Hospital and urban wastewaters shape the structure and active resistome of environmental biofilms

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Abstract

We studied hospital and urban wastewater (WW) biofilms *in situ* and in an *in vitro* laboratory model to

assess and compare their potential to actively disseminate and express ARGs. Biofilms were subjected to

confocal microscopy to qualify and quantify their extracellular polymeric substances (EPSs), including

polysaccharides, external DNA (eDNA), and lipids. We observed that antibiotic challenge had a differential

impact on the EPS and the live-dead ratios of *in vitro* hospital and urban WW biofilms. Differential

expression of a range of antibiotic resistance genes (ARGs) and mobile genetic elements (MGEs), including

those on plasmids, was assessed using a meta-transcriptomic approach. Expression of ARGs as well as

MGEs, *in situ,* was 10-fold higher in hospital WW biofilms compared with urban, suggesting a higher

frequency of ARG transfer might occur in hospital WW. *In vitro* data confirmed that the addition of

ciprofloxacin significantly increased ARG and MGE expression in hospital but not in urban WW biofilms

suggesting that the regulation of ARG and MGE expression is likely dependent on the structure and

composition of the biofilm. The different responses to antibiotic pressure in hospital *vs* urban biofilms,

coupled with differences in biofilm structure helps delineate distinct characteristics of hospital and urban

WW biofilms.

Introduction

Biofilms are the preferred lifestyle of bacteria in the natural environment. They present structures that support high densities of bacteria, exhibit high tolerance to antimicrobials, and facilitate gene transfer (1–3). In aquatic environments, biofilms are suggested to serve as hot spots for the accumulation and dissemination of antimicrobial resistance genes (ARG) (1,4). Chemical pollutants spread by urban, industrial, and agricultural waste exert a continuous selective pressure on the diverse antibiotic resistance gene (ARG) pool present in the natural environment and its microbial communities (5–7). For example, ciprofloxacin is a very stable molecule often found in anthropogenically polluted environments and enriched in wastewater (WW), particularly hospital WW (7–10). Ciprofloxacin is not efficiently removed by WW treatment through activated sludge (11); it can be detected in WW and downstream aquatic environments and was identified to pose a significant risk for selection of antimicrobial resistance (AMR) in these environments (9,10,12). Furthermore, ciprofloxacin is known to induce the SOS-response in Gram-negative bacteria. This induction of the SOS response can lead to the expression of antibiotic resistance genes as the quinolone resistance-encoding *qnrB* gene, or the expression of genetic elements as the class 1 integron integrase gene, leading to ARG cassette rearrangements (13,14). In addition, bacteria that are present in the human gut carry a wide variety of ARGs that can be spread into the environment by human feces through WW (15–18). To limit the dissemination of chemical pollutants, pathogens and ARGs into the environment, the implementation of WW treatment systems worldwide plays a key role (6,19–22). However, WW treatment plants are also regarded as selection and dissemination hot spots due to the high density of human gut, environmental bacteria, and chemical pollutants (22–24). Biofilms in WW systems represent an important source of ARGs and bacteria that may end up circulating in the environment and hence present an important model to study the dissemination of resistance in the context of discharged effluents (4,25,26). Biofilms dominate most habitats on earth, accounting for ~ 80% of all bacterial and archaeal cells (27). Biofilm communities are typified by high densities of bacteria embedded in a self-produced protective matrix, which contributes to high antibiotic tolerance, and facilitates gene transfer (4,28). Extracellular polymeric substances (EPS), such as external DNA (eDNA), proteins, lipids, and polysaccharides, are essential structural components of the biofilm matrix that are involved in resilience to antibiotics and exogenous stress, but also in horizontal gene transfer (HGT) (29–31). Previous studies have shown that hospital WW contain significantly higher amounts of resistant bacteria, ARGs and mobile genetic elements (MGEs), as well as chemical and pharmaceutical pollutants compared to urban WW. However, these quantities are generally diluted when hospital WW is mixed with urban WW in community sewer systems (7,32,33).

Demonstration of the transfer, dynamics, and regulation of ARGs and MGEs in a complex environmental

matrix is yet experimentally challenging, with many essential open questions such as how and when

transfer and dissemination of ARGs occur in nature. The extent and conditions of MGE transfer that carry

ARGs is still largely unexplored in natural environments and microbial communities. In this study we

implemented a sampling and analysis approach that allows phenotypic and genomic analyses of *in situ*

and reconstituted *in vitro* hospital and urban wastewater (WW) biofilms. To better understand the role of

hospital and urban WW biofilms in actively disseminating antimicrobial resistance in the WW system, we

explored the extracellular polymeric substances (EPS) within the biofilm matrix, the biofilm composition

(microbiota), the resistome (antibiotic, biocide, and heavy metal resistance genes) and mobilome (mobile

genetic elements), with focus on expressed genes (metatranscriptome).

Materials and Methods

Sample collection and biofilm production *in situ* and *in vitro*:

Wastewater (WW) biofilm production campaigns (*in vitro* and *in situ*) were conducted in January 2017.

For *in vitro* biofilms, 20L of untreated urban (up flow of the WWTP Limoges city), and hospital (900-bed,

Limoges teaching hospital France) WW were collected on the same day and transported (at 4°C) to the

laboratory for immediate use in continuous biofilm reactors. The remaining WW was stored at 4°C to

replace evaporated volume every two days. For each WW type, two reactors containing 3L of WW were

fed at a continuous rate, from an external container containing the same WW (5L in total), ensuring

circulation and a stable level in the biofilm reactors (closed continuous system) (Supplementary Figure 1).

Rooms were kept darkened, and the temperature was maintained at 17°C. Sixteen polystyrene slides

suitable for confocal laser scanning microscopy (CLSM) (Agar Scientific LTD, UK) were attached to a rotator

that was constantly turning, producing 16 biological replicate biofilms per reactor. After 3 days of initial

biofilm formation, 1µg/L of ciprofloxacin (considered as the minimum selective concentration (MSC) of

ciprofloxacin for most environmental bacteria (34) was added to one reactor for each biofilm type (urban,

hospital, lab duplicate). On day 7, the urban and hospital biofilms grown in the reactors containing

ciprofloxacin were challenged with a fresh dose of ciprofloxacin (1 µg/L) and harvested the same day (D7)

4 hours after ciprofloxacin challenge. Samples were processed within 30 minutes after harvesting. For

metagenomics analysis, 8 biological replica biofilms were scraped with tissue cell scrapers (Biologix®) from

the slides and washed/suspended in 1X phosphate buffered saline solution (PBS), transferred in 2ml

Eppendorf tubes, and immediately shock frozen in liquid nitrogen. Samples were stored at -80°C until

RNA/DNA extraction. For CLSM analysis, the remaining 8 slides were washed gently once with 1x PBS and

either directly stained (for life/dead staining) or fixed in 3.7% formaldehyde solution for 1 hour and

preserved for subsequent EPS staining.

