Tracking movements of an endangered bird using mark-recapture based on DNAtagging

María José Bañuelos ^{1*}, María Morán-Luis ², Patricia Mirol ³, Mario Quevedo^{1,4}

¹Department of Biology of Organisms and Systems, University of Oviedo. 33006 Oviedo, Spain

² IHCantabria - Instituto de Hidráulica Ambiental de la Universidad de Cantabria, PCTCAN. 39011 Santander, Spain

³ Grupo de Genética y Ecología en Conservación y Biodiversidad, Museo Argentino de Ciencias Naturales "Bernardino Rivadavia". C1405DJR Buenos Aires, Argentina

⁴ Biodiversity Research Institute – IMIB (University of Oviedo-CSIC-Principality of Asturias). 33600 Mieres, Spain

* Corresponding author. E-mail: banuelosmaria@uniovi.es. Tel: +34 98 510 4786

Keywords

Cantabrian capercaillie, DNA-tagging, grouse, leks, mating season, tracking movements, *Tetrao urogallus cantabricus*

Abstract

1 Knowing the location and movements of individuals at various temporal and spatial scales is an 2 important facet of behaviour and ecology. In threatened populations, movements that would en-3 sure adequate genetic flow and long term population viability are often challenged by habitat 4 fragmentation. It is also in those endangered populations where capturing and handling indi-5 viduals to equip them with transmitters or to obtain tissue samples may present additional logist-6 ical challenges. DNA-tagging, i.e. individual identification of samples obtained via non-invasive 7 approaches, can reveal certain movement patterns. We used faecal material genetically as-8 signed to individuals to indirectly track movements of a large-bodied, endangered forest bird, 9 Cantabrian capercaillie (Tetrao urogallus cantabricus), for three consecutive mating seasons. 10 We identified 127 individuals, and registered movements of 70 of them (22 females, 48 males). 11 Most movements were relatively short for capercaillie, mostly concentrated around display 12 areas. We did not find differences in movement distances between females and males within 13 mating seasons, or between them. Several longer, inter-valley movements up to 9.9 km of plani-14 metric distance linked distant display areas, showing that both females and males of Cantabrian 15 capercaillie were able to move through the landscape, complementing previous studies on gene 16 flow. Those longer movements may be taking birds outside of the study area, and into historical 17 capercaillie territories, which still include substantial forest cover. Tracking animals via DNA tag-18 ging, particularly those on endangered populations, showed clear advantages like non-intrusive-19 ness and potential for sample sizes much larger than via direct handling. However, it also 20 misses out on direct observation and natural history, which would provide key information like 21 social status and timing of movements.

Introduction

22 The study of movement patterns at various temporal and spatial scales provides basis for un-23 derstanding behaviour, ecology and, ultimately, aspects of the conservation status of most an-24 imals (Börger, Dalziel, & Fryxell, 2008). Some animal movements are tightly associated to their 25 specific habitat requirements, and in those cases navigating unfamiliar and unsuitable habitats 26 might increase risk exposure, compromising survival (Yoder et al. 2004; Bonte et al. 2012). As 27 consequence, the loss and fragmentation of habitat that affects most terrestrial ecosystems 28 (e.g. Watson et al. 2018) may result in appearance of dispersal barriers, which affect population 29 dynamics and gene flow via reduced connectivity (Ricketts 2001; Caplat et al. 2016). Indeed, 30 habitat loss and fragmentation of previous distribution ranges may yield smaller, subdivided 31 populations, if the loss of connectivity results in less frequent movements, and limits effective 32 dispersal. Those subdivided populations are more vulnerable to stochastic events that could re-33 strict their probability of survival (Lens et al., 2002).

34 The combination of the species' movement ability and the presence and shape of remnant hab-35 itat fragments together with the characteristics of the surrounding matrix, determine the species' 36 responses to landscape changes. Thus fine-scale movement behaviour of species of conserva-37 tion interest should be incorporated in conservation planning and management, particularly 38 when habitat restoration is feasible (Lechner et al., 2015). Yet, such dataset are not easy to ob-39 tain, and tend to be replaced by assumptions and simplifications (e.g. Southwell et al. 2008) de-40 rived from data from other populations, or similar species. However, movements are often par-41 tially determined by landscape characteristics, and by distinct selective pressures on different 42 populations (Baguette et al. 2013), so it is not always safe assuming that populations of a given 43 species would show equivalent movement patterns throughout their distribution range. For in-44 stance, the established knowledge on a species' behaviour and habitat requirements may have 45 been obtained from parts of its distribution range less affected by habitat loss and fragmenta-46 tion. Yet, in fragmented and peripheral areas of a distribution range, the quality of what may appear as secondary habitat is possibly key for the species' survival (Channel and Lomolino 2000; 47 48 Blanco-Fontao et al. 2010). This is particularly general for those areas within a species' range 49 where dispersal movements depend on fine-scale structural elements, acting as stepping 50 stones (Lechner et al. 2015).

