**Dynamic coastal pelagic habitat drives rapid changes in growth and condition of juvenile**

**sockeye salmon (*Oncorhynchus nerka*) during early marine migration**

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**Abstract**

Migrating marine taxa encounter diverse habitats that differ environmentally and in foraging

conditions over a range of spatial scales. We examined body (RNA/DNA, length-weight

residuals) and nutritional condition (fatty acid composition) of juvenile sockeye salmon

(*Oncorhynchus nerka*) in British Columbia, as they migrated through coastal waters that varied

oceanographically over tens of kilometers. Fish were sampled in the stratified, productive

northern Strait of Georgia (NSoG); the highly mixed, unproductive Johnstone Strait (JS); and the

transitional zone of Queen Charlotte Strait (QCS). Body and nutritional condition responded

rapidly to changes in prey availability and were lowest in JS with low prey availability, supporting

the Tropic Gauntlet Hypothesis, additionally we saw signs of compensatory growth in QCS.

Juvenile salmon leaving the SoG in 2016 had significantly lower condition than in 2015, despite

higher zooplankton biomass in 2016. We propose that this was due to the higher abundance of

low food quality southern zooplankton species in 2016. This study highlights the importance of

including food quality as a parameter to understand changes in fish condition and survival

between years. Furthermore, small scale variation in oceanographic dynamics impact foraging

conditions and need to be considered when assessing early marine survival of juvenile salmon.

**Keywords: fatty acids, RNA/DNA, zooplankton biomass, environmental variation,**

**compensatory growth, zooplankton species composition**

**Introduction**

Diverse marine taxa from fish and squid to whales undertake long-distance marine migrations. These migrations can reduce the risk of predation and diseases, provide access to spawning grounds, and increase foraging opportunities. Anadromous fish undertake long-distance marine migrations, often travelling as juveniles from the mouths of their natal watersheds to the open ocean to forage and returning to freshwater as adults to reproduce. During their coastal outmigration, anadromous fish encounter interconnected and diverse ecosystems that differ in their physics, nutrient status and productivity levels, yielding highly contrasting foraging conditions and opportunity for growth (Holt, 2008; Polis et al., 2004). The effects of spatial variation in prey availability among migration routes are important for predicting the survival of fish stocks but are commonly understudied.

Pacific salmon (*Oncorhynchus* spp.) are a group of anadromous species that experience extreme spatial variations in ecosystem attributes during their early marine life, during which they typically migrate from estuaries to the coastal ocean and then to the open ocean. The first months after Pacific salmon enter the ocean have been identified as a particularly important life phase, with mortalities up to 80-90% estimated for tagged salmon (Welch et al., 2011). Growth and mortality during this phase may contribute significantly to the interannual variability and long-term population trends of salmon stocks (Friedland et al., 2000; Mueter et al., 2005). Although there are likely numerous factors that influence juvenile salmon survival during their early marine phase, correlative evidence suggests that feeding conditions play an important role (Beamish et al., 2012; Duffy and Beauchamp, 2011). Through impacts on growth, feeding conditions have the potential to affect vulnerability to predation (Gliwicz, 2003), ability to survive stressful ocean conditions (Lasker, 1978), and the body condition required to survive their first ocean winter (Beamish and Mahnken, 2001).

Recent research from North America's west coast has drawn attention to the potential influence

of small-scale coastal marine conditions on juvenile salmon survival (Ferriss et al., 2014; Hunt

et al., 2018; Journey et al., 2018; McKinnell et al., 2014). Outmigrating juvenile Fraser River

sockeye (*Oncorhynchus nerka)* first enter the ocean into the stratified, comparatively warm, and

highly productive Strait of Georgia basin (SoG, Fig. 1). Many of the juveniles migrate northwards

through the SoG into the tidally mixed Discovery Islands (DI) and Johnstone Strait (JS) before

reaching the productive Queen Charlotte Strait (QCS) (McKinnell et al., 2014). In the Trophic

Gauntlet Hypothesis (TGH), McKinnell et al. (2014) proposed that strong tidal mixing in JS

causes light limited primary production, leading to sparse prey fields for migrating juvenile

sockeye salmon. Subsequent research has demonstrated that tidally mixed waters in JS are

indeed characterized by poor foraging conditions, with a prevalence of small zooplankton (< 2

mm; Mahara et al. (2021)), corresponding with low foraging success for migrating juvenile

sockeye salmon (James et al., 2020). Low insulin-like growth factor 1 (IGF-1) levels in JS

compared to higher levels in the northern Strait of Georgia (NSoG), Queen Charlotte Strait and

Queen Charlotte Sound supports an impact of these foraging conditions on juvenile salmon

growth (Ferriss et al., 2014; Journey et al., 2018). Poor foraging conditions can be expected to

reduce growth and increased mortality (Duffy and Beauchamp, 2011), especially when foraging

opportunities before or after JS are also poor (James et al., 2020).

Relating growth proxies to environmental conditions needs to consider the integration time of

the proxy used. IGF-1 is a biochemical proxy of growth that has been shown to integrate feeding

conditions experienced over 4-14 days (Beckman et al., 2004), relatively long when assessing

the response of fast migrating organisms over short distances (Duguid et al., 2018; Gabillard et

al., 2006; Pierce et al., 2005). The need for precise handling when sampling in the field, and the

difficulty in sampling from individuals smaller than 120mm also make the IGF-1 method difficult

to routinely apply in juvenile salmon research (Duguid et al., 2018; Ferriss et al., 2014). Another

growth proxy commonly used in fish ecology and fishery research is RNA/DNA ratio

(Clemmesen and Doan, 1996). The underlying principle of using the ratio of RNA to DNA is that

the concentration of DNA per cell is constant, whereas the RNA concentration per cell varies

with its anabolic activity. The RNA/DNA ratio integrates feeding history over a period of 1-5 days

(Buckley et al., 1999; Clemmesen and Doan, 1996; Wright and Martin, 1985), which makes it

useful for assessing fish response to rapid spatial changes in coastal migration habitat.

