**TITLE:**

Multifaceted effects of bycatch mitigation measures on target/non-target species for pelagic longline fisheries and consideration for bycatch management

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## SUMMARY

The pelagic longline fishery, in an effort to reduce bycatch of sea turtles, have developed and deployed fisheries bycatch mitigation techniques such as replacing J/tuna hooks and squid bait with circle hooks and whole fish bait. However, little emphasis has been placed on the side effects of bycatch mitigation measures on endangered species other than target bycatch species. Several previous studies of the side effects have been marred by lack of control for the covariates. Here, based on long-term data obtained from research cruises by a pelagic longline vessel, we examined the effects of using circle hooks and whole fish bait to replace squid bait on the fishing mortality of target and non-target fishes, and also bycatch species. A quantitative evaluation analysis of our results, based on a Bayesian approach, showed the use of circle hooks to increase mouth hooking in target and bycatch species, and their size to be proportional to the magnitude of the effect. Although deploying circle hooks increased catch per unit effort (CPUE) and fishing mortality per unit effort (MPUE), changing the bait species from squid to fish clearly had a far greater impact on shark mortality than the use of circle hooks. Because the impact of the introduction of bycatch mitigation measures on species other than the focused bycatch species is non-negligible, a quantitative assessment of bycatch mitigation-related fishing mortality is critical before introducing such measures.

Keywords: billfishes, bycatch, bycatch mitigation measure, fisheries management, sea turtle, sharks, tuna, longline fishery

## INTRODUCTION

Unintentional catch by fisheries is referred to as bycatch, and calls for its reduction are global, especially because the bycatch of endangered species can have devastating effects on such populations. For example, in the tuna longline fishery, concerns about the increased conservation risk for seabirds and sea turtles by unintentional and fatal catch have been a major issue among many regional fisheries management organizations (RFMOs) since the 1990s (Wallace et al. 2013; Dias et al. 2019), and in recent years, some elasmobranchs, whose populations are declining further, are also treated as bycatch species. The decline of these species has focused attention at both the national and the RFMO levels. Even for target fish that are not bycatch species, addressing the deterioration of stock status caused by overfishing requires reductions in unintentional fishing mortality (for instance, billfishes in the North Atlantic; Kerstetter & Graves 2006; Diaz 2008).

Several studies have weighed the sustainability of tuna longline fisheries against the conservation of species vulnerable to bycatch (Hall et al. 2000; Melvin et al. 2014; Clarke et al. 2015)—particularly for seabirds and marine turtles—with the development of several effective bycatch mitigation measures (Melvin et al. 2014; Swimmer et al. 2017). Bycatch mitigation measures in longline fisheries target specific animal groups and are evaluated based on their success in reducing mortality due to bycatch of specific vulnerable species. While the impact on the catch of the target fish is the primary consideration when evaluating bycatch mitigation techniques, few studies have examined the impacts and tradeoffs on species not targeted by mitigation techniques (Pacheco et al. 2011; Gilman et al. 2016). However, an ecosystem-based fisheries management approach to the introduction of bycatch reduction measures demands the assessment of the impact not only on the target bycatch species but also on the ecosystem itself in which the target species lives (Reinhardt et al. 2018).

The circle hook and finfish bait are typical sea turtle bycatch mitigation measures for pelagic longline fisheries (Watson et al. 2005; Gilman et al. 2006; Yokota et al. 2009; Stokes et al. 2011). The tip of the circle hook bends inward, and when a fish or sea turtle swallows the hooked bait, the circle hook cannot hook inside the digestive tract; instead, as the hook exits the mouth, a torque force causes it to hook through the edge of the mouth. This property allows for easy hook removal and has been reported to reduce the mortality rate of bycatch sea turtles on board and after release (Kiyota et al. 2004; Cooke & Suski 2004; Kerstetter & Graves 2006). Reports of positive effects of circle hooks include those for other

species—such as reduced haulback and post-release mortality in sharks, reduced post-release mortality in swordfish, and increased catch rates in tuna, the target fish. The use of fish bait instead of squid bait potentially reduces bycatch (Watson et al. 2005; Yokota et al. 2009, 2011) and mortality rates (Stokes et al. 2011; Parga et al. 2015) of sea turtles.

These sea turtle mitigation measures, however, exact other costs. Several meta-analytic studies have reported the circle hook-related reduction of sea turtle mortality but increase in billfish and shark catch rates (Reinhardt et al. 2018). The use of whole fish bait also reportedly increases the catch of sharks and other species (Foster et al. 2012). However, few studies have allowed for quantitative evaluation of the effects of these mitigation measures experimentally on species beyond sea turtles. This is due in part to the reliance on observer data from commercial vessels, and small sample sizes, small comparison groups, and lack of experimental rigor from research vessels. In addition, the impact assessment for sharks underestimates catch rates and mortality associated with missed catches due to “bite-off” branchline (Reinhardt et al. 2018) and several other issues. In addition, many experimental studies and meta-analyses only evaluate gear impacts at significance levels without evaluating the magnitude of the effect. Small significant differences may be judged not to matter much when assessing overall risk in bycatch species. Studies that controlled for these conditions would allow for an evaluation of adverse effects without the confounding problems described above. Also, many studies use catch/bycatch rates and mortality rates (at haulback or after release) as important impact indicators, without considering irreversible impacts, such as the number of organisms killed at the time of catch. Additionally, hooking position itself is believed to have a strong influence on mortality rate, which demands the estimation of risk under specific hooking conditions, taking causal relationships into account.

Here, we used data from controlled experiments to analyze these confounding factors and developed a Bayesian statistical model to evaluate the effects of changing hook and bait species on fish species other than turtles—particularly, tuna, swordfish, and sharks—and then verified the contribution of circle hooks and fish bait to mortality rate, catch rate, and fatal catch rate, respectively. We also discussed the appropriate assessment of bycatch mitigation measures in fisheries management.

## METHOD

### *Experimental Operations*

We analyzed data from the R/V Taikei No. 2 longline research operation conducted in the Northwest Pacific Ocean between 2002 and 2010—a typical Japanese shallow-setting operation targeting mainly swordfish and shark, using four hooks per basket, a wire leader, and a night soaking (Yokota et al. 2006a). We regarded the 4.0-sun tuna hooks as a control group against which we assessed 10 different sized circle hooks (Table 1). In Appendix 1 we describe hook shapes and other details of these hooks following the measurement method of Yokota et al. (2006b). Since the circle hooks varied in size and sample numbers were too small to analyze each hook separately, we divided them into two groups according to size. Because degree of hook offset has been reported to affect catch and hooking position (Cooke & Suski 2004), most hooks were circle hooks <10°but some were nearly 15°. The bait comprised chub mackerel *Scomber japonicus* and Japanese common squid *Todarodes pacificus* in the range of 20–30 cm in fork length or dorsal mantle length. These were frozen and stored, then completely thawed before being hooked. The sequence of line setting was divided into several experimental segments, and a different combination of hook and bait type was applied for each segment, with an alternate order of segments at each operation.

The researcher recorded catch for each operation for all catch/bycatch, and the species caught, status of haulback (alive or dead), hooking location (mouth, swallow, and body hooking), time of catch, and basket number. Hook location was recorded only for squid bait. At the start of longline operations, the researcher also collected sea surface temperature, which we subsequently used in the analysis.

Since the experiment was conducted using multiple sizes (degrees) of circle hooks in combination with whole fish bait and squid bait, we categorized hook shapes into three main types:

1. control (tuna hook 4.0-sun; tuna)
2. smaller circle hook (smaller than 5.0-sun; small-C)
3. larger circle hook (18/0 and 5.0-sun or more; large-C)

In Table 1 we show annual longline effort separated by bait and hook type, tabulated according to the above categorization. We selected the following species for analysis: blue shark *Prionace glauca*, shortfin mako shark *Isurus oxyrinchus*, bigeye tuna *Thunnus obesus*, swordfish *Xiphias gladius*, striped

marlin *Kajikia audax*, common dolphinfish *Coryphaena hippurus*, escolar *Lepidocybium flavobrunneum*and, longnosed lancetfish *Alepisaurus ferox*, and loggerhead turtles *Caretta caretta* (to confirm the effect on sea turtle bycatch).

## ***Statistical Analysis***

We conducted all analyses using a Bayesian approach to estimate parameters. We adopted haulback mortality rate, catch per unit effort (CPUE), and mortality per unit effort (MPUE) (Afonso et al. 2011) as indices to evaluate the impact of hook and bait type on fishing mortality within the analyses. We used the following data as inputs to the model: number of (units) caught, longline effort (number of hooks), fate at hauling, hooking location, year and location of operation, water temperature at operation, and soaking time.

Because of the heterogeneity of the data set due to some missing data, we split the analysis into two models. MODEL 1 relates only to relationships between hooking location and haulback mortality rate, while MODEL 2 comprises two subsets for estimating haulback mortality rate, CPUE and MPUE.

We based MODEL 1 on a logit regression using a Bernoulli distribution. The model assumes that hooking position is most directly related to mortality rate and attempts to express the indirect effect of hook type by estimating probability of hooking location for each hook type. We express the observed hooking location *H* and haulback mortality rate *M* in MODEL 1 by the following equations:

$$M \sim Bernoulli(p_{dead})\#(1)$$

$$logit(p_{dead}) = \beta_{1,hook} + \beta_{1,sst} + \beta_{1,soaktime}\#(2)$$

$$H \sim Categorical\big(softmax(\theta_{hook})\big)\#(3)$$

where $\beta_1$ is the parameter in each explanatory variable, $p_{dead}$ is the expected haulback mortality rate, and $\theta_{hook}$ is the expected probability of hooking location.

We structured MODEL 2 to calculate the expected number of mortalities per effort (MPUE) based on the parameters estimated in both the mortality and CPUE estimation subsets. In MODEL 2, due to lack of hooking location data, we calculate $p_{dead}$ from a modified equation (2) as follows;

$$logit(p_{dead}) = \beta_{2,hook} + \beta_{2,bait} + \beta_{2,sst} + \beta_{2,soaktime}\#(4)$$

where $\beta_2$ is the parameter in each explanatory variable.

The CPUE subset is almost same as a GLMM analysis in assuming the Poisson distribution as the error structure. We express the number of catches $C$ per operation using the expected CPUE $\lambda$ as follows:

$$C \sim Poisson(\lambda + \log(E))\#(5)$$

$$\lambda = \log\left(\gamma_{hook} + \gamma_{bait} + \gamma_{latitude} + \gamma_{sst} + r_{year}\right)\#(6)$$

$$r_{year} \sim N(0,\ \sigma^2)\#(7)$$

where $E$ is the longline effort, $\gamma$ is a parameter in each explanatory variable, $r_{year}$ is a random effect of annual fluctuation on CPUE, and $\sigma$ is a standard deviation.

