1	Species boundaries to the limit: validating species delimitation methods is
2	critical to avoid taxonomic inflation in the genomic era
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16 Abstract

17 With the advent of molecular phylogenetics, the number of taxonomic studies unveiling and 18 describing cryptic diversity has greatly increased. However, speciation between cryptic 19 lineages is often defined without evaluating population structure or gene flow, which can lead 20 to false claims of species status and, subsequently, taxonomic inflation. In the present study 21 we focus on the intriguing case of the Arabian gecko *Trachydactylus hajarensis* (Squamata: 22 Gekkonidae), a species for which cryptic diversity has been previously reported. We 23 generated mitochondrial data (12S rDNA) and genome-wide SNP data (ddRADseq) for 52 24 specimens to determine phylogenomic relationships, population structure and gene flow 25 within this species. Then, we applied species delimitation methods (SDMs) to evaluate 26 several competing species hypotheses through the Multispecies Coalescent model. Results 27 show that T. hajarensis is comprised by three well-defined lineages, two of them in the Hajar 28 Mountains of eastern Arabia, and one in Masirah Island, in the southeastern coast of Oman. 29 Even though high levels of past introgression and strong mitonuclear discordances were 30 found, current gene flow is scarce with clear boundaries between populations and shallow 31 levels of admixture in the contact zone between lineages. Surprisingly, species tree topology 32 differed between methods and when different individuals were used in downsampled datasets. 33 Conventional SDMs supported up to three putative new species within the group. However, 34 after species validation with the genealogical divergence index (gdi), none of the putative 35 species held. Overall, this study highlights the importance of sample choice, integrative 36 analyses, and validation methods to not incur into taxonomic inflation, providing a set of 37 already available tools to assess and validate population structure, gene flow, and SDMs 38 before describing new species.

39

40 **1. Introduction**

41 Species are one of the fundamental units in biology. However, with more than 24 different 42 definitions, the species concept is still a topic of debate, and sometimes dispute, in the 43 scientific community (Mayden, 1997). Many authors agree that species are separately 44 evolving metapopulations emerging from a speciation process (Fišer, Robinson, & Malard, 45 2018) but, since speciation is intrinsically a gradual mechanism, discriminating populations 46 from species remains a challenge. The inference of speciation events is achieved by attending 47 to different lines of evidence such as reproductive isolation, morphological differentiation or 48 monophyly (De Queiroz, 2007; Fišer et al., 2018). However, the advent of molecular 49 phylogenetics and genomics has presented a new paradigm, reshaping the tree of life as we 50 knew it. We have seen that phenotypic variation or the absence of gene flow alone do not 51 always suffice to delimit species and, in many cases, species present signals of ancestral or 52 current hybridization (Ivanov, Lee, & Mutanen, 2018). Moreover, the increase of molecular 53 species delimitation studies in the last decade has revealed that cryptic diversity is much more 54 common than previously thought (Chattopadhyay et al., 2016; Vences et al., 2022; Vilaca et 55 al., 2021).

56 Cryptic species are entities that have undergone a speciation process but remain 57 morphologically identical (Chan et al., 2020). Owing to the generalized use of molecular 58 methods in most systematic studies, it has been shown that cryptic species are widely spread across most animal phyla (Derycke et al., 2008; Fennessy et al., 2016; Riaño et al., 2022; 59 Vences et al., 2022). Some of the mechanisms that can lead to a lack of phenotypical variation 60 61 between divergent lineages include early divergence or niche conservatism (Fišer et al., 62 2018), among others. Moreover, the description of cryptic diversity is shortening the Linnean 63 shortfall and is of paramount importance for conservation efforts (Walters, Cannizzaro,

64 Trujillo, & Berg, 2021), since many species that were once thought to be homogeneous and 65 widely distributed, actually represent complexes of cryptic species with sometimes 66 endangered micro-endemic entities (Garcia-Porta, Simó-Riudalbas, Robinson, & Carranza, 2017). However, the lack of a specific threshold to distinguish separate evolutionary lineages 67 68 from populations of the same species raises another challenging question: where do we stop? 69 The advent of next generation sequencing (NGS) and the development of new species 70 delimitation methods (SDMs) have led to an outburst of taxonomic studies, especially in the 71 case of non-model organisms (Ivanov et al., 2018). The amount of data that can be obtained 72 through these techniques has massively surpassed the multi-locus approach and, in several 73 cases, signals of population structure can be interpreted as species status with conventional 74 SDMs, even when there is gene flow (Sukumaran & Knowles, 2017). This might result in 75 over-splitting and taxonomic inflation, an issue especially problematic for conservation 76 management, where limited resources have to be prioritized to the most endangered species 77 (Isaac, Mallet, & Mace, 2004).

78 Nowadays, genomic SDMs are at the forefront of an ongoing debate in the scientific 79 community (Jackson et al., 2017; Leaché et al., 2019; Sukumaran & Knowles, 2017). Within 80 those, the Multispecies Coalescent (MSC) model (Rannala & Yang, 2003), extensively used 81 with genomic data from closely related species, has been shown to capture population splits 82 rather than species divergences (Leaché et al., 2019). This is especially problematic in 83 allopatric populations where a lack of gene flow does not confirm pre- or postzygotic barriers 84 but rather results from geographic isolation. To account for this, Jackson et al. (2017) 85 proposed a heuristic criterion for species delimitation based on a genealogical divergence 86 index (gdi) which afterwards was implemented in a Bayesian parameter estimation under the MSC model (Leaché et al., 2019). With this index, two lineages can be identified either as 87

88	two distinct species or as a single one, but it also includes a range of indecision that reflects
89	the arbitrary nature of the species definition (Leaché et al., 2019).
90	In the past decades, molecular phylogenetics and the application of SDMs in
91	traditionally neglected arid regions have proven that in such areas there are still high levels of
92	undescribed diversity and many examples of cryptic species (Bray, Alagaili, & Bennett, 2014;
93	Carranza et al. 2016; Main, van Vuuren, Tilbury, & Tolley, 2022; Simó-Riudalbas et al.,
94	2017). Within the Arabian Peninsula, the Hajar Mountains rise as one of its most biodiverse
95	regions, with high levels of reptile diversity and endemicity (Burriel-Carranza, Els, &
96	Carranza, 2022; Carranza, Els, & Burriel-Carranza, 2021; Šmíd et al., 2021). Its high and
97	complex topography and the relatively low annual mean temperatures offer a spectrum of
98	diverse niches which have already been the scenery of the origin of cryptic diversity (Garcia-
99	Porta et al., 2017; Simó-Riudalbas et al., 2017, 2018; Tamar, Mitsi, & Carranza, 2019).
100	Therefore, it is a well-suited ecological system to explore the nuances of the MSC model
101	together with the new heuristic gdi with allopatric and early divergent species.
102	The ground-dwelling Arabian gecko Trachydactylus hajarensis (Arnold, 1980) is a
103	species endemic to the Hajar Mountains and to Masirah Island, a small island situated 20 km
104	off the eastern coast of Oman, but almost 200 km away from the closest continental T.
105	hajarensis (Figure 1). Even though previous data suggest a natural colonization of the island
106	instead of a human-mediated translocation (de Pous et al., 2016), the provenance of this
107	island's population still remains unclear. Previous studies on this South Arabian endemic
108	genus of only two species were carried out with multi-locus assemblies and most of the
109	analyses were based on mitochondrial sequences alone (de Pous et al., 2016). Based on those
110	results, T. hajarensis was suggested to be a species complex with at least three allopatric
111	lineages: one in the Western and Central Hajars, a second in the Central and Eastern Hajars,

112	and another one in the easternmost side of the Hajar Mountains and in Masirah Island (de
113	Pous et al., 2016). However, mitochondrial evolutionary history is not always linked to its
114	nuclear counterpart, with common cases of mito-nuclear discordances across phyla (Marshall,
115	Chambers, Matz, & Hillis, 2021; Shults et al., 2022). This stresses the need of revisiting this
116	group's systematics with NGS techniques, which recover large portions or the complete
117	nuclear genome and are key to determine whether there are mito-nuclear discordances,
118	providing a comprehensive perspective of the evolutionary history of a species.
119	In the present work, we generate <i>de novo</i> genomic reduced representations of <i>T</i> .
120	hajarensis to revisit the systematics of the group. We reconstruct the nuclear and
121	mitochondrial phylogeny of T. hajarensis together with its sister species T. spatalurus,
122	generate and test a series of species hypotheses applying the heuristic gdi and delimit the
123	population structure and evolutionary history of this species, with a special focus on the
124	colonization of Masirah Island.

126 2. Materials & Methods

127 **2.1. Sampling**

128 A total of 52 individuals of the genus *Trachydactylus* from 33 different localitites were

included in this study (Figure 1; Table S1). Specimens were collected between 2005 and 2017

along all the known distribution range of *T. hajarensis*, containing representatives of all the

131 Hajar Mountains' and Masirah Island's lineages. We also included four specimens of *T*.

132 *spatalurus* from Dhofar, South Oman, that were used as outgroup (Table S1).

133

134 **2.2. Mitochondrial analyses**

135 We sequenced 35 specimens for the 12S rDNA (12S) and 15 additional samples were 136 downloaded from GenBank. DNA extraction was done following the protocol in MacManes 137 (2013) and PCR amplification conditions and primers used were the same as described in 138 Metallinou et al. (2015). PCR products were purified and Sanger sequenced by Macrogen Inc. 139 to obtain a 523 bp fragment of the mitochondrial gene 12S. Sequences were aligned with 140 Geneious 2021.1.1 (Biomatters Ltd.) and a Bayesian Inference (BI) phylogeny was reconstructed in BEAST2 v.2.6.4 (Bouckaert et al., 2019). We calibrated the deepest node in 141 142 our phylogeny (the split between T. hajarensis and T. spatalurus) by extracting the mean 143 height from a recently published squamate phylogeny (Tejero-Cicuéndez et al., 2022) and 144 applying a normal distribution encompassing the 95% HPD intervals. We selected a 145 HKY+G+X model with four gamma categories, base frequencies were estimated, and a 146 relaxed clock LogNormal was used with a Calibrated Yule process tree prior. We conducted three independent runs of 10⁸ generations sampling every 10,000 generations. Convergence 147 148 was checked with Tracer v.1.5 (Rambaut & Drummond, 2013), a 40% burnin was applied and 149 trees were summarized with TreeAnnotator v.2.6.4 (Bouckaert et al., 2019). Then, we objectively identified and delimited deep mitochondrial lineages using the general mixed 150 151 Yule-coalescent model (GMYC; Fujisawa & Barraclough, 2013; Pons et al., 2006) 152 implemented in the R package 'splits' (Ezard, Fujisawa, & Barraclough, 2021). 153

154 **2.3.** Genomic DNA sequencing and processing

Genomic libraries were produced following Peterson's et al. (2012) protocol for double-digest restriction site-associated DNA (ddRADseq) for 52 specimens. In short, we double-digested 500 ng of genomic DNA, using a pair of rare and common restriction enzymes (Sbf1 and Msp1, respectively). The resulting fragments were ligated with barcoded Illumina adapters.

159 Fragments were then size-selected for a range between 415 – 515 bp and sequenced on an
160 Illumina NextSeq 500, for 75 bp single-end reads.

161 Raw Illumina reads were processed using Ipyrad v.0.9.62 (Eaton & Overcast, 2020) 162 discarding sites with Phred score < 33, reads with more than three missing sites, consensus 163 sequences with less than six reads, excessive heterozygous sites (more than three), or more 164 than two haplotypes. After testing several configurations, both filtered reads and consensus 165 sequences were clustered and aligned using an 89% clustering threshold. We set the minimum 166 number of samples per locus to four to retrieve the maximum number of loci possible for 167 post-processing filtering. Demultiplexed filtered reads for each individual can be found at 168 Dryad (will be added in the published version). 169 Following recommendations from O'Leary et al. (2018), we applied an iterative 170 filtering to identify and remove low quality samples and loci. We used Radiator (Gosselin, 171 Lamothe, Devloo-Delva, & Grewe, 2017), Plink2 (Chang et al., 2015) and VCFr (Knaus & 172 Grünwald, 2017) implemented in a custom script 173 (https://github.com/BernatBurriel/Post processing filtering) to filter iteratively and 174 alternatively SNP datasets. Values of missing data allowance ranged from 98% to 78% of

175 missing genotype call rate and missing data per individual, decreasing 2% between iterations.