RNA/DNA can be effectively sampled irrespective of individual size and can be sampled from

individuals frozen in the field without any degradation.

Another variable that needs to be considered when assessing fish growth response is the

available prey. Prey quantity is often considered one of the main drivers of growth, survival, and

abundance of marine organisms (Bilton and Robins, 2011), however, prey quality also plays a

critical role (Litz et al., 2017). Fatty acids (FA), and essential fatty acids (EFA) in particular, are

useful indicators of both food quality and the nutritional condition of a fish (Tocher, 2003). EFAs

are not synthesised *de novo* by vertebrates in sufficient quantities to meet their physiological

demands (Xu et al., 2018). Rather, EFAs are primarily synthesised by phytoplankton and

subsequently reach consumers through trophic transfer (Dalsgaard et al., 2003). Marine

carnivorous fish, such as sockeye salmon, have only a limited ability to convert EFAs into

important long-chained PUFAs such as docosahexaenoic acid (DHA; 22:6n-3),

eicosapentaenoic acid (EPA; 20:5n-3), and therefore must be obtained from their diets (NRC,

2011). Studies have shown that DHA and EPA concentrations measured in fish tissue are

similar to levels in the prey consumed (Jin et al., 2017). Total amounts of DHA and EPA are

known to affect the development, growth, and survival of juvenile marine fish (Mourente et al.,

1991) and are of major importance during periods of rapid growth (Jin et al., 2017) (Table 1).

Low levels of DHA and EPA can lead to reduced growth, maldevelopment, reduced

reproduction, and ultimately decreased survival (Copeman and Parrish, 2002) (Table 1). Stored

FAs, especially DHA and EPA, are also important when fish experience times of food limitation

and starvation, for energy mobilization, hyperphagia, and for compensatory growth.

Experimental studies have shown that food quality changes, especially micronutrient limitation,

and associated fatty acid changes lead to reduced RNA/DNA as lipid and protein metabolism is

compromised when supply is limited (Boersma et al., 2008). The fatty acid composition

integrates the feeding history over a period of approximately one week in juvenile Chinook

salmon (Garzke et al. in rev), and using RNA/DNA and fatty acids in combination therefore

provides a useful approach to identify changes in nutritional and physiological condition in

organisms (Clemmesen, 1994; John et al., 2001; Malzahn et al., 2007; Paulsen et al., 2014; Xu

et al., 2009)

In this study, we examine the response of juvenile sockeye salmon nutritional condition when

migrating through complex oceanographic habitats in coastal British Columbia, Canada,

including a region previously demonstrated to provide food poor conditions. We show that

juvenile sockeye salmon nutritional condition responds rapidly to changes in foraging conditions.

Furthermore, we show that the response and subsequent recovery are dependent on fish

nutritional condition before entering food poor areas. Zooplankton biomass changes alone are

not sufficient to explain changes in fish condition between years, and we suggest that the

trophic gauntlet hypothesis needs to be extended by including prey quality, here measured with

respect to fatty acids, in addition to quantity.

**Materials and Methods**

*Ethics Statement*

This study did not involve endangered or protected species and was carried out under the

guidelines of the Canadian Council on Animal Care. Juvenile salmon were collected under DFO

license number 'XR 42 2015' and 'XR 92 2016' with approval from UBC's Animal Care

Committee (Protocol A19-0025). The study was carried out in compliance with the ARRIVE

guidelines 2.0 (www.arriveguidelines.org).

*Study location*

Juvenile sockeye salmon were sampled at 6 locations between the NSoG and QCS (Fig. 1) in

2015 (6 May - 7 July) and 2016 (14 May - 6 July), in three regions along their outmigration route

(Table S1). From south to north: the Discovery Islands (DI), JS, and southern QCS (Fig. 1).

Based on oceanographic sampling (Dosser et al., 2021), the DIs have water properties of the

northern SoG and JS has water properties of QCS. The JS sites were all located within JS

proper, where water depth was > 300 m. The QCS site was located in water <150 m deep to the

north of the JS sill. Assuming RNA/DNA turnover rates of 1-5 days, the RNA/DNA ratios of fish

collected in the DI were expected to predominantly reflect their migration experience through the

northern SoG. Given this and the NSoG water properties, we henceforth refer to the DI as

NSoG. Based on RNA/DNA replacement rates, JS fish would reflect conditions in the DI-JS, and

QCS fish would reflect the transition from JS to QCS. According to tagging data, the travel rates

of juvenile sockeye between these regions are < 2 days for NSoG to DI, ca. 7 days from DI to

JS, and ca. 4 days from JS to QCS (Furey et al., 2015; Johnston, 2020; Rechisky et al., 2018;

Welch et al., 2011).

