Adaptive diversification and niche packing on rugged fitness landscapes

Ilan N. Rubin[1,*], Yaroslav Ispolatov[2], and Michael Doebeli[1,3]

January 10, 2023

1. Department of Zoology, University of British Columbia, Vancouver, British Columbia V6T 1Z4, Canada;
2. University of Santiago of Chile (USACH), Physics Department, Av. Víctor Jara 3493, Santiago, Chile;
3. Department of Mathematics, University of British Columbia, Vancouver, British Columbia V6T 1Z4, Canada.
* Corresponding author; e-mail: rubin@zoology.ubc.ca.

Keywords: Rugged fitness, adaptive landscape, frequency-dependence, niche packing, coexistence, diversification.

Abstract

Explaining the emergence of diversity and the coexistence of competing types has long been one of the main goals of ecological theory. Rugged fitness landscapes have often been used to explain diversity through the presence of local peaks, or adaptive zones, in the fitness landscape acting as available niches for different species. Alternatively, niche-packing and theories based on limiting similarity describe frequency-dependent selection leading to the organic differentiation of a continuous phenotype space into multiple coexisting types. By combining rugged carrying capacity landscapes with frequency-dependent selection, here we investigate the effects of ruggedness on adaptive diversification and stably maintained diversity. We show that while increased ruggedness often leads to a decreased opportunity for adaptive diversification, it is the shape of the global carrying capacity function, not the local ruggedness, that determines the diversity of the ESS and the total diversity a system can stably maintain.

1 Introduction

The adaptive landscape has been a staple of evolutionary thinking since its introduction by Sewall Wright in 1932 (Wright, 1932). Wright proposed that the high-dimensional network of possible gene combinations can be approximately represented by a three-dimensional landscape with two unit-less axes representing genotypes and a third, vertical axis depicting fitness. This theory was later extended to phenotypic trait spaces (e.g., Lande and Arnold, 1983; Whitlock, 1995, 1997; Whitlock et al., 1995), in which the axes represent continuous phenotypic character traits, rather than Wright's unmeasurable genetic axes, and have largely been evoked interchangeably in subsequent evolutionary theory (e.g., Whitlock, 1995). These landscapes are recognized as a convenient and intuitive tool for visualizing and modelling evolutionary dynamics, despite much debate over the years (Coyne et al., 1997; Wade and Goodnight, 1998). In particular, Wright envisioned rugged fitness landscapes, or landscapes with many local "adaptive peaks" and valleys, and stochastic shifts of a population between peaks as the model for evolutionary dynamics and speciation (Barton and Charlesworth, 1984). For an in-depth discussion of the history of adaptive landscape theory, see Svensson and Calsbeek (2013).

In Wright's classical fitness landscape, fitness is a static quantity, immutably linked to the genotype or phenotype being modeled and based on characteristics of the given trait or environment. However, it has long been appreciated that the fitness of a given type is affected by the distribution of other types in the population (Clarke, 1979; Maynard Smith, 1982). If fitness is in this way frequency-dependent, the fitness landscape now becomes dynamic and a function of the population present.

While the literature of rugged fitness landscapes explores the dynamics of how a single species adapts to and evolves on that landscape, attempts to explicitly describe adaptive diversification or stable coexistence of multiple types with rugged fitness landscapes remain largely absent. These models are unable to explicitly explain the maintenance of biodiversity if they do not include frequency dependence, a necessary condition for stable coexistence of different types (Chesson, 2000).

In contrast to the rugged fitness landscape model, niche packing and coexistence theory explains the coexistence of many types in a given environment not through the presence of different adaptive peaks, but by the partitioning of a continuous phenotype space into niches (MacArthur, 1970; Roughgarden, 1976). In these models, ecotypes are defined by a continuous phenotype that is used to calculate that type's carrying capacity (the equilibrium population size of a monomorphic population or the frequency-independent portion of fitness). Phenotype space is then partitioned into distinct niches through frequency-dependent competitive interactions (Leimar et al., 2013a; Szabó and Meszéna, 2006). Classical formulations of the carrying capacity function are unimodal (often Gaussian) or flat. While a limited number of studies consider carrying capacity functions with multiple peaks (Doebeli, 1996; Leimar et al., 2013b; Ranjan

and Klausmeier, 2022; Ruxton et al., 2008; Sasaki and Ellner, 1995), the effects of ruggedness
in the underlying frequency-independent fitness landscape on frequency-dependent ecological
dynamics and multi-species coexistence remain largely unexplored.

Here we use randomly generated rugged fitness landscapes in continuous phenotype space
to explore how the emergence and maintenance of biodiversity mediated by negative frequency-
dependent ecological interactions is affected by the ruggedness of the underlying fitness land-
scape. Ecology is described by classic Lotka-Volterra dynamics based on competition on a single,
continuous phenotype axis and a simple trait substitution process is used to model evolution.
Species are defined by a one-dimensional phenotype that controls both the species' carrying
capacity (the frequency-independent part of fitness) and the frequency-dependent competition.
The carrying capacity function has "tunable" ruggedness (the number and steepness of local
peaks can be controlled), allowing us to explore the effect of increased ruggedness on diversity.
In doing so we show that while increased ruggedness does reduce the opportunities for diversi-
fication and hinder evolutionary movement in phenotype space, it has little effect on the overall
diversity a system can support.

