

1 Evidences of a component Allee effect for an invasive pathogen: *Hymenoscyphus* 2 *fraxineus*, the ash dieback agent

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5 Abstract

6 Invasive pathogens are a major threat to forest health especially in managed forest with low diversity. The die-
7 back of European *Fraxinus spp.* caused by the fungus *Hymenoscyphus fraxineus* is the latest example of pathogen
8 invasion causing widespread damage. Host resistance and environment, in particular stand factors were shown
9 to strongly impact disease severity on European ash. The fact that *H. fraxineus* reproduce mostly through hetero-
10 thallic sexual reproduction suggest that an Allee effect could limit the mating success at low host densities, thus
11 limiting inoculum production and disease development. Populations of *H. fraxineus* were monitored during the
12 fruiting period in a network of stands across a host density gradient in forest and non-forest environment. Ash
13 dieback, basal area of ash, density of infected ash leaf debris (rachis) and apothecia in the litter and ascospores
14 load in the air were determined in the different environments during two years. We showed significant differ-
15 ences between forest and non-forest environment with ash dieback, infection rate and inoculum production
16 higher in forest settings. Host density significantly affected disease development, with crown dieback, density of
17 infected rachis in the litter and inoculum production increasing with host density. We also demonstrated that
18 fruiting rate, i.e. the number of apothecia per infected rachis dry weight, is strongly dependent on infected rachis
19 density. Inoculum production is therefore limited at low host densities. Such a component Allee effect could be
20 important in *H. fraxineus* epidemiology and invasion dynamic.

21 Key words: component Allee effect, host density, inoculum production, *H. fraxineus*

22 Introduction

23 Emerging infectious diseases are major threat to forest ecosystems. They are frequently induced by invasive fungi
24 and can have very important consequences on the biodiversity (Desprez-Loustau et al. 2007, 2009). The most
25 frequently cited examples of tree species that were drastically reduced by an invasive pathogen include chestnut
26 in USA with *Cryphonectria parasitica* or elm in Europe and the USA with *O. novo-ulmi*. The rate of invasive forest
27 pathogens arrival has been increasing from 1980 to 2008 in Europe, with a particular importance of ascomycetes
28 (Santini et al. 2013). Invasive pathogens currently represent a proportion of about 50% of the disease cases re-
29 ported by the forest health survey system in France, with a high share of currently severe epidemics such as ash
30 dieback caused by *Hymenoscyphus fraxineus* (Desprez-Loustau et al. 2016).

31 As any alien species, invasion success of exotic fungal pathogens depends on overcoming different barriers (Black-
32 burn et al. 2011) to thrive in their new environment. The pathogen needs to be transported to the new environ-
33 ment, then it must adapt to the environment, which implies finding available hosts (Engering et al. 2013). During
34 this establishment step, the invasive pathogen needs to reproduce and to survive efficiently off season to produce
35 a stable population. The next step is the spread throughout the new territory that depends on the dispersal abili-
36 ties and on the availability of favourable habitats. The establishment is a critical phase; it depends strongly on the
37 habitat suitability, but also on the pathogen population dynamic (Taylor and Hastings 2005) and on the host den-
38 sity present in the landscape (Park et al. 2001; Condeso and Meentemeyer 2007). Founding populations of invasive
39 species are often small and could be subjected to an Allee effect. Allee effect is a mechanism that may reduce the
40 population growth at a low density and thus reduce the establishment likelihood. Stephens et al. (1999) defined
41 the Allee effect as “a positive relationship between any component of individual fitness and either numbers or
42 density of conspecifics”. He further distinguishes the component Allee effect, which affects an individual fitness
43 component and demographic Allee effect which affects the total species fitness. A component Allee effect would
44 be a positive correlation between a fitness parameter such as the mating success with the population density

45 whereas the demographic Allee effect implies a relationship between the per capita growth rate and the popula-
46 tion density and is more difficult to demonstrate. An Allee effect may limit the population growth rate at low
47 density with different strength. Strong Allee effect result in negative growth rate at very low population density,
48 leading to extinction while weak Allee effect will only reduce the population growth rate through density depend-
49 ence. In invasive species, the Allee effect may result in a latency phase at the colonisation front in which population
50 growth is highly dependent on population density followed by an important growth increase when the population
51 reach a so-called Allee threshold (Hastings 1996; Veit and Lewis, 1996). In populations that are denser than the
52 Allee threshold population density is no longer the limiting factor for growth. In addition, another factor limiting
53 invasion is the strong dependence of the invasive pathogen population on the population density of its host. A
54 high host abundance in the landscape allows the pathogen to colonize the area more easily (Jules et al. 2002).
55 Moreover, as other parasites, plant pathogens are known to be a driver of host population size (Cobb et al. 2012).
56 Pathogens appear to strongly structure host populations limiting growth and regeneration or induce mortality. On
57 the contrary, low host density could limit the development of the pathogen population especially if its dynamic is
58 subject to Allee effect.

59 The invasive pathogen *H. fraxineus* was described as the responsible of ash dieback (Kowalski et al 2009; Gross et
60 al. 2014a; Baral et al 2014). First observations of European ash dieback were noted in Poland in the 90's (Przybył
61 2002) and was attributed *a posteriori* to the introduction of the ascomycete *H. fraxineus* (Gross et al. 2014a). This
62 pathogenic fungus of common ash (*Fraxinus excelsior*) and narrow-leaved ash (*Fraxinus angustifolia*) is native to
63 East Asia (Zhao et al. 2013; Gross et al. 2014b; McMullan et al. 2018) and spread through Europe to reach Ireland
64 in 2012 (Short et al. 2019), Montenegro in 2016 (Milenković et al. 2017), north of Spain in 2021 (Stroheker et al.
65 2021). This pathogen caused serious damage in the ash stand, threatening associated species and biodiversity
66 (Pautasso et al. 2013). *H. fraxineus* biological traits and the constant spread rate observed over time in France led
67 Hamelin et al (2016) to suggest the existence of a component Allee effect caused by the mating success.

68 This hypothesis is based on three facts. First, *H. fraxineus* is a heterothallic fungus, *i.e.* successful sexual reproduc-
69 tion needs presence of two mating types (Gross et al. 2012). The reproduction of this foliar pathogen occurs early
70 in summer on residual leaf debris present in the litter (rachis= petiole + central vein of the compound leave) and
71 result in formation of apothecia. Despite rare observations of apothecia production on other infected tissues by
72 *H. fraxineus* (Kirisits and Freinschlag, 2012; Kowalski and Holdenrieder, 2009; Wylder et al. 2018), the infected
73 rachis density in the litter could be considered as main sexual reproductive population. Moreover, the ascospores
74 released by apothecia are the main inoculum vector. Indeed, conidia produced on the rachis are believed to act
75 only as spermatia (Gross et al. 2012), and their spread by splashing is limited to few centimetres or few metres.
76 As a consequence, fecundation might be limited at low density of infected rachis in the litter.

