TITLE:

Multifaceted effects of bycatch mitigation measures on target/non-target species for pelagic longline fisheries and consideration for bycatch management

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SUMMARY

The pelagic longline fishery, in an effort to reduce bycatch of sea turtles, have developed and deployed fisheries bycatch mitigation techniques such as replacing J/tuna hooks and squid bait with circle hooks and whole fish bait. However, little emphasis has been placed on the side effects of bycatch mitigation measures on endangered species other than target bycatch species. Several previous studies of the side effects have been marred by lack of control for the covariates. Here, based on long-term data obtained from research cruises by a pelagic longline vessel, we examined the effects of using circle hooks and whole fish bait to replace squid bait on the fishing mortality of target and non-target fishes, and also bycatch species. A quantitative evaluation analysis of our results, based on a Bayesian approach, showed the use of circle hooks to increase mouth hooking in target and bycatch species, and their size to be proportional to the magnitude of the effect. Although deploying circle hooks increased fishing mortality per unit effort (MPUE) for shortfin mako shark, the magnitude of changing the bait species from squid to fish clearly had a far greater impact on MPUE than the use of circle hooks. Because the impact of the introduction of bycatch mitigation measures on species other than the focused bycatch species is non-negligible, a quantitative assessment of bycatch mitigation-related fishing mortality is critical before introducing such measures.

Keywords: billfishes, circle hook, finfish bait, fisheries management, sea turtle, sharks, tuna, longline fishery

INTRODUCTION

Unintentional catch by fisheries is referred to as bycatch, and bycatch of especially endangered species, can have devastating effects on such populations. Thus, efforts to minimize bycatch of endangered species are strongly encouraged at all layers from local to international in terms of species conservation. For example, in the tuna longline fishery, concerns about the increased conservation risk for seabirds and sea turtles by unintentional and fatal catch have been a major issue among many regional fisheries management organizations (RFMOs) since the 1990s (Wallace et al. 2013; Dias et al. 2019). In recent years, some elasmobranchs, whose populations are declining further, are also treated as bycatch species. The decline of these species has focused attention at both the national and the international levels. Even for target fish that are not bycatch species, addressing the deterioration of stock status caused by overfishing requires reductions in unintentional fishing mortality (for instance, billfishes in the North Atlantic; Kerstetter & Graves 2006; Diaz 2008).

Several studies have weighed the sustainability of tuna longline fisheries against the conservation of species vulnerable to bycatch (Hall et al. 2000; Melvin et al. 2014; Clarke et al. 2015)—particularly for seabirds and marine turtles—with the development of several effective bycatch mitigation measures (Melvin et al. 2014; Swimmer et al. 2017). Bycatch mitigation measures in longline fisheries target specific animal groups and are evaluated based on their success in reducing mortality due to bycatch of specific vulnerable species. While the impact on the catch of the target fish is the primary consideration when evaluating bycatch mitigation techniques, few studies have examined the impacts and tradeoffs on species not targeted by mitigation techniques (Pacheco et al. 2011; Gilman et al. 2016). However, an ecosystem-based fisheries management approach to the introduction of bycatch reduction measures demands the assessment of the impact not only on the target bycatch species but also on the ecosystem itself in which the target species lives (Reinhardt et al. 2018).

Use of circle hook and whole finfish bait are typical sea turtle bycatch mitigation measures for pelagic longline fisheries (Watson et al. 2005; Gilman et al. 2006; Yokota et al. 2009; Stokes et al. 2011). The tip of the circle hook bends inward, and when a fish or sea turtle swallows the hooked bait, the circle hook cannot hook inside the digestive tract; instead, as the hook exits the mouth, a torque force causes it to hook through the edge of the mouth. This property allows for easy hook removal and has been reported to reduce the mortality rate of bycatch sea turtles on board and after release (Kiyota et al. 2004; Cooke & Suski 2004; Kerstetter & Graves 2006). Reports of positive effects of circle hooks include those for other species—such as reduced haulback and post-release mortality in sharks, reduced post-release mortality in swordfish, and increased catch rates in tuna, the target fish. The use of fish bait instead of squid bait potentially reduces bycatch (Watson et al. 2005; Yokota et al. 2009, 2011) and mortality rates (Stokes et al. 2011; Parga et al. 2015) of sea turtles. These sea turtle mitigation measures, however, exact other costs. Several meta-analytic studies have reported the circle hook-related reduction of sea turtle mortality but increase in billfish and shark catch rates (Gilman et al. 2016; Reinhardt et al. 2018; Santos et al. 2020). The use of whole fish bait also reportedly increases the catch of sharks and other species (Foster et al. 2012). However, few studies have allowed for quantitative evaluation of the effects of these mitigation measures experimentally on species beyond sea turtles. This limited data concern is due in part to the reliance on observer data from commercial vessels, and small sample sizes, small comparison groups, and lack of experimental rigor from research vessels. In addition, the impact assessment for sharks underestimates catch rates and mortality associated with missed catches due to “bite-off” branchline (Reinhardt et al. 2018) and several other issues. In addition, many experimental studies and meta-analyses (e.g., Diaz 2008; Godin et al. 2012; Huang et al. 2016; Pacheco et al. 2011; Yokota et al. 2006a) only evaluate gear impacts at significance levels without evaluating the magnitude of the effect. Small significant differences may be judged not to matter much when assessing overall risk in bycatch species. Studies that controlled for these conditions would allow for an evaluation of adverse effects without the confounding problems described above. Also, many studies use catch/bycatch rates (Andraka et al. 2013; Foster et al. 2012; Gilman 2007; Watson et al. 2005; Yokota et al. 2006a) and mortality rates (at haulback or after release; Carruthers et al. 2009; Gallagher et al. 2014; Horodysky et al. 2005; Kerstetter et al. 2003) as important impact indicators, without considering irreversible impacts, such as the number of organisms killed at the time of catch. Additionally, hooking location itself (mouth, swallow, or external) is believed to have a strong influence on mortality rate, which demands the estimation of risk under specific hooking conditions, taking causal relationships into account.

Here, we used data from controlled experiments to analyze these confounding factors and developed a Bayesian statistical model to evaluate the effects of changing hook and bait species on fish species other than turtles—particularly, tuna, swordfish, and sharks—and then verified the contribution of circle hooks and fish bait to mortality rate, catch rate, and fatal catch rate, respectively. We also discussed the appropriate assessment of bycatch mitigation measures in fisheries management.

METHOD

Experimental Operations

We analyzed data from the R/V Taikei No. 2 longline research operation conducted in the Northwest Pacific Ocean between 2002 and 2010—a typical Japanese shallow-setting operation targeting mainly swordfish and sharks (almost <100 m depth), using four hooks per basket, a wire leader, and a night soaking (Yokota et al. 2006a). A total of 286,363 hooks from 306 operations (range: 400 – 964 hooks per operation) were deployed in the experiment (Table 1). The area of operation ranged from around the Izu Islands in Japan to off the east coast of northeastern Honshu—typical fishing ground for Japanese shallow-setting longliners (Fig. 1; Hiraoka et al. 2016). We regarded the 4.0-sun tuna hooks as a control group against which we assessed 10 different sized circle hooks (Table 2). In Appendix 1 we describe hook shapes and other details of these hooks following the measurement method of Yokota et al. (2006b). Since the circle hooks varied greatly in size (Table 2) and sample numbers were too small to analyze each hook separately, we divided them into two groups according to size for convenience (threshold: straight total length = 68 mm AND maximum total width = 80 mm; approximately equivalent to 5.0-sun or 18/0). Because degree of hook offset has been reported to affect catch and hooking location (Cooke & Suski 2004), most hooks were circle hooks <10°but some were nearly 15°. The bait comprised chub mackerel *Scomber japonicus* and Japanese common squid *Todarodes pacificus* in the range of 20–30 cm in fork length or dorsal mantle length. These were frozen and stored, then completely thawed before being hooked. The sequence of line setting was divided into several experimental segments, and a different combination of hook and bait type was applied for each segment, with an alternate order of segments at each operation.

The researcher recorded catch for each operation for all catch/bycatch, and the species caught, fate of catch (alive or dead), hooking location (mouth, swallow, and external hooking), time of catch, and float ID. The researchers determined if the catch was alive or dead based on the movement of the animals and the degree of injuries before being hauled. Float ID was recorded when a float was dropped during line setting and when it was retrieved to the deck during line hauling in order to calculate soak time. Hooking locations were recorded for catches caught using squid bait. At the start of longline operations, the researcher also collected sea surface temperature with a water thermometer (DS-1; Murayama Denki Ltd.) equipped on the vessel, which we subsequently used in the analysis.

Since the experiment was conducted using multiple sizes (degrees) of circle hooks in combination with whole fish bait and squid bait, we categorized hook shapes into three main types:

1. control (tuna hook 4.0-sun; tuna)
2. smaller circle hook (smaller than the threshold; small-C)

3. larger circle hook (the threshold or larger; large-C)

In Table 1 we show annual longline effort separated by bait and hook type, tabulated according to the above categorization. We selected the following species for analysis: blue shark *Prionace glauca*, shortfin mako shark *Isurus oxyrinchus*, bigeye tuna *Thunnus obesus*, swordfish *Xiphias gladius*, striped marlin *Kajikia audax*, common dolphinfish *Coryphaena hippurus*, escolar *Lepidocybium flavobrunneum* and, longnosed lancetfish *Alepisaurus ferox*, and loggerhead turtle *Caretta caretta*.

Statistical Analysis

We conducted all analyses using a Bayesian approach to estimate parameters. We adopted haulback mortality rate, catch per unit effort (CPUE), and mortality per unit effort (MPUE) (Afonso et al. 2011) as indices to evaluate the impact of hook and bait type on fishing mortality within the analyses. We used the following data as inputs to the model: number of caught (individuals), longline effort (number of hooks), fate at hauling, hooking location, year and location of operation, water temperature at operation, and soaking time.

Because of the heterogeneity of the data set due to not recording the hooking location when fish bait was used, we split the analysis to evaluate the impact of hook type and bait type on fishing mortality into two models. Model 1 evaluated only the effect of hook type based on capture events with squid bait and assumed that the use of circle hooks would change to more mouth hooking of each species, resulting in improved mortality rate. Model 2 evaluated both hook and bait types, and assumed that the combination of the two would result in large fluctuations in mortality rate, CPUE and MPUE for each species.

We based MODEL 1 on a logit regression using a Bernoulli distribution. We express the observed hooking location H and haulback mortality rate M in MODEL 1 by the following equations:

$$M \sim Bernoulli(p_{dead}) \#(1)$$

$$logit(p_{dead}) = \beta_{1,hook} + \beta_{1,sst} + \beta_{1,soaktime} \#(2)$$

$$H \sim Categorical\big(softmax(\theta_{hook})\big) \#(3)$$

where β_1 is the parameter in each explanatory variable, p_{dead} is the expected haulback mortality rate, and θ_{hook} is the expected probability of hooking location.