2   Model

2.1   *Tunable rugged landscapes*

Here we introduce a model for generating tunable rugged fitness landscapes in continuous trait
space. As we consider ecological dynamics including frequency dependence, the landscapes
presented here represent the carrying capacity as a function of a continuous phenotype. The
carrying capacity can also be thought of as the static portion of fitness, while the frequency
dependence due to competition contributes the dynamic portion of fitness that is a function of
the species present in the community. The actual fitness landscape is therefore dependent on
both the underlying carrying capacity landscape as well as the biological community present
and can be measured as the invasion fitness (per capita growth rate of a rare mutant) landscape
for any situation. Importantly, carrying capacity is defined as the equilibrium population size of
a monomorphic population with a given phenotype and therefore not simply the stated fitness
of an arbitrary trait, but a measurable and biologically relevant quantity.

2.1.1   *Global carrying capacity function*

79 The carrying capacity landscape can be separated into the underlying global carrying capacity
80 function and the local ruggedness. While the global carrying capacity function can take any
81 form, here we use a quartic function $K_G(x) = \exp(\frac{-x^4}{4})$ for a given trait x. The quartic function
82 has two main advantages. First, the unimodal curve naturally restricts viable phenotypes to an

₈₃ area around the origin (approximately $[-2, 2]$ for the parameterization used here). Second, the
₈₄ relatively flat shape of the curve at the origin (compared to the Gaussian competition function,
₈₅ described below) ensures the presence of a branching point at the origin (when no ruggedness
₈₆ is considered) and therefore the possibility for the coexistence of multiple species (Baptestini
₈₇ et al., 2009; Doebeli, 2011). Additionally, using quartic carrying capacity functions avoids the
₈₈ pitfalls emerging from using a Gaussian carrying capacity function with Gaussian competition
₈₉ (see below), leading to structurally unstable continuous coexistence (Gyllenberg and Meszéna,
₉₀ 2004).

2.1.2 Ruggedness model

₉₂ To generate the local ruggedness we take a sum of m Gaussian distributions along the one-
₉₃ dimensional trait axis with equal spacing. The width of each Gaussian is defined by σ_R, while
₉₄ the height of each curve, h_k, is sampled from a Gaussian distribution: $h_k \sim \mathcal{N}(\mu = 0, \sigma = 1)$.
₉₅ From here on we will refer to σ_R as the period of the local ruggedness. To control the amplitude
₉₆ of the local ruggedness and ensure that there are no negative carrying capacities, the heights are
₉₇ re-scaled so that the deepest "valley" in the local ruggedness has an amplitude of A_R. The model
₉₈ becomes:

$$K(x) = \overbrace{\left(1 + A_R \frac{R(x)}{|min\,(R(x))|}\right)}^{\text{Local ruggedness}} \underbrace{\left(e^{-\frac{x^4}{4}}\right)}_{\text{Global landscape}} \tag{1}$$

$$R(x) = \sum_k^m h_k \frac{1}{\sqrt{2\pi}} e^{-\frac{(x-x_k)^2}{2\sigma_R^2}} \tag{2}$$

where

$$h_k \sim \mathcal{N}(\mu = 0, \sigma = 1)$$

$$x_k = x_{min} + (x_{max} - x_{min})\frac{k-1}{m-1}$$

₉₉ and x_{min} and x_{max} represent the extreme range of the modelled phenotype axis.

₁₀₀   This process results in a fitness landscape with local ruggedness that is "tunable" by two
₁₀₁ parameters: A_R, the amplitude of the lowest valley relative to the non-rugged carrying capacity
₁₀₂ function, and σ_R, the width of each Gaussian curve that contributes to the ruggedness. Of
₁₀₃ particular note, the number of Gaussian curves, m, used to generate the rugged portion of the
₁₀₄ landscape only controls the "resolution" of the ruggedness and is uncorrelated to the number
₁₀₅ of peaks in the final carrying capacity function (Fig. A1). For instance, for the parameters used
₁₀₆ here ($A_R \in [0, 1]$, $\sigma_R \in (0, 0.05)$), carrying capacity landscapes we generated had between 1 and
₁₀₇ 76 local peaks, despite each being created as the sum of exactly 250 (m) Gaussian distributions

108 spaced equally between -2 and 2 on the one-dimensional trait axis.

109 A unique carrying capacity landscape is randomly generated for each simulation. For exam-
110 ples of landscapes generated with various values of A_R and σ_R see Figure 1.