77 Second, the spread of the pathogen is strongly linked to the host density. *Fraxinus* is widely present in the land-
78 scape and its distribution area extends from South of Scandinavia from North of Turkey and North of Spain (*F. ex-*
79 *celsior*) and South of Europe and Turkey (*F. angustifolia*) (EUFORGEN) making it easier for the pathogen to spread.
80 A study showed the importance of ash density and tree cover fragmentation for establishment and disease devel-
81 opment at the landscape scale (Grosdidier et al. 2020). Damage on the crown and collar was severe in dense ash
82 stands in forest while isolated ashes out the forest were less affected. Moreover, the severity of ash dieback is
83 correlated with the load of inoculum produced in the litter (Marçais et al. 2016). A component Allee effect on the
84 reproduction success could further reduce the inoculum production at low ash density.

85 Finally, the environmental conditions strongly constraint the severity of ash dieback. The inoculum production is
86 particularly affected by the level of ambient humidity. Abundant apothecia formation needs high humidity (Hietala
87 et al. 2013; Dvorak et al. 2016; Grosdidier et al. 2020) and thus, factors influencing ambient humidity such as tree
88 cover, vicinity to river, topography and precipitation will impact inoculum production and damages to the ashes
89 (Havrdová et al. 2017.; Enderle et al. 2019; Skovsgaard et al. 2017; Grosdidier et al. 2020). Further, temperatures
90 higher than 35°C are lethal for *H. fraxineus* which explains limited spread in the Mediterranean region with hot
91 summers (Grosdidier et al. 2018; Hauptman et al. 2013).

92 The objective of this study was to unravel the relationship between host density and their leaf litter production
93 (rachis density), their crown health status, infected rachis density in the litter as reproductive pathogen population
94 size and inoculum production by apothecia density in ash dieback. Assuming that the population size of *H. frax-*
95 *ineus* can be measured by infected rachis density in the litter and depends on ash density and level of infection,
96 we hypothesised that a component Allee effect on the success mating limits inoculum production and the ash
97 dieback at low population density (host and pathogen). We monitored the population dynamics of *H. fraxineus*
98 during the period of inoculum production in an ash density gradient in forest and open landscape.

99 Materials and methods

100 Stand Characterisation

101 A network consisting in 20 plots in forest stands and 10 plots in hedges and small woods was installed on the
102 village of Champenoux in North East France (WGS84 48.7521N 6.3409 E, Fig. 1). The network was established in
103 order to obtain a gradient of host density. Host density was measured by the basal area of ash (*Fraxinus excelsior*).
104 At each studied location, three concentric circular plots were established. All tree stems with a diameter at breast
105 height (DBH) over 7.5cm were measured within a radius of 7 m (154 m) while only tree stems with a DBH over
106 22.5 cm were measured in a radius of 7-16 m (805 m²) and with a DBH over 47.5 cm in a radius of 16-21 m (1386
107 m²). The basal area of ash in m². ha⁻¹ was computed by summing individual stem area at breast height weighted
108 by the sampling surface. We also measured the basal area of other trees species present in the stand and total
109 basal area was used as a proxy of canopy closure. Ash basal area was used to explain local density of rachis in the
110 litter. However, in order to evaluate the impact of ash density on ash health, we weighted basal area by the rate
111 of tree cover within the 100-m radius around the points; this enabled to account for the very patchy presence of
112 trees outside forest stands. The tree cover rate was computed with QGIS software using an IGN shape file cor-
113 rected with aerial photograph (BD Ortho® edition 2018 and BD FORET® version 2.0 available in Web Map Service
114 flow on <https://geoservices.ign.fr>). The size and health status of ash trees included in the basal area assessment
115 was recorded with their DBH measure and the following rating of crown mortality: 0-10% (healthy), 10-50% (symp-
116 tomatic), 50% - 75% (declining) and >75%. The health status of an ash stand was computed as the mean of the
117 tree ratings (using the median of their health class) and the ash size was estimated with the mean of DBH. Four
118 plots were moved by a short distance in 2021 because the 2020 location was compromised by logging. The mete-
119 orological conditions covering the sampling periods (from 1 June to 31 July 2020 and 2021) were collected at the
120 Champenoux weather station (Figure 1). The heat level was expressed by the mean of daily maximal temperature
121 and mean of daily mean temperature, the humidity level was expressed by mean of daily humidity and sum of
122 precipitation.

123 *Hymenoscyphus fraxineus* population size and inoculum production

124 In each stand, the density of ash rachises at the soil surface and their frequency of colonisation by *H. fraxineus*
125 was determined in June 2020, June 2021 and July 2021. For that, all ash rachises present along ten 0.1-m² areas
126 located along two 10-m perpendicular transects were collected (10-cm wide area along the transect on each other
127 meter). The rachises were sorted in the laboratory according to their colonisation status by *H. fraxineus*. Rachis
128 with the presence of a distinct black pseudosclerotial plate is characteristic of *H. fraxineus* infection while rachis
129 un-colonized by the pathogen remain light brown to grey with absence of a pseudosclerotial plate and are consid-
130 ered as healthy. To confirm the assessment, a part of rachis identified as infected or healthy were placed in moist
131 chamber for 8 weeks to monitor appearance of apothecia. The proportion of rachises from both categories that
132 produced *H. fraxineus* apothecia was then computed.

133 Infected and healthy rachises collected on plots and sorted in lab were dried for 48h at 50°C and then weighted.
134 The number of apothecia present on these rachises was counted during the sample collection of rachises. The
135 mean dry weight of rachis (infected and total) and the mean apothecia frequency per plot were then computed
136 in g.m⁻² and No. of units.m⁻², respectively. The fructification rate was computed as the number of apothecia per
137 infected rachis dry weight (N.g⁻¹). Data on number of apothecia and infected rachises density per m² from previous

138 work were used in the analysis of fructification rate (15 plots sampled in 2012 from Grosdidier et al. 2018, 23 plots
139 sampled in 2016 and 31 plots sampled in 2017 from Grosdidier et al. 2020). The sampling method is similar to our
140 method except that the infected rachises density is measured in length of rachis per unit surface ($\text{cm}\cdot\text{m}^{-2}$), the
141 rachis density was converted in weight per m^2 ($\text{g}\cdot\text{m}^{-2}$) according to the relationship ($L = 163.73 * DW$, where L is
142 the length in cm and DW, the dry weights of rachises, Grosdidier et al. 2020).

143 The spore trapping method developed by (Grosdidier et al. 2017) was used to determine the air load of *H. fraxineus*
144 ascospores on the studied plots. Shortly, the spore traps are passive traps composed of cellulose filter (Whatman™
145 150 mm diameter Cat No 1001-150) imbibed of 5 ml of glycerine and place on a styrifoam block at 1 meter above
146 the ground. Three spore traps were set up per plot and left exposed for two consecutive periods of 15 days (22
147 June to 8 July and from 9 to 22 July). After exposure on the plots, the filters were recovered and put individually
148 in plastic bags. In the laboratory, 30 ml of 4x TE buffer (40 mM Tris-HCl, 4 mM EDTA, pH 8.0) heated at 60°C was
149 added into each plastic bag with filter. The filters were gently hand rubbed through the plastic to separate the
150 captured particles from the filter. The TE buffer was then transferred in 50 ml vials and centrifuged 15 min at 2700
151 g. The supernatant was removed to keep approximately the bottom 3 ml of suspension containing most of the
152 particles. This 3 ml of suspension was transferred in two 2-ml microtubes, centrifuged 5 min at 18 620 g and,
153 750 μl of supernatant was removed from each tube. The remaining 750 μl were vortexed, pooled in one tube and
154 centrifugated once again 5 min at 18 620 g. The 200 μl bottom of the concentrated particles solution was kept at
155 -20°C until DNA extraction.