We obtained the expected values of hook-bait-specific MPUE $\zeta$ by multiplying CPUE by at-haulback mortality rate as follows:

$$\zeta = \widehat{p_{dead}} \times \hat{\lambda}\#(9)$$

where $\widehat{p_{dead}}$ denotes the expected mortality rate standardized for hook and bait type, and $\hat{\lambda}$ denotes the estimated CPUE standardized for hook and bait type. We used the standardization method for abundance indices used in fisheries stock assessment (Maunder & Punt 2004). In this method, explanatory variables other than the factor subject to standardization are averaged to predict the objective variable, which in stock assessment is a time scale such as years or months, but in our case we modified the standardization scale to reference hook type and bait type.

We calculated each parameter based on a Bayesian approach with Markov chain Monte Carlo (MCMC) sampling. For the MCMC sampling, we used cmdstan 2.28.2 (Stan Development Team 2021). As a prior distribution for each parameter, we used a half student-t distribution with 2 degrees of freedom, mean 0, and variance 2.5 for σ, and we used a uniform distribution for the other parameters. We computed the posterior distribution using Stan with 15000 sampling iterations including 10000 warmup iterations, number of chains as 4, and no sinning. We calculated Bayesian credible intervals based on the highest density interval (HDI) for the estimates. Although the Bayesian approach for the estimates precluded significance testing, we determined the difference between the estimates of the experimental

group and those of the control group ("swallowing" for hooking location, "tuna" hook for hook type, and "squid" for bait type), and if the lower and upper limits of HDI for the difference did not exceed 0, we considered the difference as a difference for convenience (assuming region of practical equivalence [ROPE] as 0; Kruschke 2015). In Appendix 2 we show the Stan code used to estimate each parameter in MODEL 1 and 2. For other data handling, statistical analysis, and plotting, we used R4.1.1 (R Core Team 2021) and packages "cmdstanr 0.4.0," "ggalluvial 0.12.3," "ggthemes 4.2.4," "mapdata 2.3.0," "maps 3.4.0," "sf 1.0–2," "tidybayes 3.0.1," and "tidyverse 1.3.1" (Becker & Brownrigg. 2018; Pebesma 2018; Wickham et al. 2019; Brunson 2020; Arnold 2021; Richard et al. 2018, 2021; Gabry & Češnovar 2021; Kay 2021).

## RESULTS

A total of 286,363 hooks from 612 operations were deployed in the experiment (Table 2). The area of operation ranged from around the Izu Islands in Japan to off the east coast of northeastern Honshu—typical fishing ground for Japanese shallow-setting longliners (Fig. 1; Hiraoka et al. 2016).

### *Summary Statistics*

The most common species caught was blue shark, followed by longnosed lancetfish, common dolphinfish, and shortfin mako shark (Table 3). Major miscellaneous fish species included, in descending order of abundance, salmon shark *Lamna ditropis* (N = 229), pelagic stingray *Pteroplatytrygon violacea* (N = 199), pomflets *Brama spp.* (N = 135), bigeye thresher shark *Alopias superciliosus* (N = 89), and albacore *Thunnus alalunga* (N = 69), but for these species the sample sizes were too skewed among experimental groups to converge the MCMC sampling.

In Table 4 we show the number of fish caught by hook type and bait type. Bigeye tuna had extremely low catches on large-C hooks, and loggerheads had low catches on fish bait. The most common hooking position at the time of catch was mouth hooking for all nine species, with extremely few hook locations other than mouth hooking, especially for bigeye tuna and esclolar (Table 4). The proportion of fatalities of captured species at haulback varied greatly by species (Table 5). The haulback mortality

rate was low for blue sharks, common dolphinfish, escolar, and shortfin mako sharks and, although higher for bigeye tuna, longnose lancetfish, striped marlin, and swordfish, was generally extremely low, especially in loggerhead turtles.

The size composition of the nine species used in the analysis was not statistically compared in this study and was not included in the model because it did not contribute to haulback mortality rate, but in Appendix 3 we include a histogram.

Almost all parameters in the models for the nine species were successfully converged (Rhat < 1.1) but some of the parameters for loggerhead turtle were not converged, as we describe below.

***Output of MODEL 1***

In Table 6 we show percentage estimates of hooking locations by hook type throughout the model. We observed differences in hooking location by hook type for blue shark, common dolfinfish, loggerhead turtle, and swordfish, with an increase in mouth hooking and a decrease in swallowing for large-C for loggerhead turtle and a clear increase in mouth hooking for small-C for shortfin mako shark and swordfish (Fig. 2). In blue sharks, the frequency of swallowing decreased in both small-C and large-C.

In Table 7 we show haulback mortality rates by hooking location. We observed clear differences in haulback mortality rate by hooking location for blue shark, shortfin mako shark, striped marlin, and swordfish (Fig. 3). Haulback mortality rates after hook swallowing for blue sharks were lower than those for body hooking, and higher than those for mouth hooking. We observed lower haulback mortality rates for shortfin mako shark, striped marine, and swordfish from hook swallowing than from mouth hooking.

***Output of MODEL 2***

Haulback mortality was higher for large-C than for tuna hook types in bigeye tuna, but did not differ among hook types in other species (Table 8, Fig. 4). On the other hand, the haulback mortality for two bait types differed in bigeye tuna and blue shark, with a rate reduction and increase with the use of fish bait in bigeye tuna and shortfin mako shark, respectively.

In Fig. 5 and Fig. 6 we show the effects of two environmental factors—sea surface temperature (SST) and soak time—on haulback mortality rate. The response to SST differed by species, with haulback mortality rate increasing with higher SST for shortfin mako sharks, swordfish, common dolphinfish, and

escolar, and conversely increasing with lower SST for bigeye and longnose lancetfish. We observed little fluctuation in haulback mortality rate with SST in blue shark and striped marlin. In general, haulback mortality rate increased with increasing soak time. However, for bigeye tuna, haulback mortality rate decreased with soak time, but the trend was not clear, and for blue sharks, haulback mortality rate increased only slightly with increased soak time.

In Table 9 we show standardized CPUE by hook and bait type. Differences in CPUE among hook type were found for all species except striped marlin (Fig. 7). We observed higher CPUE for small-C in bigeye tuna, escolar, longnose lancetfish, and swordfish, and higher CPUE for both large and small-C in blue shark, shortfin mako shark, and common dolphinfish. Especially in blue sharks, the difference in CPUE ranged from 10 to 20 depending on the hook type.

We observed differences in standardized CPUE by bait type in bigeye tuna, blue shark, shortfin mako shark, common dolphinfish, and escolar. CPUE decreased with whole fish bait in bigeye tuna and blue sharks, but increased in shortfin mako shark, common dolphinfish, and escolar. In the case of blue shark, escolar, shortfin mako shark, and swordfish, the bait effect could be varied with circle hooks, suppressing the CPUE-increasing effect by fish bait in bigeye tuna, blue shark, and swordfish with the use of circle hooks, while conversely this combination boosted increase of CPUE in common dolphinfish, escolar, and shortfin mako shark.

In Table 10 we show calculated MPUE values by hook and bait type. Compared to differences in MPUE among hook and bait type, those in CPUE were relatively small. We observed differences in MPUE by all hook types except for those in common dolphinfish and striped marlin, with higher MPUE for small-C in bigeye tuna, blue shark, escolar, longnosed lancetfish, and swordfish (Fig. 8). We confirmed decreases in MPUE by fish bait in bigeye tuna and blue shark, and blue sharks, and conversely, increases in MPUE in common dolphinfish, shortfin mako shark, and swordfish. In escolar, however, MPUE increased only when small-C was combined with whole fish bait. The effect of the combination of whole fish bait and circle hook varied in bigeye tuna, blue shark, common dolphinfish, shortfin mako shark, and swordfish. In bigeye tuna, blue shark, and swordfish, the effect of the circle hook on MPUE was suppressed by fish bait, while in common dolphinfish and shortfin mako shark, MPUE increased significantly when whole fish bait and circle hook were used together.

***Hook and bait effect for sea turtle bycatch***

We failed to complete our analysis for loggerhead turtles throughout the models because the bias in frequency of capture events among the experimental groups was too large and did not converge except for only a part of MODEL 1. Instead, in Table 11 we show the nominal CPUE, haulback mortality rate, and MPUE for each experimental group, and in Fig. 9, hooking location and haulback mortality rate by each hook type with squid bait. Most individuals survived regardless of hooking location. When whole fish bait was used, haulback mortality rate and associated MPUE were zero. For squid bait, large-C had the smallest CPUE, mortality, and MPUE.

## DISCUSSION

Our experimental comparisons showed that the hook shape and bait type—both considered as effective bycatch mitigation measures for sea turtles—have extremely multifaceted effects for teleost fishes and sharks and, in some species, the direction of the effects was conflicted. The results provide significant insight into two aspects of the management of vulnerable bycatch species in tuna fisheries: how the confrontational effect of bycatch mitigation measures should be managed, and in which processes of fishing mortality intervention in the management of vulnerable species should occur. As a specific concern regarding the former, when considering shortfin mako shark, which are experiencing significant stock depletion in the North Atlantic (Sims et al. 2018; ICCAT 2019), the implementation of bycatch mitigation measures for sea turtles, which are also required to reduce fishery bycatch, could conversely increase fishing mortality and become a conservation risk. Previously, most of bycatch mitigation measure assessments have focused on whether they reduce the impact of vulnerable bycatch species of concern being bycaught, with the secondary impact being from an economic perspective—in other words, whether the catch rate of commercial species is reduced or not. Here we show for the first time that widely adopted bycatch mitigation measures may have a negative impact on some endangered species. Where both species with opposing effects are at low abundance, management measures should be based on a thorough discussion identifying the optimal combination of mitigation measures, accompanied by scientific evidence. With regard to the latter issue, the mortality reduction expected from circle hooks is not very promising, especially for species with high catch mortality, since the main effect of circle hooks is to minimize internal organ damage, which is of little use for species that have died from other causes, such as heat stress or suffocation. For such species, consideration of methods to reduce the catch itself rather than the mortality rate will be of greater benefit.

Because the sea turtle bycatch events in this study were highly skewed toward some terminal gear combinations, we were unable to make many statistical comparisons regarding haulback mortality rate, CPUE, or MPUE, while the size of circle hooks was positively related to increased mouth hooking. This result is consistent with existing studies, which show that the circle hook prevents internal organ damage and improves the probability of live release (Cooke & Suski 2004). Also, although we were unable to make statistical comparisons, CPUE was low, and some combinations did not cause bycatch at all. This result is also consistent with existing studies, and is related to the lower attractant effect of whole

fish bait on marine turtles and the increased probability of swallowing caused by the difficulty to bite off the bait (Stokes et al. 2011; Parga et al. 2015).

