The evolutionary stability of antagonistic plant facilitation across environmental gradients

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Abstract

Plant interactions, understood as the net effect of an individual on the fitness of a neighbor, vary in strength and can shift from negative interference to positive facilitation as the environmental conditions change in time and space. However, the biophysical mechanisms underlying these changes are not well understood. Additionally, evolutionary theory questions the stability of antagonistic facilitation. Using a mechanistic model for belowground resource competition between individual plants, we find that, under stress conditions, antagonistic facilitation is evolutionarily stable even when both interacting plants compete for resources. This supports the theory of ecosystem engineers in primary succession and nurse plants in the stress gradient hypothesis. Furthermore, we find that the proportion of the limiting resource that spontaneously becomes available to any plant is the key environmental parameter determining the evolutionary stability of facilitation. This represents a challenge and a potential confusion factor for empirical studies.

1. Introduction

Antagonistic facilitation is the interaction in which one individual increases the fitness of a second individual, which, in return, reduces the fitness of its benefactor, for example via competition for a common resource. Antagonistic facilitation is frequent in nature, from microbial (West *et al.* 2006; Foster 2010; Tarnita 2017) to plant communities (Brooker *et al.* 2008; Schöb *et al.* 2014; Soliveres *et al.* 2015), and it contributes to the increase and maintenance of local biodiversity and ecosystem services (Schöb *et al.* 2012; Bulleri *et al.* 2016, 2018; Cavieres *et al.* 2016; Wright *et al.* 2017; Losapio *et al.* 2021a).

Antagonistic facilitation is often associated with ecosystem engineers, i.e., individuals that modify their local environment creating habitats that are more conducive for life in the long term (Hastings *et al.* 2007; Gore *et al.* 2009; Cordero *et al.* 2012; Schöb *et al.* 2014; Estrela *et al.* 2018; Losapio *et al.* 2021b). Yet, its evolution and stability pose an evolutionary

60   dilemma, which made antagonistic facilitation subject of intense research over the last decades (Nowak 2006). To resolve this question, many theoretical studies have incorporated different mechanisms, such as a population spatial structure or individual reputations, into game-theoretical frameworks (Nowak 2006). More recently, experiments in microbes have both tested these model predictions and found alternative mechanisms that stabilize

65   antagonistic facilitation (Rainey & Rainey 2003; Drescher *et al.* 2014). At larger scales, however, antagonistic facilitation remains poorly understood.

In plant communities, ecosystem engineers are usually shrubs that colonize inhospitable habitats such as sand dunes, volcanic eruptions, deglaciation lands, or unstable stony grounds during the process of primary succession (Mori *et al.* 2017; Bai *et al.* 2018;

70   Karadimou *et al.* 2018; Kjær *et al.* 2018). To grow in such conditions, plant ecosystem engineers need to evolve specific traits that allow them to colonize hostile environments where most plants cannot survive (Verdú *et al.* 2021). Once the ecosystem engineer has established, opportunistic individuals that were unable to colonize the landscape by themselves can establish and proliferate (Crain & Bertness 2006). As a result, ecosystem

75   engineers and opportunistic plants interact via antagonistic facilitation. The former acts as a nurse plant, and the latter profits from the nursing effect of the ecosystem engineer but also competes with its benefactor for shared resources. A few theoretical studies have investigated how this association is sustained in the long term, suggesting that antagonistic facilitation is

evolutionarily unstable and thus dissolves through the course of evolution (Brooker *et al.*

80   2008; Bronstein 2009) or that it stabilizes only if facilitator plants cluster and engage locally

in mutualistic interactions (Kéfi *et al.* 2008).

The mechanisms by which a nurse plant can facilitate the success of neighbors and

promote primary succession are diverse, and soil amelioration is a major one (Pugnaire *et al.*

2004; Lozano *et al.* 2017; Koffel *et al.* 2018). For example, the root exudation of organic

85   acids, chemical signals, and enzymes can have several positive effects on soil quality for

roots. It can dissolve unavailable soil nutrients such as phosphorus (P), iron (Fe), or

potassium (K), avoid toxic effects of hypoxia or an excess of aluminum (Al), and allow the

establishment of symbiotic relations with nitrogen (N)-fixing microorganisms (Jones 1998;

Dakora & Phillips 2002; Hinsinger *et al.* 2011). Because N and P are essential resources for

90   plants and are typically sparse in soils, plants that can boost the availability of these

macronutrients have received considerable attention (Lambers *et al.* 2008). Plants adapted to

grow in P-deficient soils can develop expensive structures, known as cluster roots, that allow

them to release P that has been sorbed to soil particles and make it available for locally

foraging roots (Raven *et al.* 2018; Britto Costa *et al.* 2021). Some plants growing in N-limited

95   soils can also develop root nodules, establishing mutualistic relations with N-fixing

microorganisms that trade N for photosynthates (Sprent 1989; van Velzen *et al.* 2019).

Moreover, some plants can increase soil moisture by having root systems with specific

characteristics. For example, plants with roots able to break superficial soil compaction can

increase local water infiltration to the soil (Montaña 1992; Bromley *et al.* 1997), and plants

100   with deep tap roots reaching the water table can increase water availability in superficial soil

layers by hydraulic lift (Caldwell & Richards 1989; Zapater *et al.* 2011; Prieto *et al.* 2012).

In all these situations, plant ecosystem engineers have specific root traits that allow them to locally ameliorate soil conditions at a cost to themselves, creating islands of fertility where opportunistic plants can potentially grow.

