

1 **Low diversity and microdiversity of comammox bacteria in wastewater systems suggests**  
2 **wastewater-specific adaptation within the *Ca. Nitrospira nitrosa* cluster.**

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4 Irmario Cotto<sup>1</sup>, Katherine J. Vilardi<sup>1</sup>, Linxuan Huo<sup>2</sup>, Emily C. Fogarty<sup>3</sup>, Wendell Khunjar<sup>4</sup>,  
5 Christopher Wilson<sup>5</sup>, Haydee De Clippeleir<sup>6</sup>, Kevin Gilmore<sup>7</sup>, Erika Bailey<sup>8</sup>, Sebastian Lucker<sup>9</sup>,  
6 Ameet J. Pinto<sup>2\*</sup>

7

8 1 Department of Civil and Environmental Engineering, Northeastern University, Boston, MA,  
9 USA

10 2 School of Civil and Environmental Engineering, Georgia Institute of Technology, Atlanta, GA,  
11 USA

12 3 Committee on Microbiology, The University of Chicago, Chicago, IL, USA

13 4 Hazen and Sawyer, Inc., New York, NY, USA

14 5 Hampton Roads Sanitation District, Norfolk, VA, USA

15 6 DC Water, Washington DC, USA

16 7 Department of Civil and Environmental Engineering, Bucknell University, Lewisburg, PA,  
17 USA

18 8 City of Raleigh Public Utilities, Raleigh, NC, USA

19 9 Department of Microbiology, RIBES, Radboud University, Nijmegen, the Netherlands

20

21 \*Corresponding author: [ameet.pinto@ce.gatech.edu](mailto:ameet.pinto@ce.gatech.edu)

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30 **Abstract**

31 Studies have found *Ca. Nitrospira nitrosa*-like bacteria to be the principal or sole comammox  
32 bacteria in nitrogen removal systems for wastewater treatment. In contrast, multiple populations  
33 of strict ammonia and nitrite oxidizers co-exist in similar systems. This apparent lack of diversity  
34 is surprising and could impact the feasibility of leveraging comammox bacteria for nitrogen  
35 removal. We used full-length 16S rRNA gene sequencing and genome-resolved metagenomics to  
36 compare population-level (i.e., species) diversity of comammox bacteria with that of strict  
37 nitrifiers in full-scale wastewater treatment systems and assess whether these observations were  
38 consistent or diverged at the strain-level. Full-length 16S rRNA gene sequencing indicated that  
39 while *Nitrosomonas*-like bacteria exhibited higher population-level diversity, the effective  
40 microdiversity of most *Nitrospira*-like bacteria were comparatively higher except for one  
41 *Nitrospira* Lineage II population. Comammox bacterial metagenome assembled genomes (MAGs)  
42 were associated with *Ca. Nitrospira nitrosa*. The average amino acid identity between comammox  
43 bacterial MAGs ( $93\% \pm 3$ ) across systems was significantly higher than that of the *Nitrosomonas*-  
44 like ammonia oxidizers ( $73\% \pm 8$ ) and the *Nitrospira*-like nitrite oxidizer MAGs ( $75\% \pm 13$ ),  
45 suggesting that the same comammox population was detected in all systems. Comammox bacteria  
46 and some ammonia oxidizers MAGs were significantly less microdiverse than most ammonia and  
47 nitrite oxidizers. Interestingly, strain-resolved analysis also indicates that different nitrogen  
48 removal systems harbor different comammox bacterial strains within the *Ca. Nitrospira nitrosa*  
49 cluster. These results suggest that comammox bacteria associated with *Ca. Nitrospira nitrosa* have  
50 low species- and strain-level diversity in nitrogen removal systems and may thus harbor specific  
51 adaptations to the wastewater ecosystem.

52

## 53 **Introduction**

54 Aerobic nitrification processes for nitrogen removal from wastewater are largely centered around  
55 biotransformation and growth kinetics of strict ammonia oxidizing bacteria (AOB) and nitrite  
56 oxidizing bacteria (NOB) (1). Since their discovery (2–5), several studies have detected  
57 comammox bacteria in a wide range of environmental and engineered systems (2–9), including  
58 nitrogen removal processes in laboratory and full-scale wastewater treatment bioreactors (9–20).  
59 Compared to oligotrophic engineered systems such as drinking water (9,21) and even tertiary  
60 treatment systems at full-scale wastewater treatment plants (WWTPs) (16) with low nitrogen  
61 concentrations, most studies have identified *Candidatus Nitrospira nitrosa*-like comammox  
62 bacteria as the principal or sole comammox bacteria in laboratory or full-scale secondary treatment  
63 systems (11,22,23). This apparent lack of diversity of comammox bacteria in secondary treatment  
64 processes is not only surprising but has the potential to impact treatment strategies centered around  
65 comammox bacteria (24,25) as functional outcomes reliant on low diversity communities may be  
66 less resilient to environmental fluctuations and perturbations.

67

68 *Nitrosomonas*-like AOB and *Nitrospira*-like NOB have been studied extensively in wastewater  
69 treatment systems and studies often report the coexistence of multiple populations (26–28) likely  
70 occupying different ecological niches (i.e., adapted to different conditions) within a complex  
71 community. Functional redundancy of multiple co-existing populations can confer stability to  
72 microbial ecosystems ensuring the long-term persistence (29,30). Despite the apparent low  
73 population diversity of comammox bacteria compared to strict nitrifiers, they have exhibited  
74 remarkable stability in secondary treatment systems (11,15,31). One plausible reason for their  
75 temporal persistence could be that comammox bacteria may harbor intra-population (i.e., strain)

76 level diversity as compared to population level (i.e., species). Previous literature has demonstrated  
77 that microdiversity within populations can not only allow them to adapt to rapidly changing  
78 conditions, but also contribute to their persistence (7,32). Further, small variations in metabolic  
79 capacity may underpin metabolic diversity and thus enable the co-existence of multiple closely  
80 related strains within a single population. *Nitrospira*-like bacteria have been shown to harbor high  
81 levels of microdiversity in drinking water (33,34) and wastewater (35) systems. Moreover, studies  
82 have reported that closely related co-existing *Nitrospira* strains can exhibit differences in substrate  
83 affinities, and utilization rates (35). Since, both levels of diversity (i.e., population and intra-  
84 population) have been associated with microbial persistence and ecosystem stability, it is plausible  
85 that the persistence of comammox bacteria may be associated with higher levels of microdiversity;  
86 this would be consistent with observations for *Nitrospira*-like bacteria (35).

87

88 To test this hypothesis, we used a combination of full-length 16S rRNA gene sequencing and  
89 hybrid assembly and binning approaches to systematically explore the population- and strain-level  
90 diversity of nitrifiers in three full-scale nitrogen removal systems with different process  
91 configurations. While 16S rRNA gene sequencing cannot provide information on genome content  
92 or metabolic potential, the single nucleotide resolution across the full length 16S rRNA gene can  
93 help determine differences in (micro)diversity among nitrifying populations (35,36). Nevertheless,  
94 16S rRNA gene sequences cannot be used to reliably distinguish comammox bacteria from strict  
95 *Nitrospira*-NOB. One approach to differentiate between the different guilds within *Nitrospira*  
96 bacteria and to analyze their diversity at high resolution is to leverage a genome resolved approach  
97 to obtain high quality metagenome assembled genomes (MAGs). Various methods exist to  
98 quantify diversity (e.g., average nucleotide identity [ANI] and average amino acid identity [AAI])

99 and microdiversity (e.g., average nucleotide identity from reads and nucleotide diversity from  
100 single nucleotide polymorphisms [SNPs]) within populations. Nevertheless, their applicability  
101 depends on the quality of MAGs which can be impacted when relying on short read sequencing  
102 (e.g., Illumina) in the presence of closely related strains (37); long-read sequencing on the  
103 Nanopore platform can help mitigate this constraint (38,39). However, this requires high levels of  
104 coverage to obtain polished consensus sequences with lower error rates compared to the raw data  
105 (37). Here, we utilize a hybrid metagenomic assembly approach, including long- and short-read  
106 data in an effort to assemble high quality MAGs for subsequent microdiversity analyses (37).

107

108 Thus, the overall objective of this study was to test the hypothesis that the widespread distribution  
109 and persistence of nearly identical comammox bacterial populations in multiple secondary  
110 wastewater treatment systems was due to the co-existence of multiple strains (i.e., high  
111 microdiversity). To accomplish this, we used a combination of full-length 16S rRNA gene  
112 sequencing and hybrid metagenome assembly and binning approach to systematically explore the  
113 population and strain level diversity of nitrifiers in three full-scale nitrogen removal systems with  
114 different process configurations. In doing so, we also aimed to assess factors that may influence  
115 the inter- and intra-population diversity of co-existing nitrifiers.

116

## 117 **2.0 Materials and Methods**

### 118 **2.1 Samples selection and processing**

119 Samples were selected from three nitrogen removal systems with different process configurations  
120 (i.e., sequencing batch reactor (SBR), Integrated Fixed Film Activated Sludge (IFAS), and  
121 Bardenpho 4-stages activated sludge system) with high concentrations of comammox bacteria in

122 our previous study (31). From this sample archive collected in 2017-2018, a total of thirty-three,  
 123 twenty, and six samples, respectively, were selected for full-length 16S rRNA gene sequencing on  
 124 the PacBio Sequel IIe platform, and short and long read metagenomic sequencing on the Illumina  
 125 NovaSeq and Nanopore MinION platforms, respectively (Table 1). Sample collection and  
 126 processing (including DNA extraction) and process data collection were described previously (31)  
 127 and are outlined in the supplementary text.