176 Furthermore, we applied a hard threshold of missing genotype call rate depending on the

177 dataset (Table S2), we removed non-biallelic SNPs, applied a minor allele frequency filter

178 (maf < 0.05) and removed monomorphic sites.

Dataset types varied between analyses but can be summarized into *loci* and *SNP* datasets: *loci* datasets were generated with ipyrad after removing all individuals that did not pass the previously explained filters and retaining only loci that were at least in 60% of all specimens. *SNP* datasets contained either concatenated SNPs or putatively unlinked SNPs. In

- 183 the latter only the SNP with the highest read depth of each locus was chosen. For further
- 184 specifications in each dataset, refer to Table S2.
- 185

186 **2.4. Population structure**

- 187 We used *dataset 1* (Table S2) to inferred the population ancestry of each individual with
- ADMIXTURE v.1.3.0 (Alexander & Lange, 2011; Alexander, Novembre, & Lange, 2009).
- 189 Ancestral populations ranged from K=1 to K=10, with 15 replicates for each K, and the best
- 190 K value inferred after 15 cross-validation rounds. Then, we used the same dataset to perform
- 191 a Principal Component Analysis (PCA) with Plink v2.00a2.3 (Chang et al., 2015). Finally, we
- used *dataset 2* (Table S2) for fineRADstructure (Malinsky, Trucchi, Lawson, & Falush, 2018)
- 193 to further analyze *T. hajarensis*' population structure. This analysis unravels different levels
- 194 of structure within and between populations and its robustness to missing data is optimal for
- 195 non-model organisms (Malinsky, Trucchi, Lawson, & Falush, 2018). Results from all the
- 196 former analyses were visualized with R v.4.2.1 (R Core Team, 2021).
- 197

198 2.5. Phylogenomic reconstructions

199 We conducted the phylogenomic anlayses using ML and BI on *dataset 3* (Table S2),

200 including 5,219 loci and 47 individuals. With a concatenated dataset of all loci, we generated

- 201 ML reconstructions with RAxML-ng v.1.0.2 (Kozlov et al., 2019) with a GTR+G model, a
- total of 100 starting trees (50 random and 50 parsimony) and 1,000 bootstrap replicates to
- 203 estimate branch support. We also generated individual gene trees for each locus with IQ-
- 204 TREE (Nguyen, Schmidt, von Haeseler, & Minh, 2015) with the best model obtained from
- 205 ModelFinder (TPM3+F+R2), 100 trees and 1,000 ultrafast bootstraps. Then we generated a

206 consensus tree by summarizing all trees with Astral v.5.7.8 (Zhang, Sayyari, & Mirarab, 207 2017).

208 We also estimated a time calibrated tree with BI implemented in BEAST2 v.2.6.4 209 (Bouckaert et al., 2019) applying the same priors and specifications as in the mitochondrial BI 210 approach (see methods section 2.2). In addition, we reconstructed with *dataset 3* the 211 relationships of *T. hajarensis* using unrooted phylogenetic networks implemented in 212 SplitsTree v.4.18.3 (Huson & Bryant, 2006) with the Neighbor-Net algorithm. 213

214 2.6. Coalescent-based Species Trees

215 Based on the population structure (Figure 2) and phylogenomic reconstruction (Figure 3) 216 analyses, we identified up to four well-defined monophyletic groups within T. hajarensis. 217 These groups can be geographically divided into Western Hajars lineage, Central Hajars 218 lineage, Eastern haiars lineage and Masirah Island lineage (Figure 1). We estimated a species 219 tree of the aforementioned lineages together with *T. spatalurus* to evaluate the evolutionary 220 relationships between these groups in the Multispecies Coalescent framework. The resulting 221 tree was then used as a guide tree for species delimitation methods (see methods section 2.8). 222 Since species delimitation and species tree inference tend to be computationally demanding, it 223 is common practice to downsample datasets to one or two specimens per lineage (e.g. Tonzo, 224 Papadopoulou, & Ortego, 2019). To test the effect of taxon choice when producing reduced 225 datasets, we generated four datasets with different specimen configurations that did not 226 present any signals of admixture between lineages (see Table S2 for further details); T. 227 spatalurus was used to root the phylogenetic trees: i) dataset 4 SNAPP/BPP: two specimens 228 with the highest coverage from Western, Central, Eastern, and Masirah Island lineages and 229 two specimens of T. spatalurus; ii) dataset 5 SNAPP/BPP: two specimens from Western,

230 Central, Eastern, and Masirah Island lineages and two specimens of T. spatalurus; Eastern 231 lineage specimens selected from the closest geographic region to the Masirah Island lineage 232 (CN10775-26 and CN10791-26); iii) dataset 6 SNAPP/BPP: two specimens from Western, 233 Central, Eastern, and Masirah Island lineages and two specimens of *T. spatalurus*; Eastern 234 lineage specimens selected from the farthest geographic region to the Masirah Island lineage 235 (S7150-22 and CN686-23); vi) dataset 7 SNAPP/BPP: four specimens from Western, Central, 236 Eastern, and Masirah Island lineages and four specimens of *T. spatalurus*; Eastern lineage 237 specimens selected from all its geographic range (S7150-22, S7161-24, CN4226-26 and 238 CN10775-26). For further information on the selected specimens and dataset specifications 239 refer to Tables S1 and S2.

240 Time-calibrated species trees for each of the four datasets above were inferred with 241 SNAPP v.1.5.2 (Bryant et al. 2012) twice. First, we generated a time-calibrated species tree 242 with SNAPP using the 'snapp prep.rb' script (*https://github.com/mmatschiner/tutorials*). We 243 dated the deepest node in the phylogeny as suggested by Stange et al. (2018) with a normal 244 distribution from the mean age extracted from Tejero-Cicuéndez et al. (2022). Mutation rates 245 (u & v) were fixed to 1, and a uniform distribution was set for the population mutation rate 246 theta (θ) with default boundaries (0-1,000) and was constrained to be identical on all 247 branches. The latter prior is assumed by the script 'snapp prep.rb' to decrease the 248 computational load of the analysis (Stange et al., 2018). We also repeated the same analysis 249 without linking population mutation rates and applying specific Yule (λ) and Theta priors (θ ; 250 see Coalescent-based Species Delimitation below). In both approaches we ran 4 independent 251 runs of 3,000,000 generations, sampling every 50 generations. Convergence between runs and 252 stationarity was checked with Tracer v.1.7 (Rambaut & Drummond, 2013). Posterior 253 distributions were combined with LogCombiner v.2.6.3, discarding 30% of the posterior trees

254	as burn-in and a maximum clade credibility tree was obtained calculating median heights in
255	TreeAnnotator v.2.6.3 (BEAST2 v.2.6.4; Bouckaert et al., 2019).
256	In addition, we generated species tree estimations for dataset 4 BPP to dataset 7 BPP
257	(Table S2) with BPP A01 analysis (Flouri, Jiao, Rannala, & Yang, 2018). We followed the
258	pipeline of Huang (2018) (<i>https://github.com/airbugs/Dynastes_delimitation</i>) to estimate θ
259	and τ priors (a=3, b=0.061 and a=3, b=0.131 respectively), and produced a dataset where all
260	loci were present in at least one individual of each lineage. Then, we implemented three
261	independent runs of 500,000 generations sampling every 10 generations after a burn-in of
262	50,000.
263	Finally, we generated a ML species tree with dataset 3 (Table S2). We followed the
264	same procedure in IQ-TREE (Nguyen et al., 2015) as in Methods 2.5 but when summarizing
265	all the trees with Astral v.5.7.8 (Zhang et al., 2017) we assigned each individual to its group
266	as inferred by ADMIXTURE.
267	
268	2.7. Species tree topology testing
269	As shown in results section 3.3.2, our coalescent-based species trees recovered two different
270	topologies: the first topology (from now on Topol) separates T. hajarensis into two clades:
271	the first conformed by Western and Central Hajars' lineages, and the other by Eastern Hajars
272	and Masirah Island's lineages. The second topology (from now on Topo2) places all Hajar
273	Mountains' lineages as a clade sister to Masirah Island's. Both trees were rooted with T.
274	spatalurus. We evaluated both topologies by implementing the recently published mixture
275	across sites and trees (MAST) model (Wong et al., 2022). In short, this program calculates
276	topology weights for a number of given topologies (two in our case) across an alignment,

277 calculates the most supported topology at each site, and returns an overall value for the whole

278 alignment (determined by a topology weight ranging between 0 and 1). We generated a 279 MAST analysis for each polymorphic loci present in each quintet of specimens containing 280 one individual from all T. hajarensis lineages (Western, Central, Eastern and Masirah) and 281 one individual of *T. spatalurus*. There were a total of 37,800 quintets which contained, on 282 average, 537 ± 141 loci extracted from *dataset 3* (Table S2). Then, we ran MAST 283 independently for each loci (19.35x10⁶ different MAST analyses) giving the aforementioned 284 topologies as input trees and unlinking substitution models. DNA frequencies and gamma 285 model across trees. 286 Resulting overall tree weights for each loci were extracted, weights not supporting either 287 of the given topologies with a probability above 0.55 were discarded, and the resulting

288 3.96x10⁶ weights were summarized by specimen and quintet independently. We identified the

289 most supported topology for each specimen by averaging the number of loci between all

290 quintet combinations where a specific specimen was present. When grouping our results by

291 quintet, we implemented a linear model to identify up to which point distance to the putative

292 contact zone (distance to Central Hajars in *Topo1*, and distance to Masirah in *Topo2*)

293 explained the observed shifts in topology weights. We calculated euclidean distances between

all Eastern *vs* Masirah Island and Eastern *vs* Central Hajars specimens and accounted for all

quintet combinations where Eastern and Masirah/Central specimens remained identical in two

different ways: On the one hand, we obtained a single topology weight for each locality by

averaging all pseudoreplicates. On the other, we accounted for all pseudoreplicates

implementing a nested analysis of variance.

299

295

300 **2.8. Coalescent-based Species Delimitation**

301 In the present work, we use the General Lineage Species Concept (de Queiroz, 2007). This 302 unified species concept considers species as separately evolving meta-population lineages and 303 treats this property as the single requisite for delimiting species. We consider continuously 304 distributed populations that show contemporary gene flow and no other notable forms of 305 divergence to be a metapopulation lineage of the same species (Chan et al., 2020). Other 306 properties, such as phenetic distinguishability, reciprocal monophyly, and pre- and 307 postzygotic reproductive isolation, are not part of this species concept but serve as important 308 lines of evidence relevant to assess the separation of lineages and therefore species status (de 309 Queiroz, 2007).

310 We designed and tested several coalescent-based species delimitation models by 311 splitting or lumping the previously inferred T. hajarensis lineages. Specifically, our species 312 delimitation hypotheses included: *i*) Single-species hypothesis (H_0) ; *ii*) Two species within T. 313 *hajarensis* (H_1 : Western+Central vs Masirah + Eastern lineages); *iii*) Three-species hypothesis 314 $(H_2: Western + Central vs Eastern vs Masirah lineages); and vi) All lineages as distinct$ 315 species (H_3 : Western vs Central vs Eastern vs Masirah lineages). Species delimitation analyses 316 were tested with Bayes factor delimitation (BFD* with genomic data; Leaché et al. 2014) 317 implemented in BEAST2 v.2.6.4 (Bouckaert et al., 2019), and with BPP A10 analysis (Yang 318 & Rannala, 2010).