51 We studied movements of an endangered forest bird, Cantabrian capercaillie *Tetrao urogallus* 52 *cantabricus*, living at the southern edge of the species range, under the influence of Atlantic cli-53 mate (Olson et al. 2001; Cervellini et al. 2020). This capercaillie population declined severely 54 from its known historical range in the last third of the 20th century (Pollo et al. 2005; Storch et 55 al. 2006), and its viability appears compromised (Bañuelos et al. 2019). It has been recently lis-56 ted as critically endangered in Spain (Ministerio para la Transición Ecológica 2018). Therefore, 57 to study them, it was advisable to use minimally intrusive techniques, like tagging of individuals

- 58 based on their DNA, left behind in shed tissues or scats. DNA tagging, where the repeated oc-
- 59 currence of a genotype is a direct indicator of movement, has been used to study movements at
- 60 various spatial and temporal scales (Palsbøll et al. 1997; Palsbøll 1999). It allows monitoring a
- 61 wider representation of the variability of individuals in a population, albeit with with fewer obser-
- 62 vations per individual than direct tracking techniques.

63 The remnant distribution range of Cantabrian capercaillie includes some relatively large patches 64 of forest and shrubland above the treeline, the latter being especially relevant for females with

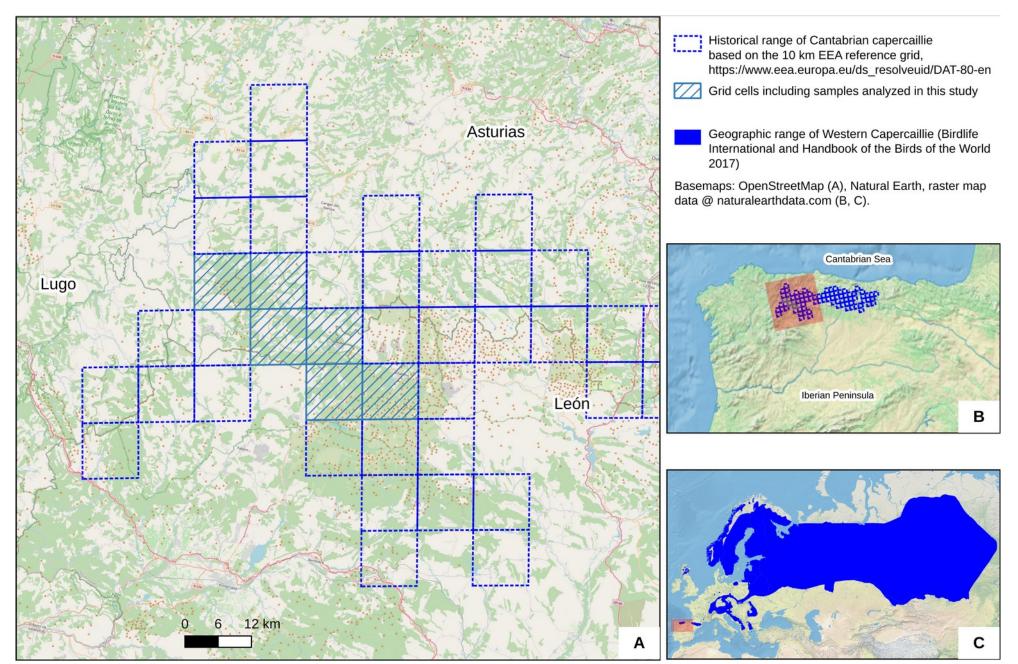
- 65 broods (Bañuelos et al. 2008). Yet, these habitats are embedded in a patchwork of disturbed
- 66 landscape, including former forest clearings and burns in various states of secondary succes-
- 67 sion, pastures, roads, villages, and industrial activities like mountaintop mining; several known
- 68 display areas are less than 500 meters from the edge of those mining operations. We wanted to
- 69 know how capercaillie moved in such landscape, and whether they would show regular, relat-
- 70 ively long movements.

Methods

Field survey

- 71 The study area (Fig. 1) lays within the region of broadleaf forests under Atlantic influence (e.g.
- 72 Olson et al. 2001). The area is rugged, with elevation ranging from 450 to 2000 m a.s.l. Larger
- 73 patches of relatively old forest are restricted to the higher slopes of the mountain range, mirror-
- 74 ing a widespread pattern of landscape configuration related to human impacts (Sandel and
- 75 Svenning 2013).
- Capercaillie gather at lekking areas in the mating season, thus making it more feasible to obtain sufficient droppings or feathers than at other times, when the birds are more dispersed and the habitat use of females and males is distinct (Bañuelos et al. 2008). We visited areas including 62 previously known capercaillie leks during the mating seasons (mid-March to early June) of 2009, 2010 and 2011 (Morán-Luis et al. 2014; Bañuelos et al. 2019 for details). Those reference locations of leks derived in fact from the period of capercaillie hunting during the mating season,
- 82 which extended for much of the 20th century (Rodríguez-Muñoz et al., 2015).

Figure 1. Study area



Of those leks, 83% showed signs of capercaillie presence at some point since year 2000; the
 rest had apparently been previously unused. The study area includes several largely forested
 areas with capercaillie presence, separated by valleys where human population and activity are
 mostly concentrated. Forest cover in the study area averaged 46%

87 Using the location of each lek from previous reports as reference, two people surveyed those 88 forest patches for 2-3 h. Locations where we did not find signs of capercaillie presence were 89 surveyed again 2-3 weeks later. The position of each sample was recorded with a GPS (± 5m) 90 and was later incorporated into a GIS. To minimize redundant samples and the risk of over-91 sampling the same individual, we selected samples based on a minimum distance of 25 m from 92 others with similar appearance (i.e. we weighed in freshness, size, shape and apparent content) 93 in the case of droppings, or from the same sex in the case of feathers. Droppings were stored in 94 tubes with silica-gel in the field, and were kept frozen at -20°C until DNA extraction. Feathers 95 were kept dry at room temperature. We assigned samples to display areas instead of the histor-96 ical leks used as reference because in some instances lek separation seemed an artifact, par-97 ticularly considering that capercaillie display do not conform to the classical lek notion (Wegge 98 et al. 2013).

