

16s rRNA gene sequence analysis of the microbial community on microplastic samples from the North Atlantic and Great Pacific Garbage Patches

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ABSTRACT

The exponential increase in plastic production has led to their accumulation in the environment, particularly in oceans, polluting these environments from the shore to the open ocean and even sea ice in the pole regions. We compared microbial communities on plastic particles, known as "Plastisphere", collected from the Atlantic and Pacific oceans gyres in the Summer of 2019 and subsequently looked for potential plastic degraders. We applied a 16S rRNA amplicon sequencing approach to decipher differences and similarities in colonization behaviour between these two gyres. Two polymer types include plastics: polyethylene (PE) and polypropylene (PP). We found that microbes differed significantly between the two oceans and identified thirty-two differentially abundant taxa at the class level. Proteobacteria, Cyanobacteria and Bacteroidota were the most prominent relative abundant phyla in the two oceans. Finally, according to the current literature, we found 40 genera documented as potential plastic degraders. This study highlights the importance of the biogeographical location with respect to microbial colonization patterns of marine plastic debris, differing even in the open oceans. Furthermore, the wide distribution of potential plastic-degrading bacteria was shown.

Keywords: Plastisphere, microbial communities, plastic degraders.

INTRODUCTION

4.8 to 12.7 million metric tons of plastics are estimated to enter the ocean yearly, mostly from land (Jambeck et al., 2015). Dris et al. (2016) related an atmospheric fallout between 2 and 355 particles/m²/day. Microscopic plastic fragments and fibres (microplastics) are widespread in the oceans. They have accumulated in the pelagic zone and sediments resulting from the degradation of macroplastic items (Thompson et al., 2004). The ubiquitous plastics in the ocean could harm the marine environment and humans through the food web. Evidence showed that microplastics could act as passive samplers for toxic compounds such as persistent organic pollutants (e.g., industrial chemicals, dioxins,

47 pesticides) and heavy metals from seawater, leading to an increased negative impact on
48 the biota (Mato et al., 2001; Horn et al., 2019). Besides that, the potential accumulation
49 of microplastics in the food chain, especially in fish and shellfish, also exposes human
50 consumers to these adsorbed chemicals (Kershaw & Rochman, 2014). Lusher et al.
51 (2013) reported that 36% of pelagic and demersal fish collected from the English Channel
52 had microplastics in their gastrointestinal tract.

53 Outreaches, national and international laws, policies, and conventions have been
54 discouraging the use of plastic and its release into the environment to face plastic
55 pollution. The African continent is at the forefront of legislative actions against plastic
56 pollution. For instance, Rwanda has banned non-biodegradable plastic since 2008 and
57 single-use plastics in 2019. The ban prohibited the manufacturing, use, import and sale
58 of plastic carrier bags and forbade travellers into Rwanda to come with such products.
59 Nigeria announced a ban on plastic bags in 2013, and in 2020, it strengthened its
60 legislation by including a fine of 1072.16 Euro or three years jail term for any store found
61 giving plastic bags to customers. In Botswana, a minimum thickness for bags was
62 established and mandated that retailers apply a minimum levy to thicker bags, which
63 would support government environmental projects. Kenya has the strictest ban on single-
64 use plastic globally and in protected areas (Greenpeace, 2021).

65 Plastic is a high molecular weight synthetic polymer of a long chain of hydrocarbons
66 derived from petrochemicals (Ahmed et al., 2018). With swift development in molecular
67 techniques, research focused on microbial communities living on plastics and their ability
68 to degrade hydrocarbons. The biological deterioration of plastic pollutants depends on
69 many factors: surface area, functional groups, molecular weight, hydrophilic and
70 hydrophobicity, melting temperature, chemical structure, crystallinity, etc. (Okada,
71 2002). Microbial degradation of plastic involves many steps: biodeterioration, bio-
72 fragmentation, assimilation, and mineralization (Purohit et al., 2020).

73 Zettler, Mincer & Amaral-Zettler (2013) coined the term "Plastisphere" to describe
74 biofilm-forming communities on marine plastic debris. They collected marine plastic
75 debris at multiple locations in the North Atlantic to analyze the microbial consortia
76 attached to it. They found diverse microbial communities, including heterotrophs,
77 autotrophs, predators, and symbionts, which they called a 'Plastisphere'. Coons et al.
78 (2021) investigated plastic-type and incubation locations in the Atlantic and Pacific
79 oceans, focusing on shore locations as drivers of marine bacterial community structure
80 development on plastic via 16S rRNA gene amplicon analysis. They found that
81 incubation location was the primary driver of the coastal Plastisphere composition. The
82 bacterial communities were consistently dominated by the classes Alphaproteobacteria,
83 Gammaproteobacteria, and Bacteroidia, irrespective of sampling location or substrate
84 type.

85 Similarly, in 2015, Amaral-Zettler et al. used next-generation DNA
86 (Deoxyribonucleic Acid) sequencing to characterize bacterial communities from the
87 Pacific and Atlantic oceans. Their objective was to determine whether the composition of
88 Plastisphere communities reflects their biogeographic origins. They found that these
89 communities differed between ocean basins and, to a lesser extent, between polymer types
90 and displayed latitudinal gradients in species richness.

91 For this work, we collected plastic particles from the North Atlantic and the Great
92 Pacific Garbage Patches in 2019 and compared microbial communities from the Atlantic
93 to the Pacific, as well as looking for potential plastic degraders. The North Atlantic and
94 the Great Pacific Garbage Patches are the biggest current patches, with a density of 10^6
95 km^{-2} (Eriksen et al., 2014) and 96 400 million metric tons of plastic (Ritchie & Roser,
96 2018), respectively. In addition to the work of Zettler et al. (2015), we pointed out

97 communities responsible for the discrepancy. Besides, our samples were composed of PP
98 and PE polymers for which only scarce information regarding environmental degradation
99 is available. To reach our goal, 16S rRNA gene amplicon sequence analysis was used on
100 the microbial community of these microplastic samples.