105   We investigate the shifts in plant interactions across environmental gradients and the evolutionary stability of antagonistic facilitation using soil-ameliorating ecosystem engineers as case study. We hypothesize that, contrary to general belief, antagonistic facilitation might be evolutionarily stable if facilitator plants find net rewards in bearing the soil ameliorating trait despite the collateral attraction of opportunistic plants. This situation is possible if the

110 self-benefits of the trait overweight the costs of competition with the opportunistic neighbors. We test this hypothesis theoretically, using a modeling framework for belowground resource competition that describes resource dynamics explicitly. Our model results can encapsulate processes happening within the life span of an individual, the evolution of species, and the geological time (**Fig. 1**). Accounting for the phenotypically plastic root distribution of plants,

115 the evolution of the ecosystem engineer trait, and the long-term improvement of the soil over the course of succession, we investigate under which conditions antagonistic facilitation may be beneficial for both interacting plants and hence evolutionarily stable.

2. Materials and Methods

120 **2.1 Model formulation**

      We extend the spatially explicit model for resource competition introduced in Cabal *et al.*, 2020 to account for facilitation (Cabal *et al.* 2021). The model assumes a series of processes governing the dynamics of a soil resource W and provides the net resource gain for

<div align="center">5</div>

each plant due to the balance between how much resource it uptakes at a soil location and
the cost of exploiting it. To investigate the evolution of plant antagonistic facilitation, we
consider two types of plants. Plant ecosystem engineers can enhance soil resource availability
by evolving a resource mining trait ϕ. This trait bears a cost that is added to the cost of
building and maintaining the root system. Opportunistic plants cannot develop the resource
mining trait.

Figure 1: The different temporal scales at which we assessed the emergence of facilitative
interactions, including the plastic response of plant root distribution R at the life span time scale, the
adaptive response of the ecosystem engineers' facilitator trait ϕ at the evolutionary time scale, and
the change in the fraction of soil resource that is spontaneously available over succession b at the
geological time scale.

Resource dynamics. We model resource dynamics in each soil point as a combination
of three processes: input at rate I, abiotic decay at rate δ, and resource uptake by the fine roots
R at a per capita rate α (see **Table 1** for a summary of the environmental parameters and their
dimensions). Lastly, we model the resource input assuming positive feedback between

resource availability and mining intensity. In soils lacking engineer plants roots, the fraction

of available resource is equal to b ($0 \le b \le 1$) and it increases with engineer root density R_E

due to the mining trait. We model this positive feedback assuming that I saturates at high

mining intensity ϕR_E,

$$I(\phi, R_E) = \omega \frac{b + \phi R_E}{1 + \phi R_E} \tag{1}$$

145 Putting all these terms together, the resource concentration changes at each soil location

according to

$$\frac{\partial W(\ell, t)}{\partial t} = I(\phi, R_E) - \alpha R W - \delta W \tag{2}$$

where we have dropped the dependence on space and time from the terms on the right of Eq.

(2) and ℓ is the spatial coordinate of the soil location measured from the the insertion point

of the focal plant's stem to the soil surface.

150 *Net resource gain.* At each soil location the net resource gain of an individual plant,

j, at time t is the balance between the resource uptake and the associated cost,

$$G_j(\ell, t) = \left[WUE \, \alpha \, W(\ell, t) - C_j(\ell, \phi) \right] R_j(\ell) \tag{3}$$

where R is the fine root density (units of fine root biomass per unit of soil volume), C is the

cost of producing and maintaining such root density, and *WUE* is the resource use efficiency

that represents the conversion factor from harvested resource to new root biomass. Because

155 the cost function depends on the distance between the soil patch and the plant insertion point

to the soil surface, the model is spatially explicit. We account for three contributions to the

cost function: the cost of producing and maintaining fine roots that absorb resources from the

soil c_r, the cost of building and maintaining transportation roots that connect the fine roots to

the plant stem c_t (Cabal *et al.* 2020), and the cost per unit density of fine roots associated to
a unit increase in the resource-mining trait c_e,

$$C_j(\ell, \phi) = c_r + c_t\,\ell^2 + c_e\,\phi \qquad (4)$$

For opportunistic plants, $\phi = 0$ and the cost function in Eq. (4) has only two contributions (see **Table 2** for a summary of the plant parameters).

2.2 Model characteristic timescales.

We parameterize and analyze the model at three nested timescales. From shorter to longer, these timescales are: (i) the lifespan of a plant individual in which the spatial distribution of fine root biomass changes, (ii) the evolutionary timescale in which species evolve, including the resource-mining trait, and adapt to the environment, and (iii) the geological scale at which the environment matures driving successional changes in the community and we expect environmental parameters to vary (**Fig. 1**). Changes occurring in each of these timescales involve changes at shorter scales according to the new conditions found in every scenario. For example, as the resource-mining trait changes in evolutionary time to adapt to fixed environmental conditions, the spatial distribution of roots will also change according to the new level of resources mined by the plant.

(i) Changes in the spatial distribution of fine root biomass at the individual lifespan timescale. The spatial distribution of fine root biomass, $R(\ell)$, is a plastic trait tuned by a single plant over the course of its life span. Because the resource dynamics is much faster than root growth, we can assume that the resource concentration is in quasi-equilibrium in each step of plant growth and write a closed expression for the net resource gain of each plant at a given soil location (see SM for a full derivation),

$$G_j(\ell) = \left[\frac{WUE\ \alpha_j\ I(\phi, R_E)}{\delta + \sum_{i=1}^{n} \alpha_i\ R_i(\ell)} - C(\ell, \phi)\right] R_j(\ell) \tag{5}$$

Where R_E is the total root biomass of engineer plants at location ℓ. To optimize G in a given environmental scenario, we look for combinations of root density profiles and resource mining intensity that optimize the net resource gain in Eq. (5).