DNA tagging and movements

99 We aimed at obtaining individual genotype profiles using nine microsatellite markers (five previ-100 ously developed for Tetrao urogallus, TUD2, TUD4, TUD5, TUT1, TUT3, Segelbacher et al. 101 2000; and four developed for the closely related Tetrao tetrix, TTD2, TTD6, BG10, BG15, Ca-102 izergues et al. 2001, Piertney and Höglund 2001), and a specific primer developed for sex as-103 signment for Cantabrian capercaillie (Pérez et al. 2011). Details of methods of DNA extraction 104 and amplification, molecular sexing, validation of genotype profiles and genotyping errors, were 105 detailed in Bañuelos et al. (2019). To ensure genotype reliability, we double-checked each gen-106 otype profile. We kept only those samples for which at least six microsatellite loci amplified cor-107 rectly, and which rendered a reliable and unequivocal consensus genotype (Morán-Luis et al. 108 2014, Bañuelos et al. 2019).

109 We recorded the repeated finding of capercaillie genotypes (hereafter *recaptures*) in success-110 ively processed samples. Note that the initial observation of an individual does not necessarily 111 imply its first presence in a lek, as scats cannot accurately be dated in this context. We used in-112 dividuals recaptured at least once for subsequent movement analyses. We focused particularly 113 on the maximum planimetric distance between recaptures of each individual, both within each 114 single mating season of 2009, 2010 and 2011, and between each mating season and the sub-115 sequent ones. We coded those recaptures as philopatric if they occurred within the same dis-116 play area.

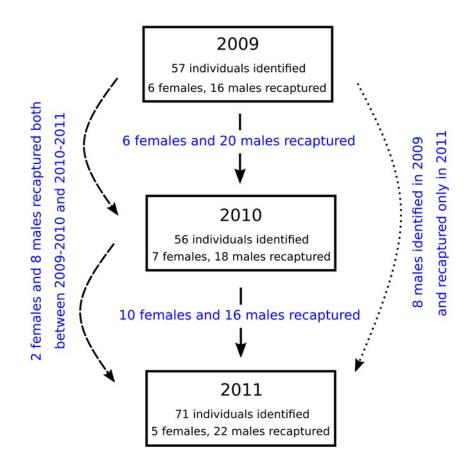
Results

- 117 We found 127 capercaillie individuals (genotypes) in the study area: 48 females, 73 males, and
- 118 6 individuals that could not be sexed. Those capercaillies were the minimum number of individu-
- als present in the area during the study. Individual identifications were derived from DNA extrac-
- 120 ted and amplified correctly in 422 samples, out of an initial set of 752 samples.
- 121 We found capercaillies in 35 display areas, which included 47 historical lek locations. In two of
- 122 those display areas we did not find males. The maximum number of males per display area was
- 123 7, with a median of 2. The maximum number of females per display area was 4 not coincident
- 124 with the display area with most males and the median was 1.
- 125 Seventy of the above capercaillies, 22 females (46%) and 48 (66%) males, were recaptured at
- 126 least once. We recaptured 53 individuals within any single mating season, and 49 individuals
- 127 between mating seasons (Fig. 2).

Figure 2. Scheme of recaptures

128 Scheme of capercaillie individuals identified and recaptured per season (boxes), recaptured in

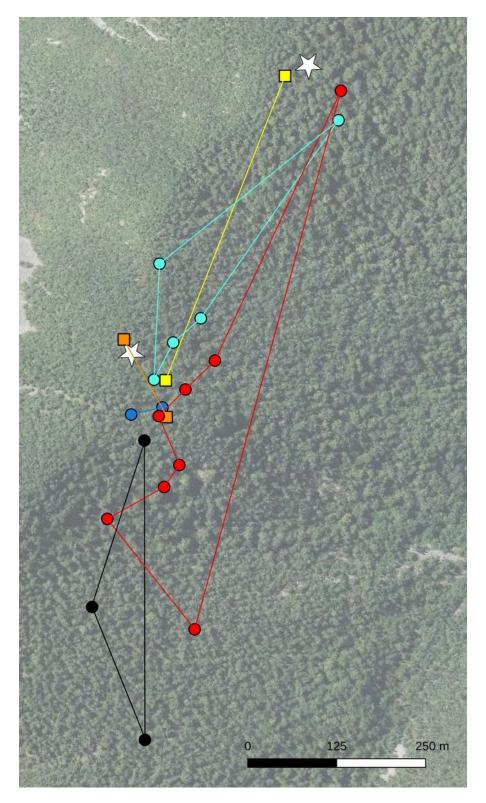
129 consecutive seasons (solid arrows), and recaptured between seasons.



