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1 Rethinking the nature of intraspecific variability and its 2 consequences on species coexistence

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25 **Abstract**

26 **Context** Intraspecific variability (IV) has been proposed to explain species coexistence in
27 diverse communities. Assuming, sometimes implicitly, that conspecific individuals can perform
28 differently in the same environment and that IV blurs species differences, previous studies have
29 found contrasting results regarding the effect of IV on species coexistence.

30 **Objective** We aim at showing that the large IV observed in data does not mean that conspecific
31 individuals are necessarily different in their response to the environment and that the role of
32 high-dimensional environmental variation in determining IV has been largely underestimated in
33 forest plant communities.

34 **Methods and Results** We first used a simulation experiment where an individual attribute is
35 derived from a high-dimensional model, representing “perfect knowledge” of individual
36 response to the environment, to illustrate how a large observed IV can result from “imperfect
37 knowledge” of the environment. Second, using growth data from clonal *Eucalyptus* plantations
38 in Brazil, we estimated a major contribution of the environment in determining individual
39 growth. Third, using tree growth data from long-term tropical forest inventories in French
40 Guiana, Panama and India, we showed that tree growth in tropical forests is structured spatially
41 and that despite a large observed IV at the population level, conspecific individuals perform
42 more similarly locally than compared with heterospecific individuals.

43 **Synthesis** As the number of environmental dimensions that are typically quantified is generally
44 much lower than the actual number of environmental dimensions influencing individual
45 attributes, a great part of observed IV might be misinterpreted as random variation across

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46 individuals when in fact it is environmentally-driven. This mis-representation has important
47 consequences for inference about community dynamics. We emphasize that observed IV does not
48 necessarily impact species coexistence *per se* but can reveal species response to high-
49 dimensional environment, which is consistent with niche theory and the observation of the many
50 differences between species in nature.

51

52 **Keywords:** competition; environmental variation; high-dimensional niche; individual variation;
53 intraspecific variability; spatial autocorrelation; spatial heterogeneity; species coexistence

54 ***Introduction***

55 Ecological communities are characterized by numerous coexisting species, for instance in
56 grasslands, coral reefs or tropical forests. Understanding how these species stably coexist while
57 competing for the same basic resources, *viz.* light, water, and nutrients (Baraloto et al. 2010), is a
58 longest-standing question in ecology (Gause 1934, Hutchinson 1961, Levine et al. 2017).
59 Although numerous mechanisms have been suggested to contribute to species coexistence
60 (Janzen 1970, Connell 1971, Chesson 2000, Hubbell 2001, Wright 2002, Levine and
61 HilleRisLambers 2009), it is unclear when and to what extent they explain the high species
62 diversity observed in nature (Clark 2010). This is especially true in forests, where tree species
63 coexist while seemingly requiring similar resources in the same location. Astonishingly, a
64 hectare of tropical forest can harbor more than 900 plant species of a diversity of forms and
65 functions (Wilson et al. 2012).

66 Many theoretical mechanisms that might explain tree species coexistence typically follow the
67 assumption that all conspecific individuals are identical. However, intraspecific variability (IV)
68 in traits, demographic rates or any proxy of performance, henceforth denoted as “attributes”, can
69 alter community structure and dynamics (Bolnick et al. 2011). Indeed, large IV has been
70 observed across a number of attributes in plant communities (Albert et al. 2012, Violle et al.
71 2012). For instance, Siefert et al. 2015 estimated that IV accounted for 25% of the variability in
72 functional traits within plant communities on average, and this proportion was even estimated at
73 44% in a tropical forest (Poorter et al. 2018). Likewise, IV in growth rates for trees of
74 standardized size, local crowding and terrain slope has been found to account for up to 58% of
75 total growth variability in a tropical forest stand (Le Bec et al. 2015).

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77 IV, as a pathway for coexistence, has so far not shared the same attention as other mechanisms.
78 This is in part because modeling studies that have explored the effect of IV on species
79 coexistence have yielded contrasting results (Stump et al. 2021). In most theoretical analyses,
80 variability in attributes among conspecific individuals has been included through independent
81 random draws (Lichstein et al. 2007, Hart et al. 2016, Barabás and D’Andrea 2016, Crawford et
82 al. 2019 but see Purves and Vanderwel 2014, Banitz 2019). Similarly, empirical studies typically
83 summarize IV as a variance around species mean attributes (Jung et al. 2010, Albert et al. 2010,
84 Siefert et al. 2015, Poorter et al. 2018). With this representation, IV can increase species niche
85 overlap and blur species differences, sometimes slowing down competitive exclusion in models
86 of community dynamics (Vieilledent et al. 2010, Crawford et al. 2019). However, in some other
87 models, non-linear responses can make such IV beneficial to the superior competitors (*i.e.* the
88 most competitive individuals of the more competitive species), thus accelerating competitive
89 exclusion (*e.g.* Courbaud et al. 2012, Hart et al. 2016). Alternatively, in particular spatial
90 configurations, more precisely when IV is greater in species preferred habitats, it has been shown
91 to foster species coexistence (Uriarte and Menge 2018). Stump et al. (2021) have proposed to
92 reconcile these contrasting results by distinguishing the effect of IV on niche traits (which
93 control individual performance response to environmental conditions) vs. hierarchical traits
94 (which control individual performance independently from environmental conditions). They
95 demonstrated with different simulation models of community dynamics that IV in traits can alter
96 stabilizing mechanisms and fitness differences in a complex way which depends upon the nature
97 of the traits (niche vs. hierarchical) and their response curve, and thus promote or not species
98 coexistence. In all the above examples however, IV, since simulated through independent

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99 random draws around species mean attributes, would be caused by differences among
100 individuals that are fully independent of the environment: differences among individuals would
101 remain unchanged even when experiencing exactly the same environmental conditions.
102 Importantly, such simulated IV thus leads to a variation among conspecific individuals that is
103 completely unstructured in space and time. New appreciation of fine-scale environmental
104 heterogeneity and structure as well as species differences in their response to the environment,
105 however, may suggest that this assumption of unstructured IV is rarely met.

106 Novel remote sensing tools such as high-spatial and -temporal resolution airborne LiDAR scans
107 (Tymen et al. 2017, Cushman et al. 2022), intensive soil samplings and metabarcoding (Zinger et
108 al. 2019), and more generally studies on the microclimate (Zellweger et al. 2019) and
109 microhabitats (Baraloto and Couteron 2010) have indeed evidenced strong environmental
110 variation operating at fine scales (*e.g.* cm to meter scales) in many dimensions (Fig. 1). These
111 environmental dimensions can be resources for which species compete (*e.g.* light, water,
112 nutrients) but also all other components that shape the environment locally in space and time
113 (*e.g.* temperature, wind, elevation, slope, soil texture, soil microorganisms *etc.*). In parallel,
114 naturalists and taxonomists have long documented species differences in many aspects of their
115 morphology and life history (Fig. 2). Such differences between species have then been specified
116 and quantified through traits that drive each species response to the environment (species
117 functional traits, McGill et al. 2006, Westoby and Wright 2006). Similar to the environment that
118 presents highly-dimensional variation at local scales, these functional species differences spread
119 along many dimensions within communities (Hutchinson 1957, 1959, Baraloto et al. 2010, Kraft
120 et al. 2015, Rüger et al. 2018, Maréchaux et al. 2020, Vleminckx et al. 2021).

121 In this paper, we explore the potential that the role of environmental variation in shaping

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122 *observed* IV has been largely underestimated with important consequences on our understanding
123 of the effect of *observed* IV on community dynamics. Indeed, a great part of *observed* IV might
124 emerge from species responses to a high-dimensional environment (Fig. 3): observed differences
125 among individuals of the same species can be caused by the (often poorly quantified) differences
126 in the micro-environment they experience. If so, variation among conspecific individuals would
127 be structured in space and time, and not necessarily by genetic variation. More specifically, we
128 present insights from a simulation experiment, experimental data, and tropical forest inventory
129 data in order to examine three hypotheses (Fig. 4): (i) the large IV observed in natural
130 communities can emerge from heterogeneity in multiple unobserved environmental dimensions;
131 (ii) because environmental variation is structured in space and time, IV is likely to be similarly
132 structured as well, suggesting that it is not appropriate to represent IV as a purely random noise
133 in models; and (iii) since a large observed IV does not necessarily imply that conspecific
134 individuals substantially differ in their fundamental niche, conspecific individuals may still
135 respond more similarly to environment than heterospecific individuals. We therefore call for a
136 reconsideration of the nature and structure of observed IV, which could shed new light on the
137 coexistence conundrum. While we acknowledge the existence of genetically-based individual
138 variations, and that plasticity has a genetic basis (Nicotra et al. 2010, Westerland et al. 2021), we
139 suggest that a substantial part of *observed* IV might result from the higher dimensionality of the
140 species niche than typically observed. Species differences along these many dimensions can lead
141 to multiple local inversions of species hierarchy in an environment varying in space and time,
142 thereby allowing the stable coexistence of numerous species.