(ii) Changes in the resource-mining trait at the evolutionary time scale. The resource mining trait is normally assumed to remain constant over the lifespan of an individual, even if in some cases plants can tune it (Barron *et al.* 2011). Therefore, we consider that the resource-mining trait ϕ is fixed by evolution and its value does not vary spatially or in time for a given plant. Resource mining may evolve in ecosystem engineer plants when they are alone (ϕ_I) or, following a primary colonization by opportunistic plants, when they interact with opportunistic plants (ϕ_{II}) (see Supplementary Material for a detailed explanation on how we calculate these values). Modeling a community considering several individual opportunistic and engineer plants in interaction with one another would make the model mathematically intractable and require many arbitrary choices about community size and spatial structure. Instead, to get a general competitive background for the evolution of ϕ_{II}, we modeled the opportunistic individual as a spreading plant that grows vegetatively covering the soil surface (de Kroons & Hutchings 1995). These spreading opportunistic plants distribute their stem everywhere in the soil surface, like a stolon or a rhizome, so their roots grow across the canopy area of the ecosystem engineer at no horizontal space-related costs. Mathematically, this growth strategy implies removing the costs associated with transportation roots, $c_t = 0$. We consider this assumption to be a more realistic and tractable

200   representation of the average background community to which ecosystem engineers may

        have adapted.

                (iii) Environmental changes in soil quality at the geological time scale. Three

        parameters determine soil resource dynamics: ω, b, and δ. Among them, the proportion of

        existing resource that becomes available spontaneously without plant mining b, may be key

205   to understanding the mechanisms that mediate facilitation. Indeed, an engineer plant cannot

        increase the total amount of physical resource available in a soil location, but only the amount

        of such resource available to roots. It is also reasonable to assume that b is the main

        environmental parameter increasing over primary succession. For example, if water is the

        limiting resource and we consider two locations at different successional stages where

210   precipitation is identical, the amount of rainwater that becomes available to plants will

        depend on water absorption rates and holding capacity of soils. While most water will quickly

        runoff in immature soils, more water will become available to plants as the substrate particle

        size decreases and the organic matter content increases in more developed soils. This increase

        in water availability is represented in our model by a larger b. In another example, rocks

215   contain a lot of mineral nutrients that are not available to plant roots except for a small

        fraction that may dissolve in the rock interface. In more developed soils, however, most of

        these mineral nutrients incorporated into soil necromass and available to plants. These two

        examples illustrate how the environmental parameter b embodies the changes happening in

        soils over primary succession and how more mature soils offer a larger fraction of resources

220   to roots relative to the resources physically present. Therefore, we mimicked a geological

        time scale varying the parameter b for fixed values of the other environmental parameters ω

        and δ.

Table 1: Parameter description and values (if not indicated otherwise) for the environmental resource dynamics.

Parameter (Environment)	Symbol	Units	Value
Potential resource input	ω	ml.day^{-1}	5
Proportion of ω abiotically available	b	-	0.05
Resource abiotic decay rate	δ	day^{-1}	0.1

Table 2: Parameter description and values for the plant functional traits in spreading opportunistic plants (SPR), normal opportunistic plants (NOP), and ecosystem engineer plants (EEP).

Parameter (Functional traits)	Symbol	Units	SOP	NOP	EEP
Uptake rate per unit fine root weight	α	ml.mg^{-1}.day^{-1}	10	10	10
Resource use efficiency	WUE	ml.mg^{-1}	1	1	1
Cost of fine roots	c_r	ml.cm^3.mg^{-1}.day^{-1}	5	5	5
Cost of transportation roots	c_t	ml.cm^3.mg^{-1}.day^{-1}	0	0.2	0.2
Cost of facilitation trait	c_e	ml.cm^3.mg^{-1}.day^{-1}	0	0	1.25
Resource-mining trait	ϕ	ml.mg^{-1}.day^{-1}	0	0	ϕ_I, ϕ_{II}

230   **2.3 Two-step model analysis**

We conducted a two-stage computational analysis. First, we parameterized the model to mimic conditions representing a specific moment in successional and evolutionary time. Then, we studied plant-plant interactions simulating experiments at the life span timescale using the previously parameterized scenarios.

235   *Model parameterization*. To mimic natual conditions in which engineer plants might have evolved the resource mining trait, we allow ϕ to evolve either assuming that soil-engineer plants grow alone, ϕ_I, or competing with spreading opportunistic plants, ϕ_{II}. We obtain these values of ϕ for a range of environmental stresses, from $b = 0$ (high stress) to $b = 1$ (low stress). This first analysis yields evolutionary stable values of the resource mining trait

240   for an ecosystem engineer plant at a given successional stage that adapted growing alone or

surrounded by opportunistic plants.

Model experiments. Then, we take individual plants from the above parameterized

conditions (i.e., using the values ϕ_I or ϕ_{II} expected to evolve at a given *b*) and simulate

experiments with such plants at the scale of their life span. We perform two classes of

245   independent experiments, a first type in which we grow the engineer or the opportunistic

plant alone, and a second type in which we grow an opportunistic-engineer pair of interacting

plants separated by a distance *d*. In this second setup, we assumed that both plants shared the

same insertion point to the soil ($d = 0$) to maximize the strength and effects of the interaction.

