<u>Transcriptional control of *hgcAB* by an ArsR-like regulator in *Pseudodesulfovibrio mercurii*</u>

<u>ND132</u>

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ABSTRACT

The *hgcAB* gene pair encodes mercury (Hg) methylation capability in a diverse group of

microorganisms, but its evolution and transcriptional regulation remain unknown. Working from

the possibility that the evolutionary function of HgcAB may not be Hg methylation, we test a

possible link to arsenic resistance. Using model Hg-methylator *Pseudodesulfovibrio mercurii*

ND132, we specifically evaluated transcriptional control of *hgcAB* by a putative ArsR encoded

upstream and co-transcribed with *hgcAB*. This regulator shares homology with ArsR repressors

of arsenic resistance and S-adenosyl-homocysteine (SAH) responsive regulators of methionine

biosynthesis but is distinct from other ArsR/SahR in *Pseudodesulfovibrio mercurii* ND132.

Using qPCR and RNA-seq analyses we confirmed this ArsR regulates *hgcAB* transcription, and

is responsive to arsenic and SAH. Additionally, RNA-seq indicated a possible link between

hgcAB activity and arsenic transformations by *Pseudodesulfovibrio mercurii* ND132, with

significant up-regulation of other ArsR-regulated arsenic resistance operons alongside *hgcAB*.

Interestingly, wild-type ND132 was less sensitive to AsV (but not AsIII) than an *hgcAB*

knockout strain, supporting the idea that *hgcAB* may be linked to arsenic resistance. Arsenic

significantly impacted Hg-methylation rates by ND132, however, responses varied with culture

conditions. Differences in growth and overall metabolic activity did not account for arsenic

impacts on methylation. One goal of this research is to better predict MeHg production in nature.

However, we found that *hgcAB* gene and transcript abundance was not a good predictor of Hg-

methylation rates. Our finding that *hgcAB* activity is linked to arsenic may hold clues to the

possible environmental drivers of horizontal transfer of *hgcAB*.

IMPORTANCE

This work reveals a link between microbial mercury methylation and arsenic resistance and may

hold clues to the evolution of mercury methylation genes (*hgcAB*). Microbes with *hgcAB*

produce methylmercury, a strong neurotoxin that readily accumulates in the food web. This

study addresses a critical gap in our understanding about the environmental factors that control

hgcAB expression. We show that *hgcAB* expression is controlled by an ArsR-like regulator

responsive to both arsenic and S-adenosyl-homocysteine in our model organism,

Pseudodesulfovibrio mercurii ND132. Exposure to arsenic also significantly impacted

Pseudodesulfovibrio mercurii ND132 mercury methylation rates. However, expression of *hgcAB*

was not always a good predictor of Hg methylation rates, highlighting the roles of Hg

bioavailability and other biochemical mechanisms in methylmercury production. This study

improves our understanding of the controls on *hgcAB* expression which is needed to better

predict environmental methylmercury production.

INTRODUCTION

Microorganisms play a major role in controlling the form and fate of mercury (Hg) in aquatic

environments. Methylmercury (MeHg), an organic Hg compound, is a neurotoxin that

accumulates in food sources and is primarily produced by microorganisms(1). The *hgcAB* gene

pair that encodes Hg methylation capability(2) is widespread in anaerobic environments and

present in genomes spanning a diverse group of metabolic clades(3-5). Hg methylation capability

is a species-specific trait that is often associated with low abundant taxa in the microbiome(3, 6).

Currently, we do not have a good metric for estimating the prevalence of the mercury

methylation trait among microbiota. However, in environmental metagenomic studies of mercury

methylator environments, sequencing coverage of *hgcAB* genes is often low compared to other

functional genes(7). For example, the coverage of dissimilatory sulfite reductase genes (*dsr*) in

metagenomic assemblies is often 3-34x higher than *hgcAB*(7). Despite the association of sulfate-

reducing bacteria (SRB) with Hg methylation, the *hgcAB* gene pair is found in only ~12% of

publicly sequenced SRB in the Deltaproteobacteria (although *hgcAB* is more common within

some genera including *Pseudodesulfovibrio*)(8-10). Of the roughly 3,078 sequenced

Deltaproteobacteria genomes in the NCBI reference database, only 368 were identified as

potential Hg-methylators when the Hg-MATE database was compiled(7, 11). Within other phyla,

especially environmental metagenome assembled genomes from the *Bacteroidetes*, *Nitrospirae*,

PVC superphylum and candidate phyla radiation, *hgcAB* distribution further diverges from

species trees but follows environmental patterns and suggests horizontal gene transfer of this

capability(4, 6). Not all Hg-methylators share the same methylation capabilities, indicating that

Hg uptake, Hg(II) binding to HgcAB, *hgcAB* transcription, enzyme regulation, or protein

structure may control Hg methylation capacity and efficiency(12-17). Currently, we have a

limited understanding of the biochemical constraints on Hg-methylation and the environmental

conditions that resulted in the conservation of *hgcAB* in Bacterial and Archaeal genomes.

Despite the identification of two genes that encode Hg methylation capability, the

mechanism by which methylation is performed within the cell is largely speculative. The genes

encode a corrinoid (i.e., vitamin B_{12}-dependent) protein (HgcA), and a 2[4Fe-4S] ferredoxin

(HgcB), that could be used as a methyl carrier and electron donor in corrinoid cofactor reduction,

respectively (2, 18). Both genes (*hgcA* and *hgcB*) are required for Hg methylation (2, 19).

Corrinoid-dependent methyltransferases are utilized in a range of microbial metabolic pathways

in the methylation of amino acids, proteins, DNA, and other intermediates(20, 21). In anaerobic

microorganisms, corrinoid-dependent enzymes drive methionine and acetyl-CoA production (22,

23). Both pathways have been linked to Hg methylation, and the methyl transfer to Hg is thought

to occur via a similar mechanism (18, 24-27). HgcA consists of a cytoplasmic corrinoid binding

domain and a transmembrane domain. The cytoplasmic domain of HgcA shares homology with

the C-terminal domain of the large subunit of the corrinoid iron-sulfur protein (CFeSP) utilized

in methyl transfer in the Wood-Ljungdahl (i.e., reductive acetyl-CoA) pathway (2, 18).

Further, the native function(s) of HgcA and HgcB remaining unknown. Although HgcAB

function has been linked to one-carbon metabolism(28), we don't know if these proteins carry

out any functions independent of Hg methylation. Redox transformations of metals and

metalloids are often part of detoxification mechanisms or energy production. However,

biomethylation of Hg does not provide any observed benefit to the cell, and Hg methylation

capability is not induced by Hg(II)(28-30). Deletion of the *hgcAB* gene pair does not appear to

hinder cell growth or produce any major phenotype in *Pseudodesulfovibrio mercurii* ND132 (2,

28). While some photosynthetic bacteria have been shown to use Hg(II) as an electron acceptor

during photosynthetic and fermentative growth(31), Hg-methylating organisms are not able to

use Hg species as growth substrates(30). Resolving the native biochemical function of HgcAB is

needed to identify the metabolic conditions that mediate HgcAB activity and function.

Understanding the biological and geochemical factors that influence MeHg production is

necessary for predicting the impacts of contamination and remediation efforts on natural

systems.

Building predictive models for net MeHg production in aquatic environments is a

challenge(32). There are several microbial, physical, and geochemical parameters that influence

mercury biogeochemistry in aquatic environments, including but not limited to Hg complexation,

availability of terminal electron acceptors, shifts in redox, the composition or availability of

organic matter, and the presence, abundance, and activity of Hg-cycling microorganisms(1, 33,

34). While the *hgcAB* gene pair can be used to identify potential Hg-methylators in a

microbiome(6), to date, *hgcAB* gene abundance alone has rarely provided a useful indicator of

environmental MeHg production(35-38). Measuring metabolic and *hgcAB* activity of Hg-

methylators has been proposed to improve predictions(36, 39-42). However, it is unclear whether

we can relate *hgcA* activity to Hg methylation rates(29). There are several significant unknowns,

including whether *hgcAB* is transcriptionally regulated, what geochemical conditions influence

hgcAB transcription, and if higher expression of *hgcAB* transcripts increases MeHg production.

Previous work indicates that *hgcAB* transcription is not regulated by Hg(28-30). A possible

overlap with microbial mechanisms for arsenic resistance has been proposed, as several studies

have shown that *hgcAB* is co-localized with genes that encode for arsenic resistance in some Hg-

methylators(4, 8, 43), including an ArsR-like regulator. In Hg-methylating members of the

Pseudodesulfovibrio, this *arsR*-like gene is often found with *hgcAB*, and has been shown to be

co-transcribed with *hgcAB*(8).

In this study, we test the ability of the ArsR-like regulator co-transcribed with *hgcAB* to

enhance both *hgcA* transcription and Hg methylation itself in *Pseudodesulfovibrio mercurii*

ND132, a model Hg methylator(9, 30). Two transcriptional responses were tested, one to arsenic

using arsenite (As(III)) and arsenate (As(V)), and the second to S-adenosyl-homocysteine (SAH)

using methionine and adenosine-2',3'-dialdehyde. We chose these test conditions due to the

shared homology between this potential *hgcAB* regulator and helix-turn-helix regulons in other

sulfate-reducing bacteria that are responsive to As(III) (ArsR) and S-adenosyl-homocysteine

(SahR). These regulate arsenic resistance or methionine biosynthesis, respectively, a(44-46).

There are possible biochemical overlaps between Hg-methylation and S-adenosyl-methionine

(SAM)- mediated methylation(47, 48) , including those utilized in methyl-arsenic or methionine

biosynthesis. Experiments were designed to test whether *hgcAB* transcription is regulated by

ArsR/SahR.

In addition to evaluating potential *hgcA* regulators, another goal of this study was to test

the idea that Hg-methylation rates are related to *hgcAB* expression rates. We hypothesized that

geochemical culture conditions that increase *hgcAB* expression in *Pseudodesulfovibrio mercurii*

ND132 will lead to higher Hg methylation rates than analogous conditions. Outcomes of this

study have important implications for modelling Hg-methylation potential in the environment.

METHODS

Bacterial strains and culture media: For this study, we used the sulfate-reducing Hg-

methylator and incomplete-oxidizer *Pseudodesulfovibrio mercurii* ND132 (previously

Desulfovibrio desulfuricans ND132) as a model organism for understanding the mechanism and

genetic regulation of Hg methylation(9, 30). *P. mercurii* ND132 has a well annotated genome,

can grow on a number of substrates including lactate/sulfate, pyruvate/fumarate,

formate/acetate/sulfate, and formate/acetate/fumarate(49). Furthermore, knock-out strains

(Δ*hgcA,B*) of *P. mercurii* ND132 have been used to show the requirement of *hgcA* and *hgcB* for

Hg methylation(2, 19). For this study, we used *P. mercurii* ND132 wild-type and mutant glycerol

stock cultures from Dr. Dwayne Elias' culture collection. Since these experiments began, wild-

type *Pseudodesulfovibrio mercurii* ND132 has been deposited into DSMZ (type strain DSM

110689) and ATCC (ATCC TSD-224) culture collections. The Hg-methylation knock-out strain

(ΔhgcAB) of *P. mercurii* ND132 was originally sourced from Judy Wall[2]. Stock cultures were

stored at -80ºC in lactate-sulfate media and 10% v/v glycerol.

ND132 is capable of fumarate respiration using pyruvate as an electron donor[9]. For

most assays, *P. mercurii* ND132 was grown in this way on either CCM[50] or Estuarine

Pyruvate Fumarate (EPF) medium[9]. Both media contain 40 mM pyruvate and 40 mM

fumarate. CCM is a defined medium, while EPF includes 0.5% yeast extract. Both media were

buffered to pH 7.2 using MOPS and include vitamin and mineral additions. Media were reduced

with 0.1 mM (CCM) or 0.05 mM (EPF) cysteine-HCl to remove trace oxygen contamination.

Other amendments varied by experiment. All media contained resazurin as a redox indicator. For

some arsenic toxicity assays, ND132 was also grown on CCM media with lactate (60 mM) and

sulfate (30 mM) as electron donor and acceptor. All culture work was done at 31 ± 1 °C.