*Juvenile Salmon Sampling*

Fish were collected from a 6 m motorized vessel at distances of 5 - 60 m from shore, using a

purse seine designed for manual deployment and retrieval (bunt: 27 x 9 m with 13 mm mesh;

tow 46 x 9 m with 76 mm mesh) (Hunt et al., 2018). Captured fish were initially held alongside

the vessel in a submerged portion of the seine's bunt, allowing fish to swim without contacting

the net and minimizing stress. Individual fish were transferred from the net by capturing them

using a seawater-filled plastic jug with the end cut off. The fish were euthanized with MS-222

solution (250 mg L$^{-1}$) in a 532 mL plastic bag with a unique identifier and frozen using liquid

nitrogen vapour in a dry-shipper (MVE Doble-47). Individuals were stored at -80°C until further

dissection in the lab. A total of 1,233 juvenile sockeye were collected across all stations and

years (weight: min. 2.8 g – 32.8 g (mean: 12.66g, median: 11.5g); length: min. 63mm – 145mm

(mean: 106.43mm, median: 106mm)), and a maximum of 20 individuals at each seine event.

Juvenile salmon were collected under Fisheries and Oceans Canada (DFO) license number 'XR

42 2015' and 'XR 92 2016' with approval from UBC's Animal Care Committee (Protocol A19-

0025). Juvenile salmon sampling and handling were in accordance with the DFO and UBC

Animal Care Committee guidelines.

*Laboratory analysis*

Five sockeye salmon were randomly selected from the total caught and were retained from

each purse seine for a given site and day for further analysis of RNA/DNA and fatty acids

(n=187). Before dissection fish were measured and weighed (mean weight = 9.07 g; range 1.27

– 38.85 g). White muscle tissue was collected from close to the dorsal fin for RNA/DNA

analysis, and 400-600 mg of muscle tissue from close proximity to the dorsal fin for fatty acid

analysis. Tissue samples were then returned to a -80 °C freezer until further analysis. In total,

187 juvenile salmon were processed across all regions and both years for fatty acids

composition and RNA/DNA ratio (2015: n=87: NSoG (n=16), JS (n=34), QCS (n=37); 2016:

n=100: NSoG (n=22), JS (n=37), QCS (n=41)).

*RNA/DNA analysis*

RNA/DNA was measured in crude muscle tissue homogenates using the non-specific, nucleic

acid intercalating fluorescence dye RiboGreen (Gorokhova and Kyle, 2002). Nucleic acid

quantification from white-muscle tissues followed a modified protocol by Caldarone et al (2006)

and Clemmesen et al. (1993). For RNA and DNA extraction, 1% N-lauroylsarcosine (VWR, Cat.

No. VWRV0719-500G) was added to the samples, followed by homogenization with glass

beads (VWR Cat. No. 12621-154), shaken for 30 min in a Retsch cell mill. Samples were

centrifuged for 10 min at 14,000 x $g$ and the supernatant diluted to match the RiboGreen

saturation range (1:20, 1:30, 1:50, 1:100 and 1:200). Triplicate aliquots from the diluted sample

were added to a black, flat-bottom 96-well microtiter plate. Standard solutions of rRNA

(*Eschericia coli* 16S and 23S rRNA, Quant-iT RiboGreen RNA assay Kit Thermo Fisher

Scientific Cat. No. R11490) and DNA (calf thymus, VWR, Supp. No. MB-102-01100) were

diluted to final concentrations between 0 µg/mL and 3 µg/mL. 1x RiboGreen was added to each

well and incubated for 5 min. Fluorescence was measured on a VarioSkan Flash Microplate

Reader (ThermoFisher Scientific; excitation = 500 nm, emission= 525 nm), with the signal

representing the fluorescence of RNA and DNA combined. Ribonuclease A (bovine pancreas,

Alpha Aesar, Cat. No. CAAAJ62232-EX3) was added to each well and incubated at 37°C for 30

min. Fluorescence of the digested samples were measured to determine the DNA only. The

RNA concentration was calculated by subtracting the DNA fluorescence and DNA+RNA

fluorescence. The RNA/DNA was calculated for each well, and samples where the coefficient of

variation (CV) for either RNA or DNA concentration in triplicate wells exceeded 15% were

excluded from further analysis following.

*Fatty acid analysis*

Fatty acid analyses were performed following a slightly modified version of the protocol

described in Puttick et al. (2009) using a one-step fatty acid methyl ester (FAME) method. Wet

weights of all muscle samples were measured, freeze-dried, and dry weight measured for

calculations of moisture content. FAMEs were obtained by reaction in a solution of 2.0 mL of 3

N HCl in $CH_3OH$ (Sigma-Aldrich cat. #90964-500ML). Nonadecanoic acid (C19:0) was used as

an internal standard (Abdulkadir and Tsuchiya, 2008). FAMEs were analysed with a gas

chromatograph (Scion 436-GC, Scion Instruments Canada, Edmonton, Alberta, Canada) using

a flame ionization detector (FID). FAMEs were separated on a 50 m column (Agilent J&W CP-

Sil 88 for FAME, Santa Clara, California, USA) using hydrogen as the carrier gas. Peaks were

identified against an external standard (GLC 455 and GLC 37 Nu-chek Prep, Inc., Elysian,

Minnesota, USA.

*Oceanographic sampling*

CTD profiles were collected every one to two weeks to characterise the regional oceanographic

conditions. CTD stations were in close proximity (0-2 km) to the salmon sampling stations (Fig.

S1A). An RBR maestro or a SeaBird 19plus V2 CTD was used (Halverson et al., 2017). For the

purpose of this study, we present temperature ($^o$C) and salinity (psu) data integrated over the

upper 10 m of the water column, where salmon were captured.

