

1 **Unravelling drivers of local adaptation through Evolutionary Functional-**

2 **Structural Plant modelling**

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## 10 Summary

- 11 1. Local adaptation to contrasting environmental conditions along environmental  
12 gradients is a widespread phenomenon in plant populations, yet we lack a mechanistic  
13 understanding of how individual agents of selection contribute to local adaptation.
- 14 2. Here, we developed a novel evolutionary functional-structural plant (E-FSP) model that  
15 simulates local adaptation of virtual plants along an environmental gradient. First, we  
16 validate the model by testing if it can recreate two elevational ecotypes of *Dianthus*  
17 *carthusianorum* occurring in the Swiss Alps. Second, we use the E-FSP model to  
18 disentangle the relative contribution of abiotic (temperature) and biotic (competition  
19 and pollination) selection pressures to elevational adaptation in *D. carthusianorum*.
- 20 3. The model reproduced the qualitative differences between the elevational ecotypes in  
21 two phenological (germination and flowering time) and one morphological trait (stalk  
22 height), as well as qualitative differences in four performance variables that emerge  
23 from GxE interactions (flowering time, number of stalks, rosette area and seed  
24 production). Our results suggest that elevational adaptation in *D. carthusianorum* is  
25 predominantly driven by the abiotic environment.
- 26 4. Our approach shows how E-FSP models incorporating physiological, ecological and  
27 evolutionary mechanisms can be used in combination with experiments to examine  
28 hypotheses about patterns of adaptation observed in the field.
- 29

## 30 **Introduction**

31 Local adaptation to contrasting environmental conditions is a widespread phenomenon in plant  
32 populations (Leimu & Fischer, 2008), resulting from divergent selection pressures imposed by  
33 variation in environmental conditions on populations occurring across a species' range. The  
34 outcome of local adaptation can be documented in field experiments that assess the  
35 performance of alternative ecotypes in contrasting environments, where local ecotypes are  
36 expected to outperform foreign ecotypes (Kawecki & Ebert, 2004). A wealth of experimental  
37 work has shown the pervasiveness of local adaptation (Leimu & Fischer, 2008), yet we often  
38 lack a mechanistic understanding of how individual agents of selection contribute to local  
39 adaptation along environmental gradients (Wadgymar *et al.*, 2017). This is caused by  
40 individual drivers of selection acting on multiple plant traits, either directly or indirectly, and  
41 by individual traits being affected by multiple drivers of selection. The interactions between  
42 drivers of selection, such as between abiotic and biotic factors (Briscoe Runquist *et al.*, 2020;  
43 Hargreaves *et al.*, 2020; Paquette & Hargreaves, 2021), further complicates disentangling the  
44 role of any individual driver in shaping local adaptation.

45 To address this, experimental studies may be complemented by mechanistic modelling  
46 approaches (Connolly *et al.*, 2017) that incorporate the eco-physiological and eco-evolutionary  
47 processes driving local adaptation. Such mechanistic modelling approaches are more  
48 commonly used in crop breeding and have proven to be powerful tools to explore the  
49 mechanistic basis of plant-environment interactions (Hammer *et al.*, 2006). However, the  
50 potential for these models to simulate the ecological complexity that drives local adaptation in  
51 natural plant communities currently remains unexplored.

52 Function-structural plant (FSP) modelling is such a mechanistic modelling approach  
53 that integrates an explicit representation of plant structure in a 3D environment, combined with  
54 functional plant responses to that environment (Evers *et al.*, 2018). The approach is particularly

55 suited to the simulation of plant-plant interactions as it explicitly simulates the spatial  
56 heterogeneity that is inherent to species mixtures and drives competitive interactions between  
57 plants (Evers *et al.*, 2019; Bongers, 2020). This explicit representation of plant form and  
58 function makes FPS modelling an excellent tool to test hypotheses about the adaptive value of  
59 functional traits in a dynamic ecological context (Bongers *et al.*, 2019; de Vries *et al.*, 2019;  
60 Douma *et al.*, 2019).

61         A novel and largely unexplored application of FSP modelling is in combination with a  
62 mechanistic model of natural selection (de Vries, 2021). Such evolutionary-FSP (E-FSP)  
63 models can simulate the combined selection pressure imposed by multiple selective agents on  
64 a population of individually distinct plants that interact with each other and with the  
65 environment (de Vries, 2021). This individual-based perspective is of particular importance to  
66 mechanistically simulate natural selection, as key mechanisms that drive selection (e.g.  
67 competition for resources) are not only driven by absolute trait values, but also by trait values  
68 relative to those of neighbouring plants (Falster & Westoby, 2003; McNickle & Dybzinski,  
69 2013). This is exemplified by competition for light, which is a pre-emptive resource (i.e. light  
70 interception by one plant also prevents light interception by other plants) whose acquisition is  
71 dependent on the height of a plant relative to the height of the surrounding vegetation, leading  
72 to competitive asymmetry (Weiner, 1990). Despite the potential for E-FSP models to simulate  
73 the mechanisms that drive local adaptation, the complexity of these mechanisms makes  
74 validation of such a model particularly challenging. As such, all E-FSP models published to  
75 date have been theoretical exercises (Renton & Poot, 2014; Yoshinaka *et al.*, 2018; de Vries *et*  
76 *al.*, 2020).

77         Here, we develop, parameterise, calibrate and validate an E-FSP model of local  
78 adaptation along an environmental gradient. As a case study, we use two elevational ecotypes  
79 of *Dianthus carthusianorum* that occur along an elevational gradient in the Swiss Alps,

80 growing at low (~1000 m.a.s.l.) and high (~2000 m.a.s.l.) elevation. These environments are  
81 characterised by commonly reported differences in (a)biotic conditions along elevational  
82 gradients (Halbritter *et al.*, 2018), resulting in a tall grassland vegetation at lower elevations,  
83 and typical alpine (i.e. shorter) vegetation at higher elevations. Elevational ecotypes of *D.*  
84 *carthusianorum* are adapted to their elevational ranges and display genetically based  
85 phenotypic divergence in phenological and morphological traits (Walther, 2020; Pålsson *et al.*,  
86 *in prep.*). The high elevation populations of *D. carthusianorum* typically exhibit lower  
87 biomass, flower earlier and produce smaller flowering stalks compared to their low elevation  
88 counterparts. Favoured by a higher energy input environment, the latter achieve larger plant  
89 size and taller flowering stalks to potentially compete for light and pollinators with the  
90 surrounding vegetation. Despite sound evidence of adaptation along an elevational gradient,  
91 the selection pressures underlying the evolution of these elevational ecotypes remains  
92 unknown. Commonly reported patterns of adaptation along elevational gradients suggest that  
93 the divergence in *D. carthusianorum* is driven by more stressful abiotic conditions at high  
94 elevations (e.g. temperature), and by biotic interactions (e.g. competition and pollination) at  
95 low elevations (Halbritter *et al.*, 2018). First, we aim to validate the E-FSP model by asking  
96 whether the E-FSP model can recreate elevational ecotypes of *D. carthusianorum*. Second, we  
97 hypothesise that temperature, competition and pollination are key agents of selection and use  
98 the E-FSP model to disentangle their relative contribution to elevational adaptation in *D.*  
99 *carthusianorum*.

