1	Running Head: Drivers of soil nematode diversity
2	
3	Title: Drivers of nematode diversity in forest soils across climatic zones
4	
5	Authors Yuanhu Shao ¹ , Zuyan Wang ^{2,3} , Tao Liu ² , Paul Kardol ⁴ , Chengen Ma ⁵ , Yonghong
6	Hu ⁶ , Yang Cui ¹ , Cancan Zhao ⁷ , Weixin Zhang ¹ , Dali Guo ^{5†} , Shenglei Fu ¹ *
7	
8	Affiliations
9	1. Key Laboratory of Geospatial Technology for the Middle and Lower Yellow River Regions,
10	Ministry of Education, College of Geography and Environmental Science, Henan University,
11	Kaifeng 475004, China.
12	2. Key Laboratory of Vegetation Restoration and Management of Degraded Ecosystems,
13	South China Botanical Garden, Chinese Academy of Sciences, Guangzhou 510650, China.
14	3. University of Chinese Academy of Sciences, Beijing 100049, China.
15	4. Department of Forest Ecology and Management, Swedish University of Agricultural
16	Sciences, 907 51, Ume å, Sweden.
17	5. Center of Forest Ecosystem Studies and Qianyanzhou Station, Key Laboratory of
18	Ecosystem Network Observation and Modeling, Institute of Geographic Sciences and Natural
19	Resources Research, Chinese Academy of Sciences, Beijing 100101, China.
20	6. Key Laboratory of Digital Earth Science, Aerospace Information Research Institute,
21	Chinese Academy of Sciences, Beijing 100094, China.
22	7. International Joint Research Laboratory for Global Change Ecology, School of Life
23	Sciences, Henan University, Kaifeng, Henan 475004, China.
24	
25	* Corresponding authors: Shenglei Fu (email: fsl@henu.edu.cn; Phone: 86-0371-23881856). 1

[†]Deceased.

28 Abstract

Nematodes are the most abundant multi-cellular animals in soil, influencing key processes and functions in terrestrial ecosystems. Yet, little is known about the drivers of nematode abundance and diversity in forest soils across climatic zones. This is despite forests cover approximately 30% of the earth's land surface, provide many crucial ecosystem services but strongly vary in hydrothermal conditions and associated ecosystem properties across climatic zones.

Here, we collected nematode samples from 13 forests across a latitudinal gradient. We 35 divided this gradient in temperate, warm-temperate, and tropical climatic zones. Using boosted 36 regression trees, we showed that across the gradient, nematode abundance and diversity were 37 mainly influenced by soil organic carbon. However, within climatic zones, other factors were 38 more important in driving nematode alpha-diversity, nematode biomass and the abundance of 39 different trophic groups: mean annual temperature and total soil phosphorus in temperate 40 zones, soil pH in warm-temperate zones, and mean annual precipitation in tropical zones. 41 Additionally, nematode beta-diversity was higher in temperate than in warm-temperate and 42 tropical zones, and we did not find significant differences among climatic zones in nematode 43 gamma diversity. 44

Together, our findings indicate a latitudinal shift in the main climatic variables controlling soil nematode communities and demonstrate that the drivers of soil nematode diversity in forested ecosystems are affected by the spatial scale and climatic conditions considered. This implies that high resolution studies are needed to accurately predict how soil functions respond if climate conditions move beyond the coping range of soil organisms. Our results also

²⁷

emphasize the importance of studying the area-diversity relationship of soil organisms under
different climatic conditions.

52

Key words: biogeographical patterns; climate; nematode trophic groups; plant diversity; soil
nematodes; soil properties; spatial scale

55

56 Introduction

Soil biodiversity accounts for a large fraction of the biodiversity on Earth and is key for the 57 58 functioning of terrestrial ecosystems (Bardgett and Wardle 2010, Bardgett and van der Putten 2014, Wagg et al. 2014), such as water infiltration, regulation of pests and pathogens, food 59 security, and the release of nutrients from soil organic matter (Wall et al. 2015). Energy, water, 60 and latitude determine the alpha-diversity of above-ground organisms (plants and animals) at 61 a global scale (Hawkins et al. 2003). However, understanding the controlling factors of soil 62 biodiversity is still a major challenge for ecologists (Decaëns 2010, van den Hoogen et al. 63 2019). For instance, a recent study found that variation in earthworm diversity was mainly 64 associated with changes in soil pH and soil carbon at small spatial scales, while precipitation 65 and temperature were more important at a global scale (Phillips et al. 2019). Furthermore, 66 plants and soil organisms are tightly linked through multiple direct and indirect pathways 67 (Wardle et al. 2004). Therefore, the uncertainty of the diversity and community composition of 68 below-ground organisms is strongly tied to the spatial scale of study due to variation in 69 climatic, soil, and vegetation properties (Bardgett and van der Putten 2014). 70

Nematodes are the most abundant multi-cellular animals in soil (Bongers and Ferris 1999) and are a highly diverse group of invertebrates (Hugot et al. 2001). It is estimated that 4.4×10^{20} nematodes inhabit surface soils across the world, with higher abundances in sub-Arctic regions than in temperate or tropical regions, and that the amount of carbon respired by soil

nematodes is equivalent to roughly 15% of carbon emissions from fossil fuel use, or around 75 76 2.2% of the total carbon emissions from soils (van den Hoogen et al. 2019). A study on the 77 community structure of soil nematodes that included different ecosystem types did not find a good predictor of family diversity at the plot level, but nematode family richness was related 78 to latitude at a global scale (Nielsen et al. 2014). Other studies showed that at regional scales, 79 soil nematode community composition was mainly influenced by climatic conditions, that is, 80 81 precipitation in grassland ecosystems (Chen et al. 2015, 2016; Xiong et al. 2020) and temperature in coastal ecosystems (Wu et al. 2016). Further, a recent study from forest 82 83 ecosystems suggested that climate factors were the drivers of nematode community composition at the regional scale, while terrain and soil characteristics were the drivers of 84 nematode community composition at local scales (Xiao et al. 2021). However, the study by 85 Xiao et al. (2021) was based on a limited number of observations (i.e., sampling at three points 86 only). So far, few studies have focused on soil nematode communities in forested ecosystems 87 along a hydrothermal gradient. Notably, no studies have comprehensively assessed nematode 88 diversity or community composition and their drivers in forest soils across climatic zones using 89 standardised sampling approaches. Understanding the determinants of biodiversity in forest 90 soils is however critical to better predict how forest ecosystems will respond to multiple 91 92 environmental changes.