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144 ***Theoretical illustration: unobserved environmental***

145 ***dimensions result in large observed IV***

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147 We first conducted a simulation experiment to illustrate how observed intraspecific variability,
148 or “individual effects”, can result from variation in unobserved environmental variables (as
149 suggested by (Clark et al. 2007). We generated simulated data of an individual attribute (here
150 tree growth) depending on a certain number of environmental variables varying in space, and
151 then analyzed the simulated data assuming that most of the environmental variables are actually
152 unobserved, as it is typically the case in the field.

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154 **A “perfect knowledge” simulation model**

155 We considered a set of J species with I individuals each, distributed in a virtual landscape. The
156 environment was assumed to be fully known and defined by N environmental variables, X_1 to X_N ,
157 that were each randomly and independently generated in the landscape, assuming spatial
158 autocorrelation. Individual location was drawn randomly in a virtual landscape defined by a $C \times$
159 C square grid, each cell corresponding to a particular environment (Fig. 5a). Individuals were
160 identical within species (same model parameters for all conspecific individuals), but different
161 between species (different model parameters between heterospecific individuals).

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163 We considered the following “perfect knowledge” mathematical model, which depicts the exact
164 attribute Y_{ijt} (e.g., growth) of an individual i of species j given its environment at time t (Eq. I,

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165 Appendix 1).

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167 $\ln(Y_{ijt}) = \beta_{0,j} + \beta_{1,j} \ln(X_{1,ijt}) + \beta_{2,j} X_{2,ijt} + \dots + \beta_{N,j} X_{N,ijt}$ (Eq. I)

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169 In this model, $\beta_j = [\beta_{0,j}, \dots, \beta_{N,j}]$ is the vector of parameters defining the response of individuals of
170 species j to the environment. Because conspecific individuals respond similarly to environmental
171 variables, variation in Y_{ijt} among them is only due to differences in the environment where and
172 when each individual is growing. Using this model, we computed the attribute Y of the $I \times J$
173 individuals at T dates, assuming that values for some of the environmental variables changed
174 between dates, and thus obtained a simulated dataset $\{Y_{ijt}, X_{1,ijt}, \dots, X_{N,ijt}\}$ with $N=10, I = 300, J =$
175 $2, C = 500$ and $T = 2$.

176

177 **An “imperfect knowledge” statistical model**

178 Second, we considered an “imperfect knowledge” statistical model for which we assumed that
179 only one explanatory variable X_1 (e.g., light) in the above simulated dataset has been measured
180 among all the environmental drivers that actually determine response variable Y (Eq. II,
181 Appendix 1). This model represents the ecologist’s imperfect understanding of attribute Y . The
182 model includes a species fixed effects on the intercept and on the slope ($\beta'_{0,j}$ and $\beta'_{1,j}$) and a
183 random individual effect $b_{0,i}$ on the intercept, $b_{0,i} \sim N(0, V_{bj})$, where V_{bj} is the intraspecific
184 variance for species j . We estimated the model parameters based on the simulated dataset
185 introduced above but considering only the first explanatory variable $\{Y_{ijt}, X_{1,ijt}\}$, the remaining
186 “unknown” environmental effects being contained in the model residuals, ε_{ijt} .

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188 $\ln(Y_{ijt}) = [\beta'_{0j} + b_{0,i}] + \beta'_{1,j} \ln(X_{1,ijt}) + \varepsilon_{ijt}, \varepsilon_{ijt} \sim N(0, V_j)$ (Eq. II)

189

190 **Apparent niche overlap and observed intraspecific variability as a result of**
191 **unobserved environmental variables**

192 Despite the fact that conspecific individuals were identical and species responses to environment
193 were different, the variance estimates \hat{V}_{bj} for individual random effects of species j were large,
194 and species responses to the environment overlapped (Fig. 5b). This is due to the contribution of
195 the unmeasured variables $\{X_{2,ijt}, \dots, X_{N,ijt}\}$ in determining the variation of Y across individuals.

196

197 Since it is driven by spatially autocorrelated variables (Eq. I), the response Y was spatially
198 autocorrelated across conspecific individuals (Fig. 6). This means that two neighboring
199 conspecific individuals have more similar attribute Y than two distant conspecific individuals.
200 Additionally, the variance of Y was lower within than between species: conspecific individuals
201 responded more similarly to the environment than heterospecific individuals did (Fig. 6).

202

203 With this simulation experiment, we simply illustrated that: (i) a high IV can emerge from
204 unobserved environmental dimensions exclusively, (ii) the spatial structure of IV follows the
205 spatial structure of the underlying environmental variables, and (iii) IV does not blur differences
206 between species (Fig. 6) despite apparent niche overlap (Fig 5b).

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208 ***Experimental insights: large observed intraspecific variability***
209 ***in a clonal tree plantation***

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211 We then moved from a theoretical to an experimental approach using census data from clonal
212 *Eucalyptus* plantations, where genetic variability among individuals growing within a single
213 same site is controlled. We explore the partitioning of IV between intrinsic (genotypes) and
214 extrinsic sources, which is often infeasible in natural settings, to demonstrate that substantial
215 observed IV can indeed emerge from genetically identical individuals in the field, even when
216 persisting in an apparently homogeneous environment.

217

218 **An extreme case of controlled genetic and environmental variation**

219 The EUCFLUX experiment (São Paulo state, Brazil) is a clonal trial with a replicated,
220 statistically-sound design (le Maire et al. 2019). It includes 14 genotypes of 5 different
221 *Eucalyptus* species or hybrids of various origins. Each genotype is planted in plots of 100 trees,
222 at a density of 1666 trees per hectare, and replicated spatially in 10 blocks (Fig. 7). The
223 experimental set-up was designed to minimize the variation in environmental factors among
224 blocks, which were separated by less than 1.5 km within a homogeneous 200-ha stand showing
225 small variation in soil properties. Tree diameter at breast height (D) has been measured over 5
226 complete censuses, spanning 6 years, age at which such plantation is generally harvested (see le
227 Maire et al. 2019 and Appendix 2 for further details on this experimental set-up).

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229 **A partitioning of observed variance among individual tree growth**

230 We computed annual diameter growth (G) in mm/yr^{-1} for each tree as well as a competition index
231 (C) as the sum of the basal area of the eight direct neighbors of each tree. The dataset included
232 64,125 growth estimates corresponding to 13,531 trees in total. To quantify the relative
233 importance of the different sources of growth variability, we used a statistical hierarchical
234 growth model (Eq. III) including an intercept (β_0), fixed effects of the log-transformed diameter
235 (β_1) and competition index (β_2), and random effects on the intercept of the block ($b_{0,b}$, with $b_{0,b} \sim$
236 $N(0, V_b)$), the genotype ($b_{0,g}$, with $b_{0,g} \sim N(0, V_g)$), the census date ($b_{0,t}$, with $b_{0,t} \sim N(0, V_t)$), and
237 the individual ($b_{0,i}$, with $b_{0,i} \sim N(0, V_i)$). All the data were log-transformed and scaled, and a
238 constant of 1 mm was added to all growth values to avoid undefined logarithms.

239

$$240 \ln(G+1)_{i,t} = [\beta_0 + b_{0,b} + b_{0,g} + b_{0,t} + b_{0,i}] + \beta_1 \ln(D)_{i,t} + \beta_2 \ln(C)_{i,t} + \varepsilon_{i,t}, \varepsilon_{i,t} \sim N(0, V) \text{ (Eq. III)}$$

241

242 We used conjugated priors with inverse-gamma distributions (with shape and scale
243 parameters= 10^{-3}) for variance parameters, and normal distributions (with mean=0 and
244 variance=1) for mean parameters. The estimation of model parameters was done using a
245 Bayesian approach using Stan software with the brms R package (Bürkner 2017, 2018). We
246 made 10,000 iterations for each MCMC with a burn-in period of 5,000 steps and a thinning rate
247 of one fifth. We obtained 1,000 estimations per parameter and examined the trace plots to check
248 convergence of the MCMC chains.

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250 We then examined the proportion of the model residual variance (variation of the response
251 variable that is not explained by the covariates) related to each random effect in order to partition

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252 the block, genotype, date and individual variances.

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254 **Variation among individuals is not explained by genotype**

255 While minor variability was associated with blocks (Table 1), confirming that they are broadly
256 homogeneous by design, the variability associated with extrinsic temporal factors was
257 predominant (Table 1). It reveals that the competition index (C) used in the analysis to
258 encapsulate the effect of progressive canopy closure does not fully encompass all temporal
259 effects.

