Lifestyle and not density of fish hosts determines parasite distribution over time and

space

Nikol Kmentová[1], Armando J. Cruz-Laufer[1], Leona Milec[1,2], Tanisha Moons[1,3], Senne

Heeren[4,5], Elze van den Hoorn[6], Kelly J. M. Thys[1], Lawrence Makasa[7], Auguste Chocha

Manda[8], Pascal Masilya Mulungula[9], Maarten Van Steenberge[1,10], Michiel W. P. Jorissen[1,]

Maarten P. M. Vanhove[1]

[1] Hasselt University, Faculty of Sciences, Centre for Environmental Sciences, Research

Group Zoology: Biodiversity and Toxicology, Agoralaan Gebouw D, 3590 Diepenbeek,

Belgium.

[2] Faculty of Biosciences and Aquaculture, Nord University, Universitetsalléen 11, 8026,

Bodø, Norway

[3] Department of Parasitology, Faculty of Science, University of South Bohemia, České

Budějovice, Czech Republic

[4] Department of Biomedical Sciences, Institute of Tropical Medicine, 2000 Antwerp, Belgium

[5] Department of Microbiology, Immunology and Transplantation, Rega Institute for Medical

Research, Katholieke Universiteit Leuven, 3000 Leuven Belgium

[6] HAS Green Academy, Applied Biology, Onderwijsboulevard 221, 5223 DE ´s-

Hertogenbosch

[7] Lake Tanganyika Research Unit, Department of Fisheries, Ministry of Fisheries and

Livestock, P. O. Box 420055, Mpulungu, Zambia

[8] Unité de recherche en Biodiversité et Exploitation durable des Zones Humides (BEZHU),

Faculté des Sciences Agronomiques, Université de Lubumbashi, Haut-Katanga, R.D. Congo

[9] Unité d'Enseignement et de Recherche en Hydrobiologie Appliquée (UERHA), Department

of Biology, ISP-Bukavu

[10] Operational Directorate Taxonomy and Phylogeny, Royal Belgian Institute of Natural

Sciences, Vautierstraat 29, B-1000 Brussels, Belgium

*corresponding author: nikol.kmentova@uhasselt.be

Abstract

1. Pelagic zones are characterised by consistent large-scale patterns of circulation on

local and global scales. Lake Tanganyika, as an enclosed ecosystem provides a

playing field to study evolutionary and ecological processes applicable to open water

areas worldwide. Despite their important role in the ecosystem, large scale patterns

of the distribution of parasites remain poorly understood. Monogenean parasites

have been proposed as tags for ecosystem dynamics because of their direct life

cycle and often high host specificity. We combined data on spatiotemporal dynamics

of parasites (*Kapentagyrus*, Monogenea) with morphological variation of clupeid

hosts to investigate general patterns of host-parasite interactions in the pelagic zone

of this ancient lake.

2. Two dominant species of clupeid fishes in the pelagic zone of Lake Tanganyika are

parasitised by two monogenean species of *Kapentagyrus* with different levels of host

specificity. The host fishes are believed to perform lake-wide migrations driven by

seasonal upwellings. To model spatiotemporal dynamics of host-parasite interaction,

we used temporal data on monogenean infection along the North-South axis of Lake

Tanganyika based on 1730 screened fishes and 3710 parasites together with

phenotypic characterisation of the clupeid hosts.

3. Infection levels are dependent on host body size with contrasting trends in the two

parasite species. We reveal temporal stability of infection with spatial distribution

restricted by host life strategies. Spatial differences between the parasite species

most likely reflect differences in migration between the host species; these are also

reflected by morphological differences between some clupeid populations. Our

results also suggest mutual facilitation of infection.

4. In conclusion, parasite infection is geographically restricted by host life strategies

even in this ecosystem lacking physical barriers. Intensity of infection seems to be

mainly driven by host size, with so far rarely reported mutual facilitation of co-

infection. Temporal stability of infection in the pelagic zone contrasts with seasonal

changes in abundance of clupeid hosts. Overall, our results suggest that parasite

infection dynamics in the open water areas are dependent on life style and not host

density.

Keywords

Lake Tanganyika, parasitic flatworms*,* clupeids, migration, population dynamics

Introduction

Trophic and host-parasite interactions in pelagic systems

The pelagic zone, also known as the open water area, represents 99% of the volume of the

biosphere. Pelagic zones are characterised by large-scale patterns of circulation and

upwelling processes (Corman et al., 2010) that support the majority of the biomass targeted

by fisheries (Pauly et al., 2002). Species aggregation is common at all trophic levels and is

driven by the patchiness of primary production (Legendre & Le Fèvre, 1991; Ritz et al.,

2011). Migration is a common strategy of pelagic species to cope with seasonal variation of

resources at both small and large geographic scales (Angel, 1993). Small pelagic fishes

(clupeids, mackerels) are known for their schooling behaviour and large distribution ranges

often linked to their long-distance migrations (Teske et al., 2021). Therefore, small pelagic

fishes have been proposed as indicator taxa for ecosystem changes related to e.g., climate

change and overfishing. However, long-distance migration and population dynamics of small

pelagic fish species are challenging to study, due to their high mobility and fragility that

restrict traditional tracking methods.

Metazoan parasites encompass a high proportion of global species diversity (Poulin, 2014;

Windsor, 1998). Despite a largely negative perception, parasites have been recognised as

ecosystem engineers that form substantial biomass in aquatic ecosystems (Kuris et al.,

2008) and alter food web topography, including competition and predation, through their

indirect effect on host abundance (Hatcher et al., 2012). They contribute to ecosystem

energy transfer between trophic levels (Lafferty et al., 2006), e.g. parasites may induce

changes in host behaviour that influence predation (Lefèvre et al., 2009), or act as mediators

in biological invasions (Blackburn et al., 2011). However, the predictability of parasite

communities at ecological timescales is often questioned as there is a lack of common

patterns found in open water areas and in host species with large distribution ranges (Timi &

Poulin, 2003).

Pelagic fish hosts are often considered as parasite species-poor compared to littoral host

communities (Marcogliese, 2002). Although the community structures of parasites have

been assessed in the context of biological tags of their hosts, including in pelagic marine

areas (reviewed in MacKenzie & Abaunza, (2014)), determinants structuring population-level

infection dynamics in pelagic environments are scarce.

Lake Tanganyika and its pelagic zone

Lake Tanganyika (LT) has been proposed as a natural study system for general principles of

evolution due to its high species diversity and endemicity of various taxa (Coulter, 1991a;

Cristescu et al., 2010). While Lake Tanganyika has all typical characteristics of a pelagic

environment (vertical stratification, seasonal upwellings, simple trophic structure), it also has

clear geographical boundaries and a species-poor pelagic community compared to marine

environments, offering simplified conditions to study pelagic ecosystem dynamics (Paugy &

Lévêque, 2017). It is dominated by two species of clupeid fishes which make up the majority

of fish biomass and are the main source of fisheries production in the four riparian countries

of the lake (Mannini et al., 1996). Despite the difficulty of tagging and effectively tracking

fragile small pelagic species, lake-scale latitudinal migrations of the endemic Tanganyika

clupeids, *Limnothrissa miodon* (Boulenger, 1906) and *Stolothrissa tanganicae* Regan, 1917

have been monitored through indirect methods, such as determining the difference in size

and abundance in relation to the season and geographic origin along the lake (Mulimbwa

N'Sibula et al., 2022; Plisnier et al., 2009). Fisheries production across landing sites varies

substantially, indicating seasonal migrations of the Tanganyika clupeids. These patterns

have been linked to predator avoidance and food-seeking behaviour (Plisnier et al., 2009).

Recent genomic studies of the Tanganyika clupeids find only a weak signal of isolation by

distance and no clear stock structure along the North-South axis of the lake (De Keyzer et

al., 2019; Junker et al., 2020).

