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Genomic health is dependent on long-term population demographic history

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Abstract

Current genetic methods of population assessment in conservation biology have been challenged by genome-scale analyses due to their quantitatively novel insights. These analyses include assessments of runs-of-homozygosity (ROH), genomic evolutionary rate profiling (GERP), and mutational load. Here, we aim to elucidate the relationships between these measures using three divergent ungulates: the white-tailed deer, caribou, and mountain goat. The white-tailed deer is currently expanding, while caribou are in the midst of a significant decline. Mountain goats remain stable, having suffered a large historical bottleneck. We assessed genome-wide signatures of inbreeding using the inbreeding coefficient F and %ROH (F_{ROH}) and identified evolutionarily constrained regions with GERP. Mutational load was estimated by identifying mutations in highly constrained elements (CEs) and sorting intolerant from tolerant (SIFT) mutations. Our results show that F and F_{ROH} are higher in mountain goats than in caribou and white-tailed deer. Given the extended bottleneck and low N_e of the mountain goat, this supports the idea that the genome-wide effects of demographic change take time to accrue. Similarly, we found that mountain goats possess more highly constrained CEs and the lowest dN/dS values, both of which are indicative of greater purifying selection; this is also reflected by fewer mutations in CEs and deleterious mutations identified by SIFT. In contrast, white-tailed deer presented the highest mutational load with both metrics, in addition to dN/dS, while caribou were intermediate. Our results demonstrate that extended bottlenecks may lead to reduced diversity and increased F_{ROH} in ungulates, but not necessarily the accumulation of deleterious alleles, likely due to the purging of deleterious alleles in small populations.

Introduction

Organizations like the International Union for Conservation of Nature (IUCN) rely primarily on population sizes and trends when assigning wildlife conservation statuses (IUCN Standards and Petitions Committee, 2019). The predictive relationship between population size (N_C) and genetic diversity is challenging, as N_C, for example, tends to be substantially more variable than diversity metrics like nucleotide diversity (π), referred to as Lewontin's paradox (Ellegren & Galtier, 2016; Lewontin, 1974), and therefore rarely accounts for recent demographic history and the genetic health of populations. While genetic diversity can correlate with IUCN status (Peart et al., 2020; Petit-Marty et al., 2021), the relationship (i.e., direction) is varied and predictive association appears low (Schmidt et al., 2022). The unprecedented loss of global biodiversity (Pascual et al., 2021) and questionable genetic diversity associations (Schmidt et al., 2022), would strongly suggest that traditional conservation genetic markers and indices must be replaced with more universally informative metrics, if they in fact exist.

Both novel genome metrics and functional versus presumed neutral variation have garnered recent attention in the conservation genetics community. The use neutral genetic diversity in conservation biology has a long history (e.g. DeWoody et al., 2021; Kohn et al., 2006), with for example specific targets of management focused on retaining 90% of initial [neutral] genetic diversity (Ballou & Foose, 2010). With the advent of whole-genome sequencing, the notion of targeted gene flow (Kelly & Phillips, 2019) and gene-editing (Phelps et al., 2020) for the purposes of conservation emerged. Teixera & Huber (2021) further suggested the focus be shifted to functional variation and gene-environment interactions, in part due to the poor associations to IUCN status noted above. Gene-targeted approaches, however, come with a suite of potential undesirable outcomes (Kardos & Shafer 2018), and genome-wide diversity still

appear to be the most indicative of population viability (Kardos et al. 2021). Likewise, Peart et al. (2020) showed that genome-wide estimates of Tajima's D, a composite statistic derived from Watterson's θ and π (Tajima, 1989), appears predictive of conservation status. Importantly, the uncertainty around which genetic marker to use reflects both a shifting understanding and nuanced patterns of genomic diversity, and further reinforces the need for identifying less ambiguous conservation-informative genetic metrics.

The general correlation between genome-wide diversity and microsatellite heterozygosity at the population-level suggests that re-sequencing data reflect smaller targeted panels of loci (Väli et al., 2008). This observation is ultimately what formed the premise of Shafer et al. (2015) that argued for the use of quantitatively novel genomic metrics for conservation. Runs-of-homozygosity (ROH), genomic evolutionary rate profiling (GERP), and mutational load are metrics dependent on genome-scale data and have only recently been applied to species and populations of concern (Grossen et al., 2020; van der Valk et al., 2019; von Seth et al., 2021). ROHs are long stretches of homozygosity that arise as a result of inbreeding (Ceballos et al., 2018), and reflect the true inbreeding coefficient and can identify genes underlying inbreeding depression (Kardos et al., 2016). GERP profiling can be used to identify evolutionarily constrained genomic elements, defined as regions with fewer substitutions than expected under neutral evolution (Davydov et al., 2010), which are often the subject of selective pressures (Cooper et al., 2005; Davydov et al., 2010). Any mutations that appear at highly conserved sites are generally assumed to be deleterious (Cooper et al., 2005; Davydov et al., 2010). Likewise, tolerant and intolerant mutations – also referred to as SIFT scores (Vaser et al., 2015) – that are considered of human clinical interest (P. Kumar et al., 2009) – have now been applied to natural

populations (Robinson et al., 2016). ROH, GERP and SIFT scores collectively allow for

inferring the mutational load and genomic health of a population or species.

While quantitatively useful, the relationships between these metrics and conservation

status and diversity metrics are generally not well defined and at-times counter to expectation.

van der Valk et al. (2019) showed a weak relationship between conservation status and

mutational load across mammalian taxa; the highest mutational loads were in fact seen in the

most abundant and diverse species. Similarly, species with the highest inbreeding (ROH) also

had the lowest mutational load and GERP scores (van der Valk et al., 2019). This pattern could

be due the purging of deleterious alleles in small populations, which is consistent with

simulations by Kyriazis et al. (2021). van der Valk et al. (2019) argued that this process has

implications for genetic rescue and suggested that individuals be selected for translocation based

on low mutational load and not diversity. Work on the Sumatran rhinoceros (*Dicerorhinus*

sumatrensis) added another layer of nuance, with populations showing low to moderate levels of

inbreeding (ROH), but a relatively high mutational load (von Seth et al., 2021). In contrast, the

Alpine ibex (*Capra ibex*) that has experienced successive demographic bottlenecks leading to

reduced diversity and increased ROH, has not accumulated highly deleterious alleles (Grossen et

al., 2020). Thus, there seems to be a link between historical demography, ROH, and mutational

load. Understanding such processes has important implications for both predicting long-term

viability, but also selecting animals for genetic rescue.

Here, we analysed signatures of genomic health across populations of three divergent

ungulates to elucidate these relationships. The caribou (*Rangifer tarandus*) is currently

experiencing significant population declines (N_C) prompting international concern (COSEWIC,

2015). In Quebec, five of the major populations have seen drastic declines from their highest

estimates ranging between 10% and 99%, depending on the herd (COSEWIC, 2015, 2017).