In the present study, we collected nematodes in forest soils from various climatic zones (i.e., temperate, warm-temperate, and tropical) to explore soil nematode distribution patterns and diversity, as well as their drivers, using boosted regression trees models (Elith et al. 2008). Boosted regression tree (BRT) models draw on insights and techniques from both statistical and machine learning methods and are a powerful alternative to other models for both explanation and prediction (Elith et al. 2008). Given that the basal resources of soil organisms are provided by plant roots, root exudates, and litter (Coleman et al. 2004), and given the slower

litter decomposition rates at higher latitudes (Meentemeyer 1978) and faster root turnover rates 100 at lower latitudes (Gill and Jackson 2000), we hypothesized that the drivers of soil nematode 101 composition and diversity would vary across climatic zones. For instance, faster root turnover 102 rates in tropical forest soils may provide more abundant food sources for plant-feeding 103 nematodes. We also hypothesized that the soil nematode communities in temperate and tropical 104 forest soils would be particularly sensitive to climatic factors. For instance, global warming 105 106 may lead to an increase in nematode diversity at higher latitudes and a decrease in nematode diversity at lower latitudes because warmer temperatures would be closer to their physiological 107 108 optima at higher latitudes and closer to their thermal limits at lower latitudes (Zhao et al. 2021).

109

110 Material and Methods

111

112 Sampling locations

A regional-scale field investigation was conducted to collect soil nematode samples from seven 113 nature reserves, including 13 forest sites in China across a latitudinal gradient ranging from 114 21 36' N to 47 11' N, covering all major zones of forest vegetation in eastern China. From 115 north to south, the seven nature reserves (and typical vegetation type) are: Liangshui (boreal 116 forest), Changbaishan (mixed coniferous-broad leaf forest), Donglingshan (deciduous broad-117 leaved forest), Jigongshan (deciduous broad-leaved forest), Tiantongshan (evergreen 118 119 broadleaved forest), Badagongshan (evergreen/deciduous broad-leaved mixed forest), and Xishuangbanna (tropical monsoon forest (Fig. 1). Generally, the northern hemisphere 120 temperate zone refers to the region between the Tropic of Cancer (approximately 23.5 °N) to 121 the Arctic Circle (approximately 66.5 °N) and the tropics is the region between the Tropic of 122 Cancer (approximately 23.5 ° N) and the Tropic of Capricorn (approximately 23.5 ° S) 123 (McColltoll 2005). Therefore, based on the climatic distribution of the nature reserves, we 124

divided the gradient into three climatic zone: the forests in Liangshui (47 °11' N) and Changbaishan (42 °22' N) were classified as temperate forests, the forests in Xishuangbanna (from 21 °36' N to 21 °57' N) were classified as tropical forests, and the other forests (from 29 °46' N to 39 °57' N) were classified as warm-temperate forests.

- 129
- 130 Soil sampling

131 In total, we selected 13 forest sites: one site in Donglingshan, Jigongshan, Tiantongshan, and Badagongshan, two sites in Liangshui and in Changbaishan, and five sites in Xishuangbanna. 132 133 This resulted in four sites in temperate forest, four sites in warm-temperate forests, and five sites in tropical forest. These forest sites were selected from larger areas of forest that are 134 representatives for the regions. The vegetation types were clearly different between any two 135 forest sites in the same nature reserve (Appendix S1: Table S4). Three 20×20 m sampling 136 plots, separated by more than 100 m, were randomly established in each forest site, resulting 137 in a total of 39 sampling plots. Sampling was conducted between August and October 2014, 138 during the growing season when productivity was highest. Five composite soil samples were 139 collected from each sampling plot. Specifically, in each plot, five subplots $(5 \times 5 \text{ m})$ were 140 established: one in each corner and one in the center. For each subplot, one composite sample 141 comprised of five soil cores was collected from the upper 0-10 cm of soil. Litter was removed 142 from the soil surface before soil samples were taken. Visible roots in the soil samples were 143 picked out by hand. As a result, we collected a total of 195 samples at the subplot level (i.e., 144 13 sites \times 3 sampling plots \times 5 composite soil samples). These samples were subsequently used 145 for analyses of physicochemical properties and soil nematodes (see below). 146

147

148 *Climate data*

6

For each site, the mean annual precipitation (MAP) and mean annual temperature (MAT) for the period 1970–2000 were obtained from the WorldClim database (http://www.worldclim.org/current) (Hijmans et al. 2005).

152

153 *Plant community and productivity data*

For each forest site in the tropical and temperate climatic zones, woody plant community data were obtained during one or two field surveys from 2004 to 2010; these data were provided by the nature reserves. Given the greater plant diversity and more diverse forest types in tropical forests, four 20×20 m plant community survey plots were randomly selected for each forest site in the tropical climatic zone while three 20×20 m plant community survey plots were randomly selected for each forest site in the temperate climatic zone. For the warm-temperate forests, plant community data were unfortunately unavailable.

We defined alpha-diversity of woody plant communities as the species richness of a single 161 20×20 m plot and gamma-diversity as the total richness of those plots at each forest site. Beta-162 diversity was defined as the species turnover among the plots at each forest site, and was 163 measured as Beta = 1-alpha_{mean}/gamma (Whittaker 1960; Kraft et al. 2011). We note that 164 diversity indices of woody plants for warm-temperate forests could not be calculated due to the 165 absence of data. The photosynthetic accumulation of carbon by plants per unit time, known as 166 gross primary productivity (GPP), was estimated from satellite images of the Moderate 167 Resolution Imaging Spectroradiometer (MODIS) according to an empirical light-use efficiency 168 (LUE) model with the following equation (Zhang et al. 2012): 169

170

171
$$GPP = LUE_{\max} \times f(VPD) \times g(T_{\min}) \times (1 - e^{-k \times LAI}) \times PAR$$

172

where LUE_{max} is the maximum light-use efficiency, f(VPD) is the scalar of daily vapor 173 pressure deficit, $g(T_{\min})$ is the scalar of daily minimum air temperature, e is the natural constant, 174 k is the empirical constant in calculating the probability of radiation transmission through the 175 canopy at zenith angle, PAR is the photosynthetically active radiation absorbed by the canopy, 176 and LAI is the leaf area index. All of these biome physical parameters are quantified based on 177 the MODIS land cover classification system using a biome property look-up table. Given that 178 179 the long-term accumulation of plant-derived carbon can also influence soil nematodes, we decided to use the dataset backwards 15 years from the date of soil sampling. Hereto, we 180 181 examined the gross primary productivity from 2000-2014 for each site by using the GPP parameters from the MOD17A2H product. This global GPP product estimates the cumulative 182 8-day composite of values with a 500 m spatial resolution based on an empirical light-use 183 efficiency model (Zhao et al. 2005, Zhao et al. 2010, Wang et al. 2017). In addition, to match 184 the time of soil sampling, we selected the GPP data from August to October of each year from 185 2000 to 2014. 186