2.2   Ecological dynamics

We model ecology using classic Lokta-Volterra dynamics with a single continuous phenotype x.
All reproduction is assumed to be asexual. The population dynamics for species i is defined as:

$$\frac{dN_i}{dt} = rN_i \left(1 - \sum_j \left(\frac{N_j \alpha(x_i, x_j)}{K(x_i)} \right) \right) \tag{3}$$

where r is the intrinsic growth rate, N_i the population size of species i, x_i its phenotype, $K(x_i)$ its
carrying capacity, and $\alpha(x_i, x_j)$ the competitive effect of species j on species i. Here we assume
competition to be Gaussian, thus the strength of the interaction between two types is symmetric
and increases with their phenotypic similarity: $\alpha(x_i, x_j) = e^{-(x_i-x_j)^2/2\sigma_\alpha^2}$. σ_α determines the width
of the competition kernel. As σ_α decreases, the interaction strength between two different types
decreases.

Ecological dynamics are thus defined by the system of ODEs ($d\vec{N}/dt$) and were numeri-
cally integrated using the Runge-Kutta-Fehlberg method. Communities are considered to have
reached an ecological equilibrium when $dN_i/dt < 5 \times 10^{-8} \ \forall i$ and any species with a population
density $N_i < 10^{-8}$ is considered extinct. All final ecological communities were confirmed to be
feasible (i.e., there exists an equilibrium of all species present with strictly positive population
sizes) (Grilli et al., 2017) as further confirmation equilibrium was reached. We practically define
community diversity as species richness, or simply the number of distinct species present in the
community.

As we only consider competitive interactions (as opposed to predator-prey interactions or
mutualism, etc.), all communities have a globally stable equilibrium and are therefore indifferent
to initial population sizes (Hernández-García et al., 2009; Rubin et al., 2022). Likewise, as we only
consider 1-dimensional phenotypes and symmetric competition, evolutionary dynamics always
result in a stationary equilibrium (for discussions of non-stationary evolutionary dynamics see
Doebeli and Ispolatov, 2014, 2017; Rubin et al., 2021).

2.3   Evolutionary dynamics

133 Evolutionary dynamics are modelled as a trait substitution process. Under the assumption of
134 rare mutations, ecological dynamics are numerically integrated until equilibrium, at which point
135 a new mutant is introduced. Species with population size less than 10^{-8} are considered extinct
136 and removed. Mutants are derived from a randomly chosen parent with probability proportional

to their population size and are given a phenotype sampled from a Gaussian distribution centered at the parent phenotype and with variance σ_{mut}^2 ($x_m \sim \mathcal{N}(x_p, \sigma_{mut})$). For computational ease, mutant phenotypes were re-chosen if they were closer than $\sigma_{mut}/5$ to any resident in the population as populations with very similar species can be exceedingly slow to reach equilibrium. If the mutant has a positive growth rate in the resident population, it is added to the population and the ecological equilibrium is recalculated. Evolutionary simulations were run until 500 consecutive mutants were unable to invade, when an evolutionary equilibrium was declared. All evolutionary simulations were initiated with a single species with a randomly chosen phenotype between -2 and 2. As with the ecological simulations, community diversity is defined as species richness.

The parameterization used here is as follows: $r = 1$, $\sigma_\alpha = 0.6$, $\sigma_{mut}^2 = 0.01$, $x \in [-2, 2]$, and $m = 250$ (the number of Gaussian curves used to generate local ruggedness). $A_R \sim [0, 1]$ and $\sigma_R^2 \sim (0, 0.05]$.

3 Results

3.1 *Saturated ecological diversity*

In order to investigate how ruggedness in the carrying capacity landscape affects the expected number of species that can coexist in a given ecosystem, we conducted numerical simulation experiments saturating a given environment with an over-abundance of species (500 randomly chosen species) and allowing the community to naturally correct down to an equilibrium community. This is a simple, but classic, method of numerically approximating the ESS of a system (e.g., D'Andrea et al., 2019; Rael et al., 2018; Sasaki and Ellner, 1995). The ESS, or evolutionary stable state, of a system is defined as a community that when at ecological equilibrium is completely resistant to invasion (Edwards et al., 2018). This simulation method will inevitably slightly overestimate the ESS (e.g., of the 500 random species, none has the exact phenotype of a species in the ESS – instead, 2 species with similar phenotypes to that ESS species can coexist), but it is still a useful method to compare the diversity different systems can support. After each simulation, the invasion fitness landscape ($x \in [-2, 2]$) was calculated. All simulations concluded with less than 4% of the available phenotype space having a positive invasion fitness, verifying this algorithm in approximating the ESS diversity (Fig. A2).

Examples of the landscapes with their final equilibrium populations can be seen in Figure 1.

