Experimental evaluation of environmental DNA detection of a rare fish in turbid

water

Running title: eDNA experiments in turbid water

Ann E. Holmes[1,2*], Melinda R. Baerwald[3], Jeff Rodzen[4], Brian M. Schreier[3], Brian

Mahardja[5], and Amanda J. Finger[2]

[1]Graduate Group in Ecology, University of California, Davis, 1 Shields Avenue, Davis,

CA, 95616 USA

[2]Genomic Variation Laboratory, Department of Animal Science, University of California,

Davis, 1 Shields Avenue, Davis, CA, 95616 USA

[3]California Department of Water Resources, Sacramento, CA 95814 USA

[4]California Department of Fish and Wildlife, Genetics Research Laboratory, 1415 North

Market Boulevard, Sacramento, CA, 95834 USA

[5]US Department of the Interior, Bureau of Reclamation, Sacramento, CA 95814 USA

*Corresponding author

Ann Holmes, aholmes@ucdavis.edu

Genomic Variation Laboratory, Department of Animal Science, University of California,

Davis, 1 Shields Avenue, Davis, CA, 95616 USA. +1 530-752-6351

Keywords: conservation; delta smelt; endangered species; estuary; environmental

DNA; particulate matter; real-time polymerase chain reaction; turbidity

Abstract

Environmental DNA (eDNA) approaches enable sensitive detection of rare aquatic

species. However, water conditions like turbidity can limit sensitivity, resulting in false

negative detections. The dynamics of eDNA detection in turbid conditions are poorly

understood, but can be better characterized through experimental work. In this study, 1-

L field-collected water samples were spiked with tank-sourced eDNA from a rare,

endangered estuarine fish at concentrations similar to eDNA samples collected from the

natural environment. Samples using non-turbid water (5 NTU), turbid water (50 NTU),

and prefiltered turbid water were filtered using four filter types (pore size range 0.45 µm-

10 µm). Detection success using a species-specific Taqman qPCR assay was assessed

as both eDNA copy number and detection/non-detection. Glass fiber filters (nominal

pore size 1.6 µm) yielded the highest number of eDNA copies and detections in non-

turbid water and the highest detection rate in turbid water when used without a prefilter.

Detection was a more robust metric for evaluating species presence across turbidity

conditions compared with eDNA copy number. Prefiltration improved detection rates for

the other filters tested (polycarbonate and cartridge filters). Filter material and design

appear to interact differently with the prefiltration step, and may be more important

considerations than pore size for eDNA capture in turbid water. Interactions between

eDNA particles, suspended particulate matter, and filters are important to consider for

eDNA methods optimization and interpretation of rare species detections in turbid

water.

Introduction

Environmental DNA (eDNA) can be an efficient tool for surveying rare species (Jerde et al. 2011; Wilcox et al. 2016) that is equally or more sensitive than conventional survey methods (Shaw et al. 2016; Sigsgaard et al. 2017). eDNA has been used to survey species that are extremely difficult to detect or are not detected using other methods (e.g., Budd et al. 2021; Lor et al. 2020; Renan et al. 2017). Indirect detection using eDNA poses little or no risk of sampling-related mortality or stress to both target and non-target organisms, an advantage when targeting or sampling in the vicinity of endangered or sensitive species. Moreover, eDNA samples can usually be collected with less risk to personnel in potentially hazardous conditions (e.g., high gradient streams and rivers). However, despite the advantages of using eDNA to survey rare species, challenging environmental conditions may adversely affect detection sensitivity, resulting in false negative detections.

The sensitivity of eDNA detection is influenced by interacting suites of biological and environmental conditions (Barnes and Turner 2016). Target organism biomass, individual body size, and eDNA shed rate may influence eDNA detection probability (e.g., Sassoubre et al. 2016). Environmental conditions such as water movement, turbidity, temperature, pH, salinity, solar radiation, or microbial community composition (or the related biotic conditions) also affect species detection (e.g., Collins et al. 2018; Jane et al. 2015; Laramie et al. 2015; Seymour et al. 2018; Shogren et al. 2018; Strickler et al. 2015; Tsuji et al. 2019). Turbidity is a previously recognized challenge for eDNA sampling and a suspected cause of reduced sensitivity and false negative detections (Egeter et al. 2018; Williams et al. 2017).

Turbidity can affect eDNA detection in a variety of ways. Turbidity is a measure

of light scatter in water and is associated with reduced water clarity, although they are

not necessarily equivalent measurements. Turbidity is caused by a variety of unrelated

phenomena including particulates spanning a range of sizes and compositions (e.g.,

sediment, inorganic material, or organic material such as plankton and plant detritus)

and decreased water clarity without particulates (e.g., staining by tannins from plant

material). Both particulates and staining can introduce PCR inhibitors (e.g., humic

compounds; Matheson et al. 2010) that interfere with molecular detection. The effects of

PCR inhibitors can be effectively removed without diluting samples using appropriate

extraction methods (Hunter et al. 2015) or with a post-extraction inhibitor removal step

(Williams et al. 2017).

Suspended particulate matter remains a major challenge for eDNA detection.

Particulates clog filters, leading to decreased filtration volumes and long filtration times.

Particulate matter can decrease sensitivity or eliminate eDNA detections altogether

even when the target species is present (Day et al. 2019). Species detection sensitivity

has been positively correlated with volume of water sampled (Hunter et al. 2019;

Schabacker et al. 2020; Sepulveda et al. 2019). Sample volume has been shown to

influence eDNA detection sensitivity more than the number of samples or quantitative

PCR (qPCR) replicates (Schultz and Lance 2015). However, long filtration times may

not be worth the wait; measurements of membrane pressure suggest diminishing

returns on eDNA detection due to increased pressure during filtration (Thomas et al.

2018). Filters with larger pores (Robson et al. 2016) and prefiltration (Takahara et al.

2012) are recommended to increase water volume and decrease filtering time in turbid

systems. While some results suggest a positive effect of prefiltration on eDNA capture

(Takahara et al. 2012), presumably due to increased volume filtered, others are

inconclusive (Majaneva et al. 2018). A better understanding of the impact of turbidity

and methodological adjustments on eDNA detection is necessary, particularly when a

rare species of interest is positively associated with turbidity (e.g., Feyrer et al. 2007;

Nobriga 2002; Sommer et al. 2011).

In this study, we examine the effects of turbidity and filtration methods for

detection of target eDNA in low concentrations. We tested four filter types (pore size

range 0.45 µm-10 µm) and the addition of a prefiltration step for turbid water. Filters

were chosen to capture fish mitochondrial eDNA particles in the size range where they

are most abundant (1-10 µm; Turner et al. 2014; Wilcox et al. 2015). We hypothesized

that (1) turbidity would decrease both eDNA copies detected and the detection rate, (2)

larger filter pore sizes would partially offset the negative effects of suspended

particulate matter on eDNA detection, and (3) prefiltration would improve detection in

turbid water.

Materials & Methods

Study species and habitat

Delta smelt (*Hypomesus transpacificus*) are small (5-7 cm), critically endangered

fish endemic to the San Francisco Estuary (SFE), California, USA. Delta smelt are

considered the sentinel species of the SFE ecosystem (Moyle et al. 2018), significant in

indigenous Miwko□ (Miwok) traditional cultural practice and law (Hankins 2018), and at

risk of extinction in the near future (Moyle et al. 2018). Delta smelt typically have an

annual life cycle and are unusually sensitive to changes in estuarine conditions (Moyle

et al. 1992). The species is protected under both the Federal Endangered Species Act

(ESA) and California Endangered Species Act (CESA) due to a 90% decline in

population abundance over the two decades prior to listing in 1993 (USFWS 1993).