For each set of experiments, we obtained the root biomass distribution $R(\ell)$ and calculated

250   the plant-level resource gain associated to such root distribution, $\mathcal{G}$, integrating G [Eq. (5)] in

space. Finally, using the results from both sets of experiments, we estimated the net

interaction between the two plants by comparing the plant-level net resource gain of solitary

and paired individuals. Following a similar procedure, we also calculated the total root

biomass of plants, $\mathfrak{R}$, by integrating the local $R(\ell)$ values across space and estimated the

255   difference in root proliferation by comparing the values for solitary and paired plants.

To measure both the net interaction outcome in pairwise experiments, and to which

extent the opportunistic plant changes its root biomass in the presence of the nurse plant, we

calculated coefficients based on normalized ratios *N* (Dohn *et al.* 2013):

$$N_X = \frac{X_{interact} - X_{solo}}{X_{interact} + X_{solo}} \tag{6}$$

Where *X* takes value $\mathfrak{R}$ or $\mathcal{G}$ depending on whether we measure changes in root production

260   or the net interaction outcome. $X_{interact}$ is the value obtained for $\mathfrak{R}$ or $\mathcal{G}$ when the opportunistic

plant grows close to nurse plant and X_{solo} is the value obtained for solitary plants. N can vary

from -1 to 1. For the net interaction outcome, $N_{\mathcal{G}} = -1$ indicates that the plant cannot coexist

with the neighbor (competitive exclusion); $-1 < N_{\mathcal{G}} < 0$ indicates net interference; $N_{\mathcal{G}} = 0$ no

effect; $0 < N_{\mathcal{G}} < 1$, net facilitation; and $N_{\mathcal{G}} = 1$ indicates that the plant cannot exist without the

265   neighbor (obligatory facilitation). For root biomass measurements, $N_{\mathcal{R}} < 0$ indicates root

underproliferation and $N_{\mathcal{R}} > 0$, overproliferation with respect to a solo plant.

3. Results

3.1 Opportunistic plants do not survive alone and engineer plants evolve in very
270   **stressful habitats.**

Only engineer plants with a mining intensity above a given threshold can colonize and

survive in stressful habitats, represented by values of b below a critical threshold $b_{c\text{-}o}$. Plants

investing in resource mining below this threshold, including opportunistic ones, cannot

survive because their root production and maintenance cost outweigh resource uptake (**Fig.**

275   **2**). The optimal resource mining intensity for solitary soil engineer plants is a finite value ϕ_I

that depends on the environmental stress. Values greater than ϕ_I are suboptimal because the

costs of increasing the mining capacity outweigh the revenue per unit increase. Indeed, at

very high values of ϕ the plant-level net resource gain and the optimal root biomass become

zero. The cost of producing roots with such level of mining ability exceed the amount of

280   resource that such roots can mine and uptake, and the plant cannot establish. The optimal

value of the resource mining trait for solitary individuals, however, does not coincide with

the value of ϕ at which root biomass is maximal for these plants. Values of ϕ lower than ϕ_I

result in more roots, but with lower marginal gain (**Fig. 2a**). Lastly, we find that the presence of spreading opportunistic plants triggers a post-colonization evolutionary drift on the resource mining trait. This shift in the resource mining trait, from ϕ_I to a lower value ϕ_{II} (**Fig. 3a**), is caused by the changes in the resource dynamics prompted by the colonization of opportunistic individuals.

Figure 2: (a)Plant-level net gain and (b) fine root biomass as a function of ϕ (log-scaled axis) and b for a soil engineer plant growing solo. Lines in both panels represent the value of the mining trait that evolves at different successional times, ϕ_I. The point $b_{c\text{-}o}$ is the stress threshold below which opportunistic plants cannot survive (light gray region).

3.2 Shifts from antagonistic facilitation to mutual interference as environmental stress decreases.

Antagonistic facilitation is evolutionary stable in stressful habitats, represented by low values of b, regardless of whether engineer plants evolved resource mining in the presence of opportunistic neighbors (ϕ_{II}) or solitary (ϕ_I) (green shaded region in **Fig. 3b**). This evolutionary stability allows opportunistic plants to survive at the shelter of engineer plants in environments where they would not survive on their own, i.e., where $b < b_{c\text{-}o}$. Moreover, at values of b higher than $b_{c\text{-}o}$ but still low, opportunistic individuals growing close to a soil engineer exhibit higher plant-level net gain than solitary ones. Engineer plants have

a strong positive effect on opportunistic neighbors before they evolve to adapt to its presence (ϕ_I). The evolutionary drift on the resource mining trait from ϕ_I to a lower value ϕ_{II} increases the engineer's net gain and decreases the opportunistic plant's net gain when they interact. However, engineer plants benefit from a non-zero ϕ_{II} value, which also benefits their competitors and allows their survival in highly stressed environments, $b < b_{c\text{-}o}$. These results show that antagonistic facilitation is evolutionary stable in our model, and that engineer plants have a positive effect on opportunistic individuals (facilitation) for any b below a critical shifting point b_s (**Fig. 3b**). We calculate this shifting point using only ϕ_{II} because ϕ evolves much faster than soil successional change and we expect the transition from ϕ_I to ϕ_{II} to be abrupt at the geological scale. Above b_s, the net gain of an opportunistic plant is higher when it grows alone than near an engineer plant. Therefore, the strength of exploitative competition exceeds the positive effect of the engineer's mining trait on the opportunistic plant and the interaction becomes mutual interference.