101 **MATERIALS AND METHODS**

102 **Plastics collection**

103

104 The samples were collected in the North Atlantic and Great Pacific Garbage Patches. The
105 pieces from the Atlantic were collected between 26-08-2019 and 04-09-2019 during the
106 POS536 cruise project 'Distribution of Plastics in the North Atlantic Garbage Patch'
107 (DIPLANOAGAP) aboard the German research vessel (R/V) Poseidon. A Neuston
108 catamaran onboard R/V Poseidon, equipped with a microplastic trawl net (mesh size 300
109 μm , mouth opening 70 cm x 40 cm) was used to collect the plastic samples from the sea
110 surface. After each tow, all microplastic fragments were removed from the trawl sample
111 and conserved in a saturated ammonium sulphate solution (700 g/l ammonium sulfate, 20
112 mM sodium citrate, 25mM EDTA, pH 5.2). This solution precipitated all proteins,
113 preventing DNA and RNA degradation for an extended time, even at room temperature.
114 Verification of plastic type by Attenuated Total Reflectance Fourier-Transform Infrared
115 (ATR-FTIR) spectroscopy analysis was subsequently performed by TUTECH GmbH in
116 Hamburg, Germany.

117 Another cruise project, MICRO-FATE, aboard another German R/V, the Sonne
118 (SO268/3), between 05-06-2019 and 27-06-2019, was used to collect plastic samples at
119 the sea surface in the Great Pacific garbage patch. Plastics were collected using a scoop
120 net sampling method. The plastic surfaces were scraped using a flame-sterilized scalpel,
121 and biofilms were transferred into microcentrifuge tubes. The sampling area was 16 x 16
122 mm, and tubes were immediately frozen in liquid nitrogen. At each station, 1 litre of
123 pacific water was filtered through a 3 μm filter (3 μm Isopore TSTP 04700 Millipore,
124 Merck KGaA, Frankfurt, Germany) and a 0.22 μm filter (0.22 μm Isopore GTTP04700
125 membrane filters Millipore, Merck KGaA, Frankfurt, Germany). Also, the filters were
126 transferred to microcentrifuge tubes and immediately frozen in liquid nitrogen.

127 **Extraction of nucleic acids from the samples**

128 For Atlantic samples, sections of the different plastic samples were cut with a sterile
129 scalpel and placed into 2 ml MP Biomedicals™ Lysing Matrix E tubes (MP Biomedicals,
130 Eschwege, Germany). Then physically disrupted using a bead-beating technique, with a
131 single cycle of 30s at a speed of 5.500 rpm in a FastPrep homogenizer (Qiagen, Hilden,
132 Germany). The DNA extraction from the lysis product was then performed using the
133 Qiagen AllPrep DNA/RNA Minikit according to the manufacturer's instructions. The
134 quality and quantity of the DNA extraction were assessed using a NanoDrop
135 Spectrophotometer (Desjardins & Conklin, 2010). The 16S rRNA gene was amplified
136 with the primer pair 27F and 1492R. Then, the quality was checked through polymerase
137 chain reaction (PCR). The sequencing of the V3-V4 region of the 16S rRNA gene was
138 performed with v3 chemistry on a MiSeq Illumina sequencing platform at the
139 Competence Centre for Genomic Analysis (CCGA) Kiel, Germany after the PCR
140 products were visually assessed using 1% gel electrophoresis. For amplicon sequencing,
141 the amplification of the V3-V4 hypervariable region of the 16S rRNA gene was
142 accomplished using primer pair 341F (50-CCTACGGGAGGCAGCAG-30; Muyzer et

143 al., 1993) and 806R (50-GGACTACHVGGGTWTCTAAT-30; Caporaso et al., 2011).
144 Raw reads were archived in NCBI under the BioProject number PRJNA901861.

145 For Pacific samples, DNA was extracted from the biofilm pellets and water filters
146 using the Macherey Nagel DNA Nucleo spin soil kit (Nucleo Spin TM Soil kit Macherey-
147 Nagel TM, Düren, Germany) according to the manufacturer's instructions. DNA
148 concentration was measured using a nano Qubit (ThermoFisher). Next-generation
149 Illumina Sequencing was performed on an Illumina MiSeq platform using a V3 (300bp
150 paired-end read) kit with a sequencing amount of 20 million reads, using the 341F
151 (CCTACGGGNGGCWGCAG) and 785R primer set
152 (GACTACHVGGGTATCTAAKCC). Raw reads were archived in NCBI under the
153 BioProject number PRJNA837054.

154 Quantitative Insights Into Microbial Ecology (QIIME2) pipeline

155 The Raw amplicon sequences were then processed using the open-source Quantitative
156 Insights into Microbial Ecology (QIIME2, version 2020.11) following a pipeline
157 developed by Kathrin Busch (Busch et al., 2021). In brief, the *cutadapt* plugin was used
158 to trim forward primers, heterogeneity spacers from forward-only single-end fastq files
159 (Martin, 2011) and the *qualityfilter* plugin (Bokulich et al., 2013) was used to check the
160 quality of the demultiplexed reads. An interactive plot served to visualize these results
161 and to determine an appropriate truncation length. Then, the reads were truncated through
162 the DADA2 algorithm to produce a total read length of 270 nucleotides. That truncation
163 significantly increased the quality of the reads, reduced the overlap between forward and
164 reverse reads, and allowed us to use only forwards reads for the analysis. Before the
165 truncation, the reads were denoised using the *denoise-single* method of the DADA2
166 algorithm (Callahan et al., 2016), which removed chimeric sequences and inferred sample
167 composition using a parametric error model.

168 The amplicon sequence variants (ASV; Callahan et al., 2017) were classified at 80%
169 confidence level using the most recent SILVA 138 16S rRNA gene reference database
170 (Quast et al., 2013; Yilmaz et al., 2014). Common eukaryotic contaminants (chloroplasts,
171 mitochondria) and unassigned sequences were removed using the *filter-features* method
172 of the *featuretable* plugin, and the resulting dataset was rarefied to 8,000 sequences.
173 Alpha rarefaction curves have an excellent saturation for 8,000 sequences. A
174 phylogenetic backbone tree was built using FastTree (Price et al., 2009; Price et al., 2010)
175 and MAFFT (Kato & Standley, 2013) alignment through the *phylogeny* plugin. The
176 resulting tree was used to compute core diversity metrics which served to compute
177 downstream analyses along with an alpha-rarefaction curve *via* the *diversity* plugin.

178 Alpha and Beta diversity measures

179 The alpha diversity was investigated according to unique ASVs per sample (species
180 richness), taking into consideration the number of times each ASV occurs in the sample
181 (Pielou's evenness) and the phylogenetic relatedness of each sample community (Faith's
182 PD). '*QIIME diversity alpha-group-significance*' plugin in QIIME2 was used to assess the
183 diversity within each area. The results were displayed through Kruskal-Wallis (all groups)
184 and Kruskal-Wallis (pairwise) results.

185 Non-phylogenetic (evenness) and phylogenetic (Faith's PD) diversity indices were
186 visualized using the online tool QIIME2 view (<https://view.qiime2.org/>). Eventually, if
187 the comparison revealed a significant difference in microbial diversity, Kruskal-Wallis
188 pairwise was considered among groups to see where the difference lies.