## 100 **Methods**

101 *Model species: D. carthusianorum*

102 *D. carthusianorum* is a primarily outcrossing gynodioecious, perennial herb that is native to  
103 Europe and is widespread on rocky slopes and dry grasslands throughout the Alps up to an  
104 elevation of 2500 meters (Bloch *et al.*, 2006). For model parameterisation, calibration and

105 validation, we used data from two elevational ecotypes of *D. carthusianorum*. These were  
106 grown in a reciprocal transplant experiment established in fall 2015 that included two replicate  
107 transplant sites at low (~1000 m.a.s.l.) and high (~2000 m.a.s.l.) elevation, respectively. We  
108 used data collected in the first growing season on fitness components (survival, flowering  
109 probability and seed count), morphological (stalks height, number of stalks, stalk leaves and  
110 flowers) and phenological (flowering time) traits (for details see: Walther, 2020; Pålsson *et al.*,  
111 *in prep.*).

### 112 *Model summary*

113 The model used in this study is based on the E-FSP model described in de Vries *et al.* (2020),  
114 which was developed in the modelling platform GroIMP (Hemmerling *et al.*, 2008) and  
115 designed to simulate adaptation to abiotic (nitrogen) and biotic (competition and herbivory)  
116 agents. Here, we expand this E-FSP model by including temperature driven plant phenology  
117 and plant-pollinator interactions. The model simulates a population of competing plants over  
118 multiple generations, with the performance of individual plants within a generation being  
119 determined by three plant traits that are subject to selection: germination time (*GM*), time to  
120 flowering (*TF*), and stalk height (*SH*). We assumed that these traits are not genetically linked  
121 so that there are no pleiotropic effects between them, and therefore the model is theoretically  
122 able to select for any combination of trait values. We simulate three environmental factors that  
123 determine plant fitness and thereby impose selection pressure; i) the difference in abiotic  
124 conditions associated with an elevational gradient (i.e. temperature and subsequently also  
125 season length and nitrogen availability), ii) interspecific competition with the surrounding  
126 vegetation, and iii) pollinator density. The model structure is summarised below (also see Fig.  
127 1). For a detailed model description, see Methods S1, which includes a list of indices used in  
128 the model description (Table S1).

### 129 *Temperature and plant development*

130 The model calculates daily average and minimum temperature as a function of elevation, based  
131 on climate data collected by weather stations at the field sites (Fig. S1, Table S2). Temperature  
132 is used to drive plant phenology (McMaster & Wilhelm, 1997), photosynthesis (Farquhar *et*  
133 *al.*, 1980), soil nitrogen availability (Rodrigo *et al.*, 1997; Kirschbaum, 2000; Guntiñas *et al.*,  
134 2012), and to calculate frost damage (Ji *et al.*, 2015).

135 Plant development is split into two stages; a vegetative and a generative stage. The *in silico*  
136 plants germinate in spring, the timing of which is a function of temperature and their  
137 germination trait (GM). During its vegetative stage, the plant invests all accumulated  
138 assimilates and nitrogen towards the growth of rosette leaves and roots. The transition to the  
139 generative stage is dependent on both the time to flowering trait (TF), and cumulative  
140 temperature time (growing degree days, McMaster & Wilhelm, 1997). In the generative stage,  
141 the plant continues to intercept light and produce assimilates through photosynthesis, but no  
142 longer grows new rosette leaves or roots. Instead, newly acquired assimilates and nitrogen are  
143 allocated to flowering stalks, stalk leaves and seed filling.

#### 144 *Plant architecture and resource capture*

145 The model uses an explicit description of plant architecture to mechanistically simulate  
146 competition for the three resources incorporated in the model; light, nitrogen and pollinators.  
147 The vegetative shoot is represented by a rosette of rectangular leaves (Fig. S5, Fig. S8) that  
148 photosynthesise based on leaf level light interception. The flowering stalks are described as a  
149 cylinder with a number of short stalk leaves that also add to assimilate production through  
150 photosynthesis and a disk at the top of the stalk that represents the flowerhead and attracts  
151 pollinators. The stalk height trait determines the position of the flower, the number of stalk  
152 leaves and the stalk diameter required to support the stalk (Fig. S7). The explicit representation  
153 of these aboveground plant parts allows for the calculation of light interception in the canopy  
154 using a Monte-Carlo ray tracer and, therefore, the outcome of competition for light between

155 individual plants (Hemmerling *et al.*, 2008; Evers *et al.*, 2010). This methodology has proven  
156 to capture the asymmetry in competition for light (de Vries *et al.*, 2018; de Vries *et al.*, 2019),  
157 making it a key model component to simulate the effect of competitive interactions on plant  
158 fitness and subsequent selection. The root architecture is described as a conical volume (Fig.  
159 S2, Table S3) from which the plant can take up nitrogen, such that a larger root system  
160 proportionally increases the potential nitrogen uptake of the plant, thus resulting in symmetric  
161 competition for nitrogen.

### 162 *Plant growth*

163 We assumed that the C:N ratio of plant tissue is conserved, so that plant growth is either limited  
164 by the plant's ability to intercept light and assimilate carbon through photosynthesis, or its  
165 ability to take up nitrogen through the root system. Photosynthesis is calculated at the leaf level  
166 using a temperature driven Farquhar, von Caemmerer and Berry photosynthesis model  
167 (Farquhar *et al.*, 1980; Yin & Struik, 2009; Yin *et al.*, 2009). The assimilates produced by the  
168 leaves are first used to pay for maintenance respiration, which is based on plant nitrogen content  
169 (Ryan, 1991), after which the remaining assimilates are allocated to growth. The potential  
170 nitrogen uptake by the root system is modelled as a function of rooting volume and soil nitrogen  
171 availability, and can be supplemented by re-allocation of nitrogen from the leaves, which is  
172 used to simulate leaf senescence at the end of the growing season (Yin & van Laar, 2005). In  
173 the vegetative stage, assimilates and nitrogen are equally allocated towards the growth of  
174 rosette leaves and roots (i.e. assuming a root:leaf ratio of 1). In the generative stage, assimilates  
175 and nitrogen are allocated to flowering stalks and seed set using a hierarchical allocation model  
176 that prioritises filling pollinated seeds (see *Pollination* below) over growing new stalks  
177 (Minchin & Thorpe, 1996).

### 178 *Pollination*

179 The number of pollinator visits is simulated as a function of flower density following the



180 correlation found by Richman *et al.* (2020). Pollination in grasslands is known to scale with  
181 stalk height in relation to the height of the surrounding vegetation (Sletvold *et al.*, 2013;  
182 Slaviero *et al.*, 2016). To simulate this, we visualise the flower heads as upwards facing disks  
183 with a diameter of 2 cm, and use the light absorbed by the flowerhead as a proxy for flower  
184 attractiveness so that flowers reaching to the top of the vegetation are the most attractive to  
185 pollinators. The pollinator visits are then distributed over the flowers in the plot based on their  
186 relative attractiveness. The relationship between the number of pollen visitations and potential  
187 seed set is based on a previous study of the pollination of *D. carthusianorum* that was  
188 conducted in the same study area (Bloch *et al.*, 2006, also see Table S4 and Fig. S4).