128 **Table 1: Overview of system process type and sub-type, operational scale and samples included in**  
 129 **this study and sample-specific sequencing strategy.**

<i>Site code</i>	<i>Process type</i>	<i>Process sub-type</i>	<i>Treatment stream</i>	<i>Selected samples (sampling month/sampling year)</i>
<b>GRE</b>	ND <sup>1</sup>	SBR <sup>2</sup>	Mainstream	06/17 <sup>+</sup> , 07/17 <sup>*†</sup> , 08/17 <sup>+</sup> , 09/17 <sup>*†</sup> , 10/17 <sup>*†</sup> , 11/17 <sup>+</sup> , 12/17 <sup>*†</sup> , 02/18 <sup>*†</sup> , 03/17 <sup>+</sup> , 04/18 <sup>*†</sup> , 05/18 <sup>+</sup> , 06/18 <sup>*†</sup>
<b>JAMMSM</b>	ND <sup>1</sup>	IFAS <sup>3</sup>	Mainstream	07/17 <sup>+</sup> , 09/17 <sup>*†</sup> , 10/17 <sup>+</sup> , 11/17 <sup>*†</sup> , 12/17 <sup>+</sup> , 01/18 <sup>*†</sup> , 02/18 <sup>+</sup> , 03/18 <sup>*†</sup> , 04/18 <sup>*†</sup> , 06/18 <sup>*†</sup>
<b>NEU</b>	ND <sup>1</sup>	Four/five stage	Mainstream	06/17 <sup>+</sup> , 07/17 <sup>*†</sup> , 08/17 <sup>+</sup> , 09/17 <sup>*†</sup> , 10/17 <sup>*†</sup> , 11/17 <sup>*†</sup> , 12/17 <sup>+</sup> , 01/18 <sup>*†</sup> , 02/17 <sup>+</sup> , 03/18 <sup>*†</sup> , 04/17 <sup>+</sup> , 05/18 <sup>*</sup> , 06/18 <sup>+</sup>

130 <sup>1</sup>Nitrification-denitrification, <sup>2</sup>Sequencing batch reactor, <sup>3</sup>Integrated Fixed-film Activated Sludge

131 \*Samples selected for Illumina NovaSeq sequencing

132 †Samples selected for PacBio full-length 16S rRNA gene sequencing

133 ‡Samples selected for nanopore sequencing on the MinION platform

134  
135

## 136 **2.2 PacBio full-length 16S rRNA gene sequencing and data analysis**

137 Sample DNA extracts (Table 1) were sent for full-length 16S rRNA gene sequencing at the Roy J.  
 138 Carver Biotechnology Center (Sequencing Core, University of Illinois Urbana-Champaign). The  
 139 full-length 16S rRNA gene amplicons were generated with universal barcoded primers (27F and  
 140 1492R)(40), PCR products were subject to library preparation and sequenced on the PacBio Sequel  
 141 IIe using the circular consensus sequencing (CCS) mode. Raw reads were demultiplexed and CCS  
 142 analyses were performed to obtain consensus reads. Further details on PCR, library preparation,  
 143 and sequencing are provided in the supplementary text. Downstream data processing was  
 144 performed on a system-by-system basis using DADA2 v1.19.2 (41) in R v4.0.4. using the sample  
 145 inference method for full-length 16S rRNA gene with single-nucleotide resolution (42). Table S1a

146 summarizes the reads per samples at different stages of data processing in DADA2. ASVs with  
147  $\geq 100$  reads in any of the three systems ( $\sim > 0.01\%$  relative abundance) were clustered into  
148 operational taxonomy units (OTUs) using the function ‘IdClusters’ in DECIPHER v2.16.1. (43)  
149 at 98.7% sequence identity, a previously recommended threshold for clustering full-length 16S  
150 rRNA gene sequences at the species level (44,45). The most abundant ASV within each OTU was  
151 used as representative sequence and its taxonomic affiliation was used as the consensus taxonomy  
152 for the OTU. Additional details on ASV to OTU data processing are provided in the supplementary  
153 text. Representative ASV sequence from all OTUs were aligned with MUSCLE v3.8.1551 (46),  
154 and the phylogenetic tree was constructed and visualized using IQ-TREE v2.0.3 (47) and iTOL  
155 v2.1.7 (48), respectively. Principal Coordinates Analysis (PCoA) was performed with the weighted  
156 and unweighted UniFrac distance metric (49) using the ‘ordinate’ function of phyloseq v1.32.0  
157 and the ‘plot\_ordination’ function of ggplot2 v3.3.5 (50).

158

159 ASVs classified at the genus level as *Nitrospira* and *Nitrosomonas* (the only two nitrifying genera  
160 detected in this study) were extracted from the rarefied ASV table (rarefied to the sample with  
161 lowest read count) and reference 16S rRNA gene sequences for the two genera were obtained from  
162 the SILVA SSU 138.1 database (Tables S2 and S3). Comammox bacterial 16S rRNA genes were  
163 extracted from references genomes downloaded from NCBI (Table S4). MUSCLE v3.8.1551 was  
164 used to align ASVs sequences from each genus with their respective references, and a maximum  
165 likelihood phylogenetic tree of each species was generated using IQ-TREE v2.0.3 and visualized  
166 in iTOL v2.1.7. To assess the population diversity (i.e., genus level) the Shannon diversity was  
167 calculated per sample from the relative abundance of the OTUs classified as the same genus (e.g.,  
168 *Nitrospira*) as the exponential of the Shannon index. To estimate intra-population diversity, the

169 effective microdiversity of each OTU was calculated on a per sample basis from the relative  
170 abundance of its ASVs as the exponential of the Shannon index, which is analogous to the effective  
171 number of strains (i.e., ASVs) within a population (i.e., OTU) (36).

172

### 173 **2.3 Illumina and Oxford Nanopore sequencing, co-assembly and hybrid assembly**

174 Seven samples from GRE and NEU and six samples from JAMMSM were sent for sequencing on  
175 the Illumina NovaSeq 6000 platform, and two samples from each system on the Oxford Nanopore  
176 Technologies GridIONx5 (Table 1) to the Roy J. Carver Biotechnology Center at the University  
177 of Illinois Urbana-Champaign Sequencing Core. Library preparation and sequencing details are  
178 provided in the supplementary text. These runs resulted in 1.78 billion paired-end short reads  
179 (2×150 nt reads) and 45.3 Gbps of long read data (Table S1b and c). Raw short reads were filtered  
180 using fastp v0.20.0 (51) and the Univec database was used to remove contamination from the  
181 filtered reads as previously described (31). The resulting bam files were sorted using SAMtools  
182 v1.9 (52) and converted into fastq files with bedtools v2.29.0 (53). All reads from the same system  
183 were co-assembled into contigs using metaSpades v3.13.0 (54). Co-assemblies were performed  
184 with kmer sizes of 21, 33, 55 and 77. Contigs smaller than 500 bp were removed from the co-  
185 assemblies using the Anvi'o v6.1 command 'anvi-script-reformat-fasta' (55). The reformatted  
186 assembly fasta files were indexed with bwa index v0.7.17 (56) and the paired end reads from each  
187 metagenome were mapped to the respective co-assembly using bwa mem v0.7.17. The resulting  
188 sam files were converted to bam files using 'samtools view -F 4 -bhS' to retain only mapped reads.  
189 Hybrid metagenomic assemblies were performed using OPERA-MS v0.9.0., which combines the  
190 advantages of short and long-read technologies to improve genome assemblies (37). The OPERA-  
191 MS inputs were short-read metagenomic assemblies to provide a good representation of the



192 sequences in the metagenome, and long and short reads to identify connections between the contigs  
193 and obtain contiguous assemblies with low base-pair error. OPERA-MS was executed with the  
194 flags `--no-ref-clustering` and `--long-read-mapper minimap2`. Both sets of co-assemblies were  
195 evaluated using QUAST v5.0.2 (57) (Table S5).

196

## 197 **2.4 Recovery, annotation, refining and dereplication of metagenome assembled genomes** 198 **(MAGs)**

199 Binning was performed separately, with co-assemblies from Illumina short reads only and hybrid  
200 assemblies from each system using MetaBAT2 v2.12.1 (58), CONCOCT v1.1.0 (59), and  
201 MaxBin2 v2.2.7 (60) using contigs greater than 2000 bp. Pilon v1.23 (61) was used for polishing  
202 to improve the draft bins obtained from the hybrid assemblies. The quality and taxonomy of the  
203 resulting bins were determined with CheckM v1.1.2 (62) and the Genome Taxonomy Database  
204 Toolkit (GTDB-Tk 1.1.1, database release r86 v3) (63), respectively. Bins were subject to gene  
205 calling using Prodigal v2.6.3 (64) and gene annotation against the KEGG database (65) using  
206 kofamscan v1.2.0 (66). Only bins that were taxonomically assigned to known nitrifying genera or  
207 those containing genes associated with nitrification (i.e., *amoA* [KO number K10944], *amoB*  
208 [K10945], *amoC* [K10946], *hao* [K10535], *nxrA* [K00370], *nxrB* [K00371]) were retained for  
209 manual refinement with Anvi'o v6.1. DASTool v1.1.2 (67) was used to combine and curate the  
210 refined bins from the three binning methods and generate a non-redundant set of bins from each  
211 co-assembly (i.e., one set of bins per system for the short-read only and for the hybrid assemblies).  
212 In total, 43 nitrifiers bins from the short-read co-assemblies and 30 from the hybrid assemblies  
213 were obtained. These bins were de-replicated using drep v2.5.4 at 95% ANI with completeness  
214 and contamination thresholds set to 50% and 10%, respectively. This resulted in 44 nitrifier MAGs

215 with completeness and redundancy estimates higher than 50% and lower than 10%, respectively,  
216 of which 21 contained genes associated with nitrification (i.e., *amoA*, *amoB*, *amoC*, *hao*, *nxrA*,  
217 and/or *nxrB*).

