Title: Spatial interactions between parrotfishes and implications for species coexistence

Keywords: coexistence, coral reefs, home range, movement, parrotfishes, spatial interaction

Authors: Manning, J.C.[1*], McCoy, S.J.[2], and Benhamou, S.[3]

[1] Florida State University, Department of Biological Sciences

319 Stadium Drive

Tallahassee, FL 32306-4295, USA

ORCID iD: 0000-0001-8130-5460

[2] University of North Carolina, Department of Biology

120 South Road, CB3280

Chapel Hill, NC 27599-3280, USA

ORCID iD: 0000-0003-1324-1578

[3] MAD Team, Centre d'Ecologie Fonctionnelle et Evolutive, CNRS

1919 Route de Mende, 34293 Montpellier Cedex 5, France

Associated to Cogitamus Lab

ORCID iD: 0000-0003-0803-8559

Corresponding Author and Email: Joshua C. Manning, jmanning@bio.fsu.edu

Abstract

Home range behavior mediates species interactions and distributions, and spatiotemporal segregation may facilitate coexistence of competing species. We investigated home range behavior and spatial interactions in four common parrotfishes on coral reefs in Bonaire, Caribbean Netherlands, to determine how spatial interactions mediate species interactions and contribute to their coexistence. We first computed home ranges for males and females of each species. We then quantified spatial overlap (i.e., static interaction) between the home ranges of neighboring male parrotfishes and their activity in shared areas to estimate interaction potential for pairs of individuals. Finally, we analyzed dynamic interactions in simultaneously tracked, spatially co-occurring interspecific pairs of parrotfishes to investigate how they interact in shared space. Generally, spatial overlap of home ranges was much lower for intraspecific pairs than for interspecific pairs, but the probability of finding males in areas shared with males of other species was species-dependent. Males in interspecific pairs moved mostly independently of each other in shared areas, but we did find some evidence of avoidance in interspecific pairs sharing the most space. We discuss our findings within the context of parrotfish social and foraging ecology to further elucidate the spatial ecology of these functionally important reef fishes.

Introduction

In many species, individuals establish home ranges, broadly defined as self-restricted areas in which animals conduct their normal daily activities. These home ranges arise from individual movement decisions made by animals in response to their environment [1–3]. Home range characteristics (e.g., size) are determined by resource abundances [4], the space use of competitors [5,6], and intrinsic characteristics of the animal (e.g., locomotion and trophic status; Harestad & Bunnell 1979; Nash *et al.* 2015; Tamburello *et al.* 2015). Additionally, some animals defend all or a portion of their home ranges as territories, when the benefit of doing so outweighs the cost (Brown 1964; Kaufmann 1983).

Home range and territorial behaviors can affect population dynamics and regulation, including by setting upper limits on population size [12–15]. Furthermore, animal movement and space use can facilitate species coexistence (reviewed in [16]). Coexistence is maintained through mechanisms that minimize average fitness differences between organisms (equalizing) or increase the strength of intraspecific competition relative to interspecific competition (stabilizing; [17]). Spatial and temporal segregation, mediated by animal movement and activity patterns, may act as a stabilizing mechanism that facilitates coexistence of competing species [16,18–21]. Yet, these mechanisms remain largely unexplored in reef fish communities.

Parrotfishes act as important agents of bioerosion and sediment production, reworking, and transport on coral reefs [22–25]; and their grazing helps to maintain reef substrates in cropped, early successional states dominated by filamentous cyanobacteria and other microscopic photoautotrophs [26,27]. Epilithic and endolithic microautotrophs, including filamentous and mat-forming benthic cyanobacteria, are the primary nutritional targets of parrotfish foraging [26,28–30]. There is strong evidence that parrotfishes partition these

resources at fine spatial scales by targeting substrata at different successional stages [28,31], and

at broader scales by foraging in different habitat types (e.g., rubble v. high-relief habitat; [32]). In

contrast, we know little about how movement and space use contribute to niche partitioning and

coexistence in parrotfishes.

Adult male (terminal phase; TP) parrotfishes of many species defend stable, exclusive

territories containing harems of intraspecific females (initial phase; IP) against intraspecific TPs,

increasing their spawning success and access to high quality foods [33–36]. Territoriality may

contribute to population regulation [12], and could impose spatial constraints on parrotfish

foraging (i.e., focused grazing within core areas; [37]) that mediate spatial patterns of benthic

community assembly [38]. Movement and space use may also influence competition and

coexistence among parrotfishes. However, studies of parrotfish movement and space-use are

often limited to correlating home range size to body size or resource abundance (e.g., [39,40]).

In this study, we investigated spatial interactions among four common Caribbean

parrotfishes. Specifically, we tested the hypothesis that parrotfishes would avoid one another

spatially and/or temporally to reduce competition for shared resources. To test this hypothesis,

we estimated home ranges for several TP and IP fishes of each species to quantify differences in

space use and the spatial overlap (i.e., static interaction) of co-occurring TP pairs (inter- and

intraspecific). Furthermore, we analyzed the movements of simultaneously tracked interspecific

pairs of TP parrotfishes to investigate the potential for dynamic interactions between these fishes

in shared space. Finally, we conducted behavioral observations of agonistic interactions to

provide context for our findings.

Material and methods

Study sites and data collection

We conducted our study in June-July 2021 at two fringing coral reef sites on the leeward coast of Bonaire, Caribbean Netherlands: Aquarius and Invisibles. These sites are characterized by relatively high coral cover and low macroalgal cover [30]. The abundance and biomass of different fish groups, including parrotfishes, are higher on Bonaire's coral reefs relative to more heavily fished reefs in the Eastern Caribbean [41]. This is likely due to fisheries management efforts, including spear gun bans in 1971, the establishment of no fishing areas in 2008, bans on parrotfishes catches in 2010, and fish trap phase-outs in 2010 [42].

