Rethinking the role of intraspecific variability in species coexistence

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Abstract

Intraspecific variability (IV) has been proposed as a new track to explain species coexistence. Previous studies generally assumed, although often non-explicitely, that IV results from intrinsic differences between conspecific individuals. Under this hypothesis, IV either (i) blurred species differences, thus promoting transient or unstable coexistence, (ii) disproportionately advantaged the strongest competitor, thus hindering coexistence, or (iii) promoted coexistence. Based on a body of evidence, we here emphasize that IV does not necessarily imply inherent differences among conspecific individuals, nor species niches overlap. Conspecific individuals can also differ in their measured attributes due to differences in the micro-environment they thrive in. If this source of variation is predominant, which we suggest it is in many cases given the variation of the environment in numerous dimensions at fine scales and the observed spatial structure of IV, it implies that conspecific individuals respond more similarly to environmental variation than heterospecific individuals, thereby concentrating competition within species -- a necessary condition for species coexistence. Importantly, as the number of environmental dimensions that are typically observed and characterized in the field is generally much lower than the actual number of environmental dimensions that influence individual attributes, a great part of observed IV can often be misinterpreted as random variation across individuals while being environmentally-driven, with radically different consequences for community dynamics. We call for new studies exploring observed IV as an outcome of species-specific responses to high-dimensional environmental variations that can lead to inversions of species hierarchy in space and time promoting stable coexistence.

46 **Keywords:** Species coexistence; intraspecific variability; individual variation; high-dimensional

47 niche; spatial heterogeneity; spatial autocorrelation; environmental variation; competition

Introduction

Hyperdiverse ecological communities characterized by numerous coexisting species are common, such as in grasslands, tropical forests, or coral reefs. For example, a hectare of tropical forest can harbor more than 900 plant species of a diversity of forms and functions (Wilson et al. 2012). Understanding how so many species stably coexist while competing for the same basic resources, *viz.* light, water, and nutrients (Baraloto et al. 2010a), is one of the longest-standing and most fundamental questions in ecology (Gause 1934, Hutchinson 1961, Levine et al. 2017). Although different mechanisms have been suggested to contribute to species coexistence (Chesson 2000, Wright 2002), such as the Janzen-Connell effect (Janzen 1970, Connell 1971) or the niche trade-offs (Levine and HilleRisLambers 2009), it is unclear if and how they can explain the high species diversity observed in the field (Clark 2010).

The coexistence conundrum has typically been addressed following the assumption that all conspecific individuals are identical. However, intraspecific variability (IV) in traits, demographic rates or any proxy of performance (such as growth, fecundity, etc.), henceforth denoted as "attributes", has been seen as a potential new track to shed light on this old puzzle (Bolnick et al. 2011). Indeed, large IV can be observed within ecological communities (Albert et al. 2012, Violle et al. 2012). For instance, Siefert et al. (2015) estimated that IV accounted for 25% of the variability in functional traits within plant communities on average, and this proportion was even estimated at 44% in a hyperdiverse tropical forest (Poorter et al. 2018). Likewise, IV in growth rates for trees of standardized size and growing conditions has been found to account for up to 58% of total growth variability in a tropical forest stand (Le Bec et al. 2015).

Modeling studies that have explored the effect of IV on species coexistence have so far yielded

contrasting results (Stump et al. 2021). In most studies, the variability in a given attribute among

conspecific individuals has been included through independent random draws from a stochastic

variable (e.g. Lichstein et al. 2007, Hart et al. 2016, Barabás and D'Andrea 2016, Crawford et al.

2019 but see Purves and Vanderwel 2014, Banitz 2019). Similarly, empirical studies typically

summarize IV as a variance around species mean attributes (e.g. Jung et al. 2010, Albert et al.

2010, Siefert et al. 2015, Poorter et al. 2018). With this representation, IV can increase species

niche overlap and blur species differences, sometimes slowing down competitive exclusion in

models of community dynamics (Vieilledent et al. 2010, Crawford et al. 2019). However, in

some other models, non-linear responses can make such IV beneficial to the superior competitors

(i.e. the most competitive individuals of the more competitive species), thus accelerating compet-

itive exclusion (e.g. Courbaud et al. 2012, Hart et al. 2016). Alternatively, in particular spatial

configurations, more precisely when IV is greater in species preferred habitats, it has been shown

to foster species coexistence (Uriarte and Menge 2018). Stump et al. (2021) have attempted to

reconcile these contrasting results by distinguishing the effect of IV on niche traits (which con-

trol individual performance response to environmental conditions) vs. hierarchical traits (which

control individual performance independently from environmental conditions). They demon-

strated with different simulation models of community dynamics that IV in traits can alter stabi-

lizing mechanisms and fitness differences in a complex way which depends upon the nature of

the traits (niche vs. hierarchical) and their response curve, and thus promote or not species coex-

istence. In all the above examples however, the way IV is represented in models, as a random

variation around the mean species trait value, implies that conspecific individuals can respond

differently to the same environmental conditions and can thus suggests that IV is the result of in-

trinsic (e.g. genetic) differences between conspecific individuals that can widen the fundamental

species niche.

In this paper, we take another line of reasoning to show that IV doesn't require intrinsic individ-

ual differences and can rather emerge from species responses to a high-dimensional environment

(Fig. 1).

After presenting a body of evidence from theoretical models, experimental data, and hyperdi-

verse tropical forest inventory data, we discuss how these evidences combine in order to examine

three hypotheses (Fig. 2): (i) the large IV observed in natural communities can emerge from het-

erogeneity in multiple unobserved environmental dimensions; (ii) because environmental varia-

tion is structured in space and time, IV is likely to be similarly structured as well, suggesting that

it it not appropriate to represent IV as a purely random noise in models; and (iii) since a large ob-

served IV does not necessarily imply that conspecific individuals substantially differ in their fun-

damental niche, conspecific individuals may still respond more similarly to environment than

heterospecific individuals. We therefore call for a reconsideration of IV as an outcome of species

responses to high-dimensional environmental variations, rather than only of intrinsic differences

among individuals, which could shed new light on the coexistence conundrum. While we ac-

knowledge the existence of genetically-based individual variations, and that plasticity often has a

genetic basis (Nicotra et al. 2010), we suggest that a substantial part of measured IV might just

be the signature of a higher dimensionality of the species niche than typically observed. Species

differences along these many dimensions can lead to multiple local inversions of species hierar-

chy in an environment varying in space and time, thereby allowing the stable coexistence of nu-

merous species.

A body of evidence on the nature of IV and its consequences

119 **on individual and species differences**

Theoretical evidence: a large intraspecific variability emerges from

unobserved environmental dimensions.

We first conducted a virtual experiment that aimed at illustrating how an observed intraspecific

variability, or "individual effects", can result from variation in unobserved environmental vari-

ables (as suggested by Clark et al. 2007). More specifically, we generated virtual data of an indi-

vidual attribute (here tree growth) depending on a certain number of environmental variables

varying in space, and then analyzed the virtual data assuming that most of the environmental

variables are actually unobserved, as it is typically the case in the field.

Methods

A "perfect knowledge" simulation model

We considered a set of J species with I individuals each, distributed in a virtual landscape. The

environment was assumed to be fully known and defined by N environmental variables, X_1 to X_N,

that were each randomly and independently generated in the landscape, assuming spatial auto-

correlation. Individual location was drawn randomly in a virtual landscape defined by a $C \times C$

square grid, each cell corresponding to a particular environment (Fig. 3a). Individuals were iden-

tical within species (same model parameters for all conspecific individuals), but different be-

tween species (different model parameters between heterospecific individuals).

We considered the following "perfect knowledge" mathematical model, which depicts the exact

attribute Y_{ijt} (e.g., growth) of an individual i of species j given its environment at time t (Eq. I,

Appendix 1).

$ln(Y_{ijt}) = \beta_{0,j} + \beta_{1,j} \, ln(X_{1,ijt}) + \beta_{2,j} \, X_{2,ijt} + \dots + \beta_{N,j} \, X_{N,ijt}$ **(Eq. I)**

In this model, $\beta_j = [\beta_{0,j}, \dots, \beta_{N,j}]$ is the vector of parameters defining the response of species j to

the environment. Because conspecific individuals respond similarly to environmental variables,

variation in Y_{ijt} among them is only due to differences in the environment where and when each

individual is growing. Using this model, we computed the attribute Y of the $I \times J$ individuals at T

dates, assuming that values for some of the environmental variables changed between dates, and

thus obtained a virtual dataset $\{Y_{ijt}, X_{1,ijt}, \dots, X_{N,ijt}\}$ simulated with $N=10$, $I = 300$, $J = 2$, $C = 500$

and $T = 2$.