We structured MODEL 2 to calculate the expected number of mortalities per effort (MPUE) based on the parameters estimated in both the mortality and CPUE estimation subsets. In MODEL 2, due to lack of hooking location data, we calculate p_{dead} from a modified equation (2) as follows;

$$logit(p_{dead}) = \beta_{2,hook} + \beta_{2,bait} + \beta_{2,sst} + \beta_{2,soaktime} \#(4)$$

where β_2 is the parameter in each explanatory variable.

The CPUE subset is almost same as a GLMM analysis in assuming the Poisson distribution as the error structure. We express the number of catches C per operation using the expected CPUE λ as follows:

$$C \sim Poisson(\lambda + \log(E)) \#(5)$$

$$\lambda = \log\big(\gamma_{hook} + \gamma_{bait} + \gamma_{latitude} + \gamma_{sst} + r_{year}\big) \#(6)$$

$$r_{year} \sim N(0, \sigma^2) \#(7)$$

where E is the longline effort, γ is a parameter in each explanatory variable, r_{year} is a random effect of annual fluctuation on CPUE, and σ is a standard deviation.

We obtained the expected values of hook-bait-specific MPUE ζ by multiplying CPUE by at-haulback mortality rate as follows:

$$\zeta = \widehat{p_{dead}} \times \hat{\lambda} \#(9)$$

where $\widehat{p_{dead}}$ denotes the expected mortality rate standardized for hook and bait type, and $\hat{\lambda}$

denotes the estimated CPUE standardized for hook and bait type. We used the standardization method for abundance indices used in fisheries stock assessment (Maunder & Punt 2004). In this method, explanatory variables other than the factor subject to standardization are averaged to predict the objective variable, which in stock assessment is a time scale such as years or months, but in our case, we modified the standardization scale to reference hook type and bait type.

We calculated each parameter based on a Bayesian approach with Markov chain Monte Carlo (MCMC) sampling. For the MCMC sampling, we used cmdstan 2.28.2 (Stan Development Team 2021). As a prior distribution for each parameter, we used a half student-t distribution with 2 degrees of freedom, mean 0, and variance 2.5 for σ, and we used a uniform distribution for the other parameters. Probability density function of half student-t distribution $p_{student-t}$ was defined as;

$$p_{student-t}\,(y|\nu,\mu,\tau) = \frac{\Gamma((\nu+1)/2)}{\Gamma(\nu/2)}\frac{1}{\sqrt{\nu\pi}\tau}\left(1+\frac{1}{\nu}\left(\frac{y-\mu}{\tau}\right)^2\right)^{-(\nu+1)/2}, \quad y \geq 0 \; \# \qquad (10)$$

where y is the random variable, Γ is the gamma function, ν is the degree of freedom, μ is the mean, and τ is the variance. We computed the posterior distribution using Stan with 15000 sampling iterations including 10000 warmup iterations, number of chains as 4, and no sinning. We calculated Bayesian credible intervals based on the highest density interval (HDI) for the estimates. Although the Bayesian approach for the estimates precluded significance testing, we determined the difference between the estimates of the experimental group and those of the control group ("swallowing" for hooking location, "tuna" hook for hook type, and "squid" for bait type), and if the lower and upper limits of HDI for the difference did not exceed 0, we considered the difference as a difference for convenience (assuming region of practical equivalence [ROPE] as 0; Kruschke 2015). In Appendix 2 we show the Stan code used to estimate each parameter in MODEL 1 and 2. For other data handling, statistical analysis, and plotting, we used R4.1.1 (R Core Team 2021) and packages "cmdstanr 0.4.0," "ggalluvial 0.12.3," "ggthemes 4.2.4," "mapdata 2.3.0," "maps 3.4.0," "sf 1.0–2," "tidybayes 3.0.1," and "tidyverse 1.3.1" (Becker & Brownrigg. 2018; Pebesma 2018; Wickham et al. 2019; Brunson

2020; Arnold 2021; Richard et al. 2018, 2021; Gabry & Češnovar 2021; Kay 2021).

RESULTS

Summary Statistics

Sufficient catches/bycatches of blue shark, longnosed lancetfish, common dolphinfish, and shortfin mako shark swordfish, bigeye tuna, loggerhead turtle, escolar and striped marlin were recorded for the later analysis (Table 3). The main species listed in Table 3 as "other species" are listed as follows; salmon shark *Lamna ditropis* (N = 229), pelagic stingray *Pteroplatytrygon violacea* (N = 199), pomflets *Brama spp.* (N = 135), bigeye thresher shark *Alopias superciliosus* (N = 89), and albacore *Thunnus alalunga* (N = 69). The sample sizes of these “other species” were too skewed among experimental groups to converge the later analysis.

In Table 3 we also show the number of fish caught by hook type and bait type. Bigeye tuna had extremely low catches on large-C hooks, and loggerheads had low catches on fish bait. The most common hooking location at the time of catch was mouth hooking for all nine species, with extremely few hook locations other than mouth hooking, especially for bigeye tuna and escolar (Table 4). The proportion of mortality of captured species at haulback varied greatly by species (Table 4). The haulback mortality rate was low for blue sharks, common dolphinfish, escolar, and shortfin mako sharks and, although higher for bigeye tuna, longnose lancetfish, striped marlin, and swordfish, was generally extremely low, especially in loggerhead turtles.

The length-based size composition of the nine species (precaudal length for blue sharks and shortfin mako sharks, straight-line carapace length for loggerhead turtles, eye-to-fork length for striped marlin and swordfish, and fork length for all other species) used in the analysis was not statistically compared in this study and was not included in the model because it did not contribute to haulback mortality rate, but in Appendix 3 we include a histogram.

Almost all parameters in the models for the nine species were successfully converged (Rhat < 1.1) but some of the parameters for loggerhead turtle were not converged, as we describe below.

Output of MODEL 1

In Table 5 we show occurrence estimates of hooking locations by hook type throughout the model. We observed differences in hooking location by hook type for blue shark, common dolphinfish, loggerhead turtle, and swordfish, with an increase in mouth hooking and a decrease in swallowing for large-C for loggerhead turtle and a clear increase in mouth hooking for small-C for shortfin mako shark and swordfish (Fig. 2). In blue sharks, the frequency of swallowing decreased in both small-C and large-C.

In Table 6 we show haulback mortality rates by hooking location. We observed clear differences in haulback mortality rate by hooking location for blue shark, shortfin mako shark, striped marlin, and swordfish (Fig. 3). Haulback mortality rates after hook swallowing for blue sharks were lower than those for external hooking, and higher than those for mouth hooking. We observed lower haulback mortality rates for shortfin mako shark, striped marlin, and swordfish from hook swallowing than from mouth hooking.

Output of MODEL 2

Haulback mortality rate was higher for large-C than for tuna hook types in bigeye tuna, but did not differ among hook types in other species (Table 7, Fig. 4). On the other hand, the mortality rate for two bait types differed in bigeye tuna and blue shark, with a rate reduction and increase with the use of fish bait in bigeye tuna and shortfin mako shark, respectively.

In Fig. 5 and Fig. 6 we show the effects of two environmental factors—sea surface temperature (SST) and soak time—on haulback mortality rate. The response to SST differed by species, with haulback mortality rate increasing with higher SST for shortfin mako sharks, swordfish, common dolphinfish, and escolar, and conversely increasing with lower SST for bigeye and longnose lancetfish. We observed little fluctuation in haulback mortality rate with SST in blue shark and striped marlin. In general, haulback mortality rate increased with increasing soak time. However, for bigeye tuna, haulback mortality rate decreased with soak time, but the trend was not clear, and for blue sharks, haulback mortality rate increased only slightly with increased soak time.

In Table 8 we show standardized CPUE by hook and bait type. Differences in CPUE among hook type were found for all species except striped marlin (Fig. 7). We observed higher CPUE for small-C in bigeye tuna, escolar, longnose lancetfish, and swordfish, and higher CPUE for both large and small-C in blue shark, shortfin mako shark, and common dolphinfish. Especially in blue sharks, the difference in CPUE ranged from 10 to 20 depending on the hook type.

We observed differences in standardized CPUE by bait type in bigeye tuna, blue shark, shortfin mako shark, common dolphinfish, and escolar. CPUE decreased with whole fish bait in bigeye tuna and blue sharks, but increased in shortfin mako shark, common dolphinfish, and escolar. In the case of blue shark, escolar, shortfin mako shark, and swordfish, the bait effect could be varied with circle hooks, suppressing the CPUE-increasing effect by fish bait in bigeye tuna, blue shark, and swordfish with the use of circle hooks, while conversely this combination boosted increase of CPUE in common dolphinfish, escolar, and shortfin mako shark.

In Table 9 we show MPUE values by hook and bait type. Compared to differences in MPUE among hook and bait type, those in CPUE were relatively small. We observed differences in MPUE by all hook types except for those in common dolphinfish and striped marlin, with higher MPUE for small-C in bigeye tuna, blue shark, escolar, longnosed lancetfish, and swordfish (Fig. 8). We confirmed decreases in MPUE by fish bait in bigeye tuna and blue shark, and blue sharks, and conversely, increases in MPUE in common dolphinfish, shortfin mako shark, and swordfish. In escolar, however, MPUE increased only when small-C was combined with whole fish bait. The effect of the combination of whole fish bait and circle hook varied in bigeye tuna, blue shark, common dolphinfish, shortfin mako shark, and swordfish. In bigeye tuna, blue shark, and swordfish, the effect of the circle hook on MPUE was suppressed by fish bait, while in common dolphinfish and shortfin mako shark, MPUE increased significantly when whole fish bait and circle hook were used together.

Hook and bait effect for sea turtle bycatch

We failed to complete our analysis for loggerhead turtles throughout the models because the bias in frequency of capture events among the experimental groups was too large and did not converge except for only a part of MODEL 1. Instead, we show the nominal CPUE, haulback mortality rate, and MPUE for each experimental group in Table 10, and hooking location and haulback mortality rate by each hook type with squid bait in Fig. 9. Most individuals survived regardless of hooking location. When whole fish bait was used, haulback mortality rate and associated MPUE were zero. For squid bait, large-C had the smallest CPUE, mortality, and MPUE.