At each site, we collected two separate datasets: (1) GPS tracking and behavioral observations of several individuals, one by one, to estimate home ranges, quantify static interactions and activity in shared space, and explore the role of agonistic behavior in driving these patterns; and (2) simultaneous GPS tracking of pairs of neighboring parrotfishes to investigate dynamic interactions between spatially co-occurring individuals. GPS tracks were recorded using a Garmin GPSMAP 78sc by a snorkeler remaining at the surface, above the tracked individual, while a SCUBA diver monitored its behavior below the surface (except during dynamic interaction tracks; see below). Divers acclimated fish to their presence for ~ 2 mins and subsequently observed fish behavior from ~ 2 m to avoid influencing their behavior. Marine fishes in protected areas and areas with high diver visitation have reduced flight initiation distances [43]. Parrotfishes in Bonaire are, thus, likely to be habituated to divers because of long-standing spearfishing bans [42] and a thriving dive tourism industry.

Space use, static interaction, and activity in shared space

At each study site, we tracked several TPs and IPs of four common parrotfishes, *Scarus*

taeniopterus, *Scarus vetula*, *Sparisoma aurforenatum*, and *Sparisoma viride,* for ~20 mins

(Supplemental Information: Table S1). We attempted to track all territorial TPs within

predetermined ~1,000 m^2 plots at each study site. We also tracked several haremic IP fishes from

different territories for all species. We verified that our tracking duration was sufficient to

capture diurnal (i.e., daytime) home range behavior by conducting visual assessments of

stationarity (Figure S1; [44]). Individual parrotfishes are known to use these diurnal home ranges

for extended periods of time [33,35,36].

For each fish, we computed home range and core areas, defined as the areas within the

95% and 50% cumulative isopleths of the utilization distribution, respectively, using movement-

based kernel density estimation [45]. We then quantified the spatial overlap of home ranges (i.e.,

static interactions) for intra- and interspecific pairs of neighboring TPs using Bhattacharyya's

Affinity (BA; [46]). We also quantified the probability of finding each TP within the home range

and core areas of neighboring fishes (i.e., the volume of the fish's utilization distribution that lies

within the home ranges and core areas of neighbors; [47]). Agonistic interactions between intra-

and interspecific TPs occurred 2.66 ± 0.09 m (mean $\pm$ SE, n = 168) from the boundaries of TP

home ranges (see below). We, therefore, defined neighboring TPs as individuals whose home

range boundaries were within 2.66 m of one another. All utilization distributions, spatial

overlaps, and probabilities of activity within home range and core areas were computed in the R

package adehabitatHR [48].

All statistical models were fit in the R package glmmTMB [49]. We fit Gamma

distributed generalized linear models with log-link functions to investigate differences in home

range and core area size as a function of site, species, phase (TP or IP), $\log_{10}$ body mass, and the

interaction of species and phase (Table S2). To assess differences in (1) spatial overlap of home

ranges for pairs of neighboring TPs and (2) the probability of finding each TP in the home range

and core areas of neighboring TPs, we fit zero-inflated beta regression models (Tables S3-S5).

For each beta regression, we fit a full model including site, the number of days between the GPS

tracks for each pair (fish were not all tracked the same day), and species pairing (e.g., *Sp. viride*

– *Sc. vetula*) as fixed factors and a reduced model including only species pairing as a fixed

factor. For the two models assessing the probability of finding TPs in the home ranges and core

areas of neighboring TPs, we also included focal fish identification as a random effect. We

conducted model selection (AICc) to determine whether to use the full or reduced model for each

response variable. After determining what fixed effects to include in our models, we fit variable

dispersion models (i.e., precision parameter allowed to vary as a function of explanatory

variables, respectively) and used model selection (AICc) to determine whether a fixed or variable

dispersion model fit best (i.e., a two-step selection process similar to [50]). The significance of

the fixed effects in the conditional portion of the three final models was tested using Type III

Wald's χ^2. We also computed marginal means for the conditional, zero-inflated, and dispersion

components of each model using the R package emmeans [51].

Spatial patterns of agonistic behavior

We concurrently video-recorded the behavior of each TP with a GoPro Hero 4 Silver

(GoPro, Inc) and analyzed these videos in the behavioral analysis software BORIS [52]. We

recorded the duration (to the closest second; those lasting less than 1 s were scored as 1 s) of all

agonistic interactions (agonisms) and species and ontogenetic phase of the interactor. Apparent

agonisms for which we were unable to identify the interactor were excluded from analyses.

Video times were synchronized with GPS time to estimate where each agonism occurred. We

assessed the predictors of agonism frequency using a generalized linear model fit to a negative

binomial distribution (Table S6). We included site, focal species, and interactor identity (e.g., TP

– intraspecific, etc.) as fixed effects and the log of observation time as an offset. We assessed the

predictors of agonism duration (log transformed to meet the model's assumptions) and the

minimum distance between the agonism and home range boundary of the focal fish with a linear

mixed model including site, focal species, and interactor identity as fixed effects and focal fish

identification as a random effect (Tables S7 and S8). We graphically assessed model residuals to

confirm that they met assumptions.

Simultaneous tracking and dynamic interactions

We investigated dynamic interactions (i.e., the tendency of two animals to move together,

to avoid each other or to move independently) between simultaneously tracked pairs of spatially

co-occurring TP *Sp. viride* and *Sc. vetula* (Video 1) at both study sites (Table S9). We chose

these species as they were the largest and easiest to track simultaneously with only snorkelers,

and because they tend to forage on similar substrates [53]. As such, we hypothesized that

individuals of these species would be most likely to avoid one another to prevent competition

within shared areas. GPS tracks for these fishes lasted 31.2 ± 2.0 min (mean $\pm$ SD, $n = 18$) and

were resampled to achieve synchronized relocations every 5s.