156 DNA was extracted from the 200 μl concentrated particles solutions using the DNeasy plant mini kit (Qiagen). Two
157 3-mm and twenty 2-mm glass beads were added to the particles solutions together with 400 μL of lysis buffer and
158 4 μl RNase. The samples were grounded twice 50 s at 6 $\text{m}\cdot\text{s}^{-1}$ with FastPrep-24 MP BIO and incubated 30 min at
159 65°C to lyse the cell. The following steps were done as described by the manufacturer. Total DNA was then eluted
160 in 200 μl AE buffer.

161 The number of *H. fraxineus* ascospores from spore traps was quantified by qPCR using the method developed by
162 loos et al. (2009). The 15 μl of reactional mix were composed of 1x Brilliant II qPCR master mix (Agilent Technolo-
163 gies), 0.03 μM ref dye provided with the master mix, 0.01 $\text{U}\cdot\mu\text{l}^{-1}$ UDG (New England BioLabs), 0.3 μM each Cfrax
164 primers (Cfrax-F, 5'-ATTATATTGTTGCTTTAGCAGGTC-3' and Cfrax-R, 5'-TCCTCTAGCAGGCACAGTC-3'), 0.1 μM
165 Cfrax-probes (Cfrax-P, 5'-FAM-CTCTGGGCGTCGGCCTCG-BHQ1-3') and 2 μl template DNA. The real time reaction
166 was performed in a Quantstudio 6 thermocycler (Applied Biosystem). The qPCR reaction was initiated by first pre-
167 cycling step at 37°C for 10 min for UDG activation and the initial denaturation step at 95°C for 15 min followed by
168 50 cycles of denaturation at 95°C for 15 sec and hybridization /elongation at 65°C for 55 sec. The ascospore quan-
169 tification was performed using ascospore solutions obtained by tenfold cascade dilution with from 50 000 to 5
170 ascospores per μl .

171 Statistical analyses

172 Crown decline, rachis densities in the litter (total and infected by *H. fraxineus*) and the infection rate of rachis in
173 the litter were analysed with generalised linear mixed models (glmm) using the R library glmmTMB. In both cases,
174 the plot was declared as random effect.

175 The crown decline rate and the rachis infection rate were modelled using a Beta-distribution that is well adapted
176 to variables lying between 0 and 1 (Figuroa-Zúñiga, et al 2013); we used the logit link function. The explicative
177 variables were the host density (ash basal area) and environmental variables, with the measured year added for
178 the rachis infection rate as fixed factor. The densities of ash rachis in the litter which is a positive continuous
179 variable was modelled with the Gamma distribution with a log link function. Ash density, crown decline and mean
180 ash diameter were included as explicative variables. The apothecia density, the amount of ascospore detected in
181 the spore traps and the infected rachis density were modelled with the negative binomial using an identity link
182 function. The sampled plots were declared as random factor.

183 These GLMs relationships were used to build a Structural Equation Modelling (SEM) using the R-package “piece-
184 wise” (Fig. 3). SEM highlighted the correlation between host variables (ash basal area, average size and health
185 status, total rachis density), *H. fraxineus* population variable (infected rachis density) and inoculum production
186 (apothecia density and amount of ascospores). The year of sampling was added to evaluated the summer varia-
187 bility and the environment to compare the forest to hedge and small wood. The coefficient estimates were stand-
188 ardisd in order to compare the effect of the different parameters.

189 To assess whether a component Allee effect was present for mating success, the relation between the apothecia
190 production rate and *H. fraxineus* population size (infected rachis density in the litter) was studied. The fruiting rate
191 (τ) was define as the number of apothecia produced per unit of dry weight of infected rachis present in the soil
192 litter:

$$193 \quad \tau = \frac{\alpha}{\varrho}$$

194 Where α , is the apothecia number per m² and ϱ , the infected rachis density (g.m⁻²). Without Allee effect, τ is
195 constant, while in presence of a component Allee effect, it should be positively correlated with the infected rachis
196 density. The Allee effect should be most effective at low rachis density, with no density dependence above a
197 threshold value (Allee threshold). Therefore, the relationship between τ and the infected rachis density can be
198 modelled with a Gompertz function (equation 1) with the hypothesis that the fruiting rate may be zero at very low
199 density, positively dependent in low density and reach an optimum at high density superior of Allee threshold
200 (Fig.2).

$$201 \quad \tau = A e^{-e^{[\frac{\mu}{A}(\lambda - \varrho) + 1]}} \text{ (equation 1)}$$

202 With A , the maximum fruiting rate (optimal mate encounter), μ , the maximum slope and λ , the strength of Allee
203 effect, i.e the value of infected rachis density under which mate encounter does not occur and no apothecia are
204 produced. The *Allee threshold* is defined as the value of infected rachis density ϱ above which the fruiting rate
205 τ is A on average and does not depend on ϱ anymore (Fig. 2).

206 The Gompertz function was fitted in a Bayesian framework using the R2jags package. The fructification rate τ was
207 assumed to follow a Gamma distribution with the mean following equation 1. Parameters A and μ were assumed
208 to depend on environment (forest or hedge/small wood) and on a year random effect. Flat priors were assumed
209 for the parameters: normal distribution N (0,0.0001) for parameters for A , uniform distribution U (20,100) for
210 parameters for μ , uniform distribution U (0,5) for λ and uniform distribution U (0,100) for variance of the random
211 factors and of the Gamma distribution. We run 3 MCMC chains for 100000 iteration with a burn-in of 75000 and
212 a thin of 10. The convergence was assessed by Gelman-Rubin tests. The fit of the model was checked by comparing
213 the mean and dispersion of observed data and of data simulated according to the model. We tested whether a
214 significant Allee effect was present by computing the value of the *Allee threshold*, defined as the value of infected
215 rachis density ϱ for which average fructification rate τ reached $0.99 * A$, and assessing whether this threshold was
216 significantly different from 0.

217 The total density of rachises in the litter needed to reach the Allee threshold depends on the infection rate (equa-
218 tion 2). On the other hand, the log of total density of rachises in the litter is also proportional to ash basal area
219 (equation 3).

$$220 \quad I_R = \frac{A_t}{T_d} \text{ (equation 2) and } \log(Td) \sim AB \text{ (equation 3)}$$

221 Where I_r , the litter rachises infection rate minimum to reach Allee threshold is determined by A_t the infected
222 rachis density of the Allee threshold divided by T_d the total rachis density provided by AB , the ash basal area

223 We estimated the litter rachises infection rate needed to reach the Allee threshold depending on the ash basal
224 area AB . For that, a bootstrap procedure was used. Values for the Allee threshold were derived from the Bayesian
225 analysis of *H. fraxineus* fruiting rate. The total ash rachis density (infected and healthy by *H. fraxineus*) was esti-
226 mated depending on the ash basal area according to the Gamma fitted regression. First, 7500 set of simulated

227 model parameters were generated assuming a multinormal distribution of the parameters using the function `mvr-`
228 `norm` of the MASS R package. Using this set of simulated parameters and average values of ash dieback rating and
229 DBH, the total rachis density was computed for plots with an ash basal area AB from 1 to 40, and ash dieback and
230 DBH values equal to the mean values of the studied plots. The litter rachises infection rate needed to reach the
231 Allee threshold was then computed as the ratio between the Allee threshold and the total rachis density. We
232 computed its mean and its 2.5% and 97.5 quantiles.