The results of MODEL 1 indicate that circle hook use, rather than tuna hook use; clearly reduced hook swallowing for blue sharks, shortfin mako sharks, and swordfish, and in these three species, haulback mortality rate after mouth hooking was also lower than that after hook swallowing. Hook swallowing has been reported to increase the likelihood of fatal damage to internal organs. Although previous studies have reported that studies to attach satellite tags to white marlin *Tetrapturus albidus* caught by recreational fishing using circle hooks and subsequently released have reduced the post-release mortality rate (Horodysky & Graves 2005), unfortunately, the present study did not corroborate this information. Since no data regarding post-release mortality were collected in our research, evaluating the overall impact of hook shape and bait type on bycatch mortality was not possible. Our results showed different effects of hooking location by hook shapes, with more frequent mouth hooking by small circle hooks in many species. Few studies have examined the relationship between hook size and hook hooking position, on the one side, and mortality of non-turtle species by circle hook, on the other, the possibility that relative differences in mouth and hook size, and differences in feeding behavior toward prey (swallowing the prey whole or biting it off) may affect the hooking position. Blue shark studies have reported the negative and long-lasting effects of gut hooking (Epperly et al. 2012; Gilman et al. 2020).

The results of MODEL 2, which examined the effects of hooks and bait, showed increases in MPUE due to the use of small-C hooks of all species except striped marlin. This result suggests that the increased CPUE by small-C hook contributed more to the increase in the MPUE than haulback mortality rate. Interestingly, we did not observe an increased MPUE with large-C hooks. There have been many previous findings on the effects of circle hook use on CPUE, with elevated CPUE for tunas and no consistent trend for sharks and other teleosts. However, although few previous studies have focused on hook size and made comparisons, catch rates for skipjack, shortbill spearfish, escolar, and lancetfish are reported to have decreased when larger hooks were used (Curran & Beverly 2012). Considering the effect of hook size in terms of the catch process, it is unlikely that catch rates increased due to swallowing, as the results of MODEL 1 indicate an increase in mouth hooking for many species. The results of MODEL 2 showed that the effects of the bait type varied by species, unlike the pattern for hook types, with

the use of whole fish bait decreasing CPUE and MPUE for blue sharks, while we observed the exact opposite effect for shortfin mako sharks, in that fish bait increased CPUE and MPUE—and even haulback mortality rate. Whole fish bait also reduced MPUE in bigeye tuna, while in common dolphinfish and escolar, it tended to increase MPUE and CPUE. We observed few effects of bait type in billfishes. Although few studies have examined the effects of bait type alone, an increase in catch rate by fish bait has been noted for shortfin mako sharks (Foster et al. 2012). In swordfish, studies evaluating the effect of switching to whole fish bait from squid bait have reported conflicting effects (increase: Santos et al. 2012; decrease: Foster et al. 2012). Very few studies have assessed haulback mortality by bait types other than those on sea turtles, but in blue sharks, the use of fish bait reduced haulback mortality (Epperly et al. 2012). The catch rates of the target species, as previously noted, were affected by bait texture, but a mechanistic explanation for mortality effects is lacking. Since the likelihood that differences in feeding behavior among species have an effect is high, this issue could be resolved through comparative studies based on observations of feeding behavior, as in the case of circle hooks.

When we considered the effects of hook and bait type simultaneously, bait types were shown to have a relatively greater impact than hook types. For example, in shortfin mako sharks and dolphinfish, the mortality-increasing effect of fish bait was further amplified by that of the circle hook, while in blue sharks the mortality-increasing effect of the circle hook was counteracted by the bait effect. Although the combination of circle hooks and fish bait is considered to be very useful for avoiding sea turtle bycatch, this very combination may pose a high mortality risk for some species, and even in the case of target fishes, may counteract the expected positive effect of the circle hook on catch rate. Although very limited studies have simultaneously examined the interrelationship between hook and bait types, all studies support the conclusion that the combination of hook type and bait type causes fluctuations in catch rates and that the direction of response varies among species (Coelho et al. 2012; Foster et al. 2012; Fernandez-Carvalho et al. 2015).

Water temperature and soak time emerged as significant factors affecting haulback mortality rate. This indicates the significance of these environmental factors, and the need to control them statistically or experimentally when assessing the effects of hook and bait type on mortality rate. The effect of water temperature—particularly during the depth and time of day when hooked—and changes in water temperature up to the time the fish is landed, are considered to be influential. In addition, in high

water temperature environments, studies have identified an increased risk of suffocation due to decreased dissolved oxygen in water and increased physiological metabolic rate (Skomal & Bernal 2010); Gallagher et al. (2014) reported an increase in haulback mortality rate for four shark species when caught during high water temperatures. In addition, for the species that adopt rum ventilation, prolonged soak time inevitably increases the risk of suffocation due to the restriction of swimming behavior by being hooked. Mortality rates of tuna, swordfish, and sharks reportedly increased with increasing soak time (Epperly et al. 2012; Gallagher et al. 2014).

Here, we quantified our data through experimental operations that standardized the various conditions, but not all aspects were completely controlled. For example, while previous studies on hook size have examined the correspondence with actual measurements (Gilman et al. 2016), several shapes of circle hook were used in the experiment in this study, precluding examination of effects of individual hook types due to sample size issues. We were also unable to examine hooking location of the catch when fish bait was used. These omissions, while having a limited impact on the present conclusions, are probably variables that should be considered for a deeper examination of the effects of terminal gear on catch and bycatch. In this experiment, wire leaders were used on all branchlines to minimize the effects of sharks' bite-off. While some studies have described concerns that wire leaders may increase catch rates, especially for rare sharks, they are considered essential for at least experimentally verifying accurate catch and mortality rates for shark species. We know from this and previous studies that haulback mortality rates for sharks are much lower than those for teleosts (Reinhardt et al. 2018), and the implementation of appropriate releases, even with wire leaders, allow for the reduction of risk for vulnerable shark species.

Here, based on a Bayesian approach, we succeeded in presenting a quantitative impact assessment of terminal gear on teleosts, sharks, and sea turtles by directly calculating the expected values for mortality rate, CPUE, and MPUE with each terminal gear. Calculating MPUE using this model can be a very useful tool because it provides a more direct estimate than does CPUE or mortality rate alone of catch/bycatch risk to populations of those species. Although we did not include post-release mortality rate in the model due to lack of data, it would be possible to estimate overall fishing mortality in the model by designing additional experiments so that mark–recapture is conducted at the same time. Even if it is not possible to use wire leaders for the proportion of “cryptic catch” due to bite-off, it is

possible to extrapolate this proportion into the model to make predictions regarding mortality—a development we anticipate.

In this study, we successfully conducted a quantitative impact assessment of bycatch integrating catch rate and mortality using a Bayesian approach that directly predicts mortality and CPUE. Although the data used in the analysis relied solely on the results of an Asian-style longline experiment in the Pacific Ocean and may therefore contain inherent biases, the same analysis method can be used in conjunction with data from other experiments conducted in other areas and fishing styles to provide a more integrated assessment.

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Table 1 List of hook types used in the experiment

| hook name | size | category |
|---|---|---|
| Komatu Keisaku tunahook | 4.0 sun | tuna |
| Doitomi Tunamutsu | 4.0 sun | small-C (circle hook) |
| Komatsu Keisaku modified | 4.0 sun | small-C (circle hook) |
| Komatsu Keisaku modified | 4.5 sun | small-C (circle hook) |
| Komatsu Keisaku modified | 4.8 sun | small-C (circle hook) |
| Komatsu Keisaku type Etsuna | 4.5 sun | small-C (circle hook) |
| Komatsu Keisaku type North America | 4.3 sun | small-C (circle hook) |
| Tankichi Uruwa | 3.8 sun | small-C (circle hook) |
| Komatsu Keisaku modified | 5.2 sun | large-C (circle hook) |
| Komatsu Keisaku type North America | 5.2 sun | large-C (circle hook) |
| Pacific Fishing Tackle circle hook | 18/0 | large-C (circle hook) |

Table 2 Fishing effort (longline hooks) in experimental operations used in the analysis.

| hook type | bait type | | total effort |
|---|---|---|---|
| | squid | fish | |
| tuna | 97,834 | 71,146 | 168,980 |
| small-C | 70,882 | 7,658 | 78,540 |
| large-C | 31,872 | 6,976 | 38,848 |
| total effort | 200,588 | 85,780 | 286,368 |

Table 3 Number of fish caught by species in the experimental operation

| species | total catch |
|---|---|
| blue shark | 13,018 |
| longnose lancetfish | 1,297 |
| common dolphinfish | 505 |
| shortfin mako shark | 485 |
| swordfish | 288 |
| bigeye tuna | 269 |
| loggerhead turtle | 268 |
| escolar | 163 |
| striped marine | 145 |
| other species | 1,578 |

Table 4 Number of catch by hook and bait type.

| species | hook type | | | bait type | |
|---|---|---|---|---|---|
| | tuna | small-C | large-C | squid | fish |
| blue shark | 8,084 | 3,562 | 1,372 | 8,903 | 4,115 |
| longnose lancetfish | 945 | 257 | 95 | 692 | 605 |
| common dolfinfish | 206 | 181 | 118 | 363 | 142 |
| shortfin mako shark | 298 | 134 | 53 | 262 | 223 |
| swordfish | 129 | 112 | 47 | 249 | 39 |
| bigeye tuna | 114 | 146 | 9 | 201 | 68 |
| loggerhead turtle | 128 | 113 | 27 | 259 | 9 |
| escolar | 79 | 56 | 28 | 108 | 55 |
| striped marine | 70 | 53 | 22 | 126 | 19 |

Table 5 Composition of hooking location and fate at hauling.

| species | hooking location | | | | fate at haulback | | |
|---|---|---|---|---|---|---|---|
| | swallowed | mouth | body | unknown(*) | alive | dead | unknown(*) |
| blue shark | 2,270 | 2,608 | 66 | 8,074 | 11,701 | 1,066 | 251 |
| longnose lancetfish | 25 | 288 | 16 | 968 | 165 | 1,010 | 122 |
| common dolphinfish | 39 | 159 | 4 | 303 | 398 | 87 | 20 |
| shortfin mako shark | 66 | 68 | 18 | 333 | 363 | 118 | 4 |
| swordfish | 59 | 118 | 18 | 93 | 55 | 226 | 7 |
| bigeye tuna | 7 | 128 | 2 | 132 | 83 | 183 | 3 |
| loggerhead turtle | 109 | 116 | 14 | 29 | 256 | 5 | 7 |
| escolar | 1 | 62 | 1 | 99 | 108 | 40 | 15 |
| striped marine | 12 | 80 | 8 | 45 | 68 | 76 | 1 |

(*) Includes catches that dropped off before reseachers checked or lack of survey.