#### 189 *Evolutionary algorithm*

190 *D. carthusianorum* is a perennial species that has a lifespan from one to several years. However,  
191 implementation of the perennial life-cycle in the model has proven challenging because we  
192 currently lack long-term data for calibration and validation. Although life history traits linked  
193 to the perennial life-cycle almost certainly contribute to adaptation in *D. carthusianorum*,  
194 phenotypic divergence and the resulting differences in performance are already apparent in the  
195 first year (Pålsson *et al.*, *in prep.*). Therefore, we opted to simulate an annual life-cycle and  
196 assume that selection acting on the first year of plant reproduction is sufficient to explain the  
197 evolution of the traits under investigation.

198 Plant fitness is composed of female and male reproductive success: female reproductive  
199 success is defined as the total number of filled seeds at the end of the growing season (i.e.  
200 fecundity), and male reproductive success is defined as the number of pollinator visits to the  
201 flowers over the course of the growing season. At the end of every generation, the model  
202 randomly selects 100 plants based on their realised female reproductive success, and 100 plants  
203 based on their male reproductive success, and recombines these to generate the 100 offspring  
204 genotypes that populate the subsequent generation of plants. In this process, the model allows

205 for a single plant to contribute multiple seeds and/or pollen to the next generation, and we  
206 assume that seeds can only germinate in the following year, so no seed bank is built up. The  
207 virtual plants were not able to self-pollinate, as selfing in *D. carthusianorum* is prevented by  
208 protandry and rarely leads to fruit formation (Bloch *et al.*, 2006). The traits of an offspring  
209 plant ( $T_o$ , dimensionless, ranging from zero to one) are randomly inherited from either of the  
210 two parental plants, effectively simulating a haploid system with traits completely determined  
211 by their genetic basis. Offspring trait values are assumed to be normally distributed around the  
212 parental trait value (mean =  $T_p$ , dimensionless, ranging from zero to one, standard deviation =  
213  $Tsd$ ).

#### 214 *Plant density and interspecific competition*

215 The model starts each generation with 100 individuals of the virtual *D. carthusianorum* plants  
216 that were randomly placed in a plot of 1 m<sup>2</sup>, meaning that individual plants can experience  
217 different levels of competition dependent on neighbour proximity. While initial plot level plant  
218 density is kept constant, the model does allow plant density to vary during the season as a result  
219 of mortality caused by frost damage or resource limitation.

220 To simulate the effect of interspecific competition for light and nutrients, particularly  
221 with the tall grasses that *D. carthusianorum* typically competes with in their low elevation  
222 habitats, we introduce 100 individuals of a second plant species designed to represent these tall  
223 grasses. This increases the initial plant density of the population from 100 plants m<sup>-2</sup> in the  
224 absence of the grass species to 200 plants m<sup>-2</sup> in its presence. The growth and development of  
225 these grasses is not simulated mechanistically, but rather described by a sigmoid function that  
226 calculates biomass as a function of temperature and time (see Methods S1). The grasses can  
227 therefore be seen as a static environmental factor that imposes competition pressure on the  
228 virtual *D. carthusianorum* plants, but is not affected by the *D. carthusianorum* plants.

229 *Model output*

230 A single simulation consists of 125 generations, by which time the simulated population had  
231 settled at an optimum through natural selection (Fig. S9). To account for random fluctuations  
232 between generations, model output was recorded at generation 105, 110, 115, 120 and 125.  
233 Model output was recorded at the end of the growing season on the level of individual plants  
234 and consisted of values for the three plant traits under selection, as well as flowering time,  
235 rosette area, the number of stalks, and fitness. We conducted no statistical analyses on the *in*  
236 *silico* data, because the sample size is so high that all treatment combinations show a high  
237 statistical significance, even if the differences between those treatment combinations are not  
238 biologically relevant. In the text, values are reported as mean  $\pm$  standard error.

239 *Model parameterisation, calibration and validation*

240 To parameterise, calibrate and validate the model, we used data collected during the first  
241 growing season (2016) of the field experiment described in Pålsson *et al.* (*in prep.*) and  
242 conducted a climate chamber experiment to measure germination times. For model  
243 parameterisation, we obtained empirical estimates of parameter values that are not readily  
244 available in published literature and that we assume to be shared between the elevational  
245 ecotypes (for a full list of parameters, see Table S4). For model validation of traits that  
246 differentiate the two elevational ecotypes, we obtained empirical estimates growing from a  
247 controlled environment (a common garden at the low elevation site for TF, SH, and two climate  
248 chambers for GM). Additionally, we obtained empirical estimates of response variables  
249 (flowering time, number of stalks, rosette area, and fitness) from the two elevational ecotypes  
250 growing in the field. For the germination experiment, we collected seeds from seven  
251 individuals from a low and a high elevation population occurring in close proximity (<1km) to  
252 the transplant sites. We vernalized the seeds for a week at -18 °C and sowed 20 seeds from each  
253 individual in compartmentalised trays with 21 cm<sup>3</sup> soil per compartment, sowing one seed per

254 compartment. These trays were placed in one of two climate chambers under a constant  
255 temperature of either 4 °C or 20 °C to measure the effect of temperature on germination rates,  
256 and a 16/8 day/night cycle. We followed a balanced randomised design for this experiment, so  
257 that each seed family was equally represented in each treatment. Thrice weekly, the seeds were  
258 watered and germination was recorded over a period of 30 days.