The expected diversity of the ESS of a system with a rugged carrying capacity landscape is relatively unaffected by the period of the ruggedness (σ_R) (Fig. 2b). However, for very small amplitude ruggedness ($A \lesssim 0.02$) the resulting diversity is slightly higher than for completely smooth carrying capacity landscapes. For landscapes with larger amplitude ruggedness the diversity of the ESS decreases. In general, systems with low amplitude ruggedness are able to

Figure 1: **Examples of rugged carrying capacity landscapes with ESS populations.** Each rugged carrying capacity landscape is shown in blue, the baseline smooth carrying capacity kernel in grey, and simulated ESS populations in red. Rows denote the amplitude of the local ruggedness (A_R between 1% and 50%) and columns denote square of the period of the local ruggedness ($\sigma_R^2 = 0.005$ to $\sigma_R^2 = 0.05$). Local ruggedness is generated randomly for each simulation. ESS are simulated by randomly selecting 500 species with phenotype $x \in [-2, 2]$ and running ecological dynamics until equilibrium (as defined by $\frac{dx_i}{dt} < 5 \times 10^{-8} \; \forall \; i$). For pairwise invasion diagrams of the same landscapes please see Fig. A3.

support a more diverse community than those with large amplitude ruggedness, regardless of the period of that local ruggedness.

3.2 Adaptive diversification

To test the effect of ruggedness on diversification dynamics we placed a single randomly chosen species in systems with the same randomly generated carrying capacity landscapes used previously. Evolution is then allowed to proceed naturally based on the trait substitution process described above. The simulation results in an evolutionarily convergent stable community in which no small-effect mutants can invade. For example communities that result from these evolutionary simulations see Figure A4.

(a) Amplitude of local ruggedness   (b) Period of local ruggedness

Figure 2: **Ecological diversity as a function of the local ruggedness.** The final diversity of ecological communities when initiated with a fully saturated initial community (500 species). The lines represent a Gaussian-weighted moving average of the final diversity for a given amplitude (A_R) or period (σ_R) of randomly generated local ruggedness. The shading represents the 95% confidence interval of the running mean (Gatz and Smith, 1995). The amplitude of the ruggedness is displayed on a log-scale. Between 250 and 500 simulations were run to generate each curve. Values of A_R and σ_R^2 were sampled randomly from $[0, 1]$ and $(0, 0.05]$ respectively. The global carrying capacity is a quartic function and the strength of competition, $\sigma_\alpha = 0.6$.

As expected for the competition kernel and carrying capacity used here, in the baseline non-rugged case, evolution always resulted in the 3 species ESS (Fig. 3b) (Rubin et al., 2022). Similarly to the saturated ecological dynamics, small amplitude ruggedness in the carrying capacity landscape resulted in a slightly increased expected equilibrium diversity. For larger amplitudes ($A_R \gtrapprox 0.02$), increasing the amplitude of the local ruggedness resulted in decreased diversity at the evolutionary equilibrium. For carrying capacity landscapes with an amplitude near 1, diversification did not occur and the final population remained only the single founding species.

While the period of ruggedness (σ_R) did not affect the saturated ecological equilibrium diversity, when considering the adaptive diversification from a single species, increasing σ_R leads to increasing diversity. For very small values of σ_R diversification does not occur, regardless of the amplitude of the ruggedness. For landscapes with small amplitude ruggedness, as σ_R increases, the evolutionary equilibrium diversity also increases until it plateaus at a slightly higher expected diversity than the non-rugged baseline case. For landscapes with larger amplitude ruggedness, increasing σ_R also results in increased expected evolutionary diversity, though we did not consider large enough values of σ_R for those curves to reach a maximal level of diversity.

(a) Amplitude of local ruggedness   (b) Period of local ruggedness

Figure 3: **Evolutionary diversity as a function of the width of local ruggedness.** The final diversity of ecological communities when initiated with a single, randomly chosen species that is then allowed to evolve. The lines represent a Gaussian-weighted moving average of the final diversity for a given amplitude (A_R) or period (σ_R) of randomly generated local ruggedness. The shading represents the 95% confidence interval of the running mean (Gatz and Smith, 1995). Of particular note, the line representing $A_R = 0$ was generated with simulations and is not simply a theoretically calculated value. All simulations with $A_R = 0$ resulted in 3 species, which is the ESS community for the non-rugged global carrying capacity function. The amplitude of the ruggedness is displayed on a log-scale. Between 250 and 500 simulations were run to generate each curve. Values of A_R and σ_R^2 were sampled randomly from $[0, 1]$ and $(0, 0.05]$ respectively. The global carrying capacity is a quartic function with $\sigma_K = 1$ and the strength of competition, $\sigma_\alpha = 0.6$.

4   Discussion

Theories based on rugged fitness landscapes and frequency dependent competition have proposed competing explanations for how diversity is generated and maintained. Rugged fitness landscape theory views local peaks as adaptive zones or available niches waiting to be filled. Evolution by natural selection (and in the asexual populations considered here) is a hill climbing process up the gradient in the fitness landscape to a local peak, at which point adaptation stalls until a stochastic "peak-shift" event. Such an event can be caused by drift in small populations (Wright, 1932), large-effect mutations (Whitlock et al., 1995), or migration (Barton and Charlesworth, 1984) – or more complicated deterministic dynamics like population variance induced peak shifts (Whitlock, 1995) or environmental change causing a shift in the location of the local peaks in phenotype space, viewing rugged landscapes as dynamic seascapes (Mustonen and Lässig, 2009; Simpson, 1953).