187

188 Soil analysis

Total soil organic carbon, total soil nitrogen, total soil phosphorus, and soil pH were measured 189 using standard methods of the Chinese Ecosystem Research Network (Liu 1996). Briefly, soil 190 water content was measured gravimetrically by drying fresh soil at 105 °C for 48 h. Soil pH 191 192 was measured with a 1:2.5 (w/v) ratio of soil to deionized water using a pH meter (Mettler Toledo, Shanghai, China). Soil organic carbon was oxidized by a solution of 0.133 M K₂Cr₂O₇-193 18.4 M H₂SO₄ in an oil bath and then the excess K₂Cr₂O₇ was titrated with 0.2 M FeSO₄. Total 194 soil nitrogen content was measured using an ultraviolet spectrophotometer after Kjeldahl 195 digestion. Total soil phosphorus content was determined by H₂SO₄–HClO₄ fusion, followed by 196 the Mo-Sb anti-spectrophotography method. 197

Soil nematodes were extracted from a 30 g subsample of each composite soil sample using 198 a sugar centrifugation method (Coleman et al. 1999) for determination of nematode abundance, 199 biomass, and community composition. After fixation in 4% formalin solution, the nematodes 200 were counted under an inverted microscope, and for each sample the first 100 individuals 201 encountered were identified to the genus or family level and were classified into trophic groups: 202 plant-feeders, bacterial feeders, fungal feeders, predators, and omnivores (Yeates et al. 1993). 203 204 All nematodes were identified when the total number of nematodes was lower than 100 individuals. In this study, we defined nematode alpha-diversity as the genus or family richness 205 206 in a single 5 \times 5 m subplot and gamma-diversity as the total richness of the five subplots in each forest plot. Beta-diversity was defined as the turnover in taxon composition among the 207 five subplots in each plot, and was measured as Beta = 1-alpha_{mean}/gamma (Whittaker 1960, 208 209 Kraft et al. 2011), where alpha_{mean} is the mean alpha diversity of each plot (N = 5).

210

211 Statistical analysis

Given the hierarchical structure of our soil sampling protocol, linear mixed effects models 212 (LMMs) were used to test the effects of climatic zone on soil and nematode properties, using 213 the R-package lmerTest (Kuznetsova et al. 2017). For models testing the effects of climatic 214 zone on soil pH, soil organic carbon, total soil nitrogen, total soil phosphorus, nematode alpha-215 diversity, total nematode biomass, and abundance and relative abundance of different nematode 216 trophic groups, we fitted forest site and plot as random effects. For models testing the effects 217 of climatic zone on nematode beta-diversity, nematode gamma-diversity, and the biomass 218 proportion of different nematode trophic groups, we fitted forest site as a random effect. Post-219 220 hoc tests were used to compare differences among climatic zones.

For all LMMs, restricted maximum likelihood estimates of the parameters were determined using the lmer function in the lme4 package in R (Bates et al. 2015). For the LMMs, we calculated the marginal and conditional R^2 values, which account for fixed and fixed plus random effects, respectively. Akaike Information Criteria (AIC) were used to test the random effects. Specifically, if there was no difference between the AIC value of the model with fixed plus random effects and the AIC value of the model with only fixed effects, then random effects were considered to be unimportant. On the other hand, if the difference in AIC value between the model with fixed plus random effects and the model with only fixed effects is large, then the model with fixed plus random effects and the model with only fixed effects is large, then the random effects cannot be ignored.

WorldClim and MODIS does not allow to obtain climate and GPP differences among subplots within a plot and among some plots within a forest site. Therefore, one-way ANOVA was used to test the effects of climatic zone on MAP, MAT and GPP. Levene's tests were performed to test for homogeneity of variance.

Furthermore, the data on woody plant alpha-diversity were derived from surveys at the 234 plot level. We therefore calculated gamma-diversity and beta-diversity of woody plants at the 235 forest site level. Thus, different from our hierarchical data on soil properties (i.e., subplots, 236 plots, sites and climatic zones), the data on plant diversity are lacking a clear hierarchy. Hence, 237 independent-samples t-tests were employed to compare differences in alpha-, beta-, and 238 gamma-diversity of woody plants between the temperate zone and tropical zone. All ANOVAs 239 and t-tests were performed using SPSS 18.0 software (SPSS Inc., Chicago, IL, USA). Statistical 240 significance was set at p < 0.05. 241

In order to determine which variables (soil pH, soil organic carbon, total soil nitrogen, total soil phosphorus, GPP, alpha-diversity of woody plants, MAP, MAT) were most influential in driving total nematode biomass, nematode alpha-diversity, total nematode abundance and the abundance of different trophic groups, the relative importance of each variable was tested using boosted regression tree (BRT) models (Elith et al. 2008) using the gbm package in R (Ridgeway 2006). Specifically, this method combines regression trees (where models are built according

to constrained clustering rules based on how recursive partitioning of a quantitative variable 248 minimizes the within-group sums of squares under the constraining of a certain explanatory 249 variable) and boosting algorithm (i.e., a numerical optimization technique for minimizing the 250 loss function by stage-wisely adding a new tree). To prevent overfitting, the model used a 251 penalized forward stepwise search and cross-validation method to identify the optimal number 252 of decision trees (Elith et al. 2008). In BRT models, the settings of tree complexity, learning 253 254 rate, bagging fraction and cross-validation folds influence the model fit (De'ath 2007; Elith et al. 2008). Therefore, multiple model processes that involved training, validation and testing 255 256 were fitted with different combinations of parameter settings. Finally, models showing the 257 smallest cross-validated relative error (CVRE) were selected as optimal models.

We ran two sets of models. First, we ran eight models (for total nematode biomass, 258 nematode alpha-diversity, the abundance of total, plant-feeding, bacterial-feeding, fungal-259 feeding, omnivorous, and predatory nematodes) based on data from all individual sub-plots 260 across all forest sites (n = 195). The optimal model settings for each dependent variable were 261 determined by running 108 BRT processes with the following combinations of settings: tree 262 complexities of 2, 3, 4 and 5; learning rate of 0.01, 0.005 and 0.001; bag fraction of 0.4, 0.5 263 and 0.6; 5-, 8- and 10-fold cross-validations. Similarly, we ran eight models (for total nematode 264 biomass, nematode alpha-diversity, the abundance of total, plant-feeding, bacterial-feeding, 265 fungal-feeding, omnivorous, and predatory nematodes) based on data from individual sub-plots 266 in temperate (n = 60), warm-temperate (n = 60), and tropical (n = 75) forest sites. For these 267 analyses, the optimal model settings for each dependent variable were determined by running 268 12 BRT processes with the following combinations of settings: tree complexity of 2; learning 269 270 rate of 0.01, 0.005, 0.001, and 0.0001; bag fraction of 0.7; 5-, 8- and 10-fold cross-validations. In all of the above analyses, we used the same climate data for all five subplots within each 271 sampling plot because WorldClim does not allow to detect climate differences among sub-plots 272

at such a small spatial scale. Similarly, the MODIS GPP data have a spatial resolution of 500 273 m, while the area of a plot is 400 m². The GPP data were obtained based on the latitude and 274 longitude of each sampling plot. We then used the same GPP data for all five subplots within 275 each sampling plot. In the present study, the plots for plant surveys do not exactly match with 276 the plots for nematode sampling. Thus, for each forest site, we averaged the plant diversity 277 values at the plot level and then used this value as a proxy of plant diversity at the subplot level. 278 279 Second, we ran eleven models (for nematode biomass, nematode alpha-, beta-, and gamma-diversity, mean alpha-diversity of nematodes at plot level, nematode richness index, 280 281 and the abundance of total, plant-feeding, bacterial-feeding, fungal-feeding, omnivorous, and predatory nematodes) based on plot-level data across temperate, warm-temperate, and tropical 282 zones (n = 39). In total, 36 BRT models for each dependent variable were fitted with the 283 combinations of the following settings: tree complexities of 2, 3, 4 and 5; learning rate of 0.01, 284 0.005 and 0.001; bag fraction of 0.75; 5-, 8- and 10-fold cross-validations. The relative 285 influence of predictor variables on the response variables was estimated based on the number 286 of times a predictor variable was selected for splitting, weighted by the squared improvement 287 to the model as a result of each split and averaged over all trees (De'ath 2007, Elith et al. 2008). 288