218

## 219 **2.5 Taxonomy aware re-assembly using nitrifying bacterial MAGs**

220 We performed re-assemblies from reads mapped to the 44 nitrifying bacterial MAGs to improve  
221 MAG quality. First, contigs associated with *Nitrospira* (including *comammox*) and *Nitrosomonas*  
222 MAGs from each system (i.e., GRE, JAMMSM, NEU) were collated into a single fasta file per  
223 system. The resulting fasta files were indexed using *bwa index v0.7.17* and short reads from the  
224 respective system were mapped using *bwa mem v0.7.17*. The resultant bam file was split into  
225 *Nitrospira* and *Nitrosomonas* specific bam files per system and respective fastq files were  
226 generated per using *samtools fastq*. Long reads were also recruited by mapping the reads from  
227 Nanopore sequencing to the fasta files with ‘*bwa mem -x ont2d*’. Short reads mapping to the  
228 contigs of the MAGs classified as *Nitrospira* and *Nitrosomonas* were re-assembled using  
229 *metaSpades v3.13.0* with *kmers 21, 33, 55 and 77* on a system-by-system basis. The new genus-  
230 specific assemblies and fastq files of mapped long reads were used as input for OPERA-MS to  
231 perform hybrid metagenomic re-assemblies (n=9). Quality assessment, binning, taxonomy  
232 annotation, manual refining and gene calling were performed as described before. This resulted  
233 in 36 nitrifier MAGs with completeness and redundancy estimates higher than 50% and lower than  
234 10%, respectively, of which 24 contained genes associated with nitrification (*amoA*, *amoB*, *amoC*,  
235 *hao*, *nxrA*, and/or *nxrB*). The abundance of each MAG per sample (reads per kilobase million,  
236 RPKM) and proportion of genome covered was calculated with *coverM (68)*. Four low abundant  
237 MAGs (3 AOB and 1 *Nitrospira*-NOB) with less than 50% genome coverage in any of the samples

238 were removed from subsequent analyses. The entire workflow for assembly and re-assembly of  
239 MAGs is outlined in Figure S1. The nitrifier MAGs were subsequently phylogenetically placed in  
240 the context of reference *Nitrospira* (Table S4) and *Nitrosomonas* (Table S6) genomes  
241 (supplementary text).

242

## 243 **2.6 Diversity and micro-diversity analysis of nitrifiers**

244 FastANI v1.3 (69) was used to calculate the pairwise average nucleotide identity (ANI) between  
245 MAGs within each functional group (i.e., AOB, *Nitrospira*-NOB, and *Nitrospira*-comammox).  
246 ANI represents the mean nucleotide identity of the orthologous genes shared between two genomes  
247 offering a robust resolution between similar or identical species (i.e., ~80-100% ANI) (69). Since  
248 ANI values lower than 80% are not reported, we also calculated the average amino acid identity  
249 (AAI) to estimate similarity between two genomes at the amino acid level using compareM (70).  
250 To determine intra-population diversity (i.e., microdiversity), we calculated the average nucleotide  
251 identity from reads (ANIr) (32) with 90% read identity threshold, as recommended for intra-  
252 population comparisons. inStrain v1.3.9 (71) was used to determine the nucleotide diversity for  
253 each MAG in each sample and their population average nucleotide identity (popANI) between the  
254 samples where they were detected. Nucleotide diversity is a measurement of genetic  
255 (micro)diversity at every position along the genome using mapped reads, while popANI is a unique  
256 ANI calculation performed by inStrain that considers both major and minor alleles. This is  
257 different from the traditional ANI (called consensus ANI [conANI] in inStrain), which only  
258 considers major alleles to call (or not) a substitution. Details on parameters and procedures  
259 associated with the implementation of ANIr and inStrain are presented in the supplementary text.  
260

## 261 **2.7 Statistical Analyses**

262 Statistical tests were performed using R v4.0.4. All correlations were performed using a linear  
263 regression model. The Spearman's rank correlation coefficients between MAG abundances were  
264 performed with the 'rcorr' function of the R package Hmisc v4.5.0. Pair-wise significances were  
265 calculated with the Kruskal-Wallis test, a nonparametric approach to the one-ANOVA. Principal  
266 coordinates analysis (PCoA) with the weighted and unweighted UniFrac distance metrics was used  
267 to compare the community composition among systems.

268

## 269 **3.0 Results**

### 270 **3.1 Community composition in nitrogen removal systems**

271 Three nitrification-denitrification systems with high absolute abundance of comammox bacteria  
272 were selected for this time-series study on the basis of our previous findings (31) since high  
273 sequencing coverage is critical for evaluation of sub-population level diversity. Specifically, over  
274 the sampling period (June 2017 to June 2018) comammox bacteria constituted approximately 0.77  
275 ( $\pm 0.32$ ), 4.7 ( $\pm 4.21$ ), and 0.45 ( $\pm 0.24$ ) % of total bacterial 16S rRNA gene abundances at GRE,  
276 JAMMSM, and NEU, respectively.

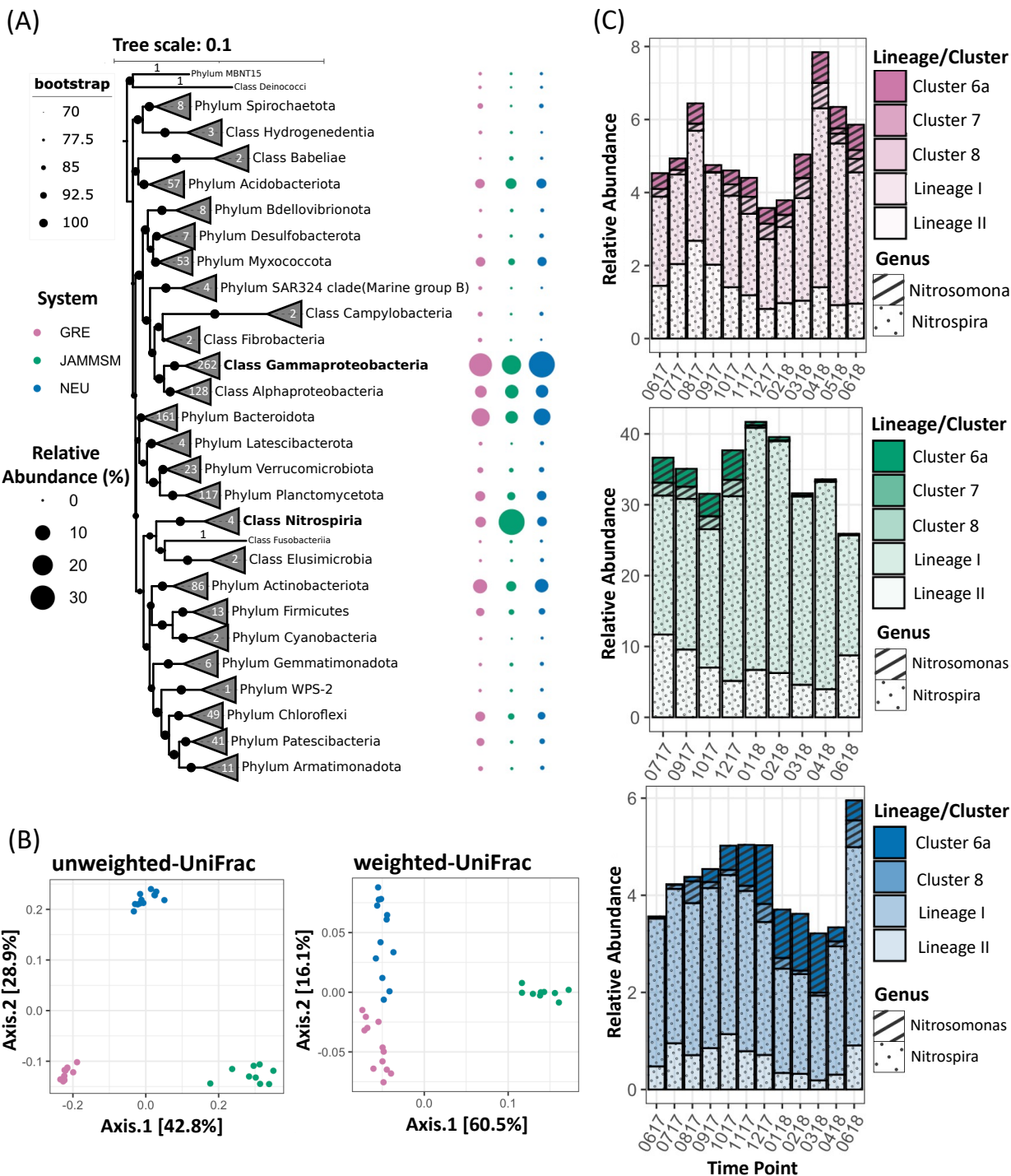
277

278 The PacBio full-length 16S rRNA gene sequencing resulted in 2,317,019 total reads (Table S1a)  
279 with 7501, 5783, and 8260 ASVs at GRE, JAMMSM and NEU, respectively. Each ASV set was  
280 rarefied to the sample with the smallest library size per system resulting in 7040, 4349, and 5749  
281 ASVs from GRE, JAMMSM, and NEU, respectively, and a total of 16651 unique ASVs across all  
282 three systems. ASVs with a total of 100 reads in each system were clustered into OTUs at 97%  
283 identity, resulting in 846 OTUs (Figure 1A; Table S7). The most abundant OTUs at GRE and NEU

284 were from the classes *Gammaproteobacteria* and *Alphaproteobacteria* and the phylum  
285 *Bacteroidota* while the class *Nitrospira* had the highest relative abundance in JAMMSM. PCoA  
286 using weighted (WUF) and unweighted UniFrac (UUF) distance metrics (Figure 1B) demonstrated  
287 that samples clustered by system and the community structure between GRE and NEU (WUF:  
288  $0.18 \pm 0.03$ ; UUF:  $0.46 \pm 0.03$ ) were significantly more similar than GRE and JAMMSM (WUF:  
289  $0.28 \pm 0.02$ ; UUF:  $0.56 \pm 0.05$ ) and NEU and JAMMSM (WUF:  $0.29 \pm 0.03$ ; UUF:  $0.52 \pm 0.04$ )  
290 ( $p < 0.05$ ; Figure S2A and B). The greater similarity between GRE and NEU is likely due to the  
291 fact that these were suspended phase communities as compared to attached phase (i.e., biofilm  
292 samples) communities collected from the IFAS system at JAMMSM.