Table 6 Estimates of the posterior distribution of the proportion of hooking location by hook when squid bait is used (median). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

| species | hook type | | |
|---|---|---|---|
| | tuna | small-C | large-C |
| **hooking place: body** | | | |
| blue shark | 0.020 (0.015 - 0.027) | 0.008 (0.005 - 0.012) | 0.005 (0.001 - 0.014) |
| loggerhead turtle | 0.025 (0.006 - 0.066) | 0.079 (0.037 - 0.143) | 0.031 (0.001 - 0.154) |
| longnose lancetfish | 0.049 (0.020 - 0.092) | 0.031 (0.011 - 0.067) | 0.074 (0.018 - 0.185) |
| shortfin mako | 0.102 (0.041 - 0.199) | 0.107 (0.052 - 0.185) | 0.136 (0.021 - 0.381) |
| striped marine | 0.083 (0.026 - 0.186) | 0.068 (0.016 - 0.173) | 0.048 (0.002 - 0.235) |
| swordfish | 0.075 (0.030 - 0.148) | 0.109 (0.057 - 0.184) | 0.031 (0.001 - 0.154) |
| **hooking place: mouth** | | | |
| blue shark | 0.497 (0.475 - 0.518) | 0.548 (0.528 - 0.568) | 0.560 (0.517 - 0.603) |
| loggerhead turtle | 0.439 (0.346 - 0.534) | 0.480 (0.381 - 0.577) | 0.746 (0.544 - 0.892) |
| longnose lancetfish | 0.857 (0.792 - 0.908) | 0.903 (0.848 - 0.943) | 0.872 (0.740 - 0.953) |
| shortfin mako | 0.356 (0.238 - 0.486) | 0.524 (0.418 - 0.630) | 0.379 (0.154 - 0.650) |
| striped marine | 0.759 (0.623 - 0.869) | 0.831 (0.696 - 0.924) | 0.883 (0.660 - 0.982) |
| swordfish | 0.528 (0.416 - 0.637) | 0.676 (0.574 - 0.767) | 0.611 (0.405 - 0.790) |
| **hooking place: swallowed** | | | |
| blue shark | 0.483 (0.461 - 0.505) | 0.444 (0.424 - 0.464) | 0.434 (0.391 - 0.478) |
| loggerhead turtle | 0.533 (0.438 - 0.626) | 0.438 (0.343 - 0.537) | 0.210 (0.078 - 0.405) |
| longnose lancetfish | 0.092 (0.052 - 0.148) | 0.064 (0.032 - 0.111) | 0.046 (0.007 - 0.145) |
| shortfin mako | 0.536 (0.405 - 0.665) | 0.365 (0.266 - 0.472) | 0.460 (0.213 - 0.724) |
| striped marine | 0.150 (0.066 - 0.273) | 0.093 (0.030 - 0.207) | 0.048 (0.002 - 0.234) |
| swordfish | 0.394 (0.289 - 0.505) | 0.211 (0.136 - 0.303) | 0.343 (0.174 - 0.550) |

Table 7 Estimated haulback mortality rates (median of posterior distribution) by hooking location when squid bait is used. Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses. Loggerhead turtles were excluded because there were no mortalities and the calculation had not been converged.

| species | body | mouth | swallowed |
|---|---|---|---|
| blue shark | 0.295 (0.185 - 0.411) | 0.052 (0.044 - 0.061) | 0.157 (0.141 - 0.172) |
| shortfin mako shark | 0.496 (0.262 - 0.729) | 0.134 (0.061 - 0.219) | 0.279 (0.176 - 0.394) |
| longnose lancetfish | 0.956 (0.810 - 1.000) | 0.860 (0.816 - 0.899) | 0.798 (0.621 - 0.934) |
| swordfish | 0.802 (0.595 - 0.959) | 0.776 (0.694 - 0.853) | 0.919 (0.842 - 0.979) |
| striped marine | 0.798 (0.488 - 0.990) | 0.431 (0.315 - 0.547) | 0.968 (0.839 - 1.000) |

Table 8 Haulback mortality rate by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

| species | hook type | | |
|---|---|---|---|
| | tuna | small-C | large-C |
| bait type : squid | | | |
| bigeye tuna | 0.436 ( 0.036 - 1.231) | 1.033 ( 0.100 - 2.938) | 0.896 ( 0.067 - 2.980) |
| blue shark | 44.457 (35.464 - 54.243) | 66.395 (53.022 - 81.268) | 56.701 (45.075 - 69.726) |
| common dolphinfish | 0.709 ( 0.336 - 1.182) | 1.676 ( 0.775 - 2.811) | 1.193 ( 0.534 - 2.020) |
| escolar | 0.213 ( 0.074 - 0.408) | 0.630 ( 0.225 - 1.216) | 0.353 ( 0.108 - 0.724) |
| longnose lancetfish | 3.326 ( 2.742 - 3.993) | 4.677 ( 3.757 - 5.683) | 4.109 ( 3.074 - 5.261) |
| shortfin mako | 1.051 ( 0.686 - 1.458) | 1.723 ( 1.084 - 2.431) | 1.547 ( 0.939 - 2.364) |
| striped marine | 0.320 ( 0.169 - 0.512) | 0.369 ( 0.181 - 0.603) | 0.356 ( 0.155 - 0.640) |
| swordfish | 0.731 ( 0.496 - 1.004) | 1.079 ( 0.706 - 1.504) | 1.115 ( 0.673 - 1.661) |
| bait type : fish | | | |
| bigeye tuna | 0.185 ( 0.014 - 0.550) | 0.441 ( 0.050 - 1.323) | 0.383 ( 0.016 - 1.342) |
| blue shark | 37.299 (29.425 - 45.517) | 55.736 (44.428 - 68.588) | 47.620 (37.760 - 58.788) |
| common dolphinfish | 1.781 ( 0.858 - 3.031) | 4.210 ( 2.019 - 7.284) | 3.000 ( 1.404 - 5.264) |
| escolar | 0.476 ( 0.153 - 0.941) | 1.410 ( 0.443 - 2.826) | 0.787 ( 0.221 - 1.712) |
| longnose lancetfish | 3.544 ( 2.849 - 4.346) | 4.982 ( 3.904 - 6.216) | 4.379 ( 3.172 - 5.761) |
| shortfin mako | 2.327 ( 1.518 - 3.322) | 3.813 ( 2.453 - 5.622) | 3.418 ( 1.913 - 5.221) |
| striped marine | 0.371 ( 0.163 - 0.663) | 0.428 ( 0.170 - 0.777) | 0.411 ( 0.144 - 0.822) |
| swordfish | 0.751 ( 0.456 - 1.116) | 1.107 ( 0.636 - 1.672) | 1.142 ( 0.618 - 1.848) |

Table 9 Standardized catch per unit effort (CPUE) by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

| species | hook type | | |
|---|---|---|---|
| | tuna | small-C | large-C |
| bait type : squid | | | |
| bigeye tuna | 0.734 (0.638 - 0.823) | 0.757 (0.679 - 0.830) | 0.960 (0.812 - 1.000) |
| blue shark | 0.084 (0.076 - 0.092) | 0.081 (0.072 - 0.091) | 0.077 (0.062 - 0.093) |
| common dolphinfish | 0.154 (0.097 - 0.214) | 0.128 (0.081 - 0.180) | 0.180 (0.107 - 0.262) |
| escolar | 0.189 (0.085 - 0.311) | 0.300 (0.163 - 0.451) | 0.181 (0.058 - 0.336) |
| longnose lancetfish | 0.901 (0.869 - 0.929) | 0.925 (0.893 - 0.954) | 0.890 (0.827 - 0.941) |
| shortfin mako | 0.161 (0.107 - 0.223) | 0.186 (0.120 - 0.253) | 0.181 (0.084 - 0.300) |
| striped marine | 0.531 (0.401 - 0.665) | 0.472 (0.334 - 0.610) | 0.530 (0.321 - 0.740) |
| swordfish | 0.829 (0.755 - 0.894) | 0.783 (0.700 - 0.859) | 0.832 (0.711 - 0.933) |
| bait type : fish | | | |
| bigeye tuna | 0.473 (0.326 - 0.627) | 0.503 (0.339 - 0.666) | 0.886 (0.570 - 1.000) |
| blue shark | 0.078 (0.068 - 0.088) | 0.075 (0.062 - 0.088) | 0.071 (0.055 - 0.088) |
| common dolphinfish | 0.237 (0.158 - 0.326) | 0.200 (0.103 - 0.312) | 0.273 (0.164 - 0.398) |
| escolar | 0.242 (0.114 - 0.389) | 0.370 (0.174 - 0.586) | 0.231 (0.061 - 0.472) |
| longnose lancetfish | 0.909 (0.880 - 0.937) | 0.931 (0.895 - 0.960) | 0.898 (0.839 - 0.945) |
| shortfin mako | 0.308 (0.242 - 0.387) | 0.347 (0.227 - 0.479) | 0.342 (0.197 - 0.488) |
| striped marine | 0.691 (0.457 - 0.883) | 0.640 (0.368 - 0.865) | 0.692 (0.384 - 0.920) |
| swordfish | 0.853 (0.728 - 0.947) | 0.811 (0.651 - 0.935) | 0.856 (0.700 - 0.967) |