Experimental assays for testing regulators of Hg methylation and *hgcAB* transcription: We

evaluated the potential transcriptional response of a putative *hgcAB* regulator to S-adenosyl-

homocysteine (SAH) and arsenic. We also evaluated potential influence of *hgcA* expression on

Hg methylation using batch culture and washed cell approaches. For all assays, cultures were

started from freezer stocks. In all cases, cells were grown on CCM or EPF media. Out of the

freezer, cells were grown to mid-late exponential phase, then inoculated at 10% v/v into medium

containing one of the four test conditions, or unamended control medium. In defined CCM media

the test amendments were: 0.2 mM sodium arsenite, 2.0 mM sodium arsenate, 1 mM methionine,

0.1 mM adenosine-2'3'-dialdehyde, and 1 nM of HgCl. In EPF media, arsenic concentrations

were reduced to 0.02 mM sodium arsenite and 0.2 mM sodium arsenate. Amendment

concentrations for expression and methylation studies were chosen based on toxicity tests (see

below). Arsenic concentrations selected were the highest levels that allowed unimpacted growth,

which was specific to each medium. In qPCR expression experiments conducted over the growth

curve, both control and amendment cultures were grown without added Hg. In the washed-cell

assays and Hg-methylation assays performed in culture, all treatments including the control were

spiked with 1 nM isotopically enriched ^{201}Hg.

Methylation assays were carried out in triplicate by the addition of 1 nM isotopically

enriched ^{201}Hg (98.1% purity; Oak Ridge National Laboratory) (in 1% v/v HCl). Uninoculated

sterile medium with and without treatment amendments served as abiotic controls. Methylation

assays ran 100 minutes, after which aliquots of unfiltered culture were immediately acidified

(0.5% v/v HCl) and frozen (-20°C) for Me^{201}Hg analysis. Experimental timeframes were selected

on the basis of previous kinetic experiments demonstrating a plateau in MeHg production by

strain ND132(17).

Methylation assays were performed using either washed cells or batch culture cells. For

washed cell assays, cells were harvested at mid log phase (OD$_{600}$ ~ 0.2) from batch culture and

washed and resuspended in a minimal wash buffer as described previously(51). The salt buffer

contained 1 mM pyruvate and 1 mM fumarate, reduced with 0.5 mM cysteine-HCl, and buffered

to pH 7.5. For batch culture experiments, culture aliquots from mid log phase (OD$_{600}$ ~ 0.2) were

directly spiked with ^{201}Hg. In most cases, the concentration of filterable and total ^{201}Hg was

measured immediately after the ^{201}Hg spike, and at the end of the methylation assays (and in

some cases through time during the assays). Samples were filtered through 13mm diameter 0.22

μm CellTreat® MCE membrane syringe filters. For most studies we also measured pH, optical

density at 600 nm wavelength (OD$_{600}$), acridine orange direct cell counts (AODC), and sulfide

concentration over the course of the assay. Additionally, for some studies we measured short-

chain fatty acids and/or CO_2 production as indicators of cellular activity. All culture

manipulations were carried out in a Coy® glove bag under an O_2-free atmosphere.

Culture aliquots were saved for RNA extraction before the assays and from each assay

replicate at the end of the time-course (t = 100 minutes). For RNA extraction, cells were

harvested from batch cultures and methylation assays at 2,800 x g for 10 minutes. The cell pellet

was resuspended in 1 mL TRIzol™ Reagent (Invitrogen), and kept at -20°C prior to nucleic acid

extraction.

Hg and MeHg analyses: Isotope-specific total Hg and MeHg concentrations were measured

using modifications of EPA Methods 1631 (EPA-821-R-01-023, Mar 2001) and 1630 (EPA-821-

R-01-020, Jan 2001) respectively, with inductively coupled plasma mass spectrometry (ICP-MS)

detection. Details of the analytical method and QA/QC (Table S1) are provided in the SI.

Other analytical methods: Sulfide was analyzed using a sulfide selective electrode with an

Ag−AgCl reference electrode (Thermo Scientific) and was calibrated against Pb-titrated Na_2S

standards made in sulfide antioxidant buffer (SAOB). Saturated Na_2S stocks were made from

washed crystals in degassed ultra-pure water and held in N_2-filled sealed serum vials until use.

SAOB contains 2 M NaOH, 0.2 M EDTA and 0.2 M ascorbic acid, in degassed ultra-pure water

and sealed under N_2 until use. Organic acids (acetate, pyruvate, fumarate, and succinate) (0.5–

200 μM) were determined with a Dionex™ Integrion HPIC equipped with an AS11-HC column

at 35 °C with a KOH gradient of 1–60 mM with an ICS-5000$^+$ EG effluent generator system

(Thermo Fisher Scientific). Sample volumes of 10 μl were injected with a Dionex™ AS-AP

auto-sampler at room temperature.

Cell counts were done manually using acridine orange staining. Cells were fixed at a

dilution of 1:100, with 0.1 mL of cell culture preserved in 9.1 mL of 20 mM phosphate buffer

solution (PBS) and 0.8 mL glutaraldehyde, and stored in the fridge prior to slide prep. Slides

were made by combining 0.2 mL of 0.1% acridine orange and 1.9 mL of fixed cells in

microcentrifuge tubes and allowing 1 hr at room temperature in the dark for staining. The stained

cells were mixed with 1 mL of 20 mM PBS and collected on filters via vacuum. Counts were

performed manually and averaged from 10 different grids over a $0.01 \ mm^2$ area.

CO_2 production was measured over time in the headspace of serum vials containing

culture aliquots, after purging headspace w/ O_2-free N_2 for ~5 min. All measurements were made

from duplicate aliquots for each treatment. The 100 ml serum vials contained either 5 ml of

culture, or 25 mL of washed cells in salt buffer. Headspace CO_2 was measured hourly after an

initial equilibration period using a LICOR LI-7000 CO_2 analyzer. Total inorganic carbon

production was calculated from headspace CO_2 concentrations based on temperature, pH, the

Henry's Law constant for CO_2 and equilibrium among dissolved inorganic carbon (DIC) species.

DIC production rates were calculated from the linear portion of the slope with time

RNA extraction: Cell cultures were pelleted at 2,800x g for 10 minutes, excess media removed,

and then resuspended in 1 mL of TRIzol™ reagent (Invitrogen), and stored at -20 ºC until

processing. RNA was extracted following TRIzol protocol, which involves a phenol-chloroform

extraction, isopropanol precipitation, and ethanol cleaning step. Extracted RNA was further

purified using RNeasy Clean-up kit (Qiagen), and DNAse treatment with TURBO DNA-free™

Kit (Thermo Fisher). RNA concentration was measured using a Qubit™ RNA HS assay kit on

the Qubit 3.0 Fluorometer (Thermo Fisher), and purity was determined by NanoDrop (Thermo

Fisher). gDNA contamination was assessed by performing a qPCR with DNase-treated RNA

prior to reverse transcription. cDNA synthesis was performed using iScript™ cDNA Synthesis

Kit (Biorad), and cDNA concentration was measured using Qubit™ ssDNA assay kit on the

Qubit 3.0 Fluorometer (Thermo Fisher). cDNA was diluted to 1 ng/uL with RNase-free water

prior to qPCR.

Real-time quantitative PCR (qPCR): qPCR reactions were set up on a 96-well plate with each

standard and biological cDNA sample run in triplicate. Each well contained 1 ng of cDNA, 0.25

µM forward and reverse primer, 0.02 µL ROX reference dye (Bio-Rad), 10 µL iQ SYBR Green

Supermix (Bio-Rad), and 1xTE Buffer (pH 8), filled to 20 µL total volume. For standard curves,

we used serial dilutions of 4.2 ng/µL (measured on Qubit™) gDNA extracted from pure culture

P. mercurii ND132 wild type, estimated to be equivalent to 10^6 copies/µL based on the size of

the genome (3.8 Mbp)(49). All standards and samples of each primer set were amplified on the

same plate, with each plate only containing one target or reference gene. All qPCR analyses from

the same experiment were performed in succession with the same standard curve. Custom-

designed oligo primers were used to amplify 107 nt bp target region of *hgcA* (DND132_1056-

hgcA-F, DND132_1056-hgcA-R) and 180 nt bp region of the *arsR*-like gene before *hgcA*

(DND132_1054-arsR-F, DND132_1054-arsR-R) (Table S2). Three single-copy housekeeping

genes that encode recombinase subunit A (*recA*), gyrase subunit A (*gyrA*), and gyrase subunit B

(*gyrB*) were used for normalization of expression of genes of interest. These housekeeping genes

were chosen for normalization due to their observed expression stability in ND132 wild-type

transcriptomes (data not shown). Custom primers were designed to amplify a 153 nt bp region of

recA (DND132_1603, DND132_1603-recA-F, DND132_1603-recA-R), 148 nt bp region of

gyrA (ND132_2285, DND132_2285-gyrA-F, DND132_2285-gyrA-R), and 182 nt bp of *gyrB*

(DND132_2284, DND132_2284-gyrB-F, DND132_2284-gyrB-R). Primers were designed to

function over the same annealing temperature (65 °C), and the same thermocycler protocol was

used to amplify each gene: polymerase activation for 3 minutes at 95 °C, followed by 32 cycles

of denaturing for 15 seconds at 95 °C, and annealing for 20 seconds at 65 °C, followed by a melt

curve of 0.1 °C/s from 65 °C to 95 °C, ending with 15 seconds at 95 °C.

RNA-seq transcriptomics: RNA sequencing was performed on clean, DNase-treated RNA

extracted from the batch culture and washed-cell Hg-methylation assays. For the washed-cell

assays, only RNA from the control, As(III), and As(V) treatments were sent for RNA-seq. All

treatments from the culture Hg methylation assays were sent for sequencing. Three biological

replicates for each treatment were sequenced. Library prep and sequencing were performed by

Genewiz (Azenta Life Sciences, NJ). Ribosomal RNA (rRNA) depletion was performed during

library preparation using three probes from QIAGEN FastSelect rRNA 5S/16S/23S Kit (Qiagen,

Hilden, Germany), respectively. RNA sequencing libraries were prepared using NEBNext Ultra

II RNA Library Preparation Kit for Illumina by following the manufacturer's recommendations

(NEB, Ipswich, MA, USA). Briefly, enriched RNAs are fragmented for 15 minutes at 94 °C,

then first and second strand cDNA are synthesized. cDNA fragments are end repaired and

adenylated at the 3' ends, and universal adapters are ligated to cDNA fragments, followed by

index addition and library enrichment with limited cycle PCR. Sequencing libraries were

validated using the Agilent Tapestation 4200 (Agilent Technologies, Palo Alto, CA, USA), and

quantified using Qubit 2.0 Fluorometer (ThermoFisher Scientific, Waltham, MA, USA) as well

as by quantitative PCR (KAPA Biosystems, Wilmington, MA, USA). Multiplexed and clustered

sequencing libraries were sequenced using Illumina Hiseq 2x150 bp paired-end configuration.

Raw sequencing files were uploaded to KBase(52). Initial read quality was assessed with

FastQC, followed by adapter and quality-based trimming with Trimmomatic(53), alignment to

the *Pseudodesulfovibrio mercurii* ND132 reference genome

(https://www.ncbi.nlm.nih.gov/assembly/GCF_000189295.2) with HISAT2(54), assembly of

transcripts from alignment using StringTie(55). Differential expression analysis was performed

with DeSeq2(56) and the results were uploaded to Galaxy web platform(57), and we used the

public server at usegalaxy.org to produce figures using the Galaxy R scripts.

Arsenic analyses: Arsenic speciation analyses were done at Brooks Applied Labs (Bothell, WA,

USA). Samples were collected from culture bottles, filtered (Whatman® cellulose acetate

membrane syringe filters 0.45 μm) into BD Vacutainers®, preserved in EDTA and stored at 4 °C

in the dark until analyzed. Arsenic (As) speciation analysis, including arsenite [As(III)], arsenate

[As(V)], monomethylarsonic acid [MMAs], dimethylarsinic acid [DMAs], and unknown As

species were determined on all samples. Arsenic species were chromatographically separated on

an ion exchange column and then quantified using inductively coupled plasma collision reaction

cell mass spectrometry (ICP- CRC-MS). The results were not method blank corrected and were

evaluated using reporting limits adjusted to account for sample aliquot size.