Theories of niche packing (Leimar et al., 2013a; MacArthur, 1970; Rubin et al., 2022), evolu-
tionary game theory (Maynard Smith, 1982), and adaptive dynamics (Doebeli, 2011; Geritz et al.,
1997) imagine sympatric speciation as an adaptive radiation that is a natural outcome of negative
frequency-dependent selection (competition) and selection for individuals to limit their similar-
ity to others in the population. These models are generally built on unimodal carrying capacity
functions that represent the frequency-independent, or static, portion of fitness and act as the
force of stabilizing selection. Evolutionary dynamics based on frequency-dependent competition
initially resemble the same hill-climbing process until the maximum of the carrying capacity
landscape is reached. While the maximum is a convergent equilibrium, it may be evolutionar-
ily unstable, i.e., prone to invasion by nearby mutants on either side, resulting in evolutionary
branching. The presence of a branching point is determined by the curvatures (as defined by
the second derivative) of the competition and carrying capacity functions at that point (or in the
one-dimensional case explored here $\left|\frac{K''(x^*)}{K(x^*)}\right| < \left|\frac{\partial^2 \alpha(x^*,y)}{\partial y^2}\right|_{y=x^*}$ with x^* representing the phenotype
of evolutionary equilibrium in question) (Dieckmann and Doebeli, 1999; Doebeli, 2011). In this
case, there will be disruptive selection at the peak driving the diversification of the population
into multiple types. Contrary to scenarios classically imagined through rugged fitness landscape
theory, now speciation does not require multiple peaks in a static fitness landscape.

By combining a rugged carrying capacity landscape with frequency-dependent competition
we consolidate the intuition and expectations derived from these two bodies of theory. We
use a model of classic Lotka-Volterra ecological dynamics with frequency-dependent selection
and an underlying rugged landscape for the carrying capacity function. These landscapes have
tunable ruggedness, allowing us to control the average curvature and number of local peaks,
while leaving the underlying global fitness landscape unchanged.

When considering both frequency-dependent selection and a rugged carrying capacity land-
scape, the evolutionary dynamics initially are the expected hill-climbing process. However, be-
cause selection is frequency-dependent, the actual fitness function does not strictly resemble the
rugged carrying capacity landscape. If a local peak and nearby valleys are shallow, the frequency-
dependent selection can "flatten" the peak in the dynamically determined fitness landscape, al-
lowing the population to deterministically cross the valley in question and continue to evolve up
the global carrying capacity gradient. Alternatively, in systems with very rugged carrying capac-
ity functions (whether because of large A_R or small σ_R), this local peak represents a convergent
and evolutionarily stable equilibrium and the endpoint of any deterministic evolutionary dy-
namics. If the landscape has some intermediate level of ruggedness such that the local peak does
represent an equilibrium but the carrying capacity is more broadly curved than the competition
function at that point in phenotype space, there will exist a convergent stable but evolutionarily
unstable equilibrium, i.e., a branching point, and the population will split. This pattern of diver-
sification will continue either until the environment is saturated or adaptation stalls because of

a valley in the carrying capacity function being too deep for frequency dependence to flatten, or because of a local peak that is too steeply curved for diversification to occur.

As the landscape becomes more rugged, more local peaks appear in the carrying capacity function and each local peak naturally becomes smaller and with a steeper curvature. Because the stability of any equilibrium in the evolutionary dynamics is dependent on the relative curvatures of the carrying capacity and competition functions at that point, smaller local peaks represent a more stable landscape that retards diversification (Fig. 3). Thus, when the carrying capacity is suitably rugged, the evolutionary dynamics resemble those predicted by Wright's shifting-balance theory. Directional selection drives the population up the carrying capacity gradient until it becomes trapped on a local peak and must then wait for a stochastic event (whether large mutation, small population-mediated drift, or something else) to cross any nearby area of low fitness (for a discussion of peak-shift dynamics see Coyne et al., 1997; Leimar et al., 2013b; Rouhani and Barton, 1987). With these rugged carrying capacity landscapes, large areas of phenotype space remain invadable but unreachable by mutations of small effect (Fig. A2). When considering frequency dependence these stochastic peak shifts can result in the stable maintenance of multiple types, formalizing the coexistence of many types on rugged fitness landscapes.