- 131 Most recaptures occurred within 500 m of the initial observation, both within and between mat-
- 132 ing seasons (Fig. 3, 4). Distributions were highly skewed, including a few relatively long-dis-
- 133 tance movements (Fig. 4, 5). Table 1 shows descriptive statistics of maximum distances of re-
- 134 capture of each individual.

Figure 3. Pattern of philopatric movements within mating seasons

- 135 Example of the short movements that where the norm in a mating season. Symbols show loca-
- 136 tions of 4 capercaillie males (circles) and 2 females (squares) at a display area in 2009. Lines
- 137 show idealized tracks as minimum concave polygons for each individual. White stars mark the
- 138 reference location of two areas historically considered as leks. Basemap: aerial image from the
- 139 Spanish Geographic Institute (<u>https://pnoa.ign.es</u>).



Period	Sex	Recaptures	Median	Median absolute deviation	Range
Within mating seasons	female	18	297	252	63 - 3443
	male	56	297	341	29 - 7618
Between mating seasons	female	16	483	599	59 - 9895
	male	44	486	460	51 - 7673

140 There were no significant differences between females and males in maximum recapture dis-

tances within mating seasons (Wilcoxon rank sum test, W = 497, p = 0.947), or between matin

seasons (Wilcoxon rank sum test, W = 375, p = 0.701). Maximum recapture distances were lar-

143 ger between mating seasons (Wilcoxon rank sum test, W = 2804, p = 0.01).

144 The farthest recapture within any mating season was that of a male that moved 7.6 km between

145 two major valleys of the study area (Fig. 5, male 13). The farthest recapture between mating

seasons was that of a female re-identified in 2011 at 9.9 km from her previous location in 2010

147 (Fig. 5, female 19). That female visited different display areas each spring, changing sub-basins

148 of the study area.

In terms of philopatry, 79 % of male recaptures and 83% of female recaptures within mating
seasons occurred at the same display area. 73 % of male recaptures and 56 of female recap-

151 tures between mating seasons occurred at the same display area.

About one quarter of the birds (32% of males, 23% of females) visited more than one display area at some point. There were no differences between females and males in the proportion of recaptures registered in display areas different from the initial ($X^2 = 0.08$, df = 1, p = 0.78 within mating seasons; $X^2 = 1.47$, df = 1, p = 0.23 between mating seasons).

156 Most birds that visited more than one display area during a single mating season were found at 157 2 different ones (3 females, 10 males); only one male was recaptured at 3 different display 158 areas. These non-philopatric recaptures during the mating season occurred essentially between 159 display areas located within sub-basins of the study area. We also found a coincident, non-160 philopatric 3.5 km movement of a female and a male in the mating season of 2010 (Fig. 5, indi-161 viduals 61 and 65). A notable outlayer to the pattern just described corresponded to the male 162 13, mentioned above, which in the mating season of 2011 moved between two major valleys, 163 separated by deforested terrain in southern exposure, and other human-modified terrain.

Figure 4. Distance of recaptures

164 Violin plots (data points plus a probability density) showing frequency distribution of maximum

165 planimetric distance of recapture (m) for each capercaillie individual, both within and between 166 mating seasons.

- male
 Within mating seasons

 female
 Between mating seasons

 male
 0

 2500
 5000

 7500
 1000
- 167 Most birds shifting display areas between mating seasons were recaptured at 2 different ones
- 168 (5 females, 8 males), with just 1 female and 2 males recaptured at 3 different display areas.

169 These recaptures also occurred mostly within major valleys, although there were three recap-

170 tures bridging sub-basins. Interestingly, two of them from 2009 to 2010, and 2010 to 2011, be-

171 longed to the same 'travelling' male 13 referred above (Fig. 5).

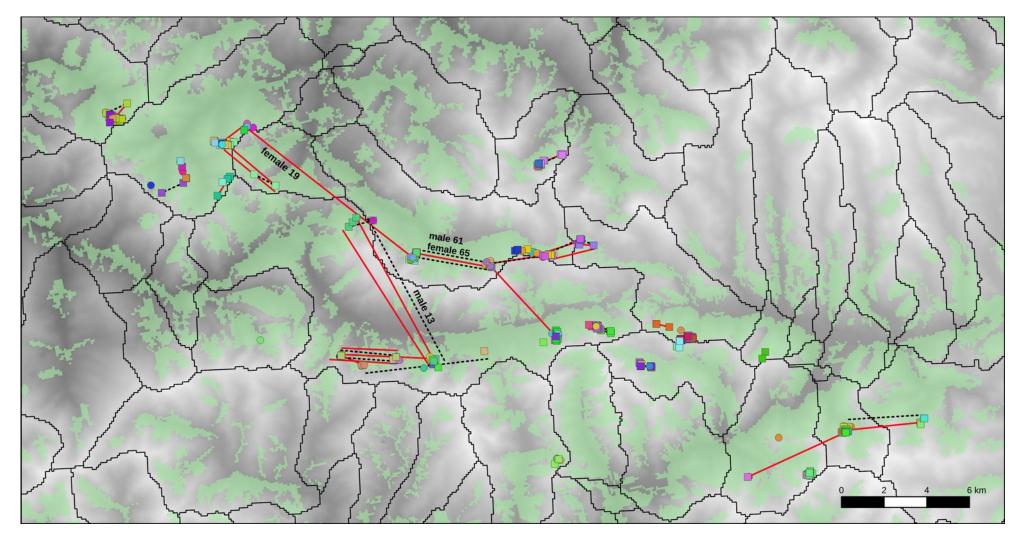
Figure 5. Non-philopatric recaptures and movements