For each pair (n = 9), we identified the area of overlap between the two home ranges

using the 'st_intersection' function in the R package sf [54]. Call $N_{s,s}$ the number of relocation

pairs for which both individuals were in the shared area, and $N_{s,e}$, $N_{e,s}$, and $N_{e,e}$ the number of

relocation pairs for which at least one of two individuals was in its respective exclusive area. We

then contrasted the actual frequency $N_{s,s}/T$ with which both individuals of a given pair were in

the shared area, where $T = N_{s,s}+N_{s,e}+N_{e,s}+N_{e,e}$ is the total number of simultaneous relocation

pairs, with the theoretical probability $(N_{s,s}+N_{s,e})(N_{s,s}+N_{e,s})/T^2$ that both individuals would

simultaneously be in the shared area if they moved independently of each other. We then tested

for avoidance or attractance within shared areas by comparing the difference in the observed and

theoretical values with a Wilcoxon signed-ranks test. For each pair, we also quantified the spatial

overlap of their home ranges and the probability of finding each fish in the home range and core

area of the other (as described above). Finally, we investigated the tendency for each pair to

move jointly, independently, or avoid one another when simultaneously present within the shared

area by computing the dynamic interaction index [55] using the 'IAB' function in the R package

wildlifeDI [56].

Results

Space use and static interactions

We found a significant interactive effect of species and phase on home range size

(Wald's $\chi^2 = 14.01$, df = 3, $p = 0.003$). There was only a marginal interactive effect of species

and phase on the sizes of core areas (Wald's $\chi^2 = 6.86$, df = 3, $p = 0.076$), but significant

differences among species (Wald's $\chi^2 = 38.78$, df = 3, $p < 0.001$). *Scarus taeniopterus* had the

smallest home ranges and core areas of all species, while *Sc. vetula* and *Sp. viride* had the largest

(Figure 1). Additionally, TP home ranges and core areas were larger than IP home ranges and

core areas for all species except *Sc. taeniopterus* (Figure 1).

The best fit model for spatial overlap of neighboring TP parrotfishes was a zero-inflated

beta regression with variable dispersion that included species pairings as an effect in each

component of the model (Table S3). We found no spatial overlap between the home ranges of

16.7 – 36.7% pairs of neighboring TPs (Figure 2a). The likelihood of neighboring TP

parrotfishes sharing space did not significantly differ among species pairs, but the home ranges

of neighboring intraspecific pairs of TPs tended to be non-overlapping more often than the home

ranges of neighboring interspecific pairs (Figure S2). When neighboring TP home ranges did

overlap, the degree of overlap differed among species pairings (Wald's χ^2 = 112.01, df = 9, p <

0.001). For *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride*, the home ranges of intraspecific TP

pairs overlapped significantly less than the home ranges of interspecific TP pairs (Figure 2a).

Overlap of intraspecific pairs of TP home ranges in *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride*

was extremely low (0.03 ± 0.01, 0.07 ± 0.02, and 0.04 ± 0.01; mean ± SE; n = 14, 25, and 11

pairs, respectively). For *Sc. taeniopterus*, there was no significant difference between the amount

of space shared by intraspecific TP pairs and the amount of space shared with TP *Sp. viride* and

TP *Sc. vetula* (Figure 2a and Figure S2).

The best fit model for the probability of finding focal TPs within the home ranges of

neighboring TPs was a zero-inflated beta regression with variable dispersion that included

species pairings as an effect in each component of the model and focal identity as a random

effect (Table S4). The proportion of neighboring TP parrotfishes that shared no space (Figure 2b)

did not differ among species pairings (Figure S3). When the home ranges of neighbors

overlapped, the probability of finding focal TPs in the home ranges of their neighbors

significantly differed among species pairings (Wald's χ^2 = 296.44, df = 15, p < 0.001). TP *Sc.*

vetula, *Sp. aurofrenatum*, and *Sp. viride* were significantly less likely to be found in intraspecific

home ranges than in interspecific home ranges (Figure 2b). The same was not true for TP *Sc.*

taeniopterus. TP *Sc. taeniopterus* were more likely to be found in the home ranges of TP *Sc.*

vetula, *Sp. aurofrenatum*, and *Sp. viride* than vice versa (Figure 2b and Figure S3).

224   The best fit model for the probability of finding focal TPs within the core areas of

neighboring TPs was a zero-inflated beta regression with variable dispersion that included

species pairings as an effect in each component of the model and focal identity as a random

effect (Table S5). The proportion of focal TPs with zero probability of being found in the core

areas of neighboring TPs was higher for intraspecific than interspecific pairs (Figure S4). In fact,

the probability of finding neighboring TP *Sp. viride* in one another's core areas was always zero

(Figure 2c). When the probability of finding neighboring TPs in one another's core areas was

greater than zero, it was still low and differed significantly among species pairings (Wald's χ^2 =

231.68, df = 14, $p < 0.001$). TP *Sc. taeniopterus* had a higher probability of being found in the

core areas of TP *Sc. vetula* and *Sp. aurofrenatum*, than the reverse (Figure 2c and Figure S4).

Additionally, though not significantly, TP *Sc. taeniopterus* were also more likely to be found in

the core areas of TP *Sp. viride* than the reverse (Figure 2c and Figure S4).

Spatial patterns of agonisms

238   Focal TPs were most frequently agonistic toward IPs and TPs belonging to the same

species (Wald's χ^2 = 160.76, df = 3, $p < 0.001$; Figure 3a). Intraspecific agonisms between TPs

lasted longer than all other agonisms (Wald's χ^2 = 260.37, df = 3, $p < 0.001$; Figure 3b) and more

frequently involved aggressive chases, rather than brief charges or displays (Fisher's Exact Test

$p < 0.001$). Intraspecific agonisms between focal TPs and IPs occurred 3.03 ± 0.07 m (mean ±

SE, n = 292) from home range boundaries, further from the home range boundary than agonisms

with other fishes (Wald's χ^2 = 11.40, df = 3, p = 0.010; Figure 3c). Focal TP *Sc. taeniopterus*

engaged in agonisms more frequently than TPs of other species (Wald's χ^2 = 27.65, df = 3, p <

0.001; Figure S5a), and their agonisms occurred closer to home range boundaries (χ^2 = 8.23, df =

3, p = 0.041; Figure S5b). Additionally, nearly all recorded interspecific agonisms (91.7%, n =

48) included TP *Sc. taeniopterus*, and the majority (52.3%) of these agonisms involved

interactions between focal TP *Sc. taeniopterus* and IP *Sc. vetula*. TP *Sc. taeniopterus* were also

involved in the majority (77.8%) of interspecific agonisms observed between TPs (n = 18). Of

the interspecific agonisms recorded between TP *Sc. taeniopterus* and other TPs (n = 14), 50%

were with TP *Scarus iseri*, 21.4% were with TP *Sc. vetula*, and the remaining were with TP *Sp.*

aurofrenatum or TP *Sp. viride* (14.3% each).