289

290 **Results**

291 Climate, plant, and soil properties

From tropical to temperate forests, mean annual precipitation (MAP) and mean annual temperature (MAT) gradually decreased, but soil organic carbon gradually increased (Table 1; Appendix S1: Tables S1, S2b). Total soil nitrogen, total soil phosphorus and nematode betadiversity differed among climatic zones. Specifically, total soil nitrogen and total soil phosphorus were greater in temperate forests than in warm-temperate and tropical forests, while there was no difference between warm-temperate and tropical forests (Table 1; Appendix 298 S1: Table S2c, S2d).

Nematode beta-diversity was greater in temperate and warm-temperate forests than in 299 tropical forests, but we did not find differences in nematode beta-diversity between warm-300 temperate and temperate forests (Table 1; Appendix S1: Table S2g). Nematode alpha-diversity 301 was greater in tropical forests than in temperate forests (Table 1; Appendix S1: Table S2e). 302 Additionally, we did not find significant differences among climatic zones in nematode 303 304 gamma-diversity and soil pH (Table 1; Appendix S1: Table S2f, S2a). In general, nematode biomass was greater in tropical forests than in warm-temperate and temperate forests (Table 1; 305 306 Appendix S1: Table S2h). Furthermore, woody plant alpha-diversity, beta-diversity and gamma-diversity were greater in tropical than in temperate forests (Table 1; Appendix S1: 307 Table S1). 308

309

310 Drivers of nematode diversity, abundance, and biomass

At a regional scale, i.e., across tropical, warm-temperate, and temperate forests, soil organic C 311 was more important in influencing nematode alpha-diversity, beta-diversity, gamma-diversity, 312 and total abundance than mean annual precipitation (MAP), mean annual temperature (MAT), 313 or vegetation properties (gross primary productivity and diversity of woody plants) (Figs. 2, 314 3). Mean annual temperature and total soil phosphorus in temperate forests and MAP in tropical 315 forests were particularly important in influencing nematode alpha-diversity and trophic groups, 316 317 while soil pH was the most important variable in influencing nematode alpha-diversity and trophic groups in warm-temperate forests. Interestingly, the alpha-diversity of woody plants 318 had a greater influence on nematode alpha-diversity in tropical forests than in temperate forests. 319 Additionally, MAP and total soil phosphorus were the most important variables in influencing 320 nematode biomass in tropical and temperate forests, respectively (Fig. 2). Furthermore, the 321 highest total abundance of nematodes was found in tropical forests, followed by warm-322

temperate forests and temperate forests (Fig. 4; Appendix S1: Table S5). Moreover, we found
a greater proportional biomass of plant-feeding nematodes in tropical forests than in warmtemperate and temperate forests (Fig. 4; Appendix S1: Table S3m). We also found that warmtemperate forests were dominated by omnivorous nematodes (Fig. 4; Appendix S1: Table S3d,
S3p), and that the biomass proportion and relative abundance of fungal-feeding nematodes was
lower in tropical forests than in warm-temperate and temperate forests (Fig. 4; Appendix S1:
Table S3h, S3n).

330

331 Discussion

In the present study, we found that the drivers of nematode diversity, biomass and abundance 332 varied across climatic zones, and that nematode diversity, biomass and the abundance of 333 different nematode trophic groups varied more strongly with climatic factors in temperate and 334 tropical forests and with soil properties in warm-temperate forests. However, latitudinal 335 variation in gross primary productivity (GPP) accumulation rate did not reflect variation in 336 nematode abundance and diversity. Interestingly, plant diversity showed similar latitudinal 337 patterns as the alpha-diversity of nematodes, but plant diversity did not reflect beta- and 338 gamma-diversity of nematodes at the regional scale. 339

Studies on nematode diversity that focussed on marine nematode assemblages have shown 340 that nematode diversity typically decreases with increasing latitude (Lambshead et al. 2002; 341 Nicholas and Trueman 2005, Lee and Riveros 2012). However, patterns observed in marine 342 environments are not easy to translate to terrestrial ecosystems. We found that nematode alpha-343 diversity was greater in tropical than in temperate forests, but we did not find statistical 344 differences in nematode alpha-diversity between warm-temperate forests and tropical or 345 temperate forests. On the contrary, we found that nematode beta-diversity was lower in tropical 346 than in temperate and warm-temperate forests, but we did not find statistical differences in 347

nematode beta-diversity between warm-temperate forests and temperate forests. Finally, we
did not find significant differences among climatic zones in nematode gamma-diversity.

The extreme heterogeneity of soil environments at different spatial and temporal scales 350 could promote the diversity of soil organisms (Bardgett et al. 2005). For instance, soil 351 biodiversity at a small scale can be affected by soil compactness (Decaëns 2010), and small-352 scale soil spatial heterogeneity has been shown to promote nematode diversity (Nielsen et al. 353 354 2010). On the other hand, at a spatial larger scale, soil fertility may largely affect the composition of soil communities (Laliberté et al. 2017). In the present study, soil organic C 355 356 increased from tropical to temperate forests. Furthermore, soil properties may be influenced in part by the structure and composition of the current vegetation through root and litter input 357 (Coleman et al. 2004), but at the same time soil formation is a long-term process and the 358 development of soil properties is partly independent of the current vegetation and may also 359 reflect past environmental conditions (Wall et al. 2005). This could explain why we found that 360 soil properties exerted a greater impact on soil nematode abundance and diversity than 361 aboveground plant GPP and diversity across climatic zones. 362