Historical estimates of N_e suggest fluctuations tracking glacial cycles, with more recent estimates

in the thousands (Taylor et al. 2021). The white-tailed deer (*Odocoileus virginianus*), in contrast,

is expanding northward and has accordingly high genomic diversity (Kessler et al., 2022). While

white-tailed deer currently number in the 10s of millions (Kauffman et al., 2018) with historic N_e

estimates in the tens of thousands (Lamb et al. 2021), deer were extirpated across much their

range due to colonial harvest and forestry practices (Kauffman et al., 2018); meaning this

demographic expansion is very recent. The mountain goat (*Oreamnos americanus*) underwent a

massive bottleneck during the last glacial maximum with N_e estimates in the hundreds

(Martchenko et al., 2020; Martchenko & Shafer 2022); mountain goats have since remained

relatively stable with a range-wide population of ~100,000 individuals (Festa-Bianchet & Côté,

2008). These three species, contrasted in their ecology but also conservation status, historical and

demographic trajectories, present useful comparisons to explore patterns of novel genomic

metrics as they pertain to conservation status. In line with recent empirical work, we predicted

that populations with low diversity and high inbreeding do not necessarily have higher

mutational load and that genome-wide effects of demographic change like ROH take time to

accrue. Similarly, we expected to see greater purging of deleterious alleles in populations with

long-term small N_e. Collectively, this means that poor genomic metrics of health would not

correlate with current negative population trends.

Methods

Caribou samples were obtained from 67 individuals across Quebec and Labrador (Dedato et al.

2022); 53 white-tailed deer were sampled across their North American range (Kessler et al.

2022). Additionally, 35 mountain goats were sampled across their native range from Alaska to as

far south as Idaho (Martchenko et al. 2022; Figure 1). All samples were sequenced with Illumina

HiSeqX technology to approximately 5X coverage. Non-target sequences were trimmed from

raw FASTQ files with Trimmomatic v.0.36 (Bolger et al., 2014), and reads were mapped to their

reference genome using BWA-MEM v.0.7.17 and SAMtools v.1.15 (Li et al., 2009; Li &

Durbin, 2009). Duplicate mapped reads were marked and removed using Picard v.2.23.0 (Broad

Institute, 2019). Uniquely mapped reads were subset by Sambamba v.0.8.2 (Streit et al., 2012),

and indels were realigned with GATK v.3.8 (Van der Auwera et al., 2013). We estimated

genotype likelihoods using ANGSD v.0.934 and the GATK model: -doVcf 1 -doGeno -4 -doPost

1 -gl 2 -SNP_pval 1e-6 -minMapQ 20 -minQ 20 -doCounts 1 -skipTriallelic 1 -doGlf 2

(Korneliussen et al., 2014). Hard calls required for analyses were performed by Beagle v.5.1

(Browning et al., 2018) with the following parameters: impute=false overlap=3000

window=100000 gprobs=TRUE parameters. Posterior genotype probabilities were estimated

using allele frequencies, and only genotypes with a posterior probability above 0.95 were called.

Tajima's D and nucleotide diversity (π) were calculated using VCFtools v.0.1.13 with windows

of 10 Kb and using all positions, regardless of allele frequency (Danecek et al., 2011).

Inbreeding and Homozygosity

A subset of SNPs in approximate linkage equilibrium was obtained using PLINK v.1.9 with a

window size of 50 SNPs, a window shift size of 10 SNPs, and a pairwise (R^2) linkage

disequilibrium threshold of 0.1 (Chang et al., 2015; Purcell et al., 2007). We then identified runs

of homozygosity (ROH) on scaffolds greater than 10 Mb in length and with a minor allele

frequency cut-off of 0.05 (Foote et al., 2021). Sliding windows of 50 SNPs were used, and each

ROH was defined as a sequence at least 100 Kb long with at least 50 homozygous SNPs (Foote

et al., 2021). Each ROH required a minimum density of 1 SNP per 50 Kb and no more than 10

missing or 3 heterozygous calls in a window. Any SNPs more than 1 Mb apart within a segment

were used to split that segment in two. Once identified, each ROH was classified by size, and the

percent ROH (F_{ROH}) for each individual was calculated. PLINK parameters were taken from

Foote et al. (2021) with output validated using the program ROHan which identifies ROHs using

genomic windows of fixed size (Renaud et al., 2019). ROHan was run on the same data with the

same ROH size definition of 100 Kb and a transition to transversion ratio (Ts/Tv) of 2.25. This

ratio was calculated using BCFtools v.1.14 and the genotypes provided by ANGSD (Li et al.,

2009). The inbreeding coefficient (F) for each individual was also estimated using PLINK and

the –het flag.

Evolutionary Constraint and Mutational Load

Twenty ungulate reference genomes from the NCBI GenBank genome database (Table S1) were

selected to represent most extant families with genomes available in both the orders of

Perissodactyla (odd-toed ungulates) and Artiodactyla (even-toed ungulates). Constrained regions

for each individual of our three focal species were identified using the Genomic Evolutionary

Rate Profiling (GERP++); this approach requires a phylogenetic tree which we constructed for

all 23 species using TimeTree (Figure S1) (Kumar et al., 2017). The raw paired-end reads of all

study individuals of the were then aligned to each of the 20 reference genomes in 50 bp segments

using BWA-MEM to produce 20 alignment files per individual (3100 in total). All duplicate

alignments were marked and removed with Picard. Then, the alignments of each study individual

were concatenated into one multiple alignment file (MAF) in pileup format using HTSbox v.1.0,

resulting in 155 total MAFs (one per individual). This method of alignment uses each study

individual as a focal reference and was first reported by von Seth et al. (2021). To eliminate bias,

each focal species was omitted from the MAF before running GERP++. The gerpcol function

included in GERP++ was run on each MAF to generate rejected substitution (RS) scores for each

alignment column. RS scores are used as a metric for constraint intensity and are calculated by

subtracting the estimated evolutionary rate at each position from the neutral rate (Cooper et al.,

2005; Davydov et al., 2010). Both rates were estimated by gerpcol using the aforementioned

phylogenetic tree (Figure S1). RS scores quantify the number of differences between the

expected and observed number of SNPs at a site, and a positive score is indicative of constraint

(Cooper et al., 2005; Davydov et al., 2010). An RS score greater than 2 represents a highly

constrained region whereas a score lower than 1 is putatively neutral and discarded (Cooper et

al., 2005; Davydov et al., 2010). gerpcol was run with the previously calculated Ts/Tv ratio of

2.25. We also used GERP++ to identify constrained elements with the gerpelem function.

To estimate mutational load, we determined the number of SNPs called by Beagle within

these constrained elements using BEDtools v.2.30.0 (Quinlan & Hall, 2010). Here the ancestral

allele at each site is considered the variant present in the phylogenetically closest outgroup (Fig.

S1), allowing derived alleles to be inferred. These derived SNPs are putatively deleterious due

to the common functional importance of constrained regions (Cooper et al., 2005; Davydov et

al., 2010). For each species, the functional consequence of each SNP called by Beagle was

determined using Ensembl's Variant Effect Predictor (VEP) with default parameters,

corresponding species reference genomes, and GFF format annotation (McLaren et al., 2016;

Table S1). The deleteriousness of each SNP across the genome was then estimated using

functional tolerance scores provided by SIFT4G with default parameters (Vaser et al., 2015).

Using the reference genomes and annotation for each study species alongside UniRef90 (Suzek

et al., 2015), protein databases were generated by SIFT for each species. SIFT calculates

normalised probabilities using these databases where substitutions with scores lower than 0.05

are predicted to be deleterious to protein function (Vaser et al., 2015).

The pN/pS ratios for each individual were calculated as the ratio of nonsynonymous to

synonymous mutations identified by SIFT. Likewise, following Jeffares et al. (2015), the

CODEML application from PAML v.4.9 was used to estimate dN/dS for each individual (Yang,

2007). We filtered each of the MAFs used in the GERP analysis to include only coding regions

using the extractcds program from the EMBOSS suite of software (Rice et al., 2000). The

positive selection model (M2a) was used in CODEML to then analyse each MAF with the

phylogenetic tree from TimeTree.