260 Importantly, the variability between individuals was almost twice as high as the variability
261 between genotypes (Table 1). Hence, even in such an extremely conservative case, where
262 environmental variation in space is minimized and genotypic variability controlled, a large part
263 of measured IV cannot be explained by purely-genetic differences among individuals that would
264 remain independent of the environment as an IV simulated through independent random draws
265 would be. This suggests an underestimated role of environmental micro-heterogeneity in shaping
266 variation among individuals, for instance inevitable spatial variation of biotic and abiotic
267 variables (soil microbiome, pathogens, soil structure and water content, light *etc.*) at fine scales
268 (e.g. cm- to m-scale, hence impacting tree-scale environment, Baraloto and Coueron 2010, Fig.
269 1) as well as potential early manipulations of the young plant, the way it was planted, *etc.*

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271 ***Empirical insights: observed intraspecific variability is high***
272 ***and spatially structured and does not blur species differences***
273 ***in tropical forests***

274 To test some of our hypotheses in natural communities, we then used data from three long-term
275 tree inventories in tropical forests, from Amazonia (Paracou, French Guiana; Gourlet-Fleury et
276 al. 2004), Central America (Barro Colorado Island, Panama; Losos and Leigh 2004) and South-
277 East Asia (Uppangala, India; Pélissier et al. 2011). More specifically, we inferred observed IV,
278 tested if individual growth showed local spatial autocorrelation, *i.e.* was structured in space, and
279 if conspecific individual growth was more similar than heterospecific individual growth locally.
280 These three sites encompass contrasting climatic conditions (rainfall ranging from 2,600 in BCI
281 to 5,100 mm.y⁻¹ in Uppangala), disturbance regimes (incl. various logging experiments in
282 Paracou) and topography (from gentle in BCI to mountainous in Uppangala), making them
283 representative of the global tropical forests. The data from these tropical forest inventories that
284 we used in this paper are summarized in Table 2.

285

286 For all three datasets, annualized growth between two censuses was computed as the difference
287 of DBH (≥ 10 cm) between two consecutive censuses, divided by the time period between those
288 two censuses. Growth estimates < -2 or > 100 mm.y⁻¹ as well as individuals from incompletely
289 identified species and individuals and species with a single observation were discarded prior to
290 analysis. Mean annual growth for each individual tree was then computed as the difference of
291 DBH between the first and the last time a tree was measured, divided by the time period between

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292 those two measurements.

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294 **High observed intraspecific variability in tree growth in tropical forests**

295 To quantify the relative importance of intra- vs. inter-specific variability in each site, we used a
296 hierarchical growth model (Eq. IV), including an intercept β_0 , a diameter (D) fixed effect β_1 , a
297 species random effect b_{oj} (with $b_{oj} \sim N(0, V_b)$) and an individual random effect d_{oi} (with $d_{oi} \sim N(0,$
298 $V_d)$) on the intercept. All data were log-transformed and scaled, and a constant of 2 mm was
299 added to all growth values to avoid undefined logarithms.

300

$$301 \ln(G_{ijt}+2) = [\beta_0 + b_{oj} + d_{oi}] + \beta_1 \times \ln(D_{ijt}) + \varepsilon_{ijt}, \varepsilon_{ijt} \sim N(0, V) \text{ (Eq. IV)}$$

302

303 For Paracou, which has a very large dataset, we sampled 100,000 growth values randomly to
304 perform inference. No sampling was done for Uppangala and BCI. We used priors with inverse-
305 gamma distributions (with shape and scale parameters= 10^{-3}) for variance parameters, and normal
306 distributions (with mean=0 and variance=1) for mean parameters. We estimated the inter- and
307 intra-specific growth variability from the variance of the species (V_b) and individual (V_d) random
308 effects, respectively. Model parameters were estimated the same Bayesian approach as for the
309 analysis of the *Eucalyptus* dataset.

310 For the three sites, IV estimated from the growth model (V_d , ranging from 0.41 to 0.66) was of
311 the same order of magnitude as the interspecific variance (V_b , ranging from 0.36 to 0.66) (Table
312 3). Overall, a large share of the variability in tree growth comes from individual effects in the
313 three sites, even after accounting for the effect of diameter on tree growth, showing a high

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314 intraspecific variability in growth in these tropical forests.

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316 **Spatial autocorrelation of individual growth within species at the local scale in**
317 **tropical forests**

318 To test whether individual growth was spatially autocorrelated, we performed in each site, spatial
319 analyses of the mean individual growth values. We chose a conservative approach based on
320 mean individual growth without accounting for the effect of diameter, thus without removing
321 ontogenetic differences and considering the pattern of individual growth as it is in the field. More
322 specifically, we performed Moran's I one-tailed tests as implemented in the *ape* R package
323 (Paradis and Schliep 2019), for pairs of conspecifics less than 100 m apart in the same plot (to
324 avoid capturing the effect of treatment in Paracou and including the spaces between the plots).
325 For the most abundant species, we sampled 3,000 individuals with a uniform probability. We
326 considered only the species with more than five conspecific neighbors less than 100 m-apart in
327 the same plot.

328 Positive spatial autocorrelation in tree growth between conspecifics was significant for 19 to
329 31% of the species in the three sites, representing between 45 and 79% of the total number of
330 individuals (Table 4). Spatial autocorrelation was however much higher in logged plots as
331 compared to unlogged in Paracou, because of a more heterogeneous light environment resulting
332 from logging history (Appendix 3).

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333 **Higher similarity of growth within conspecific- than heterospecific individuals**
334 **locally in tropical forests**

335 To test if the performance of conspecific individuals was locally more similar than the
336 performance of heterospecific individuals in the three sites, we also used mean individual
337 growth, thus ignoring ontogenetic differences. We computed the mean individual growth
338 semivariance (Baraloto and Coueron 2010) considering either conspecific or heterospecific
339 neighbors within a 100-m radius. In the first case, semivariance was estimated as the mean of the
340 squared difference in individual mean growth for all pairs of conspecific individuals. In the
341 second case, semivariance was estimated as the mean of the squared difference in individual
342 mean growth for all pairs of individuals with an individual of the focal species and one of
343 another species. We considered only the species with more than five individuals, and with more
344 than five heterospecific neighbors within the 100-m neighborhood distance. For each species, we
345 then compared the semivariances between conspecific and heterospecific individuals using a
346 Mann-Whitney test with a 0.05 alpha-risk.

347 The mean individual growth semivariance appeared significantly higher among heterospecifics
348 than among conspecifics for 42 to 61% of the species in the three sites, representing 58 to 89%
349 of the total number of individuals (Table 5). To control for a potential effect of species
350 abundance on the semivariance estimations, we replicated the analysis by sampling a maximum
351 of ten individuals per species. The results were qualitatively unchanged (Appendix 3).

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353 ***Discussion***

354 **High-dimensional environmental variation leads to large observed** 355 **intraspecific variability**

356 IV can result from intrinsic differences among individuals or from extrinsic environmental
357 variation, including biotic factors, or interactions of both (Violle et al. 2012, Moran et al. 2016,
358 Westerland et al. 2021). While much emphasis has been placed on genetically-driven IV in
359 studies on coexistence (Booth and Grime 2003, Ehlers et al. 2016, Barabás and D’Andrea 2016),
360 sometimes implicitly through the use of independent random draws across individuals (Lichstein
361 et al. 2007, Hart et al. 2016, Crawford et al. 2019), and although we acknowledge its ecological
362 and evolutionary importance, we here argue that the importance of environmentally-driven IV in
363 natural communities has been underestimated and has radically different consequences for
364 species differences and community assembly. More specifically, we argue that a large part of
365 observed IV can result from high-dimensional environmental variation in space and time.

366

367 First, using a simple simulation experiment, we illustrated how environmental variation in
368 unobserved dimensions of the environment can produce large observed IV, although conspecific
369 individuals are clones (Fig. 5). Similarly, the variance partitioning of individual tree growth
370 within a common garden of *Eucalyptus* clones (le Maire et al. 2019) shows that the variance in
371 growth between individuals is about twice as high as the variance between genotypes (Table 1).
372 This reveals that a large part of the observed IV can emerge from environmental variables, even
373 when the variation of the environment was sought to be minimized.

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375 Importantly, because IV can emerge from environmental heterogeneity without underlying
376 genetic differences, observed IV does not necessarily imply that conspecific individuals
377 substantially differ in their response to the environment, nor that species niches overlap (Fig. 3).
378 Instead, large observed IV can also reflect the projection of species' high-dimensional niches
379 within a high-dimensional environment that is variable in time and space: conspecific individuals
380 differ with each other because they each thrive in a different micro-environment. In empirically
381 observed data, such IV is therefore the result of projecting a high-dimensional response (*e.g.*,
382 physiological processes), which is controlled by multiple macro- and micro-environmental
383 variables, down to a low-dimensional, integrative response (*e.g.*, annual growth) that is poorly
384 characterized because of an incomplete view of the environmental variables that contribute to it.
385 This reassigns an important part of observed variation among individuals, often perceived as
386 neutral or random since they are seemingly unrelated to the observed dimensions of the
387 environment (Fig. 3, Fig. 5, Table 1), to the classical niche theory (Hutchinson 1957) and is in
388 agreement with natural history observations of individual trait differences that are associated to
389 species-specific ecological strategies.