172 Squares and circles show DNA-tagged male and female capercaillie, respectively. Lines show maximum straight line movements between recaptures of ca-

3 percaillie genotypes within (solid red) and between (dashed black) mating seasons. Overlapping movements were slightly displaced for clarity. Polygons show

4 sub-basins* over a digital elevation model of the study area where lighter hues indicate higher elevation (range 357 to 2094 m a.s.l.). *Ministerio para la

175 Transición Ecológica y el Reto Demográfico [https://www.miteco.gob.es/en/cartografia-y-sig/ide/descargas/agua/cuencas-y-subcuencas.aspx]



173 174

Discussion

176 Surveying display areas during three consecutive breeding seasons, we tracked different caper-177 caillie individuals in space and time. DNA-tagging recaptures provided a view of the use of 178 space of individuals of this endangered population that we did not previously have. We found 179 high capercaillie fidelity to display areas, but also rarer and much longer movements. Birds ten-180 ded to remain in one display area in spring, although a quarter of the individuals did visit 2 or 3 181 display areas. In males, the latter seemed consistent with expected exploratory movements into 182 other males' territories (Wegge et al. 2013). Overall the prevalence of short movements in the 183 mating season agrees with previous studies in capercaillie (e.g. Wegge and Larsen 1987, 184 Gjerde et al. 2000), and more broadly in lekking grouse species (Cross et al. 2017). However, 185 we expected that in this endangered population, inhabiting a fragmented forest landscape, fe-186 males visited several leks in spring due to low availability of males, as suggested in the case of 187 inter-lek movements of females in the Alps (Storch 1997). Overall, the observed pattern of 188 movement within mating seasons implies that individuals used several historically recognised 189 leks, in a pattern that reminded of the exploded leks notion (Wegge et al. 2013). Many of those 190 registered leks may in fact be merged into fewer display areas, at least in terms of planning sur-191 vey design and efforts.

Despite the prevalence of philopatric movements, non-philopatric recaptures over much longer distances showed that both females and males of Cantabrian capercaillie were able to move regularly through the landscape at ecological time scales. These results complement previous data on gene flow and thus longer time scales, which showed that there was no genetic subdivision among individuals living in relatively distant sub-basins of the landscape (Fameli et al. 2017). We did not find longer movements by females, despite that previous studies showed females dispersing farther, and moving more frequently between mating seasons (Segelbacher et

199 al. 2008; Watson and Moss 2008).

200 Our study was focused on relatively small-scale movements. Studies that explicitly attempted to 201 cover dispersal would in principle obtain much lower recapture rates, but would also identify 202 longer movement events (Koenig et al. 1996; Cross et al. 2017). It is conceivable that year-203 round movements would be longer than those we detected in the mating season (Saniga 2006, 204 Zizas et al. 2012). Possibly those would also be directed towards alternative habitats (e.g. Wat-205 son and Moss 2008), and could take birds outside of the study area, which is nested among his-206 torical capercaillie territories (Fig. 1). Thus, those historical territories, which today could be per-207 ceived as peripheral but which include substantial forest cover, should also be protected and 208 monitored, as an objective for the potential recovery of the Cantabrian capercaillie population.

209 A related question is where were the birds that we did not recapture. Some of them were likely 210 in the same area, and were not detected in our fieldwork, as reflected by mark-recapture mod-211 els of population dynamics (Bañuelos et al. 2019). But there are other, non-exclusive possibilit-212 ies. We had previously found that females showed higher turnover in the study area, and yiel-213 ded less recaptures (Morán-Luis et al. 2014; Bañuelos et al. 2019). Those results could be re-214 lated to lower female survival, but also to more frequent dispersal events that remained un-215 detected in our sampling schemes. It is possible that birds kept dispersing outside of the study 216 area, which we selected for its consistent capercaillie presence in the recent decades. Data 217 from Russian capercaillie populations suggested that younger birds switched an area of primary 218 habitat for lower quality, logged neighbouring patches, due to competition with dominant adults, 219 and some returned later as adults (Borchtchevski 1993). An equivalent pattern of sources and 220 sinks in capercaillie local populations has been described in the Bavarian Alps (Segelbacher et 221 al. 2003). We speculate that a similar process could be taking place in our study area, which in-222 cludes higher forest cover and quality than the adjacent landscape (Quevedo et al. 2006). How-223 ever, evaluating that idea would require regular, formal surveying efforts, and we are not aware 224 of any. Capercaillie management in the Cantabrian Mountains has paid little if any attention to 225 forest patches outside the historically known display areas. Moreover, as the presence of caper-226 caillie in spring declined in historically known territories, attention has been placed on a sub-227 sequently smaller fraction of those, without monitoring formerly used forests, or potential new 228 locations.