Caribou did not have an available annotation at the time of analysis which is required for

these analyses; thus, we used a mixture of *de novo* and homology-based predictions. We aligned

protein sequences from *Bos taurus, Equus caballus, Homo sapiens,* and *Ovis aries* to the caribou

genome using blastx v.2.7.1 (Altschul et al., 1990). These sequences were all obtained from

Ensembl release 89 (Hunt et al., 2018). We also aligned transcripts from the NCBI mammalian

RefSeq 10ranscript database v.1.0 using BLAT v.1.0 to help identify exons. Caribou RNAseq

data (SRX12486288) from the NCBI Sequence Read Archive were aligned to the genome with

HISAT v.2.10 for *de novo* prediction (Kim et al., 2015). Using these data, genes were predicted

by Augustus v.3.1.1 as well as GeneMarkES v.4.69 (Lomsadze et al., 2005; Stanke et al., 2006).

We generated a set of consensus genes using EvidenceModeler v.1.1.1 with the following

weighting scheme (Haas et al., 2008): transcript alignment via BLAT (10x weight); protein

alignment via blastx (5x weight); gene prediction via Augustus (1x weight); gene prediction via

GeneMarkES (1x weight). PASA v.2.3.3 was used last to refine gene identifications (Haas et al.,

2008).

For all individual metrics above we quantified the correlation to sample location (latitude,

longitude). Because white-tailed deer were extirpated across much of their range, we used

translocation data from McDonald & Miller, (2004) to determine if these metrics were correlated

to restocking efforts in the United States. We examined correlations with stock start and end

dates, estimated population lows, number of individuals stocked, and the number of states

involved in stocking.

Results

Diversity and Inbreeding

Nucleotide diversity (π) was found to be highest in deer (mean = 0.00771) and lowest in

mountain goats (mean = 0.00101) (Figure 1). Tajima's D was highest in mountain goats (mean =

0.625) and lowest in white-tailed deer (mean = -0.645) (Figure 1). For caribou, Tajima's D was

positive (mean = 0.124), but lower than in mountain goats (t-test p-value $< 2.2 \times 10^{-16}$). Mountain

goats were found to have significantly higher F_{ROH} (mean = 0.234) than both caribou (0.00184)

and white-tailed deer (mean = 0.00380) (Figure 2a). This pattern is reflected in the relative

values of F where mountain goats had the highest value (mean = -0.018) in comparison to

caribou (mean = -0.44) and deer (mean = -0.27) (Figure 2b). For the mountain goat, most ROHs

were found to be in the 500-1000 Kb range whereas caribou and deer possessed ROHs almost

exclusively below 300 Kb (Figure 3). Despite a much higher number of ROHs overall, mountain

goats possessed significantly fewer in the 100-300 Kb range than caribou (Figure 3). Both

measures of inbreeding were found to be correlated in mountain goats (Pearson's r = 0.918, $p <$

0.05), deer (Pearson's r = 0.213, $p < 0.05$), and caribou (Pearson's r = 0.199, $p < 0.05$) (Figure

S2).

Evolutionary Constraint and Mutational Load

Mountain goats had the highest number of constrained elements (Ces) (mean = 2926) while caribou have the fewest (mean = 935; Figure S3). Highly constrained elements followed a nearly identical distribution between species with all three having approximately 50% as many Ces with RS > 2 (Figure S3). In addition to having the most Ces, mountain goats showed slightly longer Ces on average (Figure S4). The three species showed little variability in average CE RS-score (Figure S4, S5). Despite having significantly fewer Ces, white-tailed deer had a larger number of SNPs in Ces (mean = 411) followed by mountain goats (mean = 212) and caribou (mean = 74) (Table 1). The relative differences between species remained the same for highly constrained Ces (Table 1). We observed that a much higher fraction of constrained positions was deleterious in deer (mean = 0.625%) than in caribou (mean = 0.326%) and mountain goats (mean = 0.121%; Figure 4a). The pN/pS ratio for each species was also highest in deer (mean = 0.446) and lowest in caribou (mean = 0.175) and mountain goats (mean = 0.182) (Figure S6); dN/dS for each species followed a similar distribution with deer having much larger ratios (mean = 0.645) than caribou (mean = 0.166) and mountain goats (mean = 0.103) (Figure 4b).

We found 42.1% of coding sequence consequences identified by VEP to be missense variants in white-tailed deer, 49.3% in caribou, and 57.4% in mountain goats (Figure 5a). Despite this distribution, white-tailed deer had the highest number of intolerant mutations (mean = 6852) identified by SIFT followed by caribou (mean = 4944) and mountain goats (mean = 3190) (Table 1). Similarly, we observed white-tailed deer to have the highest ratio of deleterious mutations to genes (mean = 0.330) while caribou (mean = 0.149) and mountain goats (mean = 0.143) had the lowest (Figure 5b). None of these metrics were correlated to the number of genes in the annotation (Table 1).

Lastly, we found F and latitude to be moderately correlated in caribou (Pearson's $r = -$
0.466, $p < 0.05$; Figure S7) and deer (Pearson's $r = -0.513$, $p < 0.05$; Figure S7), but not in
mountain goats (Pearson's $r = 0.003$, $p < 0.05$). There were no significant correlations between
the individual genome health metrics and restocking data in white-tailed deer (Pearson's $r < 0.4$,
$p > 0.05$; Table S2).

Discussion

The interplay between nucleotide diversity (π), inbreeding, mutational load, and census size (N_C),
is nuanced and full of contrasting examples. Here, we demonstrate that mutational load may be
lower than expected in populations with relatively high inbreeding and low diversity. Our
observed patterns suggest that the purging of deleterious alleles in these cases is stronger in small
populations (Huber et al., 2020), and therefore would reduce risks from inbreeding depression.
This supports our hypotheses that low genetic diversity and evidence of inbreeding do not on
their own necessitate special conservation status, and that purging is likely occurring in
populations with low N_e. We also predicted that genomic signatures of demographic events take
time to accrue, which is supported by evidence of F_{ROH} and CEs requiring extended bottlenecks
to surface and increase. Caribou have seen recent decreases in N_C yet have no detectable ROH;
in contrast, mountain goats show elevated ROH, likely due to the ancestral nature of their
bottleneck (Martchenko et al., 2020). The varied relationships between different measures of
genomic health and the lack of correlation to the population trends that are typically used for
conservation status assignment highlights the value of this comparative framework (see also
Huber et al., 2020). This information is especially valuable for small populations of conservation
concern, where mutational load combined with ROH is probably the most informative way
forward, in addition to metrics like Tajima's D (Peart et al., 2020).

Inbreeding Effects

Genomic consequences stemming from demographic change take time to accumulate (e.g. Grossen et al., 2020; von Seth et al., 2021). Historical demographic trends inferred from SMC approaches suggest both caribou and mountain goats responded negatively to the last glacial maximum (Taylor et al. 2021; Martchenko et al. 2020). More recent estimates (<10 KYA) show mountain goats with a considerably smaller N_e (<<1000) compared to caribou (Martchenko & Shafer 2022, Taylor et al. 2020). In contrast, demographic trends in white-tailed deer present as a much more subtle response to glacial cycles, with a noticeably higher recent N_e in the tens of thousands (Haworth et al., 2021; Lamb et al., 2021). Variation in temporal N_e almost certainly stems from current white-tailed deer range being below the ice sheets, suggesting they were less impacted, which was not the case for mountain goats and caribou. Importantly, these SMC-derived patterns of N_e also mirror our estimates of π (Figure 1); this is consistent with long-term estimates of N_e, derived from $\pi = 4N_e\mu$, matching the averaged temporal PSMC estimates of N_e (Peart et al. 2020).