The amount of resource that engineer plants can mine becomes progressively lower as b increases and more resource is spontaneously available for plants, which also results in smaller net gain differences between solitary engineer and opportunistic plants. We define the critical b for engineer plants $b_{c\text{-}e} < 1$ as the maximum value of b at which resource mining is evolutionarily stable and hence engineer plants can persist. Our results show that above $b_{c\text{-}e}$ the optimal value of the evolved mining trait is $\phi_{II} = 0$, suggesting that engineer plants would evolve to lose the resource mining trait and become opportunistic (**Fig. 3**). In other words, ecosystem engineers cannot persist in very mild environments, and they might lose the

325　mining trait to not be outcompeted, or specialize in resource-poor patches in the case of

heterogeneous ecosystems.

Figure 3: (a) Mining trait ϕ_I in an evolutionary equilibrium resulting from the model
parameterization step of the analysis across an environmental gradient defined by b. (b) Plant-level
330　net resource gain as a function of b for plants growing alone and engineer-opportunistic pairs.
Different lines show different evolutionary equilibria as indicated in the legend: ϕ_I (only shown for
low-b scenarios where engineer plants may have evolved alone) and ϕ_{II}.

3.3 The effects of different environmental parameters on plant interaction and root
335　proliferation

　　　　Finally, we evaluate the effects of the three environmental parameters, b, δ, and ω on

the biotic interaction. We have seen that, for fixed values of δ and ω, ϕ_{II} decreases

progressively as b increases, until $\phi_{II} = 0$ at low stress levels represented by $b = b_{c-e}$. When

the stress is caused by δ or ω, however, ϕ_{II} changes abruptly from 0, at high levels of δ- and

340　ω-stress, to a positive value when stress levels go below a threshold (**Fig. 4a, b**).

The Dohn's coefficient for the plant-level net gain N_G shows that the interaction transitions gradually from facilitation at low levels of b (high b-stress) toward higher levels of b. Indeed, above b_{c-e} the engineer loses the mining trait and hence the result is two opportunistic plants competing exploitatively. For the other environmental parameters, however, N_G shows that plants cannot survive at very high levels of δ-stress and ω-stress (**Fig 4c, d**). At lower stresses, there is an environmental threshold (solid green lines in **Fig. 4c, d**) below which pairs of plants can survive but engineers do not evolve a positive value in their mining trait. For that range of environmental values, defined by moderately high levels of δ-stress and ω-stress pairs of exploitative plants compete similarly to what we observe for $b \geq b_{c-e}$. At even lower δ- and ω-stress, we find a second threshold where engineer plants acquire their positive mining trait in an abrupt shift (dashed red lines in **Fig. 4c, d**). After passing that second threshold the net interaction, measured by N_G, depends strongly on b, but it barely changes with δ and ω. Only for a very narrow range of intermediate values of b, the interaction changes slightly as a function of the other parameters, namely from facilitation to competition as δ-stress decreases, but from competition to facilitation as ω-stress decreases. We observe that the fraction of resource that is available without the intervention of mining b, rather than the potential resource input ω or the physical decay rate of the resource δ, accounts for most of the variation in the sign and the strength of the biotic interaction. This justifies choosing b as the main environmental parameter in our study but also has important consequences in the interpretation of ecological theory.

The analysis of root production using the Dohn's coefficient for root biomass N_R shows that transition from facilitation to competition (**solid golden line in Fig. 4e, f**) does not coincide with transition from opportunistic plants root overproliferation to

underproliferation (**solid blue line in Fig 4e, f**). This result allows for three possible

interaction outcomes. First, when there is antagonistic facilitation the opportunistic plant

always overproliferates roots (F-Ov). Second, plants might interact via mutual interference,

but the opportunistic plant still overproliferate roots as compared to growing alone (I-Ov).

This overproliferation is the result of the tragedy of the commons in root competition (Cabal

2022). Finally, plants might interfere with each other and underproliferate roots (I-Und). The

latter two interaction outcomes can happen for positive values of ϕ_{II} (active mining of the

engineer) or for $\phi_{II} = 0$ (both plants compete equally).

Figure 4: Changes in the resource-mining trait ϕ_{II} (a, b); the Dohn's interaction coefficient for the plant-level net resource gain $N_{\mathcal{G}}$ (c, d); and for root biomass $N_{\mathfrak{R}}$ of the opportunistic plant in interaction with an engineer individual (e, f). We measure each of these changes at different values of the three environmental parameters b, ω and δ. The cases F-Ov, I-Ov and I-Und are describe in the main text.

4. Discussion

4.1 Successional changes in plant biotic interactions

The amount of resource available to plants without the intervention of root mining is the leading force driving the change in plant interactions over successional time. Only ecosystem engineers can colonize highly stressed ecosystems, representing the initial stages of succession ($b < b_{c\text{-}o}$) at which opportunistic plants cannot survive on their own. Nevertheless, after engineers colonize the system, opportunistic plants can establish an obligatory interaction with their benefactors and invade the ecosystem growing at the shelter of their root crown. Previous work has suggested that nurse plants may lose their mining trait over the course of evolution in the presence of opportunistic neighbors that benefit from it at no cost (Walker & Chapin III 1987; Koffel *et al.* 2018; Song *et al.* 2019). Our results, however, suggest that antagonistic facilitation is evolutionarily stable at high environmental stresses because nurse plants benefit from bearing the mining trait despite the presence of opportunistic plants. For environmental stresses higher than a critical value b_s, the engineer plant has an overall positive effect on the opportunistic because the benefits of mining overcome resource competition. In environments milder than b_s, resource competition becomes the dominant force and the net interaction between the soil-engineer and the opportunistic plant transitions from antagonistic facilitation to mutual interference. Finally, for $b > b_{c\text{-}e}$, representing low environmental stress like, for example, in well-developed soils, most of the resource becomes spontaneously available to the plants and all plants evolve an opportunistic strategy.