390

391 The “biodiversity paradox” highlights that a large number of species can coexist while
392 competing for a limited number of resources (Hutchinson 1961). This puzzling question has
393 generally been tackled considering trade-offs along a limited number of niche axes, often
394 corresponding to resources (Tilman 1982, Rees 2001). But if the number of resources may
395 indeed be relatively limited (*e.g.*, light, water, and nutrients for plants), the number of
396 independent environmental factors (*e.g.* microclimatic variables) that drive the performance of
397 individuals for a particular level of resources is not. Environments are known to vary along

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398 multiple dimensions at fine scales in space and time (Fig. 1), and in many cases, this variation
399 has been shown to influence individual attributes (*e.g.*, Fortunel et al. 2020).

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401 Nevertheless, despite technological advances, many of these abiotic and biotic environmental
402 factors are still poorly understood and monitored. As a result, the dimensionality of field
403 observations is typically low compared with the high dimensionality of the environment in
404 nature (Bramer et al. 2018, Estes et al. 2018). The variability in individual attributes due to the
405 variation of unobserved environmental variables therefore remains mostly a black box, and is
406 typically summarized in terms of residual variance in statistical models (Albert et al. 2012,
407 Siefert et al. 2015) or encapsulated into so-called “individual random effects” (Clark et al. 2007).
408 We here emphasize that even in the absence of any intrinsic differences among conspecific
409 individuals, a large IV can emerge from the imperfect characterization of the environment (Fig.
410 3, Fig. 5, Table 1), which varies in a high number of dimensions (Fig. 1).

411

412 **Intraspecific variability is structured in space and time**

413 IV has commonly been perceived and modeled through independent random draws around the
414 species mean in community ecology studies (Lichstein et al. 2007, Courbaud et al. 2012, Hart et
415 al. 2016, Barabás and D’Andrea 2016, Uriarte and Menge 2018). While this representation
416 typically results from a lack of knowledge, with randomness being used as a substitute for more
417 detailed understanding of underlying ecological processes (Clark et al. 2007), it encapsulates
418 strong hypotheses relating to the nature of IV that are rarely discussed. In contrast, we argue here
419 that IV is generally non-random and structured in both space and time.

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421 At a given time, conspecific individuals that are distributed across space can strongly vary in
422 their attributes (Violle et al. 2012, Siefert et al. 2015, Moran et al. 2016, Poorter et al. 2018).
423 While this spatial IV has often been interpreted as random, *i.e.* implying that conspecific
424 individuals can perform differently within the same environment (Fig. 3c), a large part of this
425 variability appeared in fact structured in space and likely associated with fine-scale spatial
426 changes in the environment (Fig. 3b, Moran et al. 2016). In our illustrative simulation
427 experiment, the attribute of conspecific individuals varies spatially as a result of the
428 environmental variation in space, and the spatial autocorrelation of conspecific attributes reflects
429 the spatial autocorrelation of the environmental variables (Fig. 5, Fig. 6).

430

431 Similarly, data from three long-term forest inventory sites across the tropics revealed spatial
432 autocorrelation in tree diameter growth of conspecific individuals (Table 3), suggesting that IV is
433 strongly driven by the spatial variation of the environment, which is itself highly structured (Fig.
434 1). However, we acknowledge that genetically-driven IV can also be spatially structured, for
435 instance via dispersal patterns or natural selection (Moran et al. 2016). We hypothesize that in
436 that case, attributes would likely be randomly structured in space (Getzin et al. 2014) or
437 correlated at the spatial scale of seed dispersal, typically several tens of meters in tropical forests
438 (Clark et al. 2004, Seidler and Plotkin 2006, Muller-Landau et al. 2008), while environmental
439 variables are typically highly spatially correlated at fine scales (*e.g.* meter scale, Baraloto and
440 Coutron 2010). We also acknowledge that natural selection can happen at fine scales (Marrot et
441 al. 2021), and could thus produce spatially structured IV due to local genetic adaptation.
442 Nevertheless, data documenting genetic variation within species can still reveal higher similarity

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443 between conspecifics than heterospecifics locally as well as non-overlapping species niches
444 (Schmitt et al. 2021). Importantly, any local genetic adaptation does not preclude that
445 multidimensional environmental variations generate large observed IV that is structured in space
446 and time and whose consequences cannot be well represented and understood using a random
447 variation around a species mean.

448

449 In communities of sessile organisms such as trees, IV has been commonly structured in space
450 using individual random effects, which vary among conspecific individuals but stay constant
451 through the lifetime of individuals (Clark et al. 2007, Vieilledent et al. 2010). We here argue that
452 while this approach can *reveal* the spatial structure of IV through inference, the use of the
453 resulting estimated standard deviation term to introduce individual variation in simulations of
454 community dynamics is not sufficient to *produce* a spatially structured IV, as we showed is
455 observed in natural communities.

456 Similarly, individual attributes can change over time. Because individuals within a species can
457 be measured at different points in time, as it is often the case when assembling functional trait
458 databases for example (Zanne et al. 2009, Albert et al. 2011, Kattge et al. 2020) this can lead to
459 an observed unstructured IV when characterized by a variance around a species mean (Fig. 3c).
460 But a large part of this observed IV is actually structured in time and associated with temporal
461 changes in the environment. For instance, the temporal storage effect (Chesson and Warner
462 1981), a well-known coexistence mechanism, structures species performance because species are
463 able to “store” growth during favorable timespans to overcome lean times; mast-seeding or
464 masting, which describes periodic and synchronized massive seed production of conspecific
465 individuals, would also result in a temporally structured IV (Koenig and Knops 2005). Temporal

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466 variation in individual response within a species can typically be structured with temporal
467 random effects (Clark et al. 2007). Temporal random effects have been used to estimate the
468 inter-annual variability in tree growth (Metcalf et al. 2009, Fortunel et al. 2018) and fecundity
469 (Clark et al. 2007) for example. In all those examples, temporal environmental variation affects
470 conspecific attributes in the same way (Clark 2010).

471

472 We therefore call for a reconsideration of the nature and the way of integrating IV into models of
473 community dynamics. When IV is modeled randomly with a variance around a species mean, it
474 implies that conspecific individuals can perform differently in the exact same environment, thus
475 implying intrinsic differences between conspecific individuals. This type of unstructured IV can
476 result in an overestimated increase in species niche overlap, which blurs species differences (Fig.
477 3a and 3c, Stump et al. 2021). While trait heritability has rarely been considered in studies on the
478 role of IV on coexistence (but see Barabás and D'Andrea 2016), in some studies, the random
479 variation in attributes across conspecific individuals is considered as environmental, because it is
480 not heritable in the model (*e.g.* Lichstein et al. 2007, Moran et al. 2016). However,
481 environmentally-driven IV should be structured in space and time, as the environment is (Fig. 1,
482 Fig. 6). In addition, when IV is randomly distributed among conspecific individuals, similarity
483 among conspecific individuals is systematically underestimated, which is not the case when IV is
484 structured in space and time (Purves and Vanderwel 2014, Banitz 2019), as discussed hereafter.

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486 **Conspecific individuals respond more similarly than heterospecific individuals**
487 **locally**

488 Species differ in multiple attributes, responding to a high number of environmental variables
489 (Fig. 2), but often in ways that cannot be readily observed. If observed IV results mainly from
490 high-dimensional environmental variation in space and time rather than from intrinsic
491 differences between conspecific individuals, then for a given environment, conspecific
492 individuals should respond more similarly than heterospecific individuals. This is the case in our
493 illustrative simulation experiment, where the fact that conspecific individuals have exactly the
494 same set of parameters and respond identically to spatial and temporal changes in the
495 environment results in higher inter- than intraspecific variance in the response locally (Fig. 5,
496 Fig. 6).

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498 Corroborating this point of view, pairs of conspecific individuals in 11 North-American
499 temperate forest stands showed higher correlation in their temporal variation of growth rate or
500 fecundity than pairs of heterospecific individuals on average (Clark 2010). This indicates that
501 conspecific individuals responded more similarly to environmental variation in time than
502 individuals of different species. Importantly, these results were obtained in a system with high
503 observed IV (leading to an apparent species niche overlap), where species responded in the same
504 direction to environmental changes (*e.g.* increased tree growth in climatically favorable years).
505 Hence, considering the temporal structure of IV revealed species differences that were not
506 apparent otherwise, since they led to spreading along a high number of dimensions that varied at
507 fine scales (Clark 2010). However, as well highlighted by Stump et al. 2021, these results were
508 often misinterpreted as an evidence that IV fostered coexistence. As another piece of evidence

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509 presented here, pairs of spatially proximal conspecific individuals tended to present more similar
510 temporal means in absolute tree growth than pairs of close heterospecific individuals across three
511 large contrasted tropical forest sites (Table 5). This provides new empirical evidence that,
512 although estimated IV can be substantial, conspecific individuals respond more similarly than
513 heterospecific individuals to environmental variation in space.