Structural modeling: Structural models of the putative ArsR-like transcriptional regulator from

P. mercurii ND132 (DND132_1054, Uniprot ID: F0JBE8) were generated with ColabFold(58).

ColabFold is an implementation of AlphaFold2(59) that uses MMSeqs2(60) to generate large

multiple sequence alignments by searching the UniRef and MGnify databases(61). Based on its

sequence similarity to other transcriptional regulators that are known to function as homodimers,

we modelled the homodimeric form of the protein using AlphaFold-multimer(62), which is a

version of AlphaFold2 trained specifically to model protein complexes. Models were ranked by

their average predicted local distance difference test (pLDDT) and predicted TM-score

(pTMscore) and the model with the highest pTMscore score was selected for further analysis.

Statistical methods: Statistical modelling was performed in JMP®, Version 16 (SAS Institute

Inc., Cary, NC, 1989–2022) and R statistical software (version 4.2.0)(63). Generalized linear

models were used to evaluate differences in Hg methylation rates, inorganic carbon production,

organic acid production, and growth rates among treatments. Where appropriate, models

included time, treatment and any potential interaction terms. In the absence of significant

interactions, differences among treatments were evaluated using pairwise comparison of slopes.

For qPCR results, a two-sample z-test was used to determine if the difference in *arsR* and *hgcA*

expression was significant between treatments and control. For significance testing, the mean

and standard deviation of ΔΔCq values for each target gene and treatment were compared to

control.

RESULTS AND DISCUSSION

An ArsR-like regulator encoded with HgcAB in *Pseudodesulfovibrio*: A putative arsenic

responsive regulator (ArsR) (Gene ID: DND132_1054, Uniprot: F0JBE8) is encoded directly

upstream of Hg-methylation genes *hgcA* (Gene ID: DND132_1056, Uniprot: F0JBF0) and *hgcB*

(Gene ID: DND132_1057, Uniprot: F0JBF1) in the genome of *P. mercurii* ND132. Potential

ArsR-controlled transcriptional regulation of *hgcAB* has been proposed for Hg-methylating

members of the *Pseudodesulfovibrio*(8). This potential *hgcAB* regulator shares homology to the

metalloregulator ArsR/SmtB protein family (Pfam PF01022), including a DNA-binding helix-

turn-helix (HTH) motif. This family of regulators (Pfam PF01022) are responsive to metal ion

stress (e.g., As(III), Zn(II), Cd(II)) and in the absence of coordinating metal(loid)s will repress

the expression of operons linked to heavy metal resistance(64). This transcriptional family

includes the ArsR repressor of the *ars* operon, which often encodes an arsenate reductase with

arsenic permeases and exporters(65). However, protein sequence analysis and predictive

structure modelling of the *hgcAB* ArsR-like regulator suggests differences in the metal binding

domain that may make it distinct from other ArsR family regulators.

There are five ArsR family transcriptional regulators annotated in the genome of

Pseudodesulfovibrio mercurii ND132 (Figure 1). Based on protein sequence analysis, these five

ArsR proteins are different from one another, with less than 33% pairwise identity between them

(Table S3). We numbered the *arsR* by order of appearance in the genome of *P. mercurii* ND132,

and *arsR2* is located directly upstream of *hgcA* and *hgcB* (Figure 1). While in some Hg-

methylating bacteria, *hgcAB* is located in the genome with arsenic resistance (*ars*) genes(4, 43),

no other *ars* genes are colocalized with *hgcAB* in ND132. However, three of the other *arsR*

(*arsR1*, *arsR3*, and *arsR5*) are present with *ars* genes elsewhere in the genome and likely

regulate arsenic resistance. The first set of *ars* genes encode ArsR1 (GeneID: DND132_0890)

and the first of two methylarsenite efflux permeases, ArsP (DND132_0889)(66). Another *ars*

operon encodes ArsR3 (DND132_1317) with the first of two Acr3 family As(III) permeases

(DND132_1318) and arsenate reductase, ArsC (DND 132_1319). *P. mercurii* ND132 has a

second Acr3 family As(III) efflux permease that is not part of an *ars* operon (DND132_2196).

The third *ars* operon is predicted to encode ArsR5 (DND132_2311), another ArsP

(DND132_2312), an organoarsenical resistance protein, ArsE (PFam: PF13192, thioredoxin-like

protein) (DND132_2313)(67), and a protein of unknown function with a diguanylate cyclase

(DGC) motif (Pfam: PF08859, a domain with four conserved cysteines that likely coordinate Zn)

(DND132_2314). ArsR repressors of the *ars* operon are predicted to be transcriptionally

regulated by As(III) and part of a detoxification pathway for arsenic. From protein sequence

analysis of ND132 ArsR2, it is unclear whether expression of *hgcA* and *hgcB* is part of the ArsR

regulon and controlled by arsenic.

ArsR/SmtB family regulons may not always be associated with heavy metal resistance or

arsenic transformations. In addition to regulating arsenic resistance, ArsR regulators have also

been shown to influence many other cellular functions in the presence of As(III), including

phosphate acquisition/metabolism, sugar transport, chemotaxis, copper tolerance, and iron

homeostasis(68). Regulators in the ArsR/SmtB HTH family may also be responsive to substrates

other than heavy metals and metalloids. In sulfate-reducing bacteria, an ArsR-like regulon

responsive to intracellular concentrations of S-adenosylhomocysteine (SAH) controls SAM-

dependent methyl transfer for methionine biosynthesis(44, 45). These SAH-responsive regulators

(SahR) are often encoded with adenosylhomocysteinase (AHCY), which catalyzes the hydrolysis

of SAH to homocysteine and adenosine. In *P. mercurii* ND132, *arsR4* likely encodes a SahR

(Gene ID: DND132_1481, Uniprot: F0JE80). The predicted SahR is encoded with AHCY (Gene

ID: DND132_1480, Uniprot: F0JE79) (Figure 1), and has an ArsR-like HTH motif with a SAM-

dependent methyltransferase similar to ArsM. Arsenite methyltransferase (ArsM) catalyzes the

transfer of a methyl group to arsenite from S-adenosylmethionine (SAM) to form

methylarsenicals, and is part of the ArsR regulon(69). Instead of methylarsenical production,

ND132 ArsR4 is predicted to control methionine production and be responsive to SAH. No other

homologues of ArsM are encoded in the genome of *P. mercurii* ND132. Like *arsR2* which is co-

localized with *hgcAB*, *arsR4* is not part of an *ars* operon in *P. mercurii* ND132.

Phylogenetic distribution of the ArsR2 *hgcAB* regulator: Close homologues of the ArsR2

regulator from *Pseudodesulfovibrio mercurii* ND132 are present in other *hgc+ Desulfovibrio* and

Pseudodesulfovibrio genomes (Figure S1). To identify potential ArsR2 homologues, we searched

for ArsR family proteins encoded before *hgcAB* in *Desulfovibrio* and *Pseudodesulfovibrio*

reference genomes from a previous study(9). We also used the HMM of the HTH motif of ArsR

(Pfam: PF01022, HTH_5.hmm) to broaden our search and identify all potential HTH family

transcriptional regulators encoded in the reference genomes. We used protein sequence

alignments and maximum likelihood phylogenies of HMM-hits to identify ArsR2 homologues.

Among these genomes, we found ArsR2 homologues only in *hgc+ Desulfovibrionaceae*

genomes, and always directly upstream of *hgcAB* (Figure S1). Other researchers working with a

close relative to ND132, *Pseudodesulfovibrio hydrargi* BerOc1 demonstrated that these *arsR-*

like genes and *hgcA* are co-expressed(8). Notably, all but one of the available

Pseudodesulfovibrio genomes encode the ArsR2 homologue. However, co-occurrence of *arsR*

with *hgcAB* does not appear to be phylogenetically conserved in the majority of *hgc+* genomes

of close relatives of *Desulfovibrionaceae* (Figure S1).

ArsR family regulators are also encoded with *hgcAB* in genomes from outside of the

Desulfovibrionaceae family. In *hgc+* genomes from the Hg-MATE-Db(7, 11), we identified

putative ArsR/SmtB family regulators of *hgcAB* in isolate and environmental metagenome-

406 assembled genomes (MAGs) from Deltaproteobacteria, Nitrospira, Spirochaetes,

407 Ignavibacteriae(70) and Bacteroidetes(71), including *Paludibacter jiangxiensis* NM7(72) (Table

408 S4). Based on protein sequence alignments, not all of these putative *hgcAB* regulators are close

409 homologues of the ArsR2 protein from *Pseudodesulfovibrio mercurii* ND132. In some of these

410 genomes, *hgcA* and *hgcB* are also colocalized with other *ars* genes (e.g. *arsC, arsB, arsM*). It is

411 noteworthy that many of the *hgc+* MAGs(73) of sediment and groundwater-associated bacteria

412 from the Rifle IFRC site(74) have an ArsR-like regulator encoded with *hgcAB* and *ars* genes.

413 This co-occurrence of *hgcAB* with *ars* genes in MAGs from arsenic-impacted environments has

414 been noted previously(4, 43). For this study, we focused our analytical efforts on understanding

415 possible ArsR-regulated transcription of *hgcAB* in *Pseudodesulfovibrio mercurii* ND132.

416 However, more analysis is needed to explore a possible evolutionary relationship between the

417 *ars* operon and Hg-methylation genes, and whether horizontal gene transfer of *hgcAB* may have

418 been influenced by arsenic.

420 **Structural model:** The ArsR2 homologues from *Pseudodesulfovibrio* and *Desulfovibrio* are

421 phylogenetically distinct from other ArsR/SmtB family transcriptional regulators (Figure S2).

422 The protein sequence alignment shows that ArsR2 has some of the conserved invariant residues

423 associated with the HTH family of regulators (Figure S2)(75). To provide additional insight into

424 the structure of the putative *hgcAB* transcriptional regulator (ArsR2, DND132_1054), we

425 generated a homodimeric model of the protein using AlphaFold2(58, 59) (Figure 2). The

426 cysteine-rich site associated with metal(loid) binding is different in ArsR2 compared to ArsR

427 repressors of the *ars* operon. In *E. coli* ArsR (Uniprot: P15905), the As(III) binding site is

428 associated with the three cysteines on the α3 helix (Cys32, Cys34, and Cys37)(64). The

cysteines in ND132 ArsR2 (Uniprot: F0JBE8) occur on a coil after helix 2 and a beta strand

before helix 4 (Figure S3). The ArsR2 from ND132 includes a cysteine (Cys26) near the N-

terminus and four cysteines, Cys47, Cys49, Cys86, and Cys88 (Figure 2 and Figure S3) in each

predicted metal(loid) binding site(75). Homologs from several other closely related bacteria of

the *Pseudodesulfovibrio* and *Desulfovibrio* have similar structures and cysteine positions (Figure

S4), suggesting that some of them may have similar functions. The four-cysteine motif in the

metal(loid) binding site is conserved in all ArsR2 homologs found in *Pseudodesulfovibrio* and

Desulfovibrio genomes (Figure S4) except for *Desulfovibrio brasiliensis* (WP_157054761.1),

which has a three-cysteine motif.

The metal(loid) binding domain of the putative ArsR of *hgcAB* in *Pseudodesulfovibrio*

mercurii ND132 (ArsR2, DND132_1054) differs from other ArsR repressors, containing four

rather than three cysteines. Arsenite (As(OH)$_3$) is predicted to coordinate with the three cysteine

residues of ArsR as As(III), after the loss of its hydroxide ligands(75). However, only two of the

three cysteines (Cys32, Cys34) are needed to bind arsenic to ArsR of *Escherichia coli* plasmid

R773(76). The presence of four cysteine residues in the metal(loid) binding sites of the ArsR2

regulator suggests it may preferentially bind metal(loid) cations such as Zn(II), Cd(II), and Pb(II)

like the four-cysteine residues in the binding motif of CadC(75). Alternatively, there could be 3-

coordinate binding of As(III) to three of the four cysteine residues as has been proposed with

As(III) and the SAM methyltransferase, ArsM(77).

Arsenic tolerance in ND132: The presence of a putative ArsR regulator upstream of *hgcAB*

suggests that the ability to methylate Hg may confer As resistance in some microbes. We

compared As toxicity between wild-type (WT) ND132 and the *hgcAB* deletion mutant

(ND132Δ*hgcAB*) to test the idea that the presence of *hgcAB* in the ND132 genome might

contribute to As resistance. We performed growth assays in defined (no yeast extract) CCM-PF

and CCM-LS media. Cultures were inoculated into media amended with sodium arsenate (0.2,

0.5, 1, 2, 5, or 20 mM) or sodium arsenite (0.02, 0.2, or 0.5 mM) and grown at 30 °C to

stationary phase (Figures 3 and S5).