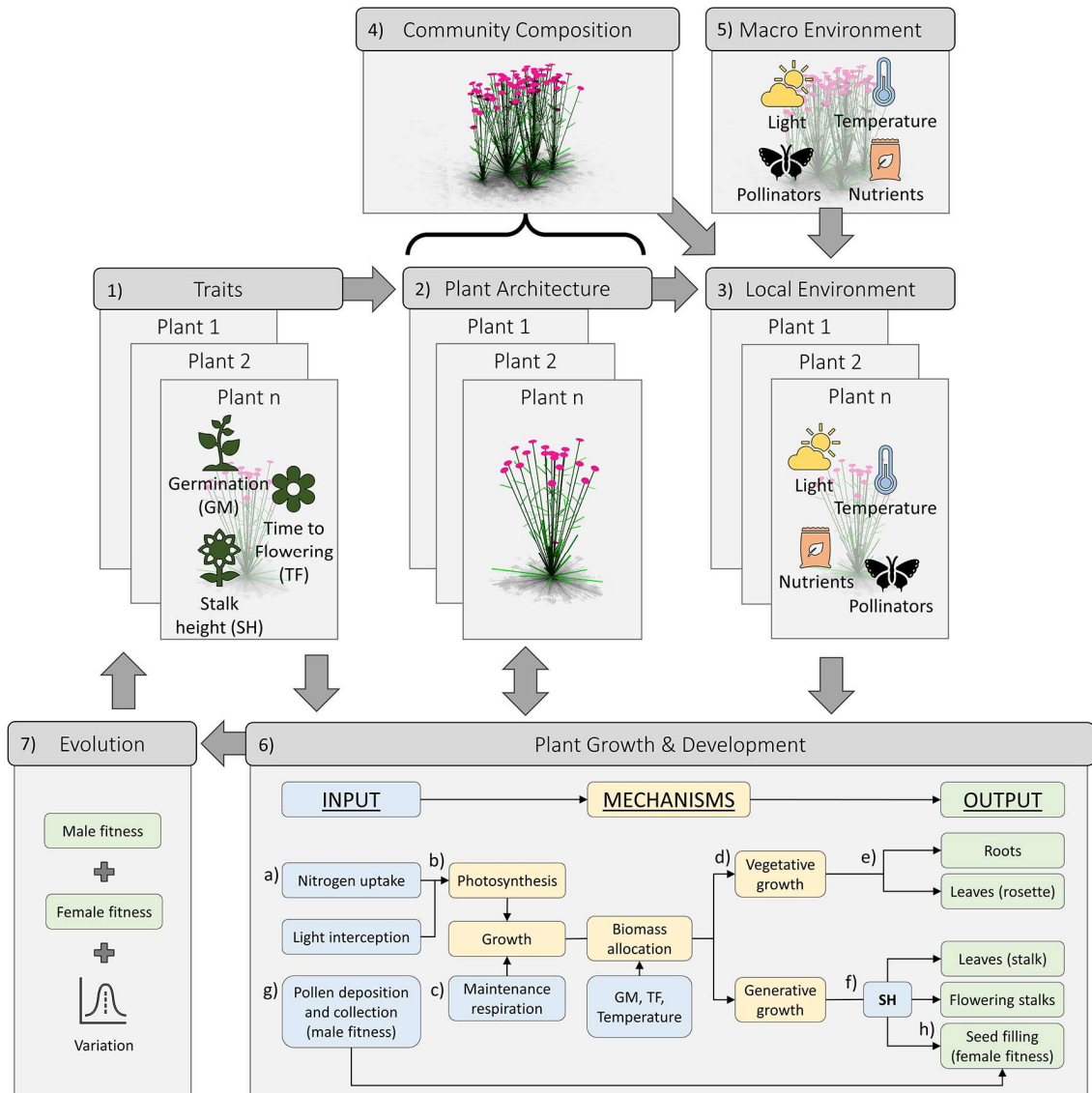
### 259 *Simulations*

260 The model incorporated three environmental factors; the difference in abiotic conditions  
261 associated with a change in **elevation** (i.e. temperature and subsequently also season length  
262 and nitrogen availability), interspecific competition with a tall grass species (hereafter named  
263 “**competition**”), and pollinator density (hereafter named “**pollination**”, see Table 1). The  
264 elevation treatments are 1000 or 2000 m, based on the elevation of the experimental sites. The  
265 competition treatments reflect the more intense competition that is generally found in low  
266 elevation habitats (Halbritter *et al.*, 2018) by simulating both intra- and interspecific  
267 competition (i.e. 100 plants m<sup>-2</sup> of *D. carthusianorum* and 100 plants m<sup>-2</sup> of a tall grass), and  
268 the shorter vegetation that is generally found in high elevation habitats (Halbritter *et al.*, 2018)  
269 by simulating only intraspecific competition (i.e. only 100 plants m<sup>-2</sup> of *D. carthusianorum*).  
270 The pollination treatments represent pollinator densities along an elevational gradient in the  
271 Swiss Alps (Richman *et al.*, 2020), with pollinators being more abundant in the low elevation  
272 habitat (0.3 pollinator visits flower<sup>-1</sup> h<sup>-1</sup>) compared to the high elevation habitat (0.03 visits  
273 flower<sup>-1</sup> h<sup>-1</sup>).

274 To validate model performance, we simulated selection in two scenarios that represent  
275 the low and high elevation habitats (**Low habitat**: 1000m elevation, 100 interspecific  
276 competitors m<sup>-2</sup>, and 0.3 pollinator visit flower<sup>-1</sup> h<sup>-1</sup>; **High habitat**: 2000m elevation, 0  
277 interspecific competitors m<sup>-2</sup>, and 0.03 pollinator visit flower<sup>-1</sup> h<sup>-1</sup>), and compared the trait  
278 variation and performance of *in silico* populations after 125 generations of selection to the trait

279 variation and performance of *in vivo* ecotypes of *D. carthusianorum* from the low and high  
280 elevation sites. To test whether the *in silico* populations of *D. carthusianorum* could be  
281 considered locally adapted, we simulated a virtual transplant experiment: 50-50 mixtures  
282 consisting of plants originating from the low and high elevation populations were grown in the  
283 low and high elevation habitats, and their seed production after one generation was used as a  
284 fitness proxy. Under the basic principles of testing adaptation in reciprocal transplant  
285 experiments, genotype x environment (GxE) interactions should result in locally adapted  
286 populations outperforming foreign populations growing under the same conditions (Blanquart  
287 *et al.*, 2013; Hargreaves & Eckert, 2019; Hargreaves *et al.*, 2020).

288 To elucidate how the different abiotic and biotic selection pressures contributed to the local  
289 adaptation of *D. carthusianorum*, we first simulated populations under control conditions (i.e.  
290 1000 m, 0 interspecific competitors m<sup>-2</sup>, 0.3 pollinator visit flower<sup>-1</sup> h<sup>-1</sup>), and then changed each  
291 environmental factor individually to assess their effects on trait variation and performance.



292

293 **Fig 1.** Visual summary of the E-FSP model used in this study. The model represents a population of individual  
 294 plants, each with their own trait values (1; Germination (GM); Time to Flowering (TF); Stalk Height (SH)), plant  
 295 architecture (2) and local environment (3; Light, nutrients, pollinators and temperature). The local environment is  
 296 dependent on the composition of the surrounding plant community (4; i.e. intra- and interspecific competition),  
 297 and the macro environment (5). To simulate plant growth and development (6), the model takes input on the level  
 298 of the individual plants (i.e. from 1,2 and 3). First, light interception and nutrient uptake (a) are used to calculate  
 299 leaf level photosynthesis (b). The respiration required to maintain the standing biomass (c) is subtracted to get the  
 300 net growth rate. These assimilates are allocated to either vegetative or generative growth (d), dependent on the  
 301 temperature and the GM and TF traits. During vegetative growth, the plant allocates assimilates to roots and  
 302 rosette leaves (e). During generative growth, the plant allocates assimilates to stalk leaves, flowering stalks and

303 seed filling, with the SH trait determining the allocation between these three (**f**). Finally, through recombination,  
304 the pollen collected by pollinators (**g**; male fitness) and the filled seeds from pollinated flowers (**h**; female fitness)  
305 determine the traits of the plants in next generation, and thus drive evolution through natural selection (**7**).