Clupeid fishes in Lake Tanganyika are parasitised by two gill-infecting species of

Kapentagyrus Kmentová, Vanhove & Gelnar, 2018 (Kmentová et al., 2018). *Kapentagyrus* is

a lineage of monogenean flatworms (Monogenea, Dactylogyridae) that has so far only been

reported from African freshwater clupeid hosts (Vanhove et al., 2021). Monogenean

flatworms are obligate parasites of mainly fishes, some of which include pelagic hosts

(Kmentová et al., 2018; Plaksina et al., 2021; Van Der Lingen et al., 2015). Given their short

generation time, high substitution rate and host dependency, obligate parasites can serve as

a magnifying glass of host population dynamics (Catalano et al., 2014; Geraerts et al.,

2022). The patterns driving the spatiotemporal distribution of directly transmitted parasites

such as monogenean flatworms may be informative for pelagic ecosystem dynamics as

mediators between the external environment and the hosts. Due to their single-host life cycle

and specificity for pelagic hosts, species of *Kapentagyrus* in Lake Tanganyika are proposed

as tags for host distribution in the pelagic zone and the dynamics of pelagic ecosystems

(Kmentová et al., 2019; Schoeman et al., 2022).

Spatiotemporal dynamics in the pelagic zone - a holistic approach

Spatiotemporal dynamics of parasites have been studied mainly in the context of host-driven

determinants of their community composition (reviewed in Lester & MacKenzie, (2009)).

Large-scale population dynamics of parasites infecting pelagic organisms are hardly ever

studied due to the considerable effort required to sample fish hosts along their long-distance

migrations. To date, determinants of population connectivity of parasitic flatworms over large

geographic distances remains to be elucidated (Poulin, 2007). The population structure of

our study species, *K. limnotrissae* (infecting *L. miodon*) and *K. tanganicanus* (infecting both

clupeid species)*,* has previously been analysed over the entire geographic range of the lake,

including temporal and seasonal sampling. Following the pattern of their clupeid hosts, no

clear differentiation across the North-South axis of LT was found (Kmentová et al., 2020). To

study the ecosystem dynamics in the pelagic zone, surveys of pelagic fishes have been

combined with analyses of their morphological variation (Muniz et al., 2020; Valentin et al.,

2014), otolith composition (Javor et al., 2011) and, more recently, population genomics

(Baltazar-Soares et al., 2018; De Keyzer et al., 2019; Junker et al., 2020). This holistic

approach has been proposed not only to study spatiotemporal dynamics in the pelagic zone

(Abaunza et al., 2008; Kerr et al., 2017), but also other aquatic ecosystems such as large

rivers (Lavoué et al., 2008).

In this study, we investigate the spatiotemporal dynamics of parasites alongside the

morphology of their clupeid hosts as a flag taxon for the pelagic zone worldwide. We

combine geographical and seasonal results on parasite infection, and morphology of the

targeted fish hosts with previously published knowledge on the biology and genetic

population structure of monogenean parasites infecting clupeids, fishery statistics, seasonal

and annual cycles in primary productivity and upwelling (schematic representation of study

design presented in Fig. 1).

159
Fig. 1: Schematic visualisation of the study system of monogenean parasites infecting

clupeid fishes in Lake Tanganyika.

We hypothesise that 1) given the suggested high level of North-South mobility of clupeid

fishes in Lake Tanganyika (Plisnier et al., 2009; De Keyzer et al., 2019; Junker et al., 2020),

the population infection dynamics of *Kapentagyrus* spp. are not related to the spatiotemporal

origins, 2) given the contrasting host range between the two species of *Kapentagyrus* and

the differences in lifestyle between the host species, spatial differences in infection will be

more pronounced in *K. limnotrissae,* and 3) following previously published results on

contrasting introduction success in different host life stages (Kmentová et al., 2019), we

expect host size-driven differential occurrence of *Kapentagyrus* spp. on *L. miodon.*

Methodology

Sample collection and species identification

In total, 1730 specimens of two endemic pelagic clupeid species, *L. miodon* (733) and *S.*

tanganicae (997), were collected along the North-South axis of Lake Tanganyika, including

all three subbasins (North, Central, and South) in two different seasons (rainy season from

October to April and dry season from May to September) within a four-year period (Fig. 2,

Table S1). Freshly caught fish specimens were either obtained in collaboration with the

experimental fishing unit of the Centre de Recherche en Hydrobiologie - Uvira (CRH) (Uvira,

Democratic Republic of the Congo) or purchased from local fishermen. We combine newly

obtained data on monogenean infection of clupeids in LT with those published in previous

studies (Kmentová et al., 2020, 2018). Host specimens were collected within a period of 2

weeks (August 2016, April 2018, and October 2019) to avoid sampling the same population

twice, because both clupeid species are highly mobile (De Keyzer et al., 2019; Mulimbwa

N'Sibula & Mannini, 1993). Whole fish and/or gills were preserved in absolute ethanol. Host

specimens were examined for the presence of monogenean parasites according to the

procedure described in Kmentová et al. (2018). Species level identification of *Kapentagyrus*

spp. was based on distinctive characters of the hard parts of the attachment organ in the

posterior part of their bodies (for more details see Kmentová et al., (2018)). In the case of *L.*

miodon, the only species of the two clupeids that hosts two monogenean species,

monogenean individuals that could not be identified at the species level were only included

in the counts of total infection intensities.

Fig. 2: Overview of screened fish hosts and incidence of infection of collected parasite populations. Season and year of origin (x-axis) and standard length of the fish host (y-axis, SL in cm) of a) *Kapentagyrus limnotrissae* ex *Limnothrissa miodon*, b) *Kapentagyrus tanganicanus* ex *Limnothrissa miodon*, c) *Kapentagyrus tanganicanus* ex *Stolothrissa tanganicae*. Infected fishes are depicted in blue, non-infected in red. The total number of fish screened at a certain time point is mentioned above each of the boxplots. Subbasin division is visible by different shapes used for sampling localities with diamonds representing northern subbasin, a circle representing central subbasin and squares representing southern subbasin. The purple colour indicates the sampling localities of fish specimens used in the geomorphometric analyses.

Parasite population dynamics

To investigate the host parameters that might influence infection levels, the dataset was divided into three host-parasite combinations according to host species (i: *K. limnotrissae* ex *L. miodon*, ii: *K. tanganicanus* ex *L. miodon*, iii: *K. tangicanus* ex *S. tanganicae*). We modelled the infection intensity per host specimen against a range of other parameters, including sampling location (as subbasin - North, Central, South), season (dry period from May to September, rainy period from October to April), and host size (standard length) as explanatory variables (Table 1) as well as infection levels of the respective other parasite species if applicable, i.e. *K. tanganicanus* in (i) and *K. limnotrissae* in (ii). Because of seasonal migration and previous records on spatiotemporal variation in body size (Plisnier et al., 2009), we expect an interaction of host size with locality and season, respectively. Several studies also suggested that infection levels of monogeneans are related to fish size (Akoll et al., 2012; Šimková et al., 2004). Therefore, we included interaction effects between host size and the remaining parameters in the initial models.

Infection parameters of parasites often present a substantial amount of zero counts (Lester, 2012; Tinsley et al., 2020). Therefore, we fitted infection levels using zero-inflated models (ZIMs) with a Poisson probability distribution that assume that the excess of zero counts is

produced by a separate process. In the present case, we hypothesised that the excess of zero counts result from a lack of contact with parasites in some specimens, while true zeros arise from host resistance (Wang, et al., 2017; Zuur et al., 2009). We fitted generalised linear models without zero-inflation to test whether these assumptions are true. The present datasets are overdispersed (residual deviance/residual degrees-of-freedom > 1.5 for a Poisson distribution). To address this overdispersion, we used a negative binomial probability distribution. Finally, host specimens most likely represent non-independent samples, as fishes belonging to the same schools may have experienced more similar parasite exposure scenarios. Therefore, we also tested whether including the sampling day and locality as random effects in a mixed model further improved the model fit. To avoid overfitting, we simplified models through a backwards elimination procedure using the function *drop1* including a χ² test. All effects that failed to significantly improve model fit were removed, starting from the interaction effects.