The WT strain was more tolerant of As(V), than the knockout. Specifically, WT ND132

growth was unaffected by As(V) levels up to 1 mM on both media, while growth of

ND132Δ*hgcAB* was significantly impacted at that level (Figure 3). Results were evaluated

statistically by normalizing OD_{600} measurements to matched unamended controls, and then

evaluating differences in normalized OD_{600} between the strains. Data from the exponential part

of the growth curves were used in the statistical tests. Differences were assessed using general

linear models that accounted for As concentration, time, and any significant interactions. The

hgcAB knockout was significantly more sensitive to As(V) at levels above 0.2 mM on PF

medium and at levels above 0.5 mM on LS medium (p <0.0001). Notably there was apparent

hormesis in growth of ND132Δ*hgcAB*, with slightly enhanced growth at the lowest tested AsV

levels.

The WT and knockout strains were equally resistant to As(III), although As(III) toxicity

to both was roughly 5-10 times higher than As(V) (Figure S5). Culture chemistry, particularly

the formation of As-sulfide complexes, impacted As toxicity. Both wild-type and

ND132Δ*hgcAB* cultures had lower tolerance to both As species when grown under sulfidogenic

conditions. A yellow mineral precipitate was observed when media with 0.5 mM As(III) was

inoculated with sulfidogenic *P. mercurii* ND132 WT and Δ*hgcAB* cultures, and likely was the

formation of orpiment (As_2S_3).

The reduced As(V) resistance of the *hgcAB* knockout strain suggests that *hgcAB* may

confer some As(V) resistance. Or alternatively, the loss the *hgcAB* from the genome may impact

other mechanisms of As(V) resistance by ND132. This observation deserves more exploration,

especially the lack of differential resistance to the more toxic As(III).

ND132 transformation of As(III) and As(V): *Pseudodesulfovibrio mercurii* ND132 has several

ArsR-regulated arsenic resistance operons (Figure 1). These operons are predicted to encode

As(V) reductase, As(III) and methylarsenite permeases, and a putative organic arsenic resistance

protein. To determine how ND132 may transform arsenic (As), we measured As speciation and

mass balance in *P. mercurii* ND132 WT cultures grown with either 0.02 mM sodium arsenite or

0.2 mM sodium arsenate in defined CCM-PF media. At these concentrations, As did not affect

cell growth rate (Figure S6A). We did not examine whether ND132 gains energy from As(V)

reduction.

Under these conditions, we found that wild-type *P. mercurii* ND132 reduced As(V) to

As(III) and produced potential thioarsenicals from both As(III) and As(V) (Figure S6, Table S5).

As concentrations and speciation were measured initially and at the end of exponential growth (7

days). ND132 reduced about 25% of added As(V) to As(III) after 7 days, and converted about

3% to an unidentified As compound. This compound was tentatively identified as a thioarsenical

based on the retention time of the peak (personal communication from Ben Wozniak at Brooks

Applied). ND132 converted a higher fraction (roughly 15%) of added As(III) to the presumed

thioarsenical compound (Figure S6).

In both As(III) and As(V) amended cultures, As speciation at the initial time point

matched the spikes, and 70-80% of the added As could be accounted for after 7 days growth.

Uninoculated medium served as abiotic controls. There was no significant loss of As or change

in As speciation in abiotic controls, including no production of unidentified As compounds. No

monomethyl- or dimethylarsenicals were detected in abiotic or cultured samples.

Potential transcriptional regulation of *hgcA* by ArsR2: To help determine the transcriptional

controls on the putative ArsR regulator of *hgcAB*, we challenged cells with As and SAH and

evaluated expression of both *arsR2* and *hgcA*. We chose these potential regulators because,

despite shared homology, the putative *hgcAB* regulator ArsR2 is distinct from other HTH

regulators of the ArsR/SmtB family, and its function is unknown. Additionally, since the four

other ArsR annotated in the genome of *P. mercurii* ND132 are predicted to regulate either

methionine production or As resistance, we speculate that there could be biochemical overlaps

between HgcAB function and As resistance or methionine production. We made this inference

based on the co-localization of *hgcAB* with other *ars* genes(4, 43), and the shared homology of

HgcA/B to corrinoid iron-sulfur proteins utilized in one-carbon metabolism(18, 78), including

methionine biosynthesis. Thus, one goal of this work was to test the idea *hgcAB* may be part of

the ArsR or SahR regulons. Because of the unusual configuration of the metal(loid) binding

motif in ArsR2, we chose to test ND132 response to As(V) as well as As(III) and SAH.

We used real-time qPCR to test relative changes in expression of *arsR2* and *hgcA* during

growth with potential transcription regulators (e.g., As(III), As(V), SAH). Tests were conducted

in defined (i.e., no yeast) CCM media or yeast-enriched EPF media. Expression was normalized

to three housekeeping genes and compared to control cultures without added treatments. We also

tested for an effect of Hg on *hgcA* transcription and found none (Figure 4). This observation is

consistent with previous work that *hgcA* expression and Hg-methylation activity are not induced

by Hg(28-30). To evaluate a potential *arsR2* transcriptional response to SAH we tested several

compounds in the methionine biosynthesis pathway. Because *P. mercurii* ND132 is not

predicted to have an SAH uptake mechanism, we tested for transcriptional response to SAH by

growing cells with methionine (1 mM), adenosine-2',3'-dialdehyde (0.1 mM), and equimolar

homocysteine with adenosine (0.2/0.2 mM, 0.5/0.5 mM, and 1/1 mM), which enabled the build-

up of SAH in the cell. Adenosine-2',3'-dialdehyde is an inhibitor of SAM-dependent

methyltransferase. Methionine, homocysteine, and adenosine are metabolites in the SAM cycle,

and higher concentrations in culture media should contribute to higher concentrations of SAH

compared to cultures grown without these substrates. Growth of *P. mercurii* ND132 was

completely inhibited at all concentrations of homocysteine with adenosine. Therefore, this

treatment was not used for further expression or Hg methylation experiments.

Transcription of *hgcA* is controlled by arsenic and SAH: We found that ArsR2 regulated

hgcA expression in *P. mercurii* ND132 and can be responsive to both As and SAH. We observed

significantly higher expression (z-test, one-tailed p-value < 0.05; Table S6) of *arsR2* and *hgcA* in

P. mercurii ND132 cultures compared to control when grown with an SAH-hydrolase blocker

(i.e., adenosine-2',3'-dialdehyde) and with As. The transcriptional response varied depending on

culture conditions (defined vs. enriched media), and growth phase (i.e., early, mid, or late

exponential) (Figure 4). In defined CCM media, we observed significantly higher expression

(Table S6) of *arsR2* and *hgcA* in cultures grown with As(V) and As(III) across the growth curve

(Figure 3). We only observed significantly higher expression of *arsR2* and *hgcA* from adenosine

dialdehyde at late exponential phase (Figure 4). Relative expression of *arsR2* and *hgcA* in

enriched media differed from defined media conditions. In enriched media, we observed

significantly higher *hgcA* expression in As(V)-amended cultures early in exponential growth, and

later in exponential growth for As(III) amended cultures (Table S6). Compared to defined media,

we observed significantly higher *hgcA* expression in methionine and adenosine amended cultures

at earlier stages of growth (Figure 4). We attribute some of these differences to the presence of

additional sulfur-containing metabolites in yeast extract (e.g., methionine) that could have

contributed to increased SAH build-up in cultures in enriched media when grown with

methionine or adenosine dialdehyde. Without yeast extract, the 1 mM methionine likely did not

cause enough SAH build-up to induce a transcriptional response. Unlike in defined media

conditions, *arsR2* transcriptional changes did not mirror *hgcA* expression patterns.

Differences in *hgcA* expression observed across the growth curve could indicate that the

transcriptional response is regulated by an unknown substrate or by an As species produced by *P.*

mercurii ND132. This effect could vary over the growth curve, as the substrate or products of As

metabolism build up in culture or the cell over time. Arsenic speciation analyses showed that *P.*

mercurii ND132 reduces As(V) to As(III), and produces an unknown thioarsenic species from

As(III) (Figure S6). Differences in expression profiles between media could also be attributed to

differences in As concentrations, with 10X less As(III) and As(V) used in enriched EPF media

than defined CCM media. Reducing the amount of starting As(III) from 0.2 mM to 0.02 mM

could have increased the amount of time needed to reach a threshold for transcriptional response.

The lower As(III) concentration used in EPF media may explain why increased *arsR2* and *hgcA*

expression was only observed in As(III)-amended EPF cultures at late exponential growth

(Figure 4). The higher starting concentration of As(V) (0.2 mM) compared to As(III) (0.02 mM)

in enriched EPF, may account for the earlier observed transcriptional response in As(V)-

amended cultures. Overall, these results show that growing *P. mercurii* ND132 with As(III) or

As(V) will significantly increase *hgcA* expression.

Metabolites and media components can confound expression analysis: When cells were

washed of culture media and then resuspended in basal salts with treatments, we observed

significantly higher *arsR2* and *hgcA* expression due to As but not SAH (Figure 5). Washing cells

of media components removed effects of adenosine dialdehyde on *arsR2* and *hgcA* expression

observed in culture. Cells that were grown in enriched media with the various treatments were

harvested at mid-log ($OD_{600} = 0.2$; Figure 4B), kept anaerobic, washed of media, and

resuspended in a phosphate-salt buffer with the different treatments and 1 nM Hg. This was the

same culture in which we measured *arsR2* and *hgcA* expression (Figure 4, EPF mid-log panel).

The difference in observed *arsR2* and *hgcA* expression in washed cells (Figure 5C), compared to

the culture (before washing) (Figure 4) indicates that build-up of SAH (or another metabolite)

during growth with methionine and adenosine dialdehyde likely causes the observed

transcriptional response of *arsR2* (DND132_1054). The transcriptional response to As, either as

As(III) and or As(V), is more direct and instantaneous. We attribute differences in the

transcriptional response to As over the growth curve as resulting from changes in As speciation

and toxicity. The qPCR results from washed cells and batch cultures indicate that the *hgcA*

regulator (DND132_1054) is responsive to both As and SAH.

As affects Hg methylation rates in ND132: We directly measured the impact of As and SAH

on Hg methylation to help confirm that changes in *hgcAB* expression affected Hg methylation

rates. Since *hgcA* transcription is regulated by *arsR2* in response to As and SAH, we expected As

and SAH should also impact Hg methylation. Arsenic significantly affected *P. mercurii* ND132

Hg-methylation rates, but effects varied depending on whether cultures were washed of media

components prior to methylation assays. We performed Hg-methylation assays in both cultures

and washed cells using mid-log cultures ($OD_{600} \sim 0.2$, Figure 4B) grown in EPF. Prior to

methylation assays, cells were inoculated into media amended with observed expression

regulators of *hgcA*: methionine, adenosine dialdehyde, As(III), or As(V). In washed cells

harvested at mid-log phase, methionine and As(V)-treated cells produced significantly more

MeHg than the control, while As(III) and adenosine significantly reduced MeHg production

(Figure 5A). In unwashed cultures, Hg methylation rates were significantly higher in As(III)-

amended cultures, but significantly lower in cultures grown with As(V) or adenosine, compared

to control (Figure 5B). In batch culture assays, there was no significant difference in MeHg

produced in cultures grown with methionine compared to control.

Differences in metabolic activity and biomass could account for some of the differences

in MeHg production between treatments. When grown with As(V) and methionine, *P. mercurii*

ND132 had a steeper growth curve and reached a higher optical density compared to control and

other treatments (Figure 4B). In the washed-cell assays, the higher metabolic activity (CO_2

respiration rate) (Figure S7) in the methionine treatment could explain some of the higher MeHg

production (Figure 5A). In As(V)-treated washed cells, higher MeHg production could not be

explained by differences in metabolic activity, biomass, or unfiltered Hg at T_0 (Figure S8, Figure

6). Normalization of methylation rates to various measures of cell growth and metabolism did

not change the relative impact of treatments (Figure 6). The measured initial concentration of

total ^{201}Hg was the same in all bottles, and Hg mass balance was maintained in all bottles (Figure

S9). However, the filterable (0.22 µm filters) ^{201}Hg concentration varied among the treatments.