Caribou populations are currently in the midst of a demographic bottleneck resulting in a low contemporary effective and census population size (N_e; Dedato et al., 2022). In contrast, mountain goats underwent a large historical bottleneck during the last glacial maximum, and while they have not recovered, they have remained relatively stable, resulting in a low N_e (Martchenko & Shafer, 2022). Low diversity stemming from bottleneck is supported by positive values of Tajima's D in these two species, which indicate a lack of rare alleles and population contraction (Tajima, 1989). The lower π in mountain goats is clearly due to their extended bottleneck that would have exacerbated any effects of inbreeding, but also their historically lower census sizes. These effects are reflected by F_{ROH}, where mountain goats were found to

have 23.4% of the genome in ROH whereas caribou have < 1%. These differences are also

unlikely to be the result of genome assembly quality, as the mean number and length of analysed

scaffolds were comparable (Table 1).

Compared to caribou and white-tailed deer, mountain goats have developed significantly

longer and more numerous ROHs, which further outline the effects of a prolonged bottleneck.

Caribou simply have not been subject to a low N_C for long enough to approach mountain goats or

even differentiate themselves from deer in their levels of inbreeding, with for example explicit

demographic models detecting historical, not recent bottlenecks (Dedato et al. 2022). Of note,

F_{ROH} was more correlated to F after an extended bottleneck (i.e., mountain goats vs. caribou and

white-tailed deer; Figure S2), which further implies that ROHs require both time and

demographic change to accumulate (see Gillespie, 2004). Overall, these results support the idea

that genome-wide effects of demographic change will not accrue unless given enough time;

genome derived F and F_{ROH} are therefore unlikely to align with IUCN conservation status

decisions based off census data, similarly noted by Schmidt et al. (2022).

Mutational Load and Purging

The relationship between N_e, inbreeding, π, and mutational load are variable (Grossen et al.,

2020; van der Valk et al., 2019; von Seth et al., 2021). Empirical studies on mammals led us to

predict that low diversity and high inbreeding do not necessarily result in the accumulation of

deleterious alleles (Grossen et al., 2020; Robinson et al., 2018, 2018; van der Valk et al., 2019).

Here, we found a much greater number of putatively deleterious mutations in white-tailed deer

than in mountain goats and caribou by identifying CE-bound SNPs with GERP. These results

were confirmed by the deleterious alleles identified by SIFT that showed a similar gap between

white-tailed deer and the other species, regardless of the annotation quality and gene number

(Table 1). This means that despite lower diversity and small N_e, caribou and mountain goats have

relatively low mutational loads; further, despite a relatively low number of CEs, deer still

possessed significantly more CE-bound SNPs indicative of deleterious mutations. Mountain

goats, in contrast, were found to have the highest amount of inbreeding, but accumulated the

fewest deleterious alleles. This reiterates the finding that inbreeding does not necessarily result in

high mutational load; in the case of caribou, if N_C can return (e.g., Peart et al. 2020) or if purging

of deleterious alleles can occur (e.g., Huber et al., 2020), the genetic cost to the current decline in

N_C could be minimized. This, however, remains an open question and likely population specific,

as for example wolves in northern Europe show fixation, not purging, of deleterious alleles

(Smeds & Ellegren, 2022).

The efficiency of purging is dependent, in part, on the rate of inbreeding and the genetic

load architecture (Hedrick & Garcia-Dorado, 2016;); there is evidence that it may occur more

efficiently in the presence of severe bottlenecks (Figure 4b; Grossen et al., 2020). Mountain

goats were shown to have low mutational load in conjunction with generally low diversity, while

caribou present intermediate profiles. This would be consistent with N_C numbers pre-20th century

for caribou, that were presumably much higher (Festa-Bianchet et al., 2011). Reductions in N_e

from severe bottlenecks can increase the fitness effects of recessive deleterious alleles by

increasing homozygosity; this strengthens the pressure of negative selection and can ultimately

lower the mutational load of a population (Garcia-Dorado, 2012, 2015; Glémin, 2003). However,

this effect can be insufficient for overcoming cases of extensive homozygosity. Sumatran rhinos

for example have a very similar historical N_e and ROH profiles to that of mountain goats (von

Seth et al., 2021); while purging might have played some role in some populations in rhinos (von

Seth et al., 2021), the loss of habitat combined with overall negative effect of autozygosity and

continued presence of high frequency deleterious alleles (e.g., Khan et al., 2021) was likely

insurmountable from a population viability standpoint.

Purifying selection should remove SNPs in functionally important regions and constrain

these areas against change, which would generate less SNPs than expected in CEs and lower

dN/dS values (Cooper et al., 2005; Davydov et al., 2010; Kryazhimskiy & Plotkin, 2008). The

mountain goat's longer and more numerous CEs, and lowest dN/dS values relative to the other

species are suggestive of stronger purifying selection. In contrast, caribou have a smaller N_e than

white-tailed deer; thus, purging has either yet to take place or selection is less efficient, or a

combination of both. The trajectory of caribou would be predicted to be towards the patterns of

mountain goat based on mammalian wide patterns and profiles (Cooper et al., 2005; Huber et al.,

2020). It is also important to note that purifying selection reduces nucleotide diversity by

eliminating sites under direct selection and those linked through background selection

(Charlesworth, 1996; Cvijović et al., 2018), which partially explains patterns of diversity seen

here (Figure 1; Buffalo, 2021).

This work has implications for conservation decisions and the expected genomic

consequences of numerical population declines. While deleterious alleles may be purged in

small, inbred populations, extremely high levels of autozygosity can render this benefit

insignificant. Thus, we posit that high mutational load must be accompanied by high ROH to be

of conservation concern, similar to what was observed in woolly mammoths (Rogers & Slatkin,

2017). Further, genetic rescue decisions should not rely solely on inbreeding and diversity

metrics, especially when their relationships to population viability (Carley et al., 2022) and

conservation status (Schmidt et al., 2022) are not straightforward. This is also true for assessing

restocking efforts, as we demonstrated that translocations are not necessarily related to

population health metrics (Table S2). Collectively, this provides further evidence of the time

needed for genome-scale changes to accrue after significant changes in demography. Therefore,

we recommend that genomic health metrics like F_{ROH}, mutational load, and evolutionary

constraint be examined alongside classic metrics like π and Tajima' D for a holistic view that can

provide more accurate assessments of genome health.

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References

Altschul, S. F., Gish, W., Miller, W., Myers, E. W., & Lipman, D. J. (1990). Basic local alignment search

tool. *Journal of Molecular Biology*, *215*(3), 403–410. https://doi.org/10.1016/S0022-

2836(05)80360-2

Ballou, J. D., & Foose, T. J. (2010). Demographic and genetic management of captive populations. In

Wild Mammals in Captivity: Principles and Techniques for Zoo Management, Second Edition

(pp. 263–283). University of Chicago Press.

Bolger, A. M., Lohse, M., & Usadel, B. (2014). Trimmomatic: A flexible trimmer for Illumina sequence

data. *Bioinformatics*, *30*(15), 2114–2120. https://doi.org/10.1093/bioinformatics/btu170

Broad Institute. (2019). *Picard Toolkit*.