233

234 Results

235 The ash basal area gradient extends in forest from 2.2 m².ha⁻¹ to 18.3 m².ha⁻¹ with 50% of stands below 4.6 m².ha⁻¹.
236 In hedge and small wood, this basal area is between 2.5 m².ha⁻¹ and 37.5 m².ha⁻¹ with a median of 10.7 m².ha⁻¹.
237 Pure ash stands were no longer present in forests around Champenoux. The highest ash basal areas were ob-
238 served in small wood, where ashes were concentrated in isolated small areas. The ash basal area obtained after
239 weighting by tree cover in a 100-m radius in hedge and small wood range between 0.12 m².ha⁻¹ and 13.5 m².ha⁻¹
240 with a median at 1.8 m².ha⁻¹. The weighted range of total basal area was between 0.12 m².ha⁻¹ and 35.8 m².ha⁻¹
241 with a median of 19.7 m².ha⁻¹.

242 Weather conditions of the summer 2020 and 2021, during the apothecia production and ascospores release pe-
243 riod (June and July), were relatively different. While average daily maximal temperature was similar ($p > 0.05$), air
244 humidity and precipitation were very different, with summer 2020 being drier than summer 2021 ($p < 0.001$) (Table
245 1).

246 The SEM analysis showed strong positive relationships between the population dynamic of *H. fraxineus* and the
247 density of its host (Fig. 3). The amount of ascospores increased with the apothecia density (0.001, $p < 0.05$), which
248 in turn increased with the infected rachis density present in the stand litter (0.017, $p < 0.001$). The infected rachis
249 density depended on the total rachis density (0.26, $p < 0.001$) which itself depended on ash basal area (0.07,
250 $p < 0.01$) and ash diameter (0.03, $p < 0.01$). Increasing crown dieback had a negative effect on total rachis density (-
251 0.05, $p < 0.01$). Crown dieback was more severe when ash density, i.e. ash basal area, was high (0.88, $p < 0.05$). Thus,
252 in dense ash stands with severe dieback, the total rachis density was reduced which could decrease the density of
253 infected rachis. The apothecia density and the amount of ascospores captured were affected by the environment
254 with lower values in hedges and small woods (respectively 0.01, $p < 0.001$ and 0.1, $p < 0.001$). The year affected the
255 apothecia density and even more the infected rachis density (0.57, $p < 0.05$), with lower value in 2020.

256 Crown decline and ashes density

257 A total of 302 ashes was assessed for crown dieback, with 30% of trees located in forest plots and 70% in hedge
258 and small wood plots. The ash trees were overall healthy with 128 trees that rated as 0.05, 120 trees as 0.3 and
259 the most severe dieback classes unfrequently observed (36 trees rated as 0.625 and 18 trees as 0.825, Fig. 4a).
260 Health status, i.e. mean plot crown decline, significantly deteriorates with increasing host density (0.14 ± 0.01 ,
261 $p < 0.001$) (Fig. 4b). Healthy ashes were predominant at ash density less than 2 and their number decreased as ash
262 density increased. By contrast, the number of symptomatic and declining trees increased with ash density. Addi-
263 tionally, the mean crown decline was greater in forest than hedge and small wood with a mean crown decline of
264 0.41 ± 0.05 (IC95%) in forest and 0.19 ± 0.03 (IC95%) in hedge and small wood (Fig. 4b $p < 0.001$). Ash trunk diameter
265 was not related to dieback severity (Fig. 4c, $p > 0.05$) and did not differ between forest and hedge/small wood.

266 *Hymenoscyphus fraxineus* population size and inoculum production

267 The fruiting test confirmed that the classification of rachises as infected / healthy was adequate. A total of $3.6 \pm$
268 1.1 % of rachises classified as healthy produced *H. fraxineus* apothecia, while 96.0 ± 1.2 % of rachises classified as
269 infected were producers.

270 Total rachis density present in the litter significantly depended on ash density, mean ash diameter and health
271 status of ash crowns (Fig. 5). The total rachis dry weight per square meter significantly increased with the ash basal
272 area and mean ash diameter (respective coefficients 0.07 ± 0.01 , $p < 0.001$ and 0.02 ± 0.01 , $p < 0.01$). The deteriora-
273 tion of ashes crowns reduced the density of rachis present in the litter (-1.25 ± 0.55 , $p < 0.05$) (Fig. 5a). The density
274 of rachis per unit ash basal area was decreased with increasing mean crown dieback rate (-1.64 ± 0.60 , $p < 0.01$)
275 (Fig. 5b). In hedge and small wood, the rachis production was slightly lower than in forest situation (-0.54 ± 0.24 ,
276 $p < 0.05$) (Fig. 5b). The rachis infection rate showed no relationship with the ash density ($p > 0.05$). The mean rate
277 of infected rachis was higher in 2020 (0.55 ± 0.07) than in 2021 (0.26 ± 0.05) ($p < 0.001$ Fig. 5c). However, the in-
278 fection rate was similar for plots in forest or in hedge / small wood ($p > 0.05$, Fig 5c).

279 In 2020, the amounts of apothecia produced were very weak. No apothecia were observed in June 2020 and only
280 four plots were sampled in July. In 2021, the apothecia appeared mid-June and all the sites were sampled both in
281 June and July. In 2012, 2016, 2017 and 2020, the observed density of infected rachis was significantly higher than
282 2021 with a mean between 4.3 and 8.7 g.m^{-2} in forest and between 2.7 and 16.0 in hedge and small wood (some
283 values over 20 g.m^{-2} in 2012), whereas the infected rachis density in 2021 was less than 5 g.m^{-2} ($1.3 \pm 0.4 \text{ g.m}^{-2}$ in
284 forest and $0.8 \pm 0.2 \text{ g.m}^{-2}$ in hedges/small woods, $p < 0.001$). The gradient of infected rachis obtained allowed us to
285 highlight a dependence of the fruiting rate of *H. fraxineus* with the density of infected rachis on the litter (Fig. 6a).
286 Indeed, the estimated Allee threshold was significantly different from 0, with very similar values in forest (1.9 IC
287 [1-2.9] g.m^{-2}) and in hedge and small wood (1.5 IC [0.8-2.5] g.m^{-2}). Below the threshold, the fruiting rate increased
288 significantly with the density of infected rachis ($\mu = 93.8$ IC [59.5- 153.3] in forest and 63 IC [39.3-93.5] in hedge
289 and small wood). For density values above the Allee threshold, the fruiting rate did not depend on the density of
290 infected rachis with a mean fruiting rate given by the parameter **A** of the Gompertz equation. The mean fruiting
291 rate was different according to the environment with a higher value in forest (88.6 IC [52.2- 124.8] apothecia.g⁻¹)
292 than in hedge and small wood (48.2 IC [30.3- 65.8] apothecia.g⁻¹). The parameter **λ** was not significantly different
293 to 0 (0.01 IC [0-0.037]) (Fig.6a). The rachis infection rate to reach the Allee threshold depended on the amount of
294 rachises produced by the ashes present in the stand, so on ash basal area (Fig.6b). The Allee threshold would be
295 reached at an infection rate inferior to 0.2 for an ash basal area higher than 20 $\text{m}^2.\text{ha}^{-1}$ and need an infection rate
296 higher than 0.5 for low ash densities ($< 5 \text{ m}^2.\text{ha}^{-1}$). In 2020, 72 % of studied ash stands had an infection rate suffi-
297 cient for a *H. fraxineus* inoculum production not subjected to the Allee effect, whereas only 27% of the ash stands
298 exceeded the Allee threshold in 2021 (Fig. 6b).