Table 10 Estimated MPUE by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

| species | hook type | | |
|---|---|---|---|
| | tuna | small-C | large-C |
| bait type：squid | | | |
| bigeye tuna | 0.318 (0.026 - 0.902) | 0.780 (0.064 - 2.214) | 0.841 (0.072 - 2.828) |
| blue shark | 3.744 (2.910 - 4.646) | 5.388 (4.152 - 6.790) | 4.370 (3.163 - 5.756) |
| common dolphinfish | 0.108 (0.043 - 0.196) | 0.213 (0.091 - 0.399) | 0.214 (0.080 - 0.405) |
| escolar | 0.040 (0.009 - 0.089) | 0.187 (0.052 - 0.403) | 0.063 (0.009 - 0.162) |
| longnose lancetfish | 2.994 (2.443 - 3.586) | 4.321 (3.440 - 5.239) | 3.641 (2.706 - 4.706) |
| shortfin mako | 0.169 (0.088 - 0.263) | 0.321 (0.156 - 0.498) | 0.278 (0.102 - 0.535) |
| striped marine | 0.169 (0.076 - 0.279) | 0.173 (0.075 - 0.299) | 0.185 (0.065 - 0.360) |
| swordfish | 0.604 (0.403 - 0.838) | 0.840 (0.538 - 1.186) | 0.919 (0.533 - 1.385) |
| bait type：fish | | | |
| bigeye tuna | 0.087 (0.007 - 0.267) | 0.219 (0.016 - 0.675) | 0.319 (0.016 - 1.158) |
| blue shark | 2.909 (2.229 - 3.684) | 4.185 (3.112 - 5.415) | 3.396 (2.408 - 4.571) |
| common dolphinfish | 0.420 (0.168 - 0.765) | 0.836 (0.288 - 1.664) | 0.812 (0.286 - 1.558) |
| escolar | 0.113 (0.026 - 0.261) | 0.511 (0.113 - 1.177) | 0.178 (0.022 - 0.512) |
| longnose lancetfish | 3.217 (2.582 - 3.953) | 4.629 (3.668 - 5.846) | 3.919 (2.800 - 5.169) |
| shortfin mako | 0.712 (0.432 - 1.063) | 1.311 (0.679 - 2.117) | 1.153 (0.527 - 2.046) |
| striped marine | 0.250 (0.089 - 0.469) | 0.264 (0.081 - 0.525) | 0.273 (0.072 - 0.589) |
| swordfish | 0.633 (0.375 - 0.959) | 0.883 (0.466 - 1.364) | 0.961 (0.497 - 1.590) |

Table 11 Nominal CPUE, Haulback mortality rate and MPUE of Loggerhead turtle。All figures are based on aggregated operational data, not estimates.

| hook type | CPUE | | haulback mortality | | MPUE | |
|---|---|---|---|---|---|---|
| | squid | fish | squid | fish | squid | fish |
| tuna | 1.247 | 0.084 | 0.0385 | 0.000 | 0.0480 | 0.000 |
| large-C | 1.034 | 0.277 | 0.0090 | 0.000 | 0.0093 | 0.000 |
| small-C | 2.148 | 0.261 | 0.0246 | 0.000 | 0.0528 | 0.000 |

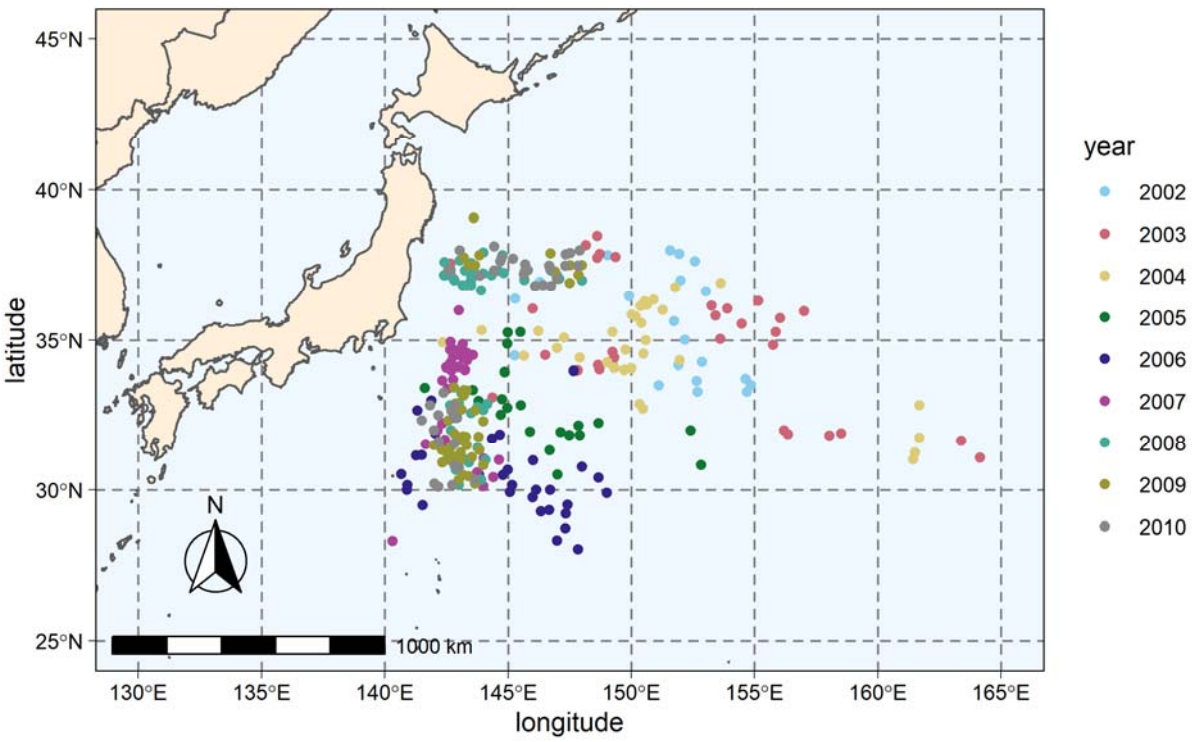



Figure 1 Locations where the longline operation experiment was conducted.

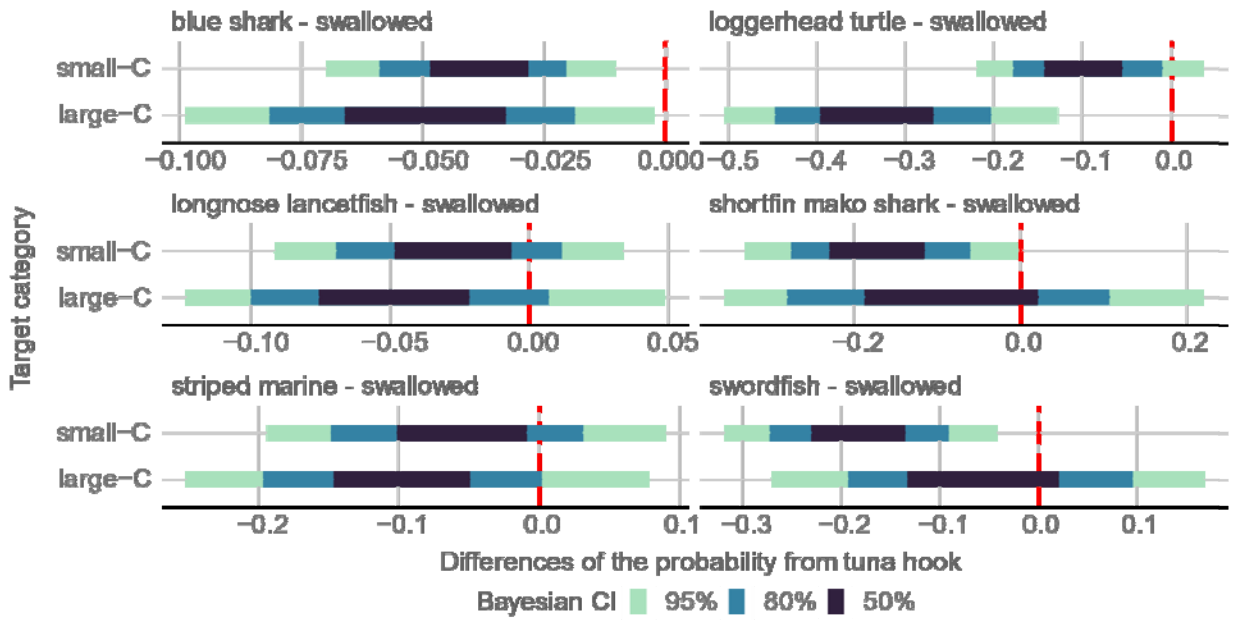



Figure 2 Differences in the estimated probability of the “swallowed” hooking position of each circle hook type from tuna hook when squid bait is used. The red dotted line indicates that the difference is zero.

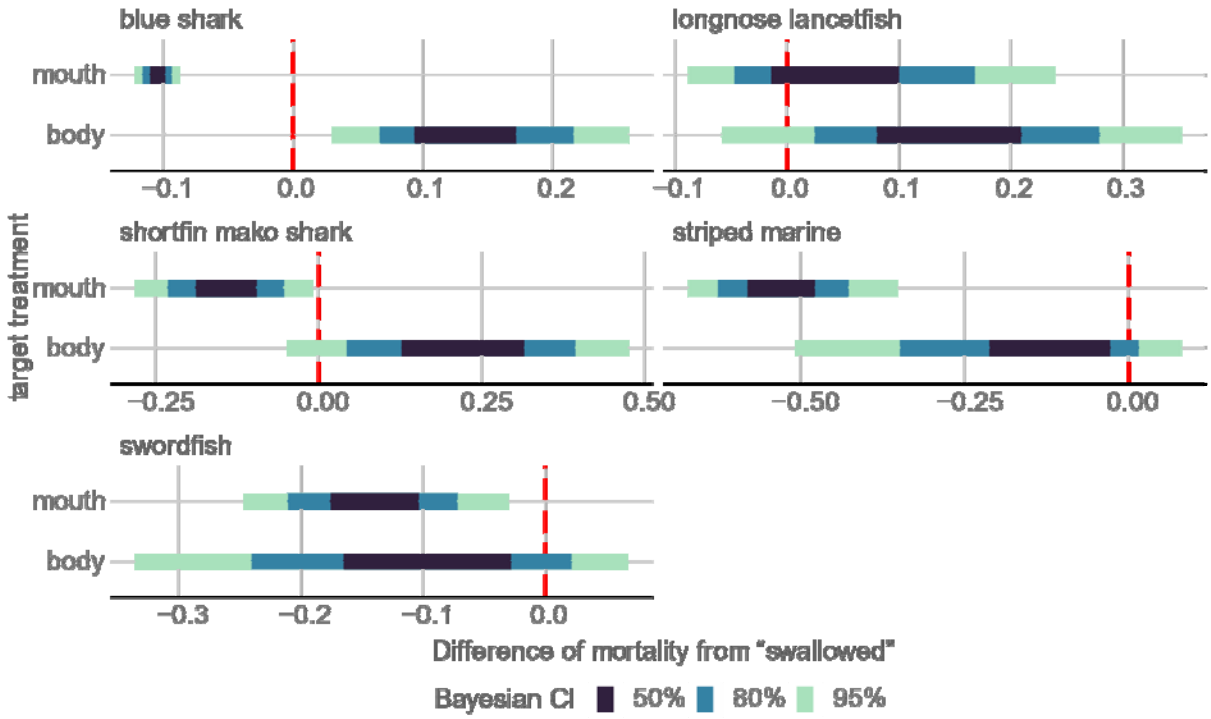



Figure 3 Differences in the estimated haulback mortality rate of each target hooking position from "swallowed" hooking position when squid bait is used. The red dotted line indicates that the difference is zero.