Nevertheless, normalization of methylation rates to initial filterable ^{201}Hg concentrations did not

alter the relative impact of treatments (Figure 6).

In the Hg-methylation assays performed with mid-log cultures, significantly more MeHg

was produced in the As(III) treatment (Figure 5B). This effect held true even when normalized to

biomass (OD_{600}, cell counts) (Figure S10), pH (Figure S11), metabolic activity (Figure S12), and

total Hg (Figure S13) (Figure 5). Hg-methylation rates in As(V)-amended cultures remained

significantly lower than control when normalized for these factors (Figure 6). In washed cells

and culture media, adenosine-2',3'-dialdehyde appears to limit cell growth and metabolic

activity of *P. mercurii* ND132 (Figure 4). MeHg production in adenosine-amended washed cells

and batch culture assays were only slightly higher than abiotic controls. It was difficult to

decouple the effects of lower metabolic activity and less biomass on limited MeHg production.

However, normalizing for biomass (as cell density) put Hg-methylation rates on par with As(V)-

amended cultures in media (Figure 6).

Differences in Hg bioavailability may contribute to the differences in MeHg production

among treatments. Sulfide has strong effects on Hg bioavailability for methylation(12, 30, 79,

80) and, although grown without sulfate, ND132 cleaves the S−C bond in cysteine provided in

growth medium to release small amounts of sulfide, generally below 10 μM(51). We consistently

measured significantly less sulfide in filtered samples from As(V)-amended washed-cell (Figure

S7) and batch culture assays (Figure S11) compared to other treatments, including As(III). This

observation indicates that As(V) impacts sulfide speciation, possibly by disrupting metabolic

pathways that produce reduced sulfide, or precipitation of As species with sulfide. The higher

concentrations of As(V) (0.2 mM) could account for the observed differences from As(III) (0.02

mM). In addition, the dissolved ^{201}Hg spike is rapidly lost to cell walls and any precipitates in the

medium. We observed differences in filterable ^{201}Hg among treatments in both washed cells

(Figure S9) and batch cultures (Figure S13). To test the impact of amendments on bioavailable

Hg, we normalized methylation rates to the initial concentration of ^{201}Hg in each treatment.

Because Me^{201}Hg is exported from cells after production(17), filterable total ^{201}Hg

concentrations after T_0 often reflect Me^{201}Hg production (Figure 6).

Hg methylation rates decoupled from *hgcA* transcription: Increased *hgcAB* transcriptional

activity did not explain increased MeHg production in most treatments. With the exception of the

As(V)-washed cells, we did not observe significantly higher Hg methylation by cells with higher

hgcA expression compared to control (Figure 5). Despite seeing significantly higher MeHg

production by washed cells amended with methionine, and cultures grown with As(III), we did

not observe significant differences in *hgcA* expression compared to control in qPCR analyses

(Figure 5). Washed cells amended with As(III) had significantly higher *hgcA* expression (Figure

5), but produced significantly less MeHg (Figure 6). Similarly, cultures grown with adenosine-

2',3'-dialdehyde and As(V) treatments had higher expression of *hgcA*, but produced significantly

less MeHg compared to control (Figure 5). This finding was surprising as we expected increased

hgcA expression to result in higher Hg-methylation rates. However, *hgcA* activity may not have

predicted higher Hg-methylation rates because of several factors that we could not account for in

this experiment, 1) increased transcription of *hgcAB* may not have resulted in increased

translation of HgcAB, 2) Hg availability and uptake by cells may be more rate-limiting than

hgcA/HgcA activity, 3) Hg complexation with As may interfere with Hg availability for

methylation(81), 4) possible competition between As and Hg for binding to HgcAB.

Furthermore, the increased Hg methylation rate observed in As(III)-amended cultures could be

from increased activity of other unknown proteins essential to the biochemical Hg methylation

reaction.

As and adenosine increase *hgcA* and *hgcB* expression in ND132 transcriptomes: RNA-seq

analyses of RNA extracted from mid-log washed cells and batch cultures confirm that both *hgcA*

and *hgcB* expression is regulated by As and SAH. Growing *P. mercurii* ND132 with As or

adenosine dialdehyde appeared to have a significant influence on the transcriptome (Figure S14).

We observed significantly higher expression ($|\log FC| > 1.5$, $p < 0.05$) of *arsR2*

(DND132_1054), *hgcA* (DND132_1056) and *hgcB* (DND132_1057) in As(III) and As(V)-

treated washed cells compared to control (Figure 7). These features were also significantly

differentially expressed ($p < 0.05$) in batch cultures amended with As(V) and adenosine

dialdehyde (Figure S15). The RNA-seq data show that *hgcB* is being up-regulated with *hgcA*,

which is important since both genes are needed for Hg methylation(19). Overall, there were more

significant shifts in batch culture transcriptomes compared to washed-cells, with many of the

same features being differentially expressed across treatments (Figure S14). Methionine

amended batch cultures had the least significant shift in its transcriptome compared to control (n

= 4; Figure S16). In *P. mercurii* ND132 cultures grown with adenosine, the predicted SahR

regulator encoded by *arsR4* (Gene ID: DND132_1481) was significantly upregulated ($\log FC =$

2.36, $p = 2.64$ E-19) along with the adenosylhomocysteinase-encoding gene *ahcY* (Gene ID:

DND132_1480, $\log FC = 2.31$, $p = 2.16E-16$) (Figure S16). The SAH responsive regulator was

not significantly upregulated in any other transcriptome, including methionine-treated cultures.

This observation indicates that there may not have been enough build-up of SAH in *P. mercurii*

ND132 cultures amended with 1 mM methionine to induce a transcriptional response from SahR

or ArsR. There appears to be a cross-reactivity to SAH and As by two ArsR encoded in the

genome of *P. mercurii* ND132 *arsR2* (DND132_1054) as well as *arsR5* (DND132_2311)

(Figure S15). However, exposure to As(III) or As(V) did not induce a transcriptional response by

SahR (*arsR4*, DND132_1481).

Co-expression of *hgcAB* with operons of arsenic resistance: *Pseudodesulfovibrio mercurii*

ND132 is resistant to As(V) and As(III), and may also be resistant to organoarsenicals and

respire As(V). Both Hg-methylation genes (*hgcA* and *hgcB*) appear to be part of the ArsR

regulon and are co-regulated with *ars* operons. Based on RNA sequencing, there were 12 shared

features (Figure S14) that had significant differential expression ($|\log FC| > 1.5$, p-value < 0.05)

in As-treated washed cells compared to control (Figure 7). These features were also significantly

differentially expressed (p < 0.05) in batch cultures amended with As(III), As(V), and adenosine

dialdehyde (Figure S15). In addition to increased *hgcA* and *hgcB* expression, two ArsR-regulated

ars operons had significantly higher expression in As-amended washed cells and batch cultures

(Figure 7, Figure S15). Operon B (Figure 8) encodes for ArsR3, an Acr3 As(III) efflux

permease, and arsenate reductase. Another set of upregulated genes (Operon C) encodes ArsR5

and enzymes homologous to those utilized in organoarsenical resistance. Operon C was also

upregulated in the transcriptome of adenosine dialdehyde cultures. The other ArsP

(DND132_0889) and Acr3 (DND132_2196) permeases encoded outside an operon were not

consistently upregulated with the *ars* operons (Figure 7, Figure S15). The only *arsR* gene of *P.*

mercurii ND132 that was not significantly differentially expressed in any transcriptome was

arsR1 (DND132_0890).

In As(III) washed cells and As(V) and adenosine batch cultures, an operon with possible

links to arsenic or polysulfide respiration was upregulated compared to control. The set of genes

(Operon D, Figure 8) encodes possible homologues of the As(V) respiratory reductase (Arr) or

polysulfide reductase (Psr) complexes. Since transcription of these genes was increased by

As(III), we speculated that their encoded function may relate to As respiration. The genes encode

homologues of ArrA (DND132_2597), ArrB (DND132_2598), and ArrC (DND132_2599), and

are respectively annotated in *P. mercurii* ND132 as a molybdopterin oxidase, an Fe-S cluster-

containing hydrogenase, and formate dehydrogenase gamma subunit protein. The Arr complex is

related to the DMSO-molybdopterin (DMSO-MPT) reductase family, with Arr subunit A having

a Mo/W-*bis*(pyranopterin guanosine dinucleotide) cofactor in its active site and a [4Fe-4S]

center, and Arr subunit B with four [4Fe-4S] clusters(82). A diverse range of microorganisms

can respire As(V) for energy using ArrAB(83), including *Chrysiogenes arsenatis*, *Shewanella*

strain ANA-3, *Alkaliphilus oremlandii* and *Geobacter* sp. Strain OR-1(84, 85). In some

microorganisms, ArrAB can function as a bidirectional enzyme(86), oxidizing As(III) to As(V).

Often, ArrAB are encoded in microbial genomes alongside genes of the *ars* operon. However, in

P. mercurii ND132 the putative operons are not co-localized. The ND132 ArrA homolog

includes the conserved molybdenum and Fe-S binding domains found in known functional

respiratory As(V) reductases but is missing some of the key active-site residues(86).

Instead, operon D (Figure 8) may encode for polysulfide or thiosulfate reduction rather

than As(V) respiration. Based on the phylogenetic relationship between the putative ArrA in *P.*

mercurii ND132 with other DMSO-MPT family proteins, the protein encoded by DND132_2597

is more closely related to bacterial polysulfide reductase (PsrA) than ArrA from known As(V)-

respiring bacteria (Figure S17). Microbes in deep-sea vents and hydrothermal springs use the

membrane protein complex PsrABC to couple periplasmic polysulfide or thiosulfate reduction

with quinone oxidation(87). The operon in *P. mercurii* ND132 also encodes homologues of

cytochrome *c* oxidase (DND132_2596) and two rhodanese-domain proteins (DND132_2600,

DND132_2601), one of which (DND132_2600) is significantly upregulated in As(III) cells. A

second cluster of putative *psrAB* genes (DND132_1314, DND132_1313) are present elsewhere

in the genome (Figure S17), and likely encode for cytoplasmic sulfur reduction(88). The

transcription of these *psrAB* genes was not significantly affected by As or SAH. The DMSO-

MPT family of proteins that includes Psr and Arr has been implicated in thioarsenic production.

The geothermal isolate *Pyrobaculum yellowstonensis* strain WP30 produces thioarsenate during

respiratory elemental sulfur and As(V) reduction(89). More work is needed to know if the

protein complex encoded by operon D interacts with As(III), As(V), or with

thiosulfates/polysulfides and enables *P. mercurii* ND132 to respire As. However, it may be

responsible for producing the unknown thioarsenic species observed in *P. mercurii* ND132

culture media.

In As(V) washed cells an operon with homology to the acetoin dehydrogenase complex

was upregulated. In batch cultures, this operon was significantly downregulated in all treatments

except methionine (Figure S15). The operon is predicted to encode an acetoin dehydrogenase

enzyme system (AoDH ES) that catalyzes the conversion of acetoin to acetyl-CoA and

acetaldehyde (Figure 8). The AoDH ES contains multiple copies of three enzymatic components:

acetoin oxidoreductase (E1), dihydrolipoamide acetyltransferase (E2) and dihydrolipoyl

dehydrogenase (E3)(90, 91). The operon (Operon E, Figure 8) encodes AoDH ES homologues

annotated as a dihydrolipoamide acetyltransferase (DND132_3103), (PFam PF00198, PF00364,

PF02817), an acetoin:2,6-dichlorophenolindophenol oxidoreductase subunit alpha

(DND132_3105) (dehydrogenase E1 component, PFam PF00676), acetoin:2,6-

dichlorophenolindophenol oxidoreductase subunit beta (DND132_3106) (transketolase central

region, PFam PF02779, PF02780), and an acetoin catabolism protein with ATP-NAD/AcoX

kinase motifs (PFam PF01513). These proteins also share homology to enzymes in the pyruvate

dehydrogenase complex that convert pyruvate into acetyl-CoA by pyruvate decarboxylation(92).