*Zooplankton sampling*

Zooplankton were collected approximately every fortnight in the NSoG, and every 4-7 days in

JS and QCS (Fig. 1). A Bongo net (250 µm mesh size, mouth diameter 0.5 m) was deployed

vertically from near-bottom or a maximum of 300 m depth to the surface. Filtered water volume

during each tow was measured using a General Oceanics mechanical flowmeter. Collected

zooplankton were preserved in 5% buffered formaldehyde-seawater solution. Zooplankton were

identified to the lowest taxonomic level possible, counted, and taxon specific biomass was

estimated by multiplying the mean dry weight (DW) of each species and stage by their

abundance (Mahara et al., 2018). Zooplankton abundance and biomass data were integrated

over the water column and presented as ind. $m^{-2}$ and mg dry weight (DW) $m^{-2}$. Since the

zooplankton samples were integrated over 0-300 m depth, they overestimate the available

zooplankton biomass and abundance for juvenile salmon, given that they likely forage in the top

30 m. Therefore, for the purpose of this study we only included zooplankton species that had

been demonstrated to be consumed by the juvenile sockeye salmon in 2015 and 2016 (see

species list Table S2, based on James (2019)).

*Statistical analyses*

Juvenile sockeye salmon length-weight residuals were calculated for all collected fish that were

used for RNA/DNA and fatty acid analyses from both 2015 and 2016 (n=187 out of n=1,233)

based on ln(weight) to ln(fork length) residuals from ordinary least squares linear regression

(Reist, 2011). Positive length-weight residuals indicate that the salmon was heavier than

expected for their fork length. Multivariate tests were performed using the R package vegan

(Oksanen et al., 2017). Two-way ANOVA for unbalanced data (type II Sum of Squares) was

used to individually test differences between years and regions for the response variables:

temperature and salinity from CTD measurements, zooplankton biomass, length-weight

residuals, RNA/DNA, total FA (TFA), saturated FA (SFA), monounsaturated FA (MUFA),

polyunsaturated FA (PUFA), EPA, and DHA. If the interaction term was non-significant, the

model was reduced by removing the interaction term and analyzed the main effects. Prior to this

analysis, normal distribution was tested using a Shapiro-Wilk test and transformed data when

necessary to achieve normal distribution (log for temperature and zooplankton biomass, and

RNA/DNA). Year and sampling region were set as fixed explanatory factors. Tukey HSD tests

were used for each significant ANOVA to distinguish region-to-region and year-to-year

differences. Differences were deemed significant where $p < 0.05$.

**Results**

Overall, we found that RNA/DNA and length-weight residuals, supported an energetic deficit

among migrating juvenile sockeye salmon in JS consistent with the TGH. We also found that

that in 2015 fish were in better nutritional condition when entering JS compared to 2016 despite

higher zooplankton biomass in 2016. Additionally, fish sampled in 2015 at QCS showed a two-

fold increase in RNA/DNA from JS whereas in 2016 the RNA/DNA only marginally increased

from 5.06 to 5.41. The fatty acid profiles showed that NSoG fish had significantly higher initial

FA concentrations in 2015 than 2016, although the entry FA concentrations were different in

between both years, all FAs groups significantly decreased in JS in both years.

*Juvenile salmon condition*

Fish length-weight residuals differed significantly between years ($p < 0.001$, Table 2) but not

between regions ($p=0.8$, Table 2). Overall, juvenile salmon were significantly heavier than

expected based on their length in 2015, whereas in 2016 juveniles were lighter than expected

(Fig. 2A). Juvenile salmon had significantly higher length-weight residuals in NSoG in 2015

compared to 2016 (Tukey HSD, $p <0.01$) and also showed significantly higher length-weight

residuals in JS in 2015 than 2016 (Tukey HSD, $p<0.05$), although no difference was detected

between both years in QCS (Tukey HSD, $p=0.14$, Fig. 2A).

Juvenile salmon RNA/DNA differed significantly by both region and year ($p <0.001$) and the

interaction of year and region ($p = 0.02$, Table 2). Across both years, mean RNA/DNA was the

lowest in JS ($5.82 \pm 0.41$), the highest in QCS ($9.01 \pm 1.01$), and intermediate in the NSoG ($5.97$

$\pm 0.38$; Fig. 2B). However, RNA/DNA was significantly higher (Tukey HSD, $p > 0.001$) in 2015

($7.90 \pm 0.61$) than 2016 ($4.96 \pm 0.24$). In the NSoG, juvenile salmon had a significantly lower

(Tukey HSD, $p = 0.003$) RNA/DNA in 2016 ($4.39 \pm 0.28$) than 2015 ($8.15 \pm 0.39$) and exited

QCS with significantly lower (Tukey HSD, $p = 0.001$) RNA/DNA in 2016 ($5.41 \pm 0.76$) than 2015

(12.59 $\pm$ 1.27). In 2015, RNA/DNA decreased from the NSoG to JS (Tukey HSD, p = 0.07), and

increased significantly from JS levels to QCS levels (Tukey HSD, p < 0.001). But in 2016, fish

arrived in NSoG with low RNA/DNA, and decreased further in JS and only increased slightly in

QCS (Tukey HSD, p = 0.98). The increase in RNA/DNA from JS to QCS was ~ 83.98% in 2015

and ~6.47% in 2016. There was no significant correlation between fish RNA/DNA ratio and fish

size (R = 0.07, p = 0.34, df = 163, Table S3). Genetic stock identification showed that 98% of

the 2015 and 100% of 2016 were Fraser River salmon (Table S4).