Around 2000, abundance of delta smelt and other pelagic fishes in the SFE again

declined dramatically, most likely due to environmental factors including changes in

water quality, habitat degradation, and effects of introduced species (Sommer et al.

2007; Moyle et al. 2016).

The presence of delta smelt is positively associated with turbid water (Feyrer et

al. 2007; Nobriga et al. 2008; Sommer et al. 2011) perhaps due to decreased predation

risk (Ferrari et al 2014; Bennett and Burau 2015) and increased larval feeding rates

(Baskerville-Bridges et al. 2004; Tigan et al. 2020). The physiological performance of

delta smelt is negatively affected by turbidity levels below 25 NTU and above 80 NTU

(Hasenbein et al. 2016). Turbidity in delta smelt habitat is attributed to suspended

sediment transported from upstream sources or resuspended in the water column due

to wind or turbulence (Schoellhamer 2002).

Delta smelt are difficult to survey due to extremely low abundance, despite

exceptional monitoring efforts by state and federal agencies. The U.S. Fish and Wildlife

Service (USFWS) began year-round, spatially extensive surveys targeting delta smelt

using multiple (conventional) gear types in late 2016 (Enhanced Delta Smelt Monitoring

program (EDSM; USFWS 2022). The EDSM provides data on distribution and

abundance of delta smelt and other species of concern for conservation and

management (Mahardja et al. 2021). Pilot eDNA surveys of delta smelt conducted

alongside EDSM trawls and indicated concordance with trawl sampling, but single

positive qPCR replicates for each sample provide weak evidence of species presence

(Supplementary File S1; Goldberg et al. 2016). Moreover, already low trawl detection

rates continued to decrease (USFWS 2022), making further field testing of eDNA

methods unfeasible. Experimental testing was undertaken to determine if turbidity and

filtration methods were major constraints on eDNA detection of delta smelt.

In vivo testing

Quantitative PCR (qPCR) detection of delta smelt eDNA used a Taqman probe

and primers previously validated using genomic DNA and tested for cross-reactivity with

congener Wakasagi smelt (*Hypomesus nipponensis*) and 21 other SFE fish species

(Baerwald et al. 2011). In this study, the Limit of Detection (LOD) and Limit of

Quantification (LOQ) were determined following guidelines for standardized analysis of

eDNA samples (Klymus et al. 2019; Merkes et al. 2019) using serial dilutions of a

synthetic oligonucleotide gBlocks Gene Fragment (Integrated DNA Technologies, San

Diego, CA) of a portion of the delta smelt cytochrome b gene assayed in 8 replicates

with a starting concentration of 0.1 pg/ul (3.5×10^{6} copies/reaction) and 1:4 subsequent

dilution.

LOD is defined as the lowest concentration in which the target molecule can be

detected in 95% of replicates (Bustin et al. 2009). The theoretical minimum LOD is 3

copies of template DNA per PCR reaction, assuming a Poisson distribution of the target

molecules in PCR reactions. The effective LOD is applied to multiple qPCR replicates,

showing a decrease in LOD with increasing replicates (Klymus et al. 2019). The LOQ

assesses precision using the coefficient of variation (CV) of the measured

concentrations of DNA standards (Kubista 2014). The LOQ is defined as the lowest

concentration at which the CV of qPCR results is less than 35% (Klymus et al. 2019).

Filtration experiment

Filtration used a peristaltic pump (Geotech Environmental Equipment Inc.,

Denver, Colorado). Bottles and other materials used for filtering were sterilized for at

least 20 min in 20% bleach then rinsed three times with clean water. Tubing was

sterilized by pumping 20% bleach through the tube for at least 60 sec then flushing the

tube with clean water for at least 60 sec. The samples were set up and filtered in a

laboratory space free from delta smelt tanks, tissue, or DNA.

Estuarine water was collected in sterilized 5-gal buckets from two sites in a

freshwater region of the upper SFE where delta smelt are not present (Figure 1).

Turbidity of the water collected at the sites was measured at ~5 NTU ("non-turbid") and

~50 NTU ("turbid") with a Hach 2100Q portable turbidimeter. The non-turbid and turbid

designations are relative measures and ecologically relevant to delta smelt; 50 NTU is

less turbid than conditions regularly observed in winter in the SFE (>100 NTU). The

buckets were covered and transported to UC Davis campus. Water was homogenized

by stirring with a sterilized implement before being transferred to sterilized 1 L bottles for

the experiment.

A schematic of the study design is shown in Figure 2. Water from a 340-L tank

(recirculating aquaculture system with daily make-up water to maintain a tank volume)

containing an estimated 186 adult delta smelt at the UC Davis Center for Aquatic

Biology and Aquaculture was collected in a sterile 1-L bottle and stored on wet ice. The

bottle was gently inverted several times to homogenize eDNA prior to pipetting 0.5 mL

tank water into each 1 L bottle of estuarine water. Pilot experiments determined that 0.5

mL tank water added to 1 L of estuary water produced Ct values similar to field

detections. The same process was repeated but 1 mL was added to each bottle in case

0.5 mL tank water was undetectable in some replicates. Adding two small but different

volumes of tank water also allowed us to assess whether small differences in eDNA

concentration can be distinguished at low concentrations. Bottles were placed in a

sterilized cooler with wet ice and filtered within ~8 hours.

Each bottle was gently inverted several times prior to filtering. Three biological

replicates were filtered using each of the four filter types in the three treatments: non-

turbid water, turbid water, and turbid water with the addition of a prefilter (Table 1). This

design resulted in 72 biological replicates (1-L bottles). Glass fiber, polycarbonate filters,

and nylon mesh prefilters were loaded into sterile filter holders (Swinnex-47,

MilliporeSigma) attached to silicon tubing. Sterivex filter cartridges were attached

directly to the tubing. Water was pumped through filters until the 1-L samples was

filtered or flow ceased (usually a maximum of ~15 min). Filtration volumes less than 1 L

reflect filter clogging in turbid water. After filtration, glass fiber and polycarbonate filters

were folded twice and placed in a sterile 2 mL tube and Sterivex filters were capped at

each end. The tubes or capped cartridges were placed in individual sterile plastic bags

and immediately frozen on dry ice. Frozen samples were transferred to -20°C for

storage until extraction. Three negative control samples of estuarine water from each

turbidity value without added tank water were processed with the field samples.

Genetics work was conducted in a dedicated eDNA laboratory space following

recommended guidelines (Goldberg et al. 2016). DNA was extracted in a dedicated

eDNA extraction hood using the DNeasy PowerWater Kit (Qiagen, Hilden, Germany),

which has been shown to effectively remove PCR inhibitors (Eichmiller et al. 2015).

DNA from whole 47 mm filters was extracted using the DNeasy PowerWater Kit and

from Sterivex cartridges were extracted using the DNeasy PowerWater Sterivex Kit.

Extraction protocols followed the manufacturer's instructions including the optional heat

lysis step. Elution buffer incubation time was extended to ~20 minutes and DNA was

eluted into LoBind tubes (Eppendorf, Hamburg, Germany).