514

515 A stronger similarity in the response to environment between conspecific than heterospecific
516 individuals locally leads to a stronger concentration of competition within species, which,
517 ultimately, can result in intraspecific competition being greater than interspecific competition, a
518 common driver of stable species coexistence (Lotka 1925, Volterra 1926, Chesson 2000).
519 Because species differ in their response to the environment, environmental variation in space and
520 time leads to local or punctual inversions of species hierarchy in performance (Fig. 3d). As
521 possibilities of hierarchy inversions between species increase rapidly with increasing
522 dimensionality of the environment (Fig. 3b), the high-dimensionality of the environment offers
523 room for the stable coexistence of numerous species (Falster et al. 2017, R uger et al. 2018). In
524 the end, we therefore argue that a substantial part of IV is not a mechanism for coexistence in
525 itself but can rather be the signature of species differences and environmental variation that allow
526 coexistence: the high-dimensional species differences, which make them respond differently in a
527 high-dimensional environment varying in space and time, can only be observed at the individual
528 scale. In the absence of precise information on the many dimensions across which species differ
529 and environment varies, large observed IV is the evidence of the niche mechanisms enabling
530 species coexistence.

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532 **Recommendations and concluding remarks**

533 Most of the theoretical studies that have explored the role of IV in species coexistence so far did
534 so by adding variances around species-specific means, thus considering IV as stochastic, which
535 implies that conspecific individuals perform differently in the same environment. Here, we
536 provide insights suggesting that large observed IV can emerge from environmental heterogeneity
537 and is structured in space and time. We stress that this interpretation has strong consequences on
538 the understanding of the effects of IV on species coexistence: (i) observed IV does not
539 necessarily imply that conspecific individuals are strongly intrinsically different nor that species
540 niches overlap, and (ii) the spatial and temporal structure of observed IV reveals stronger
541 concentration of competition within species locally in space and time, which is a frequent
542 necessary condition for stable species coexistence. We thus call for a reconsideration of the
543 nature of IV and of the way it is integrated in models, by thoroughly distinguishing its sources
544 (intrinsic vs. extrinsic, and their interactions). We acknowledge the existence of genetically-
545 driven IV, potentially due to local adaptation to the microenvironment, and its eco-evolutionary
546 importance, but suggest that multidimensional environmental variation generates a large
547 observed IV that is structured in space and time. We underline that environmentally-driven
548 structured IV has been largely overlooked in previous community ecology studies and has
549 consequences on community dynamics which cannot be represented and understood using a
550 random variation around a species mean. To this end, we recommend that empirical studies
551 explore further the spatio-temporal structure of IV and how it relates to environmental variation
552 along multiple dimensions, and, when possible, assess the relative importance of genetically and
553 environmentally driven IV, for instance by means of common garden experiments. Models of
554 community dynamics should then endeavor to structure IV in space and time so that it reflects

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555 the high-dimensional variation in both the environment and species attributes, and not only some
556 intrinsic differences between conspecific individuals (Purves and Vanderwel 2014, Moran et al.
557 2016, Banitz 2019). In both empirical studies and models, this implies that the species attributes
558 are measured at the individual level, localized in space, and repeatedly observed in time.
559 Simultaneously, the monitoring of multiple environmental variables at fine scales in space and
560 time is required in order to better capture their effect on individual attributes (such as
561 physiological or mechanistic traits, Shipley et al. 2016, Brodribb 2017), hence reducing the part
562 of unexplained IV, and ultimately to better characterize the high-dimensionality of species
563 niches. Altogether, these recommendations will enable to better account for species differences
564 that are expressed at the individual level and evidence their impacts on the community dynamics
565 in natura and in silico.

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578

579 **Supplementary information and data access**

580 The appendices and all the code used for this study are available in a GitHub repository
581 (<https://github.com/camillegirardtercieux/coexIV>) and have been permanently archived on
582 Zenodo (<https://doi.org/10.5281/zenodo.5504013>).

583

584 Appendix 1: Simulation experiment with two species.

585 Appendix 2: Analysis of an *Eucalyptus* clonal plantation dataset.

586 Appendix 3: Analysis of tropical forest inventory data.

587

588 No new data were used in this study. For access to forest plot inventory data and *Eucalyptus*

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589 plantation data used in this study, refer to the data used in Le Bec et al. 2015, Hérault and
590 Piponiot 2018, Condit et al. 2019 and le Maire et al. 2019. However, the analyses and reflections
591 presented here are original.

592

593 **Statements of author roles:**

594 *CGT, IM and GV conceived the initial ideas and coordinated the INTRACO working group. All*
595 *authors contributed to the study design and ideas within the INTRACO working group. CGT led*
596 *the analyses. CGT, IM and GV wrote the first draft of the manuscript, and all authors*
597 *contributed substantially to revisions.*

598

599 **Declaration of Interests**

600 All authors declare that they have no conflict of interest.

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825 **Table 1: Mean posteriors of the *Eucalyptus* model and their estimation errors and residual**
 826 **variance partitioning among the different random effects.**

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	Intercept	Diameter	Competition	Individual	Block	Genetic	Temporal	Residual
	(β_0)	(β_1)	(β_2)	variance	variance	variance	variance	variance
				(V_i)	(V_b)	(V_g)	(V_t)	(V)
Estimate	-3.70E-02	5.50E-01	-2.70E-01	2.30E-01	5.40E-02	1.30E-01	1.20E+00	5.10E-01
Estimation error	4.60E-01	5.10E-03	8.90E-03	4.10E-03	1.50E-02	2.90E-02	5.70E-01	2.00E-03
Percentage of unexplained variance				10.83%	2.54%	6.12%	56.50%	24.01%

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829 **Table 2: Features of the three tropical forest data sets used as empirical case studies.**

Site	Rainfall (mm.y ⁻¹)	Sampling	Mean DBH	Nb of census	Periodicity	Disturbance	Topography	Nb of species	Nb of individuals	Data source
Paracou, French Guiana	3,000	15 × 10 ha (incl. 12 logged plots)	10 cm	24	1-2 y since 1992	Natural disturbances + selective logging	flat	613	69,548	(Hérault and Piponi 2018)
BCI, Panama	2,600	50 ha	10 cm	8	5-y since 1980	Natural disturbances	hilly	225	37,224	(Condit et al. 2019)
Uppangala, India	5,100	5,92 ha (4 transects and 3 plots)	9.5 cm	20	1-yr since 1992	Natural disturbances	mountainous	102	3,789	(Le Bec et al. 2015)

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830 **Table 3: Mean posteriors of the tropical forest model and their estimation errors and**
 831 **residual variance partitioning among the different random effects.**

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	Intercept (β_0)	Diameter (β_1)	Species variance (V_b)	Individual variance (V_d)	Residual variance (V)
Paracou					
Estimate	6.70E-02	2.90E-02	4.70E-01	4.60E-01	7.60E-01
Estimation error	2.20E-02	3.80E-03	1.70E-02	3.90E-03	2.30E-03
% Variance			27.81%	27.22%	44.97%
Uppangala					
Estimate	8.40E-02	1.90E-01	3.60E-01	6.60E-01	5.90E-01
Estimation error	4.40E-02	1.20E-02	4.30E-02	8.60E-03	1.90E-03
% Variance			22.36%	40.99%	36.65%
BCI					
Estimate	1.90E-01	-2.20E-02	6.60E-01	4.10E-01	8.10E-01
Estimation error	5.00E-02	4.50E-03	3.50E-02	4.00E-03	2.00E-03
% Variance			35.11%	21.81%	43.09%

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834 **Table 4: Spatial autocorrelation of the growth of conspecific individuals in three tropical**
835 **forest sites.** Shown are the proportion of species, and of corresponding individuals, in percent,
836 for which individual growth among conspecific individuals is significantly positively spatially
837 autocorrelated. The spatial autocorrelation of individual growth was tested using Moran's I
838 index.

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	Significant	Not significant
Paracou		
% Species	31.00	69.00
% Individuals	78.90	21.10
Uppangala		
% Species	18.50	81.50
% Individuals	45.30	54.70
BCI		
% Species	20.10	79.90
% Individuals	54.70	45.30

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841 **Table 5: Comparison of local intra- and interspecific variability in individual growth for**
 842 **three tropical forest sites.** The variability was estimated with the semivariance and the
 843 comparison was performed with a Mann-Whitney’s test. The semivariances were computed for
 844 all species with > 5 individuals and > 5 heterospecific neighbors within 100 m in the same plot,
 845 and considering pairs of individuals that were less than 100 m apart and in the same plot. Shown
 846 are the proportion of species, and of corresponding individuals, for which (i) intraspecific
 847 variability was significantly lower than interspecific variability, (ii) intraspecific variability was
 848 significantly higher than interspecific variability, or (iii) the difference between inter- and
 849 intraspecific variabilities was not significant.