Dynamic interactions

The home ranges of simultaneously tracked interspecific pairs of TP *Sp. viride* and TP *Sc.*

vetula overlapped substantially (0.51 ± 0.05, mean ± SE), providing a strong basis for potential

dynamic interactions. However, the frequency of simultaneous relocations of these pairs within

shared areas did not differ from the frequency expected for independent movement (V = 24, p =

0.91). We found statistically significant evidence of dynamic interaction in one pair of TP *Sp.*

viride and TP *Sc. vetula* (Table S10; Video 1). This pair shared the most space (BA = 0.74) and

exhibited significant avoidance while moving in shared areas (Table S10). Both individuals in

this pair also had higher probabilities of being found in one another's home ranges and core areas

than individuals of most other pairs (Table S10). We also found marginally significant evidence

of avoidance in another pair of TP *Sp. viride* and TP *Sc. vetula*, which had similarly high spatial

overlap between their home ranges (BA = 0.64) and high probabilities of being found within one

another's home ranges and core areas (Table S10).

Discussion

We quantified spatial interactions between TPs (i.e., males) of four parrotfish species in

Bonaire to investigate the potential for movement and space use to mediate species interactions

and coexistence in these functionally important reef fishes. Males of multiple parrotfish species

are known to defend their daytime home ranges as fixed intraspecific territories from other males

[33–35]. Consistent with prior work on parrotfish territoriality, we found strong intraspecific

spatial segregation between neighboring TPs in *Sc. vetula, Sp. aurofrenatum,* and *Sp. viride,*

likely driven by the aggressive territorial interactions we observed. Agonistic interactions

generally occurred within 2-3 m of home range boundaries (Figure S6), suggesting that TPs

quickly detected intruders within their territories. TP core areas lie almost entirely within these

buffer zones (Figure S7), and the probability of finding neighboring TPs within core areas was

zero for the majority of intraspecific pairs for these three species. Thus, TPs likely have near-

exclusive access to the resources in these areas, aside from what they share with their harems of

IP females.

Territorial behavior evolves in response to competition for limited resources, and is only

tractable when resources are economically defendable (i.e., benefits of resource defense

outweigh the costs; [10]). The density of competitors and resource availability, among other

things, can strongly affect the cost-benefit balance of territory defense. Territory size is

negatively related to competitor density in some species and experiments have demonstrated that

removals of competitors may lead to territory expansion by remaining territory holders [34,57–

59]. Competitor density also affects patterns of spatial overlap among territory holders [57,60],

and direct encounters between intraspecific competitors can lead to home range shifts that persist

for long periods of time [61]. *Scarus taeniopterus* is the most abundant parrotfish at our study

sites [30]. The small home range sizes and increased intraspecific home range overlap that we

found for TP *Sc. taeniopterus* relative to our other species could reflect trade-offs between

territoriality and the density of intraspecific competitors. However, home range size also scales

positively with body size in many animal taxa, including parrotfishes [7,9,40,62], and *Sc.*

taeniopterus was the smallest of our study species. We cannot, therefore, rule out the effect of

body size on these patterns, nor are we able to determine the effects of resource availability on

observed spatial dynamics. Additional research is required to partition the effects of competitor

density, resource dynamics, and allometry on spacing patterns in *Sc. taeniopterus* and other

parrotfishes.

Resource partitioning is an important mechanism facilitating species coexistence. To

date, much of the literature regarding resource partitioning in parrotfishes has centered on trophic

resource partitioning. Microscopic photoautotrophs, including cyanobacteria, are the primary

nutritional resources for parrotfishes [26]. Parrotfishes partition these resources at fine spatial

scales by targeting substrates in different successional stages and with different species

compositions [28,31,63]. *Sparisoma viride* and *Sc. vetula* primarily target sparse epilithic turfs

growing atop endolithic communities, the former having increased access to endolithic

communities as an excavating species [53,64,65]. Both species primarily target turfs on high-

relief dead coral substrates, while *Sc. taeniopterus* primarily target turfs on boulder and rubble

[32]. In contrast, *Sparisoma* spp., including *Sp. aurofrenatum* and *Sp. viride* regularly target

fleshy macroalgae, though the macroalgae itself may not be the nutritional target [30,32,66].

While it is clear that parrotfishes partition trophic resources at various scales, it is less clear how

spatial partitioning influences coexistence.

Spatial segregation can mitigate interference competition and act as a stabilizing

mechanism of species coexistence [16]. For example, Berger and Gese [18] found that

differential use of shared space, as evidenced by spatial interaction analyses, could promote

coexistence in coyotes and wolves. Similarly, broad scale spatial segregation and fine scale

patterns of avoidance, which vary seasonally, have been reported as potential mechanisms of

coexistence in African carnivores and herbivores [20,67]. In this study, we found that the home

ranges of neighboring interspecific pairs of TPs overlapped more than the home ranges of

neighboring intraspecific pairs of TPs. However, spatial overlap was still relatively low (mean <

0.26), and the probability of finding neighboring TPs in one another's home range and core areas

varied by species. Specifically, TP *Sc. taeniopterus* were more likely be active within the home

ranges and core areas of other species than vice versa, and were most active in the home ranges

and core areas of TP *Sc. vetula*. TP *Sc. taeniopterus* were also involved in the majority of the

interspecific agonistic interactions that we observed, mostly interacting with IP *Sc. vetula*.