The operon is possibly regulated by a sigma(54)-dependent HTH transcriptional regulator

(DND132_3102). In closely related *Pseudodesulfovibrio hydrargi* BerOc1 this protein is

annotated as an acetoin dehydrogenase operon transcriptional activator AcoR (Uniprot:

A0A1J5NC98). No obvious link to As metabolism or resistance could be found. It is possible

that this putative AoDH ES operon is upregulated in *P. mercurii* ND132 As(V) washed cells

because of cross-reactivity of As with the HTH regulator (DND132_3102) or because of a

knock-on effect from other differentially regulated mechanisms.

Summary: The broad research aim of these experiments was to examine transcriptional

regulation of *hgcAB* in *Pseudodesulfovibrio mercurii* ND132, with a focus on the potential role

of an upstream ArsR-like regulator found in *hgc*+ Deltaproteobacteria. Additionally, we

examined the relationship between Hg methylation rates in *Pseudodesulfovibrio mercurii* ND132

with metabolic activity and *hgcAB* transcriptional activity. We have shown that *hgcAB*

expression in *Pseudodesulfovibrio mercurii* ND132 is transcriptionally regulated by an ArsR-like

regulator encoded by *arsR2* (DND132_1054) that is responsive to As and SAH. Since *P.*

mercurii ND132 can reduce As(V) to As(III), and we observed increased *hgcA* and *hgcB*

expression when grown with both species, it is likely that As(III) coordinates with ArsR2. The

hgcAB transcriptional regulator shares homology to other members of the ArsR/SmtB family, but

has a distinct four-cysteine motif in its metal(loid) binding site. This homology also distinguishes

it from other ArsR family regulators encoded in the genome of *P. mercurii* ND132. Since As(III)

preferentially binds to three cysteine residues in ArsR repressors of the *ars* operon, we cannot

exclude the possibility that another As species or heavy metal(loid) cation may coordinate with

ArsR2 and also control *hgcAB* expression. Although the transcriptional response to As was more

pronounced, under some culture conditions methionine and adenosine dialdehyde also

significantly increased *arsR2* and *hgcAB* transcription. We attribute this transcriptional response

to a build-up in SAH within the cell.

The RNA-seq analyses show that *hgcAB* transcription is co-regulated as part of the ArsR

regulon with operons that encode As resistance. We also observed significant co-expression of

hgcAB with genes that encode putative As(V) or polysulfide respiration when *P. mercurii*

ND132 is grown with As. These pathways may explain the possible thioarsenic production

observed in *P. mercurii* ND132 cultures. This indicates that As biogeochemistry may influence

hgcAB expression and Hg-methylator activity in some environmental settings. In this study, we

did not observe a significant loss in As resistance with a loss in *hgcAB* function in

Pseudodesulfovibrio mercurii ND132. However, the co-expression and co-localization of *hgcAB*

with arsenic resistance genes in some *hgc+* genomes may hold clues to *hgcAB* evolution and

possible environmental drivers of horizontal transfer of these genes. More work is needed to

tease apart possible overlaps between HgcAB function and arsenic resistance and respiration

mechanisms.

Results from these experiments do not confirm that higher *hgcAB* expression alone

increases MeHg production, but that growth and culture conditions, and the nature of *hgcAB*

transcriptional regulation probably all play roles in methylation control. Both As(III) and As(V)

impacted Hg methylation and transcription of *hgcAB* and *arsR* in *P. mercurii* ND132, although

the impacts were not always coupled. We also observed that metabolic activity and microbial

biomass can significantly influence Hg methylation rates in cultures, and counter possible

impacts of *hgcAB* activity. Hg methylation rates may also be impacted by HgcAB translation,

protein folding, co-factor and membrane insertion, and the methyl donor mechanism. Combining

information of *hgcAB* transcription with microbiome composition, metabolic activity, and Hg

speciation and geochemistry may help resolve environmental Hg methylation rates. It is

therefore important to have accurate and reliable methods for measuring these factors and using

multiple parameters when incorporating microbial dynamics into predictive models for Hg

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DATA AVAILABILITY STATEMENT

The transcriptomic sequencing data is publicly available through KBase narrative

(https://narrative.kbase.us/narrative/124642).

SUPPLEMENTARY MATERIAL

See attached document of supplementary methods, tables, and figures.

REFERENCES

1.   Branfireun BA, Cosio C, Poulain AJ, Riise G, Bravo AG. 2020. Mercury cycling in freshwater systems-An updated conceptual model. Science of the Total Environment 745:140906.

2.   Parks JM, Johs A, Podar M, Bridou R, Hurt RA, Smith SD, Tomanicek SJ, Qian Y, Brown SD, Brandt CC. 2013. The genetic basis for bacterial mercury methylation. Science 339:1332-1335.

3.   Podar M, Gilmour CC, Brandt CC, Soren A, Brown SD, Crable BR, Palumbo AV, Somenahally AC, Elias DA. 2015. Global prevalence and distribution of genes and microorganisms involved in mercury methylation. Sci Adv 1:e1500675.

4.   McDaniel EA, Peterson B, Stevens SL, Tran PQ, Anantharaman K, McMahon K. 2020. Expanded Phylogenetic Diversity and Metabolic Flexibility of Microbial Mercury Methylation. BioRxiv.

5.   Gilmour CC, Podar M, Bullock AL, Graham AM, Brown SD, Somenahally AC, Johs A, Hurt Jr RA, Bailey KL, Elias DA. 2013. Mercury methylation by novel microorganisms from new environments. Environmental science & technology 47:11810-11820.

6.   Gionfriddo CM, Wymore AM, Jones DS, Wilpiszeski RL, Lynes MM, Christensen GA,

Soren A, Gilmour CC, Podar M, Elias DA. 2020. An Improved hgcAB Primer Set and

Direct High-Throughput Sequencing Expand Hg-Methylator Diversity in Nature.

Frontiers in Microbiology 11.

7.   Capo E, Peterson BD, Kim M, Jones DS, Acinas SG, Amyot M, Bertilsson S, Björn E,

Buck M, Cosio C. 2022. A consensus protocol for the recovery of mercury methylation

genes from metagenomes. Molecular Ecology Resources.

8.   Goñi-Urriza M, Klopp C, Ranchou-Peyruse M, Ranchou-Peyruse A, Monperrus M,

Khalfaoui-Hassani B, Guyoneaud R. 2019. Genome insights of mercury methylation

among Desulfovibrio and Pseudodesulfovibrio strains. Research in microbiology.

9.   Gilmour CC, Soren AB, Gionfriddo CM, Podar M, Wall JD, Brown SD, Michener JK,

Urriza MSG, Elias DA. 2021. Pseudodesulfovibrio mercurii sp. nov., a mercury-

methylating bacterium isolated from sediment. International Journal of Systematic and

Evolutionary Microbiology 71:004697.

10.   Ranchou-Peyruse M, Goñi-Urriza M, Guignard M, Goas M, Ranchou-Peyruse A,

Guyoneaud R. 2018. Pseudodesulfovibrio hydrargyri sp. nov., a mercury-methylating

bacterium isolated from a brackish sediment. International Journal of Systematic and

Evolutionary Microbiology 68:1461-1466.

11.   Gionfriddo C, Capo E, Peterson B, Lin H, Jones D, Bravo A. 2021. Hg-cycling

Microorganisms in Aquatic and Terrestrial Ecosystems Database v1. 01142021.

12.   Gilmour CC, Bullock AL, McBurney A, Podar M, Elias DA. 2018. Robust Mercury

Methylation across Diverse Methanogenic Archaea. mBio 9.

13.   King JK, Kostka JE, Frischer ME, Saunders FM. 2000. Sulfate-reducing bacteria

865   methylate mercury at variable rates in pure culture and in marine sediments. Applied and

866   Environmental Microbiology 66:2430-2437.

867   14.   King J, Kostka J, Frischer M, Saunders FM, Jahnke R. 2001. A Quantitative Relationship

868   that Demonstrates Mercury Methylation Rates in Marine Sediments Are. Environmental

869   science & technology 35:2491.

870   15.   Date SS, Parks JM, Rush KW, Wall JD, Ragsdale SW, Johs A. 2019. Kinetics of

871   enzymatic mercury methylation at nanomolar concentrations catalyzed by HgcAB.

872   Applied and environmental microbiology 85:e00438-19.

873   16.   Schaefer JK, Rocks SS, Zheng W, Liang L, Gu B, Morel FM. 2011. Active transport,

874   substrate specificity, and methylation of Hg (II) in anaerobic bacteria. Proceedings of the

875   National Academy of Sciences 108:8714-8719.

876   17.   Graham AM, Bullock AL, Maizel AC, Elias DA, Gilmour CC. 2012. Detailed assessment

877   of the kinetics of Hg-cell association, Hg methylation, and methylmercury degradation in

878   several Desulfovibrio species. Applied and environmental microbiology 78:7337-7346.

879   18.   Cooper CJ, Zheng K, Rush KW, Johs A, Sanders BC, Pavlopoulos GA, Kyrpides NC,

880   Podar M, Ovchinnikov S, Ragsdale SW. 2020. Structure determination of the HgcAB

881   complex using metagenome sequence data: insights into microbial mercury methylation.

882   Communications biology 3:1-9.

883   19.   Smith SD, Bridou R, Johs A, Parks JM, Elias DA, Hurt Jr RA, Brown SD, Podar M, Wall

884   JD. 2015. Site-directed mutagenesis of HgcA and HgcB reveals amino acid residues

885   important for mercury methylation. Applied and environmental microbiology 81:3205-

886   3217.

887   20.   Zhang Y, Rodionov DA, Gelfand MS, Gladyshev VN. 2009. Comparative genomic

888   analyses of nickel, cobalt and vitamin B12 utilization. BMC genomics 10:1-26.

889   21.   Bridwell-Rabb J, Drennan CL. 2017. Vitamin B12 in the spotlight again. Current opinion

890   in chemical biology 37:63-70.

891   22.   Goulding CW, Postigo D, Matthews RG. 1997. Cobalamin-dependent methionine

892   synthase is a modular protein with distinct regions for binding homocysteine,

893   methyltetrahydrofolate, cobalamin, and adenosylmethionine. Biochemistry 36:8082-

894   8091.

895   23.   Ragsdale S, Lindahl P, Münck E. 1987. Mössbauer, EPR, and optical studies of the

896   corrinoid/iron-sulfur protein involved in the synthesis of acetyl coenzyme A by

897   Clostridium thermoaceticum. Journal of Biological Chemistry 262:14289-14297.

898   24.   Choi SC, Chase T, Bartha R. 1994a. Enzymatic Catalysis of Mercury Methylation by

899   Desulfovibrio-Desulfuricans LS. Applied and environmental microbiology 60:1342-

900   1346.

901   25.   Choi SC, Chase T, Bartha R. 1994b. Metabolic Pathways Leading to Mercury

902   Methylation in *Desulfovibrio-desulfuricans* LS. Applied and environmental microbiology

903   60:4072-4077.

904   26.   Siciliano SD, Lean DRS. 2002. Methyltransferase: An enzyme assay for microbial

905   methylmercury formation in acidic soils and sediments. Environmental Toxicology and

906   Chemistry 21:1184-1190.

907   27.   Zhou J, Riccardi D, Beste A, Smith JC, Parks JM. 2014. Mercury methylation by HgcA:

908   theory supports carbanion transfer to Hg (II). Inorganic chemistry 53:772-777.

28.   Qian C, Chen H, Johs A, Lu X, An J, Pierce EM, Parks JM, Elias DA, Hettich RL, Gu B.

2018. Quantitative proteomic analysis of biological processes and responses of the

bacterium Desulfovibrio desulfuricans ND132 upon deletion of its mercury methylation

genes. Proteomics 18:1700479.

29.   Goñi-Urriza M, Corsellis Y, Lanceleur L, Tessier E, Gury J, Monperrus M, Guyoneaud

R. 2015. Relationships between bacterial energetic metabolism, mercury methylation

potential, and hgcA/hgcB gene expression in Desulfovibrio dechloroacetivorans BerOc1.

Environmental Science and Pollution Research 22:13764-13771.

30.   Gilmour CC, Elias DA, Kucken AM, Brown SD, Palumbo AV, Schadt CW, Wall JD.

2011. Sulfate-reducing bacterium Desulfovibrio desulfuricans ND132 as a model for

understanding bacterial mercury methylation. Applied and environmental microbiology

77:3938-3951.

31.   Grégoire DS, Poulain AJ. 2018. Shining light on recent advances in microbial mercury

cycling. Facets 3:858-879.