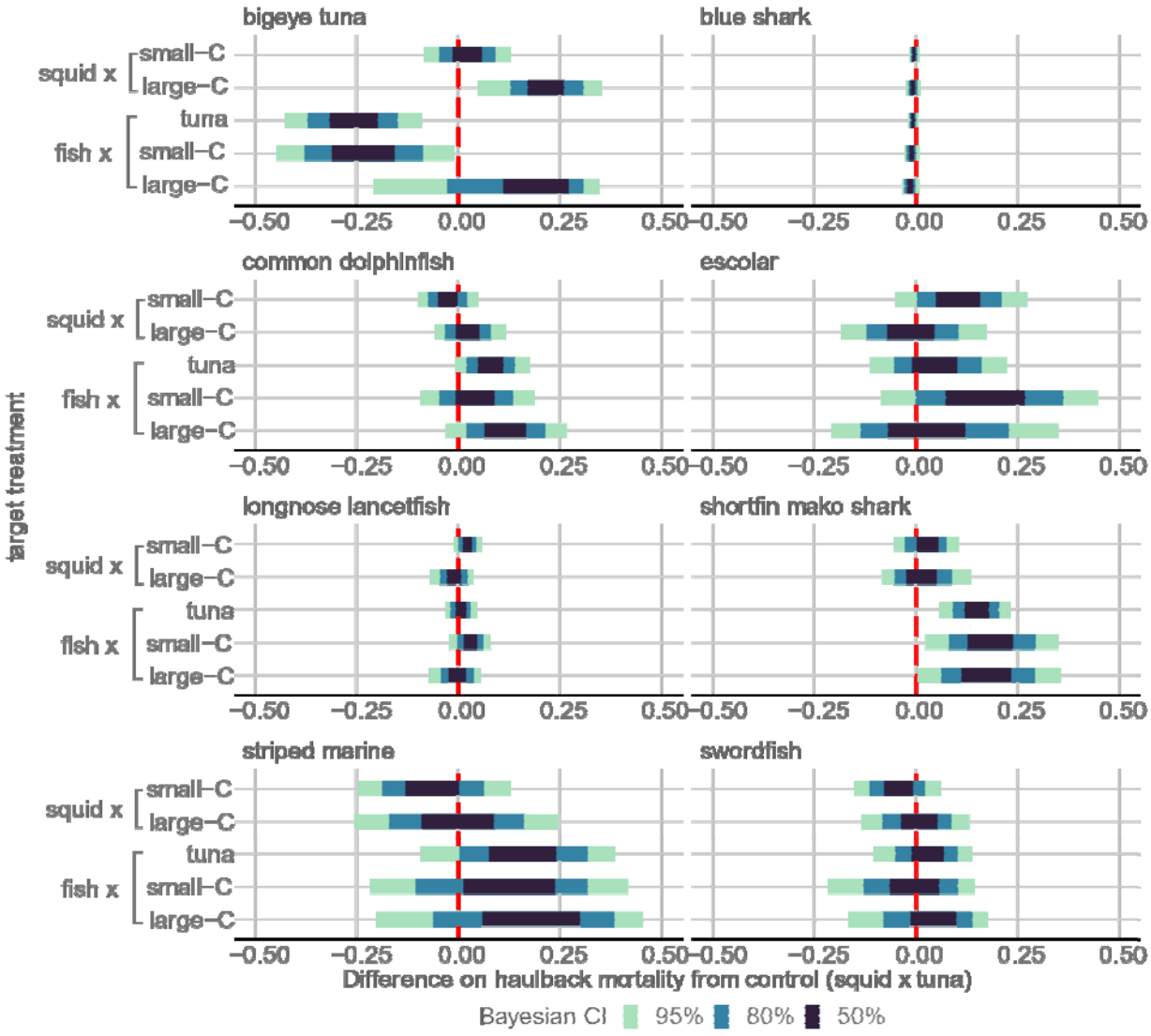



Figure 4 Differences in estimated haulback mortality between each experimental group and the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

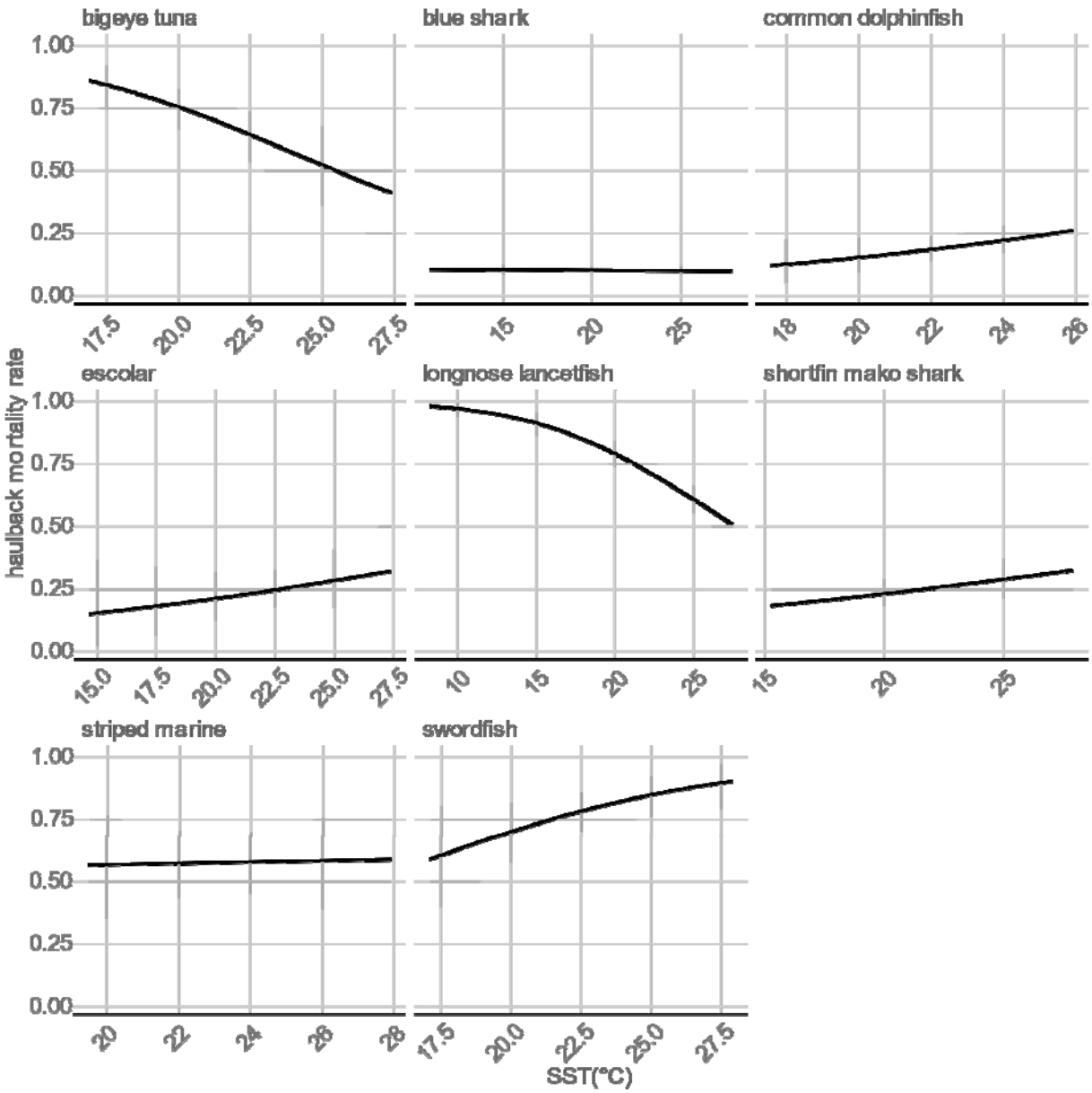



Figure 5 Relationship between sea surface temperature (SST) variability and haulback mortality rate at longline operations. Solid lines indicate median; masked areas indicate 95% Bayesian credible interval.