Buffalo, V. (2021). Quantifying the relationship between genetic diversity and population size suggests

natural selection cannot explain Lewontin's Paradox. *ELife*, *10*, e67509.

https://doi.org/10.7554/eLife.67509

Carley, L. N., Morris, W. F., Walsh, R., Riebe, D., & Mitchell-Olds, T. (2022). Are genetic variation and

demographic performance linked? *Evolutionary Applications*, *15*(11), 1888–1906.

https://doi.org/10.1111/eva.13487

Ceballos, F. C., Joshi, P. K., Clark, D. W., Ramsay, M., & Wilson, J. F. (2018). Runs of homozygosity:

Windows into population history and trait architecture. *Nature Reviews Genetics 2018 19:4*,

19(4), 220–234. https://doi.org/10.1038/nrg.2017.109

Chang, C. C., Chow, C. C., Tellier, L. C. A. M., Vattikuti, S., Purcell, S. M., & Lee, J. J. (2015). Second-

generation PLINK: rising to the challenge of larger and richer datasets. *GigaScience*, *4*(1), 7.

https://doi.org/10.1186/S13742-015-0047-8

Charlesworth, B. (1996). Background selection and patterns of genetic diversity in Drosophila

melanogaster. *Genetical Research*, *68*(2), 131–149. https://doi.org/10.1017/s0016672300034029

Cooper, G. M., Stone, E. A., Asimenos, G., Comparative, N., Program, S., Green, E. D., Batzoglou, S., &

Sidow, A. (2005). Distribution and intensity of constraint in mammalian genomic sequence.

Genome Research, *15*(7), 901–913. https://doi.org/10.1101/gr.3577405

COSEWIC. (2015). *COSEWIC Assessment and Status Report on the Caribou Rangifer tarandus,*

Newfoundland population, Atlantic-Gaspésie population, Boreal population in Canada.

https://species-registry.canada.ca/index-en.html#/documents/2769

COSEWIC. (2017). *COSEWIC Assessment and Status Report on the Caribou Rangifer tarandus, Eastern*

Migratory population and Torngat Mountains population, in Canada.

http://www.registrelepsararegistry.gc.ca/default.asp?lang=en&n=24F7211B-1

Cvijović, I., Good, B. H., & Desai, M. M. (2018). The Effect of Strong Purifying Selection on Genetic

Diversity. *Genetics*, *209*(4), 1235–1278. https://doi.org/10.1534/genetics.118.301058

Danecek, P., Auton, A., Abecasis, G., Albers, C. A., Banks, E., DePristo, M. A., Handsaker, R. E.,

Lunter, G., Marth, G. T., Sherry, S. T., McVean, G., Durbin, R., & 1000 Genomes Project

Analysis Group. (2011). The variant call format and VCFtools. *Bioinformatics*, *27*(15), 2156–

2158. https://doi.org/10.1093/bioinformatics/btr330

Davydov, E. V., Goode, D. L., Sirota, M., Cooper, G. M., Sidow, A., & Batzoglou, S. (2010). Identifying

a High Fraction of the Human Genome to be under Selective Constraint Using GERP++. *PLOS*

Computational Biology, *6*(12), e1001025. https://doi.org/10.1371/JOURNAL.PCBI.1001025

Dedato, M., Robert, C., Taillon, J., Shafer, A., & Cote, S. (2022). *Demographic history and conservation*

genomics of caribou (Rangifer tarandus) in Québec [Preprint]. Authorea.

https://doi.org/10.22541/au.164165840.02956352/v1

DeWoody, J. A., Harder, A. M., Mathur, S., & Willoughby, J. R. (2021). The long-standing significance

of genetic diversity in conservation. *Molecular Ecology*, *30*(17), 4147–4154.

https://doi.org/10.1111/mec.16051

Ellegren, H., & Galtier, N. (2016). Determinants of genetic diversity. *Nature Reviews Genetics*, *17*(7),

Article 7. https://doi.org/10.1038/nrg.2016.58

Festa-Bianchet, M., & Côté, S. D. (2008). *Mountain Goats: Ecology, Behavior, and Conservation of an*

Alpine Ungulate. Island Press.

Festa-Bianchet, M., Ray, J. C., Boutin, S., Côté, S. D., & Gunn, A. (2011). Conservation of caribou

(Rangifer tarandus) in Canada: An uncertain future1This review is part of the virtual symposium

"Flagship Species□–□Flagship Problems" that deals with ecology, biodiversity and management

issues, and climate impacts on species at risk and of Canadian importance, including the polar

bear (Ursus maritimus), Atlantic cod (Gadus morhua), Piping Plover (Charadrius melodus), and

caribou (Rangifer tarandus). *Canadian Journal of Zoology*, *89*(5), 419–434.

https://doi.org/10.1139/z11-025

Foote, A. D., Hooper, R., Alexander, A., Baird, R. W., Baker, C. S., Ballance, L., Barlow, J., Brownlow,

A., Collins, T., Constantine, R., Dalla Rosa, L., Davison, N. J., Durban, J. W., Esteban, R.,

Excoffier, L., Martin, S. L. F., Forney, K. A., Gerrodette, T., Gilbert, M. T. P., … Morin, P. A.

(2021). Runs of homozygosity in killer whale genomes provide a global record of demographic

histories. *Molecular Ecology*, *30*(23), 6162–6177. https://doi.org/10.1111/mec.16137

Garcia-Dorado, A. (2012). Understanding and Predicting the Fitness Decline of Shrunk Populations:

Inbreeding, Purging, Mutation, and Standard Selection. *Genetics*, *190*(4), 1461–1476.

https://doi.org/10.1534/genetics.111.135541

Garcia-Dorado, A. (2015). On the consequences of ignoring purging on genetic recommendations for

minimum viable population rules. *Heredity*, *115*(3), 185–187.

https://doi.org/10.1038/hdy.2015.28

Gillespie, J. H. (2004). *Population Genetics: A Concise Guide*. JHU Press.

Glémin, S. (2003). How Are Deleterious Mutations Purged? Drift Versus Nonrandom Mating. *Evolution*,

57(12), 2678–2687. https://doi.org/10.1111/j.0014-3820.2003.tb01512.x

Grossen, C., Guillaume, F., Keller, L. F., & Croll, D. (2020). Purging of highly deleterious mutations

through severe bottlenecks in Alpine ibex. *Nature Communications*, *11*(1), Article 1.

https://doi.org/10.1038/s41467-020-14803-1

Haas, B. J., Salzberg, S. L., Zhu, W., Pertea, M., Allen, J. E., Orvis, J., White, O., Buell, C. R., &

Wortman, J. R. (2008). Automated eukaryotic gene structure annotation using EVidenceModeler

and the Program to Assemble Spliced Alignments. *Genome Biology*, *9*(1), R7.

https://doi.org/10.1186/gb-2008-9-1-r7

Haworth, S. E., Nituch, L., Northrup, J. M., & Shafer, A. B. A. (2021). Characterizing the demographic

history and prion protein variation to infer susceptibility to chronic wasting disease in a naïve

population of white-tailed deer (Odocoileus virginianus). *Evolutionary Applications*, *14*(6), 1528–

1539. https://doi.org/10.1111/EVA.13214

Hedrick, P. W., & Garcia-Dorado, A. (2016). Understanding Inbreeding Depression, Purging, and

Genetic Rescue. *Trends in Ecology & Evolution*, *31*(12), 940–952.

https://doi.org/10.1016/j.tree.2016.09.005

Huber, C. D., Kim, B. Y., & Lohmueller, K. E. (2020). Population genetic models of GERP scores

suggest pervasive turnover of constrained sites across mammalian evolution. *PLoS Genetics*,

16(5), e1008827. https://doi.org/10.1371/journal.pgen.1008827

Hunt, S. E., McLaren, W., Gil, L., Thormann, A., Schuilenburg, H., Sheppard, D., Parton, A., Armean, I.