All model-based analyses were carried out in R v4.1.2 (R Core Team, 2022). Models were fitted using the package *glmmTMB* v1.1.2.3 (Brooks et al., 2017). The package *glmmTMB* offers models with two options for negative binomial distributions (options *nbinom1* and *nbinom2*) that implement linear and quadratic parameterisation, respectively (see Hardin & Hilbe, (2007)). Both options were tested here. We compared model fits using the Akaike information criterion (AIC) through the function *AICtab* in the package *bbmle* v1.0.24 (Bolker, 2017). We also checked model fits through quantile-quantile plots and residual vs. fitted plots as provided by the package *DHARMa* v0.4.5 (Hartig, 2017).

Based on the best-fitting model, we predicted infection levels for all three host-parasite combinations as a function of the subbasin and season, as well as the continuous variables the hosts' standard length and the level of co-infections through the package *emmeans* (Lenth, 2022). The resulting figures were plotted through the packages *emmeans* and *ggplot2* (Wickham, 2016).

Geomorphometrics of clupeid hosts

We expect host phenotype to be potentially linked to different environmental conditions in the subbasins. Therefore, the two host species were examined to investigate the potential link between the morphological variation and geographic origin of host specimens. In addition to classical morphological assessment, the evaluation of body shape variation can be quantified via geomorphometrics (Elewa, 2004). Specimens collected from six different localities along the lake shoreline within two weeks in October 2019 were selected (Fig. 2 and Table S1). Photographs were taken using Canon 4000D reflex camera equipped with an EF-S 18-55 mm III-lens, set on 55 mm for a total of 224 specimens of *S. tanganicae* and 195 specimens of *L. miodon*. The body shape of each specimen was captured by a set of 20 fixed landmarks. Landmarks are reference points of coordinates in 2D or 3D (2D for this study), and contain essential information on size, shape and scale (Elewa, 2004; Savriama, 2018). They were set using the *tpsDig2* software v2.31 (Rohlf, 2018) using a tps file created with the *tpsUtil* software v1.78 (Rohlf, 2018). The landmarks were defined based on previous studies conducted with other species of clupeids (De La Cruz Agüero & Rodríguez, 2004; Mounir, Ewague, Znari, & Elmghazli, 2019; Silva, 2003), see Fig. 3.

Fig. 3: Position of landmarks recovered for 2D digitisation of specimen of a) *Limnothrissa miodon* and b) *Stolothrissa tanganicae*.

Morphological variation within the two clupeid species was analysed with *MorphoJ* v2 (Klingenberg, 2011). We made a distinction between analyses based on whole-body landmarks vs. head-only landmarks. For the analyses of the whole body, we used all obtained landmarks, for the head-only analyses, landmarks 1, 2 and 11-20 were included. To extract the shape information, full Procrustes fits of landmark data were performed and aligned by longitudinal axes of the specimens. Three classifiers (species, locality of origin, and subbasin) were imported for further analyses. Principal Component Analysis (PCA) was performed on the covariance matrix to visualise the shape variation. Highly deviating specimens, identified by the PCA plot, were excluded from the analysis. Regressions against the standard length (measured separately from each specimen) of each specimen followed by a 10,000 replicate permutation test were performed on the first three individual PC axes. Due to the significant correlation between the standard length and PCA loadings/Procrustes distances (see Figs. S1 & S2), the final PCAs were performed on residuals which resulted from the regression analyses of Procrustes distances and standard length. Canonical Variate Analyses (CVA) on the residuals (see above) and permutation tests of 10,000 replicates, were performed to test for differences in morphology between specimens from different sites of origin or subbasins. The resulting figures were plotted through the *R* packages *ggplot2* (Wickham, 2016), *RColorBrewer* (Neuwirt, 2022)*, ggtext* (Wilke, 2020) and *tidyverse* (Wickham et al., 2019).

Results

Parasite population dynamics

For all three datasets (i–iii), a zero-inflated negative binomial mixed model resulted in the best fit (Table 1). In any case, seasonality failed to significantly improve model fit. Therefore, both the parameter and its interaction with the standard length were removed from all models.

For *K. tanganicanus* (i and iii), a monogenean infecting both species of clupeids, the models with a quadratic parameterisation (*nbinom2*) outperformed the models with a linear parameterisation. *Kapentagyrus limnotrissae* infecting *L. miodon* (ii), only improved the model fit marginally (ΔAIC = 2; χ^2 (1, 2) = 2310, p = 0.038). In the post-hoc analysis, we found only minor differences between infection levels of *K. tanganicanus* (Figs. 4&5), yet subbasin identity contributed significantly to the overall model fit (Table 1). Infection intensities of *K. tanganicanus* increased with host size (Figs. 4&5) and with co-infection numbers of *K. limnotrissae* (Fig. 6). For *L. miodon,* this increase was generally weaker in the southern subbasin and particularly strong in the North. The host standard length had a significant interaction with the subbasin. For *S. tanganicae*, only the standard lengths of the hosts were a determinant of infection intensity. For *K. limnotrissae* (ii), the models with a linear parameterisation (*nbinom1*) outperformed those with a quadratic parameterisation. The interaction effect of subbasin with the standard length did not improve model fit and was, therefore, removed. The three minimal adequate models can be found in Table 1. Infection levels in the South of Lake Tanganyika were significantly lower than in other subbasins (Fig. 6). Infections significantly decreased with host standard length and increased with co-infection numbers of *K. tanganicanus* (Fig. 6).

1 Table 1: Stepwise backwards selection of effects (step 1 and 2) in generalised linear (mixed) models for each and host-parasite combination.

2 Models with the lowest values of the AIC (minimal adequate models) are highlighted in bold.

		K. tanganicanus ex *L. miodon* (i)			*K. limnotrissae* ex *L. miodon* (ii)			*K. tanganicanus* ex *S. tanganicae* (iii)		
		full model	step 1	step 2	full model	step 1	step2	full model	step 1	step2
Zeroinfl	Mixed effects		-ii:SL-Season:SL	-Season		-Subbasin:SL-Season:SL	-Season		-Season:SL-Subbasin:SL	-Subbasin-Season
Poisson		3714	3770	3899	1999	2011	2012	2368	2367	2366
NBinom1		2585	2582	2610	1579	1615	1646	2053	2048	2067
NBinom2		2506	2504	2525	1638	1677	1693	2096	2094	2098
Poisson	locality, date	3078	NA	2970	1856	NA	NA	2101	2113	2107
NBinom1	locality, date	2487	2484	2484	1564	1560	**1560**	1990	1986	1981
NBinom2	locality, date	2429	2427	**2425**	1591	1591	1589	1981	1981	**1975**

3 Abbreviations: standard length (SL), zero inflated model (Zeroinfl), negative binomial distribution (NBinom1,2), Poisson distribution (Poisson).
4

Fig. 4: Predictions for infection intensity of *Kapentagyrus tanganicanus* ex *Limnothrissa miodon* with 95% confidence intervals under the minimal adequate model (Table 1). a) Infection intensity in different subbasins showed no significant differences (CIs in blue, direction of overlap in red) with estimated marginal means on the x-axis, b) Infection intensity increased with standard length (SL) of the host, c) Infection intensity increased with higher infection intensities of *Kapentagyrus limnotrissae*.

Fig. 5: Prediction of infection intensity of *Kapentagyrus tanganicanus* ex *Stolothrissa tanganicae*. The infection intensity increased with the standard length (SL) of the host.

Fig. 6: Predictions for infection intensity of *Kapentagyrus limnotrissae* ex *Limnothrissa miodon*. a) Infection intensity related to subbasin origin with the confidence interval (in blue) and direction of overlap (in red) with estimated marginal means on the x-axis, b) Infection intensity as a function of the standard length (SL) of the host, c) Infection intensity as function of co-infection by *Kapentagyrus tanganicanus*.

Geomorphometrics of clupeid hosts

Geomorphometric analyses revealed strong similarity of head morphology between specimens from different localities and subbasins of the lake in both clupeid species. In the case of *L. miodon*, the first three PC axes explained 22.1%, 17.1% and 12.3% of the variation, respectively (see Fig. 7A&B), with visible North-South gradient mainly along the PC2. In case of *S. tanganicae*, the first three PC axes explained 24.2%, 18.1% and 11.8% of the variation, respectively (see Fig. 7C&D), with no visible structuring according to locality of origin. As wireframes displayed variation in the positioning of the full body, most likely related to the preservation of specimens, results based on the whole-body shape are presented in

the supplementary information only and should be interpreted with caution (Figs. S1&2).