Six technical replicates (PCR reactions) of each sample and the estuary water

negative controls were assayed using a species-specific Taqman assay targeting delta

smelt (Baerwald et al. 2011). qPCR reactions set-up in a dedicated eDNA PCR hood

used TaqMan Environmental Master Mix 2.0 (Applied Biosystems, Waltham, MA, USA).

Reagent volumes and cycling conditions are listed in Table 2. qPCR was conducted on

a single CFX Touch Real-Time PCR instrument (Bio-Rad Laboratories, Hercules, CA,

USA) in a laboratory room separate from eDNA extraction and PCR setup hoods. No-

template qPCR controls and gBlock qPCR positive control samples were also assayed

using the same protocol.

Statistical modeling

Results were analyzed using generalized linear mixed-effect models (GLMMs)

and a model comparison approach. Analysis was conducted in R (R Core Team, 2017)

using packages lme4 (Bates et al. 2015) and bbmle (Bolker and R Development Core

Team, 2017). The models used results in non-turbid water, turbid water without a

prefilter, and turbid water with a prefilter (n=144 qPCR reactions for each treatment) to

identify factors that influence success of delta smelt detection under conditions similar

to those observed in the natural environment (turbidity and low eDNA concentrations).

Either eDNA copy number or detection/non-detection can used to evaluate the

factors that influence detection. Copy number allows for a more nuanced interpretation

of detection success, but may not be reliable for low eDNA concentrations (i.e., below

the LOQ). We modeled both eDNA copies detected (as log(copies+1)) and

detection/non-detection in each qPCR replicate as response variables in two models.

The full model for both analyses included five covariates as fixed effects: "turbidity" (a

categorical variable with two levels corresponding to 5 NTU or 50 NTU); "filter type" (a

categorical variable with four levels corresponding to the four filter types); "prefilter" (a

categorical variable with two levels), "volume filtered" (a continuous variable of the

volume of water filtered rounded to the nearest 50 mL), and "volume of tank water

added" (a categorical variable with two levels corresponding to addition of 0.5 or 1 mL of

water from the tank of delta smelt). Interactions between both turbidity and prefilter with

filter type were also considered. Full models included biological replicate (each 1-L

bottle filtered) as a random effect to account for bottle-to-bottle (biological replicate)

variation within treatments. Models were compared using Akaike's Information Criterion

corrected for small sample size (AICc).

Results

In vivo testing

The one replicate Limit of Detection (LOD) was 2.47 copies per PCR reaction

(SE 1.59) and the Limit of Quantification (LOQ) was 67 copies per PCR reaction (Figure

2; Supplementary Files S2, S3). Although less than the theoretical minimum of 3 copies

per reaction, the calculated LOD calculation is acceptable because it falls within the

calculated error (Bustin et al. 2009). The effective LOD for six qPCR replicates (the

number used in this study) was 1.02 copy per PCR reaction (SE 0.15; Supplementary

File S3).

Filtration experiment

In non-turbid water, eDNA detection was 100% for glass fiber, cartridge, and 5-

µm pore polycarbonate filters, and close to 100% for 5-µm pore polycarbonate filters

(Figure 4; Table 3; Supplemental File S4). eDNA copies were highest for glass fiber

filters and cartridge filters, despite the lower volume filtered by cartridge filters due to

clogging (Figure 5; Table 3; Supplemental File S4). In turbid water, eDNA detection was

nearly 100% for glass fiber filters, but below ~75% for Sterivex filters and at or below

50% for both polycarbonate filters (Figure 4; Table 3; Supplemental File S4). Turbidity

reduced the number of eDNA copies detected for all samples, especially the Sterivex

and polycarbonate filters (Figure 5; Table 3; Supplemental File S4). The addition of a

prefilter increased copy numbers and detection rate for all samples except those

collected on glass fiber filters, which were negatively impacted by prefiltration (Figure 4,

5; Table 3; Supplemental File S4). Delta smelt eDNA was not detected in negative

controls of estuarine water or qPCR no template controls.

 Statistical modeling

Model comparison did not support retaining a random effect term for individual

bottles (Supplemental File S5). The highest weighted models tested for both response

variables retained fixed effects filter type and prefilter, and an interaction between the

filter type and prefilter (Table 4; Supplemental File S5). For eDNA copies as the

response variable, the full model and a model where only filtration volume was missing.

For detection/non-detection as a response variable, turbidity and tank water added were

each missing from one of the top two models.

Discussion

In this study, we set out to untangle some of the challenges of eDNA detection of

a very rare target organism in turbid conditions. First, we adapted a protocol for

detection of delta smelt eDNA based on an assay developed for detection of delta smelt

tissue in predator guts (Baerwald et al. 2011). The calculated Limit of Detection (LOD)

and Limit of Quantification (LOD) indicate this part of the protocol is optimized for use

eDNA detection of delta smelt. These limits can help guide interpretation of eDNA

results (Figure 5).

We found that (1) turbidity decreased detection, (2) pore size appear less

important than filter type for increasing detection in turbid water, and (3) prefiltration has

mixed results. In non-turbid water, all filters except 10 µm polycarbonate filters had

100% detection and eDNA copies at or above the LOQ. This result is consistent with the

presumed size of eDNA particles (1-10 µm; Turner et al., 2014). Filters with smaller

pores ($\leq$1 µm) recommended to optimize eDNA capture can lead to reduced sample

volumes and longer filtration times (Li et al. 2018). We did not see a clear pattern of

larger pore sizes (as listed in filter description; Table 1) performing better in turbid water;

filter material and construction may be more important characteristics. Despite low

sample volumes, Sterivex cartridge filters used with a prefilter provided the most

consistent results in terms of eDNA copy number. Cartridge filters are easier to protect

from contamination in the field, however they are more expensive than the circular filters

and extraction is time-consuming. Experiments indicated both a reduction in eDNA

copies and detections in turbid water that could be partially mitigated with a prefilter for

some filter types; prefilters did not appear to perform well when used with glass fiber

filters. Filter type and prefilter status appeared to be the most important influences on

detection. Glass fiber filters used without a prefilter provide appear to provide more

efficient and economical detection of eDNA.

Interpretation of low concentration eDNA

LOD and LOQ help establish standard practices for reporting eDNA detections,

especially for detection of low-concentration eDNA (Klymus et al. 2019). The LOD of the

Taqman assay used for delta smelt eDNA detection (Baerwald et al. 2011) was at the

theoretical lowest limit of 3 copies per reaction, indicating that qPCR detection was well-

optimized. eDNA copy numbers were generally below the LOQ (i.e., unreliable for

quantification). In turbid water, there was not a clear distinction in detection between

samples with 0.5mL and 1L of tank water added in turbid water (Figure 5).

These results also suggest that large turbidity differences between samples can

prevent an apples-to-apples comparison of eDNA sample concentration for the same

target species. Copy number is likely an unreliable metric for using eDNA to model

abundance or biomass of a rare fish in turbid water; presence/absence is a more

straightforward signal to interpret when samples vary turbidity. Our results do not

suggest a strong impact of filtration volume on detection; it may be the interaction of

particulate matter with filters and prefilters in a manner that influences the retention of

particles. (Although the methods used in this study have been shown to effectively

remove PCR inhibition (Eichmiller et al. 2015), but we rule out the possibility of some

PCR inhibition.)

For rare species, even detection/non-detection can be challenging to interpret:

when is a weak signal considered a positive detection? High Cq values (>40) are often

treated as unreliable and therefore interpreted as potential false positive detections.