299 The amount of ascospores detected in the spore traps in 2021 significantly increased with the density of apothecia
300 observed in the plot litter in 2021 (0.01 $p < 0.05$, Fig. 7a) and no differences were observed between forest and
301 hedges/small woods ($p > 0.05$). In addition, the quantity of ascospores detected in the spore traps in 2020 and 2021
302 was positively correlated with the density of infected rachis observed in the same year and also depended on plot
303 environment and year (Fig. 7b). Indeed, the amount of trapped ascospores was higher in forests (0.26 in 2020 and
304 0.7 in 2021 $p < 0.01$) than in hedge and small woods for similar infected rachis density (-0.03 in 2020 and 0.4 in
305 2021 $p < 0.01$). Furthermore, the amount of ascospores trapped was significantly higher in 2021, than 2020 in par-
306 ticular in hedge and small wood where the amount of ascospores trapped was very weak (Fig. 7b).

307 Discussion

308 This study analyses the relationship between ash, its health status and the population dynamics of *H. fraxineus* in
309 relation to the stand environment. Our results revealed that host density was an important factor in the develop-
310 ment of *H. fraxineus* populations, their inoculum production and subsequent health impacts on ash. We show that
311 a component Allee effect on the fruiting rate exists for this fungal pathogen, confirming the hypothesis suggested
312 by Hamelin et al. (2016), and that it limits inoculum production in the ash stands studied.

313 The larger negative impact of *H. fraxineus* on crown health at higher ash density that we observe is consistent with
314 what has been reported in previous studies (Grosdidier et al. 2020; Havrdová et al. 2017), although this effect
315 could not be observed early in the ash dieback epidemic (Bakys et al, 2013; Marçais et al, 2016). The average
316 crown decline measured in the present study was in the same range as the one observed in 2016 and 2017 in the

317 same area of Champenoux (Grosdidier et al. 2020): The value in forest was of 40% in 2016, 2017 and in 2020; and
318 in none forest locations, slightly higher than 20% in the three study years of Grosdidier et al (2020), and 19% in
319 2020. This might suggest that ash health status stabilized within the past 5 years, so less than 10 years after ash
320 dieback was first observed in Champenoux (2010, see Grosdidier et al, 2020). This seems surprising because mor-
321 tality caused by ash dieback has been reported to develop late, approximately 10 years after the pathogen arrives
322 in old trees (Marçais et al. 2017; Madsen et al. 2021) and it has been reported that ash mortality does not stabilize
323 in the 15 first years of the epidemic (Coker et al. 2019).

324 Several features might explain that discrepancy. On the one hand, the ash density in the studied forest is low and
325 decreased even more as the few pure ashes stand present were clear-cut for sanitary reasons between 2017 and
326 2020. Heavy decline and mortality occurred in the first decade of the epidemic and the logging of severely dieback
327 trees probably removed the less tolerant ashes (Cleary et al. 2017; Børja et al. 2017; Skovsgaard et al. 2017). The
328 remaining ashes might be more tolerant individuals which would limit dieback severity but not *H. fraxineus* pres-
329 ence. On the other hand, severe heat-waves and droughts periods occurred in the area in 2015, 2018, 2019 and
330 2020. The development of *H. fraxineus* is strongly limited by temperature above 35°C (Hauptman et al. 2013,
331 Grosdidier et al. 2018). In addition, apothecia production and ascospore release are influenced by air and soil
332 humidity (Dvorak et al. 2016; Gross et al. 2012; Hietala et al. 2013; Kirisits and Freinschlag 2012; Schumacher
333 2011). Therefore, infection rate of ash rachis in the litter and the crown decline may have been reduced by high
334 temperatures in previous summers (Grosdidier et al. 2018). The drought observed in June and July 2020 may have
335 prevented the apothecia production which would explain the low amount of ascospores that we observed in our
336 spore traps in 2020 and the low proportion of rachis colonized by *H. fraxineus* in the litter 2021.

337 Grosdidier et al. (2020) reported that ash trees in hedges and small woods in the Champenoux area were healthier
338 than those in forested areas and that the crown decline remained stable at about 20% from 2012 to 2018. Our
339 data confirmed that the crown decline observed in 2020 and 2021 in hedge and small woods remained at about
340 20%. Furthermore, as noticed by Grosdidier et al (2020), we showed that, despite the greater decline of ash trees
341 in forest conditions, the total amount and colonization rate by *H. fraxineus* of ash rachis in the litter were similar
342 in hedges and small wood plots than in forest plots. Despite a similar infection rate of rachises, we observed lower
343 apothecia production on them in hedges and small wood compared to forest locations in 2012, 2016, 2017
344 (Grosdidier et al, 2020) and 2021. This lower apothecia production in non-forested areas, certainly due to lower
345 humidity level, may partly explain this impact difference of *H. fraxineus* on ash dieback in hedge and small wood
346 although the infection rate of rachises does not differ between the two environments.

347 The demonstration that *H. fraxineus* is subjected to a component Allee effect linked to mating success offers a
348 new insight into the invasion biology of the pathogen. Although the existence of this component Allee effect was
349 suggested by (Hamelin et al. 2016), it has never been shown that it effectively limits inoculum production in ash
350 stands. The Allee threshold was estimated to be of a similar magnitude in forests and in hedge and small woods
351 (respectively, 1.9 IC [1-2.9] and 1.5 IC [0.8-2.5] g of infected rachis.m⁻²). The parameter λ was estimated to be near
352 0 which means that, even at very low rachis density in the litter, the mating success remains greater than zero.
353 Thus, the observed effect can be considerate as a weak Allee effect. At low ash density, a strong infection rate is
354 necessary to reach the Allee threshold. According our results, in an ash stand with a density lower than 5 m².ha⁻¹,
355 the *H. fraxineus* population could be subjected to Allee effect below an infection rate of 50%. In years unfavourable
356 to leaf infection, like 2020, such an infection rate was seldom reached.