189 Beta diversity measures assessed the differences between groups following the
190 different parameters. 'Qiime diversity beta-group-significance' plugin in QIIME2 was
191 used for this analysis. The analysis was performed using the non-metric multidimensional
192 scaling method (NMDS; Kruskal, 1964) with a sample-wise unweighted UniFrac
193 distance matrix (Lozupone & Knight, 2005). Each group was assessed based on its
194 distance from the other groups in QIIME2; boxplots were displayed simultaneously with
195 the PERMANOVA results and pairwise PERMANOVA results between groups. The
196 PERMANOVA group significance and pairwise tests were run simultaneously through
197 the *betagroup-significance* method (non-parametric MANOVA; Anderson, 2001) of the
198 diversity plugin with an unweighted UniFrac matrix and 999 permutations as input.
199 We adopted the standard significant measure, p-value = 0.05, for these statistical analyses.
200 All the p-values below this standard describe a significant difference between the
201 compared parameters and vice versa.

202

203 Different taxonomic level analysis

204

205 The feature ASVs table was exported in biom format in QIIME2. Subsequently, the
206 taxonomy metadata file was added to the biom file and exported in TSV file format using
207 '*biom convert*' plugin in QIIME2. Further analyses outside the QIIME2 environment, such
208 as the share of ASVs between the samples, were performed using the resulting TSV file
209 table. Besides that, the same feature table was collapsed at the genus level (to perform the
210 sunburst plots, which helped to display microbial communities on plastics) and the class
211 level (to plot the differentially abundant taxa) using the '*qiime taxa collapse*' plugin. The
212 Linear discriminant analysis (LDA) effect size (LEfSe) helped to plot the differentially
213 abundant classes between the Atlantic and the Pacific Plastisphere, utilizing 'galaxy
214 online' (<https://huttenhower.sph.harvard.edu/galaxy/>). The level 3 data was used,
215 arranged within Excel (according to the different oceans) and imported into Galaxy for
216 LEfSe analysis. The analyses were performed on the microbial community relative
217 abundance data in both oceans. Grouped data were first analyzed using the Kruskal-
218 Wallis test with a significance threshold of 0.05 to determine if the data was differentially
219 distributed between groups.

220

221 **RESULTS & DISCUSSION**

222

223 Samples collected comprised 68 microplastic pieces from the North Atlantic and the
224 Great Pacific Garbage Patches as well as 14 water samples from the Great Pacific
225 Garbage Patch. The North Atlantic Garbage Patch accounted for 30 plastic samples
226 composed of 25 PE (polyethylene) and 5 PP (Polypropylene) particles, according to FTIR
227 analysis. In contrast, the Great Pacific Garbage Patch accounted for 38 plastic samples,
228 composed of 28 PE and 10 PP (Supplementary table S1).

229 Processing all samples in QIIME2 yielded 11,852 demultiplexed ASVs. Pacific
230 plastic accounted for 7,081 ASVs showing more richness than Atlantic plastic which
231 accounted for 4,454 ASVs and, in turn, showed more richness than Pacific water (3,623
232 ASVs). The Pacific plastisphere displayed the highest number of taxa at most taxonomic
233 levels except the phylum level (see Table 1). Here, the Atlantic plastisphere displayed the
234 highest number of phyla with 35, whereas Pacific plastic contained 33 phyla and Pacific
235 water 27 phyla. Overall, the Pacific water displayed the lowest amount of taxa irrespective
236 of the taxonomic level, which might hint towards a microhabitat formation on the plastic
237 particles as they travel across the oceans and enrich their community along the way. These
238 results also show an increasing proportion of unclassified taxa as one moves from the

239 phylum level to the genus level, which informs that much is still to be discovered. The
240 Shannon diversity indice values are between 4.88 and 8.75 in the individual samples
241 (Supplementary table S1), with no apparent large differences between the different
242 sample types and locations.

243

244 The Principal Coordinates Analysis (PCoA) of all samples

245

246 A PCoA plot, grouping all the samples, was generated from QIIME2 to infer the
247 phylogenetic relatedness between the samples' communities of both oceans. Three
248 clusters, as shown in Figure 1, were formed. It shows that the communities of each area
249 are related with no interference with the communities of the other location. However, in
250 the Pacific Ocean, two clusters were formed, which could explain that an occurrent factor
251 influences the diversity of the microbes. The outgroup in the Pacific Ocean is the only
252 sample with a noticeable amount of Archaea belonging to the class of Thermoplasmata
253 (0.6% of the reads). These Archaea were investigated by Gupta et al., 2021, and were
254 shown as acidophiles.

255

256 The ASVs distributions between the samples

257

258 The three different sampling domains shared 611 ASVs representing 5% of the total
259 reads. However, 380 ASVs representing 3%, were exclusively shared between Atlantic
260 plastic and Pacific plastic, 106 (1%) between Atlantic plastisphere and Pacific water, and
261 1598 (13%) between Pacific Plastisphere and Pacific water. Conversely, 4492 (38%) of
262 the ASVs were unique to the Pacific plastics, 3357 (29%) to the Atlantic plastics and 1308
263 (11%) to the Pacific water (Figure 2). A negligible proportion of ASVs is shared between
264 the two oceans, while each ocean showed a big proportion of unique ASVs, suggesting a
265 profound difference between their communities.

266

267 Microbial composition on the Atlantic plastics

268

269 From the analysis, the highest relative abundances were bacteria (99.91%). Three
270 bacterial phyla accounted for more than 90% of the relative abundance.
271 Verrucomicrobiota, Bdellovibrionota and Firmicutes accounted for more than 1% each,
272 while 29 other phyla (including bacterial, archaeal and eukaryal phyla) accounted for
273 4.70% of the community (each of these 29 phyla accounted for below 1% of the relative
274 abundance).

275

276 Among the abundant minor domains, Eukaryota (0.09%) were represented by the
277 phyla Amorphea (0.08%) and SAR (0.002%) and the classes of Obazoa and Alveolata.
278 Likewise, the reads of Archaea (0.0002%) were represented by the phylum of
279 Nanoarchaeota and the class of Nanoarchaeia.