*Zooplankton*

The biomass of zooplankton species that were known to be consumed by juvenile sockeye

salmon in the study region was significantly higher (F = 7.99, p < 0.01; Table 2; Fig. 3A) in QCS

(66.12 $\pm$ 10.08 mg dry weight (DW) m$^{-2}$, Tukey HSD: p < 0.001) than in JS (39.72 $\pm$ 4.02 mg DW

m$^{-2}$) and the NSoG (37.84 $\pm$ 5.53 mg DW m$^{-2}$) (Table 2, Fig. 3A). Zooplankton biomass was

significantly lower (F = 11.14, p < 0.01) in 2015 (37.50 $\pm$ 3.92 mg DW m$^{-2}$) than 2016 (57.93 $\pm$

7.12 mg DW m$^{-2}$). Contrary to biomass, the abundance of zooplankton prey species for juvenile

sockeye salmon (Table S2) was the highest in JS (346,755 $\pm$ 27,264 ind m$^{-2}$) and the lowest in

the QCS (201,159 $\pm$ 16,478 ind m$^{-2}$), and NSoG (217,241 ind $\pm$ 39,960 ind m$^{-2}$, Table 2, Fig.

3B). Zooplankton abundance was significantly higher in 2016 (F = 12.62, p <0.001, 307,546 $\pm$

26,822 ind m$^{-2}$) than 2015 (232,377 $\pm$ 26,822 ind m$^{-2}$; Table 2). Average individual zooplankton

size calculated as total prey species biomass divided by total prey species abundance (Fig. S2)

was significantly different between regions (F = 8.148, p < 0.001), with the smallest average

zooplankton size in NSoG (2.52µg DW ind$^{-1}$ $\pm$ 6.10) and largest size in QCS (5.74 µg DW ind$^{-1}$ $\pm$

11.13). There was no difference in average zooplankton size between NSoG and JS. The

composition of available zooplankton prey species for juvenile sockeye salmon also differed

considerably between the sampled regions. The NSoG and QCS had more large crustacean

species (Amphipoda, Decapoda) than JS (dominated by Copepoda, Cirripedia; Fig. S3).

*Nutritional status*

Total FA and the FA groups SFA, MUFA, PUFA, DHA+EPA had a significant interaction term of

year and region and showed comparable distribution patterns (Fig. 4A-E; Table 2). In 2015,

juvenile salmon had the highest FA levels in the NSoG (Fig. 4 A-E) and showed a significant

decrease in FA concentrations in JS and then slightly increased in QCS. In 2016, FA levels in

NSoG fish were significantly lower than in 2015 ($p < 0.05$), increased slightly but not significantly

in JS fish to levels similar to 2015 in JS, and remained constant in QCS fish (Fig. 4 A-E).

**Discussion**

Juvenile salmon encounter diverse ocean conditions during their early marine life. Research on

juvenile Fraser River sockeye salmon has measured significant regional and temporal

differences in growth (IGF-1) during their migration along the British Columbia Coast (Ferriss et

al., 2014; Journey et al., 2018). However, the potential link between juvenile salmon growth and

fish nutritional condition or oceanographic conditions was not explored. In this study, we

explicitly linked the body condition of juvenile salmon (RNA/DNA) to their nutritional condition

(fatty acids) and prey availability (zooplankton biomass). RNA/DNA and fatty acid data showed

that body condition and nutritional condition of juvenile sockeye salmon changed rapidly over a

~ 120 km section of their migration route. These variations appeared to be partially due to

differences in zooplankton availability, which was consistently low in Johnstone Strait (JS)

supporting the Trophic Gauntlet Hypothesis (TGH) of McKinnell et al. (2014), but also to

differences in prey quality. Importantly, the response of juvenile salmon to periods of starvation

was affected by their condition upon entering poor foraging areas, as was their subsequent

recovery when leaving food poor areas. Below we discuss the role of coastal oceanographic

variation in determining fish condition, the cumulative effect of fish encountering a mosaic of

habitats during migration, and the importance of taking into account spatial variability in

migration habitat when considering salmon response to climate change.

In our study system, the majority of outmigrating juvenile sockeye salmon travel northwards

from the Fraser River through the Strait of Georgia (SoG) before entering a section of tidally

mixed channels in the Discovery Islands (DI) and JS (Furey et al., 2015). In the SoG, surface

waters are relatively warm, with low salinity and stratified, which is correlated with high

phytoplankton biomass, high zooplankton biomass and high mean zooplankton size (James et

al., 2020; Mahara et al., 2021). Ocean conditions become cooler and saltier as the fish move

north into the DI, JS, and Queen Charlotte Strait (QCS). This change in ocean conditions is

driven by the deep tidal mixing that brings cold and salty deep water to the surface (McKinnell et

al., 2014). Phytoplankton biomass is low through the tidally mixed northern Discovery Islands

and Johnstone Strait, leading to both low zooplankton biomass and smaller zooplankton size.

This pattern is, however, reversed in QCS.

Across both years of this study, juvenile sockeye salmon RNA/DNA ratios supported the TGH,

with high fish condition in the NSoG and QCS, and low condition in JS. The link between

RNA/DNA and feeding success is supported by an extensive juvenile sockeye salmon stomach

fullness analysis from the same region, which found high stomach fullness in the NSoG (Gut

Fullness Index (GFI) = 1-2 %), low fullness in DI (GFI = 0.2 %), low fullness in JS (GFI = 0.48

%), and full stomachs in QCS (GFI = 3.5%) (James et al., 2020). Length-weight residuals

showed a weaker response over the migration route, reflecting their longer integration times and

highlighting the importance of more sensitive growth proxies, with faster turnover, to understand

the response of fish to small scale regional changes in conditions. Notably, RNA/DNA ratios and

length-weight residuals were significantly lower in 2016 than 2015, despite higher zooplankton

biomass and abundance in that year, indicating that another factor was contributing to juvenile

salmon nutritional and body condition. The low RNA/DNA ratios in NSoG fish in 2016 were also

reflected in low length-weight residuals, suggesting that feeding conditions may have been poor

through most of the SoG in that year.