However, the common practice of setting non-detect values to 40 may introduce bias

(McCall et al. 2014) and increase the false negative detection rate. As in many areas of

science, it is impossible to "prove the negative." In this study, 3 of 432 qPCR reactions

(less than 1%) assaying samples with the addition of delta smelt tank water generated

Cq values >40 (Table S4). The results of our analysis of these samples known to

contain target eDNA suggest that, although rare, Cq values >40 can represent true

detections. There was no evidence of contamination in negative control samples. In

addition, although the protocol ran 50 cycles, there were no detections above Cq of 42.

Similarly, an evaluation of eDNA metabarcoding laboratory protocols shows that, above

a certain threshold, additional PCR cycles do not improve species detection (Stoeckle et

al. 2022). Given the current limits of technology, interpretation of weak signals is a

balancing act between signal and noise and likely specific to each particular application

of eDNA.

Effect of suspended particulate matter on filtering and detection

As expected, filters with smaller pores filtered less water. A positive relationship

has been demonstrated between sample volume and species detection in both turbid

(Williams et al. 2017) and non-turbid water, (Wilcox et al. 2015; Wilcox et al. 2016;

Sepulvida et al. 2019; Bedwell and Goldberg 2020), although under certain conditions

there may not be a relationship between sample volume and species detection (Mächler

et al. 2016). Suspended particulate matter clogs filters and increases filtration pressure,

requiring re-optimization of the capture method (Thomas et al. 2018). Although we did

not measure membrane pressure during our experiment, our data is consistent with

poor eDNA capture due to high membrane pressure. Membrane pressure may

decrease eDNA retention by breaking apart clumps of cells or bursting cells or

mitochondria (Thomas et al. 2018). Glass fiber filters, which performed best in turbid

water, have a completely different pore type and construction than the other filters used

in this study (Table 1).

Pore sizes are not necessarily comparable between filter materials. Filter

materials have different pore types. Absolute pores (e.g., polycarbonate and

polyvinylidene fluoride (PVDF)) are uniform in size. These filters act as a screen,

retaining all particles larger than the pores on the filter surface. Glass fiber filters have

nominal pores that are irregular and retain only a percentage of particles larger than the

pore size and are depth filters with multiple layers that trap particles inside a structure.

(Cellulose and cellulose nitrate are other nominal pore filter types that are commonly

used for eDNA capture.) Despite lower capture efficiency, the thickness of depth filters

may provide relatively more space to capture particles in turbid water. Glass fiber filters

are significantly less expensive than the other filters used in this study. However,

PowerWater extractions use a bead beating step that causes glass fiber to become

sponge-like, requiring more time and care to separate the supernatant from the beads

and filter.

eDNA capture may be influenced by the characteristics of the suspended

particulates, which vary widely. In samples from experimental ponds with turbidity up to

60 NTU, turbidity was positively associated with eDNA detected using 10-μm pore

polycarbonate filters and prefiltration, suggesting that eDNA was sticking to larger

phytoplankton in the ponds (Barnes et al. 2020). Turbidity in river-dominated estuaries

like the SFE is mainly caused by river inputs of suspended particulate matter and

resuspension of bottom sediments that reduce phytoplankton biomass (Cloern 1987).

While eDNA can be detected in samples of suspended particulate matter collected in

sedimentation boxes in rivers (Díaz et al. 2020), it is not clear if eDNA co-occurs with

other particulates or is stuck to them. In our experiment, turbidity reduced detection

rates and we did not see evidence of eDNA sticking to particulates. Cellular studies

suggest that cells may be more likely to adhere to each other than foreign material

(Coman 1961).

Effect of prefiltration on detection

Prefiltration is sometimes recommended to increase sample volume and

decrease filtration time. Our results indicated a significant interaction between filter type

and prefiltration. One explanation for the negative impact of prefiltration on glass fiber

filters is that prefiltration breaks eDNA particles that cannot be efficiently captured by

glass fiber. For example, clumps of cells may be broken up into individual cells or whole

cells may be reduced to mitochondria. In cases where copy number quantification is

feasible, prefiltration in combination with end filters with small pores may provide more

consistent results across turbidity conditions. Prefiltration also increases the cost and

effort in filtering. If the study goal is to determine species presence/absence, and budget

or time is limited, then our results indicate the most practical approach for rare species

detection in turbid conditions is to use glass fiber filters without prefiltration.

Future directions

Experimental work is necessarily limited to a relatively narrow range of

treatments but can help tease out the effects of turbidity on eDNA detection without the

complexity of natural systems. eDNA is not homogeneously distributed in natural

environments, making it more difficult to determine if decreased detections in turbid

water reflect real patterns of occurrence or limitations of eDNA detection. eDNA

samples collected in relatively more turbid water (Secchi depth 73 cm; see

Supplemental File S6 for the relationship between Secchi depth and turbidity) can yield

more detections when turbid conditions are more favorable (Kumar et al. 2021).

Experimental work encompassing a wide range of turbidity conditions could help further

our understanding of eDNA detection in turbid water. A correction factor could be

developed to account for the decrease in detection observed as turbidity increases,

although such corrections may be particle- or habitat-specific.

Finally, turbidity (an optical measurement) is often approximated using other

metrics (e.g., water clarity, total suspended solids, filtration time), severely restricting the

ability to compare eDNA studies conducted in turbid water. Comparable measurements

may help optimize eDNA methods and data interpretation when turbidity is a significant

characteristic of the target species habitat. Optical measurements taken using

turbidimeters and probes are the most objective, repeatable, and accurate across a

broad range of values of turbidity and can be used in most water bodies (Pickering

1976). Water clarity measured by Secchi disk is a less expensive alternative but cannot

be used certain conditions (e.g., fast moving water) and is subject to human error

(Carlson and Simpson 1996). Filtration time as a proxy for turbidity is less useful

because it is difficult to calibrate across different studies and filtration set-ups.

Conclusions

Many critically endangered fishes are data-deficient (Guy et al. 2021), presenting

a perfect opportunity to employ high sensitivity methods like eDNA detection for surveys

and monitoring. eDNA detection methods, however, are not "one size fits all" (Barnes

and Turner 2016; Kumar et al. 2021). We use this comparison of eDNA capture

methods under controlled conditions to help guide best practices for the real-world

challenge of detecting a rare species in turbid conditions. Turbidity and filter type

influence eDNA detection success and prefiltration may not always be beneficial. These

findings provide optimism that reliable and repeatable eDNA detections of rare species

are possible in turbid when appropriately optimized methods are used.

Data accessibility

Data and R code are available at https://github.com/annholmes/eDNA-experiments-in-

turbid-water.

Acknowledgements

Thank you to Grace Auringer, Alyssa Benjamin, Alisha Goodbla, Leslie Guerrero,

and Shannon Kieran for assistance and advice on sampling and laboratory work, to

Dennis Cocherell, Brittany Davis, Luke Ellison, Nann Fangue, and Tien-Chieh Hung for

access to delta smelt tanks, and to Ted Sommer and Andrea Schreier for feedback and

support. We are grateful to Chris Hart, Jessica Adams, Denise Barnard, Bill Powell, and

the US Fish and Wildlife Service Enhanced Delta Smelt Monitoring (EDSM) Program for

facilitating paired eDNA sampling with trawl surveys. The University of California, Davis

sits on Patwin Land.