Ups and downs of studying movements with DNA tagging

229 Comparing different aspects of tracking animals based on DNA-tagging vs. direct tracking is not 230 straightforward, yet we saw both pros and cons of DNA-tagging: We identified 127 individuals, 231 and registered movements for 70 of them. Such indirect tracking thus provided a much larger 232 sample size of *tagged* individuals distances than could possibly be obtained by a standard re-233 search team working with endangered birds, and under standard Spanish research funding. In 234 addition, DNA-tagging based on scats is a non invasive technique, particularly suited to individu-235 als of endangered populations, for which handling stress would be a serious problem (Gibson et 236 al. 2013; Blomberg et al. 2018). Results may also be less influenced by behavioral bias, like 237 avoidance of the observer, or response to handling and initial mark (Miller et al. 2005; Biro and 238 Dingemanse 2009; Garamszegi et al. 2009).

239 On the other hand, what we gained in number of individuals identified, we certainly missed in

240 the number of data points per individual. Present-day GPS tracking technology could provide a

- 241 much larger datasets, thus allowing detailed inferences (e.g. de Gabriel Hernando et al. 2020).
- Another shortcoming is that using DNA-tagging based on shed tissues or scats it is not possible

- 243 detailing the sequence of the initial and subsequent "captures"; thus it is not possible tracing
- 244 precise movement paths. In addition, we lack direct information on the natural history of the
- ²⁴⁵ "tagged" birds. We lacked for instance important individual information like age and social
- status; aspects that would have been useful to interpret whether the registered movements
- 247 were dispersive or exploratory. Or to check whether results were biased to adult males. Such
- bias is certainly possible, because the presence of subordinate males may be more difficult to
- 249 detect in spring (Wegge and Larsen 1987), even though we had it in mind, and included as
- 250 much surveyed terrain as possible in the periphery of reference lek locations.

Acknowledgments

- 251 The study was funded by grants IB08-158 (FICYT, Asturian Government) to MJ Bañuelos and
- 252 CGL2010-15990 (MICINN, Spanish Government) to MJ Bañuelos and M Quevedo. We thank
- 253 Alberto Fernández-Gil, Bea Blanco Fontao, and Rolando Rodríguez Muñoz for their help in
- designing and conducting field surveys. Eduardo González, Víctor Rodríguez, and Damián
- 255 Ramos helped with the fieldwork, and Jose Carral helped with field logistics. Alberto Fameli
- greatly helped fine-tuning lab methods, and Leticia Viesca helped analysing individual geno-
- types. We acknowledge the permits granted by the environmental authorities of Asturias and
- León, required to survey capercaillie mating areas.

References

- 259 Baguette M, Blanchet S, Legrand D, Stevens VM, Turlure C. 2013. Individual dispersal,
- landscape connectivity and ecological networks. Biological Reviews 88(2):310–326.
- 261 doi:10.1111/brv.12000.
- Bañuelos MJ, Blanco-Fontao B, Fameli A, Fernández-Gil A, Mirol P, Morán-Luis M, RodríguezMuñoz R, Quevedo M. 2019. Population dynamics of an endangered forest bird using mark–
 recapture models based on DNA-tagging. Conservation Genetics 20(6):1251–1263.
 doi:10.1007/s10592-019-01208-x.
- 266 Bañuelos MJ, Quevedo M, Obeso JR. 2008. Habitat partitioning in endangered Cantabrian
- 267 capercaillie *Tetrao urogallus cantabricus*. Journal of Ornithology 149:245–252.
 268 doi:10.1007/s10336-007-0267-5.
- Biro PA, Dingemanse NJ. 2009. Sampling bias resulting from animal personality. Trends in
 Ecology & Evolution 24(2):66–67. doi:10.1016/j.tree.2008.11.001.
- Blanco-Fontao B, Fernández-Gil A, Obeso J, Quevedo M. 2010. Diet and habitat selection in
 Cantabrian Capercaillie (*Tetrao urogallus cantabricus*): ecological differentiation of a rear-
- edge population. Journal of Ornithology 151(2):269–277. doi:10.1007/s10336-009-0452-9.
- Blomberg E, Davis S, Mangelinckx J, Sullivan K. 2018. Detecting capture-related mortality in

275 radio-marked birds following release. Avian Conservation and Ecology 13(1).

276 doi:10.5751/ACE-01147-130105.