4.2 Consequences for the stress gradient hypothesis

Some ecosystems, such as deserts, Mediterranean shrublands, or tropical savannas, are more stressful and resource-poor than others, such as temperate or tropical forests, even when successionally mature. The stress gradient hypothesis (SGH) predicts that facilitation must dominate in more stressful habitats while interference must prevail in mild conditions (Bronstein 2009). Because meta-analyses of studies reporting facilitation over stress gradients did not find consistent support for this theory, several authors suggested that stress could be classified as resource-stress and non-resource-stress, and only the former leads to SGH predictions (Bertness & Callaway 1994; He *et al.* 2013). Our model further dissects resource-stress into three parameters; the total input of resource ω (e.g., levels of precipitation), the fraction of such resource that becomes available to plants spontaneously b (e.g., how much of the rainwater infiltrates in bare soil and becomes available to roots without the intervention of ecosystem engineers), and the rate of physical resource decay δ (e.g., how fast the infiltrated water leaves the soil due to evaporation or percolation). Our results support the SGH for resource stress if the stress is caused by b, but the other environmental parameters did not impact the net interaction significantly. We show that b is main driver of shifts in plant biotic interactions across resource stress gradients. Moderate changes in b would dominate and mask the effects on plant biotic interactions of changes in any of the other environmental parameters.

The humped-back SGH, an alternative theory to study plant facilitation and biotic interactions, suggests that plant facilitation maximizes at moderate, rather than highest, levels of stress (Maestre *et al.* 2005, 2009). Our results support humped-back SGH when extreme stress is caused by δ or ω. We observe an abrupt strategy shift in engineer plants that lose

their mining trait when δ or ω drive increases in environmental stress. At a narrow range of intermediate values of b, this shift can cause a sudden transition from interference between two opportunistic plants to facilitation where one plant becomes the nurse plant, followed by a gradual loss of the net facilitation.

Teasing apart the contributions of all three parameters to resource dynamics will allow researchers to elaborate more mechanistic descriptions of the resource stress and therefore develop more comprehensive theories regarding the importance of facilitation across a stress gradient.

4.3. Consequences for empirical studies

Our results have clear consequences for the interpretation of empirical studies addressing interaction shifts across environmental gradients. We have identified two major sources of potential confusion based on our model. First is the fact that b is the main environmental driver of shifts in plant interactions. This parameter, which could represent, for example, water infiltration:runoff ratio, is much harder to determine empirically than the other two environmental parameters ω or δ (e.g., rainfall or soil drying rates). Many empirical studies assess the changes in plant interaction across environmental gradients using the latter two variables, and our model suggests that this may not be the right approach. Second, many studies consider that an increase in plant biomass when plants interact indicates facilitation because they assume that plant biomass is a good proxy for plant fitness (Younginger *et al.* 2017). Our results suggest that plant biomass is a misleading fitness surrogate, despite we find that facilitated plants always overproliferate roots, because plants might also overproliferate roots due to a tragedy of the commons (Gersani *et al.* 2001; Cabal *et al.* 2020; Cabal 2022).

5. Conclusions

Antagonistic facilitation enables more diverse plant communities in stressful environments at the initial stages of primary succession and may accelerate the development of mature soils. It also may promote plant biodiversity in harsh habitats, such as deserts or Mediterranean shrublands, where nurse plants facilitate the survival of other species that depend on them. To mechanistically understand how the plant-plant interaction shifts across resource-stress gradients, it is fundamental to distinguish the biophysical parameters underpinning soil resource availability. We have proved that the main parameter explaining shifts in plant-plant interactions across resource-stress gradients is the proportion of the existing resource that becomes available to roots spontaneously (such as the proportion of rainfall that infiltrates in soil instead of running-off when the resource is water). The amount or resource input (such as rainfall when the resource is water) or the physical decay (such as loses of soil water through evaporation or percolation) play a secondary role. Accounting for this level of mechanistic detail in mathematical models for plant interactions and measuring these parameters in field studies is key to better understand the fundamental drivers of plant-plant interactions.

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References

475   Bai, Y., She, W., Michalet, R., Zheng, J., Qin, S. & Zhang, Y. (2018). Benefactor facilitation
and beneficiary feedback effects drive shrub-dominated community succession in a
semi-arid dune ecosystem. *Appl. Veg. Sci.*, 21, 595–606.

Barron, A.R., Purves, D.W. & Hedin, L.O. (2011). Facultative nitrogen fixation by canopy
legumes in a lowland tropical forest. *Oecologia*, 165, 511–520.

480   Bertness, M.D. & Callaway, R. (1994). Positive interactions in communities. *Trends Ecol.
Evol.*, 9, 191–193.

Britto Costa, P., Staudinger, C., Veneklaas, E.J., Oliveira, R.S. & Lambers, H. (2021). Root
positioning and trait shifts in Hibbertia racemosa as dependent on its neighbour's
nutrient-acquisition strategy . *Plant. Cell Environ.*, 1–11.

485   Bromley, J., Brouwer, J., Barker, A.P., Gaze, S.R. & Valentin, C. (1997). The role of surface
water redistribution in an area of patterned vegetation in a semi-arid environment, south-
west Niger. *J. Hydrol.*, 198, 1–29.

Bronstein, J.L. (2009). The evolution of facilitation and mutualism. *J. Ecol.*, 97, 1160–1170.

Brooker, R.W., Maestre, F.T., Callaway, R.M., Lortie, C.L., Cavieres, L.A., Kunstler, G., *et*

490   *al.* (2008). Facilitation in plant communities: The past, the present, and the future. *J. Ecol.*, 96, 18–34.