279

280 Proteobacteria, Cyanobacteria and Bacteroidota were the three most abundant groups
281 at the phylum level (Figure 3). The occurrent communities include Alphaproteobacteria
282 (34.60%), reported as early colonizers; Bacteroidia (17.04%), reported as secondary
283 colonizers and Gammaproteobacteria (10.9%), later-stage colonizers at the class level,
284 according to a recent 16S rRNA gene amplicon data meta-analysis from 35 Plastisphere
285 studies, which revealed the successive colonization of the Plastisphere (Wright et al.,
286 2020). So, Gammaproteobacteria's presence suggests the maturity of the biofilm,
287 indicating that the plastics have been drifting for quite some time. Meanwhile, members
288 of the phylum Cyanobacteria; have been reported as abundant components of plastic

288 debris communities (Salta et al., 2013) highly represented on PP and PE items (Zettler et
289 al., 2013).

290 Other communities at the Family level include bacteria that prefer a surface-attached
291 lifestyle, such as Saprospiraceae (McIlroy & Nielsen, 2014), Hyphomonadaceae, known
292 to be biofilm formers (Abraham & Rohde, 2014) and Rhodobacteriaceae as opportunistic
293 colonizers (Dang & Lovell, 2016). At the genus level, Lewinella and Acinetobacter were
294 described as potential plastic degraders (see Table 3).

295

296 Microbial community composition on the Pacific plastics

297

298 After processing, 99.38% of the reads belonged to the domain of Bacteria. Three phyla
299 were most abundant, with almost 91% of the total read count. The other important relative
300 abundant phyla were classified as Planctomycetota, Actinobacteriota and
301 Verrucomicrobiota. They accounted for 6.23% of the total reads. Twenty-seven phyla
302 stemming from Bacteria, Archaea and Eukaryota accounted for 2.76% (each of the 27
303 recorded below 1% of the reads).

304 Among the small percentage reads, Archaea (0.62%) showed more diversity in the
305 Pacific than within the Atlantic and were represented by the phyla Thermoplasmatota
306 (0.62%), Nanoarchaeota (0.00058%) and Halobacterota (0.00008%). At the class level,
307 Archaea were represented by Thermoplasmata, Nanoarchaeia and Methanosarcinia.
308 Meanwhile, Eukaryota (0.00018%) displayed less diversity than within the Atlantic. They
309 were represented by one phylum, SAR and one class, Stramenopiles.

310 Proteobacteria, Cyanobacteria, Bacteroidota and Actinobacteria were the most
311 abundant groups at the phylum level (Figure 4). In addition to the three highest abundant
312 phyla reads, the Pacific Plastisphere recorded Actinobacteria (2.31%), which have been
313 reported as an abundant component of plastic debris communities (Salta et al., 2013; Pinto
314 et al., 2019). Herein, Cyanobacteria and Proteobacteria showed more diversity than in the
315 Atlantic. Gammaproteobacteria were also present, suggesting the maturity of the Pacific
316 biofilms. As such, the Pacific plastics have been drifting for quite some time. At the
317 family level, the figure shows the presence of Hyphomonadaceae and Rhodobacteraceae
318 but not Saprospiraceae as in the Atlantic Plastisphere. Instead, Flavobacteraceae, bacteria
319 that prefer surface-attached lifestyles, were present herein.

320

321 Microbial community composition in the Pacific water

322

323 Pacific water sample analysis was performed to compare microbial communities on
324 Pacific plastic and its surrounding water. Many studies showed that plastic communities
325 differ from surrounding water communities.

326 From the analysis, Bacteria were the most prominent domain, with 99.62%. Its phyla
327 Proteobacteria, Cyanobacteria and Bacteroidota accounted for more than 91% of the
328 relative abundances. Actinobacteria, Verrucomicrobiota, Planctomycetota and
329 Patescibacteria accounted for 7.11%. The rest (20 phyla), stemming from Bacteria,
330 Archaea and Eukaryota, accounted for 1.83% of the reads.

331 Archaea in water (0.37%) were represented by the phylum of Thermoplasmatota and the
332 class of Thermoplasmata. Meanwhile, Eukaryota (0.0019%) were represented by the
333 phylum of Amorphea and the class of Obazoa.

334 Among the occurring phyla between Pacific Plastisphere and Pacific water,
335 Dependitiae (0.005%), PAUC34f (0.002%), Nanoarchaeota (0.0004%, from Archaea),
336 SAR (0.0001%, from Eukaryota), Latescibacterota (0.0001%), Fibrobacterota (0.0001%)

337 and Halobacterota (0.00007%) were found only on Pacific plastic. Amorphea (0.0005%,
338 from Eukaryota) was found only in water. That could probably hint toward the specificity
339 of certain microorganisms for specific substrates.

340

341 Statistical analysis of the microbial community diversity composition of the samples

342

343 The statistical analysis of the samples showed a non-significant difference in microbial
344 community diversity within the Atlantic area based on plastic polymer types as well as
345 within the Pacific area. The p-values are greater than 0.05, as shown in Table 2. Indeed,
346 some studies showed that the plastic polymer types have no effect in determining the
347 Plastisphere community composition in mature biofilms (Oberbeckmann & Labrenz,
348 2020). So, these results confirm the maturity of the biofilms in the Atlantic and Pacific
349 Plastisphere. Also, the p-values displayed (see Table 2) while assessing the diversity
350 between the Pacific plastics and its surrounding water showed no significance. The
351 Pacific Plastisphere was not significantly more or less diverse than the microbial
352 community in the Pacific water. Indeed, Oberbeckmann et al. (2014) suggested that
353 communities at early times in the colonization process are more likely to reveal polymer-
354 specificity, while communities that establish on different polymers should gradually
355 converge over time as the biofilms mature (Harrison et al., 2014).

356 Meanwhile, the diversity assessment of the Atlantic and Pacific Plastisphere showed
357 significant p-values for phylogenetic measures and beta diversity (see Table 2). So, the
358 communities in the Atlantic Plastisphere are significantly distinct from those in the
359 Pacific Plastisphere. It confirms the results obtained by Amaral-Zettler et al. seven years
360 ago on the same topic when assessing the diversity between Atlantic and Pacific
361 communities. They found the same significance level (p-value = 0.001); distinct grouping
362 based on the oceanic biogeographic zone (Atlantic versus Pacific). Biogeography is
363 incontestably a driver of microbial diversity. Similar results were also obtained by Coons
364 et al., 2021 who found that biogeography influences Plastisphere community structure
365 more than substrate type. Differences in the biofilm community composition are related
366 to different factors.

367 Some previous studies have targeted temperature as the best predictor of bacterial
368 diversity in surface waters (Ibarbalz et al., 2019). Regarding this study, the plastic
369 particles were collected at the surface of different waters. They could have attracted
370 microbial communities able to evolve at the various water surfaces.