For *in situ* biofilms, holes were drilled into polystyrene slides ends and attached with fishing line (16 slides

in total, Supplementary Figure 1). The string of polystyrene slides was immersed into the hospital WW

pipe and attached at the access point. Slides were recovered after 7 days. Slides were separated and

carefully collected in a slide box containing cooled 1xPBS solution and transported to the laboratory

immediately for processing. Slides that showed sufficient biofilm formation (by eye) and were least

polluted with residues of tissue and other large waste debris, were immediately scraped from the slides

and shock frozen in liquid nitrogen as described above. Remaining slides were washed gently in 1xPBS and

fixed in 3.7% formaldehyde solution for 1 hour for subsequent EPS staining and CLSM analysis. The same

procedure was applied for the *in situ* biofilm production in the untreated urban WW directly at the WWTP

(slides were attached and emerged in the large WW receiving pipe/body after primary water filtration

step (removal of large debris)). Slides were recovered as described above and stored in a slide box

containing cooled 1xPBS. At least four suitable biofilm slides/replicates were directly processed on site by

scraping them from the slides and shock freezing them in a portable liquid nitrogen tank. The remaining

slides were transported at 4°C to the laboratory and fixed in 3.7% formaldehyde solution for subsequent

EPS staining and CLSM analysis as described below.

Sample processing:

DNA extraction was performed with the DNA/RNA dual extraction kit (Qiagen) according to the

manufacturer's instructions and stored at -20°C. RNA quality, as measured by RNA integrity number (RIN)

was insufficient for RNA sequencing. Therefore, RNA was extracted from remaining biological replicates

and samples using the ZymoBIOMICS DNA/RNA Miniprep Kit. DNA concentration was determined by

Qubit Fluorometric Quantitation (Thermo fisher scientific, Waltham, MA USA) assays according to the

manufacturer's instructions. All DNA samples were diluted or concentrated to a final concentration of 10

ng/µl for downstream qPCR and 16S rRNA analysis. RNA concentration and integrity were analyzed by the

Agilent 2100 Bioanalyzer system.

High-throughput qPCR for clinical resistome analysis:

Nanolitre-scale quantitative PCRs to quantify genes that confer resistance to antimicrobials were

performed as described previously (7,32,35). We targeted 76 genes, grouped into 16 resistance gene

classes; targeted genes include ARGs that are prevalent in the gut microbiota of healthy individuals (15),

clinically relevant ARGs (extended-spectrum β-lactamases (ESBLs), carbapenemases, and vancomycin

resistance), and heavy metals and quaternary ammonium compounds resistance genes suggested to favor

cross and co–selection for ARGs in the environment (36,37). We also targeted 11 other targets, including

common transposase gene families and class 1, 2 and 3 integron integrase genes, that are important

vectors for ARGs in the clinics and often used as proxy for anthropogenic pollution (38). This set of 87

genes constituted what we named the targeted resistome (Supplementary Table 1). We also included

primers targeting 16S rDNA. Primer design and validation prior to and after Biomark analysis has been

done as described earlier (7,35). Real-Time PCR analysis was performed using the 96.96 BioMark™

Dynamic Array for Real-Time PCR (Fluidigm Corporation, San Francisco, CA, U.S.A) as described previously

(7). Thermal cycling and real-time imaging were performed at the Plateforme Génomique GeT – INRA

Transfert (https://get.genotoul.fr/en/), and Ct values were calculated using the BioMark Real-Time PCR

analysis software. Calculations for normalized and cumulative abundance of individual genes and

allocated gene classes were performed as previously described (7,32,35). The resistome data displayed

here and used for statistical analysis is based on resistome data that was obtained for two biological

replicates per sample.

16S rRNA gene sequencing and sequence data pre-processing

Extracted DNA samples for 16S rRNA sequencing were prepared following a dual barcoded two-step PCR

procedure for amplicon sequencing for Illumina. The V4 region was sequenced by generating 2 x 300 bp

on an Illumina MiSeq. Data was demultiplexed and filtered for quality and remaining primers according to

recommended practices as described previously (7).

16S rRNA data analysis

Illumina MiSeq forward and reverse reads were processed using the MASQUE pipeline

(https://github.com/aghozlane/masque). Briefly, raw reads are filtered and combined followed by

dereplication (dereplication is the identification of unique sequences so that only one copy of each

sequence is reported) Chimera removal and clustering are followed by taxonomic annotation of the

resulting OTUs by comparison to the SILVA database. A BIOM file is generated that combines both OTU

taxonomic assignment and the number of matching reads for each sample. Relative abundance levels for

bacterial taxa (family/genus level) were obtained and analyzed (7).

Metatranscriptomics and RNA sequencing:

Cell pellets were re-suspended in TRIzol reagent and subjected to three cycles of 1min bead-beating (0.5mm silica/zirconia beads) followed by RNA extraction using the Direct-zol RNA Kit (Zymo research) according to the manufactures' protocol. DNA was removed by TURBO DNase (Ambion) until no genomic DNA was detected by PCR. The quality and integrity of the total RNA were checked on the Bioanalyzer system (Agilent). Because RNA yield is low, only 6 samples containing the highest amount of RNA and best RIM values (>6) were selected for RNA sequencing and meta-transcriptomic analysis (data not shown). Ribosomal RNA depletion was performed using the Bacteria RiboZero kit (Illumina). From rRNA-depleted RNA, directional libraries were prepared using the TruSeq Stranded mRNA Sample preparation kit following the manufacturer's instructions (Illumina). Libraries were quantified on Bioanalyzer DNA chips (Agilent) and Qubit® dsDNA HS Assay Kit (ThermoFisher). 150 bp paired read sequences were generated on the Nextseq 500 (high output) sequencer according to manufacturer's instructions (Illumina). The multiplexing level was 10 samples per lane. Reads were cleaned for low-quality and adapter sequences using Cutadapt version 1.11.

The level of ribosomal RNAs (rRNAs) remained high in our environmental samples, even after both *in vitro* rRNA depletion step and optimized RNA extraction methods (see above). We tackled these technical shortcomings by including an *in-silico* rRNA depletion step to improve mRNA read assembly and subsequent analysis (Supplementary Fig. 2 and Supplementary Tables 2 and 3). Ribosomal RNA reads were filtered out by mapping on the Silva database with BWA-MEM v0.1.17 (default parameters) and Samtools v1.9 (36). Then, unmapped reads (mRNA) were subjected to *de novo* assembly using SPADES v3.13.1 (k-mer length 21). From 31% to 42% of trimmed reads coding for rRNA per sample were fast assigned to bacterial families with Kraken (39) against the Silva SSU/LSU-Ref database (Release 132) (40). 27% to 31% of reads corresponded to not classifiable metagenomics sequences, related to microbial dark matter (41). 0.68% to 12% of reads were classifiable on the 16S rRNA Silva database. (Supplementary Figure 6, Supplementary Table 3).

After sequencing, trimming and *in silico* ribosomal RNA depletion, we obtained 4.137.706 putative mRNA reads (57.77 % of the initial sample) for the Hospital *in situ* biofilm, and 35.861.826 putative mRNA reads (29.6% of the initial sample) for the Urban *in situ* biofilm, leading to a reasonable number of contigs to analyse, with 866.221 and 2.379.394 contigs for the hospital and urban WW *in situ* biofilms, respectively (Supplementary Table 2). For the hospital WW *in vitro* biofilms HL-1 and HL-2 (HL-2 = hospital WW *in vitro* biofilm with ciprofloxacin added) 14.860.486 and 36.452.020 mRNA reads were obtained leading to

2.332.472 (HL-1) and 3.285.208 (HL-2) assembled contigs, respectively. For the urban WW *in vitro* biofilms

UL-1 and UL-2 (UL-2= urban WW *in vitro* biofilm + ciprofloxacin), 55.513.773 and 38.054.494 mRNA reads

were obtained leading to 3.483.745 and 2.684.771 assembled contigs, respectively (Supplementary Table

2 and 3).