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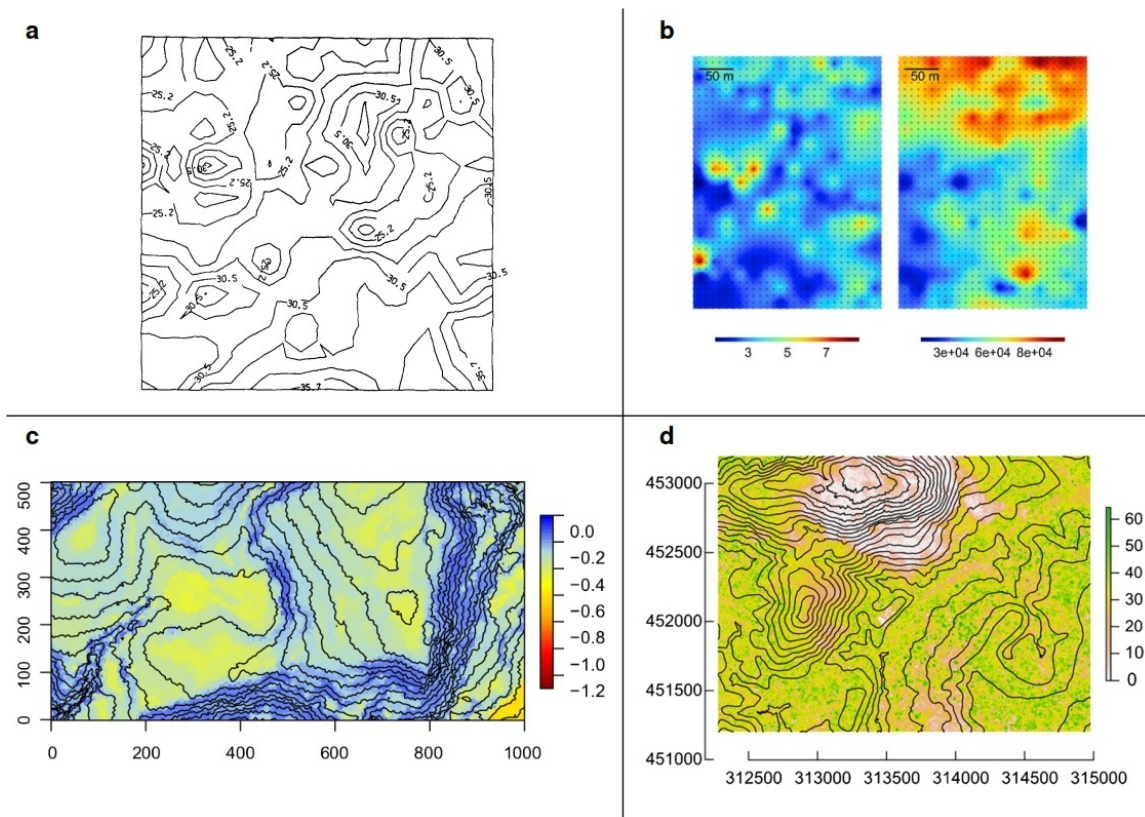
	Intraspecific variability < Interspecific variability (i)	Intraspecific variability ~ Interspecific variability (ii)	Intraspecific variability > Interspecific variability (iii)
Paracou			
% Species	60.70	40.70	0.67
% Individuals	88.80	10.90	0.28
Uppangala			
% Species	42.20	62.20	4.44
% Individuals	57.70	23.60	18.80
BCI			
% Species	46.10	47.80	3.14
% Individuals	76.00	19.30	4.69

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852 **Figure 1: High environmental variability at a small spatial scale. (a)** Soil nitrogen content in
853 a 12×12 m plot at Cedar Creek in g.kg⁻¹, (USA), Tilman 1982; **(b)** Carbon in % (left) and
854 aluminum in ppm (right) soil content in a 12-ha (250×500 m) plot at The Nouragues (French
855 Guiana), Zinger et al. 2019; **(c)** Soil water content during mid-dry season of a regular year in
856 MPa in a 50-ha (1000×500 m) forest plot at Barro Colorado Island (Panama), Kupers et al. 2019.
857 Coordinates in m.; **(d)** Canopy height in m and topography (10 m spaced elevation lines) in a 50-
858 ha (2500×2000 m) area at the Nouragues Research Field Station, Tymen et al. 2017. Coordinates
859 in m (UTM 22N).

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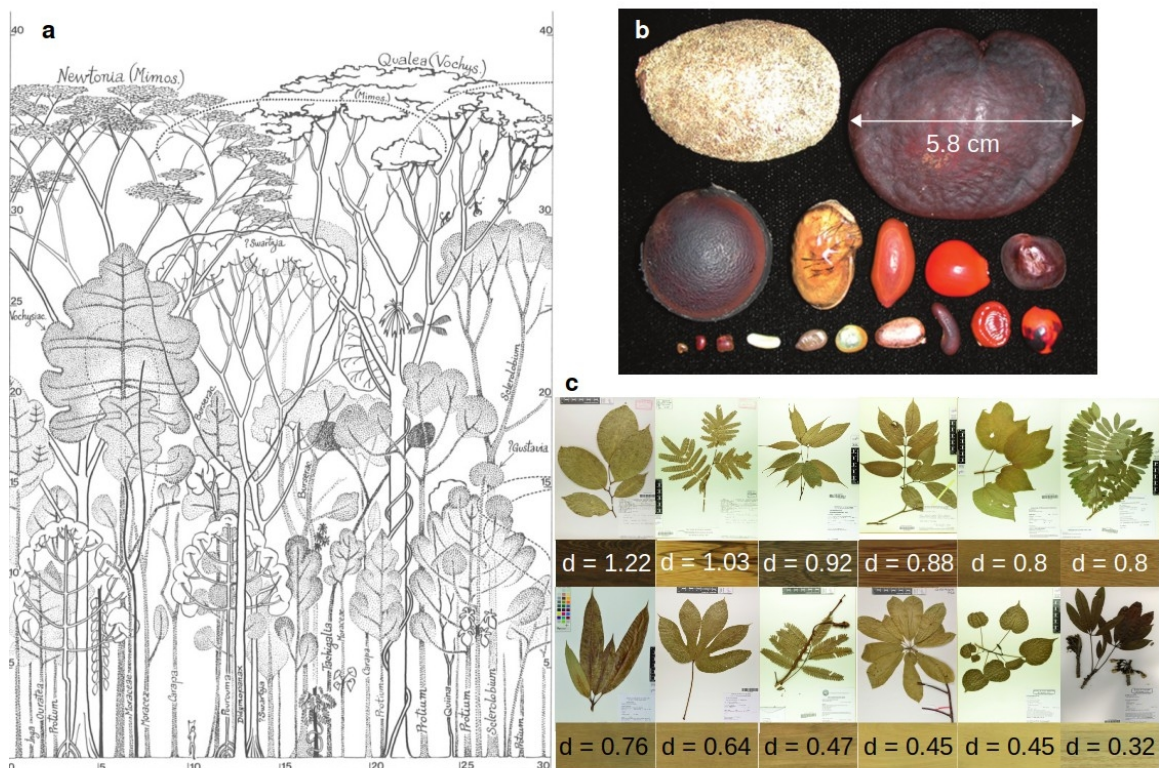
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861 **Figure 2: Morphological diversity of tree species illustrating strong differences between**
 862 **species. (a)** Diversity of tree species architecture and height in a tropical forest (Hallé et al.
 863 1978). Coordinates are in m.; **(b)** Diversity of seed size and shape from 17 tree species of the
 864 Fabaceae family in the Peruvian Amazon (Muller-Landau 2003); **(c)** Diversity of leaf size and
 865 shape (herbarium of Cayenne, Gonzalez et al. 2021) and of wood aspect (reflecting wood
 866 characteristics) and density (Normand et al. 2017) for 12 tree species in French Guiana. Species
 867 from top left to bottom right are *Bocoa prouacensis*, *Zygia racemosa*, *Vouacapoua americana*,
 868 *Eperua falcata*, *Bagassa guianensis*, *Hymenolobium excelsum*, *Mangifera indica*, *Sterculia*
 869 *pruriens*, *Parkia nitida*, *Couroupita guianensis*, *Hura crepitans*, and *Ceiba pentandra*. Black
 870 bars next to herbarium samples indicate the scale (10 cm).