Bulleri, F., Bruno, J.F., Silliman, B.R. & Stachowicz, J.J. (2016). Facilitation and the niche: Implications for coexistence, range shifts and ecosystem functioning. *Funct. Ecol.*, 30, 70–78.

495   Bulleri, F., Eriksson, B.K., Queirós, A., Airoldi, L., Arenas, F., Arvanitidis, C., *et al.* (2018). Harnessing positive species interactions as a tool against climate-driven loss of coastal biodiversity. *PLoS Biol.*, 16.

Cabal, C. (2022). Root Tragedy of the Commons: Revisiting the Mechanisms of a Misunderstood Theory. *Front. Plant Sci.*, 1–9.

500   Cabal, C., Martinez-Garcia, R., De Castro, A., Valladares, F. & Pacala, S.W. (2020). The Exploitative Segregation of Plant Roots. *Science (80-.).*, 1199, 1197–1199.

Cabal, C., Martínez-García, R., De Castro, A., Valladares, F. & Pacala, S.W. (2021). Future paths for the "expoitative segregation of plant roots" model. *Plant Signal. Behav.*

Caldwell, M.M. & Richards, J.H. (1989). Hydraulic lift: water efflux from upper roots
505   improves effectiveness of water uptake by deep roots. *Oecologia*, 79, 1–5.

Cavieres, L.A., Hernández-Fuentes, C., Sierra-Almeida, A. & Kikvidze, Z. (2016). Facilitation among plants as an insurance policy for diversity in Alpine communities. *Funct. Ecol.*, 30, 52–59.

Cordero, O.X., Ventouras, L.A., DeLong, E.F. & Polz, M.F. (2012). Public good dynamics
510   drive evolution of iron acquisition strategies in natural bacterioplankton populations.

Proc. Natl. Acad. Sci. U. S. A., 109, 20059–20064.

Crain, C.M. & Bertness, M.D. (2006). Ecosystem Engineering across Environmental Gradients: Implications for Conservation and Management. *Bioscience*, 56, 211–218.

Dakora, F.D. & Phillips, D.A. (2002). Root exudates as mediators of mineral acquisition in low-nutrient environments Felix. *Plant Soil*, 245, 35–47.

Dohn, J., Dembélé, F., Karembé, M., Moustakas, A., Amévor, K.A. & Hanan, N.P. (2013). Tree effects on grass growth in savannas: Competition, facilitation and the stress-gradient hypothesis. *J. Ecol.*, 101, 202–209.

Drescher, K., Nadell, C.D., Stone, H.A., Wingreen, N.S. & Bassler, B.L. (2014). Solutions to the public goods dilemma in bacterial biofilms. *Curr. Biol.*, 24, 50–55.

Estrela, S., Libby, E., Van Cleve, J., Débarre, F., Deforet, M., Harcombe, W.R., *et al.* (2018). Environmentally Mediated Social Dilemmas. *Trends Ecol. Evol.*, 0.

Foster, K.R. (2010). Social behaviour in microorganisms. *Soc. Behav. Genes, Ecol. Evol.*, 02138, 331–356.

Gersani, M., Brown, J.S., O'Brien, E.E., Maina, G.M. & Abramsky, Z. (2001). Tragedy of the Commons as a Result of Root Competition. *J. Ecol.*, 89, 660–669.

Gore, J., Youk, H. & van Oudenaarden, A. (2009). Snowdrift game dynamics and facultative cheating in yeast. *Nature*, 459, 253–6.

Hastings, A., Byers, J.E., Crooks, J.A., Cuddington, K., Jones, C.G., Lambrinos, J.G., *et al.* (2007). Ecosystem engineering in space and time. *Ecol. Lett.*, 10, 153–164.

Hauert, C., Wakano, J.Y. & Doebeli, M. (2008). Ecological public goods games: Cooperation

and bifurcation. *Theor. Popul. Biol.*, 73, 257–263.

He, Q., Bertness, M.D. & Altieri, A.H. (2013). Global shifts towards positive species interactions with increasing environmental stress. *Ecol. Lett.*, 16, 695–706.

535   Hinsinger, P., Brauman, A., Devau, N., Gérard, F., Jourdan, C., Laclau, J.P., *et al.* (2011). Acquisition of phosphorus and other poorly mobile nutrients by roots. Where do plant nutrition models fail? *Plant Soil*, 348, 29–61.

Jones, D.L. (1998). Organic acids in the rhizosphere - A critical review. *Plant Soil*, 205, 25–44.

540   Karadimou, E., Kallimanis, A.S., Tsiripidis, I., Raus, T., Bergmeier, E. & Dimopoulos, P. (2018). Functional diversity changes over 100 yr of primary succession on a volcanic island: insights into assembly processes. *Ecosphere*, 9.

Kéfi, S., Van Baalen, M., Rietkerk, M. & Loreau, M. (2008). Evolution of local facilitation in arid ecosystems. *Am. Nat.*, 172.

545   Kjær, U., Olsen, S.L. & Klanderud, K. (2018). Shift from facilitative to neutral interactions by the cushion plant Silene acaulis along a primary succession gradient. *J. Veg. Sci.*, 29, 42–51.

Koffel, T., Boudsocq, S., Loeuille, N. & Daufresne, T. (2018). Facilitation- vs. competition-driven succession: the key role of resource-ratio. *Ecol. Lett.*, 21, 1010–1021.

550   de Kroons, H. & Hutchings, M.J. (1995). Morphological Plasticity in Clonal Plants: The Foraging Concept Reconsidered. *J. Ecol.*, 83, 143.