Based on the wireframes of the head shape (Fig. 8), differences in the relative position of the

snout, eye and operculum are visible along the displayed PC axes.

The results of our CVAs confirmed the trend visible mainly along the second PC axis with

significant differences in the shape of the whole body and head related to the geographic

origin of *L. miodon* (Tables 2 and S1). Specifically, head shape differences increased with

geographic distance in *L. miodon*. Although the CVAs indicated significant shape differences

between some of the localities in both the full body and head datasets of *S. tanganicae* (see

Tables 2 and S1), no consistent geographical pattern was detected in the PCAs. In both

species, differences in the head shape were more pronounced in relation to geographic

origin compared to the full body.

Fig. 7: Biplots of Principal Component Analyses (PCA) showing the shape variation in the head across the sampled localities of a) *Limnothrissa miodon*, first two PCs displayed, b) *Limnothrissa miodon*, first and third PCs displayed, c) *Stolothrissa tanganicae,* first two PCs displayed, d) *Stolothrissa tanganicae,* first and third PCs displayed*.*

Fig. 8: Wireframes showing the shape variation in the head based on the coordinates of Principal Component Analyses of a) *Limnothrissa miodon,* PC1, b) *Limnothrissa miodon,* PC2, c) *Limnothrissa miodon,* PC3, d) *Stolothrissa tanganicae,* PC1, e) *Stolothrissa tanganicae,* PC2, f) *Stolothrissa tanganicae*, PC3. The target shape is presented in blue, the starting shape in pink.

Table 2: Results of Canonical Variate Analyses of head shape variation for a) *Limnothrissa miodon* b) *Stolothrissa tanganicae*. Values of Procrustes distances are displayed below the diagonal, P-values are shown above the diagonal. Significant P-values (<0.05) are indicated in bold. Number of samples are indicated between brackets behind the locality names with designation of the subbasin, N – North, C – Central, S – South).

(a)	Mulongwe (28)	Kalundu (34)	Kalemie (20)	Mpulungu (27)	Nsumbu (58)	Kalambo (23)
Mulongwe (N)	x	0.3842	**0.0207**	**0.0002**	**0.0466**	**0.0001**
Kalundu (N)	0.0252	x	**0.0029**	**0.0002**	**0.0376**	**0.0001**
Kalemie (C)	0.0447	0.0423	x	**<.0001**	**0.0355**	**0.0005**
Mpulungu (S)	0.0519	0.0472	0.0550	x	**0.0027**	0.2593
Nsumbu (S)	0.0298	0.0263	0.0308	0.0340	x	**0.0008**
Kalambo (S)	0.0603	0.0536	0.0489	0.0264	0.0379	x
(b)	Mulongwe (30)	Kalundu (38)	Kalemie (21)	Mpulungu (64)	Nsumbu (40)	Kalambo (30)
Mulongwe (N)	x	**0.0060**	0.0570	**0.0029**	**<.0001**	**0.0014**
Kalundu (N)	0.0479	x	0.0651	0.5339	0.0321	0.1954
Kalemie (C)	0.0387	0.0401	x	**0.0191**	**0.0116**	**0.0336**
Mpulungu (S)	0.0382	0.0188	0.0336	x	**0.0196**	0.3979
Nsumbu (S)	0.0584	0.0339	0.0368	0.0270	x	0.0785
Kalambo (S)	0.0467	0.0301	0.0353	0.0189	0.0274	x

Discussion

Temporal stability and spatial differences of infection

Previous studies indicate that the distribution of aquatic parasites is determined by environmental factors (Timi & Poulin, 2003). Also, many parasites are vulnerable to extinction, as a decline in population size of the host can negatively affect the often-overdispersed distribution of parasite populations (Lester, 2012). In Lake Tanganyika, annual differences in fisheries production are believed to be caused by natural cycles of clupeid species. This is visible in density changes of fish populations over a spatiotemporal gradient (Mölsä et al., 1999; Plisnier et al., 2009). Additionally, seasonal fluctuations in clupeid catches in Lake Tanganyika are driven by external factors related to the hydrodynamics and the weather. Despite the reported North-South clupeid migrations resulting in seasonal fluctuations of host population densities and mean size distribution of

the hosts (Kimirei & Mgaya, 2007; Mulimbwa N'Sibula et al., 2022), in line with our

hypothesis we found the spatiotemporal dynamics of *Kapentagyrus* spp. in the pelagic zone

to be seasonally independent. This result confirms previously suggested independence on

seasonal host population cycles of rather short-lived parasites (Lester & MacKenzie, 2009)

of which monogeneans in the tropical areas with an estimated generation time of days up to

a few weeks might be considered (Tomnatik, 1990). Alternatively, host fluctuations could

appear over longer cycles that were not covered by the sampling design of our study. While

we could not disentangle the effects of host size and geographic origin for *K. tanganicanus*,

K. limnotrissae exhibited a slight spatial differentiation along the North-South axis. We

suggest that this pattern is driven by temporal residency and patchy distribution of *L. miodon*

in the littoral habitat before reaching a certain size and becoming pelagic. In general, the

geographic distribution of parasites infecting pelagic fish hosts is linked to life cycle

complexity (number of host species with different distribution ranges) with limited spatial

distribution of monogeneans only to certain areas, as in the case of *Sprattus sprattus* L.

(Kleinertz et al., 2012) and *Clupea harengus* L. (Actinopterygii, Clupeidae) (Rahimian et al.,

1999) in the North Sea. In comparison to monogeneans infecting clupeid fishes in Lake

Tanganyika, no host size related infection intensity was observed in the case of *Mazocraes*

alosae Hermann, 1782 on two species of *Alosa* Linck, 1790 (Actinopterygii, Clupeidae) from

North-Atlantic coastal waters (Gérard et al., 2017). Host behaviour and season were shown

to drive temporal differences in infection levels of *M. alosae* on *Alosa immaculata* Bennett,

1835 in the northern Black and Azov Seas (Plaksina et al., 2021).

A holistic approach for stock identification has been highlighted before (Begg and Waldman,

1999). The combination of parasite community data and host morphometric data aided in the

discovery of multiple stocks of horse mackerel (*Trachurus trachurus* L.) in the North Atlantic

(Abaunza et al., 2008) and in the stock identification of Australian sardines (*Sardinops sagax*

(Jenyns, 1842)) along the East African coast (Van Der Lingen et al., 2015). Unlike parasites

with complex life cycles, directly transmitted parasites have the advantage that distribution is

limited to the target host enabling elucidation of large-scale and long-term population

dynamics of certain host-parasite interaction (Catalano et al., 2014). Using a

geomorphometric approach, we found intraspecific morphological differences in the LT

clupeid species linked to the geographic origin of the specimens especially in *L. miodon*.

Morphological variation can be driven by various environmental factors (Mounir et al., 2019;

Sultan & Stearns, 2005). In our study, the level of geographically determined morphological

variation differed between the studied fish species. The shape differences in the relative

position of snout and eyes may be related to the geographical distance and more concretely

subbasin origin of *L. miodon*. The environment in the middle part of the lake could serve as a

transit site between the northern and southern parts. Moreover, compared to other areas,

the northern peak can be seen as the most affected by anthropogenic pressures (Cohen et

al., 2005). Morphological variation of the head was also visible in *S. tanganicae*, but without

a clear geographical pattern. These results indicate a more profound shape differentiation in

L. miodon compared to *S. tanganicae,* despite originating from the same localities. The

absence of clearly geographically determined morphological differences in *S. tanganicae*

can be explained by the lack of residential behaviour. Although the results indicate that the

population of *L. miodon* is more structured compared to *S. tanganicae*, the lack of a clear

geographic cline supports the high level of population connectivity of this species across the

lake (Junker et al., 2020). As hypothetised, given the host life-stage dependency of *K.*

limnotrissae, contrasting infection levels related to geographic origin suggest that the spatial

distribution of the parasites and the morphological response of the clupeid hosts follow

similar geographic patterns. Overall, our results on lake-wide parasite occurrence and

seasonal dynamics combined with characterisation of clupeid hosts support the hypothesis

of largely unrestricted migration in *S. tanganicae* and the lack of barriers in the pelagic zone

of the lake (De Keyzer et al., 2019; Junker et al., 2020; Mulimbwa N'Sibula et al., 2022).