A "partial knowledge" statistical model

Second, we considered a "partial knowledge" statistical model for which we assumed that only

one explanatory variable X_1 (e.g., light) in the above virtual dataset have been measured along

with the environmental drivers that determine response variable Y (Eq. II, Appendix 1). This

model represents the ecologist's partial understanding of attribute Y. The model includes a

species fixed effects on the intercept and on the slope ($\beta'_{0,j}$ and $\beta'_{1,j}$) and a random individual ef-

fect $b_{0,i}$ on the intercept, $b_{0,i} \sim N(0, V_{bj})$, where V_{bj} is the intraspecific variance for species j. We

estimated the model parameters based on the virtual dataset introduced above but considering

only the first explanatory variable $\{Y_{ijt}, X_{1,ijt}\}$, the remaining "unknown" environmental effects

being contained in the model residuals, ε_{ijt}.

$ln(Y_{ijt}) = [\beta'_{0,j} + b_{0,i}] + \beta'_{1,j} \, ln(X_{1,ijt}) + \varepsilon_{ijt}, \, \varepsilon_{ijt} \sim N(0, V_j)$ **(Eq. II)**

Results

Despite the fact that conspecific individuals were identical and species responses to environment

were different, the variance estimates $\hat{V_{bj}}$ for individual random effects of species j were large,

and species responses to the environment overlapped (Fig. 3b). This is due to the contribution of

the unmeasured variables $\{X_{2,ijt}, \ldots, X_{N,ijt}\}$ in determining the variation of Y across individuals.

Since it is driven by spatially autocorrelated variables (Eq. I), the response Y was spatially auto-

correlated across conspecific individuals (Fig. 4). This means that two neighboring conspecific

individuals have more similar attribute Y than two distant conspecific individuals. Additionally,

the variance of Y was lower within than between species: conspecific individuals responded

more similarly to the environment than heterospecific individuals did (Fig. 4).

With this virtual experiment, we thus evidenced that: (i) a high IV can emerge from unobserved

environmental dimensions exclusively, (ii) the spatial structure of IV follows the spatial structure

of the underlying environmental variables, and (iii) IV does not blur differences between species

(Fig. 4) despite apparent niche overlap (Fig 3b).

Experimental evidence: large observed intraspecific variability in a

clonal tree plantation.

We then moved from a theoretical to an experimental approach using census data from clonal

Eucalyptus plantations, where genetic variability among individuals growing within a single

same site is controlled. We explore the partitioning of IV between intrinsic (genotypes) and ex-

trinsic sources, which is often infeasible in natural settings, to evidence that substantial observed

IV can indeed emerge from genetically identical individuals in the field, even when striving in a

similar apparent environment.

Methods

Data

The EUCFLUX experiment (São Paulo state, Brazil) is a clonal trial with a replicated, statisti-

cally-sound design (le Maire et al. 2019). It includes 14 genotypes of 5 different *Eucalyptus*

species or hybrids of various origins. Each genotype is planted in plots of 100 trees, at a density

of 1666 trees per hectare, and replicated spatially in 10 blocks (Fig. 5). The experimental set-up

was designed to minimize the variation in environmental factors among blocks, which were sep-

arated by less than 1.5 km within a homogeneous 200-ha stand showing small variation in soil

properties. Tree diameter at breast height (*D*) has been measured over 5 complete censuses, span-

ning 6 years, age at which such plantation is generally harvested (see le Maire et al. 2019 and

Appendix 2 for further details on this experimental set-up).

Analyses

We computed annual diameter growth (*G*) in mm/yr^{-1} for each tree as well as a competition index

(*C*) as the sum of the basal area of the eight direct neighbors of each tree. The dataset included

64,125 growth estimates corresponding to 13,531 trees in total. To quantify the relative impor-

tance of the different sources of growth variability, we used a statistical hierarchical growth

model (Eq. III) including an intercept (β_0), fixed effects of the log-transformed diameter (β_1) and

competition index (β_2), and random effects on the intercept of the block ($b_{0,b}$, with $b_{0,b} \sim N(0,$

$V_b)$), the genotype ($b_{0,g}$, with $b_{0,g} \sim N(0, V_g)$), the census date ($b_{0,t}$, with $b_{0,t} \sim N(0, V_t)$), and the in-

dividual ($b_{0,i}$, with $b_{0,i} \sim N(0, V_i)$). All the data were log-transformed and scaled, and a constant of

1 mm was added to all growth values to avoid undefined logarithms.

$ln(G+1)_{i,t} = [\beta_0 + b_{0,b} + b_{0,g} + b_{0,t} + b_{0,i}] + \beta_1 \, ln(D)_{i,t} + \beta_2 \, ln(C)_{i,t} + \varepsilon_{i,t}, \, \varepsilon_{i,t} \sim N(0, V)$ (**Eq. III**)

We used conjugated priors with inverse-gamma distributions (with shape and scale

parameters=10^{-3}) for variance parameters, and normal distributions (with mean=0 and

variance=1) for mean parameters. The estimation of model parameters was done using a

Bayesian approach using Stan software with the brms R package (Bürkner 2017, 2018). We

made 10,000 iterations for each MCMC with a burn-in period of 5,000 steps and a thinning rate

of one fifth. We obtained 1,000 estimations per parameter and examined the trace plots to check

convergence of the MCMC chains.

We then examined the proportion of the model residual variance (variation of the response vari-

able that is not explained by the covariates) related to each random effect in order to partition the

block, genotype, date and individual variances.

230 Results

While minor variability was associated with blocks (Table 1), confirming that they are broadly

homogeneous by design, the variability associated with extrinsic temporal factors was

predominant (Table 1). It reveals that the competition index (C) used in the analysis to

encapsulate the effect of progressive canopy closure does not fully encompass all temporal

effects.

Importantly, the variability between individuals was almost twice as high as the variability be-

tween genotypes (Table 1), which demonstrates a strong non-genetic origin of IV in tree growth.

Hence, even in this extremely conservative case, where environmental variation in space is mini-

mized and genotypic variability controlled, extrinsic IV is highly prominent over intrinsic IV.

241 *Empirical evidence: high intraspecific variability is spatially structured*

242 *and does not blur species differences in tropical forests.*

To test some of our hypotheses in natural hyperdiverse communities, we used data from three

long-term tree inventories in tropical forests, from Amazonia (Paracou, French Guiana; Gourlet-

Fleury et al. 2004), Central America (Barro Colorado Island, Panama; Losos and Leigh 2004)

and South-East Asia (Uppangala, India; Pélissier et al. 2011). More specifically, we inferred ob-

served IV, we tested if individual growth showed local spatial autocorrelation, and if conspecific

individual growth was more similar than heteropsecific individual growth locally. These three

sites encompass contrasting climatic conditions (rainfall ranging from 2,600 in BCI to 5,100

mm.y^{-1} in Uppangala), disturbance regimes (incl. various logging experiments in Paracou) and

topography (from gentle in BCI to mountainous in Uppangala), making them representative of

the global tropical forests. The data from these tropical forest inventories that we used in this pa-

per are summarized in Table 2.

Methods

For all three datasets, annualized growth between two censuses was computed as the difference

of DBH between two consecutive censuses, divided by the time period between those two cen-

suses. Growth data < -2 or > 100 mm.y^{-1} as well as individuals from incompletely identified

species were discarded prior to analysis. We also discarded individuals and species with a single

observation to avoid identifiability issues in the fitted models.

For all three datasets, annualized growth between two censuses was computed as the difference

of DBH between two consecutive censuses, divided by the time period between those two cen-

suses. Estimates with a growth < -2 mm/year or > 100 mm/year were removed from the dataset.

Unless specified, individuals from undetermined species were not taken into account in the

analyses. Mean annual growth for each individual tree was then computed as the difference of

DBH between the first and the last time a tree was measured, divided by the time period between

those two measurements.

Estimation of intraspecific variability

To quantify the relative importance of intra- vs. and inter-specific variability in each site, we

used a hierarchical growth model (Eq. IV), including an intercept β_0, a diameter (*D*) fixed effect

β_1, a species random effect b_{0j} (with $b_{0j} \sim N(0, V_b)$) and an individual random effect d_{0i} (with $d_{0i} \sim$

$N(0, V_d)$) on the intercept. All data were log-transformed and scaled, and a constant of 2 mm was

added to all growth values to avoid undefined logarithms.