Corresponding with the observed changes in growth proxies, there were significant differences

in juvenile sockeye salmon fatty acid profiles both spatially and between years. Stored lipids and

fatty acids in fish tissues are important to mobilize energy from muscle, liver and viscera by FA

β-oxidation processes during starvation phases (Halver and Hardy, 1989). During starvation,

saturated FA (SFA) are the first FAs to be mobilized for ATP production, followed by

monounsaturated FA (MUFA) and then polyunsaturated FA (PUFA) (Wen et al., 2006). In 2015,

we detected a decrease of TFA, SFA, MUFA, and PUFA along the migration corridor, with

lowest concentrations in JS, indicating that fish had oxidized FAs stored in white muscle tissue

for ATP production. In 2016, juvenile salmon sampled in the NSoG had lower concentrations of

TFA, SFA, MUFA, and PUFA than 2015, indicating that the nutritional condition of salmon was

already very low when they reached the NSoG, with less FAs stored in white muscle tissue as

an energy reserve for starvation periods. Two potential scenarios may have resulted in FA

differences of NSoG salmon in 2015 and 2016: (1) food quantity was low in the Strait of Georgia

and FA reserves for mobilizing energy had already been depleted, or (2) the food quality of the

available food in SoG was lower in 2016. Zooplankton surveys showed that biomass was in fact

significantly higher in 2016 than 2015 in the NSoG (Perry et al., 2021), suggesting that lower

food quality (DHA + EPA) in 2016 was the driver of both lower fish FA content and RNA/DNA

ratios in that year. As outlined in the introduction, low DHA and EPA availability in prey have

important implications for fish condition, including reduced growth rates, visual maldevelopment

with negative effects on hunting efficiency, an increase in FAs that are considered as pro-

inflammatory (Table 1), and reduced capacity for compensatory growth (Ballantyne et al., 2003;

Bou et al., 2017).

Compensatory growth is a phase of accelerated growth that follows a period of limited growth

once non-limiting conditions are encountered (Ali et al., 2003). When fish re-enter non-limited

food conditions, food intake increases, cell mitosis is accelerated, and food utilization is

enhanced. Compensatory growth in fish can either result in partial, total or excess body mass

recovery, and return to their normal growth rate values (Ali et al., 2003). RNA/DNA ratios in

QCS from 2015 showed compensatory growth in juvenile salmon after passing through JS,

where RNA/DNA were low and FA concentrations were lower due to needed energy

mobilization. The RNA/DNA ratio almost doubled from JS to QCS in 2015. However, in 2016

when juvenile salmon already had low condition and depleted energy reserves in the NSoG, no

compensatory growth was observed in in RNA/DNA ratios. Studies with Atlantic salmon have

shown that compensatory growth success depends on the quality of food available (Johansen et

al., 2001). Johansen et al. (2001) showed in an experimental study that Atlantic salmon were

not able to fully compensate for growth loss on a lipid poor diet but were with a lipid rich diet.

Turchini et al (2007) proposed the term 'lipo-compensatory growth' after showing that Murray

cod were able to compensate starvation and refuel fatty acid storage when fed with high fatty

acid diets.

As DHA and EPA are important biomolecules for growth and are only taken up through the diet,

high prey quality or high availability of DHA and EPA while re-feeding is necessary to support

needs during enhanced growth and mitosis. El-Sabaawi et al. (2009a) showed substantial

variability in FA content among zooplankton species in the SoG. Large calanoid copepods, such

as *Neocalanus plumchrus* and *Calanus marshallae*, and amphipods have higher DHA+EPA

concentrations than other species. Our data showed that the JS zooplankton community was

dominated by small copepod species, which tend to have low FA content (Hiltunen et al., 2021).

Both the NSoG and QCS had a high abundance of larger copepods, euphausiids, and

amphipods, with the latter two groups both having relatively large size (> 10 mm) and high DHA

and EPA content (Hiltunen et al., 2021), i.e., prey was more nutritious in the NSoG and QCS

than JS. This observation supports prey quality as an important factor in the overall better

nutritional health of fish in the NSoG and QCS. Furthermore, Galbraith et al (2017) showed that

a higher proportion of southern copepod species (*Acartia sp.*, *Clausocalanus sp.*, *Calocalanus*

*sp.*, *Ctenocalanus sp.*, *Mesocalanus sp.*, *Paracalanus sp.*) and chaetognath species

(*Mesosagitta minima*, *Serratosagitta bierii*, *Parasagitta euneritica*) occurred in the NSoG in

2016, but more local zooplankton species such as *Neocalanus plumchrus* and *Euphausia*

*pacifica* in 2015. This shift in species composition between 2015 and 2016 may account for the

lower FA content of juvenile salmon in the NSoG in 2016.

The better body and nutritional condition of juvenile salmon leaving the NSoG in 2015 appeared

to compensate for low food availability in JS. This supports the importance of good foraging

conditions, with high quality food, for enduring periods of poor foraging conditions. Protracted

periods of starvation may lead to not only slower recovery from starvation periods but may also

have long-term impacts on the physiology of the fish (Nikki et al., 2004). Our data showed that

fish with higher RNA/DNA ratio in 2015 had a higher recovery success in QCS than 2016.