References

Baerwald, M. R., Schumer, G., Schreier, B. M., & May, B. (2011). TaqMan assays for

the genetic identification of delta smelt (*Hypomesus transpacificus*) and Wakasagi

smelt (*Hypomesus nipponensis*). *Molecular Ecology Resources*, *11*(5), 784–785.

Barnes, M. A., & Turner, C. R. (2016). The ecology of environmental DNA and

implications for conservation genetics. *Conservation Genetics, 17*(1), 1–17.

https://doi.org/10.1007/s10592-015-0775-4

Baskerville-Bridges, B., Lindberg, J. C., and Doroshov, S. I. (2004). The effect of light

intensity, alga concentration, and prey density on the feeding behavior of delta

smelt larvae. *American Fisheries Society Symposium*, 39: 219–228

Bates, D., M. Maechler, B. Bolker, and S. Walker. (2015). Fitting linear mixed-effects

models using lme4. Journal of Statistical Software 67:1–48.

Bedwell, M. E., & Goldberg, C. S. (2020). Spatial and temporal patterns of

environmental DNA detection to inform sampling protocols in lentic and lotic

systems. *Ecology and Evolution, 00*, 1–11. https://doi.org/10.1002/ece3.6014

Bennett, W. A., & Burau, J. R. (2015). Riders on the Storm: Selective Tidal Movements

Facilitate the Spawning Migration of Threatened Delta Smelt in the San Francisco

Estuary. *Estuaries and Coasts, 38*(3), 826–835. https://doi.org/10.1007/s12237-

014-9877-3

Budd, A. M., Cooper, M. K., Port, A. Le, Schils, T., Mills, M. S., Deinhart, M. E., …

Strugnell, J. M. (2021). First detection of critically endangered scalloped

hammerhead sharks (Sphyrna lewini) in Guam, Micronesia, in five decades using

environmental DNA. *Ecological Indicators, 127*.

https://doi.org/10.1016/j.ecolind.2021.107649

Bustin, S. A., Benes, V., Garson, J. A., Hellemans, J., Huggett, J., Kubista, M., …

Wittwer, C. T. (2009). The MIQE Guidelines: Minimum Information for Publication of

Quantitative Real-Time PCR Experiments. *Clinical Chemistry*, *55*(4), 611–622.

https://doi.org/10.1373/clinchem.2008.112797

Carlson, R.E. and J. Simpson. 1996. *A Coordinator's Guide to Volunteer Lake*

Monitoring Methods. North American Lake Management Society. 96 pp.

Cloern, J. E. (1987). Turbidity as a control on phytoplankton biomass and productivity in

estuaries. Continental Shelf Research, 7(11–12), 1367–1381.

https://doi.org/10.1016/0278-4343(87)90042-2

Collins, R. A., Wangensteen, O. S., O'Gorman, E. J., Mariani, S., Sims, D. W., &

Genner, M. J. (2018). Persistence of environmental DNA in marine systems.

Communications Biology, *1*(1), 185. https://doi.org/10.1038/s42003-018-0192-6

Coman, D. R. (1961). Adhesiveness and Stickiness: Two Independent Properties of the

Cell Surface. *Cancer Research*, *21*(10), 1436–1438.

Díaz, C., Wege, F. F., Tang, C. Q., Crampton-Platt, A., Rüdel, H., Eilebrecht, E., &

Koschorreck, J. (2020). Aquatic suspended particulate matter as source of eDNA

for fish metabarcoding. *Scientific Reports*, *10*(1), 1–8.

https://doi.org/10.1038/s41598-020-71238-w

Day, K., Campbell, H., Fisher, A., Gibb, K., Hill, B., Rose, A., & Jarman, S. N. (2019).

Development and validation of an environmental DNA test for the endangered

Gouldian finch. *Endangered Species Research*, *40*, 171–182.

https://doi.org/10.3354/ESR00987

Egeter, B., Peixoto, S., Brito, J. C., Jarman, S., Puppo, P., & Velo-Antón, G. (2018).

Challenges for assessing vertebrate diversity in turbid Saharan water-bodies using

environmental DNA. *Genome*, *61*(11), 807–814. https://doi.org/10.1139/gen-2018-

0071

Eichmiller, J. J., Miller, L. M., & Sorensen, P. W. (2015). Optimizing techniques to

capture and extract environmental DNA for detection and quantification of fish.

Molecular Ecology Resources, n/a-n/a. https://doi.org/10.1111/1755-0998.12421

Ferrari, M. C. O., Ranåker, L., Weinersmith, K. L., Young, M. J., Sih, A., & Conrad, J. L.

(2014). Effects of turbidity and an invasive waterweed on predation by introduced

largemouth bass. *Environmental Biology of Fishes, 97*(1), 79–90.

https://doi.org/10.1007/s10641-013-0125-7

Feyrer, F., Nobriga, M. L., & Sommer, T. R. (2007). Multidecadal trends for three

declining fish species: habitat patterns and mechanisms in the San Francisco

Estuary, California, USA. *Canadian Journal of Fisheries and Aquatic Sciences*,

64(4), 723–734. https://doi.org/10.1139/f07-048

Goldberg, C. S., Turner, C. R., Deiner, K., Klymus, K. E., Francis, P., Murphy, M. A., …

Fremier, A. K. (2016). Critical considerations for the application of environmental

DNA methods to detect aquatic species. *Methods in Ecology and Evolution*, *7*,

1299–1307. https://doi.org/10.1111/2041-210X.12595

Guy, C. S., Cox, T. L., Williams, J. R., Brown, C. D., Eckelbecker, R. W., Glassic, H. C.,

… Siemiantkowski, M. J. (2021). A paradoxical knowledge gap in science for

critically endangered fishes and game fishes during the sixth mass extinction.

Scientific Reports, *11*(1), 1–9. https://doi.org/10.1038/s41598-021-87871-y

Hankins, D. L. (2018). Ecocultural equality in the Miwko☐ Waali☐ San Francisco

Estuary and Watershed Science, 16(3), 0–11.

https://doi.org/10.15447/sfews.2018v16iss3art1

Hasenbein, M., Fangue, N. A., Geist, J., Komoroske, L. M., Truong, J., McPherson, R.,

& Connon, R. E. (2016). Assessments at multiple levels of biological organization

allow for an integrative determination of physiological tolerances to turbidity in an

endangered fish species. Conservation Physiology, 4(1), 1–16.

https://doi.org/10.1093/conphys/cow004

Hunter, M. E., Ferrante, J. A., Meigs-Friend, G., & Ulmer, A. (2019). Improving eDNA

yield and inhibitor reduction through increased water volumes and multi-filter

isolation techniques. *Scientific Reports*, *9*(1), 1–9. https://doi.org/10.1038/s41598-

019-40977-w

Hunter, M. E., Oyler-McCance, S. J., Dorazio, R. M., Fike, J. A., Smith, B. J., Hunter, C.

T., Reed, R. N. & Hart, K. M. (2015). Environmental DNA (eDNA) sampling

improves occurrence and detection estimates of invasive Burmese pythons. *PLoS*

ONE, *10*(4), 1–17. https://doi.org/10.1371/journal.pone.0121655

Jane, S. F., Wilcox, T. M., Mckelvey, K. S., Young, M. K., Schwartz, K., Lowe, W. H., &

Letcher, B. H. (2015). Distance, flow and PCR inhibition: eDNA dynamics in two

headwater streams. *Molecular Ecology Resources*, *15*, 216–227.

https://doi.org/10.1111/1755-0998.12285

Jerde, C. L., Mahon, A. R., Chadderton, W. L., & Lodge, D. M. (2011). "Sight-unseen"

detection of rare aquatic species using environmental DNA. *Conservation Letters*,

4, 150–157. https://doi.org/10.1111/j.1755-263X.2010.00158.x

Klymus, K. E., Merkes, C. M., Allison, M. J., Goldberg, C. S., Helbing, C. C., Hunter, M.