Haremic IPs share a substantial amount of space with and are very active in the TP home ranges

and core areas to which they belong (Section S1, Table S11, Figures S8 and S9). Frequent

interactions between TP *Sc. taeniopterus* and IP *Sc. vetula* likely reflect competition for shared

resources in the space they share with TP *Sc. vetula*. In contrast, TP *Sc. taeniopterus* were nearly

as likely to be found in neighboring TP *Sp. aurofrenatum* and TP *Sp. viride* home range and core

areas, but were less likely to interact agonistically with these species. This may reflect a greater

degree of trophic resource partitioning between *Sc. taeniopterus* and these other species, and

reduced importance of spatial segregation to their coexistence. These results suggest that spatial

partitioning may limit agonistic interactions between species, particularly when trophic resource

partitioning is not sufficient to prevent competition for shared resources.

Animals can also partition space temporally by avoiding each other when present in

shared space, thereby limiting interference competition (e.g., [68]). We detected significant

avoidance in a single interspecific pair of TP *Sp. viride* and TP *Sc. vetula*. The home ranges of

this pair had the greatest degree of spatial overlap of all simultaneously tracked pairs of TP *Sp.*

viride and TP *Sc. vetula*. Likewise, we detected marginal avoidance in another pair of TP *Sp.*

viride and TP *Sc. vetula* that also shared a substantial amount of space, with the second highest

measure of home range overlap. It seems reasonable to assume that neighboring individuals of

different parrotfish species may interact dynamically and avoid one another in shared space

dependent upon the degree of home range overlap, but that under typical conditions spatial

overlap may be low enough to limit the need for active avoidance. It is also possible that

dynamic avoidance takes place at smaller spatiotemporal scales. In the future, it may be

necessary to investigate dynamic interactions more locally to elucidate the underlying drivers of

dynamic interactions within shared space [56]. In light of our findings that TPs of *Sc.*

taeniopterus were more likely to be found within the home ranges and core areas of interspecific

neighbors than other species, investigations of dynamic interactions between TP *Sc. taeniopterus*

and other TP parrotfishes may yield stronger evidence of temporal segregation within shared

space for Caribbean parrotfishes.

Home range sizes are correlated with the abundance and quality of resources [69,70],

including for parrotfishes [39], and animals are predicted to increase the size of their home

ranges in less productive habitats [8]. Anthropogenic stressors, including climate change, are

leading to rapid and dramatic shifts in the structure and composition of benthic communities on

coral reefs [71]. As these changes continue, we might expect the potential for spatial interactions

to change as home ranges expand, contract, and/or shift in response to changes in resource

availability. Baseline studies of spatial interactions such as this are, therefore, increasingly

important for understanding the effects of climate change on reef fish communities [72].

The results presented here elucidate a few of the mechanisms underlying movement and

space use in parrotfishes. Specifically, we found that competition drove stronger spatial

segregation between TPs belonging to the same species than to different species. Dynamic

avoidance does occur between TPs belonging to different species, though such interactions may

depend on how much space each member of a pair shares, and how such shared space is used.

Regardless the presence of dynamic avoidance suggests that spatiotemporal segregation could

act as a stabilizing mechanism mediating coexistence in parrotfishes. Our results have

implications for understanding the trophic effects of parrotfishes on coral reefs. The effects of

grazers on autotrophs depends, in part, on spatial variation in grazing pressure [73]. Home range

and territorial behavior may concentrate parrotfish grazing and bioerosion locally [37],

contributing to spatial heterogeneity in benthic reef communities. As such, studies of the spatial

ecology of parrotfishes, particularly those that investigate spatial interactions, are necessary to

provide a more complete understanding of their functional roles on coral reefs.

Acknowledgements and funding statement

This material is based upon research supported by the Chateaubriand Fellowship of the Office

for Science & Technology of the Embassy of France in the United States. The research was also

supported by Florida State University (FSU) start-up funding and the Tatelbaum Ocean Research

Fund awarded to SJM, as well as a Mote Research Assistantship and Jack Winn Gramling

Research Award from FSU, a Guy Harvey Scholarship, and a Society for Integrative and

Comparative Biology Grant-In-Aid of Research awarded to JCM. We would like to thank R.

Francisca and C. Eckrich at Stinapa Bonaire for assistance with permitting to work in the

Bonaire National Marine Park. We are also grateful for the valuable feedback that E. Cissell and

A. Rassweiler provided on earlier versions of this manuscript. Finally, we thank L. Kury and K.

Laforest for assisting with data collections.

References

1. Van Moorter B, Rolandsen CM, Basille M, Gaillard JM. 2016 Movement is the glue connecting home ranges and habitat selection. *J. Anim. Ecol.* **85**, 21–31. (doi:10.1111/1365-2656.12394)

2. Nathan R, Getz WM, Revilla E, Holyoak M, Kadmon R, Saltz D, Smouse PE. 2008 A movement ecology paradigm for unifying organismal movement research. *Proc. Natl. Acad. Sci. USA* **105**, 19052–19059.

3. Börger L, Dalziel BD, Fryxell JM. 2008 Are there general mechanisms of animal home range behaviour? A review and prospects for future research. *Ecol. Lett.* **11**, 637–650. (doi:10.1111/j.1461-0248.2008.01182.x)

4. Haskell JP, Ritchie ME, Olff H. 2002 Fractal geometry predicts varying body size scaling relationships for mammal and bird home ranges. *Nature* **418**, 527–530. (doi:10.1038/nature00840)

5. Riotte-Lambert L, Benhamou S, Chamaillé-Jammes S. 2015 How memory-based movement leads to nonterritorial spatial segregation. *Am. Nat.* **185**, E103–E116. (doi:10.1086/680009)

6. Wakefield ED *et al.* 2013 Space partitioning without territoriality in gannets. *Science (80-.).* **341**, 68–70. (doi:10.1126/science.1236077)

7. Tamburello N, Côté IM, Dulvy NK. 2015 Energy and the scaling of animal space use. *Am. Nat.* **186**, 196–211. (doi:10.1086/682070)

8. Harestad AS, Bunnell FL. 1979 Home range and body weight - a reevaluation. *Ecology*

60, 389–402.