Starvation periods longer than 3-4 weeks can decrease digestive enzyme activity, inhibiting the

ability of fish to sufficiently digest food when available again, resulting in overall lower body size

and higher mortality (Abolfathi et al., 2012). This is highly relevant to juvenile salmon migrating

through the BC coastal ocean. The migration of juvenile sockeye salmon through the DI and JS

takes approximately 14 days, while time spent in the SoG is estimated to be between 30 and 50

days (Preikshot et al., 2012). Thus, as suggested by McKinnell et al (2014), anomalously poor

foraging conditions in the SoG and/or QCS, coupled with typically poor foraging conditions in

the DI and JS, may have an additive effect that impacts early marine survival of juvenile salmon.

This study empirically links foraging conditions to the nutritional health of juvenile salmon and,

further, provides evidence that low nutritional health can reduce recovery when entering good

foraging areas.

FAs that are important to food quality are produced at the base of the food web by

phytoplankton and transferred through trophic interactions to higher trophic levels. But also, FA

composition varies between zooplankton species resulting in changes of FA availability with

zooplankton species changes for predators such as juvenile salmon. Future environmental

change leading to more abundant southern copepod species can directly affect juvenile salmon

body and nutritional condition during their migration even though zooplankton biomass might not

be affected. Therefore, it is important to consider food quality effects on salmon condition

especially for making predictions or modelling approaches.

**Conclusions**

This study demonstrates that regional differences in prey quantity and quality have important

implications for the growth and health of juvenile sockeye salmon during their early marine

migration. Poor growth through Johnstone Strait supports the Trophic Gauntlet Hypothesis of

McKinnell et al (2014), and reduction in health due to poor foraging conditions. Despite similar

food quantity in the NSoG in 2015 and 2016, juvenile sockeye salmon emerging from SoG in

2016 had low fatty acid concentrations, potentially due to a higher abundance of low-quality

southern zooplankton prey. The lower nutritional health of fish in 2016 had a negative effect on

their recovery from starvation after passing though the DI-JS corridor where food availability is

consistently low. We propose that the outmigration corridor for juvenile salmon should be

defined as a series of interconnected gauntlets which may be differently affected by inter-annual

variation driven by large-scale climate events. The recovery of juvenile salmon from regions with

naturally poor foraging conditions may depend on the availability of prey in adjacent areas and

be negatively affected by extreme climate events and climate change that reduce the quality

and quality of prey in historically productive areas.

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**Author contributions**

**JG** - Laboratory and Statistical analysis, Visualization, Writing and Editing; **BPVH** – Project Conceptualization, Methodology, Writing and Editing; **SG** – Project Conceptualization, Methodology, Field program, Investigation, Editing; **IF, BTJ** - Project Conceptualization, Methodology, Field program, Investigation, Editing **MK** - Project Conceptualization, Methodology, Field program, Investigation, Editing; **NM** - Field program, Zooplankton Data, Editing; **EAP** - Project Conceptualization, Methodology, Editing; **LR** - Project Conceptualization, Methodology, Field program, Investigation, Editing**.**

**Additional Information**

**Supplementary information** accompanies this paper

**Competing financial interest**: The authors declare no competing financial interest

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**Figure 1** Map of sampling locations (black points) for juvenile salmon. Encircled points indicate

the sampling regions, blue point indicates Fraser River outflow, and blue arrows indicate the

primary Fraser sockeye salmon migration route. Inset map shows the location of the study area

on the British Columbia coast. Figure created with R package 'PBSmapping' (Schnute et al.,

2021)[68].

**Figure 2** Annual and regional mean of A) length-weight residuals of fish sampled for RNA/DNA

(black; n=162) and all fish sampled during the study (grey, n=1,233), B) RNA / DNA ratio. NSoG

= northern Strait of Georgia; JS = Johnstone Strait; QCS = Queen Charlotte Strait. Error bars

denote $\pm$ SE

**Figure 3** Annual and regional A) biomass and B) abundance of zooplankton > 1 mm in length.

(error bars denote $\pm$ SE). NSoG = northern Strait of Georgia; JS = Johnstone Strait; QCS =

Queen Charlotte Strait; 2015 is indicated by circles and 2016 by triangles.

**Figure 4** Annual and regional mean fatty acid concentrations (µg FA per mg dry weight of white

muscle tissue) A) total fatty acids (TFA), B) saturated fatty acids (SFA), C) monounsaturated

fatty acids (MUFA), D) polyunsaturated fatty acids (PUFA), and E) DHA + EPA

(docosahexaenoic acid + eicosapentaenoic acid). Error bars denote $\pm$ SE. NSoG = northern

Strait of Georgia; JS = Johnstone Strait; QCS = Queen Charlotte Strait.

**Figure 1**

**Figure 2**

778    **Figure 3**

786    **Table 1** Summary table of fatty acid groups, names and function.

| Fatty acid group | Fatty acid name | Fatty Acid Abbreviation | Fatty acid function | Reference |
|---|---|---|---|---|
| **Highly unsaturated fatty acids** | docosahexaenoic acid | DHA; C22:6n-3 | Cell membranes, pigmentation, Eye development, brain development, gene expression, Neural cell development, Prostaglandin production | Parrish (2009) and references within |
| | eicosapentaenoic acid | EPA; C20:5n-3 | | |
| | arachidonic acid | ARA; C20:4n-6 | | |
| **Saturated Fatty Acids (SFA)** | | | SFAs can be mobilized for ATP production | Wen et al. (2006) and references within |
| **Monounsaturated Fatty Acids (MUFA)** | | | MUFAs can be mobilized for ATP production | Wen et al. (2006) and references within |
| **Polyunsaturated Fatty Acids (PUFA)** | | | PUFAs can be mobilized for ATP production | Wen et al. (2006) and references within |

| | | | Growth, survival, metamorphosis, pigmentation | Copeman et al. (2002) and references within |
|---|---|---|---|---|
| | | C16:0 (SFA), C16:1 (MUFA), C18:1n-9 *cis* (MUFA), C18:2n-6 *cis* (PUFA) | Biomarker for starvation in fish | Torstensen et al. (2000) and references within |