Habitat differences between the clupeid hosts over their lifespan most likely drive contrasting

spatial patterns of infection between closely related parasite species. This result further

supports the importance of host species with the highest dispersal capacity on the spatial

distribution of parasites, as reported for *Cichlidogyrus casuarinus* Pariselle, Muterezi

Bukinga & Vanhove, 2015 infecting pelagic cichlids in the lake (Kmentová et al., 2021).

Contrasting levels of gene flow between the two parasite species related to geographic

origin along the North-South axis of the lake (Kmentová et al., Under review) may therefore

be associated with lifestyle differences between the two clupeid species.

Host size and life-stage dependent habitat preference drive parasite occurrence

The diurnal and geographically small-scale migrations of Tanganyika clupeids are generally

better understood than the long term patterns. Vertically, these clupeids respond to light and

oxygen concentration and follow the diurnal movements of their prey. Horizontally, they

migrate between the offshore waters of the pelagic and the inshore waters of the littoral

habitat throughout their development (Matthes 1967). A recent long-term study of the

reproductive activities of the clupeids in the northern subbasin identified the littoral zone as

the main spawning ground for *L. miodon,* and the pelagic zone as the equivalent for *S.*

tanganicae (Mulimbwa N'sibula et al., 2022). Juveniles of both species occur in littoral

habitats, but in general *S. tanganicae* lives and spawns at greater distances from the shore

than *L. miodon* (Coulter, 1970, 1991b; Mannini et al., 1996). Eggs of *L. miodon can* even be

deposited just above the sandy bottom of the littoral zone. Juveniles of *L. miodon,* but not *S.*

tanganicae, occupy sandy beaches right by the shore (Mulimbwa N'Sibula et al., 2022).

In line with our hypothesis, the body size of the clupeid hosts appears to be a major

determinant of infection, as it significantly affects infection intensities for all three host-

parasite combinations. Such an overall positive correlation between fish size and

monogenean infection intensities has been associated with a larger habitat offered by larger

hosts (Alvarez-Pellitero & Gonzalez-Lanza, 1982; Poulin, 2000). Considering the age-

dependent migration of *L. miodon* from the littoral to the pelagic zone, the contrasting

infection intensities associated with host size between *K. limnotrissae* and *K. tanganicanus*

suggests a spatial stratification of infection linked with the ontogenetic migration of *L. miodon*

from the littoral to the pelagic zone. This apparent replacement contrasts with the positive

effect of each parasite species' infection intensity on that of the other species. We suggest a

combination of changes in host habitat and facilitation of infection caused by host immune

deficiency or suppression at the base of this pattern. Suppression of the fish immune system

allowing higher infection intensities has been shown in previous studies on monogeneans

(Rohlenová et al., 2011; Sitjà-Bobadilla, 2008) and other parasite taxa (Klemme et al., 2016).

Alternatively, the antagonistic interaction between the two parasite species could explain the

resulting pattern, as proposed e.g., in the monogenean communities infecting *Astyanax*

aeneus (Günther, 1860) in Neotropical rivers in Mexico (Salgado-Maldonado et al., 2019).

However, so far there is no evidence of negative or any direct interspecific monogenean

interaction between closely related species (Šimková et al., 2000; Soler-Jiménez & Fajer-

Ávila, 2012).

Conclusion

Lake Tanganyika is an ancient and pristine lake known for a spectacular level of species

diversity and endemism. The existence of schooling pelagic fishes in an enclosed ecosystem

promotes studies on general ecosystem dynamics of pelagic areas. The link between

theoretical/experimental studies and dynamics in natural ecosystems is largely unknown

(Lindegren et al., 2016) and a common framework is lacking (Ritz et al., 2011). Incorporation

of host-parasite interactions increases our knowledge on the interconnectedness of trophic

levels (Lafferty et al., 2006). Our holistic approach based on spatiotemporal distribution of

fish parasites and morphological variation of the host species reveals insights on the

ecosystem dynamics along a North-South axis, and habitat preferences of two important

fisheries targets. As restricted migration of *L. miodon* is considered as one of the drivers of

profound North-South differences in parasite occurrence and fish morphology, being also

indicated in recent studies (De Keyzer et al., 2019; Junker et al., 2020), such results should

be considered in fisheries management plans. Our results suggest that parasite population

dynamics in pelagic zones are dependent on spatial distribution of the hosts in both large

and small scales. Intensity of infection seems to be mainly driven by fish size with mutual

facilitation in case of co-infection. Lack of seasonality in infection intensity points to

environmental stability of the pelagic zone and absence of density dependent infection.

Acknowledgements

We would like to thank Stephan Koblmüller, Holger Zimmermann, Jiří Vorel, Simona

Georgieva, Cyprian Katongo, Taylor Banda, Aneesh P.H. Bose, Filip A.M. Volckaert and Els

De Keyzer for their help in organising and conducting field work. Pierre-Denis Plisnier is

acknowledged for providing bathymetric data of Lake Tanganyika. The study was supported

by the Czech Science Foundation (GACR) project no. P505/12/G112—European Centre of

Ichtyoparasitology (ECIP) and standard project GA19-13573S, research grant 1513419N of

the Research Foundation—Flanders (FWO-Vlaanderen). AJCL (BOF19OWB02), NK

(BOF21PD01) and MPMV (BOF20TT06) received support from the Special Research Fund

of Hasselt University.

Conflict of Interest

Authors declare no conflict of interest.

Author Contributions

N.K. designed the study, generated incidence data, analysed geometric morphometric data

and drafted the manuscript. A.C-L. analysed incidence data and helped draft the manuscript,

M.J. generated part of the incidence data and helped to draft the manuscript, M.V.S.

supervised geometric morphometric data analyses and interpretation of results, M.P.M.V.

discussed the results, helped draft the manuscript and supervised the study, T.M. and

E.V.H. helped with geometric morphometric data analyses, S.H. generated geometric

morphometric data, L.Mi. and K.T. helped with data interpretation and draft of the

manuscript, L.Ma. provided support in the field and knowledge on the studied ecosystem. All

authors have read and agreed to the published version of the manuscript.

Data Availability Statement

Parasite voucher material was deposited in the collection of Hasselt University under

accession numbers xx-xx and the Royal Museum for Central Africa. The geometric

morphometric data underlying the results of this article are available in Mendeley Data (xxx).

Supporting Information

Additional supporting information may be found in the online version of the article at the

publisher's website.

References

Abaunza, P., Murta, A. G., Campbell, N., Cimmaruta, R., Comesaña, A. S., Dahle, G., …

Zimmermann, C. (2008). Stock identity of horse mackerel (*Trachurus trachurus*) in the

Northeast Atlantic and Mediterranean Sea: Integrating the results from different stock

identification approaches. *Fisheries Research*, *89*, 196–209.

https://doi.org/10.1016/J.FISHRES.2007.09.022

Akoll, P., Konečný, R., Mwanja, W. W., & Schiemer, F. (2012). Risk assessment of parasitic

helminths on cultured Nile tilapia (*Oreochromis niloticus*, L.). *Aquaculture*, *356*, 123–

127. https://doi.org/10.1016/j.aquaculture.2012.05.027

Alvarez-Pellitero, M. P., & Gonzalez-Lanza, M. C. (1982). Description and population

dynamics of *Dactylogyrus legionensis* n. sp. from *Barbus barbus bocagei* Steind.

Journal of Helminthology, *56*(3), 263–273. https://doi.org/10.1017/S0022149X00034647

Angel, M. V. (1993). Biodiversity of the pelagic ocean. *Conservation Biology*, *7*, 760–772.

https://doi.org/10.1046/j.1523-1739.1993.740760.x

Baltazar-Soares, M., Hinrichsen, H. H., & Eizaguirre, C. (2018). Integrating population

genomics and biophysical models towards evolutionary-based fisheries management.