293

294 *Nitrosomonas*- (31, 25, and 14 ASVs at GRE, JAMMSM, and NEU, respectively) and *Nitrospira*-  
295 like bacteria (37, 36, and 29 ASVs at GRE, JAMMSM, and NEU, respectively) were the only  
296 nitrifiers present in the systems. *Nitrospira* lineage I ASVs were the most abundant group across  
297 all systems with average relative abundances of 2.91, 25.13, and 2.88%, followed by *Nitrospira*  
298 lineage II with average relative abundances 1.41, 7.07, and 0.64% in GRE, JAMMSM, and NEU,  
299 respectively (Figure 1C). *N. oligotropha* (Cluster 6a) and *N. communis* (Cluster 8) -like ASV's  
300 were present in all systems with relative abundances of 0.5 and 0.33, 1.68 and 0.93, and 0.58 and  
301 0.19% in GRE, JAMMSM, and NEU, respectively. *N. europaea/mobilis* (Cluster 7)-like ASVs  
302 were also detected in GRE and JAMMSM at very low abundances compared to the other two  
303 *Nitrosomonas* lineages (average  $\sim 0.032$  and 0.002%, respectively).

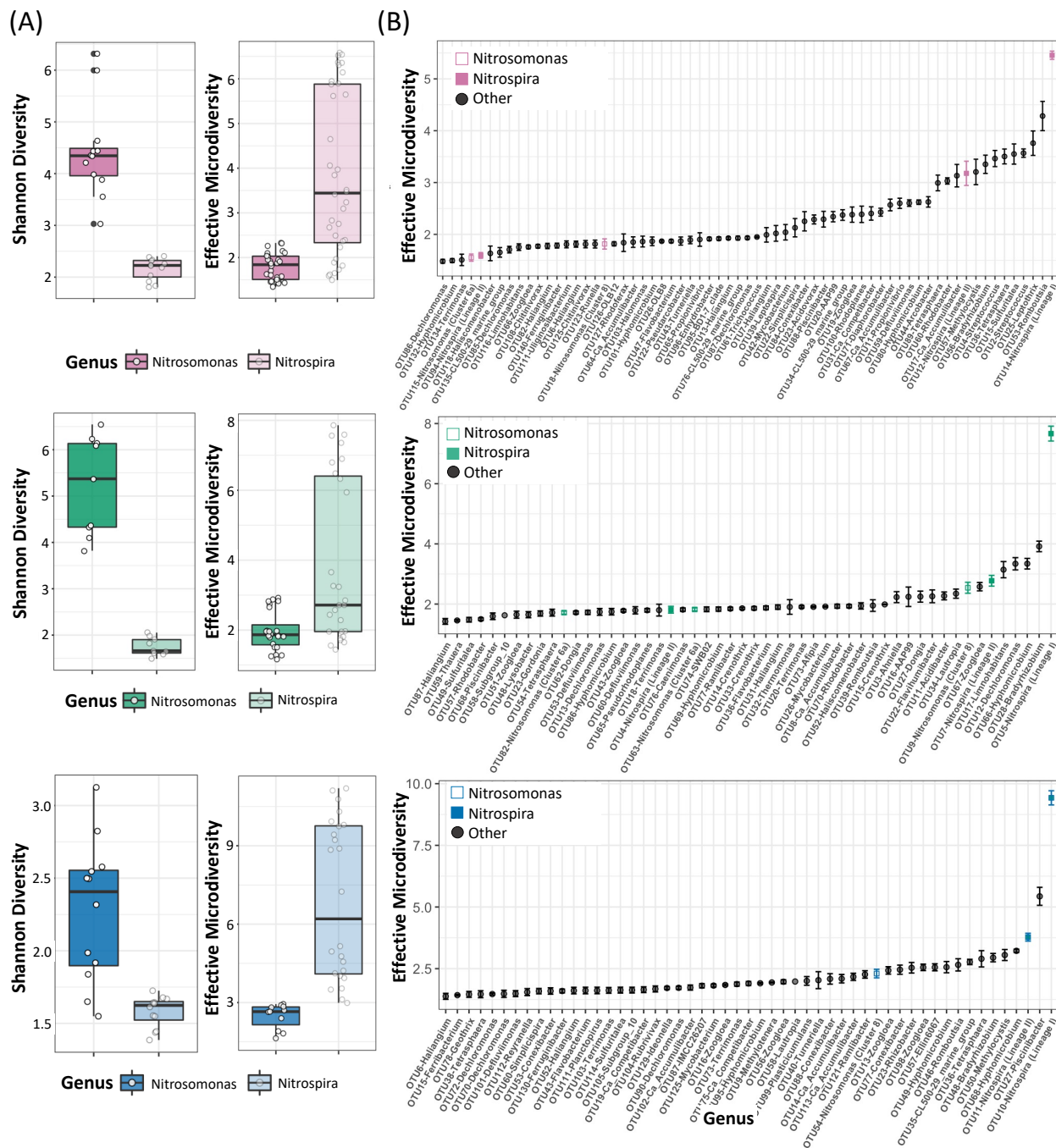


304  
 305 **Figure 1:** Community composition of the nitrogen removal systems. (A) Phylogenetic tree constructed using full-length 16S rRNA  
 306 gene sequences of the representative ASV in each OTU (98.7% sequence similarity cluster), with their corresponding relative  
 307 abundances shown at the right. Branches are collapsed at the phylum/class level and corresponding number of OTUs within each  
 308 phylum/class across all three systems are indicated. Black circles on branches designate bootstrap support. Bubbles represent  
 309 cumulative relative abundances of all OTUs within each phylum/class per system. (B) Principal Coordinate Analysis of community  
 310 composition with unweighted and weighted UniFrac distance metrics using OTU data. Points represent the samples. (C)  
 311 Cumulative relative abundances of OTUs corresponding to *Nitrospiria* Lineages and *Nitrosomonas* clusters per system at each  
 312 time point analyzed. Colors correspond to each nitrogen removal system with (GRE, pink; JAMMSM, green; NEU, blue).



### 313 3.2 16S rRNA gene sequence-based diversity and microdiversity of nitrifiers

314 The *Nitrospira*-like ASVs belonged to lineage I (20, 24, and 22 ASVs at GRE, JAMMSM, and  
315 NEU, respectively) or lineage II (17, 12, and 7 at GRE, JAMMSM, and NEU, respectively) (Figure  
316 S3A), while *Nitrosomonas*-like ASVs were associated with *N. oligotropha* (23, 21, and 10 at GRE,  
317 JAMMSM, and NEU, respectively), *N. europaea/mobilis* (5 and 1 at GRE and JAMMSM,  
318 respectively), and *N. communis* (3, 3, and 4 at GRE, JAMMSM, and NEU, respectively) lineages  
319 (Figure S3B). The *Nitrospira*- and *Nitrosomonas*-like ASVs clustered into 3, 3 and 2, and 21, 14  
320 and 10 OTUs for GRE, JAMMSM and NEU, respectively. The Shannon diversity (i.e., population  
321 diversity) of *Nitrosomonas*-like OTUs was significantly higher than *Nitrospira*-like OTUs, while  
322 the effective microdiversity of *Nitrospira*-like OTUs was significantly higher than *Nitrosomonas*-  
323 like OTUs in all systems ( $p < 0.05$ ; Figure 2A,). *Nitrospira* lineage I OTUs consistently showed the  
324 highest microdiversity (Figure 2B) while two *Nitrospira* lineage II OTUs from GRE and  
325 JAMMSM had a lower effective microdiversity than most nitrifier OTUs. These results suggest  
326 high population level diversity and low microdiversity for *Nitrosomonas*-like bacteria, while the  
327 opposite was observed for *Nitrospira*-like bacteria in Lineage I (more ASVs per OTU). The  
328 effective microdiversity of *Nitrospira*-like OTUs was positively correlated with their relative  
329 abundance (Figure S4A). However, this was neither the case for *Nitrosomonas*-like OTUs (Figure  
330 S4B) and the overall bacterial community (Figure S4C). Although the relative abundances of some  
331 other highly microdiverse OTUs were positively correlated with effective microdiversity (e.g.,  
332 *Hyphomicrobium*), this correlation was not statistically significant for most (e.g., Figure S5). This  
333 suggests that the increase in effective microdiversity of *Nitrospira*-like OTUs with increasing  
334 relative abundance is more likely to be an ecological phenomenon rather than a sampling artifact.