**Table 2.** Results of two-way ANOVA explaining the effects of region, year, and the interaction

(x) of region and year. Values in bold are significant at p < 0.05. LL and UL represent the lower-

limit and upper-limit of the partial $\eta^2$ confidence interval, respectively.

| | Predictor | Sum of Squares | df | Mean Square | F | p | partial η2 | partial η2 90% CI [LL, UL] |
|---|---|---|---|---|---|---|---|---|
| *Length-weight residuals* | Year | 0.69 | 1 | 0.69 | 30.58 | **<0.001** | 0.16 | |
| | Region | 0.01 | 2 | 0.01 | 0.22 | 0.80 | 0.00 | [0.00, 0.02] |
| | Year x region | 0.06 | 2 | 0.03 | 1.30 | 0.27 | 0.02 | [0.00, 0.05] |
| | Error | 3.51 | 156 | 0.02 | | | | |
| | | | | | | | | |
| *RNA/DNA ratio (log)* | Year | 6.54 | 1 | 6.54 | 23.57 | **<0.001** | 0.12 | |
| | Region | 4.70 | 2 | 2.35 | 8.47 | **<0.001** | 0.09 | [0.03, 0.15] |
| | Year x region | 3.27 | 2 | 1.64 | 5.88 | **0.003** | 0.06 | [0.02, 0.12] |
| | Error | 50.23 | 181 | 0.28 | | | | |
| | | | | | | | | |
| *Zoopl. Biomass (DW* | Year | 2.48 | 1 | 2.42 | 11.14 | **<0.01** | 0.15 | |

| | | | | | | | | |
|---|---|---|---|---|---|---|---|---|
| *mg⁻¹ m⁻²)* *(log+1)* | | | | | | | | |
| | Region | 3.48 | 2 | 1.74 | 7.99 | **<0.01** | 0.20 | [0.06, 0.32] |
| | Error | 13.71 | 63 | 0.22 | | | | |
| | | | | | | | | |
| *Zoopl. Abundance (ind m⁻²) (log+1)* | Year | 179917805492.14 | 1 | 179917805492.14 | 12.62 | **<0.001** | 0.17 | |
| | Region | 413096728150.15 | 2 | 206548364075.08 | 14.49 | **<0.001** | 0.32 | [0.15, 0.43] |
| | Error | 897886080448.52 | 63 | 14252160007.12 | | | | |
| | | | | | | | | |
| *TFA (µg mg⁻¹ DW) (log+1)* | Year | 0.20 | 1 | 0.20 | 0.19 | 0.66 | 0.00 | |
| | Region | 3.19 | 2 | 1.59 | 1.54 | 0.22 | 0.02 | [0.00, 0.05] |
| | Year x region | 10.38 | 2 | 5.19 | 5.02 | **0.008** | 0.05 | [0.01, 0.11] |
| | Error | 187.28 | 181 | 1.03 | | | | |
| | | | | | | | | |
| *SFA (µg mg⁻¹ DW)* | Year | 0.37 | 1 | 0.37 | 0.41 | 0.521 | 0.01 | |

| | | | | | | | | |
|---|---|---|---|---|---|---|---|---|
| *(log+1)* | | | | | | | | |
| | Region | 2.85 | 2 | 1.43 | 1.66 | 0.202 | 0.02 | [0.01, 0.05] |
| | Year x region | 8.75 | 2 | 4.38 | 4.95 | **0.008** | 0.05 | [0.01, 0.11] |
| | Error | 159.93 | 181 | 0.88 | | | | |
| | | | | | | | | |
| *MUFA (µg mg$^{-1}$ DW) (log)* | Year | 0.04 | 1 | 0.04 | 0.03 | 0.863 | 0.01 | |
| | Region | 1.55 | 2 | 0.78 | 0.61 | 0.544 | 0.01 | [0.00, 0.03] |
| | Year x region | 12.23 | 2 | 6.12 | 4.83 | **0.009** | 0.05 | [0.01, 0.10] |
| | Error | 229.23 | 181 | 1.27 | | | | |
| | | | | | | | | |
| *PUFA (µg mg$^{-1}$ DW) (log+1)* | Year | 0.24 | 1 | 0.24 | 0.24 | 0.625 | 0.01 | |
| | Region | 3.46 | 2 | 1.73 | 1.76 | 0.175 | 0.02 | [0.00, 0.06] |
| | Year x region | 9.45 | 2 | 4.72 | 4.82 | **0.009** | 0.05 | [0.01, 0.10] |
| | Error | 177.59 | 181 | 0.98 | | | | |
| | | | | | | | | |
| *DHA+EPA (log+1)* | Year | 0.33 | 1 | 0.33 | 0.36 | 0.551 | 0.00 | |
| | Region | .90 | 2 | 1.95 | 2.08 | 0.128 | 0.02 | [0.00, 0.06] |

| | | | | | | | | |
|---|---|---|---|---|---|---|---|---|
| | Year x region | 8.03 | 2 | 4.01 | 4.28 | **0.015** | 0.05 | [0.00, 0.10] |
| | Error | 169.59 | 181 | | | | | |

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