Within BFD*, we estimated for each model a species tree with SNAPP v.1.5.2 (Bryant et al., 2012) and conducted a path sampling analysis to estimate and rank marginal likelihoods between models. We then computed Bayes Factors (BF) to determine the best SDM. Since SNAPP is computationally intensive and assumes no gene flow between species, we selected 2 non-admixed individuals from each species hypothesis (*dataset 4 SNAPP*; Table S2). We also included representatives of its sister species *T. spatalurus* to account for *T. hajarensis*

325 being a single species, resulting in a dataset of 10 individuals and 2,147 unlinked SNPs. 326 Mutation rates (u & v) were fixed to 1. We set the Yule prior (λ) to a gamma distribution and 327 while alpha was set to 2, beta was estimated by calculating the expected tree height on the 328 5,219 loci dataset (maximum observed divergence between any pair of taxa divided by 2). We 329 then used 'pyule' (https://github.com/joaks1/pyule) to determine the mean value of lambda 330 and calculated beta accordingly ($\lambda = \alpha \times \beta$). Theta prior (θ) was also set to a gamma 331 distribution and the mean value of θ was estimated by averaging all genetic distances within 332 each lineage. Path sampling analyses were run for 20 steps with the following parameters: 333 500,000 MCMC generations sampling every 1,000, with an alpha of 0.3, 10% burnin and a 334 pre-burnin of 50,000. Stationarity of all runs was checked and each step was run until ESS >= 335 200. 336 We implemented further guided species delimitation analyses with BPP v.4.4.1 337 (Rannala & Yang, 2013; Yang & Rannala, 2010). We used *dataset 4 BPP* which contained 338 the same individuals as in BFD* (Table S2), used the best species tree topology (*Topol*; see 339 results 3.4) as the guide tree, and processed the dataset as described above (see methods 2.6). Then, we implemented three independent runs of BPP A10 species delimitation (Yang & 340 341 Rannala, 2010) with 1,000,000 generations, sampling every 10 generations after a burn-in of 342 100,000. 343 344 2.9. Species validation

To test the robustness of our inferred species, we calculated the genealogical divergence index (*gdi*) proposed by Jackson et al. (2017), and implemented within BPP by Leaché et al. (2019). The equation for *gdi* is $1 - e^{-2\tau AB/\theta A}$, where τ represents the divergence time between species *A* and *B* and θ represents the population size of the species *A*. Therefore, to effectively test species status for two given taxa this index has to be calculated reciprocally between taxa *A* and *B*. According to Jackson et al. (2017), low *gdi* values (*gdi* < 0.2) indicate that *A* and *B* are the same species while high values (*gdi* > 0.7) support a distinct species status of the two taxa. Values in between are considered ambiguous, lacking the support necessary to be classified as distinct species.

354 We implemented this index to the two competing species tree hypotheses (See results 355 3.3.2). We ran A00 analyses with *dataset 4 BPP* (Table S2) estimating the parameters on the 356 inferred guide trees (Topol and Topo2) to generate the posterior distributions for the most 357 recent species divergences, and then we collapsed those tips and repeated the analysis to test 358 species status for the lumped lineages (see Leaché et al., 2019 for a similar approach). Three 359 independent runs of A00 were performed for 100,000 generations, sampling every five 360 generations after a burn-in of 10,000. We calculated and visualized gdi indexes with R v.4.2.1 361 (R Core Team, 2022).

362

363 **2.10. Introgression analysis**

We used the D-Statistics (ABBA-BABA tests) implemented in Dsuite (Malinsky, Matschiner,
& Svardal, 2021) to test for past signals of introgression between *T. hajarensis* lineages. To
do so, we considered four *T. hajarensis* lineages (as supported by BFD* analysis) and
performed the analysis providing both *Topo1* and *Topo2* respectively. In all analyses, *T. spatalurus* was used as outgroup. Finally, we calculated the f-branch statistics to infer the
excess of shared derived alleles between lineages.

371 3. Results

372 **3.1. ddRADseq data processing**

Total reads obtained from sequencing added up to 11.6 x 10⁷ and after applying quality filters more than 97% of the raw reads remained, with an average of 2.1 x 10⁶ reads per individual. Post-processing filtering identified up to five individuals with low coverage levels, which were discarded from subsequent analyses. The number of loci within *loci* datasets ranged between 30,526 to 4,441 loci, and *SNP* datasets contained between 30,096 and 2,428 SNPs, depending on the number of individuals used, and the applied filters for each analysis. See Table S2 for further information on each dataset.

380

381 3.2. Population structure

382 We explored the population structure of *T. hajarensis* with *dataset 1* (Table S2), which

383 contained 2,428 unlinked SNPs present in at least 60% of all individuals (missing genotype

384 call rate set to 40). The most likely number of ancestral populations recovered from

385 Admixture was K=3 (cross-validation = 0.43). This configuration geographically segregates

386 *T. hajarensis* into Western + Central Hajars, Eastern Hajars, and Masirah Island, with almost

387 no signal of gene flow between them. Lower k numbers lumped together Masirah and Eastern

388 lineages and higher k numbers split the occidental clade into the Central and Western lineages

389 (Figure 2a & b, Figure S1). Similar results were obtained with the Principal Component

390 Analysis (PCA), which also segregated *T. hajarensis* into the three aforementioned clusters.

391 Interestingly, all continental lineages clustered closer together than to the Masirah Island

392 group (Figure 2c).

To further inspect its population structure, we implemented fineRADstructure to *dataset 2* (Table S2). Results again show a clear structure into three groups, being Masirah Island the cluster sharing the highest number of haplotypes between its specimens (up to 1,110; Figure 2d). Surprisingly, the Eastern lineage showed high levels of coancestry with

397 both Masirah Island and Western + Central lineages, suggesting a shared evolutionary history 398 with both of them. Such coancestry was stronger in specimens that where geographically 399 closer to either Masirah Island (localities 25 and 26) or the Central Hajars (localities 19 to 400 21). This can be clearly appreciated at the crossroads between the Central and Eastern 401 lineages where we find an individual (CN3750-19; Figure 2d) that almost shares the same 402 coancestry with either Central and Eastern lineages, most likely representing a hybrid 403 between them (Figure 2d). Finally, the phylogenomic network also shows the presence of all 404 three lineages and, even though the Eastern Hajars are more related to Masirah, there are 405 some reticulation events that approximate both Central and Eastern groups (Figure 2e). 406

407 **3.3.** Phylogenetic reconstructions

3.3.1. Mitochondrial tree reconstruction 408

409 The phylogeny reconstructed with the 12S mitochondrial gene (Figure 3a) is concordant with 410 the results shown in de Pous et al. (2016) with most of the Hajar Mountain's lineages forming 411 a monophyletic group (localities 1–24). In this group, there are two main lineages: the first 412 lineage spans from the Musandam Peninsula southwards into the westernmost part of the 413 Central Hajars (localities 1–16), and the second is formed by the specimens inhabiting the 414 Central and most of the Eastern Hajars (localities 17 - 24). However, not all T. hajarensis 415 inhabiting the Hajar Mountains cluster together, since the two easternmost localities 416 (localities 25 and 26 in Figure 1) were recovered as sister group to the specimens in Masirah 417 Island (Figure 3a).

418

419 3.3.2. Phylogenomic reconstructions and topological discordances

420 All phylogenomic reconstructions were discordant with the mitochondrial phylogeny and, 421 overall, they show a clear geographical structure with three distinct lineages in the Eastern, 422 Central, and Western Hajars. However, we recovered different topologies regarding the 423 position of Masirah Island's lineage and the phylogenomic relationships between the 424 aforementioned groups. The concatenation-based BI approach recovered Masirah Island and 425 Eastern Hajars lineages as reciprocally monophyletic (Figure 3b). A very similar topology 426 was obtained with the summarized ML phylogeny (Figure S2) with the exception of the 427 specimen CN3750-19, that was recovered as basal to both Masirah Island and Eastern Hajars 428 lineages, and which most likely represents a hybrid specimen between Central and Eastern 429 lineages (Figure 2d). Finally, concatenation-based ML approaches placed Masirah Island's 430 lineage within the Eastern lineage (Figure S2). All the above support a Western + Central clade as sister to an Eastern + Masirah clade. 431

Two species tree topologies were recovered across all species tree reconstruction
methods (Figure S3). The first follows the same topology obtained with BI and yields a
Western + Central clade sister to an Eastern + Masirah clade (*Topo1*). The second topology
places all Hajar Mountains' *T. hajarensis* as a monophyletic group sister to the Masirah Island
lineage (*Topo2*), and was recovered in three SNAPP and one BPP species tree analyses
(Figure S3). Both topologies are supported with a posterior probability of 1 in all nodes in at
least one analysis (Figure S3).

439

440 **3.4. Species tree topology testing**

Between the two competing species tree hypotheses, *Topo1* was consistently recovered as the
best topology for almost all specimens and quintets. Only three specimens near the contact
zone between Central and Eastern Hajar's populations presented low to negative number of

loci supporting *Topo1* (CN3750-19, OM04_2010_66-20 and OM04_2010_99-21; Figure 4a),
and it is noteworthy that in the best Admixture scenario only one of the three specimens
(CN3750-19) was recovered as admixed (Figure 2a).

447 Splitting our dataset into quintets allowed us to investigate the effects of past gene 448 flow in the observed topology of *T. hajarensis*. When evaluating topology weights through a 449 geographic gradient (either distance between Eastern Hajars and Masirah Island's specimens 450 in Figure 4b, or distance between Central Hajars and Eastern Hajars' specimens in Figure 4c) 451 we saw that support for Topol decreased when Eastern Hajars and Masirah Island's 452 specimens were farther apart (Figure 4b), and when Central and Eastern Hajar's specimens 453 were closer together (Figure 4c). Both geographic gradients significantly explain the shifts in 454 topology weights (*p*-value $\ll 0.01$; Figure 4b,c) and point towards the same direction, a shift 455 from Topol to Topo2 in the Central/Eastern contact zone. Distance between Central Hajars 456 and Eastern Hajars' specimens explained much better the observed topology weights ($R^2 =$ 457 0.82; Figure 4c) than distance between Eastern Hajars and Masirah Island's specimens ($R^2 =$ 458 0.33; Figure 4b), which is concordant with *Topol* representing the *true* evolutionary history 459 of T. hajarensis. Altogether, results suggest that Topo2 is the result of a secondary contact 460 between Central and Eastern lineages. The gradual higher support of Topo2 when Eastern 461 Hajars' specimens are closer to the Central Hajars instead of a clear-cut shift in the contact 462 zone could be indicative of past dispersals of Central Hajars' specimens into the Eastern 463 block, leaving a genomic footprint in the nowadays Eastern Hajars' populations.

Pseudoreplicate or sample choice (different Western Hajars and Central Hajars
samples for each distance between Eastern *vs* Masirah Island's specimens; Figure S4a;
different Western Hajars and Masirah Island samples for each distance between Central *vs*Eastern Hajars' specimens; Figure S4b) was recovered as significant but with less power than

468 distance between Eastern Hajars and the Masirah Island (Table S3) or Central Hajars' 469 specimens (Table S4). This means that if we downsample our dataset and select far enough 470 Eastern and Central Hajars' specimens, we will obtain *Topol* regardless of the other samples 471 in the quintet. However, if we select geographically close Eastern and Central Hajars' 472 specimens, we can obtain either *Topo1* or *Topo2* in our resulting species tree, depending on 473 the other Western Hajars and Masirah Island samples selected (Figure S4). 474 475 3.5. Species delimitation, species validation and signals of introgression 476 Based on results from our population clustering and structure analyses, we tested up to 4 477 lineages (Western, Central, Eastern and Masirah) as new putative species selecting only non-478 admixed individuals. BFD* supported the species status for all four putative species with a 479 decisive strength of support ($\ln(BF) > 5$; Kass & Raftery, 1995; Table S5). Contrastingly, 480 species delimitation analysis conducted with BPP supported the split of only three new 481 putative species, lumping the Western and Central lineages into a single species (species 482 delimitation model H_2 supported with a posterior probability of 0.95). 483 The heuristic genealogical divergence index (gdi) was implemented on dataset 4 BPP 484 (Table S2), and similar results were recovered among both species tree topologies (Figure 4). 485 First, T. hajarensis was compared to T. spatalurus to test the robustness of the current 486 systematics on the group and results supported both species as distinct from each other with 487 gdi values well above the 0.7 threshold (Figure 5c and Figure S5c). In contrast to the species 488 delimitation models, comparisons between Hajar Mountain's lineages did not support, in any 489 case, the species status of these lineages, with gdi values falling either below 0.2 or within the 490 ambiguous range. Masirah Island's group presented a *gdi* above 0.9 which would support the 491 species status of Masirah with respect to its sister group. Incongruently, Masirah's sister

492 lineage (either the Eastern Hajars lineage or the clade comprising all the Hajar Mountains' *T*.