## 306 **Results**

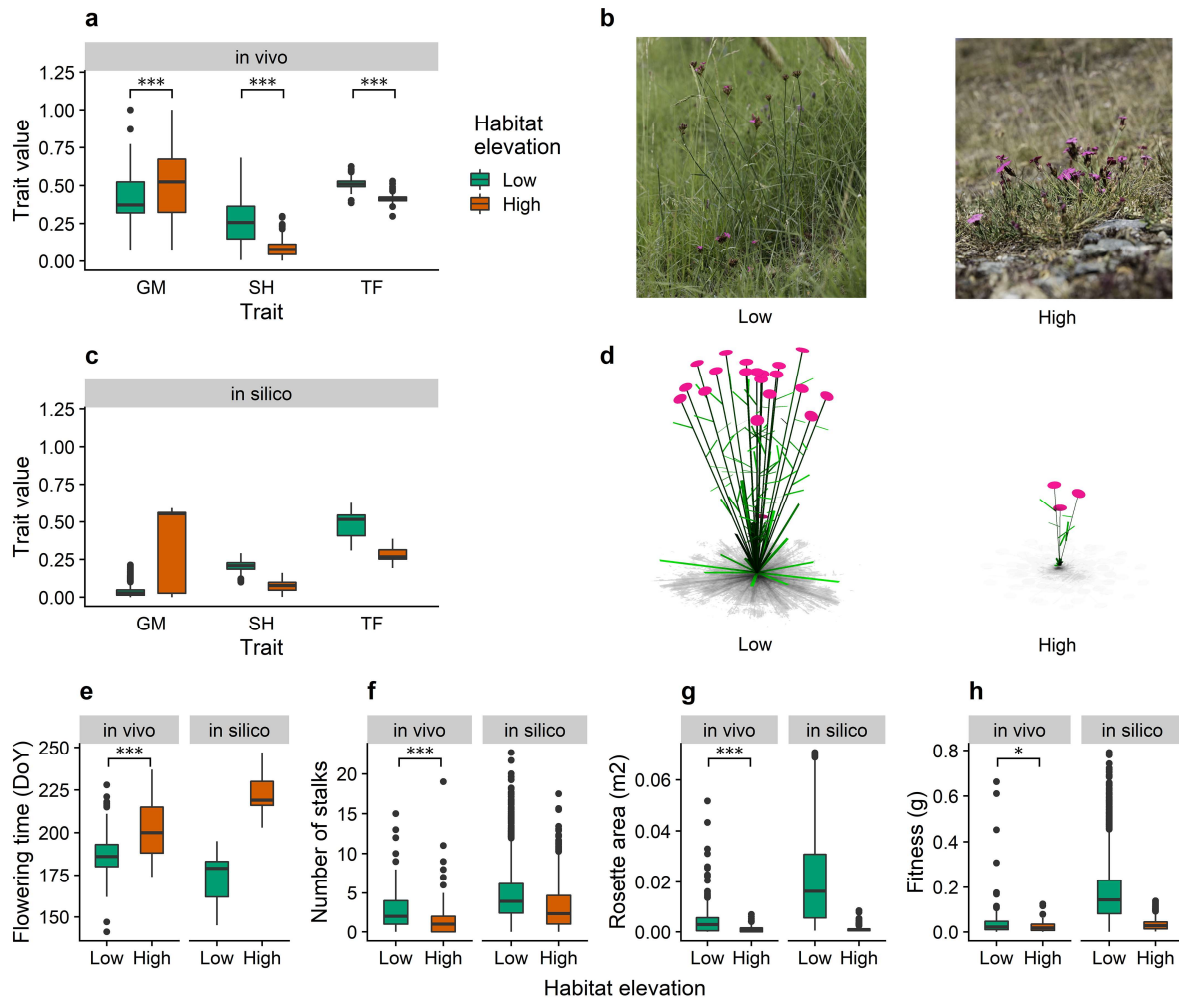
### 307 *Simulation of elevational ecotypes: growing in a common garden*

308 To validate whether the model was able to recreate the elevational ecotypes of *D.*  
309 *carthusianorum*, we compared the trait values of *in silico* populations to *in vivo* measurements  
310 of the low and high elevation ecotypes conducted on plants growing in a shared environment  
311 (i.e. the climate chamber for GM, and the low elevation habitat for SH and TF). The *in vivo*  
312 low and high elevation populations of *D. carthusianorum* expressed significant differences in  
313 each of the tree selected traits (Fig. 2a; Table S5; *in vivo*). Growing in the climate chamber,  
314 plants from the high elevation population germinated later compared to plants from the low  
315 elevation population ( $P < 0.001$ ; Fig. S6). Growing in the low elevation site, plants from the  
316 high elevation population had a shorter stalk height ( $P < 0.001$ ), and a shorter time to flowering  
317 ( $P < 0.001$ ) compared to plants from the low elevation population. The *in silico* populations  
318 showed patterns of selection that were equal to the qualitative differences in the *in vivo*  
319 populations (Fig. 2a; Table S5).

### 320 *Simulation of elevational ecotypes: growing in their home environment*

321 To validate the model's ability to capture GxE interactions, we compared the performance of  
322 *in vivo* and *in silico* populations using four performance measures that are determined by both  
323 the plant's trait values and its local environment. From here onwards, we will use the term  
324 'home environment' in relation to a plant population to refer to the environment in which  
325 selection took place. Growing in their respective home environments, the *in vivo* results show  
326 that plants from the high elevation populations flowered significantly later (Fig. 2e; Table S5;  
327  $P < 0.001$ ), produced fewer flowering stalks (Fig. 2f; Table S5;  $P < 0.001$ ), a smaller rosette area

328 (Fig. 2g; Table S5;  $P < 0.001$ ), and lower seed production (Fig. 2h; Table S5;  $P < 0.001$ )  
329 compared to plants from the low elevation populations. Again, the *in silico* results matched the  
330 qualitative patterns of the *in vivo* results (Fig. 2b,c,d,e; Table S5).