371 Other studies showed that the substratum physicochemical properties
372 (hydrophobicity, roughness, vulnerability to weather) and the surface chemodynamics
373 (surface conditioning or nutrient enrichment) play a role in microbial diversity (Dang and
374 Lovell, 2016). Besides physicochemical surface properties, it has been shown that the
375 composition of biofilm communities associated with synthetic polymers differed
376 significantly for different ocean basins (Amaral-Zettler et al., 2015) and underlay both
377 seasonal and spatial effects, e.g., in North Sea waters (Oberbeckmann et al., 2014). The
378 waters from the Atlantic and Pacific Oceans seem to have different physicochemical
379 properties, which could have impacted the properties of the plastics we collected,
380 especially since they lasted in the water.

381 Future studies on the same topic should include environmental parameters to determine
382 the likely drivers of this difference in microbial diversity composition between the
383 Atlantic and Pacific. So, the pH (as it varies between the Atlantic and the Pacific), the
384 dissolved oxygen, the salinity or the surface temperature (as it also varies between both
385 oceans) could be responsible for this difference in microbial diversity between the
386 Atlantic and Pacific oceans.

387

388 Differentially abundant classes between the Atlantic and Pacific Plastisphere

389

390 The above statistics showed that there is effectively a significant difference between the
391 Atlantic and the Pacific microbial community diversity. Linear discriminant analysis
392 (LDA) effect size (LEfSe) was used to predict the class level abundances between the
393 Atlantic and the Pacific for their different communities. It revealed 32 differentially
394 abundant classes (LDA log score $> \pm 2$) between the Atlantic and the Pacific, as displayed
395 in Figure 5. The dominant classes that made the difference between the Atlantic and the
396 Pacific (see Figure 5) belong to the phyla Proteobacteria, Bacteroidota, Planctomycetota,
397 Bdellovibrionota, Bacilli, Verrucomicrobiota and Thermoplasmata (from Archaea).
398 SAR and Amorphea (from Eukaryota) were also part of the differentially abundant
399 microorganisms.

400 The Atlantic shows 23 less abundant classes, among which Alveolata and Obazoa are
401 from Eukaryota. In comparison, the Pacific offers nine more abundant classes, among
402 which is the class Thermoplasmata from Archaea. Among these 32 classes, 12 had an
403 LDA score $> \pm 3$, including eight from the Atlantic (in ascending order Desulfuromonadia,
404 TK17, Verrucomicrobiae, Anaerolineae, Bacilli, Bdellovibrionia, Gammaproteobacteria)
405 and four from the Pacific (in ascending order Parcubacteria, Thermoplasmata,
406 Planctomycetes, and Alphaproteobacteria). Alphaproteobacteria, Gammaproteobacteria,
407 Bacteroidia had an LDA score $> \pm 4$. Thermoplasmata, ABY1 and Desulfovibrionia were
408 unique to the Pacific, while Obazoa, endosymbiont_of_Ridgeia_piscesae,
409 Vicinamibacteria, Alveolata and TK17 were unique to the Atlantic.

410

411 Potential plastic degraders within the studies Plastisphere

412

413 The plastic-degrading potential of the Plastisphere community is an ongoing topic (Zettler
414 et al., 2013). Exploring the present Plastisphere, 40 genera previously described to include
415 hydrocarbon-degrading bacteria, as shown in Table 3, were deciphered. These genera
416 represented 4.07% of the relative abundances of the whole Plastisphere and were shared
417 in 4 phyla, five classes, 21 orders and 32 families. Proteobacteria was the most
418 represented, with 22 genera. Actinobacteria came after that with eight genera,
419 Bacteroidota with seven genera and Firmicutes with three genera. Twelve genera were
420 exclusively detected in the Atlantic and three in the Pacific, while 25 were shared between
421 the two oceans.

422 Our samples were composed of PP and PE. The distribution of PE-degrading
423 microorganisms seems limited, although PP appears to be non-biodegradable. However,
424 it was reported that *Acinetobacter* sp.351 partially degraded lower molecular weight PE
425 oligomers (the genus was found herein: 1.11%) upon dispersion. In contrast, high
426 molecular weight PE could not be impaired (Tsuchii et al., 1980). The biodegradability
427 of PE could be improved by blending it with biodegradable additives, photoinitiators or
428 copolymerization (Griffin, 2007; Hakkarainen & Albertsson, 2004). A blending of PE
429 with additives generally enhances auto-oxidation, reduces the molecular weight of the
430 polymer, and then makes it easier for microorganisms to degrade the low molecular
431 weight materials.

432 Meanwhile, the possibility of degrading PP with microorganisms has been
433 investigated (Cacciari et al. 1993). In that study, it was shown that aerobic and anaerobic
434 species with different catabolic capabilities could act in close cooperation to degrade
435 polypropylene films. Some *Pseudomonas* (present in this Plastisphere) species were
436 pointed out in the process of polypropylene degradation. Besides that, many species of

437 *Pseudomonas* were indicated to degrade Polyethylene (Zheng et al., 2005), Polyvinyl
438 chloride (Danko et al., 2004), while *Rhodococcus* was shown to degrade Polyethylene
439 (Sivan et al., 2006).

440 Microbial communities associated with plastic degradation composition and species
441 richness are influenced by spatiotemporal phenomena like habitats/geographical location,
442 ecosystem, and seasonal variation (Kirstein et al., 2019; Pinto et al., 2019). Further, the
443 physiochemical nature of plastics like polyethylene, polypropylene, polystyrene, also
444 regulates this degradation (Pinnell & Turner, 2019). The composition and specificity of
445 microbial assemblage associated with polyethylene (PE) and polystyrene (PS) in the
446 marine aquatic ecosystem (coastal Baltic Sea) are indicated by an abundance of
447 Flavobacteriaceae (*Flavobacterium*), Rhodobacteraceae (*Rhodobactor*),
448 Methylophilaceae (*Methylotenera*), Planctomycetaceae (*Planctomyces*, *Pirellula*),
449 Hyphomonadaceae (*Hyphomonas*), Planctomycetaceae (*Blastopirellula*),
450 Erythrobacteraceae (*Erythrobacter*), Sphingomonadaceae (*Sphingopyxis*), etc.
451 (Oberbeckmann et al., 2018). Kirstein et al., 2019 found that the microbial community
452 composition associated with various plastics is significantly varying, and it is also
453 changing with the different phases of the plastic degradation process. In our study, the
454 genera, *Flavobacterium* (0.05%), *Hyphomonas* (0.01%) and *Erythrobacter* (0.29%) were
455 precisely found to be associated with PE (0.27%), but also PP (0.2%).