493 *hajarensis*, depending on the topology) is not supported as a distinct species from Masirah

494 $(gdi = 0.19 \ 0.01 \text{ in Figure 5c, and } gdi = 0.35 \ 0.09 \text{ in Figure S5c, respectively}).$

495 Tests of introgression were implemented through D-statistics applied to the SNPs

496 extracted from *dataset 3* (30,096 SNPs). When *Topo1* was given as a guide tree, statistically

- 497 significant introgression was revealed between the Eastern and the Central lineages (Figure
- 498 5b). Otherwise, when *Topo2* was forced, significant past introgression was recovered between
- 499 Eastern and Masirah Island lineages (Figure S5b).
- 500

501 **4. Discussion**

502 4.1. A highly structured species with past or recent gene flow

503 Our genomic analyses show that *T. hajarensis* is comprised by three well-defined lineages 504 geographically located in the Western + Central Hajar Mountains, the Eastern Hajar 505 Mountains, and in Masirah Island, respectively. However, the phylogenetic relationships of 506 these lineages differ between methodologies showing high levels of mito-nuclear discordance 507 (Figure 3) and introgression between either Central Hajars and Eastern Hajars lineages 508 (Figure 5b), or Eastern Hajars and Masirah Island lineages (Figure S5b). The mitochondrial 509 phylogenetic reconstruction showed a widely spread Central Hajars' lineage including most 510 of the Eastern Hajars, with only the easternmost region of the Hajar Mountains, near Ra's al 511 Hadd, being sister to the Masirah Island's group (Figure 3). Incongruently, none of our 512 nuclear genomic analyses recovered this topology, with a clear divergence between Central 513 and Eastern lineages in the region surrounding the Semail gap (localities 19 to 21), a well-514 known topographic feature that has already been used to delimit the separation between the 515 Eastern and Central Hajars (Garcia-Porta et al., 2017). Moreover, population structure

516 analyses showed almost no signals of admixture between Central and Eastern populations 517 (Figure 2a, b & c). However, when inspecting the fineRADstructure results (Figure 2e), the 518 Eastern lineage appeared as greatly substructured, with an increasing gradient of shared 519 alleles when geographically closer to the Central Hajars. This substructure may be caused by 520 the habitat and climatic heterogeneity, and the complex topography of the Hajar Mountains 521 (Burriel-Carranza et al., 2019; Carranza et al., 2018), which could be hindering panmixia 522 within the group. The increasing gradient of coancestry observed in fineRADstructure also 523 suggests a secondary contact between either the Eastern and Masirah lineages, or the Eastern 524 and Central lineages, which was resolved by our MAST analysis (Figure 4). The MAST model recovered Masirah Island's lineage as sister to the Eastern lineage and showed that the 525 526 topology where all Hajar Mountains' lineages form a monophyletic group (Topo2) is most 527 likely a product of a secondary contact between Eastern and Central specimens, having a 528 strong signal in the contact zone and gradually fading out with distance. The lack (or low 529 levels) of current admixture between Eastern and Central specimens suggest that the observed 530 patterns are a result of past secondary contacts or dispersal events from the Central Hajars 531 towards the Eastern Hajars, leaving genomic evidence of nuclear and mitochondrial 532 introgression between lineages.

533 Our D-statistic analysis points out that, among the number of processes leading to 534 conflicts between mitochondrial and nuclear genomes, introgression might be responsible for 535 the observed discordance in this case. Since the mitochondrial genome is maternally inherited 536 and does not segregate or recombine, introgression events can incorporate a complete foreign 537 mitochondrial genome into a population and maintain it over long periods of time, resulting 538 into extremely diverged haplotypes between specimens from the same population but from

differently-inherited mitochondrial genomes. Moreover, the lack (or low levels) of current
nuclear admixture between Central and Eastern lineages support a past hybridization event.

541

542 4.2. Caveats of downsampling individuals when building species trees

543 The high computational resources needed to infer species trees with NGS data commonly 544 leads to a dataset downsampling, selecting only some individuals for such analyses (Dufresnes et al., 2020; Kornilios et al., 2019; Thanou, Kornilios, Lymberakis, & Leaché, 545 546 2020). Moreover, since SNAPP does not assume gene flow, specimens are usually selected 547 using population structure tools such as ADMIXTURE, with the assumption that any 548 specimen from the same lineage species that is not admixed with other lineages should 549 coalesce first within its group. Here, we provide an example where, even though discarding 550 individuals with admixed proportions of their genomes, sample selection can greatly influence 551 the topology and branch lengths of the resulting trees. It is especially worrying the case of the 552 specimen OM04 2010 66-20, which even though it is not recovered as admixed in the best K 553 of ADMIXTURE, it contains more loci supporting the hybrid topology (*Topo2*; Figure 4) than 554 the *true* topology (*Topol*; Figure 4). While most of the reconstructed species trees support 555 *Topo1*, there are some cases where *Topo2* has higher node support (Figure S3).

In this case, several particular effects or the combination of them might be promoting such incongruent results. On the one hand, biological effects such as the small population size of Masirah Island, or the introgression between Eastern and Central+Western lineages might be confounding species tree estimation analyses. On the other hand, operational effects such as the usage of SNAPP linking all population sizes, a common feature used to reduce model parameters and achieve feasible run times (Stange et al., 2018) seems to be promoting a topology where Masirah Island is sister to the rest of the Hajar Mountains' *T. hajarensis*

563 (Figure S3a, c and d), while the same analyses unlinking population sizes recover a clearly564 distinct topology.

565 Overall, our study suggests that specimen selection when downsampling datasets for 566 species tree estimation should be proceeded with caution, especially if there are signs of 567 recent or past gene flow or different population sizes between taxa. Therefore, downsampled 568 datasets should not include specimens near contact zones and, if possible, include specimens 569 spanning throughout the whole distribution of each taxon.

570

571 4.3. One, two or three species? Importance of validation in species delimitation methods 572 Our analyses applying a variety of species delimitation methods yielded different results. 573 Mitochondrial-based GMYC supports T. hajarensis to be a complex of five species, BFD* 574 supports four species and BPP three. If we consider the integrated results of BFD* and BPP, 575 both support the split of Western+Central, Eastern and Masirah as three distinct species. Such 576 species would fall within the spectrum of cryptic species, since even though they are 577 genetically identifiable there are no obvious distinctive phenotypic traits between them (de 578 Pous et al. 2016; S.C. pers. observ.). Contrastingly, when using an heuristic criterion to 579 validate species status (the gdi) none of the putative species was fully recovered as a distinct 580 species (Figure 5c). The ambiguous results between Masirah Island's lineage and its 581 respective sister group are most likely due to a stated weakness of the *gdi* which, in the case 582 of populations founded by a small number of individuals or in cases where two lineages have 583 very different population sizes, results may lead to claims of species status even if the groups 584 diverged very recently (Leaché et al., 2019). Our results support a single colonization event of 585 Masirah Island and the high number of shared alleles between specimens also suggests that 586 this population was founded by few individuals. Considering the different population sizes

with its sister group and the low values of *gdi* of its counterpart test (Figure 5c), we identify Masirah Island's specific status claim as a false positive. Validating SDMs has proven crucial to not contribute to taxonomic inflation in *T. hajarensis* and we suggest that, together with an in-depth exploration of population structure and gene flow, the *gdi* (or any other similar heuristic method) should become a key tool in species delimitation validation, specially when testing the species status of allopatric early divergent lineages (Leaché et al., 2019).

593

594 **4.4. Systematics and Biogeography**

595 The present study shows that *T. hajarensis* is formed by three well supported lineages with a 596 complex and dynamic evolutionary history. The first divergence within the species separated 597 the *T. hajarensis* currently inhabiting the Western and Central Hajar Mountains from the *T*. 598 hajarensis on Masirah Island and the Eastern Hajars. The split between these clades is 599 estimated around Early to Mid-Quaternary (1.4 my; 0.7-2.13 mya HPD95%; Figure 5), setting 600 the first diversification within the group about three million years after previous estimations 601 using mitochondrial markers (de Pous et al., 2016; Figure 3). Given the distribution of the 602 different lineages we can assume that they speciated through allopatric isolation, probably 603 caused by a combination of past geographical and climatic events. Interestingly, Western + 604 Central Hajars lineage and Eastern hajars lineage are found nowadays in allopatry in the same 605 mountain range (Figure 1). However, the introgression signal recovered between the two 606 lineages and the mitochondrial phylogeny suggest that at some point they cohabited and gene 607 flow between lineages occurred. This could indicate that the Eastern lineage colonized the 608 Hajar Mountains posteriorly to the Western + Central clade and progressively displaced the 609 Western + Central populations to its current distribution exchanging genes in the process. 610 Another explanation would be that Western + Central populations in the Eastern Hajars 611 merged with the arriving Eastern lineage and the high topography of the Central Hajars

612 hindered the dispersal and complete fusion of both lineages into one. There are other 613 examples where genetic isolation is found between lineages in the Eastern and Central Hajars 614 (Carranza & Arnold, 2012; Garcia-Porta et al., 2017) and in the case of the geckos of the 615 genus Asaccus, it has even led to the description of a new species (Simó-Riudalbas et al., 616 2018). Overall, we can speculate that even though currently there is low gene flow between 617 lineages, genetic barriers are diffuse and, if at some point climatic conditions favor increased 618 dispersal of this ground-dwelling species, both groups could eventually fuse back into one. 619 In the Arabian Peninsula, Pliocene and Quaternary climatic shifts between humid and 620 arid to hyper-arid episodes have resulted from repeated high-latitude glacial events and global 621 sea level falls, promoting desert formation (Glennie, 1998). In southeastern Oman, we find 622 the Sharqiyah Sands, a sand dune desert where T. hajarensis has never been reported and 623 which separates the Masirah's lineage by roughly 200 km in a straight line to its closest sister 624 group in the Eastern Hajars (Figure 1). This desert is thought to be a result of Late Ouaternary 625 climatic shifts in southeast Arabia (although see Metallinou & Carranza, 2013) and its 626 formation is linked to the onshore-blowing SW Monsoon (Glennie, 1998) with prior fluvial 627 deposits dating back to the Plio-Pleistocene and first aeolian sands consolidating about 628 160,000 years ago (Radies, Preusser, Matter, & Mange, 2004). The divergence between Masirah and Eastern lineages dates to 1.1 mya (0.5 – 1.6 mya HPD95%; Figure 5), a time 629 630 range prior to the consolidation of the Sharqyah Sands. This would be coherent with a broader 631 historic distribution range of T. hajarensis encompassing the current region where today the 632 Sharqyah Sands reside. This would have facilitated colonization by adrift specimens or 633 clutches, since Masirah Island is only about 20 km from the closest continental shore. 634 Moreover, sea level drops of up to 130 m induced by high-latitude glacial events were 635 recurrent during the Plio-Pleistocene (Gleenie, 1998). The greatest depth between Masirah

636 Island and the coast of Oman rarely surpasses 50 m, thus such sea level drops could have 637 provided a land bridge between Masirah Island and the continent, facilitating land dispersal of 638 T. hajarensis and other reptile species towards the Island. Phylogenomic reconstructions and population structure analyses support a single colonization event of Masirah Island from few 639 640 founder individuals, with all specimens constituting a monophyletic group (Figure 3b) with 641 high levels of coancestry among them (Figure 2d). These results, together with the increased 642 sampling effort throughout several localities in the island, are in agreement with previous 643 findings (de Pous et al., 2016), and also advocate for a natural colonization of Masirah Island 644 in contrast to a human mediated introduction.

645

646 **5. Conclusions**

647 Overall, T. hajarensis seems to be conformed by a single, highly-structured species. The 648 usage of genomic techniques has proven vital to determine the population structure. 649 introgression events, and mito-nuclear discordances within this system, but further studies 650 with whole genome sequencing data might be necessary to fully understand the regions of 651 introgression within the species. Moreover, although NGS and the MSC are essential tools to 652 shorten the Linnean Shortfall and are well suited to uncover cryptic diversity, we show that 653 dataset downsampling and species delimitation methods need to be cautiously implemented 654 and validated to avoid contributing to taxonomic inflation. Altogether, this system offers the 655 rare possibility to witness both possible evolutionary outcomes of an incipient species at the 656 same time. On the one hand, if isolation is maintained between continental and Masirah 657 Island's populations, both lineages will most likely evolve into different species. On the other 658 hand, its sister lineage could potentially merge back with the other Hajar Mountain's T. 659 hajarensis, dissolving almost 1.5 my of independent evolution. However, only time will tell.