On the other hand, plant-feeding nematodes are likely to be strongly affected by the current 363 plant productivity (and biomass) and community composition as they directly feed on plant 364 roots or root hairs (Ferris and Bongers 2006, Neher 2010). However, in our dataset, plant 365 productivity did not explain much of the variation in the abundance of plant-feeding 366 nematodes. We did find however that the GPP and the biomass of nematodes in tropical forests 367 were greater than those in warm-temperate and temperate forests. Interestingly, the alpha-368 diversity of woody plants had a greater influence on nematode alpha-diversity in tropical 369 370 forests than in temperate forests. Also, the biomass of plant-feeding nematodes in tropical forests accounted for more than 90% of the total nematode biomass, and it was greater than the 371 biomass of plant-feeding nematodes in warm temperate and temperate forests. This indicates 372

that the plant-feeding nematodes and their diversity, relative to other nematode trophic groups, 373 are more likely to be controlled by the bottom-up effects of plant productivity and plant 374 diversity (Ferris and Bongers 2006, Neher 2010). Here, plant roots might represent a 375 particularly important food resource for nematodes because root turnover rates are greater at 376 low latitudes than at high latitudes (Gill and Jackson 2000). Furthermore, the diversity in root 377 morphological traits may be greatest at lower latitudes, where plant diversity is highest (Ma et 378 379 al. 2018); this may cause greater heterogeneity in soil environments and potentially stimulate greater nematode diversity in tropical forests. 380

381 Availability of food and water resources, coupled with a suitable temperature, constitute the principal factors that control nematode abundance and distribution (Yeates 2004). The 382 383 normal temperature at which nematodes are active is roughly between 5 and 30 °C with the optimal activity temperature being about 20 °C (Wharton 2004). In our study, the variation in 384 mean monthly temperature ranged from -20 to 20 °C in temperate forests, which is greater than 385 in warm-temperate forests (-8 to 26 °C) and tropical forests (about 16 to 26 °C) (Appendix S1: 386 Table S6), and this may be one of the reasons why MAT explained more of the variation in 387 nematode alpha-diversity and in the abundance of different nematode trophic groups in 388 389 temperate forests. Also, at higher latitudes, litter decomposition rate is slower (Meentemeyer 1978) and accumulated litter may create variable microclimatic conditions for soil fauna. 390 391 Meanwhile, the greater temperature fluctuations in temperate forests may result in more 392 variability in resource availability in time and space. As a result, greater litter accumulation 393 and temperature fluctuations in temperate forests may promote heterogeneity in nematodes assemblages, which could explain the increase in nematode beta-diversity with increasing 394 395 latitude.

In addition, the movement, development and survival of nematodes in soil are also regulated by soil porosity and water potential (Barbercheck and Duncan 2004). For instance,

variation in precipitation can alter the composition of soil nematode communities (Franco et 398 al. 2019), which may result in changes in nematode diversity. Reduced aeration in saturated 399 soil is often detrimental to nematodes and it has been shown that the developmental rate and 400 movement of many nematodes is highest when soil water content changes from saturated to 401 slightly drier states (Barbercheck and Duncan 2004). In our studied system, there is greater 402 variation in mean monthly precipitation in tropical forests (ranging from 16 to 315 mm) relative 403 404 to warm-temperate forests (2 to 231 mm) and temperate forests (3 to 170 mm) (Appendix S1: Table S7), and this may explain why MAP in tropical forests explains more of the variability 405 406 in soil nematode alpha-diversity and in the abundance of different nematode trophic groups compared to the other forest types. 407

408

409 Conclusions

Given that nematodes are the most abundant animals on Earth and play key roles in regulating 410 carbon and nutrient dynamics, and reflecting biological activity in soils (van den Hoogen et al. 411 2019), understanding the drivers of nematode diversity in forest soils is essential to better 412 predict how forest ecosystems respond to multiple environmental changes. Based on a large-413 scale investigation of soil nematode diversity and the abundance of different nematode trophic 414 groups and their drivers in forested ecosystems, we conclude that across a hydrothermal 415 gradient from temperate to tropical climates, nematode abundance and diversity were mainly 416 influenced by soil organic carbon. Within climatic zones, however, mean annual temperature 417 and total soil phosphorus in temperate, soil pH in warm temperate, and mean annual 418 precipitation in tropic were more important drivers. This demonstrates that the drivers of soil 419 nematode diversity in forested ecosystems are affected by the spatial scale and climatic 420 conditions considered. 421

422

In temperate forests the magnitude of variation in temperature may be an important driver

of nematode diversity, while in tropical forests the magnitude of variation in precipitation may 423 be more important. This indicates a latitudinal shift in the main climatic variables controlling 424 soil nematode communities. Our data further indicated that the alpha- and beta-diversity of 425 woody plant and nematode alpha-diversity were greater in the tropical forests than in the 426 temperate forests, while nematode beta-diversity was lower in the tropical forests than in the 427 temperate forests. This demonstrates a complex and scale-dependent relationship between 428 429 plant diversity and soil biodiversity. These findings also have consequences for the scales of observation of area-diversity relationships (Decaëns 2010). For those forests with greater 430 431 nematode beta-diversity, such as temperate forests, increasing the area of sampling may enhance the probability of detecting rare species. Therefore, for better understanding the 432 drivers of soil biodiversity distribution, it is crucial to explore the area-diversity relationships 433 of soil organisms across various ecosystems. 434

435

436 Data Availability

437 All data are presented in the paper.

438

439 Acknowledgments

We acknowledge all members from the Institute of Geographic Sciences and Natural Resources
Research, Chinese Academy of Sciences, and Peking University for their help with the
fieldwork. This study was sponsored by the Natural Science Foundation of China (31971534,
U1904204, U1804101) and the Innovation Scientists and Technicians Troop Construction
Projects of Henan Province (grant number 182101510005). We are also very grateful for the
support from the nature reserves.

446

447 Author Contributions

448	All authors contributed intellectual input and assistance to this study. D.G. and S.F. designed
449	the study. Y.S. wrote the first draft with help of P.K. Sample collection, and soil chemical and
450	nematode analysis were carried out by C.M., Z.W., T.L., and Y.S. Climate and plant data were
451	collected by Y.H., C.M. and Y.S. Statistical analyses were carried out by Y.S., Y.C. and Z.W.
452	Assistance in data interpretation was provided by Y.S., P.K., Y.H., Y.C., C.Z. and W.Z.
453	
454	Competing Interest Statement
455	The authors declare no competing interests.
456	
457	References
458	Barbercheck, M. E., and L. Duncan. 2004. Abiotic Factors. Pages 309-343 in Gaugler R., and
459	A. L. Bilgrami, editors. Nematode Behaviour. CABI Publishing, Wallingford, Oxfordshire.
460	UK.
461	Bardgett, R. D. and W. H. van der Putten 2014. Belowground biodiversity and ecosystem
462	functioning. Nature 515: 505–511.
463	Bardgett, R. D., and D. A. Wardle. 2010. Aboveground-Belowground Linkages: Biotic
464	Interactions, Ecosystem Processes, and Global Change. Oxford University Press. Oxford.
465	UK.
466	Bardgett, R. D., G. W. Yeates, and J. M. Anderson. 2005. Patterns and determinants of soil
467	biological diversity. Pages 100-118 in R. D. Bardgett, M. B. Usher, and D. W. Hopkins,
468	editors. Biological Diversity and Function in Soils. Cambridge University Press, Cambridge.
469	UK.
470	Bates, D., M. Mächler, B. Bolker, and S. Walker. 2015. "Fitting Linear Mixed-Effects Models
471	Using lme4." Journal of Statistical Software 67: 1-48. doi: 10.18637/jss.v067.i01.
472	Bongers, T. and H. Ferris. 1999. Nematode community structure as a bioindicator in