Metatranscriptomic analysis:

Mapping the mRNA assemblies to specific reference databases enables to assess the expression levels of

functionally viable ARGs within biofilms.

A specialized database of the genes targeted by the high-throughput qPCR for the targeted resistome

analysis was built (accession numbers available in Supplementary Table 1) and used for mapping

assembled contigs with BWA-MEM v0.1.17 (with default parameters) and SAMtools v1.9. Then, the

transcriptome activity and abundance of ARGs was assessed by mapping reads on assembled contigs (42).

Reads were mapped against each ARG-like sequence identified from meta-transcriptomic assembly

contigs, using Bowtie2 v2.3.4.3 with default settings. The ARG-like sequences with > 100 mapped reads

were assigned to ARG transcripts.

To increase the chance of detecting active ARGs in hospital and urban wastewater biofilms, transcripts

coding for ARGs were searched against the CARD database (release 08/21/2019) by using the Resistance

Gene Identifier (RGI 3.1.0) retaining perfect and strict hits only (43). The ARG-like sequences were

identified as ARG transcripts when their coverage was up to 80% with a minimum of 100 reads mapping.

The expression level of a particular ARG transcript in each sample was calculated by summing the number

of hits by ARG class (see list of categories in Supplementary table 3). In addition, we weighted the

identification with the mean of read coverage by ARG class.

In addition, we detected ARG-carrying MGEs in WW biofilms, namely the active resistance mobilome.

First, assembled contigs were mapped against the ACLAME database (44), using BWA and Samtools.

Subsequently, identified MGE-carrying contigs were again mapped against the CARD database to identify

ARG-carrying MGEs. If one MGE-carrying contig mapped to different ARGs, only hits with the best CIGAR

(extracted from the SAM file) were considered.

All steps described above were implanted in python-based modules, and all modules formed the

framework of the pipeline AROM, for ARg On Mobilome (Supplementary Figure 2). Lately, we added the

module get abundance to normalize gene expression levels according to the sample size (Gb), the length

of reads and the length of each ARG (45). All changes to AROM

(https://github.com/sophiaachaibou/ARGmodules) are tracked in GitHub and the versions managed using

bioconda.

Extracellular polymeric substances (EPS) staining for confocal scanning laser microscopy:

The following stains were used to fluorescently label EPS components (polysaccharides, lipids, and

external DNA) and to assess live-dead ratios of different biofilm samples. Life-dead: Propidium Iodide (red;

0.5μM working solution) and SYTO-BC (green; Life technologies™; 5uM working solution). 500μl of each

fluorescent dye applied and allowed to incubate at ambient temperature in the dark for 30 min. Slides

were subsequently rinsed with three times with PBS and immediately analyzed by CSLM. For the EPS

components, biofilms were fixed in 3.7% formaldehyde solution (see above) prior to EPS staining. For

external DNA (eDNA) the combination of TOTO-1 (green) and SYTO-60 (red) (Life technologies™) was used.

Staining was performed as described by Okshevsky et *al.* (46), by using a final working stock solution (in

phosphate buffered saline (PBS)) of TOTO-1 at 2μM and SYTO-60 at 10μM. Here, following three PBS

washes, slides kept at 4°C in the dark until CSLM analysis the same day. For polysaccharides (carbohydrate

binding proteins) and proteins a combination of Concanavalin A, tetramethylrhodamine conjugate (ConA;

Thermo Fischer Scientific) selectively binding to α-mannopyranosyl and α-glucopyranosyl residues

(carbohydrates) and Fluorescein isothiocyanate isomer I (FITC; Siegma Aldrich) selectively binding to

proteins was used. FITC working solution was concentrated at 200μg/ml. 500μl of 0.1M sodium

bicarbonate buffer was added to the biofilm slide prior to adding 500μl of FITC staining solution (2μg/ml)

and incubated at RT in the dark for 30 minutes. Afterwards the excess stain was removed by washing the

biofilm slide gently with 1 x PBS solution for three times. Subsequently 500μl of ConA staining solution

(50μg/ml) was added to the slide and incubated for 30 minutes at RT in the dark. Excess stain was removed

by washing three times with 1x PBS solution. Slides were stored at 4°C until CSLM analysis. For lipid and

internal lipid droplet staining Nile Red (0.1μg/ml) was used in combination with FITC (2μg/ml). Incubation

and staining were performed as described above by first applying 500μl of 0.1M sodium bicarbonate

buffer, subsequently 500μl of FITC working solution, incubation for 30 minutes at RT in the dark, washing

with PBS for removal of excess FITC staining and addition of 500μl of Nile Red working solution, incubation

for 30 minutes at RT in the dark, removal of excess dye by washing with PBS 3 times, and storing at 4°C

until analysis by CSLM.

Confocal scanning laser microscopy settings and analysis:

Z stacking acquisitions were acquired with a confocal microscope (ZEISS LSM880) using a x20 and x100 magnification. Five Z stacks were performed per condition. The following settings were used to measure the fluorescence from the samples: autofluorescence (λexc 405nm/λem 460nm), FITC (λexc 488nm/λem 525nm), tetramethylrhodamine (λexc 561nm/λem 580nm), Nile Red (λexc 641nm/λem 650nm), Syto60 (λexc 641nm/λem 678nm) and TOTO-1 (λexc 514nm/λem 533nm).

3D view analysis was performed with the Imaris software (Oxford Instrument) using acquisitions with the x20 magnification. Surfaces were created using a manual thresholding. The quantitative analysis was performed with the open-source software BiofilmQ (https://drescherlab.org/data/biofilmQ/) on the acquisitions made with the x100 magnification. Each fluorescent z-stack was segmented using an automatic thresholding (otsu algorithm) allowing object detection and measurements. We focused on the relative abundance parameter that is the measurement of the proportion (i.e a percentage) of one component (i.e one staining) regarding the biovolume. This biovolume must be understood as the total volume occupied by one or the other staining of interest (this is a boolean operation "OR" between these stainings). The relative abundance of the polysaccharides (corresponding to the concanavalin A staining) and the lipids (i.e nile red staining) were expressed as a percentage of the biovolume resulting from the merge between concanavalin A OR fluorescein, or between nile red OR fluorescein respectively. eDNA relative abundance was expressed as a percentage of the TOTO-1 staining (i.e eDNA) regarding the biovolume resulting from the merge between TOTO-1 OR SYTO-60 stainings.

Statistical analysis

Biofilm matrices

To test for the effect of sample origin (hospital vs urban WW), environment (*in situ* vs *in vitro*) and addition of ciprofloxacin on the composition of the biofilm matrices, the relative abundance of three features were measured, the eDNA, lipid and polysaccharides. For each of these features two models were built. The first set of models focused on data obtained without ciprofloxacin and explained the three measured features with the variables 'sample origin' and 'environment', as well as their interactions. The second set of models focused on data obtained in the in vitro environment and explained the three measured features with the variables 'sample origin' and 'ciprofloxacin', as well as their interactions. The last set of models is devised to perform a pairwise comparison of each sample, hence, a different model was built for each pair of samples, with one explanatory variable and two levels, corresponding to the sample ID.