660 Author's contributions

BB-C and SC conceived and designed the study. BB-C and ME performed laboratory work,
analyzed the data and wrote the manuscript with input from all other authors, who approved the
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664

666

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693 **References**

- Arnold, E. N. (1980). The reptiles and amphibians of Dhofar, southern Arabia. Journal of
 Oman Studies, Special Report, 2, 273–332.
- 696 Bouckaert, R., Vaughan, T. G., Barido-Sottani, J., Duchêne, S., Fourment, M., Gavryushkina,
- A., ... Drummond, A. J. (2019). BEAST 2.5: An advanced software platform for
 Bayesian evolutionary analysis. PLOS Computational Biology, 15(4), e1006650. doi:
 10.1371/JOURNAL.PCBI.1006650
- 700 Bray, T. C., Alagaili, A. N., & Bennett, N. C. (2014). A widespread problem: Cryptic
- 701 diversity in the Libyan jird. Zoological Studies 2014 53:1, 53(1), 1–8. doi:
 702 10.1186/S40555-014-0033-3
- 703 Bryant, D., Bouckaert, R., Felsenstein, J., Rosenberg, N. A., & Roychoudhury, A. (2012).
- 704 Inferring Species Trees Directly from Biallelic Genetic Markers: Bypassing Gene Trees
- in a Full Coalescent Analysis. Molecular Biology and Evolution, 29(8), 1917–1932. doi:

706 10.1093/MOLBEV/MSS086

- Burriel-Carranza, B., Els, J., & Carranza, S. (2022). Reptiles & Amphibians of the Hajar
 Mountains. Madrid: Editorial CSIC. ISBN: 978-84-00-10988-2.
- 709 Burriel-Carranza, B., Tarroso, P., Els, J., Gardner, A., Soorae, P., Mohammed, A. A., ...
- 710 Carranza, S. (2019). An integrative assessment of the diversity, phylogeny, distribution,
- and conservation of the terrestrial reptiles (Sauropsida, Squamata) of the United Arab
- 712 Emirates. PLOS ONE, 14(5), e0216273. doi: 10.1371/JOURNAL.PONE.0216273

713 Carranza, S., & Arnold, E. N. (2012). A review of the geckos of the genus Hemidactylus 714 (Squamata: Gekkonidae) from Oman based on morphology, mitochondrial and nuclear 715 data, with descriptions of eight new species. In Zootaxa (Vol. 95). 716 Carranza, S., Els, J., & Burriel-Carranza, B. (2021). A field guide to the reptiles of Oman. 717 Madrid: Digital CSIC. ISBN: 978-84-00-10876-2. 718 Carranza, S., Xipell, M., Tarroso, P., Gardner, A., Arnold, E. N., Robinson, M. D., ... Akhzami, S. N. A. (2018). Diversity, distribution and conservation of the terrestrial 719 720 reptiles of Oman (Sauropsida, Squamata). In PLoS ONE (Vol. 13). doi: 721 10.1371/journal.pone.0190389 722 Chan, K. O., Hutter, C. R., Wood, P. L., Grismer, L. L., Das, I., & Brown, R. M. (2020). Gene 723 flow creates a mirage of cryptic species in a Southeast Asian spotted stream frog complex. Molecular Ecology, 29(20), 3970-3987. doi: 10.1111/mec.15603 724 725 Chang, C. C., Chow, C. C., Tellier, L. C. A. M., Vattikuti, S., Purcell, S. M., & Lee, J. J. 726 (2015). Second-generation PLINK: Rising to the challenge of larger and richer datasets. 727 GigaScience, 4(1), 7. doi: 10.1186/S13742-015-0047-8/2707533 728 Chattopadhyay, B., Garg, K. M., Kumar, A. K. V., Doss, D. P. S., Rheindt, F. E., Kandula, S., 729 & Ramakrishnan, U. (2016). Genome-wide data reveal cryptic diversity and genetic 730 introgression in an Oriental cynopterine fruit bat radiation. BMC Evolutionary Biology, 731 16(1), 1-15. doi: 10.1186/S12862-016-0599-Y/TABLES/1 732 de Pous, P., Machado, L., Metallinou, M., Červenka, J., Kratochvíl, L., Paschou, N., ... 733 Carranza, S. (2016). Taxonomy and biogeography of Bunopus spatalurus (Reptilia; 734 Gekkonidae) from the Arabian Peninsula. Journal of Zoological Systematics and 735 Evolutionary Research, 54(1), 67–81. doi: 10.1111/jzs.12107 736 De Queiroz, K. (2007). Species Concepts and Species Delimitation. Systematic Biology, 737 56(6), 879-886. doi: 10.1080/10635150701701083 738 Derycke, S., Fonseca, G., Vierstraete, A., Vanfleteren, J., Vincx, M., & Moens, T. (2008). 739 Disentangling taxonomy within the *Rhabditis* (*Pellioditis*) marina (Nematoda, 740 Rhabditidae) species complex using molecular and morhological tools. Zoological 741 Journal of the Linnean Society, 152(1), 1–15. doi: 10.1111/J.1096-3642.2007.00365.X 742 Dufresnes, C., Pribille, M., Alard, B., Gonçalves, H., Amat, F., Crochet, P.-A., ... Martínez-743 Solano, I. (2020). Integrating hybrid zone analyses in species delimitation: Lessons

from two anuran radiations of the Western Mediterranean. Heredity, 124(3), 423–438.
doi: 10.1038/s41437-020-0294-z

Eaton, D. A. R., & Overcast, I. (2020). ipyrad: Interactive assembly and analysis of RADseq

747 datasets. Bioinformatics, 36(8), 2592–2594. doi: 10.1093/BIOINFORMATICS/BTZ966

- Ezard, T., Fujisawa, T., & Barraclough, T. (2021). splits: SPecies' LImits by Threshold
 Statistics. Retrieved from https://R-Forge.R-project.org/projects/splits/
- 750 Fennessy, J., Bidon, T., Reuss, F., Kumar, V., Elkan, P., Nilsson, M. A., ... Janke, A. (2016).
- Multi-locus Analyses Reveal Four Giraffe Species Instead of One. Current Biology,
 26(18), 2543–2549. doi: 10.1016/J.CUB.2016.07.036
- 753 Fišer, C., Robinson, C. T., & Malard, F. (2018). Cryptic species as a window into the
- paradigm shift of the species concept. Molecular Ecology, 27(3), 613–635. doi:
- 755 10.1111/MEC.14486
- Flouri, T., Jiao, X., Rannala, B., & Yang, Z. (2018). Species Tree Inference with BPP Using
 Genomic Sequences and the Multispecies Coalescent. Molecular Biology and
 Evolution, 35(10), 2585–2593. doi: 10.1093/molbev/msy147
- 759 Fujisawa, T., & Barraclough, T. G. (2013). Delimiting Species Using Single-Locus Data and
- 760 the Generalized Mixed Yule Coalescent Approach: A Revised Method and Evaluation
- 761 on Simulated Data Sets. Systematic Biology, 62(5), 707–724. doi:
- 762 10.1093/sysbio/syt033
- Garcia-Porta, J., Simó-Riudalbas, M., Robinson, M., & Carranza, S. (2017). Diversification in
- arid mountains: Biogeography and cryptic diversity of *Pristurus rupestris rupestris* in
 Arabia. Journal of Biogeography, 44(8), 1694–1704. doi: 10.1111/jbi.12929
- Glennie, K. W. (1998). The desert of southeast Arabia: A product of Quaternary climatic
 change. In Quaternary Deserts and Climatic Change. CRC Press.
- Gosselin, T., Lamothe, M., Devloo-Delva, F., & Grewe, P. (2017). radiator: RADseq data
 exploration, manipulation and visualization using R. R Package Version 0.0, 5.
- Huson, D. H., & Bryant, D. (2006). Application of Phylogenetic Networks in Evolutionary
 Studies. Molecular Biology and Evolution, 23(2), 254–267. doi:
 10.1093/molbev/msj030
- 773 Isaac, N. J. B., Mallet, J., & Mace, G. M. (2004). Taxonomic inflation: Its influence on

macroecology and conservation. Trends in Ecology & Evolution, 19(9), 464–469. doi:

775 10.1016/J.TREE.2004.06.004

- 776 Ivanov, V., Lee, K. M., & Mutanen, M. (2018). Mitonuclear discordance in wolf spiders:
- Genomic evidence for species integrity and introgression. Molecular Ecology, 27(7),
 1681–1695. doi: 10.1111/MEC.14564
- Jackson, N. D., Morales, A. E., Carstens, B. C., & O'Meara, B. C. (2017). PHRAPL:
- Phylogeographic Inference Using Approximate Likelihoods. Systematic Biology, 66(6),
 1045–1053. doi: 10.1093/SYSBIO/SYX001
- Kass, R. E., & Raftery, A. E. (1995). Bayes Factors. Journal of the American Statistical
 Association, 90(430), 773–795. doi: 10.1080/01621459.1995.10476572
- Knaus, B. J., & Grünwald, N. J. (2017). vcfr: A package to manipulate and visualize variant
 call format data in R. Molecular Ecology Resources, 17(1), 44–53. doi: 10.1111/17550998.12549
- 787 Kornilios, P., Thanou, E., Lymberakis, P., Ilgaz, Ç., Kumlutaş, Y., & Leaché, A. (2019).

788 Genome-wide markers untangle the green-lizard radiation in the Aegean Sea and

- support a rare biogeographical pattern. Journal of Biogeography, 46(3), 552–567. doi:
 10.1111/jbi.13524
- Kozlov, A. M., Darriba, D., Flouri, T., Morel, B., & Stamatakis, A. (2019). RAxML-NG: a
 fast, scalable and user-friendly tool for maximum likelihood phylogenetic inference.

793 Bioinformatics, 35(21), 4453–4455. doi: 10.1093/BIOINFORMATICS/BTZ305

- Leaché, A. D., Zhu, T., Rannala, B., & Yang, Z. (2019). The Spectre of Too Many Species.
 Systematic Biology, 68(1), 168–181. doi: 10.1093/SYSBIO/SYY051
- 796 Main, D. C., van Vuuren, B. J., Tilbury, C. R., & Tolley, K. A. (2022). Out of southern
- Africa: Origins and cryptic speciation in *Chamaeleo*, the most widespread chameleon
 genus. Molecular Phylogenetics and Evolution, 175, 107578. doi:
- 799 10.1016/J.YMPEV.2022.107578
- Malinsky, M., Matschiner, M., & Svardal, H. (2021). Dsuite—Fast D-statistics and related
 admixture evidence from VCF files. Molecular Ecology Resources, 21(2), 584–595.
 doi: 10.1111/1755-0998.13265
- 803 Marshall, T. L., Chambers, E. A., Matz, M. V., & Hillis, D. M. (2021). How mitonuclear
- 804 discordance and geographic variation have confounded species boundaries in a widely
- studied snake. Molecular Phylogenetics and Evolution, 162, 107194. doi:
- 806 10.1016/J.YMPEV.2021.107194

- Mayden, R. L. (1997). A hierarchy of species concepts: The denouement in the saga of the
 species problem (Claridge M.F, Dawah H.A, & Wilson M.R., Eds.). Retrieved from
 https://philpapers.org/rec/MAYAHO-6
- 810 Metallinou, M., & Carranza, S. (2013). New species of *Stenodactylus* (Squamata:
- 811 Gekkonidae) from the Sharqiyah Sands in northeastern Oman. Zootaxa, 3745, 449–468.
- 812 doi: 10.11646/zootaxa.3745.4.3
- 813 Metallinou, M., Červenka, J., Crochet, P. A., Kratochvíl, L., Wilms, T., Geniez, P., ...
- 814 Carranza, S. (2015). Species on the rocks: Systematics and biogeography of the rock-
- 815 dwelling *Ptyodactylus geckos* (Squamata: Phyllodactylidae) in North Africa and Arabia.
- 816 Molecular Phylogenetics and Evolution, 85, 208–220. doi:
- 817 10.1016/j.ympev.2015.02.010
- Nguyen, L.-T., Schmidt, H. A., von Haeseler, A., & Minh, B. Q. (2015). IQ-TREE: A Fast
 and Effective Stochastic Algorithm for Estimating Maximum-Likelihood Phylogenies.
 Molecular Biology and Evolution, 32(1), 268–274. doi: 10.1093/molbev/msu300
- O'Leary, S. J., Puritz, J. B., Willis, S. C., Hollenbeck, C. M., & Portnoy, D. S. (2018). These
 aren't the loci you'e looking for: Principles of effective SNP filtering for molecular
 ecologists. Molecular Ecology, 27(16), 3193–3206. doi: 10.1111/MEC.14792
- 824 Peterson, B. K., Weber, J. N., Kay, E. H., Fisher, H. S., & Hoekstra, H. E. (2012). Double
- Bigest RADseq: An Inexpensive Method for De Novo SNP Discovery and Genotyping
 in Model and Non-Model Species. PLOS ONE, 7(5), e37135. doi:
- 827 10.1371/JOURNAL.PONE.0037135
- 828 Pons, J., Barraclough, T. G., Gomez-Zurita, J., Cardoso, A., Duran, D. P., Hazell, S., ...
- 829 Vogler, A. P. (2006). Sequence-Based Species Delimitation for the DNA Taxonomy of
- 830 Undescribed Insects. Systematic Biology, 55(4), 595–609. doi:
- 831 10.1080/10635150600852011
- Radies, D., Preusser, F., Matter, A., & Mange, M. (2004). Eustatic and climatic controls on
- the development of the Wahiba Sand Sea, Sultanate of Oman. Sedimentology, 51(6),
 1359–1385. doi: 10.1111/j.1365-3091.2004.00678.x
- Rambaut, A., & Drummond, A. J. (2013). Tracer v1. 5 Available from http://beast. Bio. Ed.
 Ac. Uk/Tracer. Accessed.