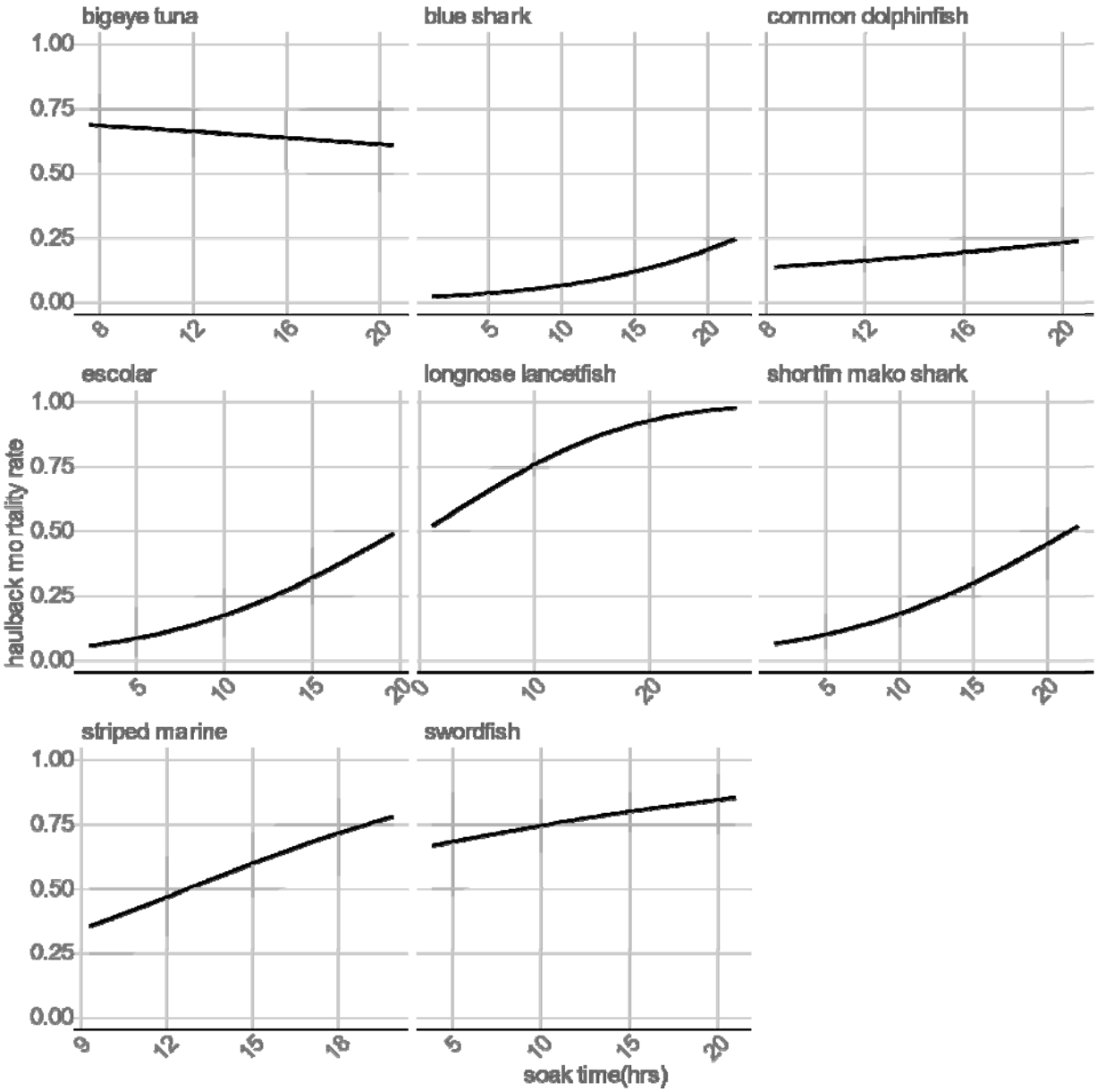



Figure 6 Relationship between soak time (time from setting the branch line to hauling) variability and haulback mortality rate at longline operations. Solid lines indicate median; masked areas indicate 95% Bayesian credible interval.

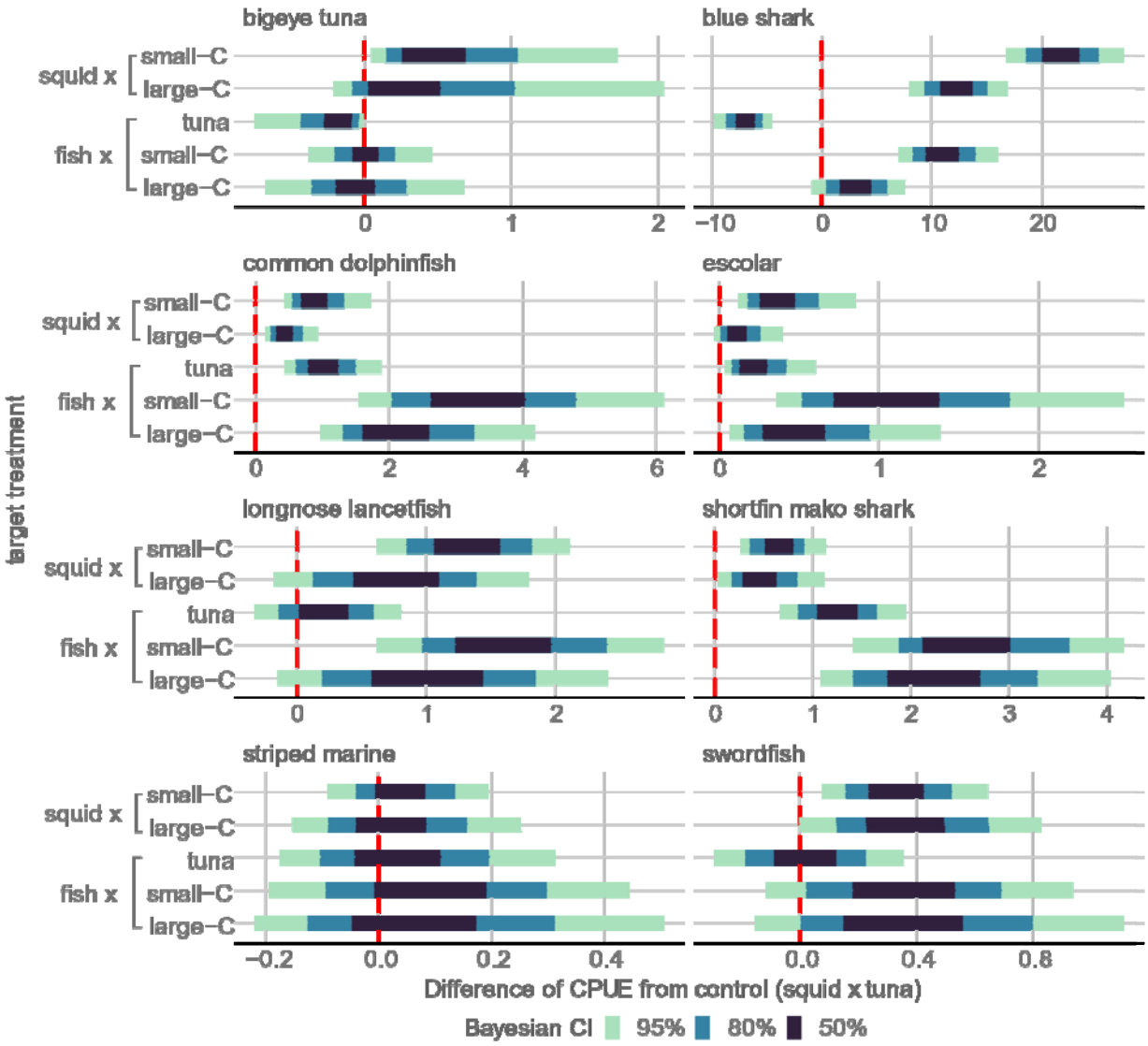



Figure 7 Differences in the standardized catch per unit effort (CPUE) for each experimental group from those for the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

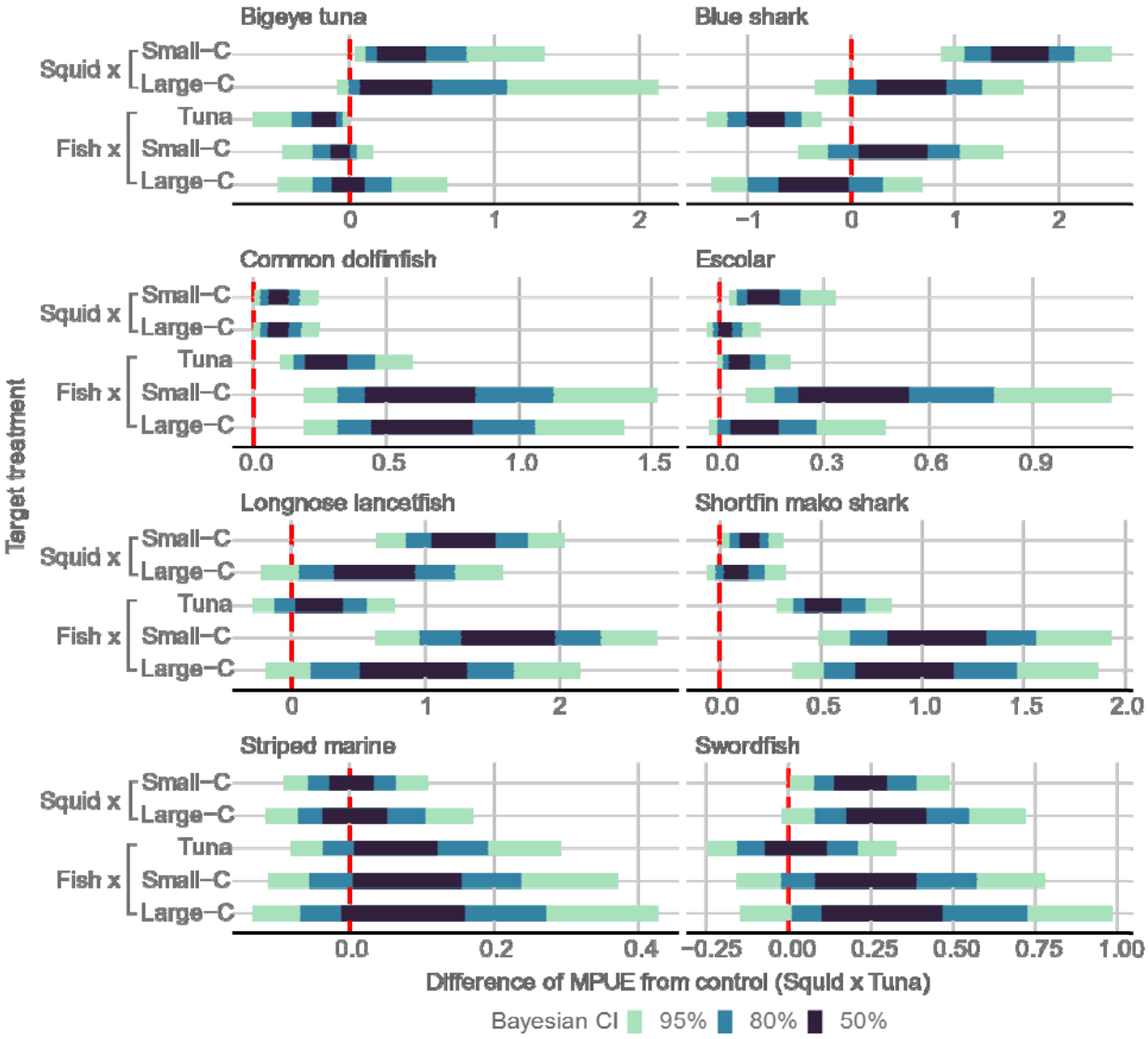



Figure 8 Differences in the estimated MPUE (mortality per unit effort) between those for each experimental group and those for the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

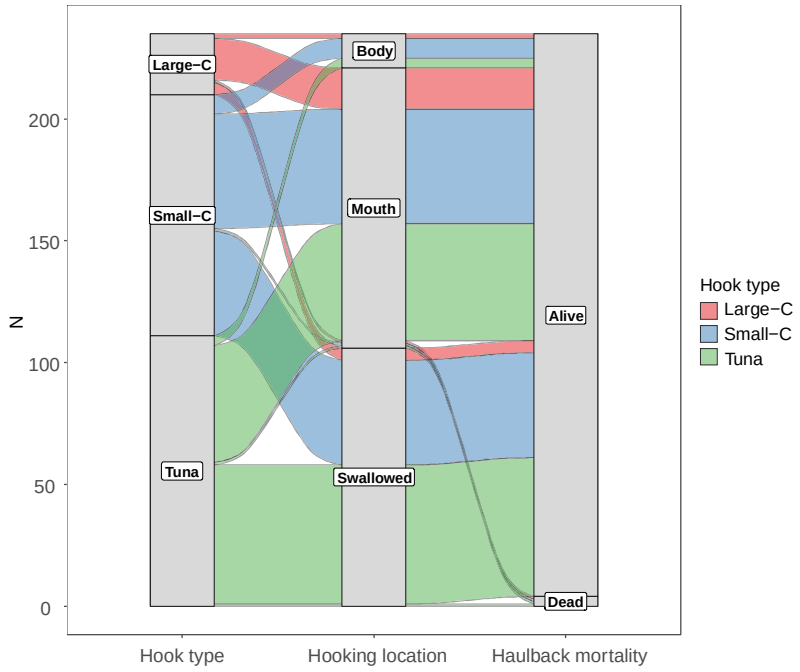
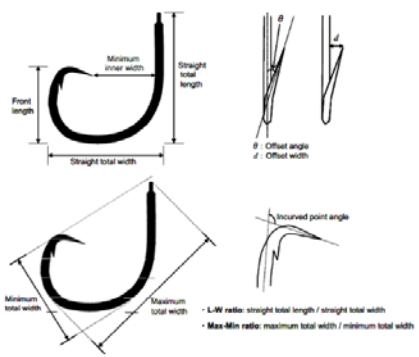



Figure 9 Alluvial plot of hooking locations and associated haulback mortality rates of loggerhead turtles by hook when squid bait is used.