All these tests were based of the Randomization of Residuals in a Permutation Procedure implemented

by the R package RRPP.

Pairwise comparison of the normalized abundance of the resistome data using the Tukey HSD test

To compare the relative abundances of genes classes between samples, for each class and biological

replicate, we summed the normalized abundances of individual genes per class. Then for each gene class,

we fitted the relative abundance with a log linear model with one explanatory variable, and one level for

each sample. Finally, we applied pairwise comparisons of samples using the Tukey HSD test implemented

by the multcomp R package and p-values were adjusted using the single step method (47).

Correlation of the normalized abundance for ARGs detected by qPCR and their likelihood to be
detected by our meta-transcriptomic pipeline

To detect the threshold for normalized abundance values of genes measured by qPCR, below which meta-

transcriptomic has a low chance of detecting the gene, we fitted a binomial generalized linear model

(GLM) to the mapped reads obtained for individual genes detected by our meta-transcriptomic pipeline,

and as explanatory variable, the normalized abundance of the respective genes measured by qPCR. This

allowed to obtain a cut-off value for normalized abundance determined by qPCR that is predicted to have

a chance of being detected by our meta-transcriptomic pipeline. We also fitted a Poisson GLMM to the

exact read maps for genes detected by our meta-transcriptomic pipeline, with the same explanatory

variable as in the binomial GLM, and in addition, an observation level random effect, to account for

overdispersion (48).

Pairwise comparison of biofilm communities, mobilome and resistome

For the six types of multivariate description of the biofilms, communities (family and genus levels),

mobilome (plasmid and gene levels) and resistome (targeted and non-targeted), we compared each pair

of samples by measuring the proportion of the Simpson's diversity that was between samples. Simpson's

diversity measures the probability that two entities taken at random from the dataset are of the same

type. It is calculated as $1 - \sum f_i^2$, where values of f are the frequencies of the entities. To compute the

proportion of this Simpson's diversity that is between a given pair samples A and B, we computed the

total diversity using the average of the frequencies of the two samples and the average diversity within

the two sample:

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$$D_{AB} = 1 - \frac{\left(1 - \sum f_{iA}^2\right) + \left(1 - \sum f_{iB}^2\right)/2}{1 - \sum \left((f_{iA} + f_{iB})/2\right)^2}$$

This divergence index turns out to also be the formula of the between population fixation index (F_{ST}) used

in population genetics, using alleles frequencies (49). To assess the significance of these divergences, we

permuted the read counts 10^5 times while keeping fixed the headcounts for each entity and sample. For

each of these permutations, we recomputed the frequencies and the divergence index, to obtain its null

distribution under the assumption that the distributions are identical in the two samples.

Based on these pairwise dissimilarities, we built UPGMA trees (unweighted pair group method with

arithmetic mean). The stability of the trees was assessed by bootstrapping the individual genes.

To test the significance of the divergence between samples we used the pair of samples with the lowest

divergence as a conservative reference. That is, we computed for each pair of samples, the proportion of

bootstrap with the lowest D_AB divergence. Finally, we assessed and tested the correlation between the

dissimilarity matrices used to build the UPGMA trees using some Mantel tests (R package ade4).

Correspondence and in between correspondence analysis for ARG detected on plasmids (mobilome)

To analyze the association between the class of the ARGs of the mobilome and the WW origin of the

biofilm and their treatment (*in situ*, *in vitro*, with or without ciprofloxacin), we first performed a

correspondence analysis (CA; function 'dudi.coa' of the ade4 R package) analyzing the link between the

relative abundance of each ARG associated to a MGE. The significance of this association was tested

through a χ^2-test with a p-value based on 10^5 Monte Carlo simulations, and its effect size through the

Cramér's V statistic, a measure of the association between two categorical variables that varies from 0

(no association) to 1 (complete association). Then ARGs were grouped into gene classes, and the

association between gene classes and biofilm was investigated by applying between class analysis

(function 'bca' of the ade4 R package) to the CA. We refer to this overall analysis as to a between

correspondence analysis (bCA) which significance was assessed though a 10^5 permutation test. This bCA

recovered 33% of the total association found in the CA.

Results

Effects of minimal selective concentrations (MSC) of ciprofloxacin, sample origin and environment, on hospital and urban WW biofilms' extracellular polymeric substances (EPS)

We first analyzed the global structure of the hospital and urban WW biofilms. The EPS (eDNA, lipids and polysaccharides) were visualized by confocal scanning laser microscopy (CSLM). 3-D imaging (IMARIS) and quantitative analysis (Biofilm Q) showed that the EPS significantly differ between their growth environment, namely whether biofilms were grown *in situ* or, as reconstituted biofilms, *in vitro*. We detected different relative abundance of eDNA (in both hospital and urban biofilms compared to their *in vitro* counterparts), polysaccharides (for urban *in situ* vs *in vitro* + cip biofilms) and lipids (for hospital *in situ* vs *in vitro* biofilms) (Figure 1 and 2, Supplementary Figure 4a and 4b, Supplementary Table 5). We also detected significant differences in the relative abundance of EPS components dependent on the sample origin, meaning between hospital and urban *in situ* WW biofilms, for eDNA, polysaccharides and lipids (Figures 1 and 2, Supplementary Figure 4a and 4b). The relative abundance of eDNA was higher for the hospital *in vitro* biofilms compared to the urban *in vitro* WW biofilms. We also assessed the impact of adding ciprofloxacin on the life-dead ratio and EPS components of the *in vitro* reconstituted hospital and urban WW biofilms. A significant increase of dead cells was detected for the hospital *in vitro* biofilm after adding ciprofloxacin, compared to its control (Supplementary Figure 3). The relative abundance of eDNA was significantly lower in the hospital *in vitro* biofilm exposed to ciprofloxacin and significantly lower for polysaccharides in the urban *in vitro* biofilm exposed to ciprofloxacin compared to their control.

These results were random permutation-based pair-wise comparisons of the average relative abundance data extracted for 5 z-stack acquisitions performed per biofilm sample. Finally, we implemented statistical models based on random permutations testing for an effect of co-founding variables, such as the environment (*in situ* vs *in vitro*), sample origin (hospital vs urban) and/or the effect of ciprofloxacin, on the EPS components of these biofilms (Supplementary Table 5). We confirmed that i) the sample origin (type of hospital vs urban WW) has a significant impact on the abundance of polysaccharides and lipids, ii) whether biofilms are grown *in situ* or *in vitro* has a significant impact on the relative abundance of eDNA

351   and iii) ciprofloxacin has a significant impact on the abundance of polysaccharides for both, hospital, and

352   urban *in vitro* biofilms (Supplementary Table 5).

Figure 1: Relative abundance of eDNA (A), polysaccharides (B) and lipids (C) in the respective biofilms. Each EPS
component was stained in a different biological replicate, and fluorescence was detected by 5 z-stack acquisitions
per biological replicate with a 100-fold magnification for the respective biofilms. Quantitative image analysis was
performed with Biofilm Q. Results of pairwise permutations tests are summarized using the compact letter display.
Samples that do not share a letter in common are significantly different from each other. +Cip= *in vitro* biofilms were
exposed after 3 days of initial biofilm formation with 1 µg/ml ciprofloxacin for 4 consecutive days and harvested
after exposure of 4 hours on the final day.