$ln(G_{ijt}+2) = [\beta_0 + b_{0j} + d_{0i}] + \beta_1 \times ln(D_{ijt}) + \varepsilon_{ijt}, \varepsilon_{ijt} \sim N(0,V)$ **(Eq. IV)**

For Paracou, which has a very large dataset, we sampled 100,000 growth values randomly to per-

form inference. No sampling was done for Uppangala and BCI. We used conjugated priors with

inverse-gamma distributions (with shape and scale parameters=10^{-3}) for variance parameters, and

normal distributions (with mean=0 and variance=1) for mean parameters. We estimated the inter-

and intra-specific growth variability from the variance of the species (V_b) and individual (V_d) ran-

dom effects, respectively. Model parameters were estimated the same Bayesian approach as for

the analysis of the *Eucalyptus* dataset.

Test of the spatial autocorrelation of individual growth

To test whether individual growth was spatially autocorrelated, we performed in each site, spatial

analyses of the mean individual growth values within a 100-m neighborhood distance. We pre-

ferred a conservative approach based on mean individual growth without accounting for the ef-

fect of diameter, thus ignoring ontogenetic differences.

We performed Moran's I one-tailed tests as implemented in the *ape* R package (Paradis and

Schliep 2019), for pairs of conspecifics less than 100 m apart in the same plot. For the most

abundant species, we sampled 3,000 individuals with a uniform probability. We considered only

the species with more than five individuals.

Comparison of between-conspecifics and between-heterospecifics similarity in growth

To test if the performance of conspecific individuals was locally more similar than the perfor-

mance of heterospecific individuals in the three sites, we also used mean individual growth, thus

ignoring ontogenetic differences.

We computed the mean individual growth semivariance (Baraloto and Couteron 2010) consider-

ing either conspecific or heterospecific neighbors within a 100-m radius. In the first case, semi-

variance was estimated as the mean of the squared difference in individual mean growth for all

pairs of conspecific individuals. In the second case, semivariance was estimated as the mean of

the squared difference in individual mean growth for all pairs of individuals with an individual of

the focal species and one of another species. We considered only the species with more than five

individuals, and with more than five heterospecific neighbors within the 100-m neighbohood dis-

tance. For each species, we then compared the semivariances between conspecific and heterospe-

cific individuals using a Mann-Whitney test with a 0.05 alpha-risk.

Results

For the three sites, IV estimated from the growth model (V_d, ranging from 0.41 to 0.66) was of the same order of magnitude as the interspecific variance (V_b, ranging from 0.36 to 0.66) (Table 3). Overall, a large share of the variability in tree growth comes from individual effects in the three sites, even after accounting for the effect of diameter on tree growth, showing a high intraspecific variability in growth in these tropical forests.

Positive spatial autocorrelation in tree growth between conspecifics was significant for 19 to 31% of the species in the three sites, representing between 45 and 79% .of the total number of individuals (Table 4). Spatial autocorrelation was however much higher in logged plots as compared to unlogged in Paracou, because of a more heterogeneous light environment resulting from logging history (Appendix 3).

The mean individual growth semivariance appeared significantly higher among heterospecifics than among conspecifics for 42 to 61% of the species in the three sites, representing 58 to 89% of the total number of individuals (Table 5). To control for a potential effect of species abundance on the semivariance estimations, we replicated the analysis by sampling a maximum of ten individuals per species. The results were qualitatively unchanged (Appendix 3).

Discussion

High-dimensional environmental variation leads to large observed intraspecific variability

IV can result from intrinsic differences among individuals (intrinsic IV) or from exogenous environmental variation, including biotic factors (extrinsic IV) (Violle et al. 2012, Moran et al. 2016). While much emphasis has been placed on the former in studies on coexistence, and although we acknowledge its ecological and evolutionary importance, we here argue that the latter is ubiquitous in natural communities and has radically different consequences for species differences and community assembly. More specifically, we argue that large observed IV can be indicative of high-dimensional environmental variation in space and time.

First, using a simple theoretical model, we showed how environmental variation in unobserved dimensions of the environment can produce large observed IV, although conspecific individuals are clones (Fig. 3). Similarly, the variance partitioning of individual tree growth within a common garden of *Eucalyptus* clones (le Maire et al. 2019) shows that the variance in growth between individuals is about twice as high as the variance between genotypes (Table 1). This reveals that a large part of the observed IV can be extrinsic and emerge from exogenous environmental variables alone.

Importantly, because IV can emerge from environmental heterogeneity without underlying genetic differences, observed IV does not necessarily imply that conspecific individuals substan-

tially differ in their response to the environment, nor that species niches overlap (Fig. 1). Instead,

large observed IV can exclusively reflect the projection of species' high-dimensional niches

within a high-dimensional environment that is variable in time and space: conspecific individuals

differ with each other because they each thrive in a different micro-environment. In empirically

observed data, such IV is therefore the result of projecting a high-dimensional response (e.g.,

physiological processes), which is controlled by multiple macro- and micro-environmental vari-

ables, down to a low-dimensional, integrative response (e.g., annual growth) that is poorly char-

acterized because of an incomplete view of the environmental variables that contribute to it. This

reassigns an important part of observed variation among individuals, often perceived as neutral

or random since they are seemingly unrelated to the observed dimensions of the environment

(Fig. 1, Fig. 3, Table 1), to the classical niche theory (Hutchinson 1957) and is in agreement with

natural history observations of individual trait differences that are associated to species-specific

ecological strategies (Box 1).

The "biodiversity paradox" highlights that a large number of species can coexist while compet-

ing for a limited number of resources (Hutchinson 1961). This puzzling question has generally

been tackled considering trade-offs along a limited number of niche axes, often corresponding to

resources (Tilman 1982, Rees 2001). But if the number of resources may indeed be relatively

limited (e.g., light, water, and nutrients for plants), the number of independent environmental

factors (e.g. microclimatic variables) that drive the performance of individuals for a particular

level of resources is not. Environments are known to vary along multiple dimensions at fine

scales in space and time (Fig. 6, Box 1), and in many cases, this variation has been shown to in-

fluence individual attributes (e.g., Fortunel et al. 2020).

Nevertheless, despite technological advances (Box 1), many of these abiotic and biotic environ-

mental factors are still poorly understood and monitored. As a result, the dimensionality of field

observations is typically low compared with the high dimensionality of the environment (Bramer

et al. 2018, Estes et al. 2018). The variability in individual attributes due to the variation of unob-

served environmental variables therefore remains mostly a black box, and is typically summa-

rized in terms of residual variance in statistical models (Albert et al. 2012, Siefert et al. 2015) or

encapsulated into so-called "individual random effects" (Clark et al. 2007). We here emphasize

that even in the absence of any intrinsic differences among conspecific individuals, a large IV

can emerge from the imperfect characterization of the environment (Fig. 1, Fig. 3, Table 1),

which varies in a high number of dimensions (Fig. 6, Box 1).

Intraspecific variability is structured in space and time

IV has commonly been perceived and modeled as a probability distribution including a variance

around the species mean (Lichstein et al. 2007, Courbaud et al. 2012, Hart et al. 2016, Barabás

and D'Andrea 2016, Uriarte and Menge 2018). While this representation typically results from a

lack of knowledge, with randomness being used as a substitute for more detailed understanding

of underlying ecological processes (Clark et al. 2007), it encapsulates strong hypotheses relating

to the nature of IV that are rarely discussed. In contrast, we argue here that IV is generally non-

random and structured in both space and time.

At a given time, conspecific individuals that are distributed across space can strongly vary in

their attributes (Violle et al. 2012, Siefert et al. 2015, Moran et al. 2016, Poorter et al. 2018).

While this spatial IV has often been interpreted as random, i.e. implying that conspecific individ-

uals can perform differently within the same environment (Fig. 1b), a large part of this variability

appeared in fact structured in space and likely associated with spatial changes in the environment

(Fig. 1c, Moran et al. 2016). In our simulated example, the attribute of conspecific individuals

varies spatially as a result of the environmental variation in space, and the spatial autocorrelation

of conspecific attributes reflects the spatial autocorrelation of the environmental variables (Fig.

3, Fig. 4).