335

336 **Figure 2:** Diversity and microdiversity based on ASV analyses. (A) Boxplot of Shannon diversity and effective microdiversity of  
 337 *Nitrospira* (lighter shade) and *Nitrosomonas* (darker shade) at each system. (B) Average effective microdiversity of major OTUs  
 338 per system. *Nitrospira* and *Nitrosomonas* OTUs are colored by system (GRE, pink; JAMMSM, green; NEU, blue), represented by  
 339 closed and open symbols, respectively.

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### 345 **3.3 Improving the recovery of metagenome assembled genomes of nitrifying bacteria**

346 A total of 43 dereplicated nitrifiers MAGS were recovered from short-read assemblies. However,  
347 several MAGs lacked key genes associated with nitrification (i.e., *amoA*, *amoB*, *amoC*, *hao*, *nxrA*,  
348 *nxB*). For example, only 2 of the 7 *Nitrosomonas* MAGs (from JAMMSM and NEU) contained  
349 the complete *amoCAB* operon and only one had the *hao* gene. To improve MAG quality, we  
350 incorporated long reads from Nanopore sequencing and a hybrid assembly using OPERA-MS (37).  
351 The hybrid approach increased the contiguity of the assembly and resulted in 30 nitrifier MAGs,  
352 with nitrification genes present in 19 MAGs (Table S8). It resulted in significant improvement of  
353 MAGs retrieved from JAMMSM but not for GRE or NEU. Thus, we pursued taxonomy aware re-  
354 assembly (72) of the 44 non-redundant nitrifier MAGs after dereplicating bins from short-read and  
355 hybrid assembly approach. The resultant 36 nitrifier MAGs had completeness greater than 70%  
356 and redundancy lower than 10% (Table S8), with nitrification genes present in 24 MAGs.  
357 Although the number of retrieved nitrifier MAGs was lower after the reassembly process, there  
358 was significant improvement in quality, especially for comammox and AOB genomes. For  
359 example, both analyses (before and after reassembly) resulted in 4 *Nitrospira*-comammox MAGs.  
360 However, 2 of the 4 comammox MAGs recovered after reassembly contained the entire genetic  
361 repertoire required for nitrification, and the other two lack only one gene. In contrast, all pre-  
362 reassembly comammox MAGs lacked at least one nitrification gene. These improvements were  
363 also observed for AOB including significant improvement in MAG quality (i.e., higher  
364 completeness, lower redundancy, and less fragmentation) (Table S8). These improvements  
365 resulted from longer contigs created during the reassembly process compared with previous co-  
366 assemblies. In contrast to comammox and AOB, despite becoming less fragmented post-  
367 reassembly, fewer *Nitrospira*-NOB MAGs contained all genes associated with nitrite oxidation.

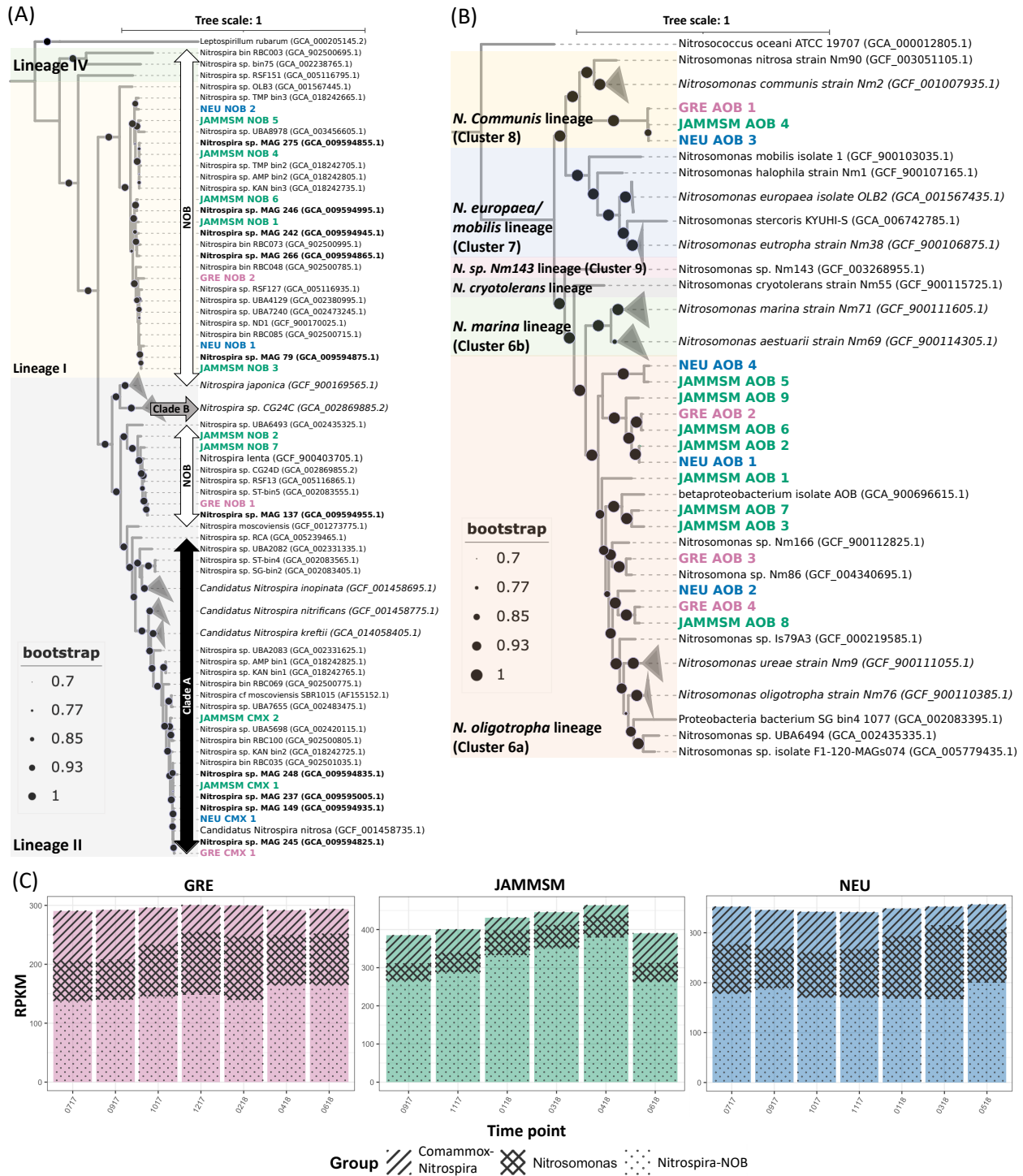
### 368 3.4 Phylogenomic placement and abundance of nitrifier MAGs

369 The 11 canonical NOB and 4 comammox were associated with *Nitrospira* lineage I (8 *Nitrospira*-  
370 NOB) and *Nitrospira* lineage II (3 *Nitrospira*-NOB and 4 *Nitrospira*-comammox), with all  
371 comammox MAGs closely related with *Ca. Nitrospira nitrosa* (Figure 3A) and belonged to  
372 *Nitrospira*-comammox clade A.1 based on *hao* gene phylogenetic inference (Figure S6). This is  
373 in agreement with a previous study that suggests comammox clade A.1 usually co-occur with the  
374 *Nitrospira* lineage I populations in wastewater treatment plants (WWTPs) while clade A.2 and  
375 most clade B are typically found in drinking water treatment plants (DWTPs) (21). While non-  
376 WWTP settings (e.g., soils, sediments, lakes, rivers, DWTPs) harbor diverse comammox bacteria  
377 (9,18,21,73,74), our findings suggest very low diversity of comammox bacteria in wastewater  
378 systems which is consistent with other wastewater studies (12,15,17,22,75–77). All AOB MAGs  
379 (n=20) were associated with the *N. oligotropha* (n=14) and *N. communis* (n=3) lineages (Figure  
380 3B). Although several members of the *N. europaea/mobilis* lineage have been detected in sewage  
381 treatment plants (78), only three AOB MAGs related to this lineage were recovered from  
382 JAMMSM at very low abundances and less than 50% genome coverage. Thus, these MAGs were  
383 excluded from subsequent analyses. The three *N. communis* cluster MAGs (GRE AOB 1,  
384 JAMMSM AOB 4, NEU AOB 3) were nearly 100% identical but shared less than 75% ANI with  
385 all reference genomes. This lineage can be divided into the *N. communis* and the *N. nitrosa* clusters  
386 (79), with *N. communis* species urease negative and detected primarily in agricultural soils, and  
387 isolates of *N. nitrosa* urease positive, preferring aquatic habitats and often found in wastewater  
388 treatment plants (80). Our three MAGs were urease positive but still distant to *N. nitrosa* in the  
389 phylogenomic analysis (Figure 3B), suggesting the presence of a novel phylogenetic lineage.  
390 Lastly, most species of the *N. oligotropha* lineage have been recovered from oligotrophic

391 freshwaters and almost all are urease positive (78,80). Accordingly, 12 of the 14 AOB MAGs  
392 placed within this *Nitrosomonas* lineage were urease positive.