Staining of eDNA   *In vitro*   *In vitro* + Cip.   *In situ*

Figure 2: **3D imaging of eDNA (green) and total nucleic acids (red), for hospital and urban *in-situ* and *in-vitro* biofilms.** Autofluorescence for unknown compounds is in blue. +Cip= *in vitro* biofilms were exposed after 3 days of initial biofilm formation with 1 µg/ml ciprofloxacin for 4 consecutive days and harvested after exposure of 4 hours on the final day.

The microbiota of hospital and urban WW *in situ* and *in vitro* biofilms

Figure 3: **Microbiota composition of hospital and urban wastewater *in situ* and *in vitro* biofilms based on**
sequencing of the 16S rRNA gene. Depicted are the top 10 taxa on order level and the top 12 on the genus level for
each sample, the remaining taxa are grouped into others. +Cip= *in vitro* biofilms were exposed after 3 days of initial
biofilm formation with 1µg/ml ciprofloxacin for 4 consecutive days and harvested after exposure of 4 hours on the
final day (D7).

The microbial compositions of WW biofilms were assessed using 16S rDNA sequencing. Strikingly and in

line with the phenotypic analysis of the biofilm structures, bacterial community profiles of hospital and

urban *in situ* biofilms were significantly different from their *in vitro* reconstituted counterparts

(Supplementary Figure 5 and Supplementary Table 6). The main genera in the *in situ* hospital WW biofilms

are Gram-negative pathogens such as *Pseudomonas*, *Acinetobacter* and *Aeromonas,* responsible for

infections in hospitalized patients (50–52) (Figure 3). *In situ* urban WW biofilms are dominated by different

genera than hospital WW biofilms, namely *Flavobacterium* and *Arcobacter* but share *Acinetobacter* with

hospital *in situ* biofilms. The genus *Arcobacter* is a common member of human WW systems and can rarely

act as animal and human pathogen (53,54). It is more abundant in urban WW biofilms, compared to the

hospital WW environment. *Pseudomonas* is a dominant genus in both urban and hospital *in vitro* biofilms*.*

Surprisingly, *Acinetobacter* decreased in the *in vitro* biofilms compared to their *in situ* counterparts. The

addition of ciprofloxacin did not change the proportional microbiota composition for *in vitro* biofilms

compared to their controls (Figure 3), suggesting that exposure of already formed WW biofilms with

ciprofloxacin in our *in vitro* model does not alter deeply the primary bacterial ecosystems. However, when

assessing Simpson's diversity index and performing pairwise comparison and mantel tests on our 16S

rDNA sequencing-based data on the family and genus level, there is evidence that a small proportion of

the total microbiota is impacted in its proportional community structure due to the addition of

ciprofloxacin in hospital *in vitro* biofilms (Supplementary Figure 5 and Supplementary Table 6).

As described in methods, a relatively large proportion of the RNA reads from the metatranscriptomic RNA

sequencing approach of the biofilm communities were ribosomal RNA (rRNA) reads. The 16S rRNA reads

were filtered out by mapping on the Silva database, allowing for the assignment of a small proportion of

those reads (between 0.68% and 12%) on family level. Hence, this step allowed for additional analysis of

the "active microbiota" of these biofilms. Among the classifiable 16S rRNA reads from the

metatranscriptomic analysis (methods), we identified a large majority of *Enterobacteriaceae* in hospital

biofilms and a mix of *Flavobacteriaceae*, *Enterobacteriaceae*, *Streptococcaceae* and *Francisellaceae* in

urban *in situ* biofilm (Supplementary Figure 6 and Supplementary Table 3).

The targeted resistome and metatranscriptome in hospital and urban WW biofilms

Figure 4: The targeted resistome and its expression in hospital and urban WW biofilms. On panel **A** and **B**,
normalized accumulated abundance of the targeted resistome (A) and their relative expression (B) were grouped
into gene families and mobile genetic elements (MGEs) and integron integrase genes. **C**: Heatmap depicting
normalized abundance of individual genes (targeted resistome) and their normalized expression levels (exact counts
of mapped reads for the respective genes) for the individual genes detected by RNA sequencing (meta-
transcriptome). Data were log2 transformed for visualization purpose. +Ab= *in vitro* biofilms were exposed after 3
days of initial biofilm formation with 1µg/ml of the antibiotic ciprofloxacin for 4 consecutive days and harvested
after exposure of 4 hours on the final day (D7). Cl= chloramphenicol; M= macrolides; E= efflux pumps/multi-drug
resistance genes; Qi= quinolones; HM= heavy metals; Q= quaternary ammonium compounds (QACs); Tet=
tetracyclines; MGEs= Mobile Genetic Elements. Int-I-Gs= integron integrase genes class I, II and III. +Cip= *in vitro*
biofilms were exposed after 3 days of initial biofilm formation with 1 µg/ml ciprofloxacin for 4 consecutive days and
harvested after exposure of 4 hours on the final day.

To assess the resistome in our samples, a targeted high-throughput qPCR approach, that detects and

quantifies 87 individual genes conferring resistance to clinically relevant antibiotics, quaternary

ammonium compounds (QACs), heavy metals, transposase genes (MGEs) and integron integrase genes,

also referred to as targeted resistome, was applied. At first, comparing the different environments in

which biofilms grew (*in situ* vs *in vitro*), we found that for the hospital WW biofilms, all gene classes had

a higher normalized abundance in *in situ* biofilm compared to the *in vitro* biofilms, except for

aminoglycosides, quinolones, QACs, sulphonamides, MGEs and integrons, which were not significantly

different (Figure 4 a and c, Supplementary Figure 7, and Supplementary Table 7). Performing the same

comparison for the urban WW *in situ* vs the *in* vitro biofilms, the normalized abundances of macrolides,

quinolones and MGEs, were significantly higher for the *in situ* biofilms, whereas genes conferring

resistance to heavy metals, vancomycin, and sulphonamides had significantly lower normalized

abundances in *in situ* urban WW biofilms. Then, comparing the different sampling origin (hospital vs urban

WW), the normalized abundance of the targeted resistome was highest in the *in situ* hospital WW biofilms

for all detected gene classes apart from macrolides, which were higher in the urban *in situ* biofilm, while

tetracyclines, quinolones and MGEs, were not significantly different in their normalized abundance

between the two *in situ* biofilm sampling origins (Supplementary Figure 7, and Supplementary Table 7).

Comparing the normalized abundance of detected gene classes for the hospital *in vitro* biofilms vs the

urban *in vitro* biofilms, it was higher in the hospital *in vitro* biofilms for most genes, except for vancomycin

and genes encoding for multi-drug efflux pumps for which the normalized abundance was higher in the

urban *in vitro* biofilms; no difference was observed for tetracyclines and heavy metal resistance genes.

Then, we analysed the metagenomic messenger RNA (mRNA) to provide information about the targeted

resistome actively being expressed in the *in situ* and *in vitro* hospital and urban WW biofilms. Overall, 39

of the 87 targeted genes by qPCR were expressed in at least one of the samples (Figure 4b and c).