Similarly, data from three long-term forest inventory plot across the tropics revealed spatial auto-

correlation in tree diameter growth of conspecific individuals (Table 3), suggesting that IV is

strongly driven by the spatial variation of the environment, which is itself highly structured (Box

1, Fig. 6). However, we acknowledge that intrinsic IV can also be spatially structured, for in-

stance via dispersal patterns or natural selection (Moran et al. 2016). Nonetheless, if observed IV

were mainly due to dispersal, it would likely be randomly structured in space (Getzin et al. 2014)

or correlated at the spatial scale of seed dispersal, typically several tens of meters in tropical

forests (Clark et al. 2004, Seidler and Plotkin 2006, Muller-Landau et al. 2008). Instead, environ-

mental variables are typically highly spatially correlated at fine scales (e.g. meter scale, Baraloto

& Couteron 2010, Box 1), and so is IV (as shown here), in accordance with our hypothesized ex-

planation. We also acknowledge that natural selection can happen at fine scales (Marrot et al.

2021), and could thus produce spatially structured IV due to local genetic adaptation. Notwith-

standing, data documenting genetic variation within species can still reveal higher similarity be-

tween conspecifics than heterospecifics locally as well as non-overlapping species niches

(Schmitt et al. 2021). Importantly, any local genetic adaptation does not preclude that multidi-

mensional environmental variations generate large observed IV that is structured in space and

time and whose consequences cannot be well represented and understood using a random varia-

tion around a species mean.

In communities of sessile organisms such as trees, IV has been commonly structured in space us-

ing individual random effects, which vary among conspecific individuals but stay constant

through the lifetime of individuals (Clark et al. 2007, 2010, Vieilledent et al. 2010). We here ar-

gue that while this approach can *reveal* the spatial structure of IV through inference, the use of

the resulting estimated standard deviation term to introduce individual variation in simulations of

community dynamics is not sufficient to *produce* a spatially structured IV, as we showed is ob-

served in natural communities

Similarly, individual attributes can change over time. Because individuals within a species can

be measured at different points in time, as it is often the case when assembling functional trait

databases for example (Zanne et al. 2009, Albert et al. 2011, Kattge et al. 2020) this can lead to

an observed unstructured IV when characterized by a variance around a species mean (Fig. 1b).

But a large part of this observed IV is actually structured in time and associated with temporal

changes in the environment. For instance, the temporal storage effect (Chesson and Warner

1981), a well-known coexistence mechanism, structures species performance because species are

able to "store" growth during favorable timespans to overcome lean times; mast-seeding or mast-

ing, which describes periodic and synchronized massive seed production of conspecific individu-

als, would also result in a temporally structured IV (Koenig and Knops 2005). Temporal varia-

tion in individual response within a species can typically be structured with temporal random ef-

fects (Clark et al. 2007). Temporal random effects have been used to estimate the inter-annual

variability in tree growth (Metcalf et al. 2009, Fortunel et al. 2018) and fecundity (Clark et al.

2007) for example. In all those examples, temporal environmental variation affects conspecific

attributes in the same way (Clark 2010).

We therefore call for a reconsideration of the nature and the way of integrating IV into models of

community dynamics. When IV is modeled randomly with a variance around a species mean, it

implies that conspecific individuals can perform differently in the exact same environment, thus

implying intrinsic differences between conspecific individuals. This type of unstructured IV can

result in an increase in species niche overlap, which blurs species differences (Fig. 1a and 1b,

Stump et al. 2021). Moreover, the effect of unstructured IV on species coexistence vary greatly

between studies, depending on the way IV is included in models (Lichstein et al. 2007, Courbaud

et al. 2012, Hart et al. 2016, Uriarte and Menge 2018, Stump et al. 2021). Note that, in some

studies, the random variation in attributes across conspecific individuals is considered as envi-

ronmental, because it is not heritable in the model (e.g. Lichstein et al. 2007, Moran et al. 2016).

However, environmentally-driven IV should be structured in space and time, as the environment

is (Fig. 4, Fig. 6). Additionally, although trait heritability has rarely been considered in studies on

the role of IV on coexistence (but see Barabás and D'Andrea 2016), the structured IV that results

from the projection of species-specific responses to high-dimensional environmental variation is

heritable, in the sense that parents and offsprings will have the same attributes in the same envi-

ronment. In addition, when IV is randomly distributed among conspecific individuals, it system-

atically increases local similarity among heterospecific individuals while decreasing similarity

among conspecific individuals, which is not the case when IV is structured in space and time

(Purves and Vanderwel 2014, Banitz 2019), as discussed now.

Conspecific individuals respond more similarly than heterospecific individuals locally

Species differ in multiple attributes, responding to a high number of environmental variables

(Fig. 7, Box 1), but often in ways that cannot be readily observed. If observed IV results mainly

from high-dimensional environmental variation in space and time rather than from intrinsic dif-

ferences between conspecific individuals, then for a given environment, conspecific individuals

should respond more similarly than heterospecific individuals. This is the case in our simulated

example, where the fact that conspecific individuals have exactly the same set of parameters and

respond identically to spatial and temporal changes in the environment results in higher inter-

than intraspecific variance in the response locally (Fig. 3, Fig. 4).

Corroborating this point of view, pairs of conspecific individuals in 11 North-American temper-

ate forest stands showed higher correlation in their temporal variation of growth rate or fecundity

than pairs of heterospecific individuals on average (Clark 2010). This indicates that conspecific

individuals responded more similarly to environmental variation in time than individuals of dif-

ferent species. Importantly, these results were obtained in a system with high observed IV (lead-

ing to an apparent species niche overlap), where species responded in the same direction to envi-

ronmental changes (e.g. increased tree growth in climatically favorable years). Hence, consider-

ing the temporal structure of IV revealed species differences that were not apparent otherwise,

since they led to spreading along a high number of dimensions that varied at fine scales (Clark

2010). However, as well highlighted by Stump et al. (2021), these results were often misinter-

preted as an evidence that IV fostered coexistence. As another piece of evidence presented here,

pairs of spatially proximal conspecific individuals tended to present more similar temporal

means in absolute tree growth than pairs of close heterospecific individuals across three large

contrasted tropical forest sites (Table 5). This provides new empirical evidence that, although es-

timated IV can be substantial, conspecific individuals respond more similarly than heterospecific

individuals to environmental variation in space.

A stronger similarity in the response to environment between conspecific than heterospecific in-

dividuals locally leads to a stronger concentration of competition within species, which, ulti-

mately, can result in intraspecific competition being greater than interspecific competition, a

common driver of stable species coexistence (Lotka 1925, Volterra 1926, Chesson 2000). Be-

cause species differ in their response to the environment, environmental variation in space and

time leads to local or punctual inversions of species hierarchy in performance (Fig. 1d). As pos-

sibilities of hierarchy inversions between species increase rapidly with increasing dimensionality

of the environment (Fig. 1c), the high-dimensionality of the environment (Box 1) offers room for

the stable coexistence of numerous species (Hurtt and Pacala 1995, Falster et al. 2017, Rüger et

al. 2018). In the end, we therefore argue that a substantial part of IV is not in itself a mechanism

for coexistence but is rather the signature of species differences and environmental variation that

allow coexistence: the high-dimensional species differences, which make them respond differ-

ently in a high-dimensional environment varying in space and time, can only be observed at the

individual scale. In the absence of precise information on the many dimensions across which

species differ and environment varies, large observed IV is the evidence of the niche mecha-

nisms enabling species coexistence.

510 *Recommendations and concluding remarks*

Most of the theoretical studies that have explored the role of IV in species coexistence so far did

so by adding variances around species-specific means, thus considering IV as stochastic, which

implies that conspecific individuals perform differently in the same environment. Here, we pro-

vide a body of evidence suggesting that large observed IV can emerge from environmental het-

erogeneity and is structured in space and time. We stress that this interpretation has strong conse-

quences on the understanding of the effects of IV on species coexistence: (i) IV does not neces-

sarily imply that individuals are intrinsically different nor that species niches overlap, and (ii) the

spatial and temporal structure of observed IV reveals stronger concentration of competition

within species locally in space and time, which is a frequent necessary condition for stable

species coexistence. Finally, a large part of observed IV does not have a direct effect on species

coexistence and is just evidence of species-specific responses to high-dimensional environmental

variations in space and time. We thus call for a reconsideration of the nature of IV and of the

way it is integrated in models, by thoroughly distinguishing its sources (intrinsic vs. extrinsic).