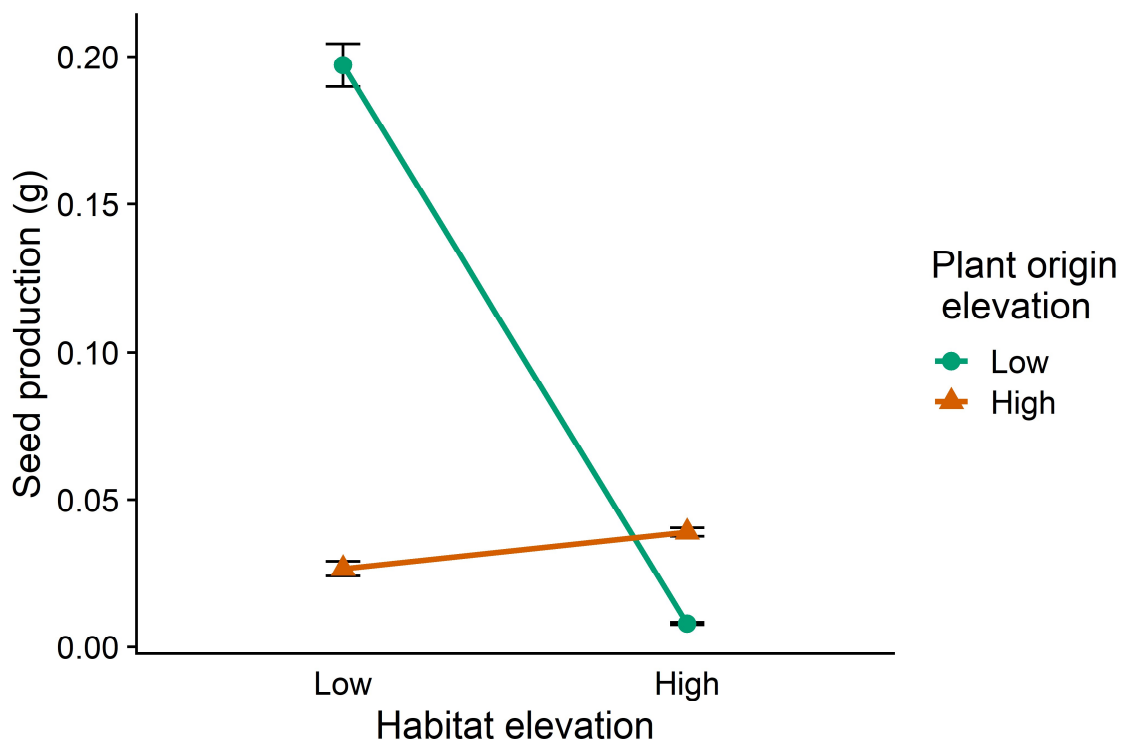
331

332 **Fig. 2.** Comparison of trait variation and performance of *in vivo* and *in silico* populations of *D. carthusianorum*  
333 in their low and high elevation habitats (Low: Green, High: Red). Trait variation (y-axis: normalized trait value  
334 (0-1)) of germination (GM), stalk height (SH) and time to flowering (TF) of *in vivo* (a,b) and *in silico* (c,d)  
335 populations subjected to selection in the low and high elevation habitats. The other panels show the variation in  
336 flowering time (e, DoY), number of stalks (f), rosette area (g, m<sup>2</sup>) and fitness (h, g) of *in vivo* and *in silico*  
337 populations growing in their home environments. Significance is shown only for the measured data (\*  $P < 0.05$ ; \*\*  
338  $P < 0.01$ ; \*\*\*  $P < 0.001$ ).



339 *Simulation of elevational ecotypes: local adaptation*

340 To test GxE interactions indicative of adaptation in the *in silico* plants, we grew 50:50 mixtures  
341 of low and high elevation genotypes under alternative environments. Plants growing in their  
342 home environment were able to outcompete the plants originating from the foreign population  
343 (Fig. 3). In the low elevation habitat, the plants originating from the low elevation habitat  
344 produced more seeds ( $0.197 \pm 0.007$  g) than the plants originating from the high elevation  
345 habitat ( $0.026 \pm 0.045$  g). In the high elevation habitat, the performance of the plants originating  
346 from the low elevation habitat saw a major decrease, resulting in them producing fewer seeds  
347 ( $0.008 \pm 0.0015$  g) than the plants originating from the high elevation habitat ( $0.039 \pm 0.0026$  g).  
348 These results fulfil both the local vs foreign and home vs away criteria forming the hallmarks  
349 of local adaptation (Kawecki & Ebert, 2004; Savolainen *et al.*, 2013).

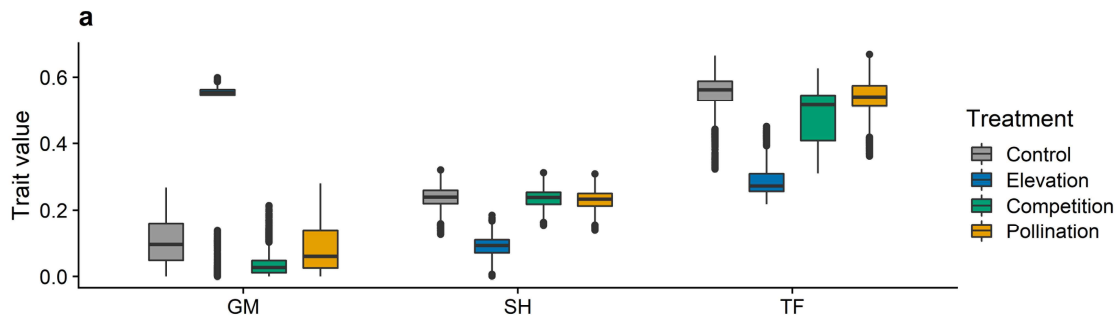


350

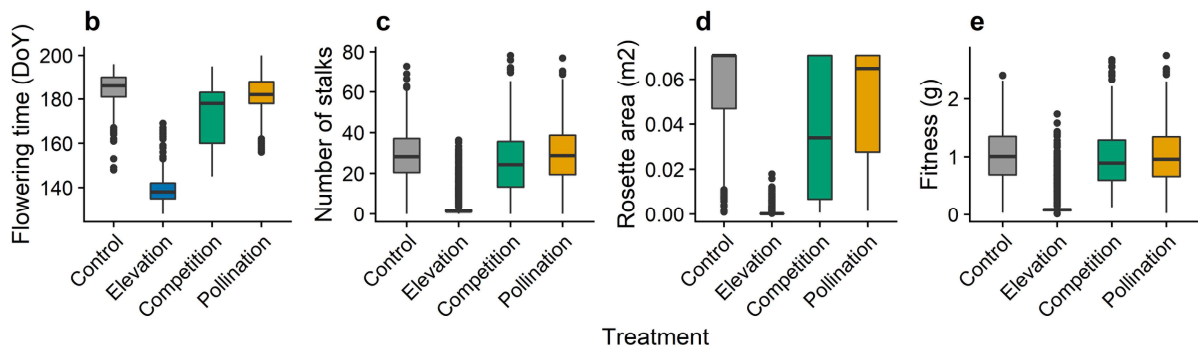
351 **Fig. 3.** Seed production of a virtual transplant experiment in the low and high elevation habitats. We simulated  
352 plant populations consisting of a 50:50 mixture of plants originating from the *in silico* low (green) and high (red)  
353 elevation habitats competing in either the low and high elevation habitats. Error bars show the standard error of

354 the mean.

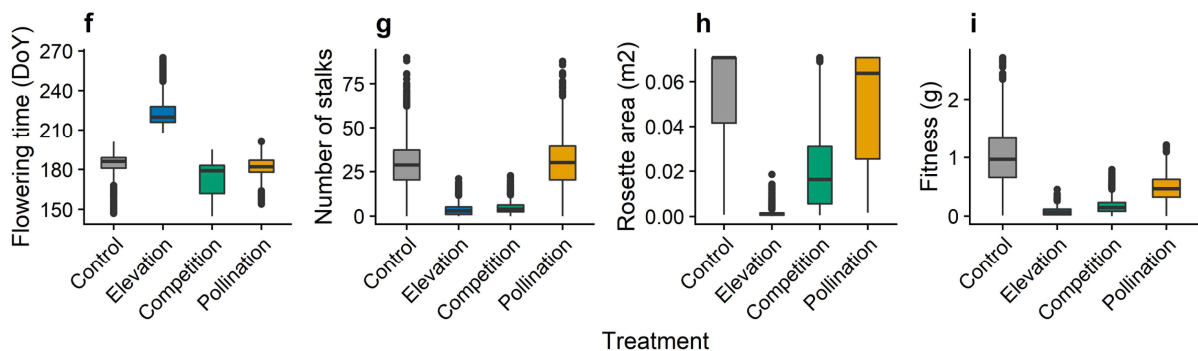
## Plant Traits



## Performance in the control environment



## Performance in the treatment environments



355

356 **Fig. 4.** Trait selection on *in silico* populations of *D. carthusianorum*. Panel a shows the trait values of three plant  
357 traits (a; normalized trait value (0-1); germination (GM), stalk height (SH) and time to flowering (TF)) resulting  
358 from selection imposed by treatments where we varied individual environmental factors (**Control**, grey; changes  
359 in abiotic conditions associated with an increased **Elevation**, blue; increased **Competition**, green; or decreased  
360 **Pollination**, yellow). Panels b-e show the performance of the adapted plant populations in the control environment

361 to show the direct effect of the trait changes on plant performance. Panels **f-i** show the performance of the adapted  
362 plant populations in their respective home environments (i.e. the environment in which selection took place),  
363 which shows the combined effect of changes in environment and plant traits on plant performance.