32.   Rathore SS, Schwartz GE, Brooks SC, Painter SL. 2022. Joint estimation of

biogeochemical model parameters from multiple experiments: A bayesian approach

applied to mercury methylation. Environmental Modelling & Software:105453.

33.   Bravo AG, Cosio C. 2019. Biotic formation of methylmercury: A bio–physico–chemical

conundrum. Limnology and Oceanography.

34.   Hsu-Kim H, Kucharzyk KH, Zhang T, Deshusses MA. 2013. Mechanisms regulating

mercury bioavailability for methylating microorganisms in the aquatic environment: A

critical review. Environmental Science & Technology 47:2441-2456.

35. Christensen GA, Gionfriddo CM, King AJ, Moberly JG, Miller CL, Somenahally AC,

Callister SJ, Brewer H, Podar M, Brown SD. 2019. Determining the Reliability of

Measuring Mercury Cycling Gene Abundance with Correlations with Mercury and

Methylmercury Concentrations. Environmental science & technology 53:8649-8663.

36. Schaefer JK, Kronberg RM, Björn E, Skyllberg U. 2020. Anaerobic guilds responsible

for mercury methylation in boreal wetlands of varied trophic status serving as either a

methylmercury source or sink. Environmental microbiology 22:3685-3699.

37. Rocca JD, Hall EK, Lennon JT, Evans SE, Waldrop MP, Cotner JB, Nemergut DR,

Graham EB, Wallenstein MD. 2015. Relationships between protein-encoding gene

abundance and corresponding process are commonly assumed yet rarely observed. The

ISME journal 9:1693-1699.

38. Liu Y-R, Yang Z, Zhou X, Qu X, Li Z, Zhong H. 2019. Overlooked role of putative non-

Hg methylators in predicting methylmercury production in paddy soils. Environmental

Science & Technology 53:12330-12338.

39. Capo E, Feng C, Bravo AG, Bertilsson S, Soerensen AL, Pinhassi J, Buck M, Karlsson

C, Hawkes J, Björn E. 2022. Abundance and expression of hgcAB genes and mercury

availability jointly explain methylmercury formation in stratified brackish waters.

bioRxiv.

40. Bouchet S, Goñi-Urriza M, Monperrus M, Guyoneaud Rm, Fernandez P, Heredia C,

Tessier E, Gassie C, Point D, Guédron S. 2018. Linking microbial activities and low-

molecular-weight thiols to Hg methylation in biofilms and periphyton from high-altitude

tropical lakes in the Bolivian Altiplano. Environmental science & technology 52:9758-

9767.

41.   Lin H, Ascher DB, Myung Y, Lamborg CH, Hallam SJ, Gionfriddo CM, Holt KE,

Moreau JW. 2020. Mercury methylation by metabolically versatile and cosmopolitan

marine bacteria. bioRxiv.

42.   Peterson BD, McDaniel EA, Schmidt AG, Lepak RF, Janssen SE, Tran PQ, Marick RA,

Ogorek JM, DeWild JF, Krabbenhoft DP. 2020. Mercury methylation genes identified

across diverse anaerobic microbial guilds in a eutrophic sulfate-enriched lake.

Environmental Science & Technology 54:15840-15851.

43.   Gionfriddo CM, Stott MB, Power JF, Ogorek JM, Krabbenhoft DP, Wick R, Holt K,

Chen L-X, Thomas BC, Banfield JF. 2020. Genome-resolved metagenomics and detailed

geochemical speciation analyses yield new insights into microbial mercury cycling in

geothermal springs. Applied and environmental microbiology 86.

44.   Leyn SA, Suvorova IA, Kholina TD, Sherstneva SS, Novichkov PS, Gelfand MS,

Rodionov DA. 2014. Comparative Genomics of Transcriptional Regulation of

Methionine Metabolism in Proteobacteria. PLOS ONE 9:e113714.

45.   Novichkov PS, Li X, Kuehl JV, Deutschbauer AM, Arkin AP, Price MN, Rodionov DA.

2014. Control of methionine metabolism by the SahR transcriptional regulator in

Proteobacteria. Environ Microbiol 16:1-8.

46.   Li X, Krumholz LR. 2007. Regulation of arsenate resistance in Desulfovibrio

desulfuricans G20 by an arsRBCC operon and an arsC gene. J Bacteriol 189:3705-11.

47.   Ekstrom EB, Morel FMM. 2008. Cobalt limitation of growth and mercury methylation in

sulfate-reducing bacteria. Environmental science & technology 42:93-99.

48.   Smith SD. 2015. Protein components of the microbial mercury methylation

pathwayUniversity of Missouri--Columbia.

49.   Brown SD, Gilmour CC, Kucken AM, Wall JD, Elias DA, Brandt CC, Podar M,

978   Chertkov O, Held B, Bruce DC. 2011. Genome sequence of the mercury-methylating

979   strain Desulfovibrio desulfuricans ND132. Am Soc Microbiol.

980   50.   Walker CB, He Z, Yang ZK, Ringbauer Jr JA, He Q, Zhou J, Voordouw G, Wall JD,

981   Arkin AP, Hazen TC. 2009. The electron transfer system of syntrophically grown

982   Desulfovibrio vulgaris. Journal of bacteriology 191:5793-5801.

983   51.   Graham AM, Aiken GR, Gilmour CC. 2012. Dissolved Organic Matter Enhances

984   Microbial Mercury Methylation Under Sulfidic Conditions. Environmental Science &

985   Technology 46:2715-2723.

986   52.   Arkin AP, Cottingham RW, Henry CS, Harris NL, Stevens RL, Maslov S, Dehal P, Ware

987   D, Perez F, Canon S, Sneddon MW, Henderson ML, Riehl WJ, Murphy-Olson D, Chan

988   SY, Kamimura RT, Kumari S, Drake MM, Brettin TS, Glass EM, Chivian D, Gunter D,

989   Weston DJ, Allen BH, Baumohl J, Best AA, Bowen B, Brenner SE, Bun CC, Chandonia

990   J-M, Chia J-M, Colasanti R, Conrad N, Davis JJ, Davison BH, DeJongh M, Devoid S,

991   Dietrich E, Dubchak I, Edirisinghe JN, Fang G, Faria JP, Frybarger PM, Gerlach W,

992   Gerstein M, Greiner A, Gurtowski J, Haun HL, He F, Jain R, et al. 2018. KBase: The

993   United States Department of Energy Systems Biology Knowledgebase. Nature

994   Biotechnology 36:566-569.

995   53.   Bolger AM, Lohse M, Usadel B. 2014. Trimmomatic: a flexible trimmer for Illumina

996   sequence data. Bioinformatics 30:2114-2120.

997   54.   Kim D, Paggi JM, Park C, Bennett C, Salzberg SL. 2019. Graph-based genome

998   alignment and genotyping with HISAT2 and HISAT-genotype. Nature Biotechnology

999   37:907-915.

55.   Pertea M, Pertea GM, Antonescu CM, Chang T-C, Mendell JT, Salzberg SL. 2015.

StringTie enables improved reconstruction of a transcriptome from RNA-seq reads.

Nature Biotechnology 33:290-295.

56.   Love MI, Huber W, Anders S. 2014. Moderated estimation of fold change and dispersion

for RNA-seq data with DESeq2. Genome Biology 15:550.

57.   Afgan E, Baker D, Batut B, van den Beek M, Bouvier D, Čech M, Chilton J, Clements D,

Coraor N, Grüning BA, Guerler A, Hillman-Jackson J, Hiltemann S, Jalili V, Rasche H,

Soranzo N, Goecks J, Taylor J, Nekrutenko A, Blankenberg D. 2018. The Galaxy

platform for accessible, reproducible and collaborative biomedical analyses: 2018 update.

Nucleic Acids Research 46:W537-W544.

58.   Mirdita M, Schütze K, Moriwaki Y, Heo L, Ovchinnikov S, Steinegger M. 2022.

ColabFold: making protein folding accessible to all. Nature Methods 19:679-682.

59.   Jumper J, Evans R, Pritzel A, Green T, Figurnov M, Ronneberger O, Tunyasuvunakool

K, Bates R, Žídek A, Potapenko A, Bridgland A, Meyer C, Kohl SAA, Ballard AJ,

Cowie A, Romera-Paredes B, Nikolov S, Jain R, Adler J, Back T, Petersen S, Reiman D,

Clancy E, Zielinski M, Steinegger M, Pacholska M, Berghammer T, Bodenstein S, Silver

D, Vinyals O, Senior AW, Kavukcuoglu K, Kohli P, Hassabis D. 2021. Highly accurate

protein structure prediction with AlphaFold. Nature 596:583-589.

60.   Mirdita M, Steinegger M, Söding J. 2019. MMseqs2 desktop and local web server app for

fast, interactive sequence searches. Bioinformatics 35:2856-2858.

61.   Mitchell AL, Almeida A, Beracochea M, Boland M, Burgin J, Cochrane G, Crusoe MR,

Kale V, Potter SC, Richardson LJ, Sakharova E, Scheremetjew M, Korobeynikov A,

1022   Shlemov A, Kunyavskaya O, Lapidus A, Finn RD. 2020. MGnify: the microbiome

1023   analysis resource in 2020. Nucleic Acids Research 48:D570-D578.

1024 62. Evans R, O'Neill M, Pritzel A, Antropova N, Senior A, Green T, Žídek A, Bates R,

1025   Blackwell S, Yim J, Ronneberger O, Bodenstein S, Zielinski M, Bridgland A, Potapenko

1026   A, Cowie A, Tunyasuvunakool K, Jain R, Clancy E, Kohli P, Jumper J, Hassabis D.

1027   2022. Protein complex prediction with AlphaFold-Multimer. bioRxiv

1028   doi:10.1101/2021.10.04.463034:2021.10.04.463034.

1029 63. Team RC. 2013. R: A language and environment for statistical computing.

1030 64. Busenlehner LS, Pennella MA, Giedroc DP. 2003. The SmtB/ArsR family of

1031   metalloregulatory transcriptional repressors: Structural insights into prokaryotic metal

1032   resistance. FEMS Microbiol Rev 27:131-43.

1033 65. Andres J, Bertin PN. 2016. The microbial genomics of arsenic. FEMS Microbiology

1034   Reviews 40:299-322.

1035 66. Chen J, Madegowda M, Bhattacharjee H, Rosen BP. 2015. ArsP: a methylarsenite efflux

1036   permease. Molecular Microbiology 98:625-635.

1037 67. Chen J, Zhang J, Rosen BP. 2019. Role of ArsEFG in Roxarsone and Nitarsone

1038   Detoxification and Resistance. Environmental Science & Technology 53:6182-6191.

1039 68. Rawle R, Saley TC, Kang Y-S, Wang Q, Walk S, Bothner B, McDermott TR. 2021.

1040   Introducing the ArsR-Regulated Arsenic Stimulon. Frontiers in Microbiology 12.

1041 69. Qin J, Rosen BP, Zhang Y, Wang G, Franke S, Rensing C. 2006. Arsenic detoxification

1042   and evolution of trimethylarsine gas by a microbial arsenite *S*-

1043   adenosylmethionine methyltransferase. Proceedings of the National Academy of Sciences

1044   103:2075-2080.

1045   70.   Woodcroft BJ, Singleton CM, Boyd JA, Evans PN, Emerson JB, Zayed AAF, Hoelzle

1046   RD, Lamberton TO, McCalley CK, Hodgkins SB, Wilson RM, Purvine SO, Nicora CD,

1047   Li C, Frolking S, Chanton JP, Crill PM, Saleska SR, Rich VI, Tyson GW. 2018.

1048   Genome-centric view of carbon processing in thawing permafrost. Nature 560:49-54.

1049   71.   Haas S, Desai DK, LaRoche J, Pawlowicz R, Wallace DWR. 2019. Geomicrobiology of

1050   the carbon, nitrogen and sulphur cycles in Powell Lake: a permanently stratified water

1051   column containing ancient seawater. Environmental Microbiology 21:3927-3952.