We acknowledge the existence of intrinsic IV, potentially due to local adaptation to the microen-

vironment, and its eco-evolutionary importance, but still affirm that multidimensionnal environ-

mental variation generates a large observed IV that is structured in space and time, and whose

consequences cannot be well represented and understood using a random variation around a

species mean. To this end, we recommend that empirical studies explore further the spatio-tem-

poral structure of IV and how it relates to environmental variation along multiple dimensions,

and, when possible, assess the relative importance of genetically vs. environmentally driven IV,

for instance by means of common garden experiments. Models of community dynamics should

then endeavor to structure IV in space and time so that it reflects the high-dimensional variation

in both the environment and species attributes, and not only some intrinsic differences between

conspecific individuals (Purves and Vanderwel 2014, Moran et al. 2016, Banitz 2019). In both

empirical studies and models, this implies that the species attributes are measured at the individ-

ual level, localized in space, and repeatedly observed in time. Simultaneously, the monitoring of

multiple environmental variables at fine scales in space and time is required in order to better

capture their effect on individual attributes (such as physiological or mechanistic traits, Shipley

et al. 2016, Brodribb 2017), hence reducing the part of unexplained IV, and ultimately to better

characterize the high-dimensionality of species niches. Altogether, these recommendations will

enable to better account for species differences that are expressed at the individual level and evi-

dence their impacts on the community dynamics in natura and in silico.

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Supplementary information and data access

The appendices and all the code used for this study are available in a GitHub repository (https://github.com/camillegirardtercieux/coexIV) and have been permanently archived on Zenodo (https://doi.org/10.5281/zenodo.5504013).

Appendix 1: Virtual experiment with two species.

Appendix 2: Analysis of an *Eucalyptus* clonal plantation dataset.

Appendix 3: Analysis of tropical forest inventory data.

No new data were used in this study. For access to forest plot inventory data and *Eucalyptus* plantation data used in this study, refer to the data used in Le Bec et al. 2015, Hérault and Piponiot 2018, Condit et al. 2019 and le Maire et al. 2019. However, the analyses and reflections

566 presented here are original.

Statements of author roles:

569 CGT, IM and GV conceived the initial ideas and coordinated the INTRACO working group. All

570 authors contributed to the study design and ideas within the INTRACO working group. CGT led

571 the analyses. CGT, IM and GV wrote the first draft of the manuscript, and all authors contributed

572 substantially to revisions.

Declaration of Interests

575 All authors declare that they have no conflict of interest.

Literature cited

Albert, C. H., F. de Bello, I. Boulangeat, G. Pellet, S. Lavorel, and W. Thuiller. 2012. On the importance of intraspecific variability for the quantification of functional diversity. Oikos 121:116–126.

Albert, C. H., F. Grassein, F. M. Schurr, G. Vieilledent, and C. Violle. 2011. When and how should intraspecific variability be considered in trait-based plant ecology? Perspectives in Plant Ecology, Evolution and Systematics 13:217–225.

Albert, C. H., W. Thuiller, N. G. Yoccoz, A. Soudant, F. Boucher, P. Saccone, and S. Lavorel. 2010. Intraspecific functional variability: extent, structure and sources of variation. Journal of Ecology 98:604–613.

Banitz, T. 2019. Spatially structured intraspecific trait variation can foster biodiversity in disturbed, heterogeneous environments. Oikos 128:1478–1491.

Barabás, G., and R. D'Andrea. 2016. The effect of intraspecific variation and heritability on community pattern and robustness. Ecology Letters 19:977–986.

Baraloto, C., and P. Couteron. 2010. Fine-scale Microhabitat Heterogeneity in a French Guianan Forest: Tropical Forest Microhabitat Heterogeneity. Biotropica 42:420–428.

Baraloto, C., C. E. Timothy Paine, S. Patiño, D. Bonal, B. Hérault, and J. Chave. 2010a. Functional trait variation and sampling strategies in species-rich plant communities. Functional Ecology 24:208–216.

Baraloto, C., C. E. Timothy Paine, L. Poorter, J. Beauchene, D. Bonal, A.-M. Domenach, B. Hérault, S. Patiño, J.-C. Roggy, and J. Chave. 2010b. Decoupled leaf and stem economics in rain forest trees: Decoupled leaf and stem economics spectra. Ecology Letters 13:1338–1347.

Bolnick, D. I., P. Amarasekare, M. S. Araújo, R. Bürger, J. M. Levine, M. Novak, V. H. W.

602 Rudolf, S. J. Schreiber, M. C. Urban, and D. A. Vasseur. 2011. Why intraspecific trait

603 variation matters in community ecology. Trends in Ecology & Evolution 26:183–192.

604 Bramer, I., B. J. Anderson, J. Bennie, A. J. Bladon, P. De Frenne, D. Hemming, R. A. Hill, M. R.

605 Kearney, C. Körner, A. H. Korstjens, J. Lenoir, I. M. D. Maclean, C. D. Marsh, M. D.

606 Morecroft, R. Ohlemüller, H. D. Slater, A. J. Suggitt, F. Zellweger, and P. K. Gillingham.

607 2018. Advances in Monitoring and Modelling Climate at Ecologically Relevant Scales.

608 Pages 101–161 Advances in Ecological Research. Elsevier.

609 Brodribb, T. J. 2017. Progressing from 'functional' to mechanistic traits. New Phytologist

610 215:9–11.

611 Bürkner, P.-C. 2017. brms : An R Package for Bayesian Multilevel Models Using Stan. Journal

612 of Statistical Software 80.

613 Bürkner, P.-C. 2018. Advanced Bayesian Multilevel Modeling with the R Package brms. The R

614 Journal 10:395.

615 Chesson, P. 2000. Mechanisms of Maintenance of Species Diversity. Annual Review of Ecology

616 and Systematics 31:343–366.

617 Chesson, P. L., and R. R. Warner. 1981. Environmental Variability Promotes Coexistence in

618 Lottery Competitive Systems. The American Naturalist 117:923–943.

619 Clark, J. S. 2010. Individuals and the variation needed for high species diversity in forest trees.

620 Science 327:1129–1132.

621 Clark, J. S., D. Bell, C. Chu, B. Courbaud, M. Dietze, M. Hersh, J. Hillerislambers, I. Ibáñez, S.

622 Ladeau, S. McMahon, J. Metcalf, J. Mohan, E. Moran, L. Pangle, S. Pearson, C. Salk, Z.

623 Shen, D. Valle, and P. Wyckoff. 2010. High-dimensional coexistence based on individual

624 variation: A synthesis of evidence. Ecological Monographs 80:569–608.

625 Clark, J. S., M. Dietze, S. Chakraborty, P. K. Agarwal, I. Ibanez, S. LaDeau, and M. Wolosin.

626 2007. Resolving the biodiversity paradox. Ecology Letters 10:647–659.

627 Clark, J. S., S. LaDeau, and I. Ibanez. 2004. Fecundity of trees and the colonization-competition

628 hypothesis. Ecological Monographs 74:415–442.

Condit, R., R. Pérez, S. Aguilar, S. Lao, R. Foster, and S. Hubbell. 2019. Complete data from the

Barro Colorado 50-ha plot: 423617 trees, 35 years. Dryad.

Connell, J. H. 1971. On the role of natural enemies in preventing competitive exclusion in somf

marine animals and in rain forest trees. Dynamics of populations: Proceedings of the

advanced study institute on dynamics of numbers in populations:298–312.

Courbaud, B., G. Vieilledent, and G. Kunstler. 2012. Intra-specific variability and the

competition–colonisation trade-off: coexistence, abundance and stability patterns.

Theoretical Ecology 5:61–71.

Crawford, M., F. Jeltsch, F. May, V. Grimm, and U. E. Schlägel. 2019. Intraspecific trait

variation increases species diversity in a trait-based grassland model. Oikos 128:441–455.

Cushman, K. C., M. Detto, M. García, and H. C. Muller-Landau. 2022. Soils and topography

control natural disturbance rates and thereby forest structure in a lowland tropical

landscape. Ecology Letters:ele.13978.

Estes, L., P. R. Elsen, T. Treuer, L. Ahmed, K. Caylor, J. Chang, J. J. Choi, and E. C. Ellis. 2018.

The spatial and temporal domains of modern ecology. Nature Ecology & Evolution

2:819–826.

Falster, D. S., Å. Brännström, M. Westoby, and U. Dieckmann. 2017. Multitrait successional

forest dynamics enable diverse competitive coexistence. Proceedings of the National

Academy of Sciences 114:E2719–E2728.

Fortunel, C., J. R. Lasky, M. Uriarte, R. Valencia, S. J. Wright, N. C. Garwood, and N. J. B.