E., … Richter, C. A. (2019). Reporting the limits of detection and quantification for

environmental DNA assays. *Environmental DNA*, *2*(3), 271–282.

https://doi.org/10.1002/edn3.29

Kubista, M. (2014). Prime time for qPCR - Raising the quality bar. *European*

Pharmaceutical Review, 19(3), 63–67.

Kumar, G., Farrell, E., Reaume, A. M., Eble, J. A., & Gaither, M. R. (2021). One size

does not fit all: Tuning eDNA protocols for high- and low-turbidity water sampling.

Environmental DNA, (May), 1–14. https://doi.org/10.1002/edn3.235

Laramie, M. B., Pilliod, D. S., Goldberg, C. S., & Strickler, K. M. (2015). Environmental

DNA sampling protocol - filtering water to capture DNA from aquatic organisms.

U.S Geological Survey Techniques and Methods, *Book 2*(Chapter A13), 15 p.

https://doi.org/10.3133/TM2A13

Li, J., Lawson Handley, L.-J., Read, D. S., & Hänfling, B. (2018). The effect of filtration

method on the efficiency of environmental DNA capture and quantification via

metabarcoding. *Molecular Ecology Resources*. https://doi.org/10.1111/1755-

0998.12899

Lor, Y., Schreier, T. M., Waller, D. L., & Merkes, C. M. (2020). Using environmental

DNA (eDNA) to detect the endangered Spectaclecase mussel (*Margaritifera*

monodonta). *Freshwater Science*, *39*(4), 837–847. https://doi.org/10.1086/71167

Mächler, E., Deiner, K., Spahn, F., & Altermatt, F. (2016). Fishing in the Water: Effect of

Sampled Water Volume on Environmental DNA-Based Detection of

Macroinvertebrates. *Environmental Science and Technology*, *50*(1), 305–312.

https://doi.org/10.1021/acs.est.5b04188

Mahardja, B., Mitchell, L., Beakes, M., Johnston, C., Graham, C., Goertler, P., …

Matthias, B. (2021). Leveraging Delta Smelt Monitoring for Detecting Juvenile

Chinook Salmon in the San Francisco Estuary. *San Francisco Estuary and*

Watershed Science, *19*(1), 1–24. https://doi.org/10.15447/sfews.2021v19iss1art2

Matheson, C. D., Gurney, C., Esau, N., & Lehto, R. (2010). Assessing PCR Inhibition

from Humic Substances. *The Open Enzyme Inhibition Journal*, *3*, 38–45.

https://doi.org/10.2174/1874940201003010038

McCall, M. N., McMurray, H. R., Land, H., & Almudevar, A. (2014). On non-detects in

qPCR data. *Bioinformatics*, *30*(16), 2310–2316.

https://doi.org/10.1093/bioinformatics/btu239

Merkes C. M., Klymus K. E., Allison M. J., Goldberg C., Helbing C. C., Hunter M. E.,

Jackson C. A., Lance R. F., Mangan A. M., Monroe E. M., Piaggio A. J., Stokdyk J.

P., Wilson C. C., & Richter C. (2019). Generic qPCR Limit of Detection (LOD) /

Limit of Quantification (LOQ) calculator. R Script. Available at:

https://github.com/cmerkes/qPCR_LOD_Calc. https://doi.org/10.5066/P9GT00GB

(Accessed: 2022-07-26).

Moyle, P. B., Bennett, W. A., Fleenor, W. E., & Lund, J. R. (2010). Habitat Variability

and Complexity in the Upper San Francisco Estuary. *San Francisco Estuary and*

Watershed Science, *8*(3), 0–24. https://doi.org/10.15447/sfews.2010v8iss3art1

Moyle, P. B., & Bennett, W. A. (2008). The Future of the Delta Ecosystem and Its Fish.

Technical Appendix D. In Lund J, Hanak H, Fleenor W, Bennett, W, Howitt R,

589 Mount J, Moyle P. Comparing futures for the Sacramento–San Joaquin Delta.

590 *Public Policy Institute of California*, 46.

591 Moyle, P., Brown, L., Durand, J., & Hobbs, J. (2016). Delta Smelt: Life History and

592 Decline of a Once-Abundant Species in the San Francisco Estuary. *San Francisco*

593 *Estuary and Watershed Science*, *14*(2).

594 https://doi.org/10.15447/sfews.2016v14iss2art6

595 Moyle, P. B., Herbold, B., Stevens, D. E., & Miller, L. W. (1992). Life History and Status

596 of Delta Smelt in the Sacramento-San Joaquin Estuary, California. *Transactions of*

597 *the American Fisheries Society*, *121*(1), 67–77. https://doi.org/10.1577/1548-

598 8659(1992)121<0067:LHASOD>2.3.CO;2

599 Moyle, P. B., Hobbs, J. A., & Durand, J. R. (2018). Delta Smelt and Water Politics in

600 California. *Fisheries*, *43*(1), 42–51. https://doi.org/10.1002/fsh.10014

601 Nobriga, M., Feyrer, F., Baxter, R. D., & Chotkowski, M. (2005). Fish Community

602 Ecology in an Altered River Delta: Spatial Patterns in Species Composition, Life

603 History Strategies, and Biomass. *Estuaries*, *28*(5), 776–785.

604 Nobriga, M. L. (2002). Larval Delta smelt diet composition and feeding incidence:

605 Environmental and ontogenetic influences. *California Fish and Game*, *88*(4), 149–

606 164.

607 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., and the R Development Core Team

608 (2013). nlme: Linear and Nonlinear Mixed Effects Models. R package version 3.1-

609 108. https://cran.r-project.org/web/packages/nlme/index.html

610 Pickering, R. (1976). *Measurement of "turbidity" and related characteristics of natural*

611 *waters.*

Renan, S., Gafny, S., Perl, R. G. B., Roll, U., Malka, Y., Vences, M., & Geffen, E.

(2017). Living quarters of a living fossil — Uncovering the current distribution

pattern of the rediscovered Hula painted frog (*Latonia nigriventer*) using

environmental DNA. *Molecular Ecology*, *26*, 6801–6812.

https://doi.org/10.1111/mec.14420

Robson, H. L. A., Noble, T. H., Saunders, R. J., Robson, S. K. A., Burrows, D. W., &

Jerry, D. R. (2016). Fine-tuning for the tropics: application of eDNA technology for

invasive fish detection in tropical freshwater ecosystems. *Molecular Ecology*

Resources, *16*(4), 922–932. https://doi.org/10.1111/1755-0998.12505

Sassoubre, L. M., Yamahara, K. M., Gardner, L. D., Block, B. A., & Boehm, A. B.