ICES Journal of Marine Science, *75*, 1245–1257.

https://doi.org/10.1093/ICESJMS/FSX244

Blackburn, T. M., Pyšek, P., Bacher, S., Carlton, J. T., Duncan, R. P., Jarošík, V., …

Richardson, D. M. (2011). A proposed unified framework for biological invasions.

Trends in Ecology and Evolution, *26*, 333–339.

https://doi.org/10.1016/j.tree.2011.03.023

Bolker, B. (2017). *Package 'bbmle'." Tools for General Maximum Likelihood Estimation*.

Brooks, M. E. ;, Kristensen, K. ;, Van Benthem, K. J. ;, Magnusson, A. ;, Berg, C. W. ;,

Nielsen, A. ;, … Mächler, M. (2017). glmmTMB balances speed and flexibility among

packages for zero-inflated generalized linear mixed modeling. *The R Journal*, *9*, 378–

400. https://doi.org/10.3929/ETHZ-B-000240890

Catalano, S. R., Whittington, I. D., Donnellan, S. C., & Gillanders, B. M. (2014). Parasites as

biological tags to assess host population structure: Guidelines, recent genetic advances

and comments on a holistic approach. *International Journal for Parasitology: Parasites*

and Wildlife, *3*, 220–226. https://doi.org/10.1016/j.ijppaw.2013.11.001

Cohen, A. S., Palacios-Fest, M. R., Msaky, E. S., Alin, S. R., McKee, B., O'Reilly, C. M., …

Lezzar, K. E. (2005). Paleolimnological investigations of anthropogenic environmental

change in Lake Tanganyika: IX. Summary of paleorecords of environmental change

and catchment deforestation at Lake Tanganyika and impacts on the Lake Tanganyika

ecosystem. *Journal of Paleolimnology 34,* 125–145. https://doi.org/10.1007/S10933-

005-2422-4

Corman, J. R., McIntyre, P. B., Kuboja, B., Mbemba, W., Fink, D., Wheeler, C. W., …

Flecker, A. S. (2010). Upwelling couples chemical and biological dynamics across the

littoral and pelagic zones of Lake Tanganyika, East Africa. *Limnology and*

Oceanography, *55*, 214–224. https://doi.org/10.4319/LO.2010.55.1.0214

Coulter, G. W. (1970). Population changes within a group of fish species in Lake Tanganyika

following their exploitation. *Journal of Fish Biology*, *2*, 329–353.

https://doi.org/10.1111/j.1095-8649.1970.tb03292.x

Coulter, G. W. (1991a). Introduction. In G. W. Coulter (Ed.), *Lake Tanganyika and Its Life*

(pp. 1–6). London Oxford & New York: Natural History Museum Publications & Oxford

University Press.

Coulter, G. W. (1991b). Pelagic Fish. In G. W. Coulter (Ed.), *Lake Tanganyika and Its Life*

(pp. 111–150). London Oxford & New York: Natural History Musem & Oxford University

Press.

Cristescu, M. E., Adamowicz, S. J., Vaillant, J. J., & Haffner, D. G. (2010). Ancient lakes

revisited: from the ecology to the genetics of speciation. *Molecular Ecology*, *19*, 4837–

4851. https://doi.org/10.1111/J.1365-294X.2010.04832.X

De Keyzer, E. L. R., De Corte, Z., Van Steenberge, M., Raeymaekers, J. A. M., Calboli, F. C.

F., Kmentová, N., … Vanhove, M. P. M. (2019). First genomic study on Lake

Tanganyika sprat *Stolothrissa tanganicae*: A lack of population structure calls for

integrated management of this important fisheries target species. *BMC Evolutionary*

Biology, *19*, 6. https://doi.org/10.1186/s12862-018-1325-8

De La Cruz Agüero, J., & Rodríguez, F. J. G. (2004). Morphometric stock structure of the

Pacific sardine *Sardinops sagax* (Jenyns, 1842) off Baja California, Mexico. In A. M. T.

Elewa (Ed.), *Morphometrics: Applications in Biology and Paleontology* (pp. 115–127).

https://doi.org/10.1007/978-3-662-08865-4_9

Elewa, A. (2004). *Morphometrics: applications in biology and paleontology. Vol.14*. Springer

Science & Business Media.

Geraerts, M., Huyse, T., Barson, M., Bassirou, H., Bilong Bilong, C. F., Bitja Nyom, A. R., …

Vanhove, M. P. M. (2022). Mosaic or melting pot: The use of monogeneans as a

biological tag and magnifying glass to discriminate introduced populations of Nile tilapia

in sub-Saharan Africa. *Genomics*, *114*, 110328.

https://doi.org/10.1016/J.YGENO.2022.110328

Gérard, C., Hervé, M., Gay, M., Bourgau, O., Feunteun, E., Acou, A., & Réveillac, E. (2017).

Metazoan parasite communities in *Alosa alosa* (Linnaeus, 1758) and *Alosa fallax*

(Lacépède, 1803) (Clupeidae) from North-East Atlantic coastal waters and connected

rivers. *Parasitology Research*, *116*, 2211–2230. https://doi.org/10.1007/S00436-017-

5525-8/TABLES/12

Hardin, J. W., & Hilbe, J. M. (2007). *Generalized linear models and extensions*. Stata press.

Hartig, F. (2017). *DHARMa: Residual Diagnostics for Hierarchical (Multi-Level/Mixed)*

Regression Models. R package version 0.4.5.

Hatcher, M. J., Dick, J. T. A., & Dunn, A. M. (2012). Disease emergence and invasions.

Functional Ecology, *26*, 1275–1287. https://doi.org/10.1111/j.1365-2435.2012.02031.x

Javor, B., Lo, N., & Vetter, R. (2011). Otolith morphometrics and population structure of

Pacific Sardine (*Sardinops sagax*) along the west coast of North America. *Fishery*

Bulletin, *109*, 402–415.

Junker, J., Rick, J. A., McIntyre, P. B., Kimirei, I., Sweke, E. A., Mosille, J. B., … Wagner, C.

E. (2020). Structural genomic variation leads to genetic differentiation in Lake

Tanganyika's sardines. *Molecular Ecology*, *29*, 3277–3298.

https://doi.org/10.1111/MEC.15559

Kerr, L. A., Hintzen, N. T., Cadrin, S. X., Clausen, L. W., Dickey-Collas, M., Goethel, D. R.,

… Nash, R. D. M. (2017). Lessons learned from practical approaches to reconcile

mismatches between biological population structure and stock units of marine fish.

ICES Journal of Marine Science, *74*, 1708–1722.

https://doi.org/10.1093/ICESJMS/FSW188

Kimirei, I., & Mgaya, Y. (2007). Influence of environmental factors on seasonal changes in

clupeid catches in the Kigoma area of Lake Tanganyika. *African Journal of Aquatic*

Science, *32*, 291–298. https://doi.org/10.2989/AJAS.2007.32.3.9.308

Kleinertz, S., Klimpel, S., & Palm, H. W. (2012). Parasite communities and feeding ecology

of the European sprat (*Sprattus sprattus* L.) over its range of distribution. *Parasitology*

Research, *110*, 1147–1157. https://doi.org/10.1007/S00436-011-2605-Z/TABLES/3

Klemme, I., Louhi, K. R., & Karvonen, A. (2016). Host infection history modifies co-infection

success of multiple parasite genotypes. *Journal of Animal Ecology*, *85*, 591–597.

https://doi.org/10.1111/1365-2656.12472

Klingenberg, C. P. (2011). MorphoJ: an integrated software package for geometric

morphometrics. *Molecular Ecology Resources*, *11*, 353–357.

https://doi.org/10.1111/j.1755-0998.2010.02924.x

Kmentová, N., Van Steenberge, M., Raeymaekers, J. A. R., Koblmüller, S., Hablützel, P. I.,

Muterezi Bukinga, F., … Vanhove, M. P. M. (2018). Monogenean parasites of sardines

in Lake Tanganyika: Diversity, origin and intra-specific variability. *Contributions to*

Zoology, *87*, 105–132.