Kraft. 2018. Topography and neighborhood crowding can interact to shape species

growth and distribution in a diverse Amazonian forest. Ecology 99:2272–2283.

Fortunel, C., C. Stahl, P. Heuret, E. Nicolini, and C. Baraloto. 2020. Disentangling the effects of

environment and ontogeny on tree functional dimensions for congeneric species in

tropical forests. New Phytologist 226:385–395.

Gause, G. F. 1934. The Struggle for Existence.

Getzin, S., T. Wiegand, and S. P. Hubbell. 2014. Stochastically driven adult–recruit associations

of tree species on Barro Colorado Island. Proceedings of the Royal Society B: Biological

Sciences 281:20140922.

Gonzalez, S., V. Bilot-Guérin, P. Delprete, C. Geniez, J.-F. Molino, J.-L. Smock, F. Théveny,

IRD, CIRAD, INRAE, Université de Montpellier, Herbier de Guyane, Cayenne, Guyane

française, and CNRS. 2021. L'herbier IRD de Guyane.

Gourlet-Fleury, S., J.-M. Guehl, O. Laroussinie, and ECOFOR (Group), editors. 2004. Ecology

and management of a neotropical rainforest: lessons drawn from Paracou, a long-term

experimental research site in French Guiana. Elsevier, Paris.

Grisebach, A. 1872. Die Vegetation der Erde nach ihrer klimatischen Anordnung : ein Abriss der

vergleichenden Geographie der Pflanzen. Verlag von Wilhelm Engelmann, Leipzig.

Hallé, F., R. A. A. Oldeman, and P. B. Tomlinson. 1978. Tropical Trees and Forests. Springer

Berlin Heidelberg, Berlin, Heidelberg.

Hart, S. P., S. J. Schreiber, and J. M. Levine. 2016. How variation between individuals affects

species coexistence. Ecology letters 19:825–838.

Hérault, B., and C. Piponiot. 2018. Key drivers of ecosystem recovery after disturbance in a

neotropical forest: Long-term lessons from the Paracou experiment, French Guiana.

Forest Ecosystems 5:2.

Hurtt, G. C., and S. W. Pacala. 1995. The consequences of recruitment limitation: reconciling

chance, history and competitive differences between plants. Journal of Theoretical

Biology 176:1–12.

Hutchinson, G. E. 1957. Concluding Remarks. Cold Spring Harbor Symposia on Quantitative

Biology 22:415–427.

Hutchinson, G. E. 1959. Homage to Santa Rosalia or Why Are There So Many Kinds of

Animals? The American Naturalist 93:145–159.

Hutchinson, G. E. 1961. The Paradox of the Plankton. The American Naturalist 95:137–145.

Janzen, D. H. 1970. Herbivores and the Number of Tree Species in Tropical Forests. The

American Naturalist 104:501–528.

Jung, V., C. Violle, C. Mondy, L. Hoffmann, and S. Muller. 2010. Intraspecific variability and

trait-based community assembly. Journal of Ecology 98:1134–1140.

Kattge, J., G. Bönisch, S. Díaz, S. Lavorel, and I. Colin. 2020. TRY plant trait database –

enhanced coverage and open access. Global Change Biology 26:119–188.

Koenig, W. D., and J. M. H. Knops. 2005. The Mystery of Masting in Trees: Some trees

reproduce synchronously over large areas, with widespread ecological effects, but how

and why? American Scientist 93:340–347.

Kraft, N. J. B., O. Godoy, and J. M. Levine. 2015. Plant functional traits and the

multidimensional nature of species coexistence. Proceedings of the National Academy of

Sciences 112:797–802.

Kupers, S. J., C. Wirth, B. M. J. Engelbrecht, and N. Rüger. 2019. Dry season soil water

potential maps of a 50 hectare tropical forest plot on Barro Colorado Island, Panama.

Scientific Data 6:63.

Le Bec, J. L., B. Courbaud, G. L. Moguédec, and R. Pélissier. 2015. Characterizing tropical tree

species growth strategies: Learning from inter-individual variability and scale invariance.

PLoS ONE 10.

Levine, J. M., J. Bascompte, P. B. Adler, and S. Allesina. 2017. Beyond pairwise mechanisms of

species coexistence in complex communities. Nature 546:56–64.

Levine, J. M., and J. HilleRisLambers. 2009. The importance of niches for the maintenance of

species diversity. Nature 461:254–257.

Lichstein, J. W., J. Dushoff, S. A. Levin, and S. W. Pacala. 2007. Intraspecific variation and

species coexistence. American Naturalist 170:807–818.

Losos, E. C., and E. G. Leigh, editors. 2004. Tropical forest diversity and dynamism: findings

from a large-scale plot network. University of Chicago Press, Chicago.

Lotka, A. J. 1925. Elements of physical biology. Williams&Wilkins Co.

le Maire, G., J. Guillemot, O. C. Campoe, J.-L. Stape, J.-P. Laclau, and Y. Nouvellon. 2019.

Light absorption, light use efficiency and productivity of 16 contrasted genotypes of

several Eucalyptus species along a 6-year rotation in Brazil. Forest Ecology and

Management 449:117443.

Maréchaux, I., L. Saint-André, M. K. Bartlett, L. Sack, and J. Chave. 2020. Leaf drought

tolerance cannot be inferred from classic leaf traits in a tropical rainforest. Journal of

Ecology 108:1030–1045.

Marrot, P., M. Latutrie, J. Piquet, and B. Pujol. 2021. Natural selection fluctuates at an extremely

fine spatial scale inside a wild population of snapdragon plants. Evolution:evo.14359.

McGill, B. J., B. A. Maurer, and M. D. Weiser. 2006. Empirical evaluation of neutral theory.

Ecology 87:1411–1423.

Metcalf, C. J. E., J. S. Clark, and D. A. Clark. 2009. Tree Growth Inference and Prediction When

the Point of Measurement Changes: Modelling around Buttresses in Tropical Forests.

Journal of Tropical Ecology 25:1–12.

Moran, E. V., F. Hartig, and D. M. Bell. 2016. Intraspecific trait variation across scales:

implications for understanding global change responses. Global Change Biology 22:137–

150.

Muller-Landau, H. C. 2003. Seeds of understanding of plant diversity. Proceedings of the

National Academy of Sciences 100:1469–1471.

Muller-Landau, H. C., S. J. Wright, O. Calderón, R. Condit, and S. P. Hubbell. 2008.

Interspecific variation in primary seed dispersal in a tropical forest. Journal of Ecology

96:653–667.

Nicotra, A. B., O. K. Atkin, S. P. Bonser, A. M. Davidson, E. J. Finnegan, U. Mathesius, P. Poot,

M. D. Purugganan, C. L. Richards, F. Valladares, and M. van Kleunen. 2010. Plant

phenotypic plasticity in a changing climate. Trends in Plant Science 15:684–692.

Normand, D., A. Mariaux, P. Détienne, and P. Langbour. 2017. CIRAD's xylotheque.

Paradis, E., and K. Schliep. 2019. ape 5.0: an environment for modern phylogenetics and

evolutionary analyses in R. Bioinformatics 35:526–528.

Pélissier, R., J.-P. Pascal, N. Ayyappan, B. R. Ramesh, S. Aravajy, and S. R. Ramalingam. 2011.

Tree demography in an undisturbed Dipterocarp permanent sample plot at Uppangala,

Western Ghats of India: *Ecological Archives* E092-115. Ecology 92:1376–1376.

Poorter, L., C. V. Castilho, J. Schietti, R. S. Oliveira, and F. R. C. Costa. 2018. Can traits predict

individual growth performance? A test in a hyperdiverse tropical forest. New Phytologist

219:109–121.

Purves, D. W., and M. C. Vanderwel. 2014. Traits States and Rates: Understanding Coexistence

in Forests. Page Forests and Global Change. British Ecological Society. Cambridge

University Press.

Rees, M. 2001. Long-Term Studies of Vegetation Dynamics. Science 293:650–655.

Rüger, N., L. S. Comita, R. Condit, D. Purves, B. Rosenbaum, M. D. Visser, S. J. Wright, and C.

Wirth. 2018. Beyond the fast-slow continuum: demographic dimensions structuring a

tropical tree community. Ecology Letters 21:1075–1084.

Schmitt, S., N. Tysklind, B. Hérault, and M. Heuertz. 2021. Topography drives microgeographic

adaptations of closely related species in two tropical tree species complexes. Molecular

Ecology 30:5080–5093.