- 277 Bonte D, Van Dyck H, Bullock JM, Coulon A, Delgado M, Gibbs M, Lehouck V, Matthysen E, 278 Mustin K, Saastamoinen M. 2012. Costs of dispersal. Biological Reviews 87(2):290-312. doi:
- 279 10.1111/j.1469-185X.2011.00201.x
- 280 Borchtchevski VG. 1993. Population biology of the capercaillie. Principles of the structural 281 organization [in Russian; tables and plots in English]. Moscow: Central Laboratory of the 282 Management of Hunting and Nature Reserves.
- 283 Börger L, Dalziel BD, Fryxell JM. 2008. Are there general mechanisms of animal home range 284 behaviour? A review and prospects for future research. Ecology Letters 11(6):637-650. 285 doi:10.1111/j.1461-0248.2008.01182.x.
- 286 Caizerques A, Dubois S, Loiseau A, Mondor G, Rasplus J-Y. 2001. Isolation and 287 characterization of microsatellite loci in black grouse (Tetrao tetrix). Molecular Ecology Notes 288 1(1-2):36-38. doi:10.1046/j.1471-8278.2000.00015.x.
- 289 Caplat P, Edelaar P, Dudaniec RY, Green AJ, Okamura B, Cote J, Ekroos J, Jonsson PR, 290 Löndahl J, Tesson SV, et al. 2016. Looking beyond the mountain: dispersal barriers in a 291 changing world. Frontiers in Ecology and the Environment 14(5):261-268.
- 292 doi:10.1002/fee.1280.
- 293 Cervellini M, Zannini P, Musciano MD, Fattorini S, Jiménez-Alfaro B, Rocchini D, Field R, 294 Vetaas OR, Irl SDH, Beierkuhnlein C, et al. 2020. A grid-based map for the Biogeographical 295 Regions of Europe. Biodiversity Data Journal 8:e53720. doi:10.3897/BDJ.8.e53720.
- 296 Channel R, Lomolino MV. 2000. Dynamic biogeography and conservation of endangered 297 species. Nature 403 (6765):84-86. doi:10.1038/47487.
- 298 Cross TB. Naugle DE, Carlson JC. Schwartz MK. 2017. Genetic recapture identifies long-299 distance breeding dispersal in Greater Sage-Grouse (Centrocercus urophasianus). The 300 Condor 119(1):155-166. doi:10.1650/CONDOR-16-178.1.
- 301 de Gabriel Hernando M, Karamanlidis AA, Grivas K, Krambokoukis L, Papakostas G, Beecham 302 J. 2020. Reduced movement of wildlife in Mediterranean landscapes: a case study of brown
- 303 bears in Greece. Journal of Zoology 311(2):126-136. doi: 10.1111/jzo.12768
- 304 Fameli A, Morán-Luis M, Rodríguez-Muñoz R, Bañuelos MJ, Quevedo M, Mirol P. 2017.
- 305 Conservation in the southern edge of Tetrao urogallus distribution: Gene flow despite
- 306 fragmentation in the stronghold of the Cantabrian capercaillie. European Journal of Wildlife
- 307 Research 63(3). doi:10.1007/s10344-017-1110-9.
- 308 Garamszegi LZ, Eens M, Török J. 2009. Behavioural syndromes and trappability in free-living 309 collared flycatchers, Ficedula albicollis. Animal Behaviour 77(4):803-812.

- doi:10.1016/j.anbehav.2008.12.012.
- Gibson D, Blomberg EJ, Patricelli GL, Krakauer AH, Atamian MT, Sedinger JS. 2013. Effects of
- radio collars on survival and lekking behavior of male Greater Sage-Srouse. The Condor.
- 313 115(4):769–776. doi:10.1525/cond.2013.120176.
- Gjerde I, Wegge P, Rolstad J. 2000. Lost hotspots and passive female preference: the dynamic
 process of lek formation in capercaillie *Tetrao urogallus*. Wildlife Biology 6(4):291–298.
 doi:10.2981/wlb.2000.029.
- 317 Koenig WD, Van Vuren D, Hooge PN. 1996. Detectability, philopatry, and the distribution of
- dispersal distances in vertebrates. Trends in Ecology & Evolution 11(12):514–517.
- doi:10.1016/S0169-5347(96)20074-6.
- Lechner AM, Doerr V, Harris RMB, Doerr E, Lefroy EC. 2015. A framework for incorporating
 fine-scale dispersal behaviour into biodiversity conservation planning. Landscape and Urban
 Planning 141:11–23. doi:10.1016/j.landurbplan.2015.04.008.
- Lens L, Van Dongen S, Norris K, Githiru M, Matthysen E. 2002. Avian persistence in
 fragmented rainforest. Science 298(5596):1236–1238. doi:10.1126/science.1075664.
- Miller CR, Joyce P, Waits LP. 2005. A new method for estimating the size of small populations
 from genetic mark–recapture data. Molecular Ecology 14(7):1991–2005. doi:10.1111/j.1365 294X.2005.02577.x.
- 328 Ministerio para la Transición Ecológica. 2018. Declaración de situación crítica de *Cistus*
- 329 heterophyllus subsp. carthaginensis, Lanius minor, Margaritifera auricularia, Marmaronetta

330 angustirostris, Mustela lutreola, Pinna nobilis y Tetrao urogallus cantabricus en España.

- 331 https://www.boe.es/eli/es/o/2018/09/28/tec1078.
- Morán-Luis M, Fameli A, Blanco-Fontao B, Fernández-Gil A, Rodríguez-Muñoz R, Quevedo M,
 Mirol P, Bañuelos MJ. 2014. Demographic status and genetic tagging of endangered
- capercaillie in NW Spain. PLOS ONE 9(6):e99799. doi:10.1371/journal.pone.0099799.
- 335 Olson DM, Dinerstein E, Wikramanayake ED, Burgess ND, Powell GVN, Underwood EC,
- D'amico JA, Itoua I, Strand HE, Morrison JC, et al. 2001. Terrestrial ecoregions of the world:
- a new map of life on Earth. BioScience 51(11):933. doi:10.1641/0006-
- 338 3568(2001)051[0933:TEOTWA]2.0.CO;2.
- Palsbøll PJ. 1999. Genetic tagging: contemporary molecular ecology. Biological Journal of the
 Linnean Society 68(1–2):3–22. doi:10.1111/j.1095-8312.1999.tb01155.x.
- 341 Palsbøll PJ, Allen J, Bérubé M, Clapham PJ, Feddersen TP, Hammond PS, Hudson RR,
- Jørgensen H, Katona S, Larsen AH, et al. 1997. Genetic tagging of humpback whales.
- 343 Nature 388(6644):767–769. doi:10.1038/42005.
- Pérez T, Vázquez JF, Quirós F, Domínguez A. 2011. Improving non-invasive genotyping in

- 345 capercaillie (*Tetrao urogallus*): redesigning sexing and microsatellite primers to increase
- 346 efficiency on faeces samples. Conservation Genetics Resources 3(3):483–487.