Following the same comparison pattern, when first comparing the biofilm environments (*in situ* vs *in*

vitro), beta-lactam, chloramphenicol and sulphonamide resistance genes were 112, 5, and 47-fold higher

expressed in the hospital *in situ* biofilm vs *in vitro* biofilm, whereas macrolide resistance genes were 49-

fold higher expressed in the urban *in situ* biofilm compared to *in vitro* biofilm, while for all other gene

classes were gene expression could be detected, these detection levels were lower in the urban *in situ*

biofilm vs *in vitro* (Figure 4b and c, Supplementary Table 8 and Supplementary 8). For the different sample

origins (hospital vs urban WW) of the studied biofilms, expression levels were higher for 7 gene classes,

from 5.5-fold higher for genes conferring resistance to aminoglycosides to 957-fold higher for genes

conferring resistance to beta-lactams, (Supplementary Table 8 and Supplementary 8) for the *in situ*
hospital vs urban *in situ* WW biofilms.

We studied the impact of ciprofloxacin in the *in vitro* biofilm models. When looking at normalized
abundance in hospital and urban *in vitro* biofilms exposed to ciprofloxacin, no differences could be
detected, except for genes conferring resistance to macrolides in the hospital *in vitro* biofilm exposed to
ciprofloxacin compared to its control. (Supplementary Figure 7, and Supplementary Table 7). Interestingly,
we observed an enhanced expression of the targeted resistome (Figure 4b and c, Supplementary Table 8
and Supplementary Figure 8) after ciprofloxacin addition in the hospital biofilm. For example, expression
of ARGs conferring resistance to beta-lactams was 26-fold higher and aminoglycoside resistance genes
were expressed 11-fold higher and represented the class of ARGs that was most abundantly expressed
proportionally within the hospital biofilm (Figure 4b and c, Supplementary Table 8 and Supplementary
Figure 8). Furthermore, ciprofloxacin enhanced expression (4-fold) of the quinolone-resistance genes,
qnrB and *qnrS* genes, and integron integrase genes (*IntI1* and *IntI3*) by 6-fold and the grouped transposase
genes by 1.7-fold compared to its control (Figure 4b and c, Supplementary Table 8 and Supplementary
Figure 8). When looking at the detected expression levels of ARG classes in the urban *in vitro* biofilm
exposed to ciprofloxacin, expression of ARGs was lower for all detected gene classes apart from the
integron integrase and aminoglycoside genes, were expression slightly increased by 1.1 and 1.7-fold
respectively, compared to its control (Figure 4b and c, Supplementary Table 8 and Supplementary Figure
8). The *ISS1N* gene was specifically expressed in the *in situ* urban WW biofilm.

Overall, these findings showed that 45% of the targeted resistome is actively expressed in the WW biofilms
and that the expression of ARGs and integron integrase genes in the *in vitro* hospital WW biofilm is
enhanced by the addition of ciprofloxacin. Interestingly, the expression level of ARGs could be correlated
(linear regression analysis to be R=0.77, Supplementary Figure 10 and Supplementary Table 10) to the
normalized abundance of genes determined by the targeted resistome approach until a numerical
threshold of 0.057 for the normalized abundance of the individual genes (Supplementary Figure 10 and
Supplementary Table 10). Below this threshold, the detection of ARG expression could no longer be
achieved, suggesting a lack of transcripts. However, technical limitations such as the low sensitivity of
meta-transcriptomics and/or the detection of DNA from dead cells and eDNA by qPCR might also be
possible explanations as to why not all the targeted genes could be detected by the metatranscriptomic
approach (55–57).

Assessing the expression of ARGs based on the CARD database in hospital and urban WW biofilms

To identify additional ARGs that were not included in our targeted resistome approach, mRNA reads of our samples were mapped against the Comprehensive Antibiotic Resistance Database (CARD) that archives thousands of gene sequences encoding resistance determinants to antimicrobial agents.

ARG expression was higher for *in situ* hospital WW biofilms compared to their *in vitro* counterpart (10.4% for hospital *in situ* biofilm vs about 0.5% for *in vitro* of total mRNA reads in the respective samples (Supplementary Table 1)). The detected expression levels of ARGs in the *in situ* urban WW biofilms was lower compared to their *in vitro* counterpart (0.04% of mRNA reads for the urban *in situ* vs urban *in vitro* 0.4% biofilm). A tenfold higher expression level of ARGs was detected in the *in situ* hospital WW biofilm (10.4 % of all mRNA reads coding for ARGs) compared to the urban WW *in situ* biofilm (0.04 % of mRNA reads coding for ARGs) (Supplementary Table1). For the *in vitro* hospital biofilm exposed to ciprofloxacin, expression level of ARGs was about 4.4-fold higher compared to the non- exposed one (2.4% vs 0.55% vs, (Supplementary Table 1)). No difference in ARG expression level based on total ARG mRNA reads could be observed between urban *in vitro* biofilms exposed or not to ciprofloxacin for the normalized data (0.41% vs 0.37% of total mRNA reads mapping to ARGs, respectively, Supplementary Table 1).

Figure 5: ARG metatranscriptome based on the CARD database. Means of mapped reads per antimicrobial resistance gene were summed by ARG class and are displayed per sample.

In total we detected the expression of 70 individual resistance genes conferring resistance to 16 classes

of antibacterial drugs across all biofilm samples (Figure 5, Supplementary Tables 4 and 9). This approach

led to the detection of the expression of additional genes in hospital and urban WW biofilms that were

not detected by the targeted approach (Supplementary Table 4). For example, active beta-lactamase such

as *bla*[RCP], *bla*[LCR], *bla*[FOX], *bla*[NPS], four aminoglycoside-modifying enzymes (*aac(3)-Ib/aac(6')-Ib*, *aad*B,

aph(3''), *aph*(6)) and a macrolide phosphotransferase (MPH) were identified. Genes that confer resistance

to the antimicrobial classes fusidic acid, elfamycin, rifamycin and antitumoral glycopeptides were found

to be actively expressed in all the investigated samples. Genes that confer resistance to isoniazid and

triclosan were found to be expressed in the *in vitro* hospital biofilm exposed to ciprofloxacin, and in *in*

vitro urban WW biofilms in both conditions. The expression of an enterobacterial fosfomycin resistance

gene (*ptsI*) was detected in the *in vitro* hospital WW biofilm exposed to ciprofloxacin (Supplementary

Table 4). Furthermore, the expression of multidrug efflux pumps, specifically RND and MFS types, and the

expression of the *bla*[OXA] beta-lactamases was detected in all the samples.

Overall the composition of the ARG metatranscriptome of the CARD resistome analysis correlated to the

targeted resistome analysis (Supplementary Figure 5, Supplementary Table 4 and 9). Comparing the

different environments for hospital biofilms (*in situ* vs *in vitro*), higher gene expression levels were

detected for the *in situ* hospital biofilm between 3.2-fold and 15.2-fold (macrolides and beta-lactamases

respectively). For the *in situ* vs *in vitro* urban biofilms, expression levels were lower for all detected ARG

classes, apart from macrolide ARGs, which were expressed at a higher level in urban *in situ* biofilm

(Supplementary Figure 9, Supplementary Table 4 and 9). For hospital *in situ* biofilm, six ARG classes had

between 1.8 and 81.2-fold higher expression levels (genes encoding resistance to fluoroquinolones and

aminoglycosides respectively) vs urban *in situ* biofilm. Comparing hospital *in vitro* biofilm exposed to

ciprofloxacin to its control, overall, increasing fold-change for seven ARG classes, between 3- and 7.5-fold

change for genes encoding for multi-drug efflux pumps and genes conferring resistance to beta-lactams,

respectively, was detected. Interestingly, 5 ARG classes, namely aminoglycosides, beta lactams,

quinolones, sulphonamides, and genes encoding for multi-drug efflux pumps, were detected to have

higher gene expression levels in both, the targeted and non-targeted resistome approach, for the hospital

in vitro biofilm exposed to ciprofloxacin compared to its control (Supplementary Figures 8 and 9,

Supplementary Table 4, 8 and 9). For the urban *in vitro* biofilm exposed to ciprofloxacin, expression levels

are lower, expect for genes encoding for beta-lactams were a slight increase (1.18-fold) could be detected

compared to its control (Supplementary Figure 9, Supplementary Table 4 and 9).