DISCUSSION

Our experimental comparisons showed that the hook shape and bait type—both considered as effective bycatch mitigation measures for sea turtles—have extremely multifaceted effects for teleost fishes and sharks and, in some species, the direction of the effects was conflicted. The results provide significant insight into two aspects of the management of vulnerable bycatch species in tuna fisheries: how the confrontational effect of bycatch mitigation measures should be managed, and in which processes of fishing mortality intervention in the management of vulnerable species should occur. As a specific concern regarding the former, when considering shortfin mako shark, which are experiencing significant stock depletion in the North Atlantic (Sims et al. 2018; ICCAT 2019), the implementation of bycatch mitigation measures for sea turtles, which are also required to reduce fishery bycatch, could conversely increase fishing mortality and become a conservation risk. Previously, most of bycatch mitigation measure assessments have focused on whether they reduce the impact of vulnerable bycatch species of concern being bycaught, with the secondary impact being from an economic perspective—in other words, whether the catch rate of commercial species is reduced or not. Here we show for the first time that widely adopted bycatch mitigation measures may have a negative impact on some endangered species. Where both loggerhead turtle and shortfin mako with opposing effects are at low abundance, management measures should be based on a thorough discussion identifying the optimal combination of mitigation measures, accompanied by scientific evidence. With regard to the latter issue, the mortality reduction expected from circle hooks is not very promising, especially for species with high catch mortality, since the main effect of circle hooks is to minimize internal organ damage, which is of little use for species that have died from other causes, such as heat stress or suffocation. For such species, consideration of methods to reduce the catch itself rather than the mortality rate will be of greater benefit.

Analysis of MODEL 1 revealed that use of large circle hooks increased mouth hooking at capture events of loggerhead turtle. This result is consistent with existing studies, which show that the circle hook prevents internal organ damage and improves the probability of live release (Cooke & Suski 2004) while the impact of circle hooks on haulback mortality in this study could not be evaluated due to skewed data about mortality events (Fig. 9). For the same reason, the MODEL 2 analysis could not evaluate the effects of circle hook and bait type on CPUE, mortality rate, and MPUE of loggerhead turtle. However, since mortality event of loggerhead turtle did not occur at all when fish bait was used, it may be assumed that there is an effect of mortality reduction by using fish bait. This result is also consistent with existing studies, and is related to the lower attractant effect of whole fish bait on marine turtles and the increased probability of swallowing caused by the difficulty to bite off the bait (Stokes et al. 2011; Parga et al. 2015).

The results of MODEL 1 indicate that circle hook use, rather than tuna hook use clearly reduced hook swallowing for blue sharks, shortfin mako sharks, and swordfish, and in these three species, haulback mortality rate after mouth hooking was also lower than that after hook swallowing. Hook swallowing has been reported to increase the likelihood of fatal damage to internal organs. Although previous studies have reported that studies to attach satellite tags to white marlin *Tetrapturus albidus* caught by recreational fishing using circle hooks and subsequently released have reduced the post-release mortality rate (Horodysky & Graves 2005), unfortunately, the present study did not corroborate this information. Our results showed different effects of hooking location by hook types, with more frequent mouth hooking by small circle hooks in many species. Few studies have examined the relationship among size of circle hook, hooking location and haulback mortality of non-turtle species. However, two studies discussed about the possibility that relative differences in mouth and hook size, and differences in feeding behavior toward prey (swallowing the prey whole or biting it off) may affect the hooking location (Epperly et al. 2012; Gilman et al. 2020).

The results of MODEL 2, which examined the effects of hooks and bait, showed increases in MPUE due to the use of small-C hooks of all species except striped marlin. This result suggests that the increased CPUE by small-C hook contributed more to the increase in the MPUE than haulback mortality rate. Interestingly, we did not observe an increased MPUE with large-C hooks. There have been many previous findings on the effects of circle hook use on CPUE, with elevated CPUE for tunas and no consistent trend for sharks and other teleosts. However, although few previous studies have focused on hook size and made comparisons, catch rates for skipjack, shortbill spearfish, escolar, and lancetfish are reported to have decreased when larger hooks were used (Curran & Beverly 2012). Considering the effect of hook size in terms of the catch process, it is unlikely that catch rates increased due to swallowing, as the results of MODEL 1 indicate an increase in mouth hooking for many species. The results of MODEL 2 showed that the effects of the bait type varied by species, unlike the pattern for hook types, with the use of whole fish bait decreasing CPUE and MPUE for blue sharks, while we observed the exact opposite effect for shortfin mako sharks, in that fish bait increased CPUE and MPUE—and even haulback mortality rate. Whole fish bait also reduced MPUE in bigeye tuna, while in common dolphinfish and escolar, it tended to increase MPUE and CPUE. We observed few effects of bait type in billfishes. Several studies have examined the effects of fish bait without circle hook and have reported reduced catch rates for tropical tunas, blue sharks, and escolar and increased catch rates for shortfin mako, porbeagle shark, and white marlin (Fernandez-Carvalho et al. 2015; Foster et al. 2012; Watson et al. 2005; Yokota et al. 2009). In swordfish, some previous studies evaluating the effect of switching to whole fish bait from squid bait have reported conflicting effects (increase: Santos et al. 2012; Foster et al. 2012; decrease: Fernandez-Carvalho et al. 2015). Very few studies have assessed haulback mortality by bait types other than those on sea turtles, but in blue sharks, the use of fish bait reduced haulback mortality (Epperly et al. 2012). The catch rates of the target species, as previously noted, were affected by bait texture, but a mechanistic explanation for mortality effects is lacking. Since the likelihood that differences in feeding behavior among species have an effect is high, this issue could be resolved through comparative studies based on observations of feeding behavior, as in the case of circle hooks.

When we considered the effects of hook and bait type simultaneously, bait types were shown to have a relatively greater impact than hook types. Results for MPUE in Model 2 (Table 8, Fig. 8) show that in shortfin mako sharks and dolphinfish, the mortality-increasing effect of fish bait was further accumulated by that of the circle hook, while in blue sharks the mortality-increasing effect of the circle hook was counteracted by the bait effect. Although the combination of circle hooks and fish bait is considered to be very useful for avoiding sea turtle bycatch, this combination may pose a high mortality risk for endangered species like shortfin mako, and even in the case of target fishes like bigeye tuna, may counteract the expected positive effect of the circle hook on catch rate. In the case of shortfin mako, for example, changing from tuna hooks to small-C hooks increases MPUE by about 1.8 times, and changing the bait from squid to fish increases MPUE by over 7 times

(Table 9), and in the case of bigeye tuna, the CPUE estimate, which increased by 2.3 times with small-C hooks, returned to the same level as tuna hooks by changing from squid to fish bait (Table 8). Such substantial changes in CPUE and MPUE would not be ignored when managing fisheries for those species. Although very limited studies have simultaneously examined the interrelationship between hook and bait types, all studies support the conclusion that the combination of hook type and bait type causes fluctuations in catch rates and that the direction of response varies among species (Coelho et al. 2012; Foster et al. 2012; Fernandez-Carvalho et al. 2015).

Water temperature and soak time emerged as significant factors affecting haulback mortality rate in MODEL 2, which had been reported in sharks (Carruthers et al. 2009; Gallasher et al. 2014) and sea turtles (Watson et al. 2005). This indicates the these environmental factors need to be controlled statistically or experimentally when assessing the effects of hook and bait type on mortality rate. The effect of water temperature—particularly during the depth and time of day when hooked—and changes in water temperature up to the time the fish is landed, are considered to be influential. In addition, in high water temperature environments, studies have identified an increased risk of suffocation due to decreased dissolved oxygen in water and increased physiological metabolic rate (Skomal & Bernal 2010). Gallagher et al. (2014) reported an increase in haulback mortality rate for four shark species when caught during high water temperatures. In addition, for the species that adopt rum ventilation, prolonged soak time inevitably increases the risk of suffocation due to the restriction of swimming behavior by being hooked. Mortality rates of tuna, swordfish, and sharks reportedly increased with increasing soak time (Epperly et al. 2012; Gallagher et al. 2014).

Here, we quantified our data through experimental operations that standardized the various conditions, but not all aspects were completely controlled. For example, while previous studies on hook size have examined the correspondence with actual measurements (Gilman et al. 2016), several shapes of circle hook were used in the experiment in this study, precluding examination of effects of individual hook types due to sample size issues. We were also unable to examine hooking location of the catch when fish bait was used. These omissions, while having a limited impact on the present conclusions, are probably variables that should be considered for a deeper examination of the effects of terminal gear on catch and bycatch. In this experiment, wire leaders were used on all branchlines to minimize the effects of sharks' bite-off. While some studies have described concerns that wire leaders may increase catch rates, especially for rare sharks, they are considered essential for at least experimentally verifying accurate catch and mortality rates for shark species. We know from this and previous studies that haulback mortality rates for sharks are much lower than those for teleosts (Reinhardt et al. 2018), and the implementation of appropriate releases, even with wire leaders, allow for the reduction of risk for vulnerable shark species.

Here, based on a Bayesian approach, we succeeded in presenting a quantitative impact assessment of terminal gear on teleosts, sharks, and sea turtles by directly calculating the expected values for mortality rate, CPUE, and MPUE with each terminal gear. Calculating MPUE using this model can be a very useful tool because it provides a more direct estimate than does CPUE or mortality rate alone of catch/bycatch risk to populations of those species. Although we did not include post-release mortality rate in the model due to lack of data, it would be possible to estimate overall fishing mortality in the model by designing additional experiments so that mark–recapture is conducted at the same time. Even if it is not possible to use wire leaders for the proportion of “cryptic catch” due to bite-off, it is possible to extrapolate this proportion into the model to make predictions regarding mortality—a development we anticipate.

In this study, we successfully conducted a quantitative impact assessment of bycatch integrating catch rate and mortality using a Bayesian approach that directly predicts mortality and CPUE. Although the data used in the analysis relied solely on the results of an Asian-style longline experiment in the Pacific Ocean and may therefore contain inherent biases, the same analysis method can be used in conjunction with data from other experiments conducted in other areas and fishing styles to provide a more integrated assessment.