9.   Nash KL, Welsh JQ, Graham NAJ, Bellwood DR. 2015 Home-range allometry in coral

reef fishes: comparison to other vertebrates, methodological issues and management

implications. *Oecologia* **177**, 73–83. (doi:10.1007/s00442-014-3152-y)

10.   Brown JL. 1964 The evolution of diversity in avian territorial systems. *Wilson Bull.* **76**,

160–169.

11.   Kaufmann JH. 1983 On the definitions and functions of dominance and territoriality. *Biol.*

Rev. **58**, 1–20.

12.   López-Sepulcre A, Kokko H. 2005 Territorial defense, territory size, and population

regulation. *Am. Nat.* **166**, 317–329.

13.   Wang M, Grimm V. 2007 Home range dynamics and population regulation: An

individual-based model of the common shrew *Sorex araneus*. *Ecol. Modell.* **205**, 397–409.

(doi:10.1016/j.ecolmodel.2007.03.003)

14.   Wood JLA, Grant JWA, Belanger MH. 2012 Population density and territory size in

juvenile rainbow trout, *Oncorhynchus mykiss*: implications for population regulation. *Can.*

J. Fish. Aquat. Sci. **69**, 1121–1128. (doi:10.1139/F2012-048)

15.   Newton I. 1992 Experiments on the limitation of bird numbers by territorial behaviour.

Biol. Rev. **67**, 129–173.

16.   Schlägel UE *et al.* 2020 Movement-mediated community assembly and coexistence. *Biol.*

Rev. **95**, 1073–1096. (doi:10.1111/brv.12600)

17.   Chesson P. 2000 Mechanisms of maintenance of species diversity. *Annu. Rev. Ecol. Syst.*

31, 343–366. (doi:10.1146/annurev.ecolsys.31.1.343)

18. Berger KM, Gese EM. 2007 Does interference competition with wolves limit the

distribution and abundance of coyotes? *J. Anim. Ecol.* **76**, 1075–1085.

(doi:10.1111/j.1365-2656.2007.01287.x)

19. Papastamatiou YP, Bodey TW, Friedlander AM, Lowe CG, Bradley D, Weng K, Priestley

V, Caselle JE. 2018 Spatial separation without territoriality in shark communities. *Oikos*

127, 767–779. (doi:10.1111/oik.04289)

20. Macandza VA, Owen-Smith N, Cain JW. 2012 Dynamic spatial partitioning and

coexistence among tall grass grazers in an African savanna. *Oikos* **121**, 891–898.

(doi:10.1111/j.1600-0706.2012.20250.x)

21. Lear KO, Whitney NM, Morris JJ, Gleiss AC. 2021 Temporal niche partitioning as a

novel mechanism promoting co-existence of sympatric predators in marine systems. *Proc.*

R. Soc. B Biol. Sci. **288**. (doi:10.1098/rspb.2021.0816)

22. Bruggemann JH, Van Kessel AM, Van Rooij JM, Breeman AM. 1996 Bioerosion and

sediment ingestion by the caribbean parrotfish *Scarus vetula* and *Sparisoma viride*:

Implications of fish size, feeding mode and habitat use. *Mar. Ecol. Prog. Ser.* **134**, 59–71.

(doi:10.3354/meps134059)

23. Bellwood DR. 1995 Carbonate transport and within-reef patterns of bioerosion and

sediment release by parrotfishes (family Scaridae) on the Great Barrier Reef. *Mar. Ecol.*

Prog. Ser. **117**, 127–136.

24. Bellwood DR. 1995 Direct estimate of bioerosion by two parrotfish species, *Chlorurus*

gibbus and *C. sordidus*, on the Great Barrier Reef, Australia. *Mar. Biol.* **121**, 419–429.

(doi:10.1007/BF00349451)

25.   Alwany MA, Thaler E, Stachowitsch M. 2009 Parrotfish bioerosion on Egyptian Red Sea

reefs. *J. Exp. Mar. Bio. Ecol.* **371**, 170–176. (doi:10.1016/j.jembe.2009.01.019)

26.   Clements KD, German DP, Piché J, Tribollet A, Choat JH. 2017 Integrating ecological

roles and trophic diversification on coral reefs: multiple lines of evidence identify

parrotfishes as microphages. *Biol. J. Linn. Soc.* **120**, 729–751. (doi:10.1111/bij.12914)

27.   Grange JS, Rybarczyk H, Tribollet A. 2015 The three steps of the carbonate biogenic

dissolution process by microborers in coral reefs (New Caledonia). *Environ. Sci. Pollut.*

Res. **22**, 13625–13637. (doi:10.1007/s11356-014-4069-z)

28.   Nicholson GM, Clements KD. 2020 Resolving resource partitioning in parrotfishes

(Scarini) using microhistology of feeding substrata. *Coral Reefs* **39**, 1313–1327.

(doi:https://doi.org/10.1007/s00338-020-01964-0)

29.   Cissell EC, Manning JC, McCoy SJ. 2019 Consumption of benthic cyanobacterial mats on

a Caribbean coral reef. *Sci. Rep.* **9**, 12693. (doi:10.1038/s41598-019-49126-9)

30.   Manning JC, McCoy SJ. 2023 Preferential consumption of benthic cyanobacterial mats by

Caribbean parrotfishes. *Coral Reefs* , 1–9. (doi:https://doi.org/10.1007/s00338-023-02404-

5)

31.   Nicholson GM, Clements KD. 2021 Ecomorphological divergence and trophic resource

partitioning in 15 syntopic Indo-Pacific parrotfishes (Labridae: Scarini). *Biol. J. Linn. Soc.*

132, 590–611. (doi:10.1093/biolinnean/blaa210)

32.   Adam TC, Kelley M, Ruttenberg BI, Burkepile DE. 2015 Resource partitioning along

multiple niche axes drives functional diversity in parrotfishes on Caribbean coral reefs.