Kmentová, N., Van Steenberge, M., Thys van den Audenaerde, D. F. E., Nhiwatiwa, T.,

Muterezi Bukinga, F., Mulimbwa N'sibula, T., … Vanhove, M. P. M. (2019). Co-

introduction success of monogeneans infecting the fisheries target *Limnothrissa miodon*

differs between two non-native areas: The potential of parasites as a tag for introduction

pathway. *Biological Invasions*, *21*, 757–773. https://doi.org/10.1007/s10530-018-1856-3

Kmentová, N., Koblmüller, S., Van Steenberge, M., Raeymaekers, J. A. M., Artois, T., De

Keyzer, E. L. R., … Vanhove, M. P. M. (2020). Weak population structure and recent

demographic expansion of the monogenean parasite *Kapentagyrus* spp. infecting

clupeid fishes of Lake Tanganyika, East Africa. *International Journal for Parasitology*,

50, 471–486. https://doi.org/10.1016/j.ijpara.2020.02.002

Kmentová, N., Hahn, C., Koblmüller, S., Zimmermann, H., Vorel, J., Artois, T., … Vanhove,

M. P. M. (2021). Contrasting host-parasite population structure: Morphology and

mitogenomics of a parasitic flatworm on pelagic deepwater cichlid fishes from Lake

Tanganyika. *Biology*, *10,* 797. https://doi.org/10.3390/biology10080797

Kmentová, N., Thys, K., Hahn, C., Vorel, J., Koblmüller, S., Chocha Manda, A., Makasa, L.,

… Vanhove, M. P. M. Comparative population mitogenomics of fish parasites reveals

contrasting geographic pattern in the pelagic zone of Lake Tanganyika. Under review in

Molecular Ecology; preprint available on *Authorea*.

https://doi.org/10.22541/au.167274604.41657501/v1.

Kuris, A. M., Hechinger, R. F., Shaw, J. C., Whitney, K. L., Aguirre-Macedo, L., Boch, C. A.,

… Lafferty, K. D. (2008). Ecosystem energetic implications of parasite and free-living

biomass in three estuaries. *Nature*, *454*, 515–518. https://doi.org/10.1038/nature06970

Lafferty, K. D., Dobson, A. P., & Kuris, A. M. (2006). Parasites dominate food web links.

Proceedings of the National Academy of Sciences, *103*, 11211–11216.

https://doi.org/10.1073/PNAS.0604755103

Lavoué, S., Sullivan, J. P., Arnegard, M. E., & Hopkins, C. D. (2008). Differentiation of

morphology, genetics and electric signals in a region of sympatry between sister

species of African electric fish (Mormyridae). *Journal of Evolutionary Biology*, *21*, 1030–

1045. https://doi.org/10.1111/J.1420-9101.2008.01544.X

Lefèvre, T., Lebarbenchon, C., Gauthier-Clerc, M., Missé, D., Poulin, R., & Thomas, F.

(2009). The ecological significance of manipulative parasites. *Trends in Ecology &*

Evolution, *24*, 41–48. https://doi.org/10.1016/J.TREE.2008.08.007

Legendre, L., & Le Fèvre, J. (1991). From individual plankton cells to pelagic marine

ecosystems and to global biogeochemical cycles. *Particle Analysis in Oceanography*,

261–300. https://doi.org/10.1007/978-3-642-75121-9_11

Lenth, R. V. (2022). *emmeans: Estimated Marginal Means, aka Least-Squares Means. R*

package version 1.7.5.

Lester, R. J. G. (2012). Overdispersion in marine fish parasites. *Journal of Parasitology*, *98*,

718–721. https://doi.org/10.1645/GE-3017.1

Lester, R. J. G., & MacKenzie, K. (2009). The use and abuse of parasites as stock markers

for fish. *Fisheries Research*, *97*, 1–2. https://doi.org/10.1016/J.FISHRES.2008.12.016

Lindegren, M., Checkley, D. M., Ohman, M. D., Koslow, J. A., & Goericke, R. (2016).

Resilience and stability of a pelagic marine ecosystem. *Proceedings of the Royal*

Society B: Biological Sciences, *283*, 20151931.

https://doi.org/10.1098/RSPB.2015.1931

MacKenzie, K., & Abaunza, P. (2014). Parasites as Biological Tags. *Stock Identification*

Methods: Applications in Fishery Science: Second Edition, 185–203.

https://doi.org/10.1016/B978-0-12-397003-9.00010-2

Mannini, P., Aro, E., Katonda, K. I., Kassaka, B., Mambona, C., Milindi, G., … Verburg, P.

(1996). Pelagic fish stocks of Lake Tanganyika: Biology and exploitation. FAO/FINNIDA

research for the management of the fisheries of Lake Tanganyika. In

GCP/RAF/271/FIN—TD/53 (En). Bujumbura.

Marcogliese, D. J. (2002). Food webs and the transmission of parasites to marine fish.

Parasitology, *124*, 83–99. https://doi.org/10.1017/S003118200200149X

Mölsä, H., Reynolds, J. E., Coenen, E. J., & Lindqvist, O. V. (1999). Fisheries research

towards resource management on Lake Tanganyika. *Hydrobiologia*, *407*, 1–24.

https://doi.org/10.1023/A:1003712708969

Mounir, A., Ewague, A., Znari, M., & Elmghazli, H. (2019). Discrimination of the phenotypic

sardine *Sardina pilchardus* stocks off the Moroccan Atlantic coast using a morphometric

analysis. *Https://Doi.Org/10.2989/1814232X.2019.1597765*, *41*, 137–144.

https://doi.org/10.2989/1814232X.2019.1597765

Mulimbwa N'Sibula, T., & Mannini, P. (1993). Demographic characteristics of *Stolothrissa*

tanganicae, Limnothrissa miodon and *Lates stappersii* in the Northwestern (Zairean)

waters of Lake Tanganyika. *CIFA Occasional Paper (FAO)*. Retrieved from

http://agris.fao.org/agris-search/search.do?recordID=XF9543811

Mulimbwa N'Sibula, T., Milec, L. J. M., Raeymaekers, J. A. M., Sarvala, J., Plisnier, P. D.,

Marwa, B., & Micha, J. C. (2022). Spatial and seasonal variation in reproductive indices

of the clupeids *Limnothrissa miodon* and *Stolothrissa tanganicae* in the Congolese

waters of northern Lake Tanganyika. *Belgian Journal of Zoology*, *152*, 13–31.

https://doi.org/10.26496/BJZ.2022.96

Muniz, A. A., Moura, A., Triay-Portella, R., Moreira, C., Santos, P. T., & Correia, A. T.

(2020). Population structure of the chub mackerel (*Scomber colias*) in the North-east

Atlantic inferred from otolith shape and body morphometrics. *Marine and Freshwater*

Research, *72*, 341–352. https://doi.org/10.1071/MF19389

Neuwirt, E. (2022). *RColorBrewer: ColorBrewer Palettes. R package version 1.1-3*.

Paugy, D., & Lévêque, C. (2017). Diets and food webs. In D. Paugy, C. Lévêque, & O. Otero

(Eds.), *The inland water fishes of Africa: Diversity, ecology and human use* (pp. 233–

257).

Pauly, D., Christensen, V., Guénette, S., Pitcher, T. J., Sumaila, U. R., Walters, C. J., …

Zeller, D. (2002). Towards sustainability in world fisheries. *Nature*, *418*(6898), 689–695.

https://doi.org/10.1038/NATURE01017

Plaksina, M. P., Gibson, D. I., & Dmitrieva, E. V. (2021). Life-history studies on

infrapopulations of *Mazocraes alosae* (Monogenea) parasitising *Alosa immaculata*

(Actinopterygii) in the northern Black and Azov Seas.

Http://Folia.Paru.Cas.Cz/Doi/10.14411/Fp.2021.009.Html, *68*, 1–10.

https://doi.org/10.14411/FP.2021.009

Plisnier, P. D., Mgana, H., Kimirei, I., Chande, A., Makasa, L., Chimanga, J., … Cornet, Y.