- 473 environmental monitoring. Trends in Ecology & Evolution 14: 224–228.
- 474 Chen, D., J. Cheng, P. Chu, J. Mi, and Y. Bai. 2016. Effect of diversity on biomass across
- grasslands on the mongolian plateau: contrasting effects between plants and soil nematodes.
- 476 Journal of Biogeography 43: 955–966.
- Chen, D., J. Cheng, P. Chu, S. Hu, Y. Xie, I. Tuvshintogtokh, and Y. Bai. 2015. Regional-scale
 patterns of soil microbes and nematodes across grasslands on the Mongolian plateau:
- relationships with climate, soil, and plants. Ecography 38: 622-631.
- 480 Coleman, D.C., J. M. Blair, E. T. Elliott, and D. H. Wall. 1999. Soil invertebrates. Pages 349-
- 481 377 in G. P. Robertson, D. C. Coleman, C. S. Bedsoe, and P. Sollins, editors. Standard Soil
- 482 Methods for Long-term Ecological Research. Oxford University Press. New York. USA.
- 483 Coleman, D.C., D. A. Crossley Jr., and P. F. Hendrix. 2004. Fundamentals of Soil Ecology.
- 484 Second edition. Elsevier Academic Press, Burlington, San Diego, London.
- 485 De'ath, G.2007. Boosted trees for ecological modeling and prediction, Ecology 88, 243–251.
- 486 Decaëns, T. 2010. Macroecological patterns in soil communities. Global Ecology and
 487 Biogeography 19: 287–302.
- Elith, J., J. R. Leathwick, and T. Hastie. 2008. A working guide to boosted regression trees.
 Journal of Animal Ecology 77: 802–813.
- Ferris, H., and T. Bongers. 2006. Nematode indicators of organic enrichment. Journal of
 Nematology 38: 3–12.
- 492 Franco, A.L.C., L. A. Gherardi, C. M. D. Tomasel, W. S. Andriuzzi, and D. H. Wall. 2019.
- 493 Drought suppresses soil predators and promotes root herbivores in mesic, but not in xeric
- 494 grasslands. Proceedings of the National Academy of Sciences USA 116: 12883–12888. doi:
 495 10.1073/pnas.1900572116.
- 496 Gill, R. A., and R. B. Jackson. 2000. Global patterns of root turnover for terrestrial ecosystems.
- 497 New Phytologist 147: 13–31.

- Hawkins, B.A. et al. 2003. Energy, water, and broad-scale geographic patterns of species
 richness. Ecology 84: 3105–3117.
- 500 Hijmans, R. J., S. E. Cameron, J. L. Parra, P. G. Jones, and A. Jarvis. 2005. Very high resolution
- interpolated climate surfaces for global land areas. International Journal of Climatology 25:
 1965–1978.
- Hugot, J. P., P. Baujard, and S. Morand. 2001. Biodiversity in helminths and nematodes as a
 field of study: an overview. Nematology 3: 199–208.
- 505 Kraft, N. J. et al. 2011. Disentangling the drivers of β diversity along latitudinal and elevational 506 gradients. Science 333: 1755–1758.
- 507 Kuznetsova, A., P. B. Brockhoff, and R. H. B. Christensen. 2017. "ImerTest Package: Tests in
- 508 Linear Mixed Effects Models." Journal of Statistical Software 82: 1–26. doi:
 509 10.18637/jss.v082.i13.
- Laliberté, E., P. Kardol., R. K. Didham, F. P. Teste, B. L. Turner, and D. A. Wardle. 2017. Soil
 fertility shapes belowground food webs across a regional climate gradient. Ecology Letters
 20: 1273–1284.
- Lambshead, P., C. J. Brown, T. J. Ferrero, N. J. Mitchell, C. R. Smith, L. E. Hawkins, and J.
- 514 Tietjen. 2002. Latitudinal diversity patterns of deep-sea marine nematodes and organic
- fluxes: A test from the central equatorial Pacific. Marine Ecology Progress Series 236: 129–
 135.
- Lee, M. R., and M. Riveros. 2012. Latitudinal trends in the species richness of free-living
 marine nematode assemblages from exposed sandy beaches along the coast of Chile (18–
 42°S). Marine Ecology 33: 317–325.
- Liu, G. 1996. Soil Physical and Chemical Analysis & Description of Soil Profiles. China
 Standard, Beijing, China (in Chinese).
- 522 Ma, Z.Q. et al. 2018. Evolutionary history resolves global organization of root functional traits.

- 523 Nature 555: 94-97.
- McColltoll, R. W. 2005. Encyclopedia of World Geography, Volume 1. Facts on File Library
 of World Geography. Facts on File, New York, USA.
- Meentemeyer, V. 1978. Macroclimate and lignin control of litter decomposition rates. Ecology
 59, 465–472.
- Neher, D. A. 2010. Ecology of plant and free-living nematodes in natural and agricultural soil.
 Annual Review of Phytopathology 48: 371–394.
- 530 Nicholas, W. L., and J. Trueman. 2005. Biodiversity of marine nematodes in Australian sandy
- beaches from tropical and temperate regions. Biodiversity and Conservation 14: 823–839.
- 532 Nielsen, U.N., E. Ayres, D. H. Wall, G. Li, R. D. Bardgett, T. Wu, and J. R. Garey. 2014. Global-
- scale patterns of assemblage structure of soil nematodes in relation to climate and ecosystem
- properties. Global Ecology and Biogeography 23: 968–978.
- 535 Nielsen, U.N., G. H. R. Osler, C. D. Campbell, R. Neilson, D. F. R. P. Burslem, and R. van
- der Wal. 2010. The enigma of soil animal species diversity revisited: The role of small-scale
 heterogeneity. PLoS One 5: e11567.
- 538 Phillips, H.R.P. et al. 2019. Global distribution of earthworm diversity. Science 366: 480–485.
- Ridgeway, G. 2006. Generalized boosted regression models. Documentation on the R Package
 'gbm', version 1.5–7.
- van den Hoogen, J. et al. 2019. Soil nematode abundance and functional group composition at
 a global scale. Nature 572: 194–198.
- 543 Wagg, C., S. F. Bender, F. Widmer, and M. G. A. van der Heijden. 2014. Soil biodiversity and
- soil community composition determine ecosystem multifunctionality. Proceedings of the
 National Academy of Sciences USA 111: 5266–5270.
- 546 Wall, D. H., A. H. Fitter, and E. A. Paul. 2005. Developing new perspectives from advances in
- 547 soil biodiversity research. Pages 3–27 in R. D. Bardgett, M. B. Usher, and D.W. Hopkins,