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Table 1 Fishing effort (longline hooks) in experimental operations used in the analysis.

hook type	bait type		total effort
	squid	fish	
tuna	97,834	71,146	168,980
small-C	70,882	7,658	78,540
large-C	31,872	6,976	38,848
total effort	200,588	85,780	286,368

Table 2 List of hook types used in the experiment

hook name	size	category	straight total length(mm)	maximum total width(mm)
Komatu Keisaku tunahook	4.0 sun	tuna	63	63
Doitomi Tunamutsu	4.0 sun	small-C (circle hook)	58	69
Komatsu Keisaku modified	4.0 sun	small-C (circle hook)	60	68
Komatsu Keisaku modified	4.5 sun	small-C (circle hook)	63	71
Komatsu Keisaku modified	4.8 sun	small-C (circle hook)	N/A	N/A
Komatsu Keisaku type Etsuna	4.5 sun	small-C (circle hook)	62	76
Komatsu Keisaku type North America	4.3 sun	small-C (circle hook)	57	63
Tankichi Uruwa	3.8 sun	small-C (circle hook)	56	64
Komatsu Keisaku modified	5.2 sun	large-C (circle hook)	74	81
Komatsu Keisaku type North America	5.2 sun	large-C (circle hook)	76	85
Pacific Fishing Tackle circle hook	18/0	large-C (circle hook)	68	80

Table 3 Number of fish caught by hook and bait type in the experimental operation

species	total catch	hook type			bait type	
		tuna	small-C	large-C	squid	fish
blue shark	13,018	8,084	3,562	1,372	8,903	4,115
longnose lancetfish	1,297	945	257	95	692	605
common dolphinfish	505	206	181	118	363	142
shortfin mako shark	485	298	134	53	262	223
swordfish	288	129	112	47	249	39
bigeye tuna	269	114	146	9	201	68
loggerhead turtle	268	128	113	27	259	9
escolar	163	79	56	28	108	55
striped marlin	145	70	53	22	126	19
other species	1,578	-	-	-	-	-

Table 4 Composition of hooking location and fate at hauling.

species	hooking location				fate at haulback		
	swallowed	mouth	external	unknown (*)	alive	dead	unknown (*)
blue shark	2,270	2,608	66	8,074	11,701	1,066	251
longnose lancetfish	25	288	16	968	165	1,010	122
common dolphinfish	39	159	4	303	398	87	20
shortfin mako shark	66	68	18	333	363	118	4
swordfish	59	118	18	93	55	226	7
bigeye tuna	7	128	2	132	83	183	3
loggerhead turtle	109	116	14	29	256	5	7
escolar	1	62	1	99	108	40	15
striped marlin	12	80	8	45	68	76	1

(*) Includes catches that dropped off before reseachers checked or lack of survey.

Table 5 Estimates of the posterior distribution of the proportion of hooking location by hook when squid bait is used (median). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

species	hook type		
	tuna	small-C	large-C
hooking place: external			
blue shark	0.020 (0.015 - 0.027)	0.008 (0.005 - 0.012)	0.005 (0.001 - 0.014)
loggerhead turtle	0.025 (0.006 - 0.066)	0.079 (0.037 - 0.143)	0.031 (0.001 - 0.154)
longnose lancetfish	0.049 (0.020 - 0.092)	0.031 (0.011 - 0.067)	0.074 (0.018 - 0.185)
shortfin mako	0.102 (0.041 - 0.199)	0.107 (0.052 - 0.185)	0.136 (0.021 - 0.381)
striped marlin	0.083 (0.026 - 0.186)	0.068 (0.016 - 0.173)	0.048 (0.002 - 0.235)
swordfish	0.075 (0.030 - 0.148)	0.109 (0.057 - 0.184)	0.031 (0.001 - 0.154)
hooking place: mouth			
blue shark	0.497 (0.475 - 0.518)	0.548 (0.528 - 0.568)	0.560 (0.517 - 0.603)
loggerhead turtle	0.439 (0.346 - 0.534)	0.480 (0.381 - 0.577)	0.746 (0.544 - 0.892)
longnose lancetfish	0.857 (0.792 - 0.908)	0.903 (0.848 - 0.943)	0.872 (0.740 - 0.953)
shortfin mako	0.356 (0.238 - 0.486)	0.524 (0.418 - 0.630)	0.379 (0.154 - 0.650)
striped marlin	0.759 (0.623 - 0.869)	0.831 (0.696 - 0.924)	0.883 (0.660 - 0.982)
swordfish	0.528 (0.416 - 0.637)	0.676 (0.574 - 0.767)	0.611 (0.405 - 0.790)
hooking place: swallowed			
blue shark	0.483 (0.461 - 0.505)	0.444 (0.424 - 0.464)	0.434 (0.391 - 0.478)
loggerhead turtle	0.533 (0.438 - 0.626)	0.438 (0.343 - 0.537)	0.210 (0.078 - 0.405)
longnose lancetfish	0.092 (0.052 - 0.148)	0.064 (0.032 - 0.111)	0.046 (0.007 - 0.145)
shortfin mako	0.536 (0.405 - 0.665)	0.365 (0.266 - 0.472)	0.460 (0.213 - 0.724)
striped marlin	0.150 (0.066 - 0.273)	0.093 (0.030 - 0.207)	0.048 (0.002 - 0.234)
swordfish	0.394 (0.289 - 0.505)	0.211 (0.136 - 0.303)	0.343 (0.174 - 0.550)

Table 6 Estimated haulback mortality rates (median of posterior distribution) by hooking location when squid bait is used. Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses. Loggerhead turtles were excluded because there were no mortalities and the calculation had not been converged.

species	external	mouth	swallowed
blue shark	0.295 (0.185 - 0.411)	0.052 (0.044 - 0.061)	0.157 (0.141 - 0.172)
shortfin mako shark	0.496 (0.262 - 0.729)	0.134 (0.061 - 0.219)	0.279 (0.176 - 0.394)
longnose lancetfish	0.956 (0.810 - 1.000)	0.860 (0.816 - 0.899)	0.798 (0.621 - 0.934)
swordfish	0.802 (0.595 - 0.959)	0.776 (0.694 - 0.853)	0.919 (0.842 - 0.979)
striped marlin	0.798 (0.488 - 0.990)	0.431 (0.315 - 0.547)	0.968 (0.839 - 1.000)

Table 7 Haulback mortality rate by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

species	hook type		
	tuna	small-C	large-C
bait type: squid			
bigeye tuna	0.734 (0.638 - 0.823)	0.757 (0.679 - 0.830)	0.960 (0.812 - 1.000)
blue shark	0.084 (0.076 - 0.092)	0.081 (0.072 - 0.091)	0.077 (0.062 - 0.093)
common dolphinfish	0.154 (0.097 - 0.214)	0.128 (0.081 - 0.180)	0.180 (0.107 - 0.262)
escolar	0.189 (0.085 - 0.311)	0.300 (0.163 - 0.451)	0.181 (0.058 - 0.336)
longnose lancetfish	0.901 (0.869 - 0.929)	0.925 (0.893 - 0.954)	0.890 (0.827 - 0.941)
shortfin mako	0.161 (0.107 - 0.223)	0.186 (0.120 - 0.253)	0.181 (0.084 - 0.300)
striped marlin	0.531 (0.401 - 0.665)	0.472 (0.334 - 0.610)	0.530 (0.321 - 0.740)
swordfish	0.829 (0.755 - 0.894)	0.783 (0.700 - 0.859)	0.832 (0.711 - 0.933)
bait type: fish			
bigeye tuna	0.473 (0.326 - 0.627)	0.503 (0.339 - 0.666)	0.886 (0.570 - 1.000)
blue shark	0.078 (0.068 - 0.088)	0.075 (0.062 - 0.088)	0.071 (0.055 - 0.088)
common dolphinfish	0.237 (0.158 - 0.326)	0.200 (0.103 - 0.312)	0.273 (0.164 - 0.398)
escolar	0.242 (0.114 - 0.389)	0.370 (0.174 - 0.586)	0.231 (0.061 - 0.472)
longnose lancetfish	0.909 (0.880 - 0.937)	0.931 (0.895 - 0.960)	0.898 (0.839 - 0.945)
shortfin mako	0.308 (0.242 - 0.387)	0.347 (0.227 - 0.479)	0.342 (0.197 - 0.488)
striped marlin	0.691 (0.457 - 0.883)	0.640 (0.368 - 0.865)	0.692 (0.384 - 0.920)
swordfish	0.853 (0.728 - 0.947)	0.811 (0.651 - 0.935)	0.856 (0.700 - 0.967)

Table 8 Standardized catch per unit effort (CPUE) by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

species	hook type		
	tuna	small-C	large-C
bait type: squid			
bigeye tuna	0.436 (0.036 - 1.231)	1.033 (0.100 - 2.938)	0.896 (0.067 - 2.980)
blue shark	44.457 (35.464 - 54.243)	66.395 (53.022 - 81.268)	56.701 (45.075 - 69.726)
common dolphinfish	0.709 (0.336 - 1.182)	1.676 (0.775 - 2.811)	1.193 (0.534 - 2.020)
escolar	0.213 (0.074 - 0.408)	0.630 (0.225 - 1.216)	0.353 (0.108 - 0.724)
longnose lancetfish	3.326 (2.742 - 3.993)	4.677 (3.757 - 5.683)	4.109 (3.074 - 5.261)
shortfin mako	1.051 (0.686 - 1.458)	1.723 (1.084 - 2.431)	1.547 (0.939 - 2.364)
striped marlin	0.320 (0.169 - 0.512)	0.369 (0.181 - 0.603)	0.356 (0.155 - 0.640)
swordfish	0.731 (0.496 - 1.004)	1.079 (0.706 - 1.504)	1.115 (0.673 - 1.661)
bait type: fish			
bigeye tuna	0.185 (0.014 - 0.550)	0.441 (0.050 - 1.323)	0.383 (0.016 - 1.342)
blue shark	37.299 (29.425 - 45.517)	55.736 (44.428 - 68.588)	47.620 (37.760 - 58.788)
common dolphinfish	1.781 (0.858 - 3.031)	4.210 (2.019 - 7.284)	3.000 (1.404 - 5.264)
escolar	0.476 (0.153 - 0.941)	1.410 (0.443 - 2.826)	0.787 (0.221 - 1.712)
longnose lancetfish	3.544 (2.849 - 4.346)	4.982 (3.904 - 6.216)	4.379 (3.172 - 5.761)
shortfin mako	2.327 (1.518 - 3.322)	3.813 (2.453 - 5.622)	3.418 (1.913 - 5.221)
striped marlin	0.371 (0.163 - 0.663)	0.428 (0.170 - 0.777)	0.411 (0.144 - 0.822)
swordfish	0.751 (0.456 - 1.116)	1.107 (0.636 - 1.672)	1.142 (0.618 - 1.848)

Table 9 Estimated MPUE by hook and bait type (median of posterior distribution). Lower and upper limits of Bayesian credible interval (95% highest density interval [HDI]) are shown in parentheses.