M., Trevanion, S. J., Flicek, P., & Cunningham, F. (2018). Ensembl variation resources.

Database: The Journal of Biological Databases and Curation, *2018*.

https://doi.org/10.1093/database/bay119

IUCN Standards and Petitions Committee. (2019). *Guidelines for Using the IUCN Red List Categories*

and Criteria (No. 14; pp. 1–60). IUCN Red List.

Jeffares, D. C., Tomiczek, B., Sojo, V., & dos Reis, M. (2015). A beginners guide to estimating the non-

synonymous to synonymous rate ratio of all protein-coding genes in a genome. *Methods in*

Molecular Biology (Clifton, N.J.), *1201*, 65–90. https://doi.org/10.1007/978-1-4939-1438-8_4

Kardos, M., Taylor, H. R., Ellegren, H., Luikart, G., & Allendorf, F. W. (2016). Genomics advances the

study of inbreeding depression in the wild. *Evolutionary Applications*, *9*(10), 1205–1218.

https://doi.org/10.1111/eva.12414

Kauffman, M. J., Meacham, J. E., Hall, S., Steingisser, A. Y., Rudd, W. J., & Ostlind, E. (2018). *Wild*

Migrations: Atlas of Wyoming's Ungulates (Vol. 31). University of Nebraska Press.

https://muse.jhu.edu/article/789690

Kelly, E., & Phillips, B. L. (2019). Targeted gene flow and rapid adaptation in an endangered marsupial.

Conservation Biology, *33*(1), 112–121. https://doi.org/10.1111/cobi.13149

Kessler, C., Wootton, E., & Shafer, A. B. A. (2022). *Speciation without gene-flow in hybridizing deer.*

https://doi.org/10.1101/2022.04.20.488928

Khan, A., Patel, K., Shukla, H., Viswanathan, A., van der Valk, T., Borthakur, U., Nigam, P., Zachariah,

A., Jhala, Y. V., Kardos, M., & Ramakrishnan, U. (2021). Genomic evidence for inbreeding

depression and purging of deleterious genetic variation in Indian tigers. *Proceedings of the*

National Academy of Sciences, *118*(49), e2023018118. https://doi.org/10.1073/pnas.2023018118

Kim, D., Langmead, B., & Salzberg, S. L. (2015). HISAT: A fast spliced aligner with low memory

requirements. *Nature Methods*, *12*(4), 357–360. https://doi.org/10.1038/nmeth.3317

Kohn, M. H., Murphy, W. J., Ostrander, E. A., & Wayne, R. K. (2006). Genomics and conservation

genetics. *Trends in Ecology & Evolution*, *21*(11), 629–637.

https://doi.org/10.1016/J.TREE.2006.08.001

Korneliussen, T. S., Albrechtsen, A., & Nielsen, R. (2014). ANGSD: Analysis of Next Generation

Sequencing Data. *BMC Bioinformatics 2014 15:1*, *15*(1), 1–13. https://doi.org/10.1186/S12859-

014-0356-4

Kryazhimskiy, S., & Plotkin, J. B. (2008). The Population Genetics of dN/dS. *PLOS Genetics*, *4*(12),

e1000304. https://doi.org/10.1371/journal.pgen.1000304

Kumar, P., Henikoff, S., & Ng, P. C. (2009). Predicting the effects of coding non-synonymous variants on

protein function using the SIFT algorithm. *Nature Protocols*, *4*(7), Article 7.

https://doi.org/10.1038/nprot.2009.86

Kumar, S., Stecher, G., Suleski, M., & Hedges, S. B. (2017). TimeTree: A Resource for Timelines,

Timetrees, and Divergence Times. *Molecular Biology and Evolution*, *34*(7), 1812–1819.

https://doi.org/10.1093/MOLBEV/MSX116

Kyriazis, C. C., Wayne, R. K., & Lohmueller, K. E. (2021). Strongly deleterious mutations are a primary

determinant of extinction risk due to inbreeding depression. *Evolution Letters*, *5*(1), 33–47.

https://doi.org/10.1002/evl3.209

Lamb, S., Taylor, A. M., Hughes, T. A., McMillan, B. R., Larsen, R. T., Khan, R., Weisz, D.,

Dudchenko, O., Aiden, E. L., Edelman, N. B., & Frandsen, P. B. (2021). De novo chromosome-

length assembly of the mule deer (Odocoileus hemionus) genome. *Gigabyte*, *2021*, 1–13.

https://doi.org/10.46471/gigabyte.34

Lewontin, R. C. (1974). *The Genetic Basis of Evolutionary Change*. Columbia University Press.

Li, H., & Durbin, R. (2009). Fast and accurate short read alignment with Burrows-Wheeler transform.

Bioinformatics, *25*(14), 1754–1760. https://doi.org/10.1093/BIOINFORMATICS/BTP324

Li, H., Handsaker, B., Wysoker, A., Fennell, T., Ruan, J., Homer, N., Marth, G., Abecasis, G., & Durbin,

R. (2009). The Sequence Alignment/Map format and SAMtools. *Bioinformatics (Oxford,*

England), *25*(16), 2078–2079. https://doi.org/10.1093/BIOINFORMATICS/BTP352

Lomsadze, A., Ter-Hovhannisyan, V., Chernoff, Y. O., & Borodovsky, M. (2005). Gene identification in

novel eukaryotic genomes by self-training algorithm. *Nucleic Acids Research*, *33*(20), 6494–

6506. https://doi.org/10.1093/nar/gki937

Martchenko, D., Chikhi, R., & Shafer, A. B. A. (2020). Genome Assembly and Analysis of the North

American Mountain Goat (Oreamnos americanus) Reveals Species-Level Responses to Extreme

Environments. *G3: Genes, Genomes, Genetics*, *10*(2), 437–442.

https://doi.org/10.1534/G3.119.400747

Martchenko, D., & Shafer, A. (2022). *Contrasting whole-genome and reduced representation sequencing*

for population demographic inference: An alpine mammal exampl.

https://doi.org/10.22541/au.164190604.45542461/v1

McDonald, J. S., & Miller, K. V. (2004). *A History of White-tailed Deer Restocking in the United States,*

1878 to 2004. Quality Deer Management Assoc.

McLaren, W., Gil, L., Hunt, S. E., Riat, H. S., Ritchie, G. R. S., Thormann, A., Flicek, P., &

Cunningham, F. (2016). The Ensembl Variant Effect Predictor. *Genome Biology 2016 17:1*,

17(1), 1–14. https://doi.org/10.1186/S13059-016-0974-4

Pascual, U., Adams, W. M., Díaz, S., Lele, S., Mace, G. M., & Turnhout, E. (2021). Biodiversity and the

challenge of pluralism. *Nature Sustainability*, *4*(7), Article 7. https://doi.org/10.1038/s41893-021-

00694-7

Peart, C. R., Tusso, S., Pophaly, S. D., Botero-Castro, F., Wu, C.-C., Aurioles-Gamboa, D., Baird, A. B.,

Bickham, J. W., Forcada, J., Galimberti, F., Gemmell, N. J., Hoffman, J. I., Kovacs, K. M.,

Kunnasranta, M., Lydersen, C., Nyman, T., de Oliveira, L. R., Orr, A. J., Sanvito, S., … Wolf, J.