Unlike the evolutionary dynamics, for the ecological dynamics of fully saturated communities, the underlying global carrying capacity landscape is the greatest determinant of the level of diversity a system can support, not the nature or severity of the ruggedness of that landscape (Fig. 2). Increasing the amplitude of local ruggedness does result in a less diverse equilibrium community, though the effect is slight compared to the stark effect of ruggedness on diversification dynamics. The total number of local peaks in a landscape has little predictive power for the number of distinct types that the environment can stably support. In very rugged landscapes many of the local peaks in the carrying capacity landscape remain unoccupied and uninvasible, while in less rugged landscapes multiple types may share a single peak.

Interestingly, while increased ruggedness generally depresses diversity in a system relative to the smooth global landscape, small rugged perturbations can actually lead to an increase in community diversity. This can be clearly seen in the small amplitude diversification dynamics (Fig. 3a) and a slight effect in the small amplitude saturated ecological equilibrium (Fig. 2a). This small amplitude ruggedness is not large enough to hinder any evolutionary movement of the population in phenotype space, but does add just enough variation in the carrying capacity to allow diversity to be maintained at phenotypes not included in the smooth carrying capacity ESS.

Of course, a well-known critique of describing evolutionary dynamics on low-dimensional landscapes is that an area of low fitness in one dimension may be circumnavigated through adaptation in higher dimensions (Gavrilets, 1999). While we will not fully comment on higher

dimensional landscapes, the general results we focus on here largely stem from the relative curvature of the carrying capacity and competition functions and are thus likely replicated in higher dimensions. However, multi-dimensional evolutionary dynamics are notoriously complex and can result in cyclic and chaotic dynamics (Doebeli and Ispolatov, 2014, 2017; Rubin et al., 2021). How these complex evolutionary dynamics interact with rugged carrying capacity landscapes remains unexplored and could prove insightful to both the peak-shift and adaptive dynamics theories.

4.1 Conclusions

Our results affirm that suitably rugged carrying capacity landscapes generate similar diversification dynamics as predicted by peak-shift theory, but not the common narrative that the individual local peaks each represent a specific niche waiting to be filled. Increased carrying capacity ruggedness does impede evolutionary movement in phenotype space and retard diversification. However, the saturated ecosystem diversity is still largely a function of the width of the global niche space compared to the width of the competition, as predicted by niche-packing theory (Doebeli, 2011; Leimar et al., 2013a), rather than the local ruggedness.

Acknowledgements

YI acknowledges support from FONDECYT (The National Fund for Scientific and Technological Development of Chile) project no. 1200708 (`https://www.conicyt.cl/fondecyt/fondecyt-program/`).

MD acknowledges support from NSERC (The Natural Sciences and Engineering Research Council of Canada) grant no. 219930 (`https://www.nserc-crsng.gc.ca/NSERC-CRSNG/Index_eng.asp`)

Literature Cited

Baptestini, E.M., de Aguiar, M.A., Bolnick, D.I., Araújo, M.S., 2009. The shape of the competition and carrying capacity kernels affects the likelihood of disruptive selection. Journal of Theoretical Biology 259, 5–11. doi:10.1016/j.jtbi.2009.02.023.

Barton, N.H., Charlesworth, B., 1984. Genetic revolutions, founder effects, and speciation. Annual Review of Ecology and Systematics 15, 133–164. doi:10.1146/annurev.es.15.110184.001025.

Chesson, P., 2000. Mechanisms of maintenance of species diversity. Annual Review of Ecology and Systematics 31, 343–366. doi:10.1146/annurev.ecolsys.31.1.343.

Clarke, B.C., 1979. The evolution of genetic diversity. Proceedings of the Royal Society of London. Series B. Biological Sciences 205, 453–474. doi:10.1098/rspb.1979.0079.

Coyne, J.A., Barton, N.H., Turelli, M., 1997. Perspective: A critique of sewall wright's shifting balance theory of evolution. Evolution 51, 643–671. doi:10.1111/j.1558-5646.1997.tb03650.x.

D'Andrea, R., Riolo, M., Ostling, A.M., 2019. Generalizing clusters of similar species as a signature of coexistence under competition. PLOS Computational Biology 15, e1006688. doi:10.1371/journal.pcbi.1006688.

Dieckmann, U., Doebeli, M., 1999. On the origin of species by sympatric speciation. Nature 400, 354–357. doi:10.1038/22521.

Doebeli, M., 1996. A quantitative genetic competition model for sympatric speciation. Journal of Evolutionary Biology 9, 893–909. doi:10.1046/j.1420-9101.1996.9060893.x.

Doebeli, M., 2011. Adaptive Diversification (MPB-48). Princeton University Press. doi:10.2307/j.ctt7rgw4.

Doebeli, M., Ispolatov, I., 2014. Chaos and unpredictability in evolution. Evolution 68, 1365–1373. doi:10.1111/evo.12354.

Doebeli, M., Ispolatov, I., 2017. Diversity and coevolutionary dynamics in high-dimensional phenotype spaces. The American Naturalist 189, 105–120. doi:10.1086/689891.