The active resistance mobilome

To study the expression of MGEs and more specifically MGEs associated with ARGs, assembled contigs were mapped against the plasmid sequence collection v0.4 available at the ACLAME database (44). Subsequently, plasmid associated contigs identified by this approach were mapped back to the CARD database to identify specifically plasmids that carry ARGs. These ARG associated plasmids are referred to as the "active resistance mobilome" in the respective samples.

Global expression of plasmids was very low in all the samples studied (0.48% of all assembled contigs for hospital *in situ* biofilm and 0.14% for urban *in situ* biofilm (Supplementary Table 1)). In total, we counted respectively 115 and 66 plasmid-ARG associated contigs for the hospital and urban *in situ* biofilms and identified a diverse range of expressed plasmids (71 individual plasmids identified across all samples) expressing a diverse range of ARGs that could be grouped into 10 classes conferring resistance to antimicrobials, quaternary ammonium compounds and heavy metals. The highest expression level for the active resistance mobilome was observed in the *in vitro* hospital biofilm exposed to ciprofloxacin (Supplementary Figure 11).

55 different plasmids carrying a broad range of ARGs were identified as the active resistance mobilome across all hospital WW biofilm samples. These plasmids were mainly resistance plasmids found in Gram-negative pathogens such as the *bla-$_{VIM7}$* carrying plasmid *pmATVIM-7* first identified in *P. aeruginosa* (50); the IncP-6 plasmid *Rms149,* a small resistant plasmid associated with *P. aeruginosa* (58) (detected only in hospital *in situ* biofilm); several small resistance plasmids (e.g. *pKMA757, pARD3079, pKMA202,* and *pKMA5)* associated with *Actinobacilli* (59) and plasmid families representing the groups IncP, F, N and Q (Supplementary Figure 11). They primarily expressed aminoglycoside, sulphonamide, chloramphenicol, and beta-lactam resistance genes (Supplementary Figure 11, Figure 6). 14 plasmids were detected in all hospital WW biofilm samples, and 7 additional plasmids were only shared between the *in situ* hospital WW biofilms and the *in vitro* hospital WW biofilm exposed to ciprofloxacin (Supplementary Figure 11). Interestingly, 22 plasmids, carrying between 1 and 5 ARGs, were only identified in the *in vitro* hospital WW biofilm exposed to ciprofloxacin (Supplementary Figure 11).

14 plasmids were unique for urban WW biofilm samples (*in situ* and *in vitro*). These plasmids mainly carried macrolide and tetracycline resistance genes found in Gram-positive bacteria, such as Lactobacilli (56). In addition, the *Enterococcus faecium* and *faecalis* plasmids *pRUM* and *pTEF1* were identified as expressing both the macrolide resistance gene *ermB* (60–62) (Supplementary Figure 11). The broad host

range IncP-1 plasmid pKJK5 was also detected, expressing an aminoglycoside (*aadA6*) and a sulphonamide

(*sul1*) resistance gene (63,64). One Gram-negative multidrug resistance plasmid *pJR1* (65) was detected,

carrying the chloramphenicol resistance gene *catB2* and the tetracycline resistance gene *tet(G)*

(Supplementary Figure 11). Fifteen plasmids were uniquely identified in the *in situ* urban WW biofilm and

not in its *in vitro* counterpart. Five of these plasmids were also detected in the hospital WW biofilms

expressing different ARGs (e.g. *pJHCMW1* and *R46*), whereas they only carried a single ARG when detected

in the *in situ* urban biofilm (Supplementary Figure 11).

In the hospital WW biofilms, many plasmids were identified to carry multiple resistance genes (up to 5

ARGs carried on one MGE contig, Supplementary Figure 11). These genes mainly encoded resistance to

aminoglycosides, beta-lactams, sulphonamides, chloramphenicol, and quaternary ammonium

compounds (Figure 6, Supplementary Figure 11). For the urban WW biofilms, MGEs that express multiple

resistance genes were also identified, but with less ARGs associated (maximum three ARGs on one plasmid

in the *in situ* urban WW biofilm) (Supplementary Figure 11). The ARGs mainly detected in the active

resistance mobilome of urban WW biofilms were macrolide, tetracyclines and multi-drug efflux pumps

(Figure 6, Supplementary Figure 11). Contrary to what is observed for the effect of ciprofloxacin on the

active resistance mobilome of the *in vitro* hospital biofilm, less MGE-ARG associated contigs were

identified, and a much lower diversity of ARGs were expressed in the *in vitro* urban biofilm exposed to

ciprofloxacin (mainly macrolide resistance genes) (Supplementary Figure 11, Figure 6).

Overall, multivariate analyses of our mobilome data (correspondence analysis (CA) and in between CA

(bCA)), revealed a prominent association between the type of ARGs classes present on plasmids and their

biofilm origin. (*p*-value=0.026; Figure 6).

Figure 6: Between correspondence analysis (bCA) assessing the association between the type of ARGs classes present on plasmids, in terms of ARG class abundance (numbers of contigs) and the biofilm origin. This association represents 33% of the total association between biofilm origin and the ARGs composition of the mobilome identified by the CA (Supplementary Figure 12).

Discussion

The goal of this study was to develop a comprehensive experimental approach to assess WW biofilms as putative hotspots for actual transfer of ARGs via MGEs within these communities using metagenomic approaches. We furthermore wanted to quantify and visualize the EPS components of these biofilms, due to the proposed role of the EPS in horizontal gene transfer within biofilms, as well as their relevance for biofilms to tolerate antibiotics and exogeneous stressors (29,31,66–68).