Seidler, T. G., and J. B. Plotkin. 2006. Seed Dispersal and Spatial Pattern in Tropical Trees.

PLoS Biology 4:e344.

Shipley, B., F. De Bello, J. H. C. Cornelissen, E. Laliberté, D. C. Laughlin, and P. B. Reich.

2016. Reinforcing loose foundation stones in trait-based plant ecology. Oecologia

180:923–931.

Siefert, A., C. Violle, L. Chalmandrier, C. H. Albert, A. Taudiere, A. Fajardo, L. W. Aarssen, C.

Baraloto, M. B. Carlucci, M. V. Cianciaruso, V. de L. Dantas, F. de Bello, L. D. S.

Duarte, C. R. Fonseca, G. T. Freschet, S. Gaucherand, N. Gross, K. Hikosaka, B.

Jackson, V. Jung, C. Kamiyama, M. Katabuchi, S. W. Kembel, E. Kichenin, N. J. B.

Kraft, A. Lagerström, Y. L. Bagousse-Pinguet, Y. Li, N. Mason, J. Messier, T.

Nakashizuka, J. M. Overton, D. A. Peltzer, I. M. Pérez-Ramos, V. D. Pillar, H. C.

Prentice, S. Richardson, T. Sasaki, B. S. Schamp, C. Schöb, B. Shipley, M. Sundqvist, M.

T. Sykes, M. Vandewalle, and D. A. Wardle. 2015. A global meta-analysis of the relative

extent of intraspecific trait variation in plant communities. Ecology Letters 18:1406–

1419.

Stump, S. M., C. Song, S. Saavedra, J. M. Levine, and D. A. Vasseur. 2021. Synthesizing the

effects of individual-level variation on coexistence. Ecological Monographs.

Tilman, D. 1982. Resource competition and community structure. Princeton University Press,

Princeton, N.J.

Tymen, B., G. Vincent, E. A. Courtois, J. Heurtebize, J. Dauzat, I. Marechaux, and J. Chave.

2017. Quantifying micro-environmental variation in tropical rainforest understory at

landscape scale by combining airborne LiDAR scanning and a sensor network. Annals of

Forest Science 74:32.

Uriarte, M., and D. Menge. 2018. Variation between individuals fosters regional species

coexistence. Ecology Letters 21:1496–1504.

Vieilledent, G., B. Courbaud, G. Kunstler, J.-F. Dhôte, and J. S. Clark. 2010. Individual

variability in tree allometry determines light resource allocation in forest ecosystems: a

hierarchical Bayesian approach. Oecologia 163:759–773.

Violle, C., B. J. Enquist, B. J. McGill, L. Jiang, C. H. Albert, C. Hulshof, V. Jung, and J.

Messier. 2012. The return of the variance: Intraspecific variability in community ecology.

Trends in Ecology and Evolution 27:244–252.

Vleminckx, J., C. Fortunel, O. Valverde-Barrantes, C. E. Timothy Paine, J. Engel, P. Petronelli,

A. K. Dourdain, J. Guevara, S. Béroujon, and C. Baraloto. 2021. Resolving whole-plant

economics from leaf, stem and root traits of 1467 Amazonian tree species.

Oikos:oik.08284.

Volterra, V. 1926. Fluctuations in the Abundance of a Species considered Mathematically.

Nature 118:558–560.

Von Humboldt, A. 1849. Aspects of Nature, in Different Lands and Different Climates; with

Scientific Elucidations. Lea and Blanchard, Philadelphia.

Westoby, M., and I. J. Wright. 2006. Land-plant ecology on the basis of functional traits. Trends

in Ecology and Evolution 21:261–268.

Wilson, J. B., R. K. Peet, J. Dengler, and M. Pärtel. 2012. Plant species richness: the world

records. Journal of Vegetation Science 23:796–802.

Wright, J. S. 2002. Plant diversity in tropical forests: a review of mechanisms of species

coexistence. Oecologia 130:1–14.

Zanne, A. E., G. Lopez-Gonzalez, D. A. Coomes, J. Ilic, S. Jansen, S. L. Lewis, R. B. Miller, N.

G. Swenson, M. C. Wiemann, and J. Chave. 2009. Data from: Towards a worldwide

wood economics spectrum. Dryad.

Zellweger, F., P. De Frenne, J. Lenoir, D. Rocchini, and D. Coomes. 2019. Advances in

Microclimate Ecology Arising from Remote Sensing. Trends in Ecology & Evolution

34:327–341.

Zinger, L., P. Taberlet, H. Schimann, A. Bonin, F. Boyer, M. De Barba, P. Gaucher, L. Gielly, C.

Giguet-Covex, A. Iribar, M. Réjou-Méchain, G. Rayé, D. Rioux, V. Schilling, B. Tymen,

J. Viers, C. Zouiten, W. Thuiller, E. Coissac, and J. Chave. 2019. Body size determines

soil community assembly in a tropical forest. Molecular Ecology 28:528–543.

Box 1: High dimensional environmental variation and species differences

The environment varies strongly in space and time, and this temporal and spatial heterogeneity

of the environment has long been recognized as a key driver of species distributions and commu-

nity assembly (Von Humboldt 1849, Grisebach 1872). Novel remote sensing tools such as high-

spatial and temporal resolution airborne LiDAR scans (Tymen et al. 2017, Cushman et al. 2022),

intensive soil samplings and metabarcoding (Zinger et al. 2019), and more generally studies on

the microclimate (Zellweger et al. 2019) and microhabitats (Baraloto and Couteron 2010) have

also evidenced strong environmental variation operating at fine scales (e.g. cm to meter scales) in

many dimensions (Fig. 6). These environmental dimensions can be resources for which species

compete (e.g. light, water, nutrients) but also all other components that shape the environment lo-

cally (e.g. temperature, wind, elevation, slope, soil texture, soil microorganisms etc.).

In parallel, naturalists and taxonomists have long documented species differences in many as-

pects of their morphology and life history (Fig. 7). Such differences between species have then

been specified and quantified through traits that drive each species response to the environment

(species functional traits, McGill et al. 2006; Westoby & Wright 2006). Similar to the environ-

ment that presents highly-dimensional variation at local scales, these functional species differ-

ences spread along many dimensions within communities (Hutchinson 1957, 1959, Baraloto et

al. 2010b, Kraft et al. 2015, Rüger et al. 2018, Maréchaux et al. 2020, Vleminckx et al. 2021).

The adaptive significance of trait differences between species is widely appreciated (Westoby

and Wright 2006, Falster et al. 2017); many of these species traits could never have evolved if

they did not contribute to species fitness. Because both species characteristics and environment

830 vary in a high number of dimensions, each species can outcompete the others in a particular set

831 of environmental conditions, at least as defined at one point in space and time.

832 **Table 1: Mean posteriors of the *Eucalyptus* model and their estimation errors and residual**

833 **variance partitioning among the different random effects.**

	Intercept (β_0)	Diameter (β_1)	Competition (β_2)	Individual variance (V_i)	Block variance (V_b)	Genetic variance (V_g)	Temporal variance (V_t)	Residual variance (V)
Estimate	-3.70E-02	5.50E-01	-2.70E-01	2.30E-01	5.40E-02	1.30E-01	1.20E+00	5.10E-01
Estimation error	4.60E-01	5.10E-03	8.90E-03	4.10E-03	1.50E-02	2.90E-02	5.70E-01	2.00E-03
Percentage of unexplained variance				10.83%	2.54%	6.12%	56.50%	24.01%

836 **Table 2: Features of the three tropical forest data sets used as empirical case studies.**

Site	Rainfall (mm.y^{-1})	Sampling	Min DBH	Nb of censuses	Periodicity	Disturbance	Topography	Nb of species	Nb of individuals	Data source
Paracou, French Guiana	3,000	15 × 6.25 ha (incl. 12 logged plots)	10 cm	24	1-2 y since 1992	Natural disturbances + selective logging	flat	613	69,548	(Hérault and Piponiot 2018)
BCI, Panama	2,600	50 ha	10 cm	8	5-y since 1980	Natural disturbances	hilly	225	37,224	(Condit et al. 2019)
Uppangala, India	5,100	5,92 ha (4 transects and 3 plots)	9.5 cm	20	1-yr since 1992	Natural disturbances	mountainous	102	3,789	(Le Bec et al. 2015)

838 **Table 3: Mean posteriors of the tropical forest model and their estimation errors and resid-**