B. W. (2020). Determinants of genetic variation across eco-evolutionary scales in pinnipeds.

Nature Ecology & Evolution, *4*(8), Article 8. https://doi.org/10.1038/s41559-020-1215-5

Petit-Marty, N., Vázquez-Luis, M., & Hendriks, I. E. (2021). Use of the nucleotide diversity in COI

mitochondrial gene as an early diagnostic of conservation status of animal species. *Conservation*

Letters, *14*(1), e12756. https://doi.org/10.1111/conl.12756

Phelps, M. P., Seeb, L. W., & Seeb, J. E. (2020). Transforming ecology and conservation biology through

genome editing. *Conservation Biology: The Journal of the Society for Conservation Biology*,

34(1), 54–65. https://doi.org/10.1111/cobi.13292

Purcell, S., Neale, B., Todd-Brown, K., Thomas, L., Ferreira, M. A. R., Bender, D., Maller, J., Sklar, P.,

De Bakker, P. I. W., Daly, M. J., & Sham, P. C. (2007). PLINK: A Tool Set for Whole-Genome

Association and Population-Based Linkage Analyses. *The American Journal of Human Genetics*,

81(3), 559–575. https://doi.org/10.1086/519795

Quinlan, A. R., & Hall, I. M. (2010). BEDTools: A flexible suite of utilities for comparing genomic

features. *Bioinformatics*, *26*(6), 841–842. https://doi.org/10.1093/BIOINFORMATICS/BTQ033

Renaud, G., Hanghøj, K., Korneliussen, T. S., Willerslev, E., & Orlando, L. (2019). Joint Estimates of

Heterozygosity and Runs of Homozygosity for Modern and Ancient Samples. *Genetics*, *212*(3),

587–614. https://doi.org/10.1534/GENETICS.119.302057

Rice, P., Longden, I., & Bleasby, A. (2000). EMBOSS: The European Molecular Biology Open Software

Suite. *Trends in Genetics*, *16*(6), 276–277. https://doi.org/10.1016/S0168-9525(00)02024-2

Robinson, J. A., Brown, C., Kim, B. Y., Lohmueller, K. E., & Wayne, R. K. (2018). Purging of Strongly

Deleterious Mutations Explains Long-Term Persistence and Absence of Inbreeding Depression in

Island Foxes. *Current Biology*, *28*(21), 3487-3494.e4. https://doi.org/10.1016/j.cub.2018.08.066

Robinson, J. A., Ortega-Del Vecchyo, D., Fan, Z., Kim, B. Y., vonHoldt, B. M., Marsden, C. D.,

Lohmueller, K. E., & Wayne, R. K. (2016). Genomic Flatlining in the Endangered Island Fox.

Current Biology, *26*(9), 1183–1189. https://doi.org/10.1016/j.cub.2016.02.062

Rogers, R. L., & Slatkin, M. (2017). Excess of genomic defects in a woolly mammoth on Wrangel island.

PLOS Genetics, *13*(3), e1006601. https://doi.org/10.1371/journal.pgen.1006601

Schmidt, C., Hoban, S., Hunter, M., Paz-Vinas, I., & Garroway, C. J. (2022). *The IUCN Red List is not*

sufficient to protect genetic diversity. EcoEvoRxiv. https://doi.org/10.32942/osf.io/hn4by

Shafer, A. B. A., Wolf, J. B. W., Alves, P. C., Bergström, L., Bruford, M. W., Brännström, I., Colling, G.,

Dalén, L., Meester, L. D., Ekblom, R., Fawcett, K. D., Fior, S., Hajibabaei, M., Hill, J. A.,

Hoezel, A. R., Höglund, J., Jensen, E. L., Krause, J., Kristensen, T. N., … Zieliński, P. (2015).

Genomics and the challenging translation into conservation practice. *Trends in Ecology &*

Evolution, *30*(2), 78–87. https://doi.org/10.1016/j.tree.2014.11.009

Smeds, L., & Ellegren, H. (2022). *From high masked to high realized genetic load in inbred*

Scandinavian wolves. Authorea. https://doi.org/10.22541/au.165907442.23809998/v1

Stanke, M., Tzvetkova, A., & Morgenstern, B. (2006). AUGUSTUS at EGASP: Using EST, protein and

genomic alignments for improved gene prediction in the human genome. *Genome Biology*, *7*

Suppl 1, S11.1-8. https://doi.org/10.1186/gb-2006-7-s1-s11

Streit, K., Hammacher, C., Zeller, A., & Hack, S. (2012). Sambamba: A Runtime System for Online

Adaptive Parallelization. In M. O'Boyle (Ed.), *Compiler Construction* (pp. 240–243). Springer.

https://doi.org/10.1007/978-3-642-28652-0_13

Suzek, B. E., Wang, Y., Huang, H., McGarvey, P. B., Wu, C. H., & the UniProt Consortium. (2015).

UniRef clusters: A comprehensive and scalable alternative for improving sequence similarity

searches. *Bioinformatics*, *31*(6), 926–932. https://doi.org/10.1093/bioinformatics/btu739

Tajima, F. (1989). Statistical Method for Testing the Neutral Mutation Hypothesis by DNA

Polymorphism. *Genetics*, *123*(3), 585–595.

Väli, Ü., Einarsson, A., Waits, L., & Ellegren, H. (2008). To what extent do microsatellite markers reflect

genome-wide genetic diversity in natural populations? *Molecular Ecology*, *17*(17), 3808–3817.

https://doi.org/10.1111/j.1365-294X.2008.03876.x

Van der Auwera, G. A., Carneiro, M. O., Hartl, C., Poplin, R., del Angel, G., Levy-Moonshine, A.,

Jordan, T., Shakir, K., Roazen, D., Thibault, J., Banks, E., Garimella, K. V., Altshuler, D.,

Gabriel, S., & DePristo, M. A. (2013). From FastQ Data to High-Confidence Variant Calls: The

Genome Analysis Toolkit Best Practices Pipeline. *Current Protocols in Bioinformatics*, *43*(1),

11.10.1-11.10.33. https://doi.org/10.1002/0471250953.bi1110s43

van der Valk, T., Díez-del-Molino, D., Marques-Bonet, T., Guschanski, K., & Dalén, L. (2019).

Historical Genomes Reveal the Genomic Consequences of Recent Population Decline in Eastern

Gorillas. *Current Biology : CB*, *29*(1), 165-170.e6. https://doi.org/10.1016/J.CUB.2018.11.055

Vaser, R., Adusumalli, S., Ngak Leng, S., Sikic, M., & Ng, P. C. (2015). SIFT missense predictions for

genomes. *Nature Protocols*, *11*. https://doi.org/10.1038/nprot.2015.123

von Seth, J., Dussex, N., Díez-del-Molino, D., van der Valk, T., Kutschera, V. E., Kierczak, M., Steiner,

C. C., Liu, S., Gilbert, M. T. P., Sinding, M. H. S., Prost, S., Guschanski, K., Nathan, S. K. S. S.,

Brace, S., Chan, Y. L., Wheat, C. W., Skoglund, P., Ryder, O. A., Goossens, B., … Dalén, L.