Lambers, H., Raven, J.A., Shaver, G.R. & Smith, S.E. (2008). Plant nutrient-acquisition

strategies change with soil age. *Trends Ecol. Evol.*, 23, 95–103.

Losapio, G., Norton Hasday, E., Espadaler, X., Germann, C., Ortiz-Sánchez, F.J., Pont, A., *et al.* (2021a). Facilitation and biodiversity jointly drive mutualistic networks. *J. Ecol.*, 109, 2029–2037.

Losapio, G., Schmid, B., Bascompte, J., Michalet, R., Cerretti, P., Germann, C., *et al.* (2021b). An experimental approach to assessing the impact of ecosystem engineers on biodiversity and ecosystem functions. *Ecology*, 102, 1–12.

Lozano, Y.M., Armas, C., Hortal, S., Casanoves, F. & Pugnaire, F.I. (2017). Disentangling above- and below-ground facilitation drivers in arid environments: the role of soil microorganisms, soil properties and microhabitat. *New Phytol.*, 216, 1236–1246.

Maestre, F.T., Callaway, R.M., Valladares, F. & Lortie, C.J. (2009). Refining the stress-gradient hypothesis for competition and facilitation in plant communities. *J. Ecol.*, 97, 199–205.

Maestre, F.T., Valladares, F. & Reynolds, J.F. (2005). Is the change of plant-plant interactions with abiotic stress predictable? A meta-analysis of field results in arid environments. *J. Ecol.*, 93, 748–757.

Montaña, C. (1992). The Colonization of Bare Areas in Two-Phase Mosaics of an Arid Ecosystem. *J. Ecol.*, 80, 315–327.

Mori, A.S., Osono, T., Cornelissen, J.H.C., Craine, J. & Uchida, M. (2017). Biodiversity–ecosystem function relationships change through primary succession. *Oikos*, 126, 1637–1649.

Nowak, M.A. (2006). Five rules for the evolution of cooperation. *Science*, 314, 1560–1563.

575   Prieto, I., Armas, C. & Pugnaire, F.I. (2012). Water release through plant roots: New insights into its consequences at the plant and ecosystem level. *New Phytol.*, 193, 830–841.

Pugnaire, F.I., Armas, C. & Valladares, F. (2004). Soil as a mediator in plant-plant interactions in a semi-arid community. *J. Veg. Sci.*, 15, 85–92.

Rainey, P.B. & Rainey, K. (2003). Evolution of cooperation and conflict in experimental
580   bacterial populations. *Nature*, 425, 72–74.

Raven, J.A., Lambers, H., Smith, S.E. & Westoby, M. (2018). Costs of acquiring phosphorus by vascular land plants: patterns and implications for plant coexistence. *New Phytol.*, 217, 1420–1427.

Schöb, C., Butterfield, B.J. & Pugnaire, F.I. (2012). Foundation species influence trait-based
585   community assembly. *New Phytol.*, 196, 824–834.

Schöb, C., Prieto, I., Armas, C. & Pugnaire, F.I. (2014). Consequences of facilitation: One plant's benefit is another plant's cost. *Funct. Ecol.*, 28, 500–508.

Soliveres, S., Smit, C. & Maestre, F.T. (2015). Moving forward on facilitation research: response to changing environments and effects on the diversity, functioning and
590   evolution of plant communities. *Biol. Rev. Camb. Philos. Soc.*, 90, 297–313.

Song, M., Yu, L., Jiang, Y., Korpelainen, H. & Li, C. (2019). Increasing soil age drives shifts in plant-plant interactions from positive to negative and affects primary succession dynamics in a subalpine glacier forefield. *Geoderma*, 353, 435–448.

Sprent, J.I. (1989). Which steps are essential for the formation of functional legume nodules ?

595   *New*, 111, 129–153.

Tarnita, C.E. (2017). The ecology and evolution of social behavior in microbes. *J. Exp. Biol.*, 220, 18–24.

van Velzen, R., Doyle, J.J. & Geurts, R. (2019). A Resurrected Scenario: Single Gain and Massive Loss of Nitrogen-Fixing Nodulation. *Trends Plant Sci.*, 24, 49–57.

600   Verdú, M., Gómez, J.M., Valiente-Banuet, A. & Schöb, C. (2021). Facilitation and plant phenotypic evolution. *Trends Plant Sci.*, 26, 913–923.

Wakano, J.Y., Nowak, M.A. & Hauert, C. (2009). Spatial dynamics of ecological public goods. *Proc. Natl. Acad. Sci. U. S. A.*, 106, 7910–7914.

Walker, L.R. & Chapin III, F.S. (1987). Interactions among Processes Controlling
605   Successional Change. *Oikos*, 50, 131–135.

West, S.A., Griffin, A.S., Gardner, A. & Diggle, S.P. (2006). Social evolution theory for microorganisms. *Nat. Rev. Microbiol.*, 4, 597–607.

Wright, A.J., Wardle, D.A., Callaway, R.M. & Gaxiola, A. (2017). The Overlooked Role of Facilitation in Biodiversity Experiments. *Trends Ecol. Evol.*, 32, 383–390.

610   Younginger, B.S., Sirová, D., Cruzan, M.B. & Ballhorn, D.J. (2017). Is Biomass a Reliable Estimate of Plant Fitness? *Appl. Plant Sci.*, 5, 1–8.

Zapater, M., Hossann, C., Bréda, N., Bréchet, C., Bonal, D. & Granier, A. (2011). Evidence of hydraulic lift in a young beech and oak mixed forest using 18O soil water labelling.*Trees - Struct. Funct.*, 25, 885–894.

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