(2016). Quantification of environmental DNA (eDNA) shedding and decay rates for

three marine fish. *Environmental Science and Technology*, *50*, 10456–10464.

https://doi.org/10.1021/acs.est.6b03114

Schabacker, J. C., Amish, S. J., Ellis, B. K., Gardner, B., Miller, D. L., Rutledge, E. A.,

Sepulvida, A. J. & Luikart, G. (2020). Increased eDNA detection sensitivity using a

novel high-volume water sampling method. *Environmental DNA*, *2*, 244-251.

https://doi.org/10.1002/edn3.63

Schultz, M. T., & Lance, R. F. (2015). Modeling the sensitivity of field surveys for

detection of environmental DNA (eDNA). *PLoS ONE*, *10*(10), 1–16.

https://doi.org/10.1371/journal.pone.0141503

Schoellhamer, D. H. (2002). Variability of suspended-sediment concentration at tidal to

annual time scales in San Francisco Bay, USA. Continental Shelf Research,

22(11–13), 1857–1866. https://doi.org/10.1016/S0278-4343(02)00042-0

Sepulveda, A. J., Schabacker, J., Smith, S., Al-Chokhachy, R., Luikart, G., & Amish, S.

J. (2019). Improved detection of rare, endangered and invasive trout in using a new

large-volume sampling method for eDNA capture. *Environmental DNA*, *1*(3), 227–

237. https://doi.org/10.1002/edn3.23

Seymour, M., Durance, I., Cosby, B. J., Ransom-Jones, E., Deiner, K., Ormerod, S. J.,

… Creer, S. (2018). Acidity promotes degradation of multi-species environmental

DNA in lotic mesocosms. *Communications Biology*, *1*(1), 4.

https://doi.org/10.1038/s42003-017-0005-3

Shaw, J. L. A., Clarke, L. J., Wedderburn, S. D., Barnes, T. C., Weyrich, L. S., &

Cooper, A. (2016). Comparison of environmental DNA metabarcoding and

conventional fish survey methods in a river system. *Biological Conservation*, *197*,

131–138. https://doi.org/10.1016/j.biocon.2016.03.010

Shogren, A. J., Tank, J. L., Egan, S. P., August, O., Rosi, E. J., Hanrahan, B. R., …

Bolster, D. (2018). Water Flow and Biofilm Cover Influence Environmental DNA

Detection in Recirculating Streams. *Environmental Science and Technology*,

52(15), 8530–8537. https://doi.org/10.1021/acs.est.8b01822

Sigsgaard, E. E., Nielsen, I. B., Carl, H., Krag, M. A., Knudsen, S. W., Xing, Y., …

Thomsen, P. F. (2017). Seawater environmental DNA reflects seasonality of a

coastal fish community. *Marine Biology*, *164*(6), 1–15.

https://doi.org/10.1007/s00227-017-3147-4

Sommer, T., Armor, C., Baxter, R., Breuer, R., Brown, L., Chotkowski, M., … Souza, K.

(2007). The collapse of pelagic fishes in the upper San Francisco estuary.

Fisheries, *32*(6), 270–277.

Sommer, T., Mejia, F., Nobriga, M., Feyrer, F., & Grimaldo, L. (2011). The spawning

migration of delta smelt in the Upper San Francisco Estuary. *San Francisco*

Estuary & Watershed Science, *9*(2), 1–16.

https://doi.org/10.5811/westjem.2011.5.6700

Strickler, K. M., Fremier, A., & Goldberg, C. S. (2015). Quantifying effects of UV-B,

temperature, and pH on eDNA degradation in aquatic microcosms. *Biological*

Conservation, *183*, 85–92.

Stoeckle, M. Y., Adolf, J., Ausubel, J. H., Charlop-Powers, Z., Dunton, K. J., & Hinks, G.

(2022). Current laboratory protocols for detecting fish species with environmental

DNA optimize sensitivity and reproducibility, especially for more abundant

populations. *ICES Journal of Marine Science*.

https://doi.org/10.1093/icesjms/fsab273

Takahara, T., Minamoto, T., Yamanaka, H., Doi, H., & Kawabata, Z. (2012). Estimation

of fish biomass using environmental DNA. *PLoS ONE*, *7*(4), e35868.

https://doi.org/10.1371/journal.pone.0035868

Thomas, A. C., Howard, J., Nguyen, P. L., Seimon, T. A., & Goldberg, C. S. (2018).

eDNA Sampler: A fully integrated environmental DNA sampling system. *Methods in*

Ecology and Evolution, *9*, 1379–1385. https://doi.org/10.1111/2041-210X.12994

Tigan, G., Mulvaney, W., Ellison, L., Schultz, A., & Hung, T. C. (2020). Effects of light

and turbidity on feeding, growth, and survival of larval Delta Smelt (*Hypomesus*

transpacificus, Actinopterygii, Osmeridae). *Hydrobiologia, 847*(13), 2883–2894.

https://doi.org/10.1007/s10750-020-04280-4

Tsuji, S., Takahara, T., Doi, H., Shibata, N., & Yamanaka, H. (2019). The detection of

aquatic macroorganisms using environmental DNA analysis-A review of methods

for collection, extraction, and detection. *Environmental DNA*, (April), 1–10.

https://doi.org/10.1002/edn3.21

Turner, C. R., Barnes, M. A., Xu, C. C. Y., Jones, S. E., Jerde, C. L., & Lodge, D. M.

(2014). Particle size distribution and optimal capture of aqueous macrobial eDNA.

Methods in Ecology and Evolution, *5*(7), 676–684. https://doi.org/10.1111/2041-

210X.12206

USFWS (United States Fish and Wildlife Service). (1993). Determination of threatened

status for the delta smelt. Federal Register 58:12854-12864.

USFWS (United States Fish and Wildlife Service), Senegal, T., Mckenzie, R., Speegle,

J., Perales, B., Bridgman, D., Erly, K., Staiger, S., Arrambide, A., and Gilbert, M.

(2022). Interagency Ecological Program and US Fish and Wildlife Service: San

Francisco Estuary Enhanced Delta Smelt Monitoring Program data, 2016-2021 ver

8. Environmental Data Initiative.

https://doi.org/10.6073/pasta/e1a540c161b7be56b941df50fd7b44c5 (Accessed

2022-07-26).

Wilcox, T. M., McKelvey, K. S., Young, M. K., Lowe, W. H., & Schwartz, M. K. (2015).

Environmental DNA particle size distribution from Brook Trout (Salvelinus

fontinalis). *Conservation Genetics Resources*, *7*(3), 639–641.

https://doi.org/10.1007/s12686-015-0465-z

Wilcox, T. M., McKelvey, K. S., Young, M. K., Sepulveda, A. J., Shepard, B. B., Jane, S.

F., … Schwartz, M. K. (2016). Understanding environmental DNA detection

probabilities: A case study using a stream-dwelling char Salvelinus fontinalis.

Biological Conservation, *194*, 209–216.

https://doi.org/10.1016/j.biocon.2015.12.023

Williams, K. E., Huyvaert, K. P., & Piaggio, A. J. (2017). Clearing muddied waters:

Capture of environmental DNA from turbid waters. *PLoS ONE, 12*(7), 1–17.

https://doi.org/10.1371/journal.pone.0179282

Author contributions

Conception and design: AH, MB, JR, BS, AF

Experiments and laboratory work: AH

Field sampling: AH, BM

Data analysis: AH with input from BM

Prepared first draft of manuscript: AH

Revised and approved: AH, MB, JR, BS, BM, AF

Table and Figure Legends

Table 1. Characteristics of the filters and prefilter used in the filtration experiment.