Here we introduced a method by confocal laser scanning microscopy (CLSM) to visualize and quantify the EPS of hospital and urban WW biofilms which will allow to study EPS components in these biofilms over time and under different anthropogenic constraints. We showed that i) the EPS components investigated are differentially abundant dependent on their growth environment (*in in* situ and *in vitro*) and their sampling origin (hospital vs urban WW) and ii) ciprofloxacin at MSC impacted the quantity of the measured EPS components. The decrease in the quantity of polysaccharides in the urban and in hospital *in vitro* biofilms, and for eDNA in the hospital *in vitro* biofilm, due to the presence of ciprofloxacin might be linked to a regulatory response of the biofilm community. In fact, it has been shown that exogeneous stress in multi-species biofilms leads to decreased EPS production (69). External DNA was more abundant in *in vitro* hospital WW biofilms compared to *in vitro* urban WW biofilms, which could be linked to higher amounts of surfactants, antibiotics, and other pharmaceutical and chemical pollution in hospital WW (7) that in turn may enhance bacterial lysis and active eDNA secretion in those biofilms (70). eDNA is known to promote tolerance to antimicrobial agents in biofilms by inhibiting the diffusion of cationic molecules (71) and by increasing the resistance to aminoglycosides in *Pseudomonas aeruginosa* biofilms (72,73). It has also been shown that eDNA is involved in transfer of the ARG *tetM* by the conjugative transposon Tn916 (68), through transformation. Transformation is suspected to be an important mechanism for gene transfer between bacterial communities in biofilms (68). We showed that the microbiota of hospital and urban WW biofilms is distinct, which is likely correlated to the different organization and relative abundances detected for the EPS in the two WW biofilm niches, as shown by our analysis (Supplementary Figure 5 and Supplementary Table 6). The significant increase of dead cells observed for the *in vitro* hospital WW biofilm exposed to ciprofloxacin could be linked to the predominant Gram-negative genera in hospital WW biofilms (Figure 3).

By implementing a meta-transcriptomic approach, at first, we wanted to assess the basic differential expression levels of mobile genetic elements (MGEs) associated with ARGs as a proxy for actual transfer of ARGs within hospital and urban WW biofilms. Second, we wanted to compare these levels to *in vitro*

reconstituted biofilms and third, we wanted to see whether these "basic" expression levels do change in

response to a single antibiotic added to the WW environment. Indeed, we showed that a) basic ARG and

MGEs expression levels are higher in *in situ* than *in vitro* WW hospital biofilms; b) expression levels of

ARGs and MGEs increase upon adding a minimal selective concentration (MSC) of ciprofloxacin to the *in*

vitro hospital WW biofilm environment and c) that expression of plasmids carrying specific ARG classes

and multiple ARGs at the same time, is associated with specific biofilm types (Figure 6). We were able to

detect a broad range of plasmids that were uniquely detected in either, hospital or urban WW biofilms,

hence correlated to their biofilm origin. Interestingly, plasmids that were detected in both, hospital and

urban *in situ* and *in vitro* biofilms, carried different types and quantities of ARGs (Supplementary Figure

9). This allowed us to identify a correlation between the nature of ARGs that are present on plasmids and

the biofilm environment and origin, and their likelihood to be actively expressed.

Interestingly, the reconstituted *in vitro* hospital and urban WW biofilms were more like each other

concerning their microbiota composition than their natural (*in situ*) biofilm counterparts. They differed

mainly in the relative abundance of their 3 main families (*Flavobacteriaceae*, *Comamonadaceae* and

Pseudomonadaceae) or genera (*Acinetobacter, Pseudomonas* and *Comamonas* for hospital *in vitro* WW

biofilms *and Acinetobacter, Pseudomonas* and *Flavobacterium* for urban *in vitro* biofilms, Figure 3). This

may suggest that the *in vitro* laboratory conditions promote similar biofilm formation for hospital and

urban WW biofilms, despite their originally quite different WW microbiota composition, which questions

our experimental set-up to implement microcosms in the laboratory that are capable to reproduce

ecosystem conditions identical to a natural environment. However, for example more frequent

refreshment of the WW alimentation in the *in vitro* system could significantly improve our experimental

model.

Among the classifiable rRNA reads from the meta-transcriptomic analysis (methods) we identified a large

majority of *Enterobacteriaceae* in the hospital biofilms, and a mix of *Flavobacteriaceae*,

Enterobacteriaceae, Streptococaceae and *Francisellaceae* in the urban *in situ* biofilm (Supplementary

Figure 6 and Supplementary Table 2). The fact that these rRNA reads were obtained from the RNA

sequencing run may indicate that these families were highly "active" in these biofilms although not highly

abundant based on 16S rDNA sequencing (Figure 3). The highly active proportion of *Enterobacteriaceae*

in hospital *in situ* and *in vitro* biofilms might indeed explain the higher expression levels and the types of

MGEs and ARGs detected in these samples, and their response to antibiotic stress. This further highlights

the need to study the 'active biofilm community' and might point towards the fact that the key players

for accumulation and transfer of ARGs and MGEs in highly anthropized sites such as WW are indeed
human-gut associated bacteria.

In a recent study, the Plascad tool was designed for plasmid classification, AMR gene annotation, and
plasmid visualization (74) and They showed that: i) most plasmid-borne ARGs, including those localized
on class-1 integrons, are enriched in conjugative plasmids in their datasets, and ii) transfer between AMR
plasmids and bacterial chromosomes was mediated by insertion sequences, mainly those belonging to
the IS6 family, IS26 and IS6100 (74). We have shown previously that IS26 and IS6100 were both abundant
in WW (7) and in this study we show that IS6100 is highly expressed in hospital WW biofilms. IS26 is
frequently detected in plasmids of clinical importance associated with antibiotic resistance genes.
Furthermore, IS26 has also been associated with the expression of antibiotic resistance genes (75). Finally,
the distinct plasmid profiles (Figure 6 and Supplementary Figure 11) and their specific ARG expression
profiles for hospital and urban WW biofilms point towards a strong environmental selection pressure
shaping the active resistance mobilome.

The strong correlation between high normalized abundance values of ARGs and the likelihood to be
detected by our meta-transcriptomic approach points towards limitations of RNA sequencing and
metatranscriptomics for the accurate quantification of ARGs and MGEs expression levels in complex multi-
species biofilm communities. The normalization of total ARG and MGEs reads allowed for the estimation
of those transcripts in the respective samples, and the comparison between the samples. However,
whether the "absence" of mRNA reads for ARGs, MGEs and specially ARGs detected to be on MGEs is due
to an insufficient sequencing depth cannot be excluded. Therefore, our analysis focused on the presence
of the ARG-MGE associated contigs, and their qualitative characteristics, which allowed for the detection
of significant correlations between the WW biofilm environment and sample origin, and to a slight degree,
the impact of antibiotic pressure on their presence and active transcription in the respective biofilms. Fast
progress is being made in the improvement of long-read third generation sequencing technologies, which
will allow for future validation of our findings, and the more accurate and in-depth study of hospital and
urban WW biofilms (76,77) and their differential expression of the "active resistance mobilome" in
response to exogeneous stress.

Here we provide an approach to comprehensively study multi-species biofilms that form in highly
anthropized environments such as WW. We show that hospital and urban wastewaters shape the
structure and active resistome of environmental biofilms, and we confirmed that hospital WW is an
important hot spot for the dissemination and selection of AMR. Hence, we propose that studying hospital

WW biofilms may present an ideal model to further unravel the complex correlation between pollution

and AMR selection and the structural, ecological, functional, and genetic organization of multi-species

biofilms as antibiotic resistance factories.

Data availability

16S rRNA sequence data are available at the European Nucleotide Archive (ENA) under the project
number PRJEB59054. RNA sequencing data (metatranscriptomic data) will be available at the NCBI
under submission number SUB12549421. All other important raw data needed to reconstruct the
findings of our study are made available in the supplementary material.

Funding

E. Buelow has received funding from the European Union's Horizon 2020 research and innovation
program under the Marie Skłodowska-Curie grant agreement RESOLVE 707999-standard EF.

Declaration of competing interest

The authors declare that they have no known competing financial interests or personal relationships
that could have appeared to influence the work reported in this paper.

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