837 Rannala, B., & Yang, Z. (2003). Bayes Estimation of Species Divergence Times and

838 Ancestral Population Sizes Using DNA Sequences From Multiple Loci. Genetics,

839 164(4), 1645–1656. doi: 10.1093/GENETICS/164.4.1645

- Rannala, B., & Yang, Z. (2013). Improved Reversible Jump Algorithms for Bayesian Species
 Delimitation. Genetics, 194(1), 245–253. doi: 10.1534/genetics.112.149039
- 842 Riaño, G., Fontsere, C., Manuel, M. de, Talavera, A., Burriel-Carranza, B., Tejero-Cicuéndez,
- H., ... Carranza, S. (2022). Genomics reveals introgression and purging of deleterious
 mutations in the Arabian leopard (*Panthera pardus nimr*). p. 2022.11.08.515636.
 bioRxiv. doi: 10.1101/2022.11.08.515636
- 846 Shults, P., Hopken, M., Eyer, P. A., Blumenfeld, A., Mateos, M., Cohnstaedt, L. W., &
- 847 Vargo, E. L. (2022). Species delimitation and mitonuclear discordance within a species
- complex of biting midges. Scientific Reports 2022 12:1, 12(1), 1–13. doi:
- 849 10.1038/s41598-022-05856-x
- 850 Simó-Riudalbas, M., Pous, P. de, Els, J., Jayasinghe, S., Péntek-Zakar, E., Wilms, T., ...
- 851 Carranza, S. (2017). Cryptic diversity in *Ptyodactylus* (reptilia: Gekkonidae) from the
 852 northern hajar mountains of Oman and the United Arab Emirates uncovered by an
- 853 integrative taxonomic approach. PLoS ONE, 12(8), 1–25. doi:
- 854 10.1371/journal.pone.0180397
- 855 Simó-Riudalbas, M., Tarroso, P., Papenfuss, T., Al-Sariri, T., & Carranza, S. (2018).
- 856 Systematics, biogeography and evolution of *Asaccus gallagheri* (Squamata,
- 857 Phyllodactylidae) with the description of a new endemic species from Oman.
- 858 Systematics and Biodiversity, 16(4), 323–339. doi: 10.1080/14772000.2017.1403496
- 859 Šmíd, J., Sindaco, R., Shobrak, M., Busais, S., Tamar, K., Aghová, T., ... Carranza, S. (2021).
- 860 Diversity patterns and evolutionary history of Arabian squamates. Journal of

861 Biogeography, 48(5), 1183–1199. doi: 10.1111/jbi.14070

862 Stange, M., Sánchez-Villagra, M. R., Salzburger, W., & Matschiner, M. (2018). Bayesian

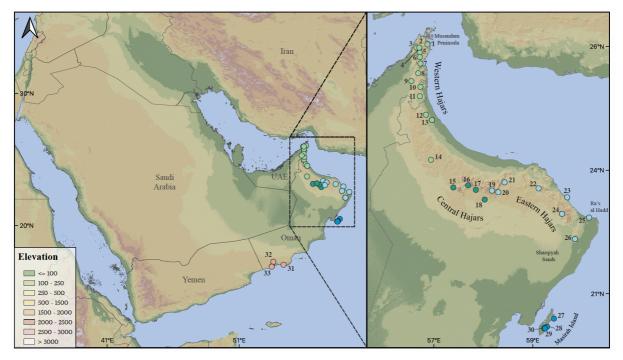
- 863 Divergence-Time Estimation with Genome-Wide Single-Nucleotide Polymorphism
- Bata of Sea Catfishes (Ariidae) Supports Miocene Closure of the Panamanian Isthmus.
 Systematic Biology, 67(4), 681–699. doi: 10.1093/SYSBIO/SYY006
- 866 Sukumaran, J., & Knowles, L. L. (2017). Multispecies coalescent delimits structure, not
- species. Proceedings of the National Academy of Sciences, 114(7), 1607–1612. doi:
- 868 10.1073/PNAS.1607921114

869	Tamar, K., Chirio, L., Shobrak, M., Busais, S., & Carranza, S. (2019). Using multilocus
870	approach to uncover cryptic diversity within Pseudotrapelus lizards from Saudi Arabia.
871	Saudi Journal of Biological Sciences, 26(7), 1442-1449. doi:
872	10.1016/j.sjbs.2019.05.006
873	Tamar, K., Mitsi, P., & Carranza, S. (2019). Cryptic diversity revealed in the leaf-toed gecko
874	Asaccus montanus (Squamata, Phyllodactylidae) from the Hajar Mountains of Arabia.
875	Journal of Zoological Systematics and Evolutionary Research, 57(2), 369-382. doi:
876	10.1111/jzs.12258
877	Tejero-Cicuéndez, H., Patton, A. H., Caetano, D. S., Šmíd, J., Harmon, L. J., & Carranza, S.
878	(2022). Reconstructing Squamate Biogeography in Afro-Arabia Reveals the Influence
879	of a Complex and Dynamic Geologic Past. Systematic Biology, 71(2), 261-272. doi:
880	10.1093/SYSBIO/SYAB025
881	Thanou, E., Kornilios, P., Lymberakis, P., & Leaché, A. D. (2020). Genomic and
882	mitochondrial evidence of ancient isolations and extreme introgression in the four-lined
883	snake. Current Zoology, 66(1), 99-111. doi: 10.1093/cz/zoz018
884	Tonzo, V., Papadopoulou, A., & Ortego, J. (2019). Genomic data reveal deep genetic
885	structure but no support for current taxonomic designation in a grasshopper species
886	complex. Molecular Ecology, 28(17), 3869–3886. doi: 10.1111/mec.15189
887	Vences, M., Multzsch, M., Gippner, S., Miralles, A., Crottini, A., Gehring, PS., Scherz,
888	M. D. (2022). Integrative revision of the Lygodactylus madagascariensis group reveals
889	an unexpected diversity of little brown geckos in Madagascar's rainforest. Zootaxa,
890	5179(1), 1-61. doi: 10.11646/ZOOTAXA.5179.1.1
891	Vilaça, S. T., Piccinno, R., Rota-Stabelli, O., Gabrielli, M., Benazzo, A., Matschiner, M.,
892	Bertorelle, G. (2021). Divergence and hybridization in sea turtles: Inferences from
893	genome data show evidence of ancient gene flow between species. Molecular Ecology,
894	30(23), 6178–6192. doi: 10.1111/MEC.16113
895	Walters, A. D., Cannizzaro, A. G., Trujillo, D. A., & Berg, D. J. (2021). Addressing the
896	Linnean shortfall in a cryptic species complex. Zoological Journal of the Linnean
897	Society, 192(2), 277-305. doi: 10.1093/ZOOLINNEAN/ZLAA099
898	Wong, T. K., Cherryh, C., Rodrigo, A. G., Hahn, M. W., Minh, B. Q., & Lanfear, R. (2022,
899	October 8). MAST: Phylogenetic Inference with Mixtures Across Sites and Trees (p.

900 2022.10.06.511210). p. 2022.10.06.511210. bioRxiv. doi: 10.1101/2022.10.06.511210

- 901 Yang, Z., & Rannala, B. (2010). Bayesian species delimitation using multilocus sequence
- 902 data. Proceedings of the National Academy of Sciences, 107(20), 9264–9269. doi:
- 903 10.1073/pnas.0913022107
- 904 Zhang, C., Sayyari, E., & Mirarab, S. (2017). ASTRAL-III: Increased scalability and impacts
- 905 of contracting low support branches. Lecture Notes in Computer Science (Including
- 906 Subseries Lecture Notes in Artificial Intelligence and Lecture Notes in Bioinformatics),
- 907 10562 LNBI, 53–75. doi: 10.1007/978-3-319-67979-2_4/TABLES/3
- 908

909 FIGURES





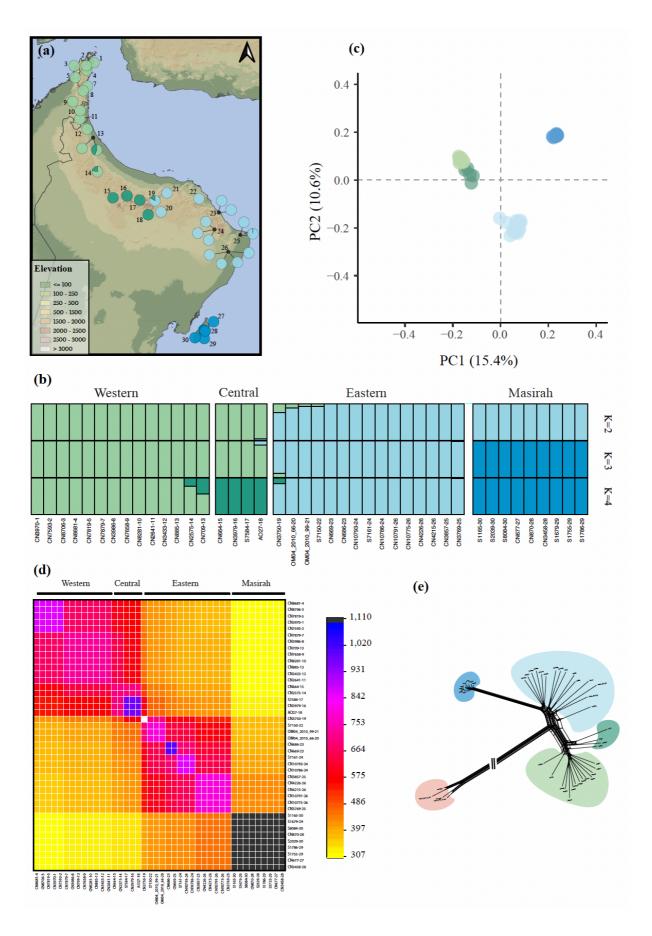
911 Figure 1. Topographic map of the study area showing the 33 localities of *Trachydactylus*

sampled for this study (see Table S1 for correspondece between the 33 localities and the 53

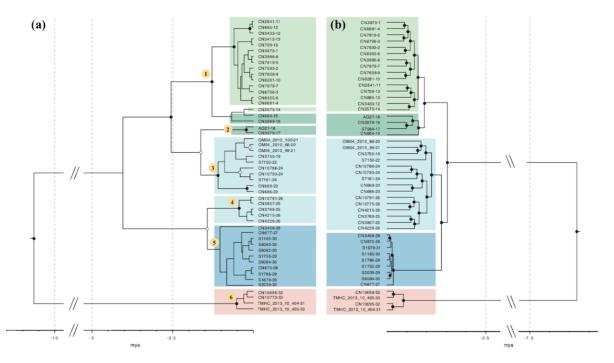
913 sampled specimens); colors represent T. spatalurus (pink) and T. hajarensis Western (light

green), Central (dark green), Eastern (light blue) and Masirah Island (dark blue) lineages

915 according to our ddRADseq genomic analyses.



918 Figure 2. Population structure analyses of *T. hajarensis*. a: Geographic distribution of the 919 different individuals analyzed, with pies representing the percentage of admixture for K=3; b: 920 individual assignments of ADMIXTURE nuclear clusters (K=2-4); Numbers after last dash 921 correspond to locality number from Figure 1 and Table S1; c: Principal Component Analysis 922 (PCA), with points representing the different individuals colored by taxa; d: Co-ancestry 923 matrix generated in fineRADstructure, indicating pairwise genetic similarity between all T. 924 hajarensis specimens. To the right, legend shows the number of shared alleles between 925 individuals. Darker colors representent higher inter-individual co-ancestry; e: Phylogenomic 926 network constructed with the Neighbor-Net algorithm in SplitsTree. T. spatalurus (pink); T. 927 hajarensis Western (light green), Central (dark green), Eastern (light blue) and Masirah Island 928 (dark blue) lineages. Datasets contain 2,428 unlinked SNPs (a,b & c; dataset 1), 30,526 loci 929 (d; dataset 2) and 5,219 loci (e; dataset 3). See table S2 for further specifications on each 930 dataset.