357 Consequently, the development of ash dieback is highly dependent on ash density. At low ash density, significant
358 apothecia and inoculum production occurs only when the infection rate of rachis is high. The component Allee
359 effect might be expected to reduce inoculum production when infection is scattered in newly infected areas, and
360 thus should reduce the overall dispersion rate (Lewis and Kareiva 1993; Taylor and Hastings 2005). This is con-
361 sistent with the hypothesis raised by Hamelin et al (2016) and may explain why the dispersal speed observed in
362 France has remained constant over time, *i.e.* around 60 km per year (Grosdidier 2017). This dispersal speed is very
363 similar to what has been observed elsewhere in Europe (between 30 and 75 km per year Børja et al. 2017; Queloz
364 et al. 2017; Ghelardini et al. 2017). This component Allee effect could also explain that the introduction of the

365 pathogen through the planting of infected seedlings may result in foci that remain limited for an extended period
366 of time. This was observed in central England, where dendrochronological analyses revealed the presence of *H.*
367 *fraxineus* as early as 2005 in ash tree plantations remote from any other sources of inoculum, that is seven years
368 before the pathogen was first reported in the country (Wylder et al. 2018).

369 In addition, the increase of the disease severity could lead to contain the pathogen population growth. Indeed, we
370 observed, as expected, a positive relationship between total rachis density in the litter and ash basal area. How-
371 ever, the rachis production decreases with crown decline, the total rachis density of declining stands were strongly
372 reduced compared to a healthier stand. As a consequence, the suitable reproduction substrate available for *H.*
373 *fraxineus* becomes scarcer as the dieback progress. The density of infected rachis in the litter, which is a good
374 measure of *H. fraxineus* population size, may also be reduced by the tree dieback. This could lead to infected rachis
375 densities below the Allee threshold for mating success. A population density below the Allee threshold has a lower
376 fruiting rate, the amount of ascospores release is weaker, the infection rate will decrease and this could lead local
377 population extinction.

378 The last point concerns the potential impact of an Allee effect on the pathogen genetic diversity during the inva-
379 sion process. Indeed, it was shown through modelling by Roques et al. (2012) that a population subjected to an
380 Allee effect spreads as a pushed wave which results in a genetic diversity that remains stable throughout the
381 colonization process. In this case, the bottleneck-induced loss of genetic diversity for a long-dispersal founder
382 event is rare. Such founder events are usually characterized by low population density and strongly limit the
383 growth of populations subject to the Allee effect (Roques et al. 2012). Noteworthy, Burokiene et al (2015) showed
384 that *H. fraxineus* populations present in eastern Europe, in anciently colonized areas present similar genetic diver-
385 sity compared to populations present on the expansion front. This is in line with the hypothesis of the invasion of
386 Europe by pushed waves mechanism.

387 The management suggested to date to control ash dieback disease in affected stands is to thin stands to reduce
388 host density (Skovsgaard et al. 2017; Enderle et al. 2019; Short and Hawe 2019). The known consequences were
389 a drier and warmer microclimate due to reduced tree cover and openness to light. But, as we have shown, lower
390 ash density also results in lower total rachis density in the litter and, as a consequence, in lower infected rachis
391 density. However, inoculum production depends on this density of infected rachis, especially when it is inferior to
392 the Allee threshold. Therefore, a reduction in rachis density by thinning ash trees will decrease inoculum produc-
393 tion and the Allee effect component may exacerbate this mechanism. This kind of management of ash stands to
394 reduce the production of *H. fraxineus* inoculum could be beneficial to the health of the ash trees. Indeed, the
395 severity of damage is directly related to the level of inoculum present in the area. In case of high inoculum pres-
396 ence, the pathogen induces not only crown dieback but also collar necrosis, pathway to the establishment of other
397 aggressors such as *Armillaria spp* (Husson et al. 2012; Madsen et al. 2021; Marçais et al. 2016).

398 Regarding the future of ash in Europe, it appears that after 10 years of the epidemic, some ash trees remain
399 relatively healthy. As tolerant ash trees have been shown to produce more seeds than declining individuals
400 (Semizer-Cuming et al. 2019), their offspring may be increasingly adapted to the disease, which should be good
401 news for future stands in the region. Moreover, non-forest environment seems have some conditions unfavoura-
402 ble to the disease development, despite an infection rate similar to forest environment, which allows to ash to
403 stay a structuring species of landscape within hedge and small wood.

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540

541 Statements & Declarations

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546 Conflict of interest

547 The authors declare to have not conflict of interests.

548 Author Contributions

549 The design of this study was developed by Benoit Marçais and Simon Laubray. The data were collected by Simon
550 Laubray. The statistical analyses were performed by Benoît Marçais and Simon Laubray. The first draft of the man-
551 uscript was written by Simon Laubray and all author commented and corrected it. The final version was revised
552 and agreed by all author.

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554 Author information

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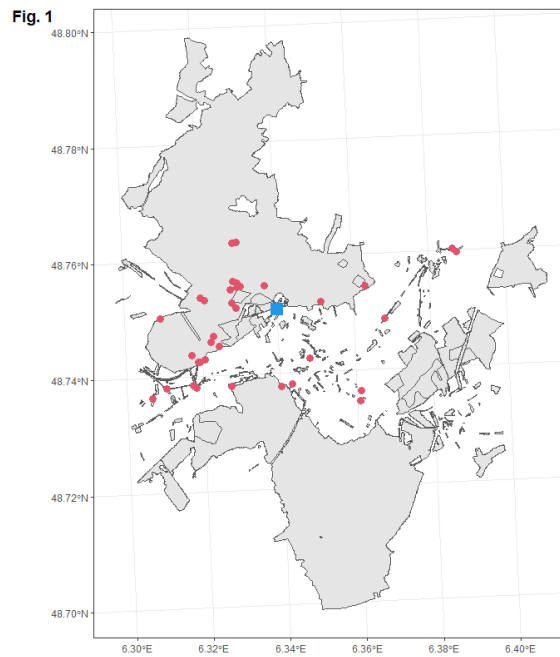
563 **Table 1** Weather conditions during the apothecia production (June and July). Max.Temp : average daily maximal
564 temperature, Moy.temp: average daily temperature, Moy.Hum: average air humidity, Sum.Prec: precipitations
565 sum.

Year	Max.Temp	Moy.Temp	Moy.Hum	Sum.Prec
2020	24.6±1.1°C	18.4±0.7°C	65.9±3.3%***	76.0mm
2021	24.3±0.9°C	18.6±0.6°C	79.7±2.8%***	221.5mm

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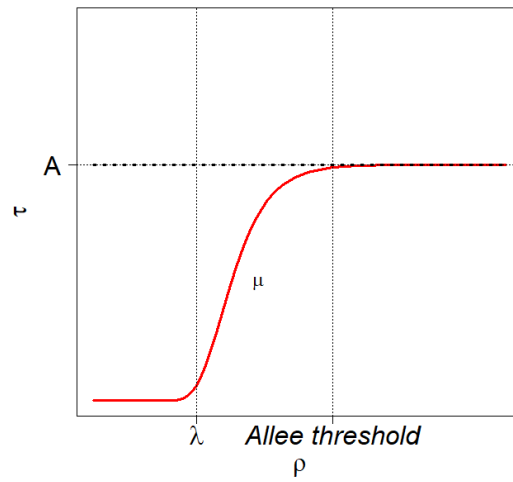
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570 **Fig. 1** Plot distribution around Champenoux village of ash stand sampled for ash density, health status, rachises
571 densities (total and infected), apothecia density and ascospores trapping (red point). Meteorological station (blue
572 square). Wooded area appears in grey on the map.