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872 **Figure 3: Reinterpreting observed intraspecific variability (IV): from niche widening to**
873 **niche projection into a high-dimensional environment.** In **(a)**, within a given environment E_0
874 defined along an environmental axis X_1 ($E_0=E(X_1,0)$), conspecific individuals are identical and
875 have the same performance p_A and p_B , for species A (blue) and species B (orange). Species A
876 outcompetes species B in E_0 . Actual measured differences among conspecific individuals,
877 shown in **(c)**, can be interpreted in different ways. First, as conspecific individuals exhibit
878 contrasting attributes in E_0 , they become more different. This can result in some heterospecific
879 individuals having similar performances: IV would blur species differences. Alternatively, IV
880 measured in E_0 results from the variation of unobserved environmental variables ($E_0=E(X_1,0,$
881 $X_2)$; **(b)**). Contrasting performances among conspecific individuals in E_0 do not result from
882 intrinsic differences among them but from differences in the local environment they experience
883 and that was poorly characterized, *i.e.* the number of observed dimensions is lower than the
884 actual number of environmental dimensions. Similarly, although species niches present some
885 overlap when projected on one dimension **(d)**, they do not overlap in the two-dimensional space
886 **(b)**. Moreover, while species A outcompetes species B on average when $X_1=X_{1,0}$, the opposite
887 occurs when $X_1=X_{1,1}$ **(d)**, leading to an inversion of species hierarchy between different
888 environments. Similarly, while species A outcompetes species B in $E(X_1,0, X_2,0)$, the opposite
889 occurs in $E(X_1,0, X_2,1)$. Although only two dimensions are shown, species respond to many
890 environmental variables varying in space and time, multiplying the possibilities of niche
891 segregation and hierarchy inversions between species, offering room for species coexistence in a
892 variable high-dimensional environment. The code used to generate this figure is available online.

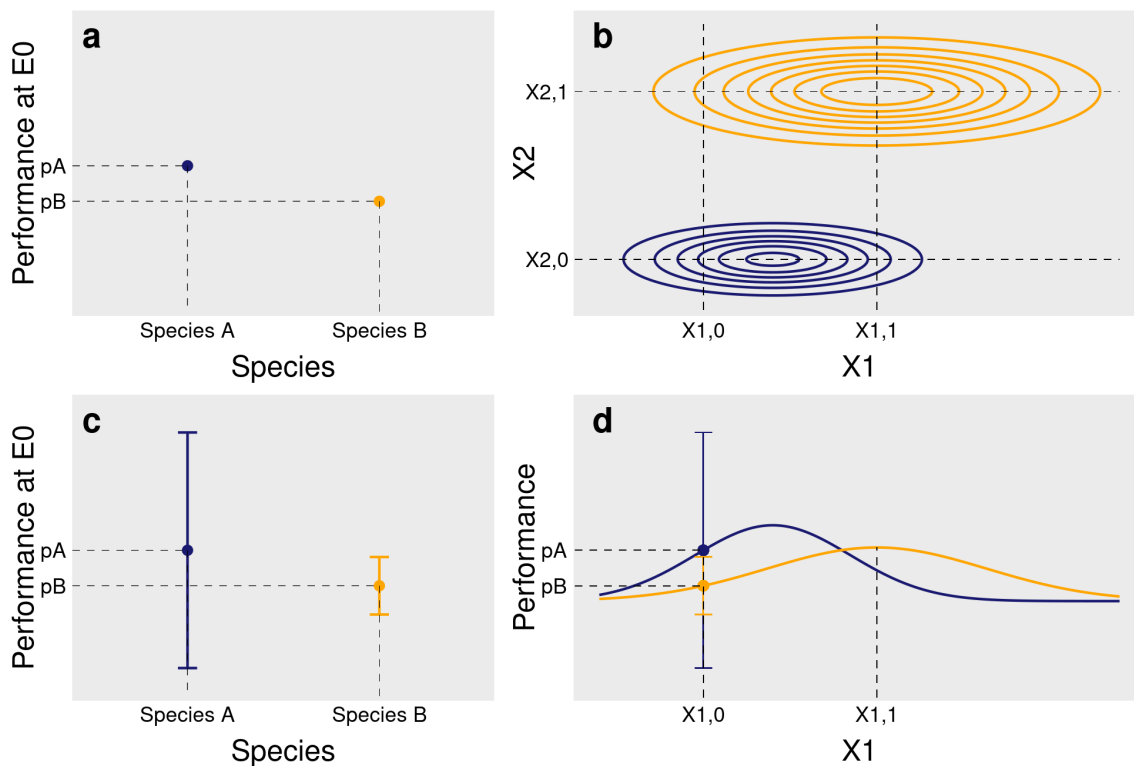
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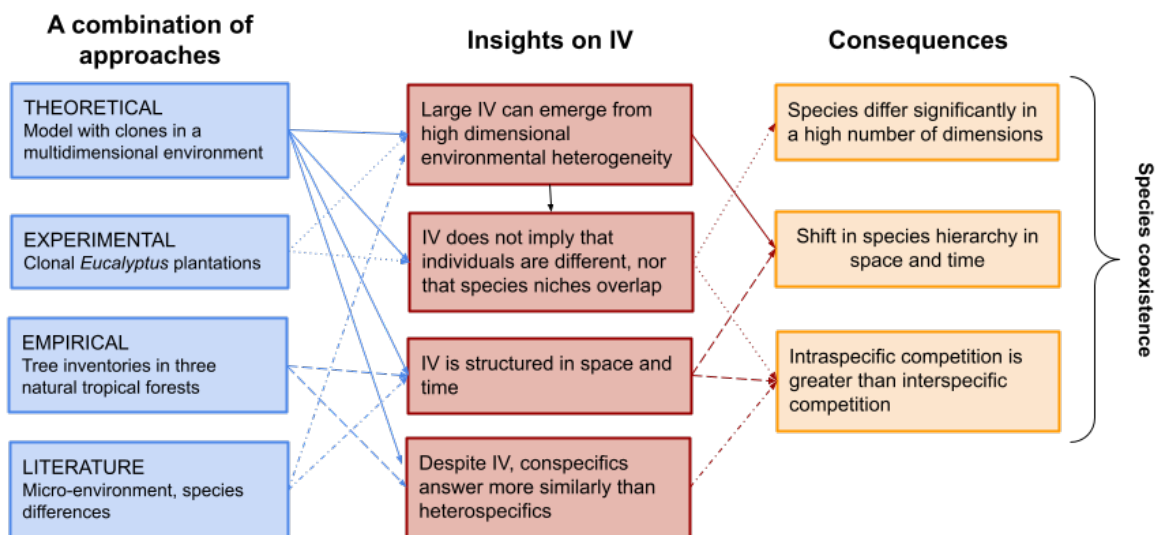


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895 **Figure 4: Multiple insights on the nature of IV and its consequences on individual and**
896 **species differences.** We used literature and data analyses of various nature to support the
897 hypothesis that a large part of observed IV can result from multidimensional environmental
898 variations that are spatially and temporally structured rather than by intrinsic and spatio-
899 temporally unstructured differences between conspecific individuals, with radically different
900 consequences on species coexistence.



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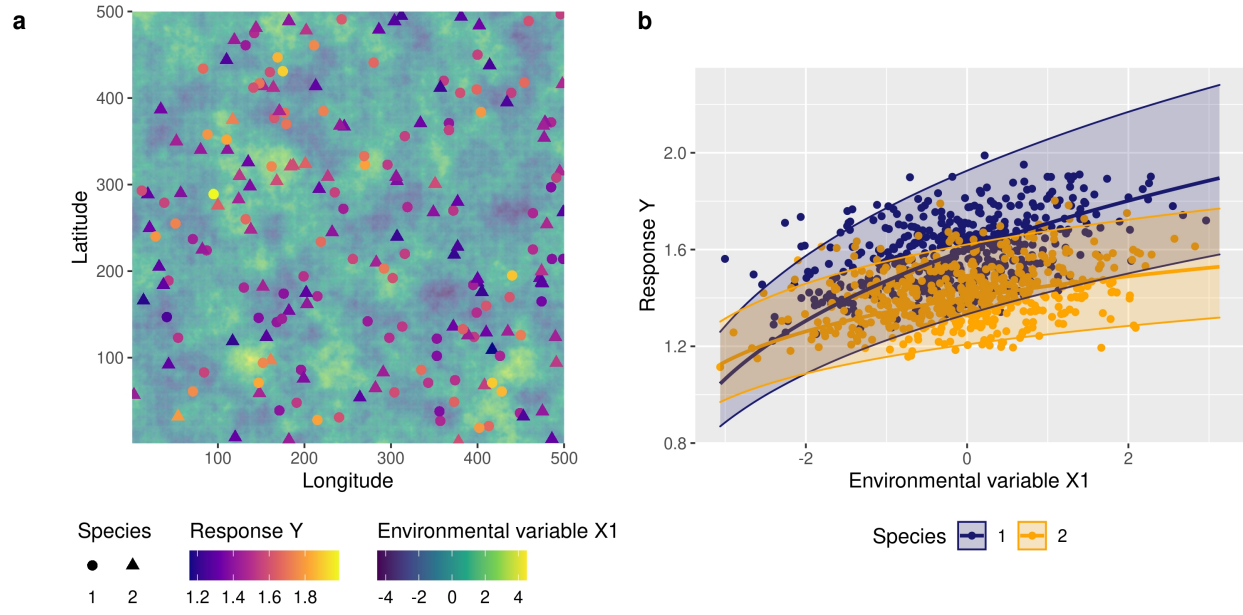
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902 **Figure 5: Observed intraspecific variability as a result of the imperfect characterization of**
903 **the environment.** A simulated response variable (Y , *e.g.* growth) is generated for individual
904 clones of two species thriving in a high-dimensional environment. This response variable was
905 first computed as a function of ten environmental variables (“perfect knowledge” model, Eq. I),
906 but is then analyzed using a statistical model that accounts for the unique environmental variable
907 that was assumed to be observed in the field (X_1 , *e.g.* light) and includes a random individual
908 effect (“imperfect knowledge” model, Eq. II). The intraspecific variability estimated with these
909 random individual effects is then due to the variation in space and time of the nine unobserved
910 environmental variables. **(a)** Positions of a sample of $I=600$ individuals from $J=2$ species in a
911 landscape defined by a square grid of $C \times C$ cells ($C=500$). The background color indicates the
912 value of the observed environmental variable X_1 on each cell at date t . The response Y of each
913 individual, which depends on the environment within each cell (Eq. I), is also indicated by a
914 color scale. **(b)** Response Y as a function of the observed environmental variable X_1 for the two
915 species. Points represent the data $\{Y_{ijt}, X_{1,ijt}\}$. Thick lines represent the predictive posterior means
916 for the two species. The envelopes delimited by two thin lines represent the 95% credible
917 intervals of the predictive posterior marginalized over individuals (taking into account \hat{V}_{bj}). The
918 envelopes thus represent the intraspecific variability which is due to the $N-1$ unobserved
919 environmental variables.