Edwards, K.F., Kremer, C.T., Miller, E.T., Osmond, M.M., Litchman, E., Klausmeier, C.A., 2018. Evolutionarily stable communities: a framework for understanding the role of trait evolution in the maintenance of diversity. Ecology Letters 21, 1853–1868. doi:10.1111/ele.13142.

Gatz, D.F., Smith, L., 1995. The standard error of a weighted mean concentration—i. bootstrapping vs other methods. Atmospheric Environment 29, 1185–1193. doi:10.1016/1352-2310(94)00210-C.

Gavrilets, S., 1999. A dynamical theory of speciation on holey adaptive landscapes. The American Naturalist 154, 1–22. doi:10.1086/303217.

Geritz, S.A.H., Metz, J.A.J., Kisdi, É., Meszéna, G., 1997. The dynamics of adaptation and evolutionary branching. Physical Review Letters 78, 1–4.

Grilli, J., Barabás, G., Michalska-Smith, M.J., Allesina, S., 2017. Higher-order interactions stabilize dynamics in competitive network models. Nature 548, 210–213. doi:10.1038/nature23273.

Gyllenberg, M., Meszéna, G., 2004. On the impossibility of coexistence of infinitely many strategies. Journal of Mathematical Biology 50, 133–160. doi:10.1007/s00285-004-0283-5.

Hernández-García, E., López, C., Pigolotti, S., Andersen, K.H., 2009. Species competition: coexistence, exclusion and clustering. Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences 367, 3183–3195. doi:10.1098/rsta.2009.0086.

Lande, R., Arnold, S.J., 1983. The measurement of selection on correlated characters. Evolution 37, 1210–1226. doi:10.2307/2408842.

Leimar, O., Sasaki, A., Doebeli, M., Dieckmann, U., 2013a. Limiting similarity, species packing, and the shape of competition kernels. Journal of Theoretical Biology 339, 3–13. doi:10.1016/j.jtbi.2013.08.005.

Leimar, O., Tullberg, B.S., Mallet, J., 2013b. Mimicry, saltational evolution, and the crossing of fitness valleys, in: The Adaptive Landscape in Evolutionary Biology. Oxford University Press, pp. 259–270. doi:10.1093/acprof:oso/9780199595372.003.0016.

MacArthur, R., 1970. Species packing and competitive equilibrium for many species. Theoretical Population Biology 1, 1–11. doi:10.1016/0040-5809(70)90039-0.

Maynard Smith, J., 1982. Evolution and the Theory of Games. Cambridge University Press.

Mustonen, V., Lässig, M., 2009. From fitness landscapes to seascapes: non-equilibrium dynamics of selection and adaptation. Trends in Genetics 25, 111–119. doi:10.1016/j.tig.2009.01.002.

Rael, R.C., D'Andrea, R., Barabás, G., Ostling, A., 2018. Emergent niche structuring leads to increased differences from neutrality in species abundance distributions. Ecology 99, 1633–1643. doi:10.1002/ecy.2238.

Ranjan, R., Klausmeier, C.A., 2022. How the resource supply distribution structures competitive communities. Journal of Theoretical Biology 538, 111054. doi:10.1016/j.jtbi.2022.111054.

Roughgarden, J., 1976. Resource partitioning among competing species—a coevolutionary approach. Theoretical Population Biology 9, 388–424. doi:10.1016/0040-5809(76)90054-X.

Rouhani, S., Barton, N., 1987. Speciation and the "shifting balance" in a continuous population. Theoretical Population Biology 31, 465–492. doi:10.1016/0040-5809(87)90016-5.

Rubin, I., Ispolatov, Y., Doebeli, M., 2022. Ecological diversity exceeds evolutionary diversity in model ecosystems. bioRxiv doi:10.1101/2022.03.24.485441.

Rubin, I.N., Ispolatov, I., Doebeli, M., 2021. Evolution to alternative levels of stable diversity leaves areas of niche space unexplored. PLOS Computational Biology 17, e1008650. doi:10.1371/journal.pcbi.1008650.

Ruxton, G.D., Franks, D.W., Balogh, A.C.V., Leimar, O., 2008. Evolutionary implications of the form of predator generalization for aposematic signals and mimicry in prey. Evolution 62, 2913–2921. doi:10.1111/j.1558-5646.2008.00485.x.

Sasaki, A., Ellner, S., 1995. The evolutionarily stable phenotype distribution in a random environment. Evolution 49, 337. doi:10.1111/j.1558-5646.1995.tb02246.x.

Simpson, G.G., 1953. The Major Features of Evolution. Columbia University Press, New York.

Svensson, E., Calsbeek, R. (Eds.), 2013. The Adaptive Landscape in Evolutionary Biology. Oxford University Press. doi:10.1093/acprof:oso/9780199595372.001.0001.

Szabó, P., Meszéna, G., 2006. Limiting similarity revisited. Oikos 112, 612–619. doi:10.1111/j.0030-1299.2006.14128.x.