Oecologia **179**, 1173–1185. (doi:10.1007/s00442-015-3406-3)

33.   van Rooij JM, Kroon FJ, Videler JJ. 1996 The social and mating system of the

herbivorous reef fish *Sparisoma viride*: one-male versus multi-male groups. *Environ. Biol.*

Fishes **47**, 353–378. (doi:10.1007/BF00005050)

34.   Mumby PJ, Wabnitz CCC. 2002 Spatial patterns of aggression, territory size, and harem

size in five sympatric Caribbean parrotfish species. *Environ. Biol. Fishes* **63**, 265–279.

35.   Manning JC, McCoy SJ. 2023 Territoriality drives patterns of fixed space use in

Caribbean parrotfishes. *Ecol. Evol.* **13**, e9833. (doi:https://doi.org/10.1002/ece3.9833)

36.   Pickholtz R, Kiflawi M, Crossin GT, Pickholtz E, Zamsky R, Kahan I, Gavriel T,

Belmaker J. 2022 Highly repetitive space-use dynamics in parrotfishes. *Coral Reefs*

(doi:10.1007/s00338-022-02258-3)

37.   Welsh JQ, Bellwood DR. 2012 Spatial ecology of the steephead parrotfish (*Chlorurus*

microrhinos): An evaluation using acoustic telemetry. *Coral Reefs* **31**, 55–65.

(doi:10.1007/s00338-011-0813-8)

38.   Sandin SA, McNamara DE. 2012 Spatial dynamics of benthic competition on coral reefs.

Oecologia **168**, 1079–1090. (doi:10.1007/s00442-011-2156-0)

39.   Carlson PM, Davis K, Warner RR, Caselle JE. 2017 Fine-scale spatial patterns of

parrotfish herbivory are shaped by resource availability. *Mar. Ecol. Prog. Ser.* **577**, 165–

176. (doi:10.3354/meps12258)

40.   Welsh JQ, Goatley CHR, Bellwood DR. 2013 The ontogeny of home ranges: Evidence

from coral reef fishes. *Proc. R. Soc. B Biol. Sci.* **280**, 20132066.

(doi:10.1098/rspb.2013.2066)

41.   Steneck RS, Arnold SN, Boenish R, de León R, Mumby PJ, Rasher DB, Wilson MW.

2019 Managing recovery resilience in coral reefs against climate-induced bleaching and

hurricanes: A 15 year case study from Bonaire, Dutch Caribbean. *Front. Mar. Sci.* **6**, 1–

12. (doi:10.3389/fmars.2019.00265)

42.   Jackson J, Donovan M, Cramer K, Lam V, editors. 2014 *Status and Trends of Caribbean*

Coral Reefs: 1970-2012. Gland, Switzerland. See

http://wedocs.unep.org/bitstream/handle/20.500.11822/9230/-Status and Trends of

Caribbean Coral Reefs%3A 1970-2012-2014Caribbean Coral Reefs - Status Report 1970-

2012.pdf?sequence=3&isAllowed=y.

43.   Kennedy Rhoades O, Lonhart SI, Stachowicz JJ. 2018 Fished species uniformly reduced

escape behaviors in response to protection. *Biol. Conserv.* **226**, 238–246.

(doi:10.1016/j.biocon.2018.06.030)

44.   Benhamou S. 2014 Of scales and stationarity in animal movements. *Ecol. Lett.* **17**, 261–

272. (doi:10.1111/ele.12225)

45.   Benhamou S. 2011 Dynamic approach to space and habitat use based on biased random

bridges. *PLoS One* **6**, e14592. (doi:10.1371/journal.pone.0014592)

46.   Bhattacharyya A. 1943 On a measure of divergence between two statistical populations

defined by their probability distributions. *Bull. Calcutta Math. Soc.* **35**, 99–109.

47.   Fieberg J, Kochanny CO. 2005 Quantifying home-range overlap: The importance of the

utilization distribution. *J. Wildl. Manage.* **69**, 1346–1359.

48.   Calenge C. 2006 The package adehabitat for the R software: tool for the analysis of space

and habitat use by animals. *Ecol. Modell.* **197**, 1035.

49.   Brooks ME, Kristensen K, van Benthem KJ, Magnusson A, Berg CW, Nielsen A, Skaug

HJ, Maechler M, Bolker BM. 2017 glmmTMB Balances Speed and Flexibility Among

Packages for Zero-inflated Generalized Linear Mixed Modeling. *R J.* **9**, 378–400.

50.   Bayer FM, Cribari-Neto F. 2017 Model selection criteria in beta regression with varying

dispersion. *Commun. Stat. - Simul. Comput.* **46**, 729–746.

(doi:10.1080/03610918.2014.977918)

51.   Lenth R V. 2020 emmeans: Estimated Marginal Means, aka Least-Squares Means.

52.   Friard O, Gamba M. 2016 BORIS: a free, versatile open-source event-logging software

for video/audio coding and live observations. *Methods Ecol. Evol.* **7**, 1325–1330.

(doi:10.1111/2041-210X.12584)

53.   Bruggemann JH, Kuyper MWM, Breeman AM. 1994 Comparative-analysis of foraging

and habitat use by the sympatric Caribbean parrotfish *Scarus vetula* and *Sparisoma viride*

(Scaridae). *Mar. Ecol. Prog. Ser.* **112**, 51–66. (doi:10.3354/meps112051)

54.   Pebesma E. 2018 Simple features for R: Standardized support for spatial vector data. *R J.*

10, 439–446.