(2009). Limnological variability and pelagic fish abundance (*Stolothrissa tanganicae* and

Lates stappersii) in Lake Tanganyika. *Hydrobiologia*, *625*, 117–134.

https://doi.org/10.1007/s10750-009-9701-4

Poulin, R. (2000). Variation in the intraspecific relationship between fish length and intensity

of parasitic infection: Biological and statistical causes. *Journal of Fish Biology*, *56*, 123–

137. https://doi.org/10.1006/jfbi.1999.1146

Poulin, R. (2007). *Evolutionary ecology of parasites. 2nd ed.* Princeton University Press,

New Jersey, USA.

Poulin, R. (2014). Parasite biodiversity revisited: Frontiers and constraints. *International*

Journal for Parasitology, *44*, 581–589. https://doi.org/10.1016/J.IJPARA.2014.02.003

R Core Team (2022). R: A language and environment for statistical computing. Foundation

for Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.

Rahimian, H., Longshaw, M., MacKenzie, K., & Thulin, J. (1999). *Pseudanthocotyloides*

heterocotyle (van Beneden, 1871) Euzet & Prost, 1969 (Monogenea:

Polyopisthocotylea: Mazocraeidae), a parasite of herring *Clupea harengus* L. and sprat

Sprattus sprattus L. (Teleostei: Clupeidae). *Systematic Parasitology, 42*, 193–201.

https://doi.org/10.1023/A:1006042302301

Ritz, D. A., Hobday, A. J., Montgomery, J. C., & Ward, A. J. W. (2011). Social aggregation in

the pelagic zone with special reference to fish and invertebrates. *Advances in Marine*

Biology, *60*, 161–227. https://doi.org/10.1016/B978-0-12-385529-9.00004-4

Rohlenová, K., Morand, S., Hyršl, P., Tolarová, S., Flajšhans, M., & Šimková, A. (2011). Are

fish immune systems really affected by parasites? An immunoecological study of

common carp (*Cyprinus carpio*). *Parasites & Vectors*, *4*, 120.

https://doi.org/10.1186/1756-3305-4-120

Salgado-Maldonado, G., Mendoza-Franco, E. F., Caspeta-Mandujano, J. M., & Ramírez-

Martínez, C. (2019). Aggregation and negative interactions in low-diversity and

unsaturated monogenean (Platyhelminthes) communities in *Astyanax aeneus*

(Teleostei) populations in a neotropical river of Mexico. *International Journal for*

Parasitology: Parasites and Wildlife, *8*, 203–215.

https://doi.org/10.1016/j.ijppaw.2019.02.005

Savriama, Y. (2018). A Step-by-step guide for geometric morphometrics of floral symmetry.

Frontiers in Plant Science, *9*, 1433. https://doi.org/10.3389/FPLS.2018.01433/BIBTEX

Schoeman, A. L., du Preez, L. H., Kmentová, N., & Vanhove, M. P. M. (2022). A

monogenean parasite reveals the widespread translocation of the African clawed frog in

its native range. *Journal of Applied Ecology*, *59*, 2670–2687.

https://doi.org/10.1111/1365-2664.14271

Silva, A. (2003). Morphometric variation among sardine (*Sardina pilchardus*) populations

from the northeastern Atlantic and the western Mediterranean. *ICES Journal of Marine*

Science, *60*, 1352–1360. https://doi.org/10.1016/S1054-3139(03)00141-3

Šimková, A., Desdevises, Y., Gelnar, M., & Morand, S. (2000). Co-existence of nine gill

ectoparasites (*Dactylogyrus*: Monogenea) parasitising the roach (*Rutilus rutilus* L.):

History and present ecology. *International Journal for Parasitology*, *30*, 1077–1088.

https://doi.org/10.1016/S0020-7519(00)00098-9

Šimková, A., Morand, S., Jobet, E., Gelnar, M., & Verneau, O. (2004). Molecular phylogeny

of congeneric monogenean parasites (*Dactylogyrus*): A case of intrahost speciation.

Evolution; International Journal of Organic Evolution, *58*, 1001–1018.

https://doi.org/10.1111/j.0014-3820.2004.tb00434.x

Sitjà-Bobadilla, A. (2008). Living off a fish: A trade-off between parasites and the immune

system. *Fish & Shellfish Immunology*, *25*, 358–372.

https://doi.org/10.1016/J.FSI.2008.03.018

Soler-Jiménez, L. C., & Fajer-Ávila, E. J. (2012). The microecology of dactylogyrids

(Monogenea: Dactylogyridae) on the gills of wild spotted rose snapper *Lutjanus guttatus*

(Lutjanidae) from Mazatlan Bay, Mexico. *Folia Parasitologica*, *59*, 53–58.

https://doi.org/10.14411/fp.2012.008

Sultan, S. E., & Stearns, S. C. (2005). Environmentally contingent variation: Phenotypic

plasticity and norms of reaction. *Variation*, 303–332. https://doi.org/10.1016/B978-

012088777-4/50016-8

Teske, P. R., Emami-Khoyi, A., Golla, T. R., Sandoval-Castillo, J., Lamont, T., Chiazzari, B.,

… van der Lingen, C. D. (2021). The sardine run in southeastern Africa is a mass

migration into an ecological trap. *Science Advances*, *7,* eabf4514.

https://doi.org/10.1126/sciadv.abf4514

Timi, J. T., & Poulin, R. (2003). Parasite community structure within and across host

populations of a marine pelagic fish: How repeatable is it? *International Journal for*

Parasitology, *33*, 1353–1362. https://doi.org/10.1016/S0020-7519(03)00203-0

Tinsley, R. C., Rose Vineer, H., Grainger-Wood, R., & Morgan, E. R. (2020). Heterogeneity

in helminth infections: Factors influencing aggregation in a simple host-parasite system.

Parasitology, *147*, 65–77. https://doi.org/10.1017/S003118201900129X

Tomnatik, V. E. (1990). The influence of water temperature on the sexual maturation of

Dactylogyrus vastator. *Parazitologiya*, *24*, 235–238.

Valentin, A. E., Penin, X., Chanut, J. P., Power, D., & Sévigny, J. M. (2014). Combining

microsatellites and geometric morphometrics for the study of redfish (*Sebastes* spp.)

population structure in the Northwest Atlantic. *Fisheries Research*, *154*, 102–119.

https://doi.org/10.1016/J.FISHRES.2014.02.008

Van Der Lingen, C. D., Weston, L. F., Ssempa, N. N., & Reed, C. C. (2015). Incorporating

parasite data in population structure studies of South African sardine *Sardinops sagax*.

Parasitology, *142*, 156–167. https://doi.org/10.1017/S0031182014000018

Vanhove, M. P. M., Hermans, R., Artois, T., & Kmentová, N. (2021). From the Atlantic coast

to Lake Tanganyika: Gill-infecting flatworms of freshwater pellonuline clupeid fishes in

West and Central Africa, with description of eleven new species and key to

Kapentagyrus (Monogenea, Dactylogyridae). *Animals*, *11*, 3578.

https://doi.org/10.3390/ANI11123578/S1

Wang, C., Torgerson, P. R., Höglund, J., & Furrer, R. (2017). Zero-inflated hierarchical

models for faecal egg counts to assess anthelmintic efficacy. *Veterinary Parasitology*,

235, 20–28. https://doi.org/10.1016/J.VETPAR.2016.12.007

Wickham, H. (2016). *ggplot2: Elegant Graphics for Data Analysis*. New York: Springer-

Verlag.

Wickham, H., Averick, M., Bryan, J., Chang, W., McGowan, L., François, R., … Yutani, H.

(2019). Welcome to the Tidyverse. *Journal of Open Source Software*, *4*, 1686.

https://doi.org/10.21105/joss.01686

Wilke, C. (2020). *ggtext: Improved Text Rendering Support for "ggplot2". R package version*

0.1.1.

Windsor, D. A. (1998). Controversies in parasitology, most of the species on Earth are

parasites. *Int J Parasitol*, *28,* 1939-1941. https://doi.org/10.1016/S0020-7519(98)00153-

2

Zuur, A. F., Ieno, E. N., Walker, N. J., Saveliev, A. A., & Smith, G. M. (2009). *Zero-Truncated*

and Zero-Inflated Models for Count Data. In Mixed effects models and extensions in

ecology with R. Statistics for Biology and Health (pp. 261-293).

https://doi.org/10.1007/978-0-387-87458-6_11