1052   72.   Qiu YL, Kuang XZ, Shi XS, Yuan XZ, Guo RB. 2014. Paludibacter jiangxiensis sp. nov.,

1053   a strictly anaerobic, propionate-producing bacterium isolated from rice paddy field. Arch

1054   Microbiol 196:149-55.

1055   73.   Anantharaman K, Brown CT, Hug LA, Sharon I, Castelle CJ, Probst AJ, Thomas BC,

1056   Singh A, Wilkins MJ, Karaoz U. 2016. Thousands of microbial genomes shed light on

1057   interconnected biogeochemical processes in an aquifer system. Nature communications

1058   7:13219.

1059   74.   Giloteaux L, Holmes DE, Williams KH, Wrighton KC, Wilkins MJ, Montgomery AP,

1060   Smith JA, Orellana R, Thompson CA, Roper TJ, Long PE, Lovley DR. 2013.

1061   Characterization and transcription of arsenic respiration and resistance genes during in

1062   situ uranium bioremediation. The ISME Journal 7:370-383.

1063   75.   Prabaharan C, Kandavelu P, Packianathan C, Rosen BP, Thiyagarajan S. 2019. Structures

1064   of two ArsR As(III)-responsive transcriptional repressors: Implications for the

1065   mechanism of derepression. Journal of Structural Biology 207:209-217.

1066   76.   Shi W, Dong J, Scott RA, Ksenzenko MY, Rosen BP. 1996. The Role of Arsenic-Thiol

1067   Interactions in Metalloregulation of the ars Operon (∗). Journal of Biological Chemistry

1068   271:9291-9297.

1069   77.   Packianathan C, Kandavelu P, Rosen BP. 2018. The Structure of an As (III) S-

1070   adenosylmethionine methyltransferase with 3-coordinately bound As (III) depicts the first

1071   step in catalysis. Biochemistry 57:4083-4092.

1072   78.   Rush K. 2018. Expression and characterization of HgcA and HgcB, two proteins

1073   involved in methylmercury biosynthesis.

1074   79.   Barrouilhet S, Monperrus M, Tessier E, Khalfaoui-Hassani B, Guyoneaud R, Isaure M-P,

1075   Goñi-Urriza M. 2021. Effect of Exogenous and Endogenous Sulfide on The Production

1076   and The Export of Methylmercury by Sulfate Reducing Bacteria.

1077   80.   Benoit JM, Gilmour CC, Mason RP. 2001. The influence of sulfide on solid phase

1078   mercury bioavailability for methylation by pure cultures of *Desulfobulbus propionicus*

1079   (1pr3). Environmental Science & Technology 35:127-132.

1080   81.   Mija A, Jules M B, Alexandre J P. 2022. Arsenate decreases production of

1081   methylmercury across increasing sulfate concentration amendments in freshwater lake

1082   sediments. Environmental Science: Processes & Impacts doi:10.1039/D1EM00543J.

1083   82.   Glasser NR, Oyala PH, Osborne TH, Santini JM, Newman DK. 2018. Structural and

1084   mechanistic analysis of the arsenate respiratory reductase provides insight into

1085   environmental arsenic transformations. Proceedings of the National Academy of Sciences

1086   115:E8614-E8623.

1087   83.   Stolz JF, Basu P. 2018. Unraveling the inner workings of respiratory arsenate reductase.

1088   Proceedings of the National Academy of Sciences 115:9051-9053.

84.   Malasarn D, Keeffe JR, Newman DK. 2008. Characterization of the Arsenate Respiratory

Reductase from *Shewanella* sp. Strain ANA-3. Journal of Bacteriology 190:135-

142.

85.   Tsuchiya T, Ehara A, Kasahara Y, Hamamura N, Amachi S, Zhou N-Y. 2019. Expression

of Genes and Proteins Involved in Arsenic Respiration and Resistance in Dissimilatory

Arsenate-Reducing *Geobacter* sp. Strain OR-1. Applied and Environmental

Microbiology 85:e00763-19.

86.   Richey C, Chovanec P, Hoeft SE, Oremland RS, Basu P, Stolz JF. 2009. Respiratory

arsenate reductase as a bidirectional enzyme. Biochemical and Biophysical Research

Communications 382:298-302.

87.   Jormakka M, Yokoyama K, Yano T, Tamakoshi M, Akimoto S, Shimamura T, Curmi P,

Iwata S. 2008. Molecular mechanism of energy conservation in polysulfide respiration.

Nat Struct Mol Biol 15:730-7.

88.   Wang S, Jiang L, Hu Q, Liu X, Yang S, Shao Z. 2021. Elemental sulfur reduction by a

deep-sea hydrothermal vent Campylobacterium Sulfurimonas sp. NW10. Environ

Microbiol 23:965-979.

89.   Jay ZJ, Beam JP, Dohnalkova A, Lohmayer R, Bodle B, Planer-Friedrich B, Romine M,

Inskeep WP. 2015. Pyrobaculum yellowstonensis Strain WP30 Respires on Elemental

Sulfur and/or Arsenate in Circumneutral Sulfidic Geothermal Sediments of Yellowstone

National Park. Applied and environmental microbiology 81:5907-5916.

90.   Oppermann FB, Schmidt B, Steinbüchel A. 1991. Purification and characterization of

acetoin:2,6-dichlorophenolindophenol oxidoreductase, dihydrolipoamide dehydrogenase,

and dihydrolipoamide acetyltransferase of the Pelobacter carbinolicus acetoin

dehydrogenase enzyme system. Journal of Bacteriology 173:757-767.

91.   Oppermann FB, Steinbüchel A. 1994. Identification and molecular characterization of the

aco genes encoding the Pelobacter carbinolicus acetoin dehydrogenase enzyme system.

Journal of Bacteriology 176:469-485.

92.   de Kok A, Hengeveld AF, Martin A, Westphal AH. 1998. The pyruvate dehydrogenase

multi-enzyme complex from Gram-negative bacteria. Biochimica et Biophysica Acta

(BBA) - Protein Structure and Molecular Enzymology 1385:353-366.

93.   Harrison KJ, Crécy-Lagard Vd, Zallot R. 2017. Gene Graphics: a genomic neighborhood

data visualization web application. Bioinformatics 34:1406-1408.

94.   Sprouffske K, Wagner A. 2016. Growthcurver: an R package for obtaining interpretable

metrics from microbial growth curves. BMC Bioinformatics 17:172.

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1125   **MAIN FIGURES**

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1128   Figure 1. ArsR/SmtB family transcriptional regulators encoded in the genome of

1129   *Pseudodesulfovibrio mercurii* ND132. The five *arsR* genes are numbered by order of appearance

1130   in genome, with gene ID in arrow (e.g. HgcA enoded by DND132_1056), gene orientation

1131   shown by arrows, and color coded by protein family. Figure made using Gene Graphics(93).

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Figure 2. (A) AlphaFold2 model of the putative ArsR-like transcriptional regulator encoded by

arsR2 (Uniprot ID: F0JBE8) and co-localized with *hgcAB* in the genome of *Pseudodesulfovibrio*

mercurii ND132. The model is colored by model confidence (i.e., predicted LDDT score). Dark

blue = very high confidence (pLDDT > 90), cyan = confident (90 > pLDDT > 70), yellow = low

confidence (70 > pLDDT > 50), and orange = very low confidence (pLDDT < 50). (B) Close-up

view colored by chain with cysteine residues shown in the ball-and-stick representation.

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1142 Figure 3. Growth assay tests of arsenate toxicity to *Pseudodesulfovibrio mercurii* ND132 wild-

1143 type (WT) and *hgcAB* deletion mutant (Δ*hgcAB*) in defined CCM media with pyruvate-fumarate

1144 (PF) or lactate-sulfate (LS) and sodium arsenate, As(V). Growth curves fitted to optical density

1145 measured at 600 nm (OD$_{600}$) absorbance of biological triplicates using 'ggplot' and 'dplyr' in R.

1146 Shaded area shows the standard error of average OD$_{600}$ measurements for control replicates.

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Figure 4. Regulation of *arsR* and *hgcA* expression in wild type *Pseudodesulfovibrio mercurii*

ND132. Top: Growth curves of for cell grown in defined CCM-PF (A) and enriched EPF media

(B) with various potential transcriptional regulators. Growth plotted as change in optical density

at 600 nm absorbance (OD_{600}) from inoculation at T_0, using R program, Growthcurver(94). Gray

dashed lines indicate when cells were harvested for RNA extraction for qPCR at early, mid, and

late exponential growth. Bottom: Relative expression ($\Delta\Delta$Cq) of *arsR2* (DND132_1054) and

hgcA (DND132_1056) from qPCR of cDNA (from RNA). Expression of the target genes was

normalized to three housekeeping genes (*gyrA*, *gyrB*, *recA*) and to control (i.e., culture with no

treatment). Treatments: 0.1 mM (CCM and EPF) adenosine-2',3'-dialdehyde, 0.2 mM (CCM)

and 0.02 mM sodium arsenite, 2.0 mM (CCM) and 0.2 mM (EPF) sodium arsenate, 1.0 mM

HgCl (CCM only), 1.0 mM (CCM and EPF) methionine. Stars indicate significant difference

between $\Delta\Delta$Cq of treatment compared to control (z-Test, $p < 0.05$). qPCR was performed on

biological triplicate cultures (CCM) and single culture (EPF) and included three analytical

replicates per cDNA sample.

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Figure 5. Impact of potential *hgcA* transcription regulators on MeHg production by

Pseudodesulfovibrio mercurii ND132. Top: MeHg production in washed-cell (A) and batch

cultures (B) with treatments: 0.1 mM adenosine-2',3'-dialdehyde, 0.02 mM sodium arsenite, 0.2

mM sodium arsenate, 1.0 mM methionine, control (cells in unamended washed buffer or EPF

media), and abiotic (no cells, unamended wash buffer or EPF media). Linear regression fit to

data using R. Hg methylation assays were performed in triplicates, with sampling at 0, 10, 20,

40, 100 minutes after being spiked with 1 nM of [201]HgCl. Error bars represent +/-1 standard error

of the mean. Cells were harvested from assays at t = 100 minutes for qPCR of cDNA (from

1174 RNA) to measure normalized expression ($\Delta\Delta$Cq) of *arsR* (DND132_1054) and *hgcA*

1175 (DND132_1056) in washed cells (C) and batch culture (D). Expression normalized to three

1176 housekeeping genes (*gyrA, gyrB, recA*) and control (Supp. Methods 1.1). Stars indicate

1177 significant difference between $\Delta\Delta$Cq of treatment compared to control (z-Test, $p < 0.05$).

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Figure 6. Normalized mercury methylation rates from culture and washed-cell assays. Mercury

methylation rates in *Pseudodesulfovibrio mercurii* ND132 washed-cell (top panel) and batch

culture (bottom panel) assays with test treatments compared to control when normalized to

dissolved inorganic carbon (DIC) production (i.e., CO_2 respiration), optical density (OD_{600}) at

T_0, cell density at T_0, filter-passable total Hg at T_0, and total unfiltered Hg at T_0.

Figure 7. Differential expression of significant features (|log FC| > 1.5, p < 0.05) from RNA-seq analyses of As(V) (0.2 mM) and As(III) (0.02 mM) treated washed cells of wild-type *Pseudodesulfovibrio mercurii* ND132. Several putative operons (A-E) had significant differential expression compared to control washed-cells (i.e., no arsenic treatment).

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Figure 8. Proposed biochemical reactions for operons that were significantly differentially

expressed in *Pseudodesulfovibrio mercurii* ND132 transcriptomes grown with arsenic. This

includes genes that are part of the ArsR regulon in ND132, including (A) Hg-methylation genes

hgcA and *hgcB,* (B) operon that encodes for arsenate reductase (ArsC) and As(III) efflux pump

Acr3, and (C) an operon that includes homologues of organoarsenic resistance proteins ArsE and

ArsP. Also significantly differentially expressed was an operon (D) that encodes homologues of

the polysulfide reductase (Psr) or arsenic respiration complex (Arr), which couples menaquinone

(MK) oxidation to periplasmic AsV or sulfur reduction, and (E) a set of genes that encode

homologues of the acetoin dehydrogenase enzyme system (AoDH ES). *P. mercurii* ND132 is a

confirmed Hg-methylator and can reduce As(V) to As(III). The biochemical mechanism that

produces the unknown thioarsenic compound is unknown. To date, *P. mercurii* ND132 has not

been tested for organoarsenic resistance or for acetoin degradation. The mechanism for uptake of

As(V) and As(III) by ND132 is unknown. Operon graphic made using Gene Graphics(93).