933 Figure 3. Bayesian inference time calibrated trees. a: Phylogenetic tree based on the 12S

934 mitochondrial gene. Numbers in yellow circles correspond to GMYC species level

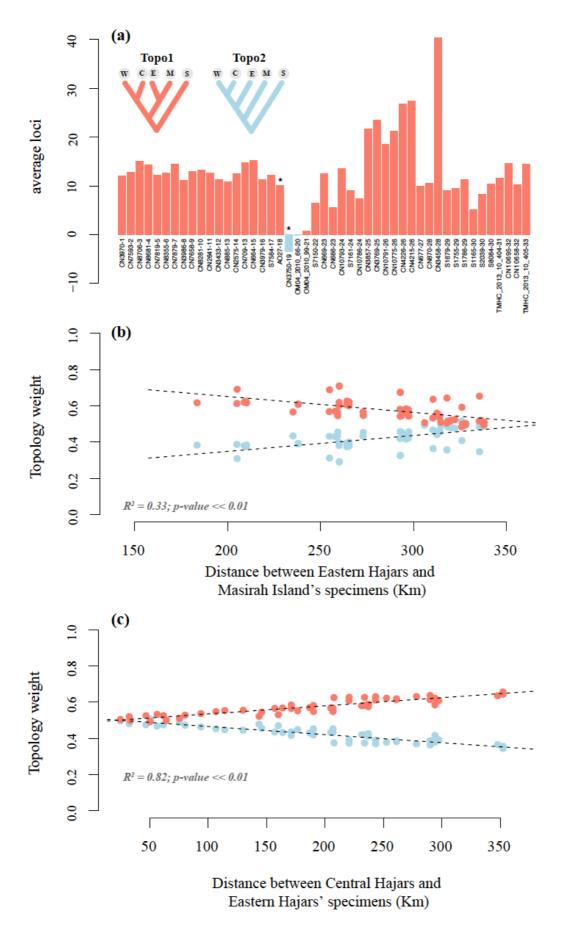
935 assignment; **b**: Phylogenomic tree inferred with a concatenated dataset of 5,219 nuclear loci.

Posterior probability (pp) above 0.95 and pp > 0.85 are shown with black and white dots at

937 each node respectively; Numbers after last dash correspond to locality number in Figure 1 and

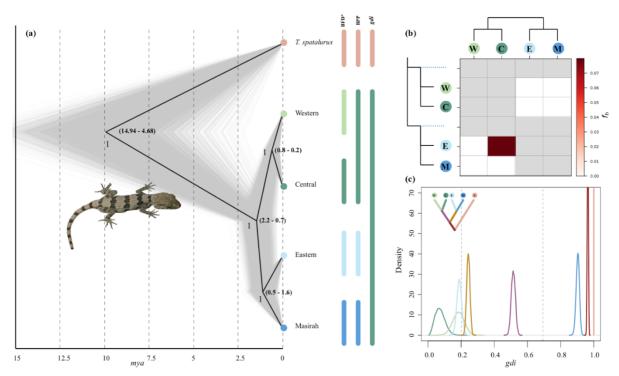
938 Table S1; T. spatalurus (pink), T. hajarensis Western (light green), Central (dark green),

939 Eastern (light blue) and Masirah Island (dark blue) lineages.



942 Figure 4. Species tree topology testing. a: Number of loci averaged across all quintets that

- 943 support each topology (shown as *Topo1-Topo2*). Numbers after last dash correspond to
- 944 locality number from Figure 1. Asterisks highlight admixed specimens in Figure 2a. T.
- 945 spatalurus (S), T. hajarensis Western (W), Central (C), Eastern (E), and Masirah Island (M)
- 946 lineages; **b**: Shifts in *Topo1* (salmon) and *Topo2* (blue) topology weights explained by the
- 947 distance from each Eastern Hajars' specimen to each Masirah Island's specimen; c: Shifts in
- 948 *Topol* (salmon) and *Topo2* (blue) topology weights explained by the distance from each
- 949 Eastern Hajars' specimen to each Central Hajars' individual. Each point corresponds to the
- 950 averaged topology weight across all quintet combinations with the same Eastern and Masirah
- 951 specimens (see Figure S4). The sum of topology weights (*Topol* and *Topo2*) of each MAST
- analysis always adds up to 1.0, thus plotted regression lines have inverse directions but
- 953 identical R^2 and *p*-values.

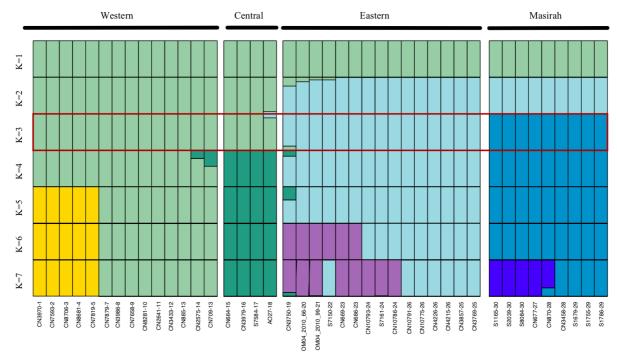




956 Figure 5. Species tree, gene flow and species delimitation analyses. a: Most supported species 957 tree topology between SNAPP, BPP and IQ-TREE analyses. Species tree was inferred with 958 SNAPP from *dataset 5 SNAPP* and both posterior (grey) and consensus trees (black) are 959 shown. To the right, different coloured bars represent BFD*, BPP and gdi species level 960 assignment; **b**: F-branch statistic analysis showing introgression between *Trachydactylus* 961 hajarensis Eastern and Central clades; c: Posterior distribution for the gdi values between 962 every pair of sister taxa within Trachydactylus hajarensis, including T. spatalurus. Colors in 963 the internal branches of the upper-left corner species tree represent subsequent A00 BPP 964 analyses where the descendant tips of the branch were lumped together and compared to its 965 closest sister group. Values below 0.2 support a single species hypothesis while above 0.7 966 supports distinct species status. 967

- 0.00
- 968





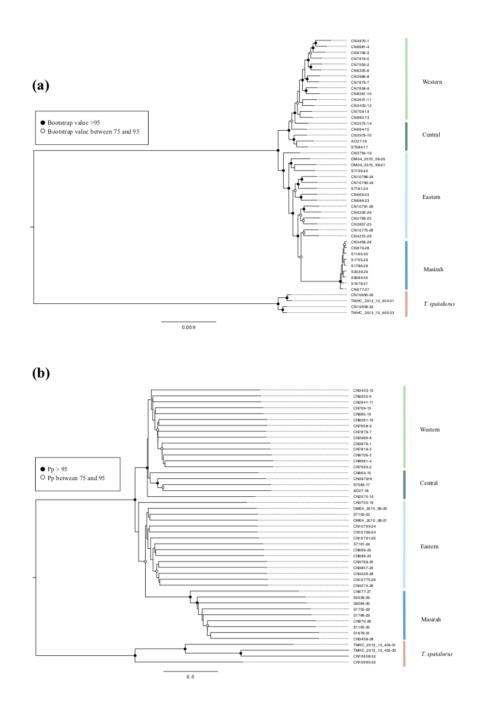


971 Figure S1. Individual assignments of ADMIXTURE nuclear clusters from K = 1 to K = 7.

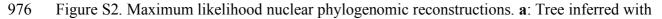
972 The lowest cross validation result, depicting the most probable number of populations, is

973 highlighted in red. Numbers after last dash correspond to locality number from Figure 1 and

Table S1. Information on all the specimens analyzed can be found in Table S1.



975

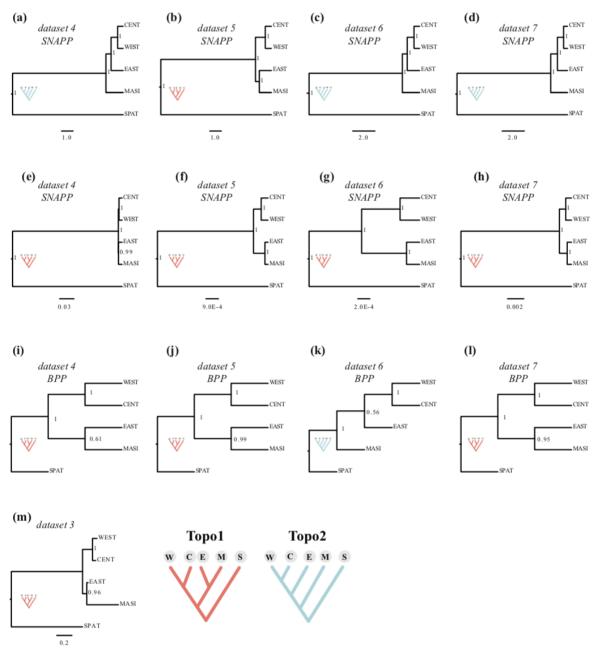


977 RAxML-NG with a concatenated dataset of 5,219 loci; b: Consensus tree inferred with Astral

978 after computing 5,219 gene trees in IQ-TREE. Western, Central, Eastern and Masirah

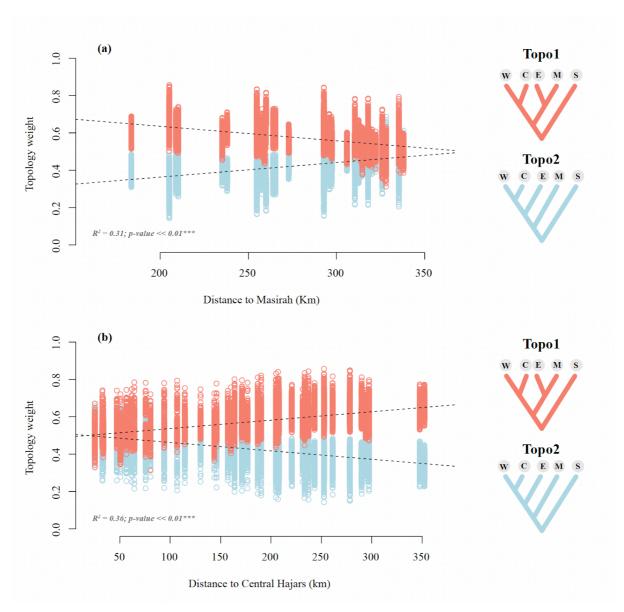
979 represent the different *Trachydactylus hajarensis* clades. Numbers after last dash correspond

980 to locality number from Figure 1 and Table S1.





982 Figure S3. Species tree topologies with different sample selection. Samples included in each 983 analysis range between 10–20 specimens in *datasets* 4–7, and 47 samples (all available samples) in dataset 3. For more information on each dataset see Table S2. Species trees were 984 985 inferred with SNAPP linking all population sizes (a–d), with SNAPP without linking 986 population sizes (e-h), with BPP A01 approach (i-l), and with Astral after summarizing all 987 5,219 gene trees obtained with IQtree (m). Abbreviations are as follow: WEST, Western 988 Clade of T. hajarensis; CENT, Central Clade of T. hajarensis; EAST, Eastern Clade of T. 989 hajarensis; MASI, Masirah Island Clade of T. hajarensis; SPAT, T. spatalurus.



991 Figure S4. a: Shifts in *Topol* (salmon) and *Topo2* (blue) topology weights explained by the 992 distance from each Eastern Hajars' specimen to each Masirah Island's specimen; b: Shifts in 993 *Topol* (salmon) and *Topo2* (blue) topology weights explained by the distance from each 994 Eastern Hajars' specimen to each Central Hajars' individual. Each point represents the 995 averaged topology weight of a quintet across all its loci. Points distributed across vertical 996 lines correspond to quintet combinations where Eastern and Masirah Island specimens 997 remained identical but the individuals from the other lineages varied. We accounted for such 998 pseudoreplicates by averaging each locality's topology weight (Figure 3), or by including 999 them into a nested ANOVA (Tables S3 and S4). The sum of topology weights (Topol and 1000 *Topo2*) of each MAST analysis always adds up to 1.0, thus plotted regression lines have 1001 inverse directions but identical R² and p-values.