456
457 Biotechnologies are leading scientists toward an alternative to mechanical or chemical
458 degradation of plastic waste for a more sustainable end life of plastics. Finding a natural
459 solution to the man-made plastic problem is an urgent issue, but the work keeps going.
460 Once a candidate has revealed plastic degrading potential, the responsible gene must be
461 uncovered and cloned into a host organism (e.g., *Bacillus subtilis*; Austin et al., 2018) to
462 optimize the secreted protein (Wang et al., 2020). Thus, an efficient natural solution
463 against plastic pollution can be found.

464

465 **CONFLICT OF INTEREST**

466

467 The authors have not declared any conflict of interest

468

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470

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- 834
- 835 **TABLES**

836

837 **Table 1.** Taxonomic rank abundance distribution per sample type and percental display
838 of unclassified ASVs per taxonomic rank.

Taxonomic level	Atlantic plastic		Pacific plastic		Water	
	Classified	%Unclassified	Classified	%Unclassified	Classified	%Unclassified
Kingdom	3	0	3	0	3	0
Phylum	35	0.51	33	0.12	27	0.13
Class	74	0.57	74	0.26	55	0.21
Order	161	4.19	172	2.30	142	2.91
Family	241	5.75	252	5.35	206	5.67
Genus	369	34.09	400	35.33	323	38.56

839

840 **Table 2.** Statistical analysis of the samples: table displaying p-values from Kruskal-
841 Wallis and PERMANOVA analysis.

		Atlantic Plastisphere	Pacific Plastisphere	Pacific plastics & water	Atlantic & Pacific plastics
Alpha diversity	Non-phylogenetic measure				
	Considering the polymer types	0.8	0.6	0.96	0.82
	Regardless of the polymer types			0.88	0.38
	Phylogenetic measure				
	Considering the polymer types	0.67	0.57	0.4	0.002
	Regardless of the polymer types			0.23	0.000063
Beta diversity	Considering the polymer types	0.35	0.84	0.8	0.001
	Regardless of the polymer types			0.59	0.001

842

843 **Table 3.** Genera of potential plastic degraders within the studied Plastispheres. Genera in
844 Bold are those detected only in one area, relative abundances are indicated in each ocean
845 and on each plastic type, relative abundances below 0.01 are indicated as <0.01.

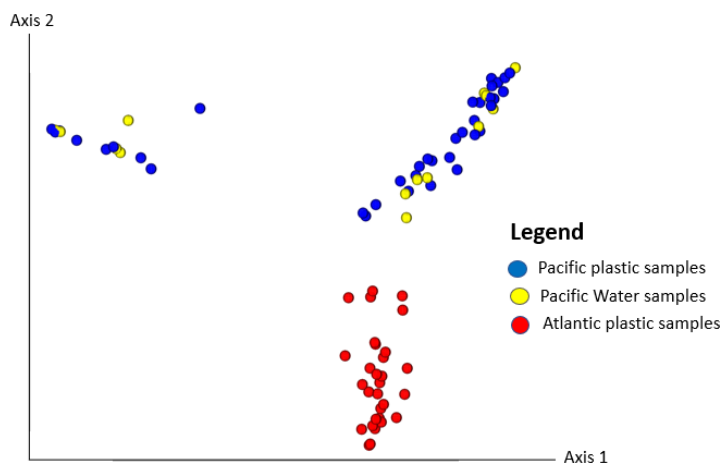
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Genus&Reference	Atlantic	Pacific	PP	PE	Total
<i>Lewinella</i> (Vaksmas et al., 2021)	0.73	0.42	0.19	0.97	1.16
<i>Acinetobacter</i> (Chaîneau et al., 1999)	1.10	0.01	0.41	0.69	1.11
<i>Erythrobacter</i> (Harwati et al. 2007)	0.08	0.29	0.18	0.18	0.37
<i>Algimonas</i> (Vaksmas et al., 2021)	0.12	0.14	0.06	0.2	0.26
<i>Vibrio</i> (Hedlund and Staley, 2001)	0.18	0.032	0.03	0.18	0.21
<i>Winogradskyella</i> (Wang et al. 2014)	0.03	0.16	0.07	0.13	0.19
<i>Tenacibaculum</i> (Wang et al. 2014)	0.07	0.08	0.02	0.12	0.14
<i>Alteromonas</i> (Iwabuchi et al., 2002)	0.09	0.04	0.02	0.10	0.12
<i>Brevundimonas</i> (Chaîneau et al., 1999)	0.1	0.002	0.05	0.05	0.1
<i>Roseovarius</i> (Peeb et al., 2022)	0.007	0.08	0.03	0.06	0.09
<i>Pseudomonas</i> (Le Petit et al., 1975)	0.06	<0.01	0.04	0.03	0.07
<i>Hyphomonas</i> (Yakimov et al., 2005)	0.04	0.01	<0.01	0.04	0.05
<i>Flavobacterium</i> (Stucki and Alexander, 1987)	0.05	<0.01	0.01	0.04	0.05