- editors. Biological Diversity and Function in Soils. Cambridge University Press, Cambridge,
- 549 UK.
- Wall, D.H., U.N. Nielsen, and J. Six. 2015. Soil biodiversity and human health. Nature 528:
 69–76.
- Wang, L, H. Zhu, A. Lin, L. Zou, W. Qin, and Q. Du. 2017. Evaluation of the latest MODIS
 GPP products across multiple biomes using global eddy covariance flux data. Remote
- 554 Sensing 9: 418; doi:10.3390/rs9050418.
- 555 Wardle, D. A., R. D. Bardgett, J. N. Klironomos, H. Setälä, W. H. van der Putten, and D. H.
- Wall. 2004. Ecological linkages between aboveground and belowground biota. Science 304:
 1629–1633.
- Wharton, D.A. 2004. Survival Strategies. Pages 371–399 *in* R. Gaugler, and A. L. Bilgrami,
 editors. Nematode Behaviour. CABI Publishing, Wallingford, Oxfordshire. UK.
- Whittaker, R.H. 1960. Vegetation of the Siskiyou Mountains, Oregon and California.
 Ecological Monographs 30: 279–338.
- Wu, J., H. Chen, and Y. Zhang 2016. Latitudinal variation in nematode diversity and ecological
 roles along the Chinese coast. Ecology and Evolution 6: 8018–8027.
- Xiao, H., W. Wang, S. Xia, Z. Li, J. Gan, and X. Yang. 2021. Distributional patterns of soil
 nematodes in relation to environmental variables in forest ecosystems. Soil Ecology Letters
 3: 115–124.
- Xiong, D. et al. 2020. Nonlinear responses of soil nematode community composition to
 increasing aridity. Global Ecology and Biogeography 29: 117–126.
- 569 Yeates, G.W. 2004. Ecological and Behavioural Adaptations. Pages 1–24 in R. Gaugler, and A.
- 570 L. Bilgrami, editors. Nematode Behaviour. CABI Publishing, Wallingford, Oxfordshire. UK.
- 571 Yeates, G.W., T. Bongers, R. G. M. De Goede, D. W. Freckman, and S. S. Georgieva. 1993.
- 572 Feeding habits in soil nematode families and genera—an outline for soil ecologists. Journal

573 of Nematology 25: 315–331.

- Zhang, F, J. M. Chen, J. Chen, C. M. Gough, T. A. Martin, and D. Dragoni. 2012. Evaluating
- spatial and temporal patterns of MODIS GPP over the conterminous U.S. against flux
- 576 measurements and a process model. Remote Sensing of Environment 124: 717–729.
- 577 Zhao, M, F. A. Heinsch, R. R. Nemani, and S.W. Running. 2005. Improvements of the MODIS
- 578 terrestrial gross and net primary production global data set. Remote Sensing of Environment
- 579 **95**: 164–176.
- 580 Zhao, M, and S.W. Running. 2010. Drought-induced reduction in global terrestrial net primary
- 581 production from 2000 through 2009. Science 329: 940–943.
- 582 Zhao, L. et al. 2021. The effects of plant resource inputs on the energy flux of soil nematodes
- are affected by climate and plant resource type. Soil Ecology Letters 3: 134–144.

584 Table 1. Climate, soil, plant and nematode properties for tropical, warm-temperate and temperate forest sites. Values are means ±SE for nematode beta- and gamma-diversity, and woody plant beta- and gamma diversity (n = 4 in temperate forests (4 sites), and n = 5 in tropical forests (5 sites)); 585 for woody plant alpha-diversity (n = 12 in temperate forests (4 sites \times 3 plots), and n = 20 in tropical forests (5 sites \times 4 plots)); for climatic 586 properties, soil physio-chemical properties, GPP, nematode alpha-diversity and nematode biomass (n = 12 in temperate and warm-temperate forests 587 (4 sites \times 3 plots), and n = 15 in tropical forests (5 sites \times 3 plots). Different letters in a row indicate significant differences among tropical, warm-588 temperate, and temperate forests (post-hoc tests after one-way ANOVA for MAP, MAT and GPP, independent-samples t-test for alpha-, beta- and 589 gamma-diversity of woody plants, and post-hoc test after linear mixed effects models for soil and nematode properties, P < 0.05). MAP, mean 590 annual precipitation; MAT, mean annual temperature; *, diversity indices for warm-temperate could not be calculated due to the absence of plant 591 592 data.

			Tro	pical	Warm	-ten	nperate	Temp	oera	te
Climate	MAP (mm)	1645.9	±	8.2ª	1116.3	±	115.1 ^b	661.7	±	9.9°
	MAT (°C)	21.9	±	0.1 ^a	12.1	±	1.1 ^b	2.2	±	0.2 ^c
Soil	pH	5.3	±	0.3ª	5.1	±	0.3ª	5.7	±	0.1ª
	Soil organic C (mg g ⁻¹)	28.1	±	2.2 ^c	71.8	±	11.3 ^b	150.0	±	12.2ª
	Total soil N (mg g ⁻¹)	2.0	±	0.3 ^b	3.77	±	0.49 ^b	8.71	±	1.01 ^a
	Total soil P (mg g ⁻¹)	0.3	±	0.0^{b}	0.44	±	0.08^{b}	0.92	±	0.09 ^a
Plant	Alpha-diversity	44.6	±	4.9 ^a		*		13.3	±	1.3 ^b
	Gamma-diversity	106.2	±	22.7 ^a		*		19.0	±	2.3 ^b
	Beta-diversity	0.58	±	0.01 ^a		*		0.31	±	0.03 ^b
	GPP	229.4	±	0.3 ^a	137.1	±	6.4 ^b	126.3	±	2.5 ^b
Nematode	Alpha-diversity	42.8	±	1.2ª	38.8	±	3.6 ^{ab}	37.3	±	3.3 ^b
	Gamma-diversity	59.0	±	2.4 ^a	55.8	±	5.2ª	61.0	±	5.5 ^a
	Beta-diversity	0.27	±	0.02 ^b	0.31	±	0.04 ^a	0.39	±	0.04 ^a
	Biomass	3737.2	±	592.5ª	490.5	±	108.8 ^b	250.2	±	50.4 ^b

594	J	9	4
-----	---	---	---

Figure Legends

595

Fig. 1. The forest sampling sites across different climatic zones in China. Dots refer to the
sampling sites for field investigation, which spanned across tropical forests (Xishuangbanna,
XSBN), warm-temperate forests (Donglingshan, DLS; Jigongshan, JGS; Tiantongshan, TTS;