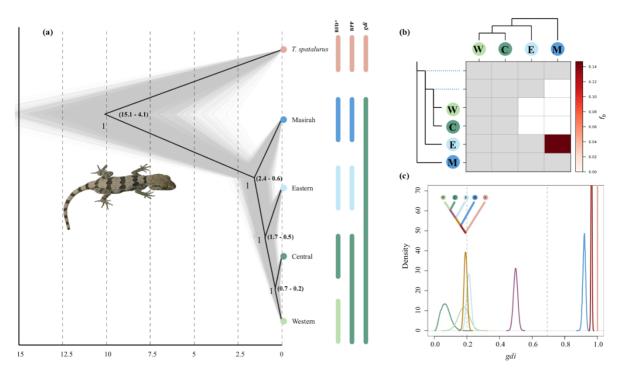


Figure S5. Species tree, gene flow and species delimitation analyses. a: Second species tree 1003 topology obtained with SNAPP & BPP. Species tree was inferred with SNAPP from dataset 4 1004 1005 SNAPP and posterior trees (grey) and consensus tree (black) are shown. To the right, different 1006 coloured bars represent BFD*, BPP and gdi species level assignment, respectively; b: F-1007 branch statistic analysis showing introgression between Trachydactylus hajarensis Eastern and Masirah Island clades; c: Posterior distribution for the gdi values between every pair of 1008 1009 sister taxa within Trachydactylus hajarensis, including T. spatalurus. Colors in the internal 1010 branches of the upper-left corner species tree represent subsequent A00 BPP analyses where 1011 the descendant tips of the branch were lumped together and compared to its closest sister 1012 group. Values below 0.2 support a single species hypothesis while above 0.7 supports distinct 1013 species status.

1014

1015 SUPPLEMENTARY TABLES

Species	Specimen code	Locality	Country	Latitude	Longitude	Clade	12S Accession nº	Raw reads	Filtered reads	Post-processing	dataset 1	dataset 2	dataset 3	dataset 4	dataset 5	dataset 6	dataset 7
T. hajarensis	CN3970	1	Oman	26.042	56.37	Western	KT302084	1,191,821	1,189,784	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN7593	2	Oman	25.978	56.205	Western	XXX	2,829,981	2,820,714	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN8706	3	UAE	25.976	56.15	Western	KT302083	932,759	930,372	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN8681	4	Oman	25.957	56.203	Western	XXX	5,840,800	5,823,408	YES	YES	YES	YES	YES	YES	YES	YES
T. hajarensis	CN7819	5	Oman	25.88	56.214	Western	KT302080	1,583,871	1,580,461	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN8355	6	Oman	25.786	56.217	Western	XXX	416,106	415,073	YES	NO	NO	YES	NO	NO	NO	NO
T. hajarensis	CN7879	7	Oman	25.656	56.229	Western	XXX	1,338,816	1,335,631	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN3986	8	UAE	25.459	56.183	Western	KT302078	1,515,929	1,511,634	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	CN7658	9	UAE	25.3	56.045	Western	KT302079	4,384,451	4,369,626	YES	YES	YES	YES	YES	YES	YES	YES
T. hajarensis	CN8281	10	UAE	25.182	56.229	Western	KT302081	7,996,636	7,984,850	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN2641	11	UAE	24.994	56.217	Western	XXX	1,283,653	1,281,350	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN3433	12	Oman	24.621	56.34	Western	XXX	568,169	566,799	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN3412	13	Oman	24.513	56.463	Western	XXX	229,466	229,09	NO	NO	NO	NO	NO	NO	NO	NO
T. hajarensis	CN885	13	Oman	24.513	56.463	Western	XXX	506,362	505,264	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN709	13	Oman	24.513	56.463	Western	XXX	1,109,312	1,106,557	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	CN2575	14	Oman	23.71	56.443	Western	XXX	332,487	331,462	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN664	15	Oman	23.15	56.894	Central	XXX	1,600,422	1,597,354	YES	YES	YES	YES	YES	NO	NO	YES
T. hajarensis	CN3979	16	Oman	23.193	57.196	Central	XXX	3,993,190	3,983,810	YES	YES	YES	YES	YES	YES	YES	YES
T. hajarensis	S7584	17	Oman	23.101	57.35	Central	-	1,713,240	1,708,957	YES	YES	YES	YES	NO	YES	YES	YES
T. hajarensis	AO27	18	Oman	22.905	57.53	Central	KT302062	1,775,777	1,769,136	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	CN3750	19	Oman	23.087	57.676	Eastern	XXX	242,966	242,194	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	OM04_2010_66	20	Oman	23.053	57.805	Eastern	XXX	1,081,611	1,079,368	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	OM04_2010_99	21	Oman	23.254	57.931	Eastern	KT302069	2,058,902	2,056,273	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	OM04_2010_100	21	Oman	23.254	57.931	Eastern	XXX	21,346	21,272	NO	NO	NO	NO	NO	NO	NO	NO
T. hajarensis	\$7150	22	Oman	23.132	58.619	Eastern	KT302066	3,949,614	3,938,428	YES	YES	YES	YES	NO	NO	YES	YES
T. hajarensis	CN686	23	Oman	22.949	59.198	Eastern	KT302064	4,623,864	4,611,556	YES	YES	YES	YES	YES	NO	YES	NO
T. hajarensis	CN669	23	Oman	22.95	59.198	Eastern	XXX	4,842,487	4,829,145	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	S7161	24	Oman	22.616	59.094	Eastern	KT302067	6,787,797	6,777,425	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	CN10786	24	Oman	22.616	59.094	Eastern	XXX	477,79	476,704	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN10793	24	Oman	22.616	59.094	Eastern	XXX	743,334	742,037	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN3857	25	Oman	22.541	59.641	Eastern	XXX	936,911	934,448	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN3769	25	Oman	22.541	59.642	Eastern	XXX	558,198	556,537	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN4226	26	Oman	22.107	59.357	Eastern	KT302089	9,600,663	9,587,672	YES	YES	YES	YES	YES	NO	NO	YES
T. hajarensis	CN4215	26	Oman	22.107	59.357	Eastern	XXX	726,547	725,041	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	CN10775	26	Oman	22.107	59.357	Eastern	-	4,403,945	4,394,347	YES	YES	YES	YES	NO	YES	NO	YES
T. hajarensis	CN10791	26	Oman	22.107	59.357	Eastern	XXX	5,603,416	5,593,156	YES	YES	YES	YES	NO	YES	NO	NO
T. hajarensis	CN677	27	Oman	20.498	58.931	Masirah	KT302085	2,521,949	2,519,348	YES	YES	YES	YES	YES	YES	YES	YES
T. hajarensis	CN3458	28	Oman	20.334	58.789	Masirah	XXX	3,505,223	3,498,383	YES	YES	YES	YES	YES	YES	YES	YES
T. hajarensis	CN870	28	Oman	20.333	58.789	Masirah	XXX	1,303,376	1,300,819	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	81755	29	Oman	20.335	58.737	Masirah	XXX	943,7	941,956	YES	YES	YES	YES	NO	NO	NO	NO

Table S1. Table of all specimens included in this study information regarding location, clade, GenBank accession number, ddRADseq raw and filtered reads as well as information on which individuals are included in each dataset. GenBank accession numbers labled XXXX were de novo produced and will be uploaded; datasets 4-7 include both SNAPP and BPP datasets in Table S2.

T. hajarensis	S1679	29	Oman	20.312	58.737	Masirah	XXX	1,192,372	1,189,767	YES	YES	YES	YES	NO	NO	NO	YES
T. hajarensis	S1786	29	Oman	20.312	58.737	Masirah	XXX	924,669	922,689	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	S8082	30	Oman	20.299	58.74	Masirah	XXX	163,003	162,547	NO							
T. hajarensis	S8083	30	Oman	20.299	58.74	Masirah	XXX	27,807	27,751	NO							
T. hajarensis	S8084	30	Oman	20.299	58.74	Masirah	XXX	194,522	194,234	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	S1165	30	Oman	20.299	58.75	Masirah	XXX	597,319	592,252	YES	YES	YES	YES	NO	NO	NO	NO
T. hajarensis	S2039	30	Oman	20.299	58.75	Masirah	XXX	1,507,855	1,505,424	YES	YES	YES	YES	NO	NO	NO	NO
T. spatalurus	TMHC_2013_10_404	31	Oman	17.029	54.666	SPAT	KT302092	2,732,990	2,726,923	YES	YES	YES	YES	NO	NO	NO	YES
T. spatalurus	CN10773	32	Oman	17.252	53.891	SPAT	XXX	107,794	107,492	NO							
T. spatalurus	CN10658	32	Oman	17.252	53.891	SPAT	-	7,878,774	7,851,814	YES							
T. spatalurus	CN10695	33	Oman	16.899	53.772	SPAT	XXX	575,237	573,924	YES	YES	YES	YES	NO	NO	NO	YES
T. spatalurus	TMHC_2013_10_405	33	Oman	16.884	53.773	SPAT	KT302093	3,841,169	3,828,560	YES							

dataset name	dataset type	Analysis	Postprocessing filtering (%)	n° of individuals	dataset length (bp)	n° of loci	n° of SNPs	Missingness (%)
dataset 1	uSNPs	AdMIXTURE & PCA	78 ind - 40 gen	42	2,428	2,428	2,428	18.57
dataset 2	loci	fineRAdstructure	Not filtered	42	1,973,898	30,526	99,828	63.98
dataset 3	loci	raxml-ng, IQTREE, BEAST, SplitsTree, MAST & dsuite	78 ind - 40 gen	47	338,507	5,219	30,096	22.21
dataset 4 SNAPP	uSNPs	Species tree inference and Species delimitation	10 gen	10	2,147	2,147	2,147	4.23
dataset 5 SNAPP	uSNPs	Species tree inference	10 gen	10	3,210	3,210	3,210	4.88
dataset 6 SNAPP	uSNPs	Species tree inference	10 gen	10	2,897	2,897	2,897	4.84
dataset 7 SNAPP	uSNPs	Species tree inference	10 gen	20	2,731	2,731	2,731	9.72
dataset 4 BPP	loci	Species tree inference, Species delimitation and gdi	60 gen	10	261,420	4,357	20,246	15.54
dataset 5 BPP	loci	Species tree inference	60 gen	10	271,980	4,533	19,966	13.7
dataset 6 BPP	loci	Species tree inference	60 gen	10	266,460	4,441	20,262	15.3
dataset 7 BPP	loci	Species tree inference	60 gen	20	315,240	5,254	26,103	27.2

Table S2: dataset specifications. uSNPs: Unlinked Single Nucleotide Polymorphisms; Ind: Missing data allowed per individual; gen: Missing genotype call rate allowed; bp: Base pairs.

Table S4. Nested analysis of variance table. Response variable shown in this nested ANOVA correspond to topology weigths 1. However, the exact same results were obtanied when using topology weigths 2. Dist ME: Distance between Masirah and Eastern specimens

Response Topo1							
	DF	Sum_sq	Mean_Sq	F value	Pr(>F)		
Dist ME	1	45.892	45.89	11,399.25	<< 0.01		
Dist ME:replicate	3,359	13.523	0.004	1,585	<< 0.01		
Residuals	34,439	87.472	0.003				

Table S4. Nested analysis of variance table. Response variable shown in this nested ANOVA correspond to topology weigths 1. However, the exact same results were obtanied when using topology weigths 2. Dist CE: Distance between Central and Eastern specimens

Response Topo1					
	DF	Sum_sq	Mean_Sq	F value	Pr(>F)
Dist CE	1	53.144	53.144	3,801.42	<< 0.01
Dist CE:replicate	2,015	28.17	0.014	7.629	<< 0.01
Residuals	35,783	65.572	0.002		

Table S5: Results of BFD* testing the support of the competing species hypotheses. For each hypothesis we show Maximum Likelihood Estimates (MLE), their Bayes Factors ($2 \times (H0 - Hi)$), the rank and the strength of support ln(BF). Additionally, we show the specimens assigned to each putative species hypothesis.

Species delimitation hypothesis	MLE	BF	Rank	strength of support
H0: T. hajarensis vs T. spatalurus	12979,10	0	4	-
H1: T. spatalurus vs Western+Central vs Masirah + Eastern clades	12110,69	1736,82	3	7,46
H2: T. spatalurus vs Western + Central vs Eastern vs Masirah clades	11515,77	2926,65	2	7,98
H3: T. spatalurus vs Western vs Central vs Eastern vs Masirah clades	11438,92	3080,36	1	8,03

	Specimen assignment
Western:	CN7658, CN8681
Central:	CN3979, CN664
Eastern:	CN4226, CN686
Masirah:	CN3458, CN677
T. spatalurus:	CN10658, TMHC_2013_10_405