<i>Fabibacter</i> (Wang et al. 2014)	0.02	<0.01	<0.01	0.02	0.03
<i>Dokdonia</i> (González et al., 2011)	0.02	<0.01	<0.01	0.014	0.02
<i>Stenotrophomonas</i> (Juhász et al., 2000)	0.02	-	0.01	<0.01	0.02
<i>Marinobacter</i> (Gauthier et al., 1992)	<0.01	0.01	<0.01	0.01	0.01
<i>Halomonas</i> (Wang et al., 2007)	<0.01	<0.01	<0.01	0.01	0.01
<i>Oleiphilus</i> (Golyshin et al., 2002)	0.01	-	-	0.01	0.01
<i>Methylobacterium-Methylorubrum</i> (Bodour et al., 2003)	<0.01	<0.01	<0.01	<0.01	0.01
<i>Staphylococcus</i> (Saadoun et al., 1999)	<0.01	<0.01	<0.01	-	<0.01
<i>Hyphomicrobium</i> (Ozaki et al., 2006)	-	<0.01	<0.01	<0.01	<0.01
<i>Corynebacterium</i> (Chaineau et al., 1999)	<0.01	<0.01	<0.01	<0.01	<0.01
<i>Pseudoxanthomonas</i> (Yue et al., 2021)	<0.01	-	<0.01	<0.01	<0.01
<i>Chryseobacterium</i> (Szoboszlay et al., 2008)	<0.01	<0.01	<0.01	<0.01	<0.01
<i>Thalassospira</i> (Kodama et al., 2008)	-	<0.01	<0.01	<0.01	<0.01
<i>Alkanindiges</i> (Bogan et al., 2003)	<0.01	-	<0.01	<0.01	<0.01
<i>Alcanivorax</i> (Yakimov et al., 1998)	<0.01	<0.01	-	<0.01	<0.01
<i>Micrococcus</i> (Ilori et al., 2000)	<0.01	<0.01	<0.01	<0.01	<0.01
<i>Kocuria</i> (Dashti et al., 2009)	<0.01	-	-	<0.01	<0.01
<i>Rhodococcus</i> (Meyer et al., 1999)	<0.01	-	<0.01	<0.01	<0.01
<i>Methylophaga</i> (Mishamandani et al. 2014)	-	<0.01	-	-	<0.01
<i>Oleispira</i> (Yakimov et al., 2003)	<0.01	<0.01	-	<0.01	<0.01
<i>Mycobacterium</i> (Willumsen et al., 2001)	<0.01	<0.01	<0.01	<0.01	<0.01
<i>Nocardioides</i> (Hamamura and Arp, 2000)	<0.01	-	<0.01	<0.01	<0.01
<i>Arthrobacter</i> (Le Petit et al., 1975)	<0.01	-	-	<0.01	<0.01
<i>Actinomyces</i> (ZoBell, 1946)	<0.01	-	<0.01	<0.01	<0.01
<i>Achromobacter</i> (Le Petit et al., 1975)	<0.01	-	-	<0.01	<0.01
<i>Lactobacillus</i> (Floodgate, 1984)	<0.01	-	-	<0.01	<0.01
<i>Bacillus</i> (Li et al., 2008)	<0.01	-	-	<0.01	<0.01

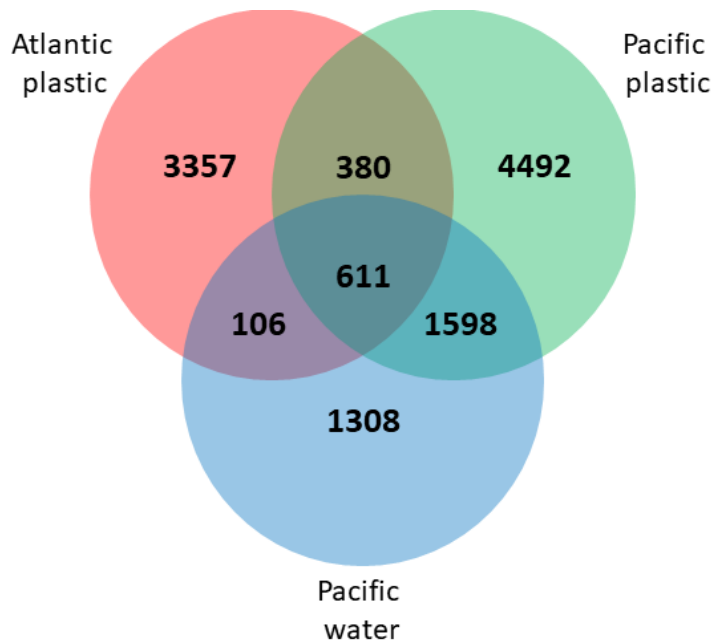
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848 FIGURES

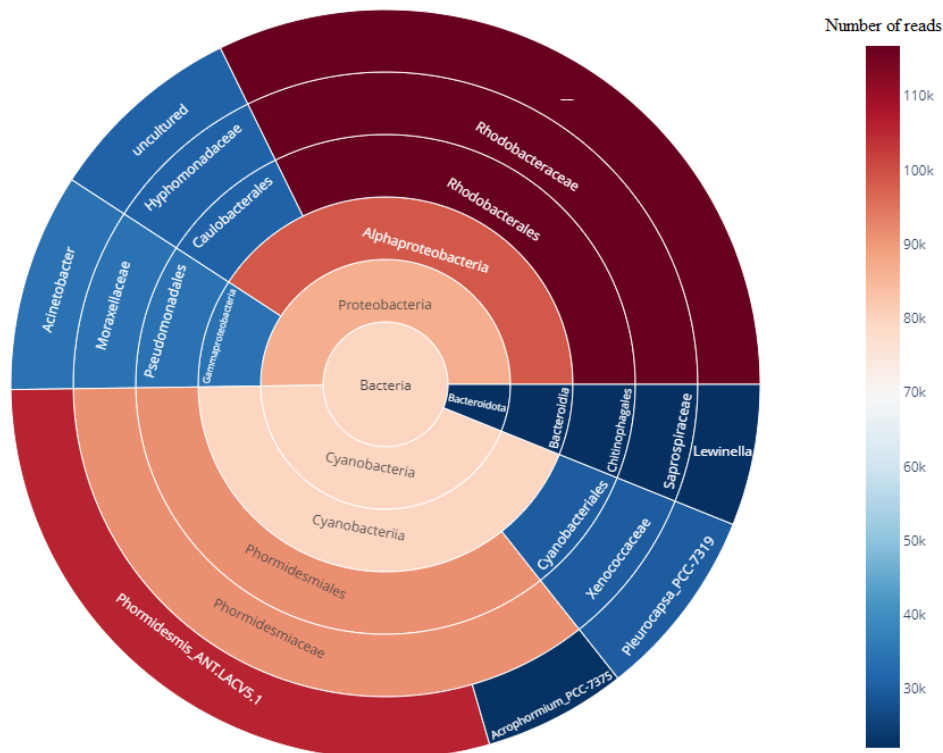


849

850 **Figure 1.** Principal coordinates analysis (PCoA) of all samples showing a specific
851 clustering pattern of the associated microbial communities from the different oceans and
852 samples. The phylogenetic distances calculated within the dataset indicate three clusters
853 showing their level of relatedness. The plot was generated in QIIME2.