species	hook type		
	tuna	small-C	large-C
bait type: squid			
bigeye tuna	0.318 (0.026 - 0.902)	0.780 (0.064 - 2.214)	0.841 (0.072 - 2.828)
blue shark	3.744 (2.910 - 4.646)	5.388 (4.152 - 6.790)	4.370 (3.163 - 5.756)
common dolphinfish	0.108 (0.043 - 0.196)	0.213 (0.091 - 0.399)	0.214 (0.080 - 0.405)
escolar	0.040 (0.009 - 0.089)	0.187 (0.052 - 0.403)	0.063 (0.009 - 0.162)
longnose lancetfish	2.994 (2.443 - 3.586)	4.321 (3.440 - 5.239)	3.641 (2.706 - 4.706)
shortfin mako	0.169 (0.088 - 0.263)	0.321 (0.156 - 0.498)	0.278 (0.102 - 0.535)
striped marlin	0.169 (0.076 - 0.279)	0.173 (0.075 - 0.299)	0.185 (0.065 - 0.360)
swordfish	0.604 (0.403 - 0.838)	0.840 (0.538 - 1.186)	0.919 (0.533 - 1.385)
bait type: fish			
bigeye tuna	0.087 (0.007 - 0.267)	0.219 (0.016 - 0.675)	0.319 (0.016 - 1.158)
blue shark	2.909 (2.229 - 3.684)	4.185 (3.112 - 5.415)	3.396 (2.408 - 4.571)
common dolphinfish	0.420 (0.168 - 0.765)	0.836 (0.288 - 1.664)	0.812 (0.286 - 1.558)
escolar	0.113 (0.026 - 0.261)	0.511 (0.113 - 1.177)	0.178 (0.022 - 0.512)
longnose lancetfish	3.217 (2.582 - 3.953)	4.629 (3.668 - 5.846)	3.919 (2.800 - 5.169)
shortfin mako	0.712 (0.432 - 1.063)	1.311 (0.679 - 2.117)	1.153 (0.527 - 2.046)
striped marlin	0.250 (0.089 - 0.469)	0.264 (0.081 - 0.525)	0.273 (0.072 - 0.589)
swordfish	0.633 (0.375 - 0.959)	0.883 (0.466 - 1.364)	0.961 (0.497 - 1.590)

Table 10 Nominal CPUE, Haulback mortality rate and MPUE of Loggerhead turtle。 All figures are based on aggregated operational data, not estimates.

hook type	CPUE		haulback mortality		MPUE	
	squid	fish	squid	fish	squid	fish
tuna	1.247	0.084	0.0385	0.000	0.0480	0.000
large-C	1.034	0.277	0.0090	0.000	0.0093	0.000
small-C	2.148	0.261	0.0246	0.000	0.0528	0.000

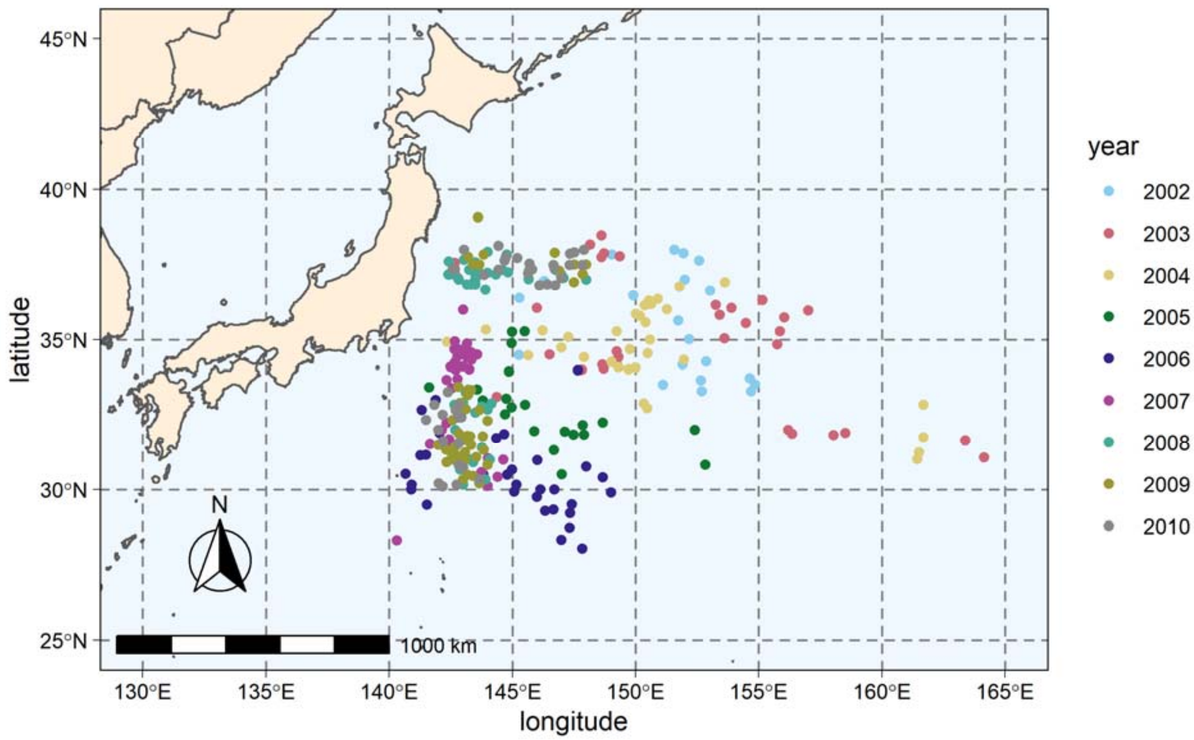


Figure 1 Locations where the longline operation experiment was conducted.

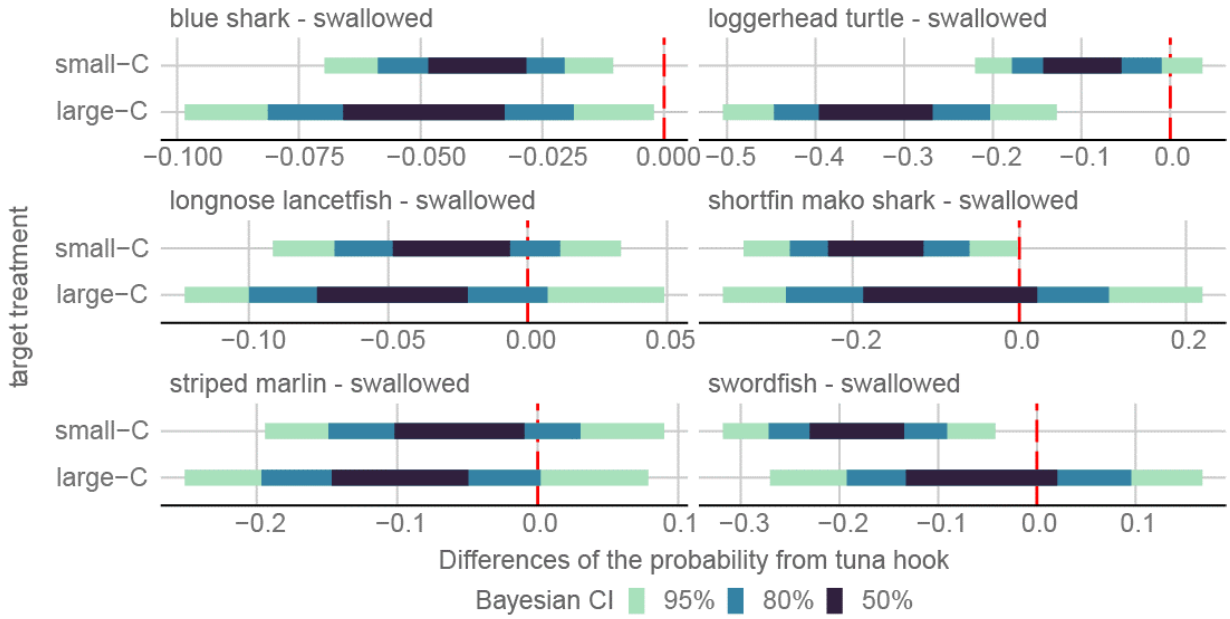


Figure 2 Differences in the estimated probability of the "swallowed" hooking location of each circle hook type from tuna hook when squid bait is used. The red dotted line indicates that the difference is zero.

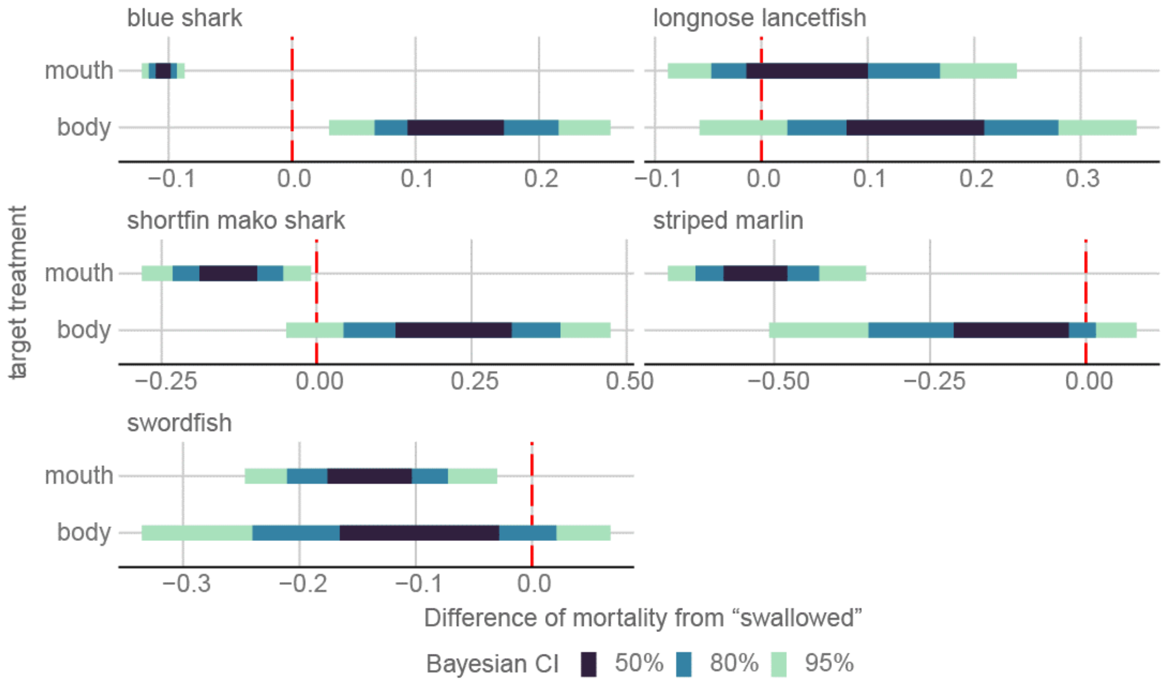


Figure 3 Differences in the estimated haulback mortality rate of each target hooking location from “swallowed” hooking location when squid bait is used. The red dotted line indicates that the difference is zero.