599 Badagongshan, BDGS), and temperate forests (Liangshui, LS; Changbaishan CBS).

600

Fig. 2. The relative importance of the seven or eight variables from the eight models based on 601 602 subplot-level data at the regional scale (n = 195) (A) and separately for tropical (n = 75), warmtemperate (n = 60), and temperate (n=60) forests (B-D) in explaining soil nematode variables. 603 Rows show the results of each model (from top to bottom, nematode biomass (Biomass), 604 nematode alpha-diversity (Alpha), the total abundance of nematodes (Total), and the 605 abundance of plant-feeding (Pl), omnivorous (Om), predatory (Pr), fungal-feeding (Fu), and 606 bacterial-feeding (Ba) nematodes). Columns represent the variables that are present in the 607 model. Within each row, the size of the circles is proportional to the relative importance of the 608 variables. Palpha refers to the alpha-diversity of woody plants. 609

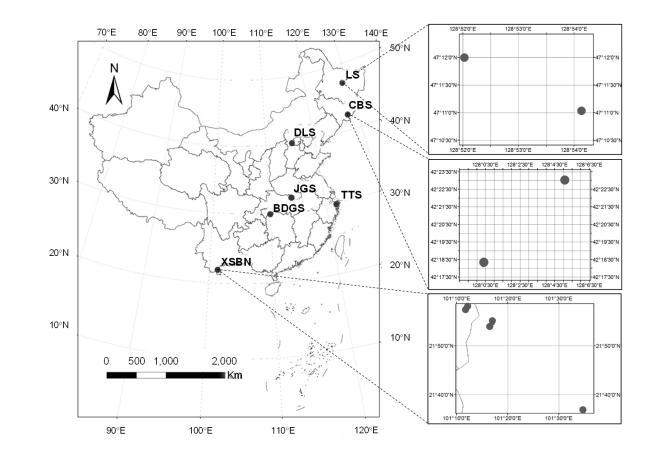
610

Fig. 3. The importance of the eight variables from the eleven models based on plot-level data across temperate, warm-temperate, and tropical forest sites (n = 39) in explaining soil nematode variables. Rows show the results of each model (from top to bottom, nematode biomass (Biomass), nematode gamma-diversity (Gamma), nematode beta-diversity (Beta), nematode alpha-diversity (Alpha.mean), nematode taxon richness (Richness), the total abundance of nematodes (Total), and the abundance of plant-feeding (Pl), omnivorous (Om), predatory (Pr), fungal-feeding (Fu), and bacterial-feeding (Ba) nematodes). Columns represent the variables

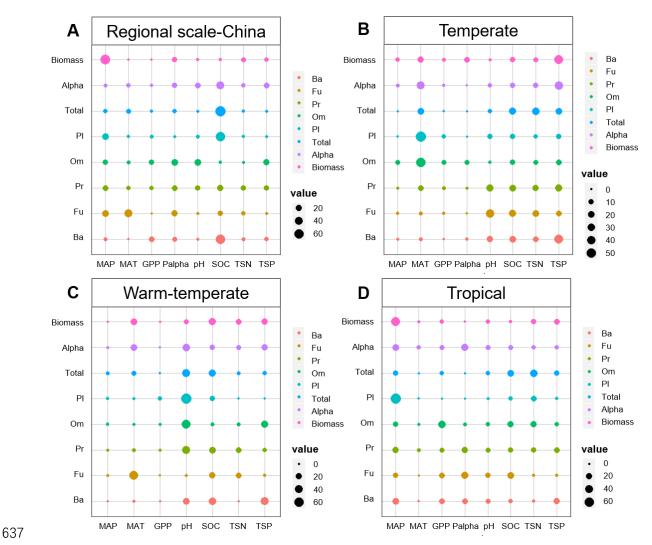
618	that are present in the model. Within each row, the size of the circles is proportional to the
619	relative importance of the variables. Palpha refers to the alpha-diversity of woody plants.
620	

- 621 Fig. 4. The abundance (a), relative abundance (b), and biomass proportion (c) of plant-
- 622 feeding (Pl), fungal-feeding (Fu), bacterial-feeding (Ba), omnivorous (Om), unknown and
- 623 predatory (Pr) nematodes for tropical, warm-temperate, and temperate forests. Notably,
- 624 'unknown' nematodes referred to unidentified nematodes and were not included in panel c
- because nematode biomass was estimated based on the identified nematodes, and biomass of
- 626 unknown nematodes could not be estimated accurately. Data are means \pm SE (n = 12 for
- 627 temperate and warm- temperate forest plots, n = 15 for tropical forest plots). Within each
- panel, different lowercase letters denote significant (P < 0.05) differences among climatic
- 629 zones based on post-hoc tests after linear mixed effects models.

630

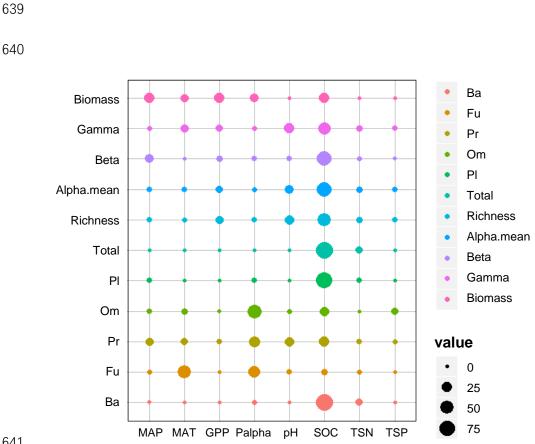


634 Fig. 1



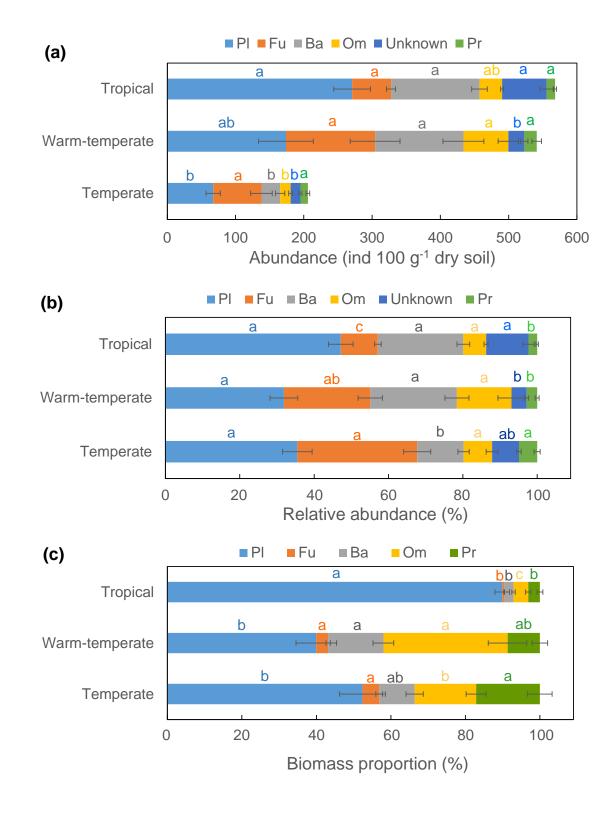












644

645 Fig. 4