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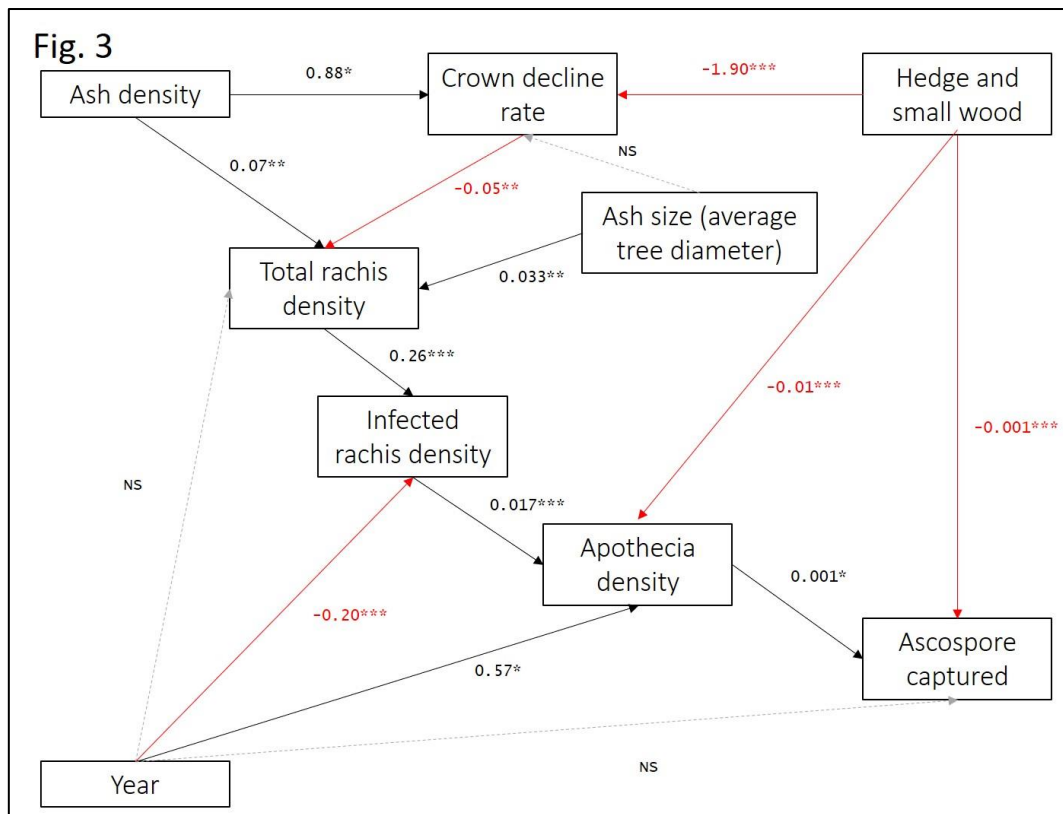
Fig. 2



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576 **Fig. 2** Theoric curve of (ρ) with in bold dashed black line no Allee effect and in red line presence of Allee effect
577 according to Gompertz equation. With the parameter of Gompertz equation: **A** as the average τ reached for den-
578 sity $\rho >$ **Allee threshold**, μ as the maximal slope and λ the minimum value of density ρ where $\tau > 0$

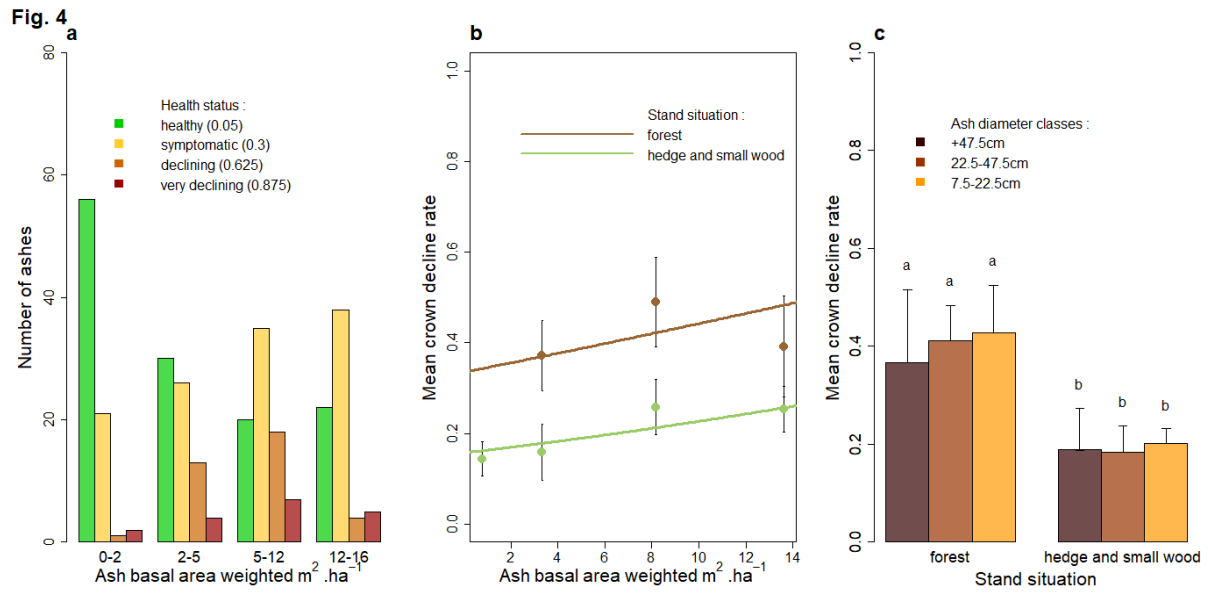
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581 **Fig. 3** Structural equation modeling (SEM) to highlight the relation between hosts parameters (ash density and
582 total rachis density), tree health status (crown decline rate), pathogen population (rachis density infected *H. frax-*
583 *ineus*) and fungal inoculum production (apothecia density and captured ascospores) according to year and envi-
584 ronment. Black line positive correlations, red line negative correlations and dotted line no correlations, the coef-
585 ficients of correlation estimated were scaled with * $p < 0.05$, ** $p < 0.01$ and *** $p < 0.001$.