393

394 *Nitrospira*-NOB was the most abundant functional group in all three systems for the duration of  
395 the study with higher abundances in JAMMSM ( $314.5 \pm 70.6$  RPKM) as compared to NEU ( $178.2$   
396  $\pm 12.5$  RPKM) and GRE ( $149.3 \pm 11.7$  RPKM). *Nitrosomonas*-like MAGs were more abundant in  
397 NEU ( $105.2 \pm 23$  RPKM) and GRE ( $86 \pm 15.8$  RPKM) as compared to JAMMSM ( $53.2 \pm 27.4$ ).  
398 This suggests that *Nitrospira*-like bacteria may prefer biofilm-based growth over suspended phase  
399 (81). *Nitrospira*-comammox relative abundances were very similar in all systems, ranging from  
400 55.5 to 64.9 RPKM in JAMMSM and NEU, respectively. The abundances of nitrifying MAGs  
401 were weakly correlated with each other, with only 37 of 195 correlations significant (p-value  
402  $< 0.05$ ). Most of the significant correlations between AOB and *Nitrospira*-NOB were positive  
403 (Figure S7), likely arising as consequence of metabolic interactions of these groups within the  
404 nitrification process. Comammox bacteria had very few correlations with other nitrifiers, with  
405 primarily negative correlations with MAGs from *N. oligotropha*-like bacteria (cluster 6a).



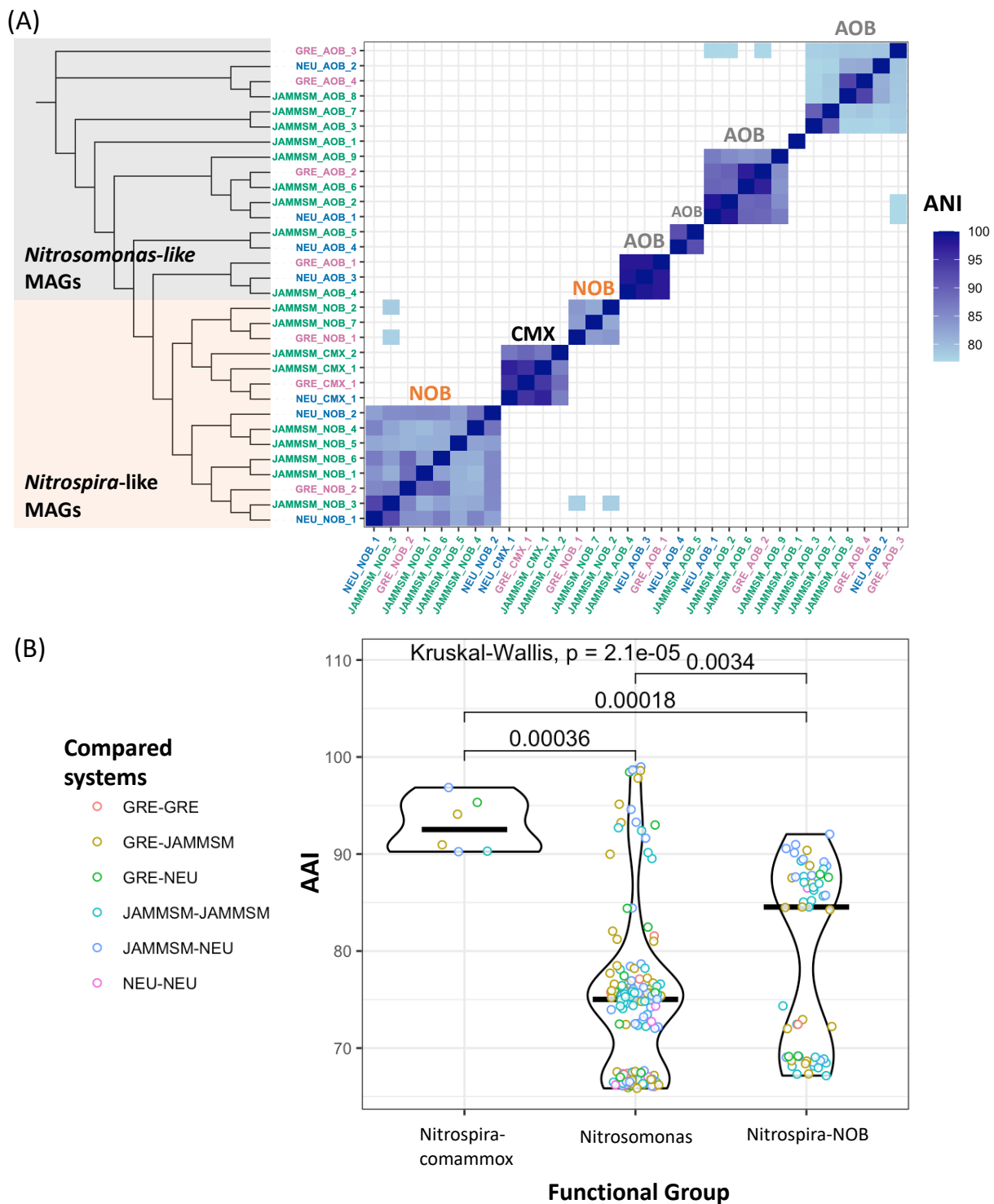
406

407 **Figure 3:** Phylogenetic placement of (A) *Nitrospira* and (B) *Nitrosomonas* MAGs (GRE, pink; JAMMSM, green and NEU, blue)  
 408 with 78 and 48 reference genomes (black), respectively. Comammox clade A (black arrow), clade B (grey arrow) and canonical  
 409 NOB (white arrow) are indicated by arrows and phylogenetic lineages by colored boxes. Branches with reference genomes  
 410 that did not include MAGs from this study were collapsed and labeled with a representative reference genome. All reference genomes  
 411 used for the reconstruction of the trees are listed in Tables S4 and S6. (C) Cumulative abundances (RPKM) of *Nitrospira*-  
 412 comammox, *Nitrospira*-NOB and *Nitrosomonas* MAGs at each time point per system.

413

### 414 3.5 Population-level diversity of nitrifying bacteria within and between systems

415 Post dereplication at 95% ANI (69,82), pairwise ANI and AAI were calculated between all MAGs  
416 within the same functional group (i.e., AOB, *Nitrospira*-NOB, and *Nitrospira*-comammox)  
417 (Figure 4A-B, Tables S9a-b). Of the 136 pairwise comparisons between the 17 AOB MAGs, only  
418 18 (13.2%) had ANI values higher than 80% suggesting high population-level diversity within and  
419 between nitrogen removal systems for the *Nitrosomonas*-like AOB (Figure 4A). The pairwise AAI  
420 for *Nitrospira*-NOB MAGs showed a bimodal distribution (Figure 4B) reflective of the two  
421 lineages detected in each system (i.e., Lineage I and II). Pairwise AAIs for *Nitrospira*-NOB MAGs  
422 of the same lineage within or across systems were typically >85% (Table S9b). These results  
423 suggest that while multiple AOB and *Nitrospira*-NOB coexist in each system, AOB were  
424 substantially more diverse than *Nitrospira*-NOB. In contrast, three of the four comammox MAGs  
425 (one from each system) share ANI values between 95-97% and are closely related with *Ca.*  
426 *Nitrospira nitrosa* (31). Although a comammox MAG from JAMMSM (JAMMSM CMX 2) has  
427 lower than 95% with the other three MAGs, it still exhibited greater than 85% AAI (Figure 4B)  
428 with them and also falls within the *Ca. Nitrospira nitrosa* cluster (Figure 3A)(83). Pairwise AAI  
429 comparisons (which are highly correlated with ANI – Figure S8) indicated that *Nitrospira*-  
430 comammox MAGs have significantly higher AAI compared with AOB and NOB, while AOB  
431 were the most diverse functional group at the population level ( $p < 0.05$ ; Figure 4B). Moreover,  
432 similar to ASV analysis, *Nitrospira*-NOB MAGs were associated with one of two clusters  
433 (*Nitrospira* lineage I or II), with highly similar populations within each cluster (Figure 4B) being  
434 detected across systems. These findings further confirm that in contrast to AOB that demonstrate  
435 high population diversity, comammox bacteria show very low population diversity and, in fact,  
436 may all be associated with the same population across multiple wastewater systems



437

438 **Figure 4:** (A) ANI values calculated between *Nitrosomonas* (AOB), *Nitrospira*-NOB and *Nitrospira*-comammox (CMX) MAGs along  
 439 with the phylogenetic placement of the MAGs. MAG labels are colored by system (GRE, pink; JAMMSM, green and NEU, blue).  
 440 (B) AAI values per functional group. Black lines represent the average, points represent the AAI value between two MAGs and  
 441 colors decode the systems that are compared. Pairwise statistical comparisons were performed with the Kruskal-Wallis test.  
 442  
 443



