** OP exposure (OP PTx), OP selection regime (OP Reg), Bt exposure (BtTx), and the

interaction between OP regime and exposure (OP Reg:PTx) and **B.** Pyr exposure (Pyr PTx), Pyr

selection regime (Pyr Reg), Bt exposure (BtTx), and the interaction between Pyr regime and

exposure (Pyr Reg:PTx).

Suppl. Fig. 7. Cuticle-associated genes significantly differentially expressed (padj < 0.05) with

A. OP exposure (OP PTx), OP selection regime (OP Reg), Bt exposure (BtTx), the interaction

between OP regime and exposure (OP Reg:PTx), and the interaction between OP and Bt

exposure (OP PTx:BtTx), and **B.** Pyr exposure (Pyr PTx), Pyr selection regime (Pyr Reg), Bt

exposure (BtTx), the interaction between Pyr regime and exposure (Pyr Reg:PTx), the interaction

between Pyr regime and Bt exposure (Pyr Reg:BtTx), and the interaction between Pyr and Bt

exposure (Pyr PTx:BtTx).

Suppl. Fig. 8. WGCNA gene modules and significant relationships with experimental factors.

The total number of transcripts (n) and number of differentially expressed (DE) transcripts in

1004 each module are listed, and significant relationships with experimental factors are indicated with

1005 black stars.

Figures

1008   **Figure 1.**

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1010 **Figure 2.**

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1012   **Figure 3.**

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1014 **Figure 4.**

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1016   **Figure 5.**

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1018 **Suppl. Figure 1.**

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1020 **Suppl. Fig. 2.**

A. Btt growth rate in response to OP

B. Btt growth rate in response to Pyr

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1022 **Suppl. Figure 3.**

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1024 **Suppl. Fig. 4.**

A.

B.

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1026   **Suppl. Figure 5.**

A.

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1029 **Suppl. Fig. 6.**

A.

OP− Cuticle-associated Genes

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Suppl. Fig. 7.

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1034 **Suppl. Fig. 8.**