#### 364 *Disentangling the role of individual selection pressures: Elevation*

365 Increasing elevation from 1000m to 2000m decreased daily average temperature by  $\sim 5^{\circ}\text{C}$ ,  
366 which decreased the season length by 108 days (assuming a base temperature of zero) and  
367 decreased nitrogen mineralisation over the year from  $329 \text{ g N m}^{-2}$  to  $182 \text{ g N m}^{-2}$ . Additionally,  
368 this increase in elevation increased the variation in minimum temperature (eq. S30), increasing  
369 the frequency and strength of freezing events that potentially lead to frost damage. These  
370 changes in the environment led to selection for plants that germinated later compared to the  
371 control treatment (Fig. 4a, GM), which allowed the plants to escape the increased risk of frost  
372 damage early in the season. These environmental changes also selected for shorter flowering  
373 stalks (Fig. 4a, SH), and earlier flowering time compared to the control treatment (Fig. 4a, TF).  
374 These changes in plant traits led to plants that, following adaptation to high elevation and  
375 grown in a control environment, flowered earlier compared to plants from the control treatment  
376 (Fig. 4b), and also produced fewer stalks (Fig. 4c), a smaller rosette (Fig. 4d) and lower fitness  
377 (Fig. 4e). However, when growing in their home environment, these plants still flowered later  
378 than the control plants in the control treatment because of the late start of the season at high  
379 elevation (Fig. 4f). The decrease in temperature associated with the increase in elevation led to  
380 a decrease in productivity through a decrease in photosynthetic rates, a shorter growing season  
381 and lower nitrogen availability. This lower productivity in combination with the trait changes  
382 led to the plants having smaller rosettes (Fig. 4g), fewer flowering stalks (Fig. 4h), and lower  
383 seed production (Fig. 4i).

#### 384 *Disentangling the role of individual selection pressures: Interspecific competition*

385 In the competition treatment, the *in silico D. carthusianorum* plants competed for light and

386 nitrogen with an equal density of a tall grass species. This interspecific competition selected  
387 for earlier germination (Fig. 4a, GM), and a small decrease in the time to flowering (Fig. 4a,  
388 TF), but not for a change in stalk height compared to the control treatment (Fig. 4a, SH).  
389 Growing in the control environment, these trait changes resulted in a slightly earlier flowering  
390 time (Fig. 4b) and small decreases in the number of stalks (Fig. 4c), the rosette area (Fig. 4d),  
391 and fitness (Fig. 4e). In their high competition home environment, the plants flowered slightly  
392 earlier (Fig. 4f), and increased competition led to a major reduction in productivity, resulting  
393 in major decreases in the number of stalks (Fig. 4g), rosette area (Fig. 4h), and fitness (Fig. 4i)  
394 compared to the plants growing in the control treatment.

#### 395 *Disentangling the role of individual selection pressures: Pollination*

396 In the pollination treatment, the decrease in pollinator abundance led to a shift from seed  
397 production being mostly carbon limited, to seed production being more pollen limited and an  
398 increase in unfilled seeds (Fig. S10). This shift in limitation leads to a decrease in fitness (Fig.  
399 4i) without a decrease in productivity, as the number of stalks (Fig. 4g) and rosette area (Fig.  
400 4h) did not change relative to the control treatment. The decrease in pollinator density selected  
401 for slightly earlier germination (Fig. 4a, GM), but there were no changes in selection for stalk  
402 height and time to flowering (Fig. 4a, SH, TF). This resulted in plants that flowered slightly  
403 earlier compared to plants from the control treatment (Fig. 4b), but achieved an equal number  
404 of stalks (Fig. 4c), rosette area (Fig. 4d) and fitness (Fig. 4e) in the control environment.

## 405 **Discussion**

### 406 *Simulation of elevational ecotypes*

407 Adaptation to local conditions is a key mechanism in the evolution and diversification of plant  
408 species (Hargreaves & Eckert, 2019; Hargreaves *et al.*, 2020). Our E-FSP model was able to  
409 reproduce the patterns of local adaptation along an elevational gradient found in *D.*

410 *carthusianorum*. The model reproduced the qualitative differences between two elevational  
411 ecotypes in two phenological (germination and time to flowering) and one morphological trait  
412 (stalk height), as well as qualitative differences in four variables related to plant performance  
413 that emerge from GxE interactions (flowering time, number of stalks, rosette area and seed  
414 production). Moreover, the model satisfied the home vs away and local vs foreign criteria that  
415 indicate populations are locally adapted to their home environments, in line with empirical  
416 evidence (Pålsson *et al.*, *in prep.*). It is remarkable that the model was able to recreate these  
417 patterns of local adaptation in a complex natural system where selection is driven by multiple  
418 abiotic and biotic agents.

419         So far, FSP models have mostly focussed on agricultural (Lopez *et al.*, 2010; Zhu *et*  
420 *al.*, 2015; Evers & Bastiaans, 2016; Coussement *et al.*, 2020), horticultural (Sarlikioti *et al.*,  
421 2011; Chen *et al.*, 2014; Dieleman *et al.*, 2019; Zhang *et al.*, 2020), and model systems  
422 (Bongers *et al.*, 2018). FSP models that simulate natural systems with an increased ecological  
423 complexity are seeing recent development, yet these models are still being validated on data  
424 collected under controlled experimental conditions (de Vries *et al.*, 2018; Faverjon *et al.*, 2019)  
425 and still lack the ecological variability and complexity that shape plant communities (Bongers,  
426 2020; de Vries, 2021). Here, we validated our model on data from a transplant experiment  
427 where plants grew under natural conditions, which, to our knowledge, is the first time an E-  
428 FSP model has been validated to empirical data collected under natural conditions. The model's  
429 ability to recreate the patterns of selection exerted by such a complex environment highlights  
430 the potential of this approach for learning more about selection and to study the complex eco-  
431 evolutionary dynamics that shape natural plant communities.

#### 432 *Disentangling the role of individual selection pressures: the abiotic environment*

433 Our results suggest that in the case of *D. carthusianorum*, the abiotic environment is the most  
434 important driver of elevational adaptation, imposing strong selection pressure on both the

435 phenological (germination and flowering times) and morphological (stalk height) traits. These  
436 selection pressures resulted in high elevation plants that, growing in a shared environment  
437 following adaptation, flowered earlier, were shorter and accumulated less biomass than plants  
438 from the low elevation population. These results match commonly reported trends in studies  
439 of plant adaptation along elevational gradients (Halbritter *et al.*, 2018). Evolution towards  
440 smaller size is generally assumed to be advantageous in alpine environments due to warmer  
441 microclimates close to the ground, increased protection from wind, or a result of selection for  
442 increased stress resistance (Körner, 2003). Interestingly, our model did not implement  
443 microclimate, wind or stress resistance, yet reproduced this pattern of adaptation through  
444 divergence in phenological traits. High elevation genotypes germinate late to avoid frost  
445 damage and flower fast to complete the reproductive cycle within the summer season, resulting  
446 in a plant phenotype that has a shorter vegetative stage. This shorter period of vegetative growth  
447 leads to lower potential for biomass accumulation, a decrease in the number of stalks and a  
448 decrease in reproductive fitness. Thus, the shift in phenology may be responsible for lower  
449 stalk height, because of the decrease in potential biomass accumulation. Overall, our results  
450 suggest that divergence in phenological traits potentially compound direct selection pressure  
451 for shorter and smaller phenotypes at high elevations.