Appendix 1 Detailed measurements of hooks used in the experiment and a figure explaining measurement points of hooks (copied from Yokota et al. 2006b).

| **manufacture** | **Komatsu Keisaku** | **Hisamatsu Tankichi** | **Doitomi** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Komatsu Keisaku** | **Pacific Fishing Tackle MFG., CO.** |
|---|---|---|---|---|---|---|---|---|---|---|---|
| **hook name** | tuna hook | Uruwa hook BKN | tuna circle hook SS-170 | modified circle hook | modified circle hook | modified circle hook | cirlce hook type Etsuna | cirlce hook type North America | cirlce hook type North America | modified circle hook | circle hook |
| **standardized size** | 4.0 sun | 3.8 sun | #4 | 4.0 sun | 4.5 sun | 4.8 sun | 4.5 sun | 4.3 sun | 5.2 sun | 5.2 sun | 18/0 |
| **material** | stainless steel | hard steel | stainless steel | stainless steel | stainless steel | stainless steel | stainless steel | hard steel | hard steel | stainless steel | stainless steel |
| **hook eye** | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring | yes, with ring |
| **shank thickness(mm)** | 5.3 | 4.0 | 4.1 | 5.3 | 5.2 | N/A | 4.9 | 5.3 | 5.7 | 5.3 | 5.1 |
| **straight total length(mm)** | 63 | 56 | 58 | 60 | 63 | N/A | 62 | 57 | 74 | 76 | 68 |
| **straight total width(mm)** | 38 | 44 | 49 | 47 | 49 | N/A | 56 | 45 | 56 | 54 | 59 |
| **minimum total width(mm)** | 38 | 36 | 39 | 41 | 45 | N/A | 51 | 41 | 52 | 48 | 51 |
| **maximum total width(mm)** | 63 | 64 | 69 | 68 | 71 | N/A | 76 | 63 | 81 | 85 | 80 |
| **front length(mm)** | 41 | 33 | 38 | 35 | 44 | N/A | 47 | 39 | 49 | 47 | 45 |
| **minimum inner width(mm)** | 27 | 20 | 15 | 24 | 25 | N/A | 26 | 20 | 27 | 26 | 27 |
| **L-W ratio** | 1.7 | 1.3 | 1.2 | 1.3 | 1.3 | N/A | 1.1 | 1.3 | 1.3 | 1.4 | 1.2 |
| **max-min ratio** | 1.7 | 1.8 | 1.8 | 1.7 | 1.6 | N/A | 1.5 | 1.5 | 1.6 | 1.8 | 1.6 |
| **incurved point angle** | | 80 | 70 | 90 | 70 | N/A | 70 | 65 | 75 | 80 | 80 |
| **offset angle** | $5° \leqq \theta < 10°$ | $\theta \approx 0°$ | $\theta \approx 0°$ | $\theta \approx 10°$ | $5° \leqq \theta < 10°$ | $\theta < 10°$ | $\theta \approx 10°$ | $5° \leqq \theta < 10°$ | $10° \leqq \theta < 15°$ | $\theta \approx 5°$ | $10° \leqq \theta < 15°$ |
| **offset width(mm)** | 0.9 | $\approx 0$ | $\approx 0$ | 1.8 | 2.6 | N/A | 2.1 | 1.5 | 3.7 | 1.0 | 5.4 |
| **weight (g)** | 19.9 | 12.2 | 15.0 | 19.7 | 21.6 | N/A | 21.4 | 19.4 | 30.3 | 25.5 | 23.2 |



Appendix 2 Stan code used for MCMC sampling of (a) Model 1 and (b) Model 2

(a) MODEL 1

```
data {
  int HK_cat;
  int<lower=1> N1;
  int<lower=1> LOC_cat;
  int<lower=1, upper=LOC_cat> LOC[N1];
  int<lower=1, upper=HK_cat> HK[N1];
  int<lower=0, upper=1> M[N1];
  int L;
  matrix[N1, L] X1;
  int N2;
  int  Catch[N2];
  int L2;
  matrix[N2, L2] X2;
  vector[N2] O;
  int YRID;
  int YR[N2];
}

transformed data{
  vector[HK_cat] Zeros;
  Zeros = rep_vector(0, HK_cat);
}

parameters {
  matrix[HK_cat, LOC_cat - 1] theta_raw;
  vector[L] beta;
  vector[L2] beta2;
  vector[YRID] r;
  real<lower= 0> sigma;

}

transformed parameters {
  vector[N2] mu;
  vector[N1] phi;
  matrix[HK_cat, LOC_cat] theta;
  phi = inv_logit(X1 * beta);
  theta = append_col(Zeros, theta_raw);
  for(i in 1:N2)
  mu[i] = X2[i,] * beta2 + r[YR[i]];

model {
  sigma ~ student_t(4, 0, 2.5);
  r ~ normal(0, sigma);
  for(n in 1:N1){
  M[n] ~ bernoulli(phi[n]);
  target += categorical_lpmf(LOC[n]|softmax(theta[HK[n],]'));
  }
  Catch ~ poisson(exp(mu + log(O)));
}
```

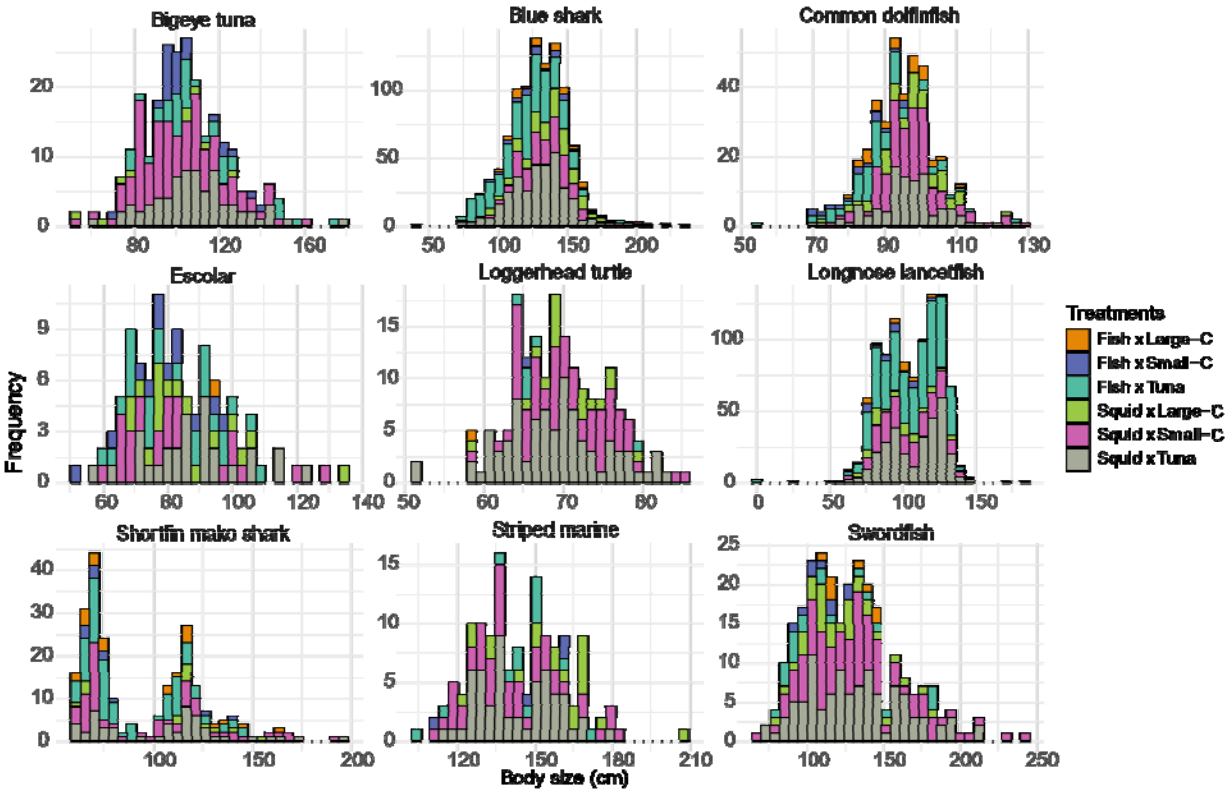
(b) MODEL 2

```
data {
  int HK_BAIT_comb;
  int<lower=1> N1;
  int<lower=0, upper=1> M[N1];
  int L;
  matrix[N1, L] X1;
  int N2;
  int  Catch[N2];
  int L2;
  matrix[N2, L2] X2;
  vector[N2] O;
  int YRID;
  int YR[N2];
}

parameters {
  vector[L] beta;
  vector[L2] beta2;
  vector[YRID] r;
  real<lower= 0> sigma;
}

transformed parameters {
  vector[N2] mu;
  vector[N1] phi;
  phi = inv_logit(X1 * beta);
  for(i in 1:N2)
  mu[i] = X2[i,] * beta2 + r[YR[i]];
}

model {
  sigma ~ student_t(4, 0, 2.5);
  r ~ normal(0, sigma);
  for(n in 1:N1){
  M[n] ~ bernoulli(phi[n]);
  }
  Catch ~ poisson(exp(mu + log(O)));
}
```




Appendix 3 Size distributions by species for the major species captured in the study, with body size as an index of precaudal length for blue sharks and shortfin mako sharks, straight-line carapace length for loggerhead turtles, eye-to-fork length for striped marine and swordfish, and fork length for all other species.