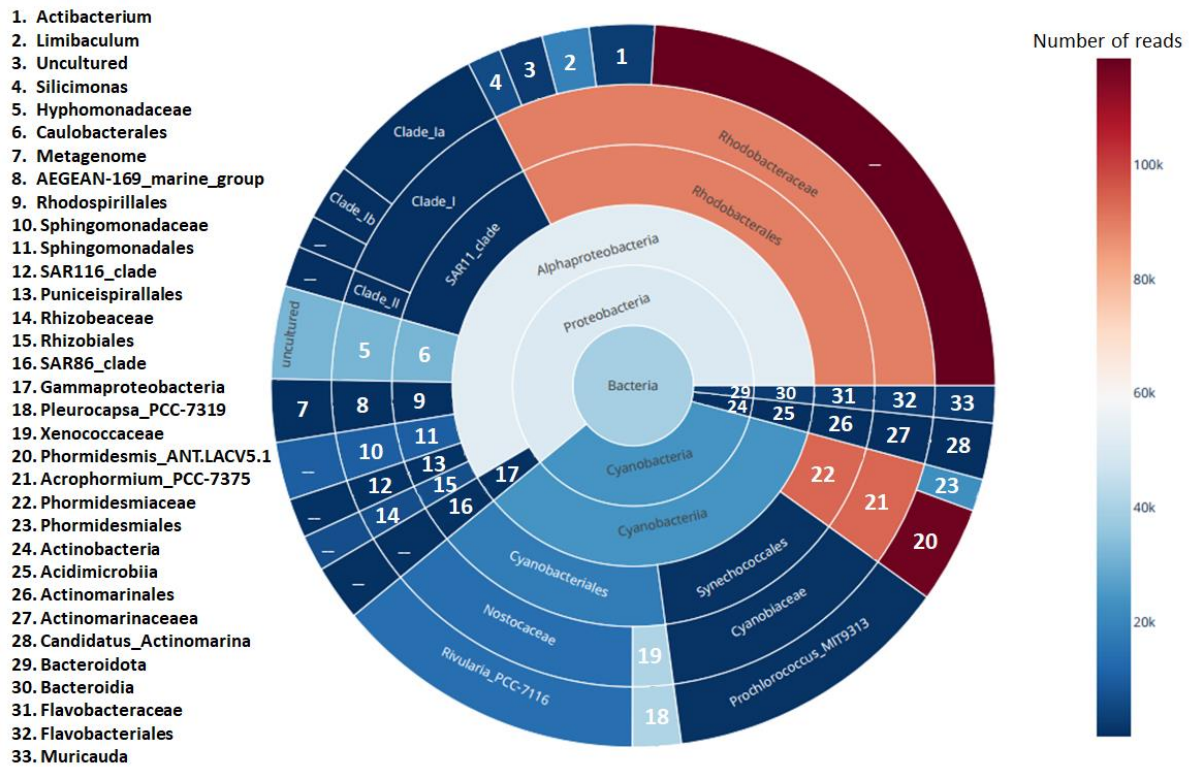


854 **Figure 2.** Distribution of ASVs between the different sample types. Unique and shared
855 ASVs between Atlantic plastic, Pacific plastic, and Pacific water samples. The plot was
856 made in R using the 'Venn diagram' package.
857

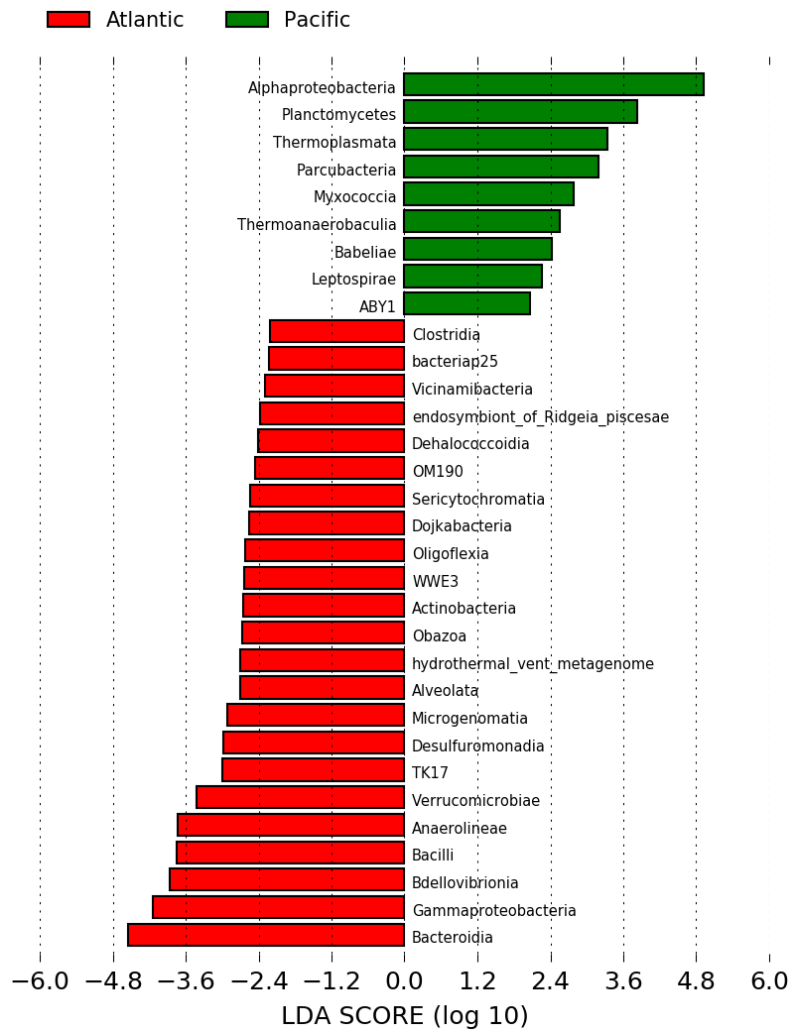


858 **Figure 3.** Reads and taxonomic affiliation of recurring communities on the Atlantic
859 Plastisphere. Sunburst chart displaying the affiliations of genera that reached values
860

861 above 20,000 reads. Each plot crown represents one taxonomic level from the Kingdom
 862 to the genus. The plot was implemented using 'plotly.express' package in Python.



863
 864 **Figure 4.** Reads and taxonomic affiliation of recurring communities on the Pacific
 865 Plastisphere. Sunburst chart displaying the affiliations of genera that reached values
 866 above 20,000 reads. Each plot crown represents one taxonomic level from the Kingdom
 867 to the genus. The plot was implemented using 'plotly.express' package in Python.



868

869

870 **Figure 5.** Differentially abundant classes between the Atlantic and the Pacific oceans.

871 Linear discriminant analysis (LDA) effect size (LEfSe) results per ocean. Bar plots depict

872 all classes which had an LDA log score $> \pm 2$ between all plastic samples ($N = 68$) in the

873 Atlantic ($n=30$) or Pacific ($n=38$) oceans. The plot was made by utilizing galaxy online

874 (<https://huttenhower.sph.harvard.edu/galaxy/>). Grouped data were first analyzed using

875 the Kruskal-Wallis test with a significance threshold of 0.05 to determine if the data was

876 differentially distributed between groups.

877

Supplementary material

878

879

Supplementary table S1: Samples used in this study with their respective geographic

880

location, plastic type and shannon diversity indice indicated.

881