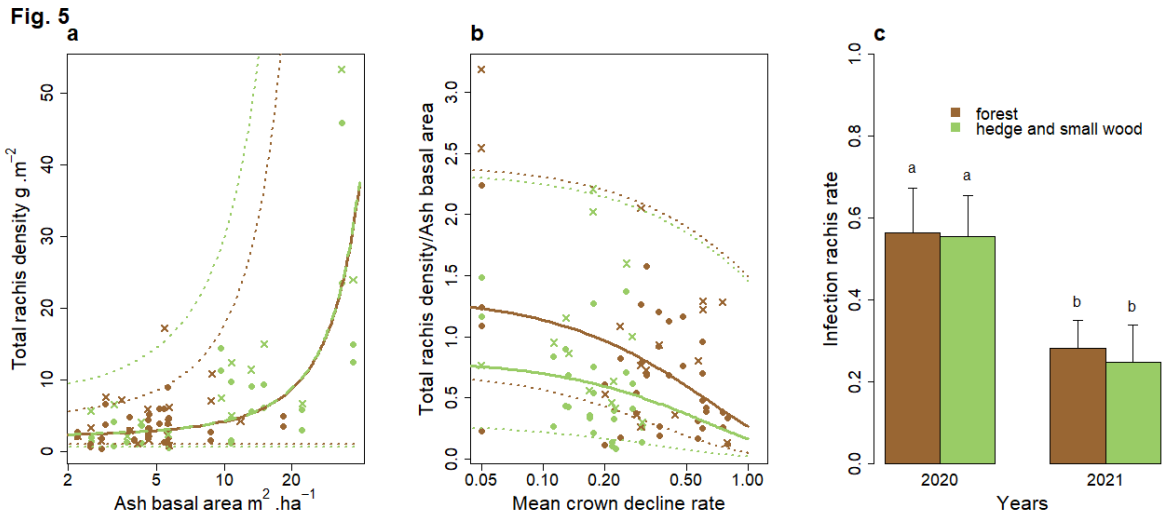
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588 **Fig. 4 a** Ash distribution according to ash density and health status class (with median crown decline rate for each
589 class). **b** Mean crown decline rate of the ash stand according to ash density and the environment. **c** Mean crown
590 decline rate in different ash diameter classes in forest and hedge and small wood.

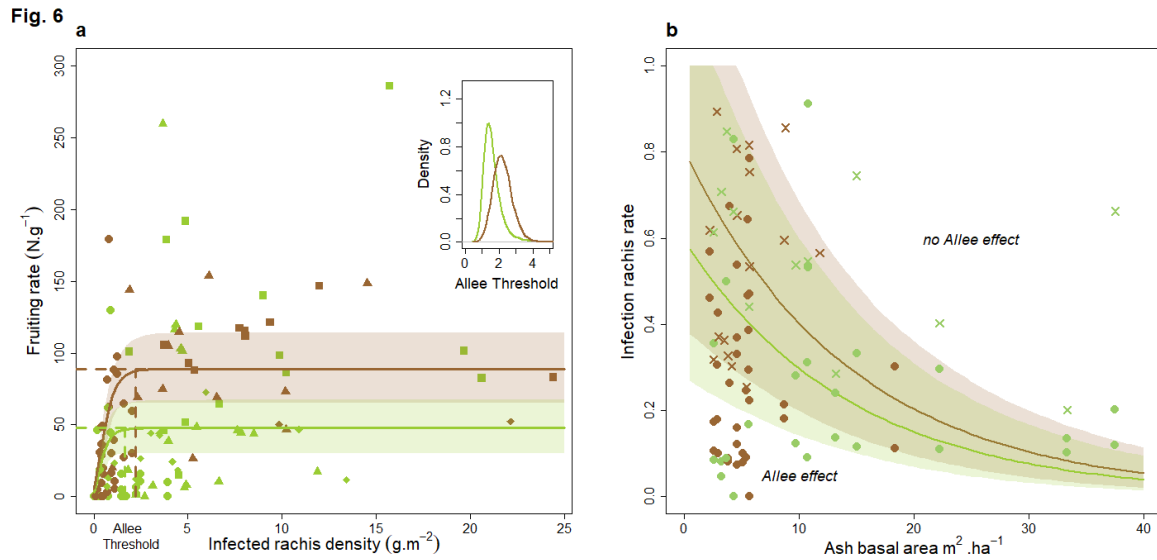
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593 **Fig. 5 a** Total ash rachis density collected in the litter in forest (brown points) and in hedge and small wood (green
594 points) according to ash density and their associated total rachis density, confidence intervals at 97.5% (dotted
595 lines). **b** Rachis production by ashes according to mean crown decline rate for forest situation (brown), hedge and
596 small wood (green) and the fitted regression model (line) and confidence intervals at 97.5% (dotted lines). **c** Infec-
597 tion rate according to ash density for forest situation (brown), and hedge and small wood (green), for year 2020.

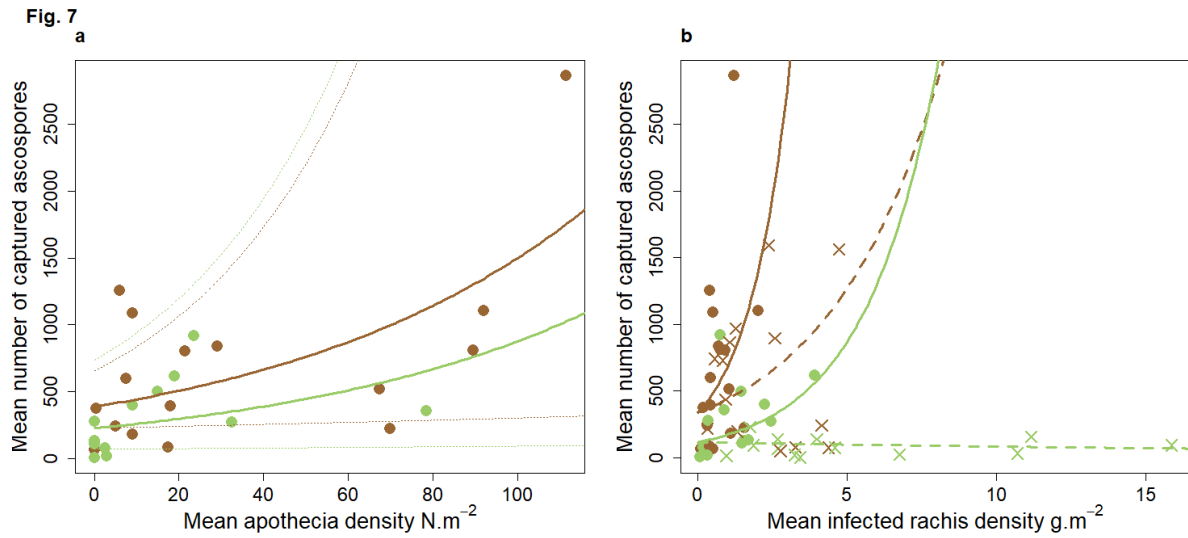
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600 **Fig. 6 a** Evolution of the fruiting rate as a function of the infected rachis density. In brown, measure from forest, in
601 green, measure from hedge and small wood with their Gompertz function associated with year 2012 (diamonds),
602 year 2016 (squares), year 2017 (triangles) and year 2021 (circles). **b** Estimated infection rate as a function of ash
603 basal area to reach the Allee threshold in forest (brown curve) and in hedge and small wood (green curve), with
604 measured infection rate in each situation in 2020 (cross) and 2021 (circle). Shaded areas correspond to 97.5%
605 confidence intervals

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608 **Fig. 7 a** Mean of ascospores captured in the ash stand according to the apothecia density for the year 2021. Brown
609 points represent forest environment and green points represent the hedge and small wood environment. **b** Rela-
610 tion between mean of ascospores captured and infected rachis density in forest environment (brown) and hedge
611 and small wood (green) during summer 2020 (cross points) and summer 2021 (circle points). Line represents the
612 regression for 2021 and dashed line year 2020 (brown: forest, green: hedge and small wood)

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