(2021). Genomic insights into the conservation status of the world's last remaining Sumatran

rhinoceros populations. *Nature Communications 2021 12:1*, *12*(1), 1–11.

https://doi.org/10.1038/s41467-021-22386-8

Xue, Y., Prado-Martinez, J., Sudmant, P. H., Narasimhan, V., Ayub, Q., Szpak, M., Frandsen, P., Chen,

Y., Yngvadottir, B., Cooper, D. N., de Manuel, M., Hernandez-Rodriguez, J., Lobon, I.,

Siegismund, H. R., Pagani, L., Quail, M. A., Hvilsom, C., Mudakikwa, A., Eichler, E. E., …

Scally, A. (2015). Mountain gorilla genomes reveal the impact of long-term population decline

and inbreeding. *Science*, *348*(6231), 242–245. https://doi.org/10.1126/science.aaa3952

Yang, Z. (2007). PAML 4: Phylogenetic Analysis by Maximum Likelihood. *Molecular Biology and*

Evolution, *24*(8), 1586–1591. https://doi.org/10.1093/molbev/msm088

Data Availability and Benefit Sharing

All genome assemblies can be found on GenBank (details in Appendix Table S1). Re-sequencing data has been submitted to SRA and accession numbers are pending. Bash and R scripts for all analyses are available on GitLab: https://gitlab.com/WiDGeT_TrentU/undergrad-theses/-/tree/master/Wootton_2022

Benefits Generated: This research addresses important questions about the genomic health of populations of priority concern and illustrates relationships between genomic health and long-term demographic history. Collaborations were formed with researchers across North America, and all major collaborators are included as co-authors. All contributors are listed in the Acknowledgements. We are committed to institutional capacity-building and have shared the results of our research with our collaborators and the broader scientific community.

Author Contributions

ABAS and EW designed the research project. EW conducted all analyses. ABAS and EW wrote the manuscript using comments and contributions from all coauthors.

Tables and Figures

Table 1: The number of scaffolds > 10 Mb, mean scaffold length > 10 Mb, number of genes, and average number of deleterious mutations per species identified by GERP and SIFT (total and homozygous). The number of genes was taken from the annotation of each species.

Species	Caribou	Mountain Goat	White-Tailed Deer
Scaffolds > 10 Mb (n)	68	49	86
Mean scaffold length > 10 Mb	19,723,800	49,667,287	21,282,945
Genes (n)	33,242	22,292	20,750
SNPs in CEs (GERP)	74	212	411
SNPs in CEs with RS > 2 (GERP)	32	74	187
Deleterious mutations (SIFT)	4,943	3,190	6,834
Homozygous deleterious mutations (SIFT)	2,509	1,701	3,570

Figure 1: Sample locations and images of each species. (Red) *Rangifer tarandus* (N=67), (Yellow) *Oreamnos americanus* (N=35), (Blue) *Odocoileus virginianus* (N=53). Current demographic trends are outlined in the top left corner of each image. Boxplots showing Tajima's D and nucleotide diversity (π) per species are also pictured.

Figure 2: (A) Boxplot showing F_{ROH} as calculated by PLINK for all three species. Only ROHs greater than 100 Kb are included in each estimate. (B) Boxplot showing the method-of-moments inbreeding coefficient (*F*) for each species as calculated by PLINK.

Figure 3: (A) Bar plot showing the total number of ROHs in each species stratified by ROH size category in Kb. (B) Bar plot showing the total length (Mb) of all ROHs in each size category (Kb).

Figure 4: (A) Boxplot showing the fraction of constrained positions with SNPs by species. Such SNPs are putatively deleterious, making this the fraction of positions that are deleterious in constrained elements (CEs). (B) Boxplot showing dN/dS calculated for each individual by CODEML.

Figure 5: (A) Bar plot showing the fractions of each coding consequence identified by VEP in each species. (B) Boxplot showing the ratio of deleterious mutations identified by SIFT to the number of genes in each species.

     Supplemental Information

747   **Table S1:** GenBank assembly accession numbers for each species used by GERP.

Species	GenBank Assembly Accession
Equus caballus	GCA_002863925.1
Equus asinus	GCA_016077325.2
Tapirus terrestris	GCA_004025025.1
Diceros bicornis	GCA_004027315.2
Ceratotherium simum	GCA_000283155.1
Camelus dromedarius	GCA_000803125.3
Camelus bactrianus	GCA_000767855.1
Hippopotamus amphibius	GCA_004027065.2
Tragulus javanicus	GCA_004024965.2
Moschus moschiferus	GCA_004024705.2
Moschus chryogaster	GCA_006461725.1
Bos taurus	GCA_002263795.2
Connochaetes taurinus	GCA_006408615.1
Ovis canadensis	GCA_004026945.1
Giraffa camelopardalis	GCA_017591445.1
Antilocapra americana	GCA_007570785.1
Alces alces	GCA_007570765.1
Cervus elaphus	GCA_910594005.1
Phacochoerus africanus	GCA_016906955.1
Sus scrofa	GCA_000003025.6
Rangifer tarandus	GCA_019903745.1
Oreamnos americanus	GCA_009758055.1
Odocoileus virginianus	GCA_014726795.1

749   **Table S2:** Correlation matrix between genomic health metrics and translocation data for white-tailed

750   deer.

	F_{ROH}	F	Fraction of deleterious CEs	Fraction of deleterious genes
Stock start	0.032	0.127	0.076	-0.225
Stock end	0.372	0.392	0.205	0.347
Population low	-0.182	-0.059	-0.177	0.308
Number stocked	-0.102	-0.062	0.162	-0.103
Number of states	-0.161	-0.164	-0.093	0.096

Figure S1: Phylogenetic tree generated by TimeTree showing all 23 ungulate species analysed by GERP++ and estimated times of divergence in MYA (Kumar et al., 2017). Study species with sample data are starred in red (*Odocoileus virginianus*, *Rangifer tarandus*, and *Oreamnos americanus*).

Figure S2: (A) Scatterplot and linear regression line showing the correlation between F and F_{ROH} in caribou (Pearson's $r = 0.199$, $p < 0.05$). (B) Scatterplot and linear regression line showing the correlation between F and F_{ROH} in mountain goats (Pearson's $r = 0.918$, $p < 0.05$). (C) Scatterplot and linear regression line showing the correlation between F and F_{ROH} in white-tailed deer (Pearson's $r = 0.213$, $p < 0.05$).

Figure S3: (A) Boxplot showing the number of constrained elements (CEs) identified per species by
GERP++'s gerpelem function. (B) Boxplot showing the number of highly constrained CEs (RS > 2)
identified in each species by gerpelem.

Figure S4: (A) Boxplot showing the average length of constrained elements (CEs) identified per species.
(B) Boxplot showing the average intensity of constraint (RS-score) of CEs identified in each species.

Figure S5: Histogram showing the fractional distribution of constrained element (CE) RS-scores.

Figure S6: Boxplot showing the ratio of nonsynonymous to synonymous mutations (pN/pS) identified by SIFT in each species.

Figure S7: (A) Scatterplot and linear regression line showing the correlation between *F* and latitude in caribou (Pearson's $r = -0.466$, $p < 0.05$). (B) Scatterplot and linear regression line showing the correlation between *F* and latitude in white-tailed deer (Pearson's $r = -0.513$, $p < 0.05$).