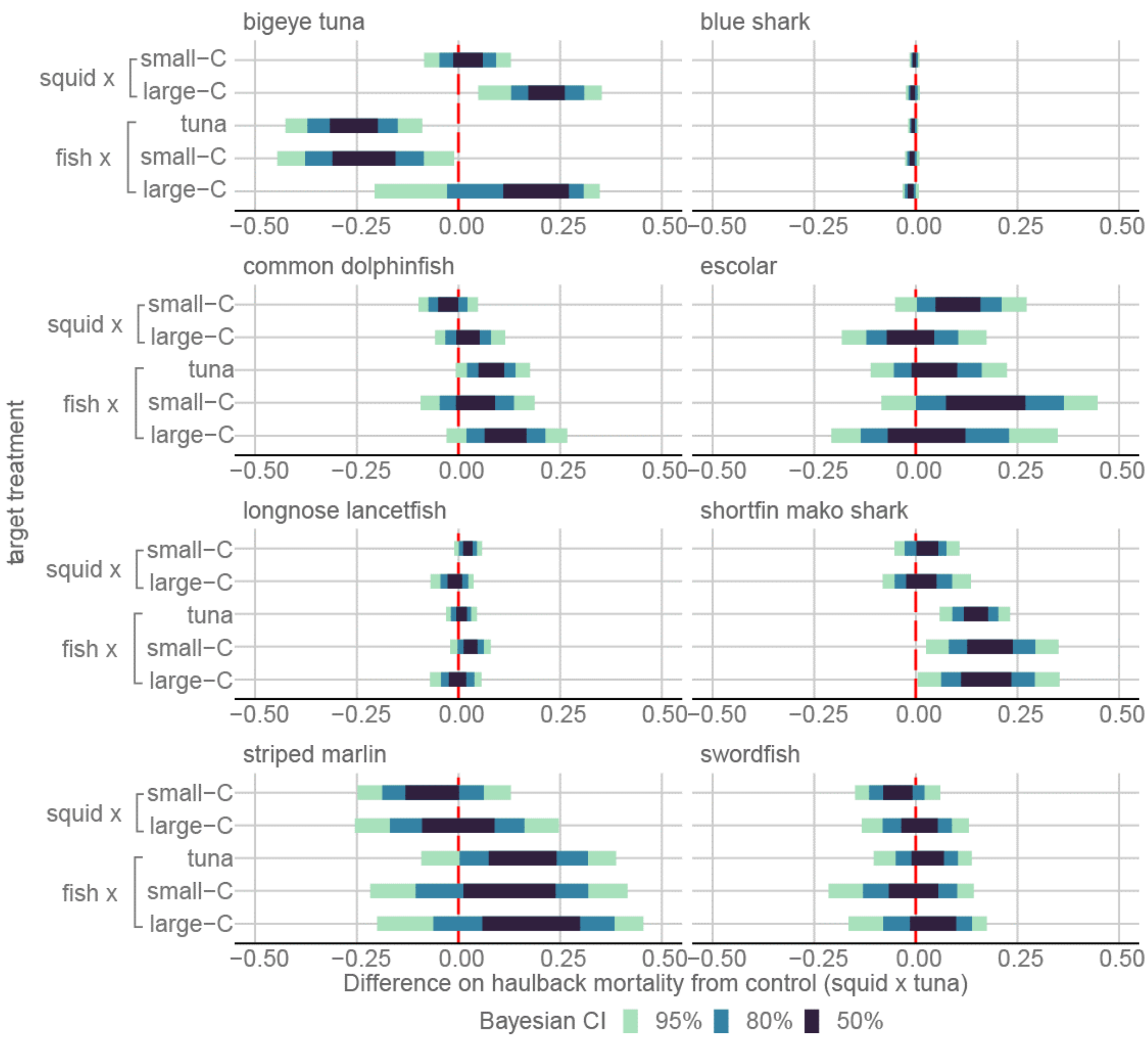


Figure 4 Differences in estimated haulback mortality between each experimental group and the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

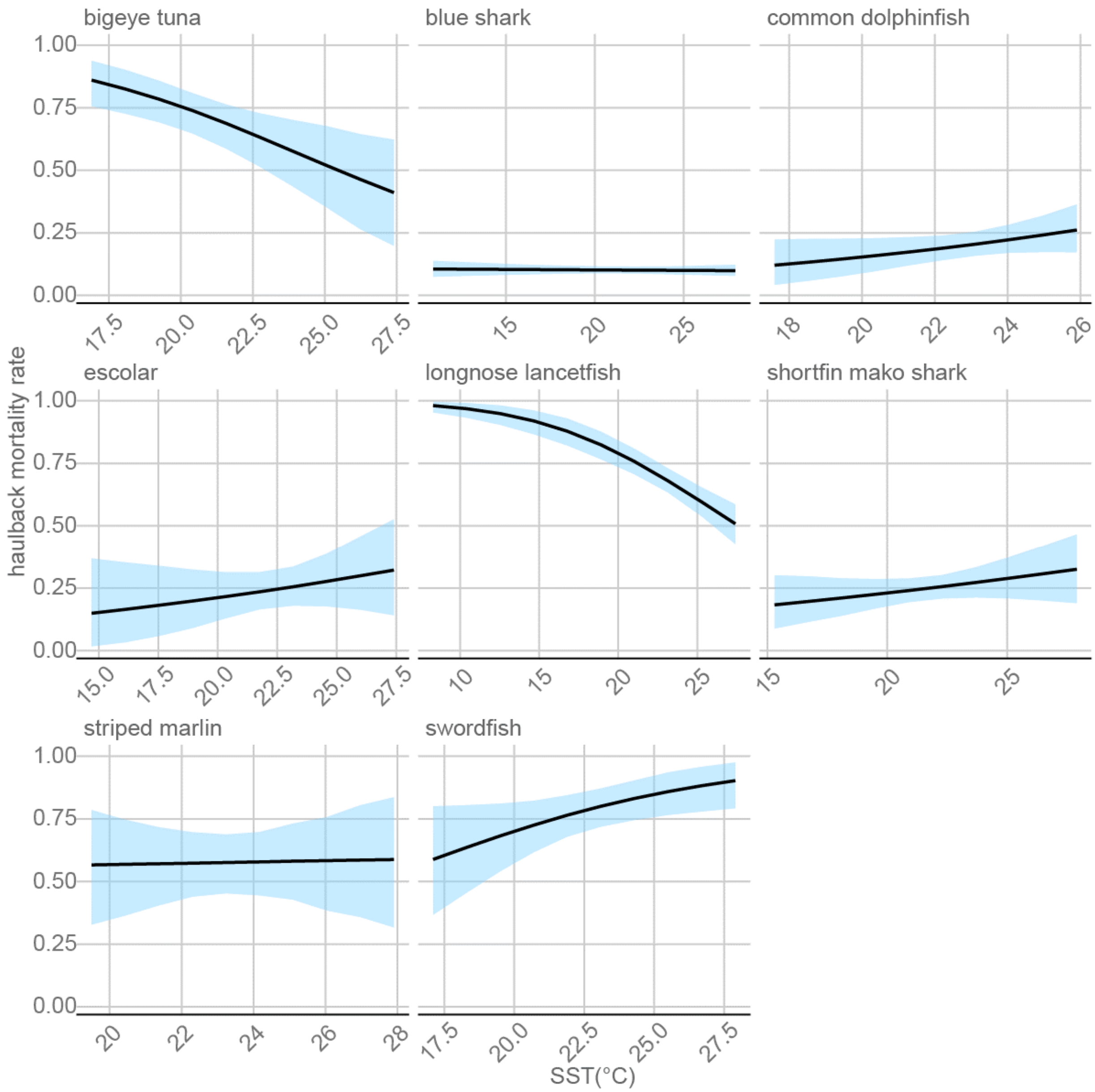


Figure 5 Relationship between sea surface temperature (SST) variability and haulback mortality rate at longline operations. Solid lines indicate median; masked areas indicate 95% Bayesian credible interval.

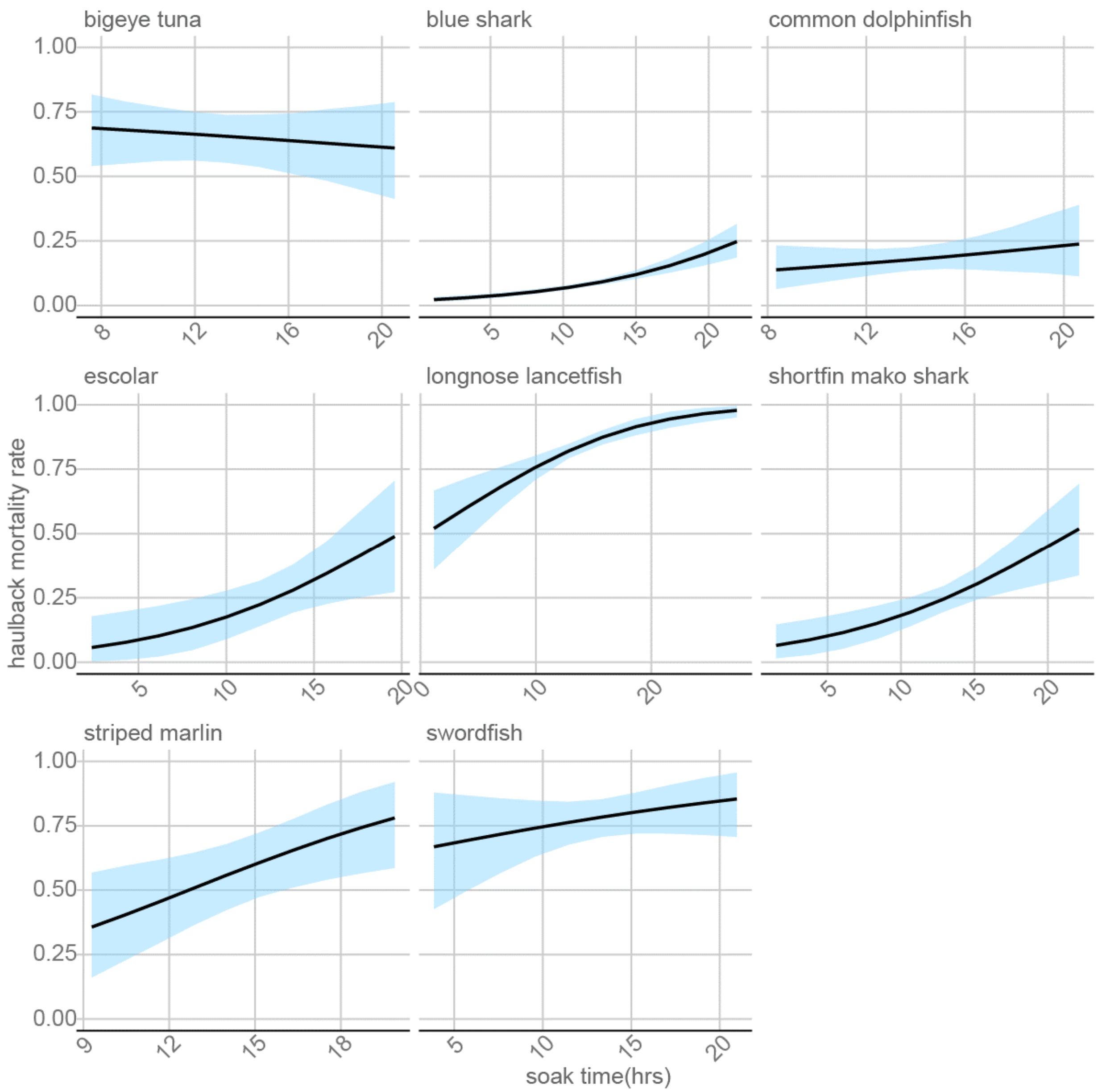


Figure 6 Relationship between soak time (time from setting the branch line to hauling) variability and haulback mortality rate at longline operations. Solid lines indicate median; masked areas indicate 95% Bayesian credible interval.