### 444 3.6 . Microdiversity of nitrifying bacteria

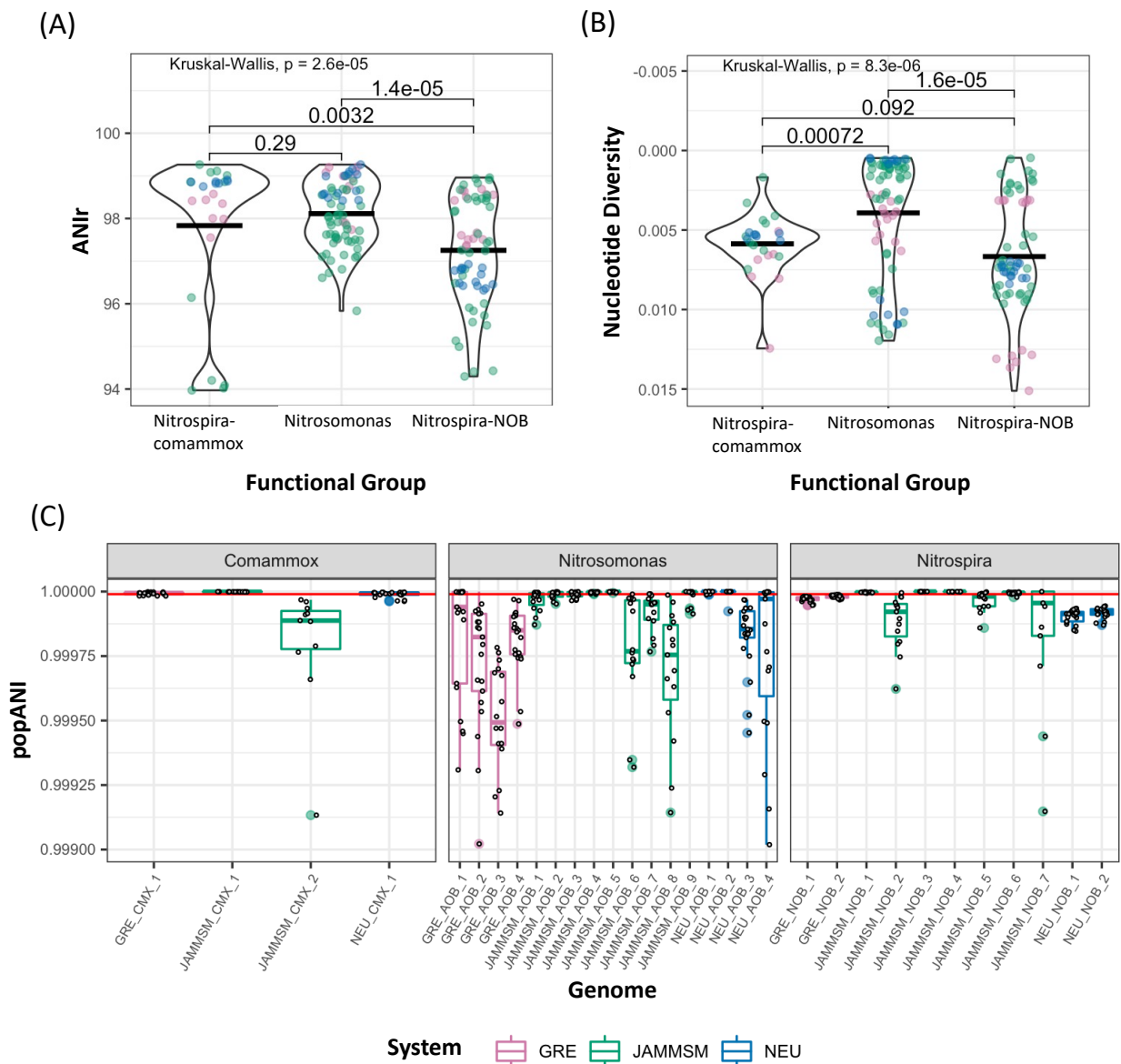
445 The ANIr distributions for comammox and AOB MAGs were not significantly different from each  
446 other, and significantly higher than for *Nitrospira*-NOB MAGs (Figure 5A). While *Nitrospira*-  
447 NOB overall demonstrated lower ANIr indicating higher microdiversity compared to AOB and  
448 comammox bacteria, there was some variability within each system (Figure S9). These observed  
449 trends were similar between ANIr and inStrain estimated nucleotide diversity (Figure 5B), with  
450 AOB showing on average lowest microdiversity and *Nitrospira*-NOB the highest. Specifically, the  
451 average nucleotide diversity of *Nitrosomonas*-like AOB was significantly lower than of both  
452 *Nitrospira* groups, i.e., NOB and comammox ( $p < 0.05$ ), while with ANIr only *Nitrospira*-NOB  
453 were significantly more microdiverse than the two other functional groups. This difference arises  
454 from the fact that the two measures of intra-population diversity (i.e., ANIr and nucleotide  
455 diversity) are calculated differently. For instance, ANIr considers only the major alleles in the  
456 consensus sequence to call a substitution, while nucleotide diversity is calculated using base pair  
457 frequencies at each position. Nonetheless, both approaches indicate that *Nitrospira*-NOB exhibit  
458 significantly higher microdiversity as compared to *Nitrosomonas*-like AOB and *Nitrospira*-  
459 comammox bacteria. These results are consistent with the 16S rRNA gene analysis where  
460 *Nitrospira* lineage I and most lineage II OTUs had significantly higher effective microdiversity  
461 than the *Nitrosomonas*-like OTUs and the remaining *Nitrospira* lineage II OTUs. Based on these  
462 results, we speculate that the *Nitrospira* lineage II OTUs with low effective microdiversity likely  
463 correspond to comammox bacteria (Figure 2B).

464

465 Finally, popANI, a microdiversity-aware ANI calculation, was used to discriminate between  
466 strains across samples using the recommended popANI threshold of 99.999% (Olm et al., 2021).

467 All popANI values from GRE\_CMX\_1, JAMMSM\_CMX\_1, and NEU\_CMX\_1 were above or  
468 very close to the 99.999% threshold, suggesting low strain-level diversity of comammox bacteria  
469 (Figure 5C). Only one comammox genome retrieved from JAMMSM (JAMMSM\_CMX\_2) had  
470 popANI values lower than the recommended threshold. However, the average relative abundance  
471 of this comammox (RPKM =  $7 \pm 2$ ) was particularly low as compared to the more abundant  
472 comammox in JAMMSM (RPKM =  $48 \pm 21$ ) and those retrieved from GRE (RPKM =  $60 \pm 19$ ) and  
473 NEU (RPKM =  $65 \pm 17$ ). Therefore, the principal or sole comammox bacteria in each system is  
474 not only a single population but likely also a specific strain. The ANI<sub>r</sub> for each comammox MAG  
475 estimated by mapping reads from a different system (e.g., mapping GRE or NEU reads to MAGs  
476 assembled from JAMMSM) were lower than when mapping of reads from the system from where  
477 the MAG was obtained (e.g., mapping JAMMSM reads to MAGs assembled from JAMMSM;  
478 Figure S10). This suggests that on the one hand the comammox bacterial population within a single  
479 system is restricted to a specific *Ca. Nitrospira nitrosa*-like strain, but on the other hand that  
480 different systems contain different strains.





481

482 **Figure 5:** (A) ANI<sub>r</sub> and (B) nucleotide diversity values from each functional group. Points represent the average value of each MAG  
 483 and are colored according to the systems (GRE, pink; JAMMSM, green and NEU, blue) from which the MAGs were recovered. (C)  
 484 popANI values from each MAG. MAGs are colored by system and points represent the popANI value of the MAG between two  
 485 samples.

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## 495 4 Discussion

### 496 4.1 Hybrid assembly and re-assembly processes improve the quality of metagenome

#### 497 assembled genomes

498 Reliance on short reads for genome assembly from complex metagenomes can lead to incomplete  
499 and fragmented MAGs especially in systems containing closely related populations (84). Indeed,  
500 short read analysis resulted in highly fragmented nitrifier MAGs lacking several nitrification  
501 genes. To improve the short read assemblies, we used long reads generated from Nanopore  
502 sequencing with a hybrid assembly approach (38). Recent studies have demonstrated the ability of  
503 hybrid assemblies to significantly improve the recovery of complex bacterial genomes (85,86) by  
504 increasing the contiguity and accuracy of metagenome assembly (87). Combining the two  
505 sequencing techniques (i.e., short and long reads) with the hybrid assembler OPERA-MS resulted  
506 in modest assembly and binning improvements regarding genome contiguity and completeness.  
507 However, the bead beating based DNA extraction protocol may have resulted in shearing of  
508 genomic DNA into smaller fragments which is not ideal for long-read sequencing (88) and resulted  
509 in lower read lengths. In an attempt to further improve MAG quality, we performed taxonomy  
510 aware reassemblies (72) as described in the materials and methods section. Since a hybrid approach  
511 resulted in some improvements in MAG quality, we also adopted a hybrid strategy with OPERA-  
512 MS for taxonomy aware reassembly. This significantly improved the completeness and contiguity  
513 of MAGs and number of annotated nitrifier genes (especially for *Nitrosomonas* and comammox  
514 *Nitrospira*) binned within them. While the *Nitrospira*-NOB MAGs also became less fragmented,  
515 they showed fewer improvements with respect to the presence of nitrite oxidation genes. This is  
516 likely caused by the high levels of microdiversity within *Nitrospira*-NOB (Figures 2 and 5), which  
517 likely impacts both the read mapping and de novo assembly process. In the future, this issue might

518 be improved by incorporating more long-read data into the re-assemblies, especially once the  
519 accuracy of long-read sequencing technologies further increases.

520

521 **4.2 *Ca.* Nitrospira nitrosa-like bacteria are the primary comammox bacteria in secondary**  
522 **wastewater treatment for nitrogen removal**

523 Structural diversity and functional redundancy are inherently linked to process stability when  
524 environmental and process conditions vary (28). Both, metagenomics and 16S rRNA gene  
525 sequencing, indicated that *Nitrosomonas* and *Nitrospira*-like bacteria were the only known  
526 nitrifiers present in the systems investigated. This is consistent with numerous studies  
527 (26,35,89,90) suggesting the specific adaptation of these genera to the wastewater environment.  
528 The majority of the *Nitrosomonas*-like AOB MAGs, within and between systems, had AAI values  
529 below 85% (Figure 4) indicating that this functional group exhibited high species level diversity.  
530 Meanwhile, *Nitrospira*-like MAGs within the same lineage reported AAI values above 85% in  
531 most cases. The high population level diversity of the *Nitrosomonas*-like bacteria was also  
532 confirmed with full length 16S rRNA gene sequencing (Figure S3B) and corroborates earlier  
533 studies on their diversity in wastewater (26,90,91). Although *N. europaea*, *N. oligotropha* and the  
534 *Nitrosomonas*-like clusters closely associated with *N. communis/nitrosa* are frequently the most  
535 abundant groups in wastewater, different *Nitrosomonas* populations dominate different processes  
536 and the extent of their diversity is system dependent (26). For example, *N. europaea* are often the  
537 major AOB in high ammonia concentration environments (e.g., anammox reactors) due to low  
538 ammonia affinity and low DO level adaptation (91). The observed dominance of *N. oligotropha*  
539 and *N. communis* in this study are consistent with previous studies of WWTPs with relatively  
540 lower ammonia concentrations (26,91,92).