#### 452 *Disentangling the role of individual selection pressures: the biotic environment*

453 In our model, both interspecific competition and decreased pollination strongly decreased plant  
454 fitness, but contributed comparatively little to local adaptation. This is in concordance with the  
455 findings of a recent meta-analysis that showed biotic interactions generally do not make  
456 patterns of local adaption stronger or more common (Hargreaves *et al.*, 2020), despite having  
457 a strong and well documented effect on plant performance (e.g. Weiner, 1990).

458 In our model, interspecific competition selected for earlier germination and earlier  
459 flowering, increasing the resource capture early in the season when interspecific competition

460 was lower, but also increasing the risk of frost damage early in the season. This highlights how  
461 trade-offs between different components of plant performance (e.g. biomass accumulation and  
462 survival) can drive selection in opposite directions, potentially resulting in stabilising selection.  
463 Surprisingly, interspecific competition did not select for an increase in stalk height compared  
464 to the control treatment (i.e. Control vs Competition, see Fig. 4a). This contradicts expectations,  
465 as increased height is a well-known response of plants that grow in a competitive environment  
466 (Ballaré *et al.*, 1990; Falster & Westoby, 2003). Traits such as leaf angle, leaf shape and  
467 especially stem elongation are known to be key determinants of the outcome of competition  
468 for light as they determine leaf light interception by mediating the position of leaves relative to  
469 the surrounding vegetation (Franklin, 2008; Ballaré & Pierik, 2017). However, plants growing  
470 in competition can be equally tall as plants growing in the absence of competition, yet with a  
471 much higher height to biomass ratio caused by decreased biomass accumulation under  
472 competition, making the same investment in height growth relatively more costly (de Vries *et*  
473 *al.*, 2018). When considering the investment in height relative to plant biomass, both the *in*  
474 *vivo* and *in silico* populations of *D. carthusianorum* growing with interspecific competition  
475 show the increased investment in height growth that is expected in a competitive environment.

476 In our model, reduced pollinator abundance decreases plant fitness, but does not affect  
477 selection. The *in silico* plants can increase their competitiveness for pollinators in one of three  
478 ways; produce more flowers, produce taller stalks, or have a longer generative stage of  
479 development and thus have a longer period in which to attract pollinators. While each of these  
480 traits will increase male fitness and potential female fitness, they also severely restrict the  
481 plant's ability to accumulate biomass, compete for light and nutrients, and to fill seeds. The  
482 model did not include flower traits, which are known to be strong drivers of pollinators  
483 visitation (Fornoff *et al.*, 2017; Walther, 2020), and are known to differ along elevational  
484 gradients (Fabbro & Körner, 2004). Flower traits may be the main mechanism for plants to

485 increase pollinator attraction as they potentially come at lower (opportunity) costs than the  
486 mechanisms included in our model.

#### 487 *Future model development*

488 Here, we have shown the potential for E-FSP modelling to simulate the emergent behaviour of  
489 a complex natural system that includes abiotic and biotic agents, and integrates physiological,  
490 ecological and evolutionary mechanisms. E-FSP modelling is a promising and versatile tool  
491 that is capable of simulating more complex and dynamic systems than is common in FSP  
492 modelling, and integrates more physiological and spatial detail than commonly used eco-  
493 evolutionary modelling approaches. We would like to highlight two avenues of future model  
494 development for E-FSP models: simulation of multi-species communities with different life-  
495 history strategies, and simulation of complex genetic and demographic processes that shape  
496 local adaptation.

497 FSP modelling has proven to be capable of simulating the growth and development of  
498 a wide range of plant species (Dunbabin *et al.*, 2013; Pagès *et al.*, 2014; Louarn & Song, 2020),  
499 and FSP modelling is being used to simulate multi-species systems in an agricultural setting  
500 (Evers *et al.*, 2019). Conversely, FSP models that focus on natural systems are often used to  
501 simulate single plant species rather than diverse mixed-species communities, which have  
502 received only recent attention (Faverjon *et al.*, 2019; Bongers, 2020; de Vries, 2021). Here, we  
503 have focussed on a single plant species, but have shown the model's ability to simulate the  
504 diversifying forces of selection. A key point of focus for the future development of E-FSP  
505 models is the simulation of different life-history traits and species co-existence. By simulating  
506 a community consisting of multiple species, the model can theoretically select for different life-  
507 history traits that fill different niches. The main challenge for the implementation of different  
508 life-history traits lies in the complexity of carbohydrate and nitrogen cycles in perennial plants,  
509 which leads to difficulties in linking theory to observations and formulating a comprehensive



510 mechanistic model (Monson *et al.*, 2006).

511         Future development of E-FSP modelling can see the incorporation of more detail in  
512 genetic and demographic processes that drive population and community dynamics (Lowe *et*  
513 *al.*, 2017). In particular, gene flow between populations is known to play a complex eco-  
514 evolutionary role as it can either promote or constrain adaptation, dependent on the migration-  
515 selection balance (Garant *et al.*, 2007). Gene flow is traditionally seen as a force that  
516 homogenises populations by working against the diversifying forces of selection, which drive  
517 local adaptation (Haldane, 1930; García-Ramos & Kirkpatrick, 1997). However, recent studies  
518 show that local adaptations can be maintained despite high gene flow provided that selection  
519 coefficients can sustain ecotypic divergence (Gonzalo-Turpin & Hazard, 2009; Fitzpatrick *et*  
520 *al.*, 2015; Tigano & Friesen, 2016; Luqman *et al.*, 2021). On the other hand, low amounts of  
521 gene flow between locally adapted populations can be beneficial as they allow adaptive alleles  
522 to spread across populations and lead to genetic rescue in the face of rapid environmental  
523 change (Slatkin, 1987; Rieseberg & Burke, 2001; Tallmon *et al.*, 2004). E-FSP models can  
524 contribute to our understanding of the role gene flow plays in mediating the responses of plant  
525 communities to environmental change, particularly because the strength of selection, and thus  
526 the migration-selection balance, emerges naturally from interactions between mechanisms  
527 implemented in the FSP model.

528 The model presented here represents a major advance in the development of mechanistic  
529 models that incorporate physiological, ecological and evolutionary mechanisms to simulate the  
530 complexity of plant phenotypic variation. We have shown the promise of this methodology to  
531 explore the ecological complexity that drives local adaptation in natural plant communities,  
532 thereby complementing experimental and statistical modelling approaches. The approach  
533 offers a tool to better understand what mechanisms and selective agents drive local adaptation,  
534 and how local adaptation mediates the response of plant communities to rapid environmental

535 change.

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