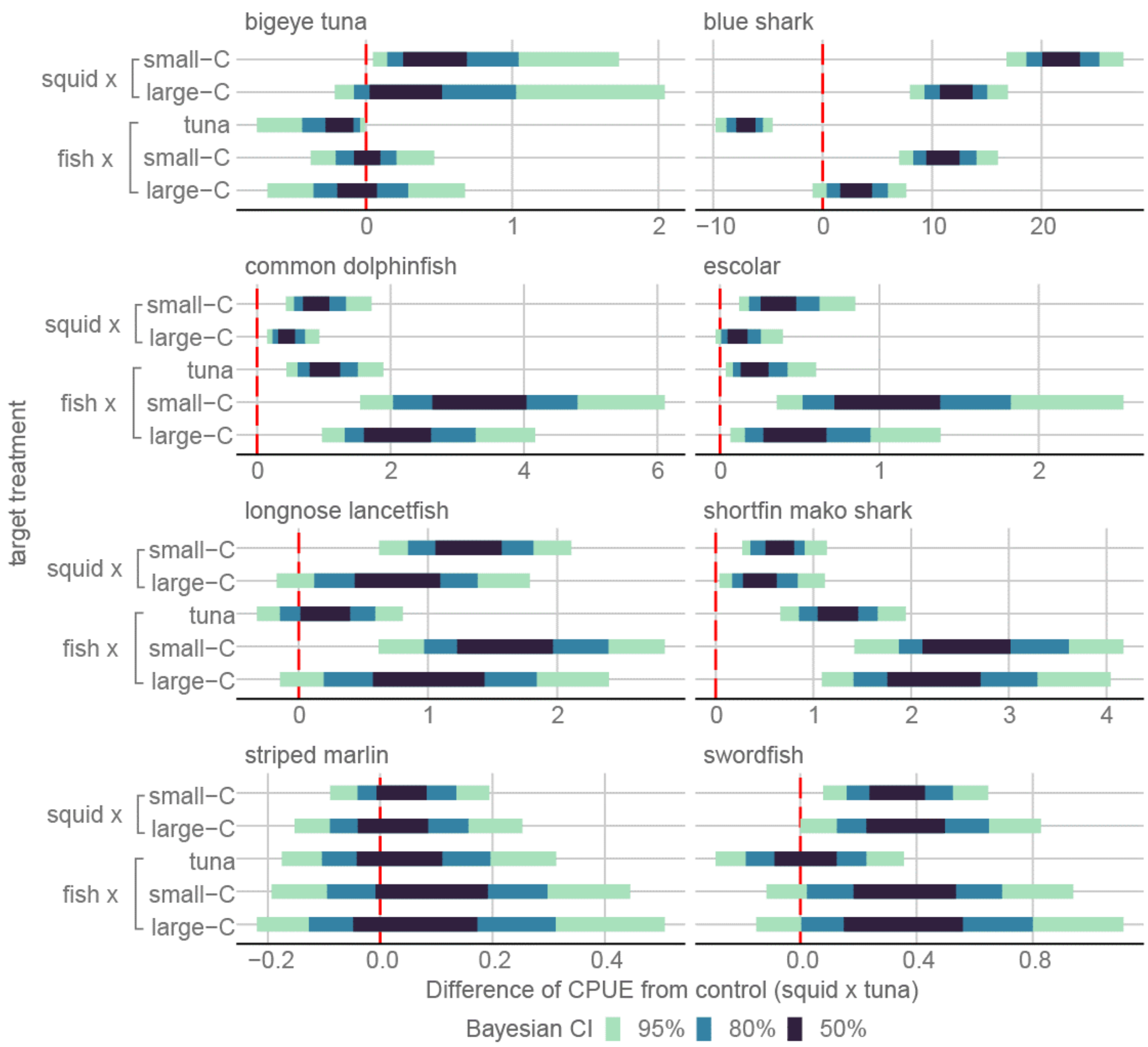


Figure 7 Differences in the standardized catch per unit effort (CPUE) for each experimental group from those for the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

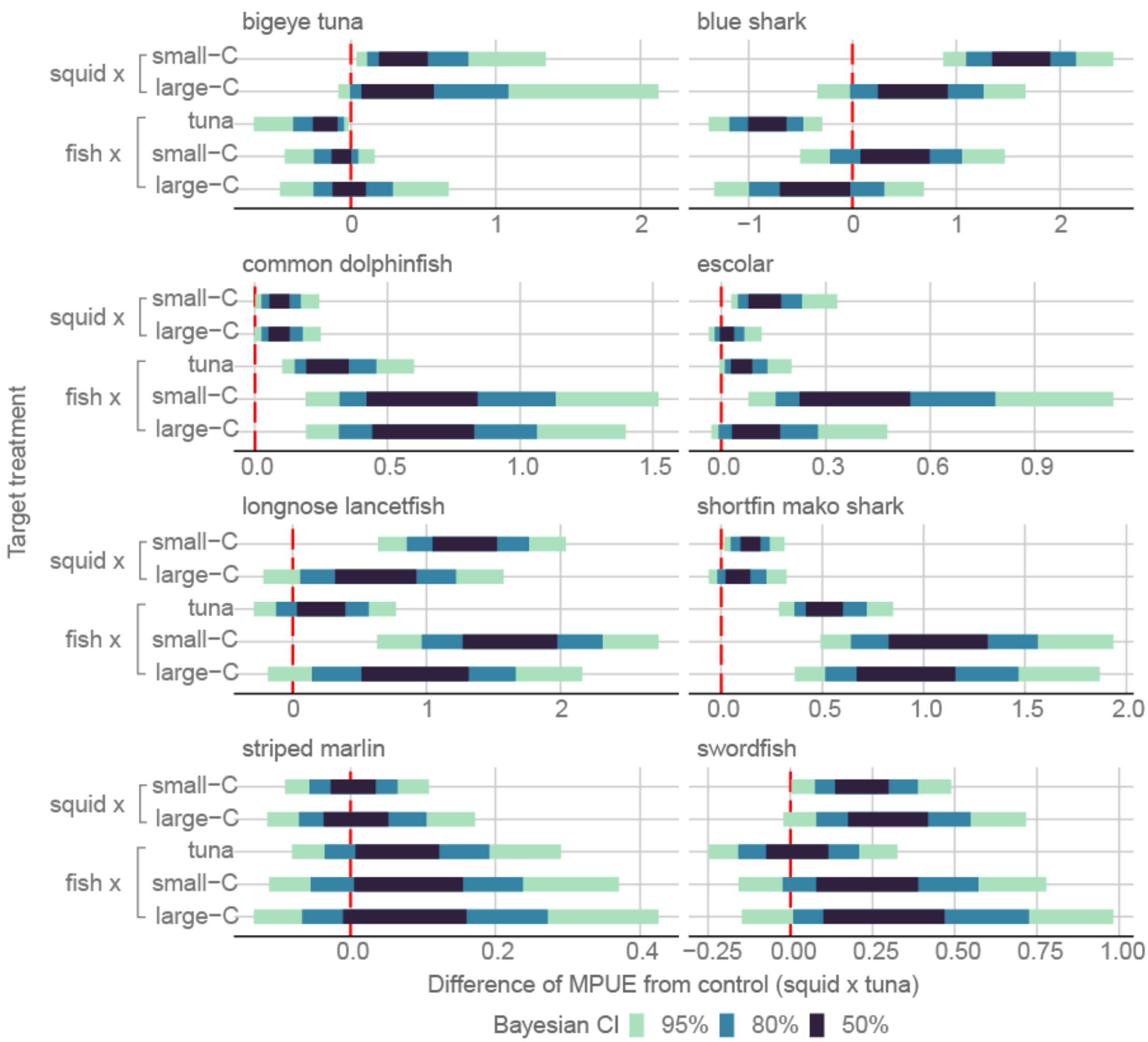


Figure 8 Differences in the estimated MPUE (mortality per unit effort) between those for each experimental group and those for the control group (squid x tuna hook). The red dotted line indicates that the difference is zero.