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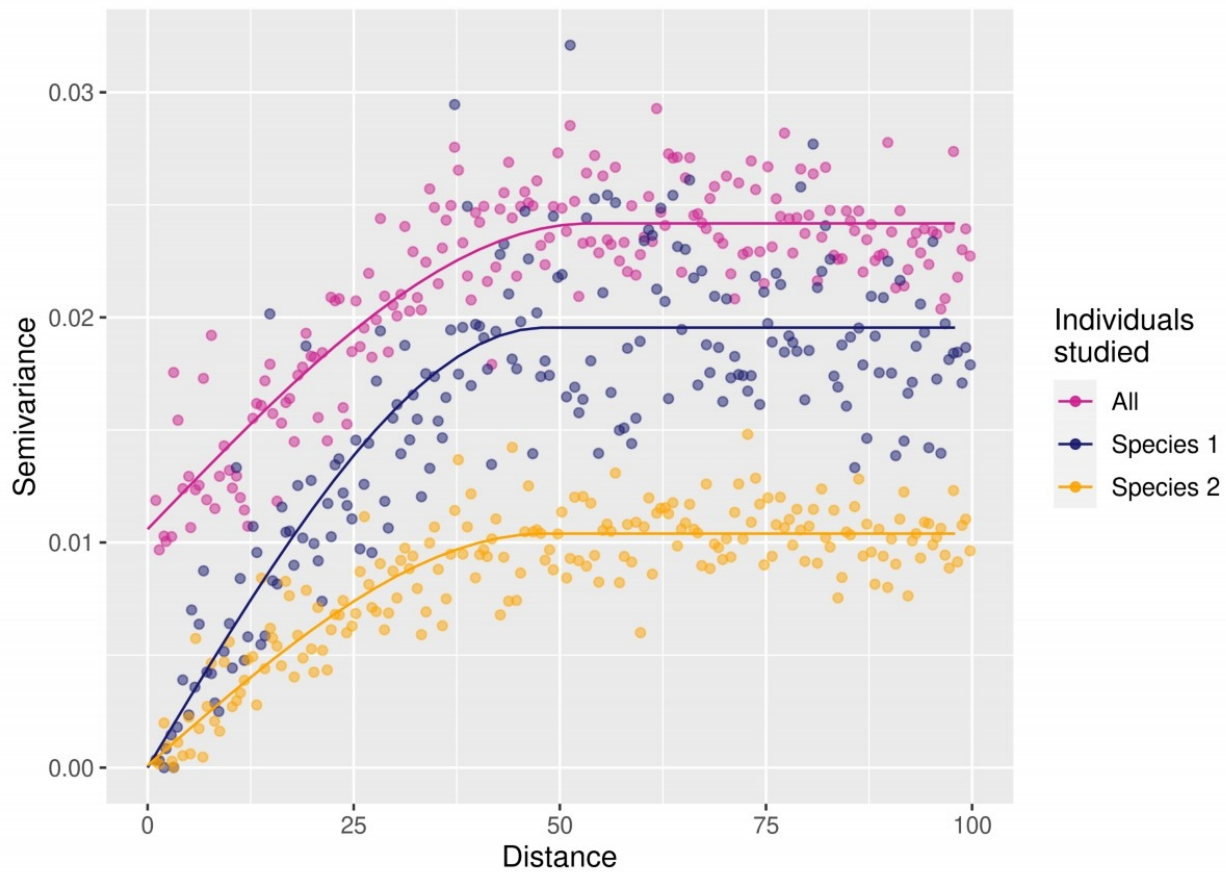


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921 **Figure 6: Spatial autocorrelation of attribute Y across individuals within and between**
922 **species (J=2) in a simulation experiment.** This semivariogram represents the semivariance of
923 the individual mean attribute Y as a function of the distance between individuals. The increasing
924 curves evidence spatial autocorrelation in Y (similar results using Moran's I test). The
925 semivariance of all individuals taken together (purple curve) is higher than the semivariance of
926 conspecific individuals for the two species (orange and blue curves), which means that
927 intraspecific variability is lower than interspecific variability.

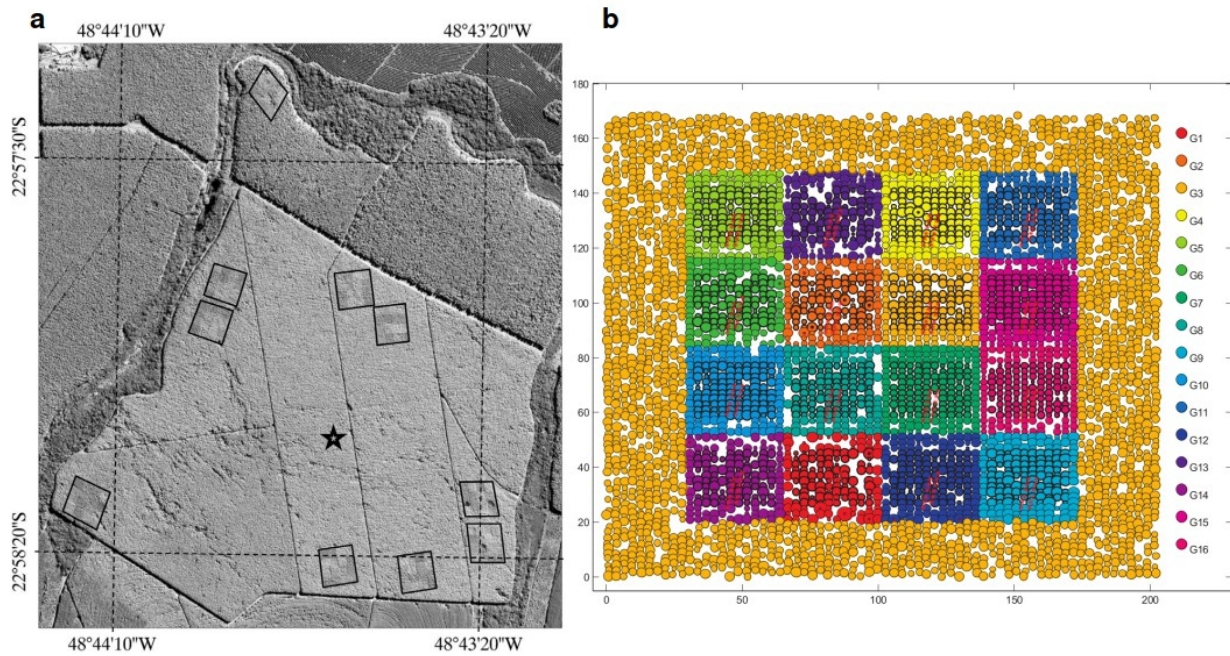


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929 **Figure 7: Experimental setup of the EUCFLUX experiment.** The ten blocks **(a)** and the
930 organization of the 16 genotypes within a block **(b)**. In our analyses, two genotypes were
931 discarded because they were obtained from seeds and not clones and therefore included some
932 genetic variability. A more complete figure legend can be found in le Maire et al. 2019.



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