839 **ual variance partitioning among the different random effects.**

	Intercept (β_0)	Diameter (β_1)	Species variance (V_b)	Individual variance (V_d)	Residual variance (V)
Paracou					
Estimate	6.70E-02	2.90E-02	4.70E-01	4.60E-01	7.60E-01
Estimation error	2.20E-02	3.80E-03	1.70E-02	3.90E-03	2.30E-03
% Variance			27.81%	27.22%	44.97%
Uppangala					
Estimate	8.40E-02	1.90E-01	3.60E-01	6.60E-01	5.90E-01
Estimation error	4.40E-02	1.20E-02	4.30E-02	8.60E-03	1.90E-03
% Variance			22.36%	40.99%	36.65%
BCI					
Estimate	1.90E-01	-2.20E-02	6.60E-01	4.10E-01	8.10E-01
Estimation error	5.00E-02	4.50E-03	3.50E-02	4.00E-03	2.00E-03
% Variance			35.11%	21.81%	43.09%

842   **Table 4**: **Spatial autocorrelation of the growth of conspecific individuals in three tropical**

843   **forest sites.**

845   Shown are the proportion of species, and of corresponding individuals, in percent, for which in-

846   dividual growth among conspecific individuals is significantly positively spatially autocorre-

847   lated. The spatial autocorrelation of individual growth was tested using Moran's I index.

	Significant	Not Significant
Paracou		
% Species	31.00	69.00
% Individuals	78.90	21.10
Uppangala		
% Species	18.50	81.50
% Individuals	45.30	54.70
BCI		
% Species	20.10	79.90
% Individuals	54.70	45.30

Table 5: **Comparison of local intra- and interspecific variability in individual growth for**

three tropical forest sites.

The variability was estimated with the semivariance and the comparison was performed with a

Mann-Whitney's test. The semivariances were computed for all species with > 5 individuals and

> 5 heterospecific neighbors within 100 m in the same plot, and considering pairs of individuals

that were less than 100 m apart and in the same plot. Shown are the proportion of species, and of

corresponding individuals, for which (i) intraspecific variability was significantly lower than in-

terspecific variability, (ii) intraspecific variability was significantly higher than interspecific vari-

ability, or (iii) the difference between inter- and intraspecific variabilities was not significant.

	Intraspecific variability < Interspecific variability (i)	Intraspecific variability ~ Interspecific variability (ii)	Intraspecific variability > Interspecific variability (iii)
Paracou			
% Species	60.70	40.70	0.67
% Individuals	88.80	10.90	0.28
Uppangala			
% Species	42.20	62.20	4.44
% Individuals	57.70	23.60	18.80
BCI			
% Species	46.10	47.80	3.14
% Individuals	76.00	19.30	4.69

Figure 1: **Reinterpreting observed intraspecific variability (IV): from niche widening to**

niche projection into a high-dimensional environment.

In **(a)**, within a given environment E0 defined along an environmental axis x (E0=E(x0)), con-

specific individuals are identical and have the same performance p1 and p2, for species 1 (blue)

and species 2 (orange). Species 2 outcompetes species 1 in E0. Actual measured differences

among conspecific individuals, shown in **(b)**, can be interpreted in different ways. First, as con-

specific individuals exhibit contrasting attributes in E0, they become more different. This can re-

sult in some heterospecific individuals having similar performances: IV would blur species dif-

ferences. Alternatively, IV measured in E0 results from the variation of unobserved environmen-

tal variables (E0=E(x0, y); **(c)**). Contrasting performances among conspecific individuals in E0

do not result from intrinsic differences among them but from differences in the local environ-

ment they experience and that was poorly characterized, i.e. the number of observed dimensions

is lower than the actual number of environmental dimensions. Similarly, although species niches

present some overlap when projected on one dimension **(d)**, they do not overlap in the two-di-

mensional space (c). Moreover, while species 2 outcompetes species 1 on average when x=x0,

the opposite occurs when x=x1 (d), leading to an inversion of species hierarchy between differ-

ent environments. Similarly, while species 1 outcompetes species 2 in E(x0, y0), the opposite oc-

curs in E(x0, y1). Although only two dimensions are shown, species respond to many environ-

mental variables varying in space and time, multiplying the possibilities of niche segregation and

hierarchy inversions between species, offering room for species coexistence in a variable high-

dimensional environment. The code used to generate this figure is available online.

Figure 2: Body of evidence on the nature of IV and its consequences on individual and

species differences.

We used literature and data analyses of various nature to support the hypothesis that a large part

of observed IV is caused by multidimensional environmental variations that are spatially and

temporally structured rather than by intrinsic differences between conspecific individuals, with

radically different consequences on species coexistence.

Figure 3: Observed intraspecific variability as a result of the imperfect characterization of

the environment.

A virtual response variable (Y, e.g. growth) is generated for individual clones of two species

thriving in a high-dimensional environment. This response variable was first computed as a func-

tion of ten environmental variables (perfect knowledge model, Eq. I), but is then analyzed using

a statistical model that accounts for the unique environmental variable that was assumed to be

observed in the field (X_1, e.g. light) and includes a random individual effect (partial knowledge

model, Eq. II). The intraspecific variability estimated with these random individual effects is

then due to the variation in space and time of the nine unobserved environmental variables. (a)

Positions of a sample of I=600 individuals from J=2 species in a landscape defined by a square

grid of C × C cells (C=500). The background color indicates the value of the observed environ-

mental variable X_1 on each cell at date t. The response Y of each individual, which depends on

the environment within each cell (Eq. I), is also indicated by a color scale. (b) Response Y as a

function of the observed environmental variable X_1 for the two species. Points represent the data

$\{Y_{ijt}, X_{1,ijt}\}$. Thick lines represent the predictive posterior means for the two species. The en-

velopes delimited by two thin lines represent the 95% credible intervals of the predictive poste-

rior marginalized over individuals (taking into account V_{bj}). The envelopes thus represent the in-

traspecific variability which is due to the N-1 unobserved environmental variables.

Figure 4: Spatial autocorrelation of attribute Y across individuals within and between

species (J=2).

This semivariogram represents the semivariance of the individual mean attribute Y as a function

of the distance between individuals. The increasing curves evidence spatial autocorrelation in Y

(similar results using Moran's I test). The semivariance of all individuals taken together (purple

curve) is higher than the semivariance of conspecific individuals for the two species (orange and

blue curves), which means that intraspecific variability is lower than interspecific variability.

Figure 5: Experimental setup of the EUCFLUX experiment

The ten blocks (a) and the organization of the 16 genotypes within a block (b). In our analyses,

two genotypes were discarded because they were obtained from seeds and not clones and there-

fore included some genetic variability. A more complete figure legend can be found in le Maire

et al. 2019.

Figure 6: High environmental variability at a small spatial scale.

(a) Soil nitrogen content in a 12×12 m plot at Cedar Creek in g.kg⁻¹, (USA), (Tilman 1982); **(b)** Carbon in % (left) and aluminum in ppm (right) soil content in a 12-ha (250×500 m) plot at The Nouragues (French Guiana), (Zinger et al. 2019); **(c)** Soil water content during mid-dry season of a regular year in MPa in a 50-ha (1000×500 m) forest plot at Barro Colorado Island (Panama), (Kupers et al. 2019). Coordinates in m.; **(d)** Canopy height in m and topography (10 m spaced elevation lines) in a 50-ha (2500×2000 m) area at the Nouragues Research Field Station, (Tymen et al. 2017). Coordinates in m (UTM 22N).

Figure 7: Morphological diversity of tree species illustrating strong differences between

species.

(a) Diversity of tree species architecture and height in a tropical forest (Hallé et al. 1978). Coor-

dinates are in m.; (b) Diversity of seed size and shape from 17 tree species of the Fabaceae fam-

ily in the Peruvian Amazon (Muller-Landau 2003); (c) Diversity of leaf size and shape (herbar-

ium of Cayenne, Gonzalez et al. 2021) and of wood aspect (reflecting wood characteristics) and

density (Normand et al. 2017) for 12 tree species in French Guiana. Species from top left to bot-

tom right are Bocoa prouacensis, Zygia racemosa, Vouacapoua americana, Eperua falcata,

Bagassa guianensis, Hymenolobium excelsum, Mangifera indica, Sterculia pruriens, Parkia ni-

tida, Couroupita guianensis, Hura crepitans, and Ceiba pentandra. Black bars next to herbarium

samples indicate the scale (10 cm).