Wade, M.J., Goodnight, C.J., 1998. Perspective: The theories of fisher and wright in the context of metapopulations: when nature does many small experiments. Evolution 52, 1537–1553. doi:10.1111/j.1558-5646.1998.tb02235.x.

Whitlock, M.C., 1995. Variance-induced peak shifts. Evolution 49, 252. doi:10.2307/2410335.

Whitlock, M.C., 1997. Founder effects and peak shifts without genetic drift: Adaptive peak shifts occur easily when environments fluctuate slightly. Evolution 51, 1044. doi:10.2307/2411033.

Whitlock, M.C., Phillips, P.C., Moore, F.B., Tonsor, S.J., 1995. Multiple fitness peaks and epistasis. Annual Review of Ecology and Systematics 26, 601–629. doi:10.1146/annurev.es.26.110195.003125.

Wright, S., 1932. The roles of mutation, inbreeding, crossbreeding and selection in evolution, in: Proceedings of the VI International Congress of Genetics, pp. 356–366.

5   Appendix A: Supplementary figures

(a) Fixed $\sigma_R^2 = 0.02$ (b) Fixed $A_R = 0.1$

Figure A1: **The number of peaks in the carrying capacity landscape.** Three parameters used to generate the random ruggedness, the number of Gaussian curves that are laid out at equal intervals between -2 and 2 (m), the width of each Gaussian ($\sigma_R^2 \in \mathbb{R}_{\geq 0}$), and the amplitude of the lowest valley in the randomly generated ruggedness ($A_R \in [0, 1]$). The number of peaks in the rugged carrying capacity function is relatively indifferent to the number of Gaussian curves used in the generation function (m). The number of local peaks was determined numerically by breaking the rugged landscape into 1000 sequential bins. A peak was defined as any bin with a larger carrying capacity than both of its neighbors. For panel A σ_R^2 was set to 0.02 and A_R to 0.01, 0.05, 0.1, 0.25, or 0.5. For panel B A_R was set to 0.1 and σ_R^2 to 0.005, 0.01, 0.025, or 0.05. The number of Gaussians used in the algorithm, m, was then chosen randomly between 3 and 1000. Curves represent the running mean with Gaussian weighting and shading the 95% confidence interval of that running mean. 50 landscapes with different values of m were used to generate each curve. The global carrying capacity is a quartic function.

(a) Saturated ecology   (b) Evolution (Adaptive radiation)

Figure A2: **Proportion of phenotype space with positive invasion fitness after ecological saturation.** Histogram of the proportion of the phenotype space with positive invasion fitness at the end of the simulation. The ecological simulations starting with a saturated population are shown in panel A and the evolutionary simulations starting with a single species in panel B. Note the x-axis for panel A goes from 0 to 0.06, while the x-axis for panel B is from 0 to 1. Evolutionary simulations with a proportion of invasion fitness equal to 0 represent populations at the ESS. Evolutionary simulations with a very large proportion of the phenotype space open to invasion are due to local ruggedness restricting diversification and evolution in phenotype space.

Figure A3: **Examples pairwise invasion diagrams for rugged carrying capacity landscapes.** Pairwise invasion diagrams for the same landscapes included in figure 1. The x-axis represents the resident phenotype and y-axis the mutant phenotype. The black areas indicate mutants that have a positive growth rate when introduced to a monomorphic population of the resident. If assuming small mutations, the $y = x$ line is indicative of the evolutionary dynamics. If there is black above and white below $y = x$, selective pressure will push the population to evolve to a larger phenotype and vice-versa. Places where the area of positive invasion fitness crosses the diagonal indicate peaks and valleys in the landscape.

Figure A4: **Examples of rugged carrying capacity landscapes with final population after evolution starting from founding species.** Each rugged carrying capacity landscape is shown in blue, the baseline smooth carrying capacity kernel in grey, the initial population as a large hollow, red point, and the population after evolution as solid, red points. Rows denote the amplitude of the local ruggedness (A_R between 1% and 50%) and columns denote the period of the local ruggedness ($\sigma_R^2 = 0.005$ to $\sigma_R^2 = 0.05$). Local ruggedness is generated by adding together 250 Gaussian distributions that are laid out on an equal interval between -2 and 2, with the width defined by σ_R^2 and heights randomly chosen from a Gaussian distribution with $\mu = 1, \sigma = 1$. The generated local ruggedness is then re-scaled such that the minimum is equal to the amplitude and then multiplied by the global landscape kernel ($K(x) = exp(-x^4/4)$) Thus, an amplitude $A_R = 0$ results in no ruggedness while $A_R = 1$ means that the lowest valley in the landscape will have a carrying capacity equal to 0. An amplitude $A_R = 0.5$ means that the lowest valley in the landscape has a carrying capacity equal to half what it would be with a purely smooth carrying capacity function. A single species was randomly selected ($x \sim [-2, 2]$) as the initial species. Evolution was then simulated through a trait substitution process until 500 consecutive mutants were unable to invade. For pairwise invasion diagrams of the same landscapes please see Fig. A3.