55.   Benhamou S, Valeix M, Chamaillé-Jammes S, Macdonald DW, Loveridge AJ. 2014

Movement-based analysis of interactions in African lions. *Anim. Behav.* **90**, 171–180.

(doi:10.1016/j.anbehav.2014.01.030)

56.   Long JA, Nelson TA, Webb SL, Gee KL. 2014 A critical examination of indices of

dynamic interaction for wildlife telemetry studies. *J. Anim. Ecol.* **83**, 1216–1233.

(doi:10.1111/1365-2656.12198)

57.   Stamps JA. 1990 The effect of contender pressure on territory size and overlap in

seasonally territorial species. *Am. Nat.* **135**, 614–632. (doi:10.1086/285065)

58.   Norman MD, Jones GP. 1984 Determinants of territory size in the pomacentrid reef fish,

Parma victoriae. *Oecologia* **61**, 60–69. (doi:10.1007/BF00379090)

59.   Tricas TC. 1989 Determinants of feeding territory size in the corallivorous butterflyfish,

Chaetodon multicinctus. *Anim. Behav.* **37**, 830–841.

60.   Mazerolle DF, Hobson KA. 2004 Territory size and overlap in male Ovenbirds:

Contrasting a fragmented and contiguous boreal forest. *Can. J. Zool.* **82**, 1774–1781.

(doi:10.1139/Z04-175)

61.   Fagan WF *et al.* 2023 Intraspecific encounters can induce home-range shifts. *bioRxiv* , 1–

30.

62.   McNab BK. 1963 Bioenergetics and the determination of home range size. *Am. Nat.* **97**,

133–140.

63.   Nicholson GM, Clements KD. 2022 *Scarus spinus*, crustose coralline algae and

cyanobacteria: an example of dietary specialization in the parrotfishes. *Coral Reefs*

(doi:10.1007/s00338-022-02295-y)

64.   Bruggemann JH, van Oppen MJH, Breeman AM. 1994 Foraging by the stoplight

parrotfish *Sparisoma viride*. I . Food selection in different, socially determined habitats.

Mar. Ecol. Prog. Ser. **106**, 41–55. (doi:10.3354/meps106041)

65.   Bruggemann JH, Begeman J, Bosma EM, Verburg P, Breeman AM. 1994 Foraging by the

stoplight parrotfish *Sparisoma viride*. II. Intake and assimilation of food, protein and

energy. *Mar. Ecol. Prog. Ser.* **106**, 57–72. (doi:10.3354/meps106057)

66.   Dell CLA, Longo GO, Burkepile DE, Manfrino C. 2020 Few herbivore species consume

dominant macroalgae on a Caribbean coral reef. *Front. Mar. Sci.* **7**, 1–12.

(doi:10.3389/fmars.2020.00676)

67.   Vanak AT, Fortin D, Thaker M, Ogden M, Owen C, Greatwood S, Slotow R. 2013

Moving to stay in place: Behavioral mechanisms for coexistence of African large

carnivores. *Ecology* **94**, 2619–2631. (doi:10.1890/13-0217.1)

68.   Kortello AD, Hurd TE, Murray DL. 2007 Interactions between cougars (*Puma concolor*)

and gray wolves (*Canis lupus*) in Banff National Park, Alberta. *Ecoscience* **14**, 214–222.

69.   Relyea RA, Lawrence RK, Demarais S. 2000 Home range of desert mule deer: testing the

body-size and habitat-productivity hypotheses. *J. Wildl. Manage.* **64**, 146–153.

70.   Saïd S, Gaillard JM, Widmer O, Débias F, Bourgoin G, Delorme D, Roux C. 2009 What

shapes intra-specific variation in home range size? A case study of female roe deer. *Oikos*

118, 1299–1306. (doi:10.1111/j.1600-0706.2009.17346.x)

71.   Hughes TP. 1994 Catastrophes, phase shifts, and large-scale degradation of a Caribbean

coral reef. *Science (80-.).* **265**, 1547–1551. (doi:10.1126/science.265.5178.1547)

72.   Manning JC. 2022 Movement, space use, and the responses of coral reef fishes to climate

577   change. *Integr. Comp. Biol.* , icac128. (doi:10.1093/icb/icac128)

578 73. Olff H, Ritchie ME. 1998 Effects of herbivores on grassland plant diversity. *Trends Ecol.*

579   *Evol.* **13**, 261–265. (doi:10.1016/S0169-5347(98)01364-0)

Figure Captions

Figure 1: Mean (±SE) home range and core areas for TP (dark) and IP (light) of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride*.

Figure 2: (a) A heatmap of the mean spatial overlap of the home ranges of neighboring TPs of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride*. Lighter shades of blue indicate less spatial overlap. The mean (±SE) spatial overlap and the proportion of neighboring pairs with zero spatial overlap are displayed for each species pairing. (b) A heatmap of the mean probability of finding TPs of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride* (Species 2) in the home ranges of neighboring TPs of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride* (Species 1). Lighter shades of red indicate lower probabilities. The mean (±SE) probabilities and the proportions of neighboring pairs where the probability is zero are displayed for each species pairing. (c) A heatmap of the mean probability of finding TP of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride* (Species 2) in the core areas of neighboring TPs of *Sc. taeniopterus*, *Sc. vetula*, *Sp. aurofrenatum*, and *Sp. viride* (Species 1). Lighter shades of purple indicate lower probabilities. The mean (±SE) probabilities and the proportions of neighboring pairs where the probability is zero are displayed for each species pairing.

Figure 3: Boxplots of (a) the number of agonisms, (b) agonism durations (s), and (c) the distances (m) of those agonisms from the focal home range boundary for agonisms between focal TPs and IPs belonging to either the same species (IP-intraspecific) or another species (IP-interspecific), and TPs belonging either to the same species (TP-intraspecific) or another species (TP-interspecific).

604 **Figures**

606 Figure 1

608   Figure 2

611   Figure 3