347 doi:10.1007/s12686-011-9385-8.

- Piertney SB, Höglund J. 2001. Polymorphic microsatellite DNA markers in black grouse (*Tetrao tetrix*). Molecular Ecology Notes 1(4):303–304. doi:10.1046/j.1471-8278.2001.00118.x.
- Pollo C, Robles L, Seijas JM, García-Miranda A, Otero R. 2005. Trends in the abundance of

351 Cantabrian Capercaillie *Tetrao urogallus cantabricus* at leks on the southern slope of the

352 Cantabrian Mountains, north-west Spain. Bird Conservation International 15(04):397–409.

- 353 doi:10.1017/S0959270905000626.
- Quevedo M, Bañuelos MJ, Obeso JR. 2006. The decline of Cantabrian capercaillie: How much
 does habitat configuration matter? Biological Conservation 127(2):190–200.
- 356 doi:10.1016/j.biocon.2005.07.019.
- Ricketts TH. 2001. The matrix matters: effective isolation in fragmented landscapes. The
 American Naturalist 158(1):87–99. doi:10.1086/320863.
- Rodríguez-Muñoz R, del Valle CR, Bañuelos MJ, Mirol P. 2015. Revealing the consequences of
 male-biased trophy hunting on the maintenance of genetic variation. Conservation Genetics.
 16:1375–1394. doi:10.1007/s10592-015-0747-8.
- Sandel B, Svenning J-C. 2013. Human impacts drive a global topographic signature in tree
 cover. Nature Communications 4. doi:10.1038/ncomms3474.
- Saniga M. 2006. Home range sizes and roosting places in capercaillie (*Tetrao urogallus* L.)
 cocks living solitary in the West Carpathians. Folia Oecologica 33(2):121–128.
- 366 Segelbacher G, Manel S, Tomiuk J. 2008. Temporal and spatial analyses disclose
- consequences of habitat fragmentation on the genetic diversity in capercaillie (*Tetrao urogallus*). Molecular Ecology 17:2356–2367. doi:10.1111/j.1365-294X.2008.03767.x.
- 369 Segelbacher G, Paxton RJ, Steinbrück G, Trontelj P, Storch I. 2000. Characterization of
- microsatellites in capercaillie *Tetrao urogallus* (Aves). Molecular Ecology 9:1934–1936.
 doi:10.1046/j.1365-294x.2000.0090111934.x.
- Segelbacher G, Storch I, Tomiuk J. 2003. Genetic evidence of capercaillie *Tetrao urogallus* dispersal sources and sinks in the Alps. Wildlife Biology 9(4):267–273.
- doi:10.2981/wlb.2003.014.
- 375 Southwell DM, Lechner AM, Coates T, Wintle BA. 2008. The sensitivity of population viability
- analysis to uncertainty about habitat requirements: implications for the management of the
- endangered southern brown bandicoot. Conservation Biology 22(4):1045–1054.
- 378 doi:10.1111/j.1523-1739.2008.00934.x.
- 379 Storch I. 1997. Male territoriality, female range use, and spatial organisation of capercaillie

- 380 *Tetrao urogallus* leks. Wildlife Biology 3(3–4):149–161. doi:10.2981/wlb.1997.019.
- 381 Storch I, Bañuelos MJ, Fernández-Gil A, Obeso JR, Quevedo M, Rodríguez-Muñoz R. 2006.
- 382 Subspecies Cantabrian capercaillie *Tetrao urogallus cantabricus* endangered according to 383 IUCN criteria. Journal of Ornithology 147(4):653–655. doi:10.1007/s10336-006-0101-5.
- Watson A, Moss R. 2008. Grouse: The Natural History of British and Irish Species. 1st edition.
 London: Collins.
- 386 Watson JEM, Venter O, Lee J, Jones KR, Robinson JG, Possingham HP, Allan JR. 2018.
- 387 Protect the last of the wild. Nature 563(7729): 27. doi:10.1038/d41586-018-07183-6.
- Wegge P, Larsen B. 1987. Spacing of adult and subadult male common Capercaillie during the
 breeding season. The Auk 104:481–490. doi:10.2307/4087547.
- Wegge P, Rolstad J, Storaunet KO. 2013. On the spatial relationship of males on 'exploded'
- 391 leks: the case of Capercaillie grouse *Tetrao urogallus* examined by GPS satellite telemetry.
- 392 Ornis Fennica 90(4):222–235.
- 393 Yoder JM, Marschall EA, Swanson DA. 2004. The cost of dispersal: predation as a function of
- 394 movement and site familiarity in ruffed grouse. Behavioral Ecology 15(3):469–476.
- 395 doi:10.1093/beheco/arh037.
- 396 Zizas R, Shamovich D, Kurlavičius P, Belova O, Brazaitis G. 2012. Radio-tracking of
- 397 Capercaillie (*Tetrao urogallus* L.) in North Belarus. Baltic Forestry 18(2):270–277.