Table 2. Reagent volumes (a) and qPCR thermocycling protocol (b) for delta smelt

eDNA detection.

Table 3. Summary of detection rate and DNA copies detected for each treatment in the

filtration experiment (turbidity, prefilter, and amount of tank water added (mL)). Further

details in

Supplementary Data S4. GF, glass fiber filter; PC, polycarbonate filter; ST, Sterivex

PVDF filter.

Table 4. Summary of best models for success of delta smelt eDNA detection using (a)

eDNA copies and (b) detection/non-detection as the response variables. Results of all

models tested are provided Tables S6 and S7.

Figure 1. Upper San Francisco Estuary (California, USA) collection sites (inset) for

water used in filtration experiments. Non-turbid water (~5 NTU) was collected from the

upper Sacramento Deep Water Shipping Channel (38.5653, -121.5539) and turbid

water (~50 NTU) was collected from upper Prospect Slough (35.5299, -121.589),

adjacent to the shipping channel. (Map made using kepler.gl and mapbox.)

Figure 2. Schematic representation of filtration experiment. Three biological replicates

of 6 treatments were each filtered on 4 different filter types and assessed in 6 qPCR

replicates using a species-specific assay (Baerwald et al. 2011). GF, glass fiber filter;

PC, polycarbonate filter; ST, Sterivex cartridge filter.

Figure 3. Calibration curve showing Limit of Detection (LOD) for one replicate (2.47

copies) and **t**he Limit of Quantification (LOQ; 67 copies) for the delta smelt Taqman

assay (Baerwald et al. 2011). Calculations follow standard methods for validating eDNA

assays (Klymus et al. 2019; Merkes et al. 2019). Only points in the middle 2 quartiles of

standards with at least 50% detection (black circles) are included in the calculations.

Figure 4. Results of filtration experiments as detection/non-detection of qPCR

replicates (n=6) within 1-L biological replicates (n=3) for each treatment. Rows are

treatment (turbidity and prefiltration), columns are filter type, and amount of delta smelt

tank water added is within each box. GF, glass fiber filter; PC, polycarbonate filter; ST,

Sterivex cartridge filter.

Figure 5. Results of filtration experiments as eDNA copies in qPCR replicates (n=6)

within 1-L biological replicates (n=3) for each treatment. Rows are treatment (turbidity

and prefiltration), columns are filter type, and amount of delta smelt tank water added is

within each box. The black dotted line is the Limit of Quantification (LOQ) and the red

dashed line is the Limit of Detection (LOD). GF, glass fiber filter; PC, polycarbonate

filter; ST, Sterivex cartridge filter.

Filters	Abbreviation	Filter shape	Filter type	Pore size	Pore type
Sterivex polyvinylidene fluoride (PVDF; MilliporeSigma)	ST	cartridge	screen	0.45 µm	absolute
Glass fiber (Whatman)	GF	47mm diameter	depth	1.6 µm	nominal
Polycarbonate track-etched (MilliporeSigma)	PC	47mm diameter	screen	5 µm	absolute
Polycarbonate track-etched (MilliporeSigma)	PC	47mm diameter	screen	10 µm	absolute
Prefilter					
Nylon net (MilliporeSigma)	NN	47mm diameter	screen	20 µm	mesh

Reagent	Volume
Taqman Environmental Master Mix 2.0 (Applied Biosystems)	10 µl
Primer CytB-Htr-F (10 µM)	1.8 µl
Primer CytB-Htr-R (10 µM)	1.8 µl
Probe CytB-Htr-P	0.3 µl
DNA template	6.1 µl
Total	**20 µl**

Step	Time	Temperature	Cycles
Initial denaturation	10 min	95℃	1
Denaturation	15 sec	95℃	50
Annealing/Extension	1 min	63℃	

Treatment	Detection	DNA copies per reaction (6.1 µl template)					Approximate volume filtered (mL)	Mean DNA copies adjusted for volume filtered
Non-turbid	Positive qPCR replicates: 6 replicates x 3 bottles	Mean	Median	SD	Min	Max		
GF+500	18/18	233.4	203.7	94.3	141	489.8	1000	233.4
PC10+500	16/18	24.2	11.4	22.8	0	63.8	1000	24.2
PC5+500	18/18	68.6	60.3	46.1	9.6	161.4	1000	68.6
ST+500	18/18	106.6	92.7	98.9	26.8	476.8	500	213.2
GF+1000	18/18	324.7	310.1	111.9	128.4	560.4	1000	324.7
PC10+1000	17/18	53.3	48.3	34.9	0	135.5	1000	53.3
PC5+1000	18/18	106.4	103.5	40.4	58.8	187.1	1000	106.4
ST+1000	18/18	238.1	173.3	157.2	73.4	506.6	500	476.2
Turbid without prefilter								
GF+500	18/18	30.8	34.7	15.8	5.9	55	1000	30.8
PC10+500	8/18	3.3	0	4.7	0	15.7	450	7.3
PC5+500	10/18	5.8	3.9	6.6	0	20.2	200	29
ST+500	11/18	8.3	5	9.5	0	31	100	83
GF+1000	17/18	46.9	47.4	33.7	0	99.4	1000	46.9
PC10+1000	9/18	7.1	2.5	11.3	0	45.5	450	15.8
PC5+1000	5/18	5.5	0	10	0	31.5	200	27.5
ST+1000	12/18	8.4	6.5	8.3	0	22.5	100	84
Turbid with prefilter								
GF+500	9/18	23.2	3.3	34.9	0	94.2	1000	23.2
PC10+500	17/18	49.3	49	27.2	0	96.5	750	65.7
PC5+500	17/18	31.7	36.8	23.5	0	76.5	200	158.5
ST+500	17/18	21.9	20.4	12.5	0	41.4	150	146
GF+1000	9/18	29.3	19.9	33	0	98.7	1000	29.3
PC10+1000	17/18	38.8	29.1	25.7	8.3	80.1	750	51.7
PC5+1000	17/18	99.4	64.2	89.6	4.6	269.1	200	497
ST+1000	17/18	44.7	46.8	22.2	8.4	80.7	150	298

Fixed effects model structure for Log(eDNA Copies + 1)	ΔAICc	W$_i$	Cumulative W$_i$
Turbidity + Filter + Prefilter + Volume + Tank water added + Filter*Turbidity + Filter*Prefilter	0	0.4974	0.4974
Turbidity + Filter + Prefilter + Tank water added + Filter*Turbidity + Filter*Prefilter	0	0.4974	0.9948

Fixed effects model structure for Detection/Non-detection	ΔAICc	W$_i$	Cumulative W$_i$
Turbidity + Filter + Prefilter + Volume + Filter*Prefilter	0	0.381	0.381
Filter + Prefilter + Volume + Tank water added + Filter*Prefilter	0.1	0.356	0.737
Turbidity + Filter + Prefilter + Volume + Tank water added + Filter*Prefilter	1.7	0.16	0.897
Turbidity + Filter + Prefilter + Volume + Filter*Prefilter + Filter*Turbidity	3.8	0.056	0.953
Turbidity + Filter + Prefilter + Volume + Tank water added + Filter*Prefilter + Filter*Turbidity	5.6	0.023	0.976
Turbidity + Filter + Prefilter + Tank water added + Filter*Prefilter + Filter*Turbidity	5.6	0.023	0.999

Standard curve for: DSM

R-squared: 0.9971
y = -3.4368x + 41.7245