541

542 Comammox bacteria exhibited significantly lower population level diversity as compared to strict  
543 AOB and NOB (Figure 4). All comammox *Nitrospira* MAGs, independent of the system they were  
544 recovered from, belong to *Nitrospira* lineage II and were closely associated with *Ca. Nitrospira*  
545 *nitrosa*. These findings are consistent with several other studies that have found *Ca. Nitrospira*  
546 *nitrosa*-like bacteria as the unique or principal comammox bacteria present in their laboratory and  
547 full-scale wastewater treatment systems (11,15,17,20,22,75,76), both using metagenomics and  
548 gene targeted analyses. In contrast, other engineered and natural habitats typically harbor multiple  
549 co-existing species across the two primary clades (i.e., comammox clade A and B). For example,  
550 a study of multiple groundwater-fed sand filters recovered 12 comammox MAGs (5 clade A and  
551 7 clade B) and revealed the co-occurrence of at least five comammox species (10 on average) per  
552 system (34). Other studies in soils, sediments, lakes, drinking water, groundwater, and estuaries  
553 have also reported the co-existence of multiple comammox species (9,73,93,94). However, this  
554 co-existence of comammox populations across clades does not seem to be prevalent in wastewater  
555 treatment systems. For instance, a study of 8 WWTPs found 9 of 14 comammox OTUs were  
556 associated with the *Ca. Nitrospira nitrosa* cluster and accounted for 94.34% of all comammox  
557 *amoA* sequences (17) with different OTUs representing the dominant comammox species in  
558 different systems. Moreover, Beach and Noguera (75) reported the dominance of *Ca. Nitrospira*  
559 *nitrosa* in five low DO nitrogen removal bioreactors using qPCR with species-specific primers.  
560 These findings suggest that the diversity of comammox bacteria in WWTPs is lower than in other  
561 habitats where several comammox species co-exist.

562

563 In contrast, Spasov et al. (16) reported the detection of multiple comammox MAGs in samples  
564 collected from rotating biological contactors (RBCs) used for tertiary treatment at a municipal  
565 WWTP in Ontario, Canada. The difference in population diversity between that study and other  
566 WWTP investigations, including this one, may be associated with ammonia availability.  
567 Specifically, tertiary treatment systems are typically employed as a polishing step; they are  
568 designed and operated to remove low residual amounts of nitrogen. The influent ammonium  
569 concentrations reported for the RBC tertiary treatment system range from 0.2 to 16.3  $\mu\text{M}$  (16),  
570 while the three systems in this study have influent ammonia concentrations between 2 and 3 mM.  
571 The comammox *Nitrospira* MAGs with overall higher abundances in the tertiary plant (RBC001  
572 and RBC083) were less abundant at the beginning of the train (higher ammonium concentrations)  
573 than at the end of the train (lower ammonium concentration). In contrast, the abundance of  
574 RBC035, the only MAG phylogenetically associated with *Ca. Nitrospira nitrosa* (ANI >95%),  
575 decreased in abundance with decreasing ammonium concentrations. The high comammox  
576 diversity in the RBCs and other ammonium-limited systems suggest that ammonium concentration  
577 and/or flux through the system may be a key factor driving comammox bacteria diversity in  
578 nitrification systems. In fact, Palomo et al. (34) reported that the influent ammonium concentration  
579 was the key explanatory variable associated with comammox bacterial diversity in 12  
580 groundwater-fed sand filters when higher comammox population diversity was detected in systems  
581 receiving lower ammonium concentrations. *Ca. Nitrospira inopinata* and *Ca. Nitrospira kreftii*,  
582 two comammox species with known kinetic parameters, have high apparent ammonia affinities  
583 (95,96); assuming this is a conserved trait in comammox bacteria, this explains their prevalence  
584 and diversity in ammonium-limited systems. However, the kinetic parameters (i.e., affinity  
585 constant, growth rate) of other comammox species especially of the *Ca. Nitrospira nitrosa* clade

586 have not yet been reported, and our findings suggest that at least this comammox clade may have  
587 substantially different kinetic traits that allow it to thrive in systems with higher ammonium  
588 concentrations. Low DO has also been associated with the prevalence of comammox bacteria over  
589 canonical nitrifiers in wastewater systems (15,75). However, the high DO concentrations (>2  
590 mg/L) in this and other studies (17,20) suggest DO is not a crucial factor for the proliferation of  
591 *Ca. Nitrospira nitrosa* in nitrogen removal systems.

592

#### 593 **4.3 Comammox bacteria not only exhibit low diversity at the species/population level but** 594 **also at the strain/intra-population level across wastewater systems**

595 The high abundance and microdiversity of *Nitrospira*-NOB have been previously reported in  
596 wastewater and other engineered and natural environments (4,33,35). However, it is unclear  
597 whether the high microdiversity of *Nitrospira* is associated with high functional diversity or is  
598 indicative of the coexistence of functionally identical *Nitrospira* with allelic diversity. Gruber-  
599 Dorninger et al. (35) reported variable responses of closely related bacteria within *Nitrospira*  
600 lineage I (sequence identities ranging from 95.8 to 99.6%) to different nitrite availabilities.  
601 Similarly, ecological niche partitioning was also identified as potential mechanisms for co-  
602 existence of three *Nitrospira* lineage I strains that used formate under different conditions: two  
603 used formate when incubated with nitrite and ammonia, respectively, while the third used formate  
604 efficiently as the sole substrate (35). Although these results shed light on the possible niche  
605 partitioning of *Nitrospira* sub-species, this does not shed light on whether the microdiversity in  
606 our study represents coexisting *Nitrospira* strains with slight functional differences. Although  
607 functional implications of such high microdiversity are not clear, the persistence of high intra-  
608 population diversity may suggest it plays an important role in the distribution and success of

609 *Nitrospira* populations in wastewater systems. A few studies in other environments have shown  
610 that persistent populations exhibit increased intra-population sequence diversity (7,32). A study in  
611 a saltern pond suggested that the ecologically important genes of the major archaeal sequence-  
612 discrete population were carried by distinct sub-populations (strains), indicating that the adaptation  
613 to different salinity concentrations had led to sub-population differentiation and speciation (niche  
614 partitioning) (82). Moreover, a study in eight different temperate bog lakes concluded that high  
615 microdiversity is associated with the maintenance of functional microbial communities during  
616 changes in environmental conditions (36).

617

618 Contrastingly, comammox bacteria not only exhibited an unexpected low population-level  
619 diversity within and between systems but also showed significantly lower microdiversity as  
620 compared to strict NOB within the genus *Nitrospira*. These results may indicate specific  
621 adaptations of comammox bacteria within the *Ca. Nitrospira nitrosa* cluster to the wastewater  
622 environment and particularly secondary treatment systems. Palomo et al. (34) observed a negative  
623 correlation between the species-level diversity of comammox *Nitrospira* and ammonium  
624 concentrations. However, that was not the case for microdiversity, suggesting that different  
625 mechanisms may shape inter- versus intra-population diversities, or that the range of ammonium  
626 concentrations in the investigated drinking water systems was too narrow to capture any  
627 underlying associations. For instance, the comammox nucleotide diversity in our study ranged  
628 from 0.002 to 0.008, while this was from ~0.005 to ~0.013 for the MAGs obtained from these  
629 drinking water systems, with only three of the twelve comammox MAGs presenting nucleotide  
630 diversities lower than 0.008 (34). This may suggest that comammox microdiversity is also  
631 associated with ammonium concentrations and/or flux through the system, with the observed low

632 microdiversity in wastewater systems associated with the higher prevailing ammonium  
633 concentrations.

634 Despite the broad detection of comammox bacteria in WWTPs, their role and process relevance  
635 are as yet unclear. In contrast to other environments, comammox *Nitrospira* present very low  
636 population and intra-population diversity levels in wastewater treatment systems for nitrogen  
637 removal. This observed lack of diversity and, consequently, lack of functional redundancy may  
638 influence the feasibility of potential design and operational strategies relying primarily on  
639 comammox *Nitrospira* for nitrogen removal. Although ammonium concentration and/or  
640 availability apparently influences the diversity of comammox bacteria, further studies are  
641 necessary to determine other factors driving the success of their clonal community in wastewater.  
642 Moreover, studies are required to estimate the activity of comammox bacteria, specifically *Ca.*  
643 *Nitrospira nitrosa*, and to assess their contribution to nitrification in full-scale nitrogen removal  
644 systems. It is also important to note that the intra-population diversity cannot be solely studied  
645 with shotgun metagenomic methods since high microdiversity can prevent robust assembly of  
646 individual genomes (82). Therefore, more accurate techniques are needed to obtain high-quality  
647 strain-level genome assemblies for *Nitrospira* and other highly microdiverse bacteria.

648

#### 649 **4 Acknowledgements**

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652 acknowledge the operational personnel for assistance with sampling and process data sharing. The  
653 authors also thanks A.M. Eren for assistance with metagenomic co-assemblies.

654



655 **5 Data availability**

656 All raw sequencing data from Illumina and Nanopore platforms and nitrifiers MAGs are

657 available on NCBI under bioproject number PRJNA846349.

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