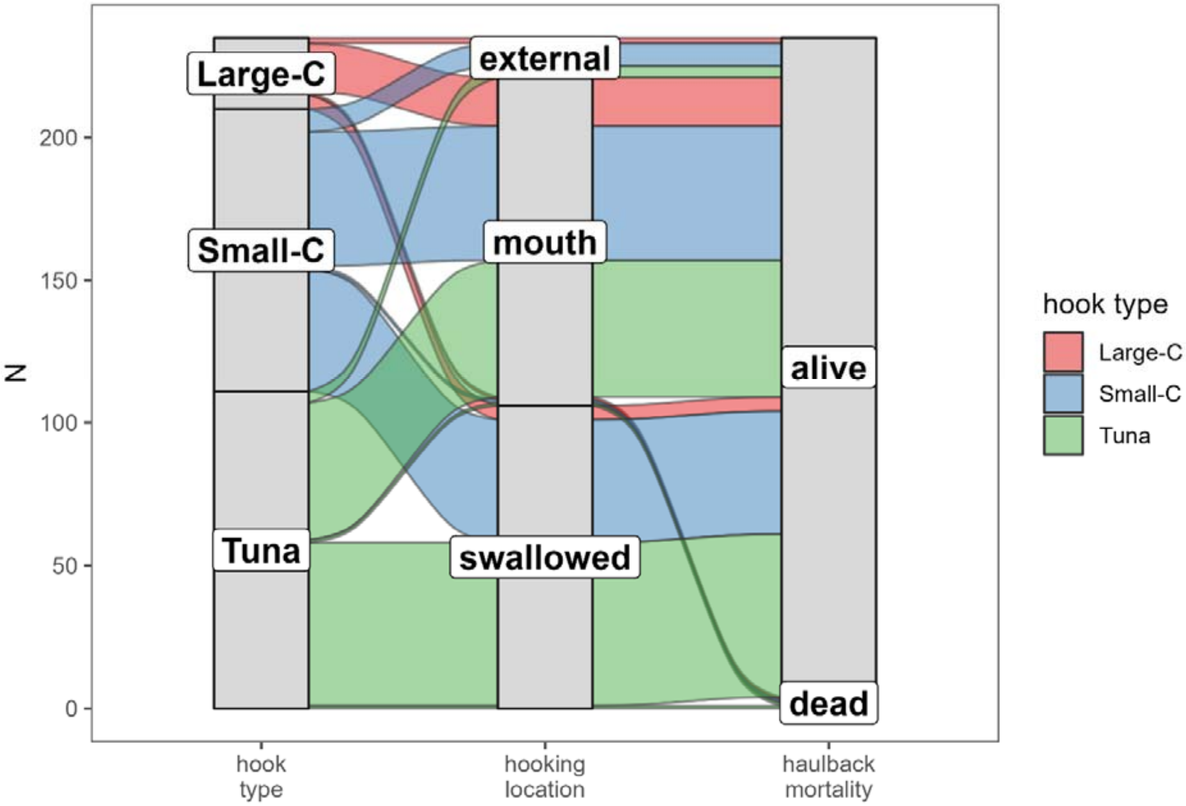
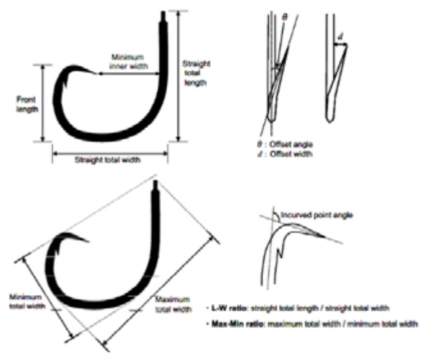


Figure 9 Alluvial plot of hooking locations and associated haulback mortality rates of loggerhead turtles by hook when squid bait is used.

Appendix 1 Detailed measurements of hooks used in the experiment and a figure explaining measurement points of hooks (copied from Yokota et al. 2006b).

manufacture	Komatsu Keisaku	Hisamatsu Tankichi	Doitomi	Komatsu Keisaku	Komatsu Keisaku	Komatsu Keisaku	Komatsu Keisaku	Komatsu Keisaku	Komatsu Keisaku	Komatsu Keisaku	Pacific Fishing Tackle MFG., CO.
hook name	tuna hook	Uruwa hook BKN	tuna circle hook SS-170	modified circle hook	modified circle hook	modified circle hook	cirlce hook type Etsuna	cirlce hook type North America	cirlce hook type North America	modified circle hook	circle hook
standardized size	4.0 sun	3.8 sun	#4	4.0 sun	4.5 sun	4.8 sun	4.5 sun	4.3 sun	5.2 sun	5.2 sun	18/0
material	stainless steel	hard steel	stainless steel	stainless steel	stainless steel	stainless steel	stainless steel	hard steel	hard steel	stainless steel	stainless steel
hook eye	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring	yes, with ring
shank thickness(mm)	5.3	4.0	4.1	5.3	5.2	N/A	4.9	5.3	5.7	5.3	5.1
straight total length(mm)	63	56	58	60	63	N/A	62	57	74	76	68
straight total width(mm)	38	44	49	47	49	N/A	56	45	56	54	59
minimum total width(mm)	38	36	39	41	45	N/A	51	41	52	48	51
maximum total width(mm)	63	64	69	68	71	N/A	76	63	81	85	80
front length(mm)	41	33	38	35	44	N/A	47	39	49	47	45
minimum inner width(mm)	27	20	15	24	25	N/A	26	20	27	26	27
L-W ratio	1.7	1.3	1.2	1.3	1.3	N/A	1.1	1.3	1.3	1.4	1.2
max-min ratio	1.7	1.8	1.8	1.7	1.6	N/A	1.5	1.5	1.6	1.8	1.6
incurved point angle		80	70	90	70	N/A	70	65	75	80	80
offset angle	5°≦θ<10°	θ≈ 0°	θ≈ 0°	θ≈ 10°	5°≦θ<10°	θ<10°	θ≈ 10°	5°≦θ<10°	10°≦θ<15°	θ≈ 5°	10°≦θ<15°
offset width(mm)	0.9	≈ 0	≈ 0	1.8	2.6	N/A	2.1	1.5	3.7	1.0	5.4
weight (g)	19.9	12.2	15.0	19.7	21.6	N/A	21.4	19.4	30.3	25.5	23.2



Appendix 2 Stan code used for MCMC sampling of (a) Model 1 and (b) Model 2

(a) MODEL 1

```
data {
 int HK_cat;
 int<lower=1> N1;
 Int<lower=1> LOC_cat;
 int<lower=1, upper=LOC_cat> LOC[N1];
 int<lower=1, upper=HK_cat> HK[N1];
 int<lower=0, upper=1> M[N1];
 int L;
 matrix[N1, L] X1;
 Int N2;
 Int  Catch[N2];
 Int L2;
 matrix[N2, L2] X2;
 vector[N2] O;
 Int YRID;
 int YR[N2];
}

transformed data{
 vector[HK_cat] Zeros;
 Zeros = rep_vector(0, HK_cat);
}

parameters {
 matrix[HK_cat, LOC_cat - 1] theta_raw;
 vector[L] beta;
 vector[L2] beta2;
 vector[YRID] r;
 real<lower= 0> sigma;

}

transformed parameters {
 vector[N2] mu;
 vector[N1] phi;
 matrix[HK_cat, LOC_cat] theta;
 phi = inv_logit(X1 * beta);
 theta = append_col(Zeros, theta_raw);
 for(i in 1:N2)
 mu[i] = X2[i,] * beta2 + r[YR[i]];

model {
 sigma ~ student_t(4, 0, 2.5);
 r ~ normal(0, sigma);
 for(n in 1:N1){
 M[n] ~ bernoulli(phi[n]);
 target += categorical_lpmf(LOC[n]|softmax(theta[HK[n],]'));
 }
 Catch ~ poisson(exp(mu + log(O)));
}
```

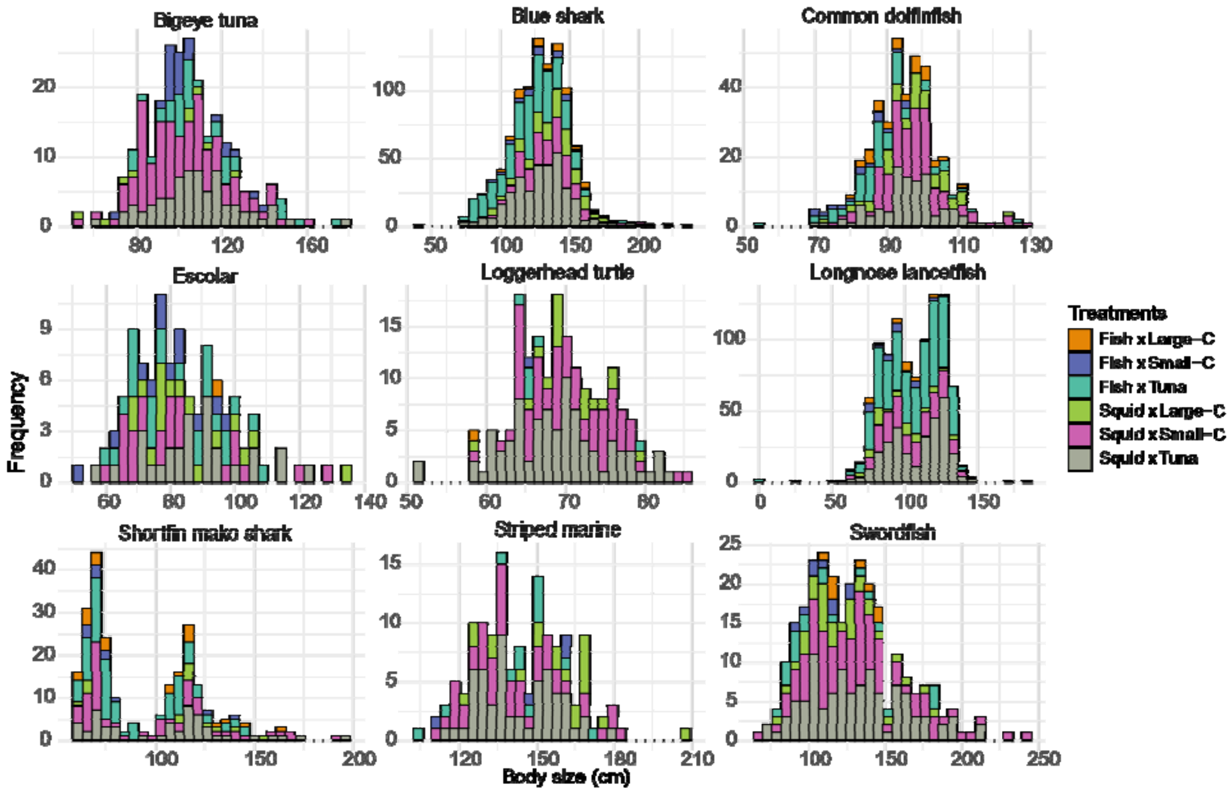
(b) MODEL 2

```
data {
 int HK_BAIT_comb;
 int<lower=1> N1;
 int<lower=0, upper=1> M[N1];
 int L;
 matrix[N1, L] X1;
 int N2;
 int  Catch[N2];
 int L2;
 matrix[N2, L2[ X2;
 vector[N2] O;
 int YRID;
 int YR[N2];
}

parameters {
 vector[L] beta;
 vector[L2] beta2;
 vector[YRID] r;
 real<lower= 0> sigma;
}

transformed parameters {
 vector[N2] mu;
 vector[N1] phi;
 phi = inv_logit(X1 * beta);
 for(i in 1:N2)
 mu[i] = X2[i,] * beta2 + r[YR[i]];
}

model {
 sigma ~ student_t(4, 0, 2.5);
 r ~ normal(0, sigma);
 for(n in 1:N1){
 M[n] ~ bernoulli(phi[n]);
 }
 Catch ~ poisson(exp(mu + log(O)));
}
```



Appendix 3 Size distributions by species for the major species captured in the study, with body length as an index of precaudal length for blue sharks and shortfin mako sharks, straight-line carapace length for loggerhead turtles, eye-to-fork length for striped marlin and swordfish, and fork length for all other species.