Bayesian hierarchical modeling of size spectra

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Abstract

A fundamental pattern in ecology is that smaller organisms are more abundant than larger

organisms. This pattern is known as the individual size distribution (ISD), which is the frequency

of all individual sizes in an ecosystem, regardless of taxon. The ISD is described by power law

distribution with the form $f(x) = Cx^\lambda$, and a major goal of size spectra analyses is to estimate

the ISD parameter λ. However, while numerous methods have been developed to do this, they

have focused almost exclusively on estimating λ from single samples. Here, we develop an

extension of the truncated Pareto distribution within the probabilistic modeling language Stan.

We use it to estimate multiple ISD parameters simultaneously with a hierarchical modeling

approach. The most important result is the ability to examine hypotheses related to size spectra,

including the assessment of fixed and random effects, within a single Bayesian generalized

(non)-linear mixed model.

Keywords: *Bayesian, body size spectra, hierarchical, Pareto, power law, Stan*

Introduction

In any ecosystem, large individuals are typically more rare than small individuals. This fundamental feature of ecosystems leads to a remarkably common pattern in which relative abundance declines with individual body size, generating the individual size distribution (ISD), also called the community size spectrum (Sprules et al. 1983, White et al. 2008). Understanding how body sizes are distributed has been a focus in ecology for over a century (Peters and Wassenberg 1983), in part because they represent an ataxic approach that reflects fundamental measures of ecosystem structure and function, such as trophic transfer efficiency. (Kerr and Dickie 2001, White et al. 2007, Perkins et al. 2019). Individual size distributions are also predicted as a result of physiological limits associated with body size, thereby emerging from predictions of metabolic theory and energetic equivalence (Brown et al. 2004).

More formally, the ISD is a frequency distribution that can be approximated by a bounded power law with a single free parameter λ, corresponding to the following probability density function (Edwards et al. 2020):

$$f(x) = Cx^{\lambda}, x_{min} \leq x \geq x_{max}$$

where x is the body size (e.g., mass or volume) of an individual regardless of taxon, x_{min} is the smallest individual attainable and x_{max} is the largest possible individual (White et al. 2008). C is a constant equal to:

$$C = \begin{cases} \dfrac{\lambda + 1}{x_{max}^{\lambda+1} - x_{min}^{\lambda+1}}, \lambda \neq -1 \\ \dfrac{1}{log x_{max} - log x_{min}}, \lambda = -1 \end{cases}$$

This model is also known as a bounded power law or truncated Pareto distribution. The terms

"bounded" or "truncated" refer to the limits of x_{min} and x_{max}, which represent the minimum and

maximum attainable body size values (White et al. 2008). In practice, values of x_{min} and x_{max}

often come from the minimum and maximum body sizes in a data set or are estimated

statistically (White et al. 2008, Edwards et al. 2017).

A compelling feature of size spectra is that λ may vary little across ecosystems as a result of

physiological constraints that lead to size-abundance patterns more broadly. Metabolic scaling

theory predicts $\lambda + 1 = \frac{log10\alpha}{log10\beta} - 3/4$, where α is trophic transfer efficiency in the food web and

β is the mean predator-prey mass ratio (Reuman et al. 2008). The value of $-3/4$ is the predicted

scaling exponent of log abundance and log mass (Damuth 1981, Peters and Wassenberg 1983). It

is the reciprocal of scaling coefficient of metabolic rate and mass (0.75) (Brown et al. 2004) and

as a result, values of $\lambda + 1$ have been used to estimate metabolic scaling across ecosystems

(Reuman et al. 2008, Perkins et al. 2018, 2019). Because $\frac{log10\alpha}{log10\beta}$ is typically <<0.01, this implies

that a λ values of ~-1.75 represent a reasonable first guess of expected ISD exponents, with

values of ranging from -1.2 to -2 often appearing in the literature (Andersen and Beyer 2006,

Blanchard et al. 2009, Pomeranz et al. 2020b).

Whether λ represents a fixed or variable value is debated, but it often varies among samples

and ecosystems (Blanchard et al. 2009, Perkins et al. 2018, Pomeranz et al. 2020b). It is often

described by its "steepness", with more negative values (i.e., "steeper") indicating lower

abundance of large relative to small individuals, and vice versa. These patterns of size frequency

are an emergent property of demographic processes (e.g., age-dependent mortality), ecological

interactions (e.g., size-structured predation, trophic transfer efficiency), and physiological

constraints (e.g., size-dependent metabolic rates) (Muller-Landau et al. 2006, Andersen and

Beyer 2006, White et al. 2008). As a result, variation in λ across ecosystems or across time can

indicate fundamental shifts in community structure or ecosystem functioning. For example,

overfishing in marine communities has been detected using size spectra in which λ was steeper

than expected, indicating fewer large fish than expected (Jennings and Blanchard 2004). Shifts in

λ have also been used to document responses to acid mine drainage in streams (Pomeranz et al.

2019, 2020a), land use (Martínez et al. 2016), resource subsidies (Perkins et al. 2018), and

temperature (O'Gorman et al. 2017, Pomeranz et al. 2022).

Given the ecological information it conveys, the data required to estimate size spectra are

deceptively simple; only a single column of data are needed, in which each data point is a single

measure of the body size of an individual. As long as the body sizes are collected systematically

and without bias towards certain taxa or phenotypes, there is no need to know any more

ecological information about the data points (e.g., taxon, trophic position, age, abundance).

However, despite the simple data requirement, the statistical models used to estimate λ are

diverse. Edwards et al. (2017) documented 8 different analytical methods. Six involved binning,

in which the body sizes are grouped into size bins (e.g., 2-49 mg, 50-150 mg, etc.) and then

counted, generating values for abundance within each size bin. When both axes are log-

transformed, binning allows λ to be estimated using simple linear regression. Unfortunately, the

binning process also removes most of the variation in the data, collapsing information about

1000's of individuals into just 6 or so bins. Doing so can lead to the wrong values of λ,

sometimes drastically so (White et al. 2008, Edwards et al. 2017, 2020).

An improved alternative to binning and linear regression is to fit the body size data to a

power law probability distribution (White et al. 2008, Edwards et al. 2017, 2020). This method

uses all raw data observations directly to estimate λ, typically using the maximum likelihood

estimation method (Edwards et al. 2017). In addition to estimating size spectra of single samples,

ecologists have used this method to examine how λ varies across environmental gradients

(Perkins et al. 2019, Pomeranz et al. 2022). However, these analyses typically proceed in two

steps. First, λ estimates are obtained individually from each collection (e.g., each site or year,

etc.). Second, these estimates are used as response variables in a linear model to examine how

they relate to corresponding predictor variables (Edwards et al. 2020). A downside to this

approach is that it treats body sizes (and subsequent λ's) as independent samples, even if they

come from the same site or time. It also removes information on sample size (number of

individuals) used to derive λ. As a result, the approach not only separates the data generation

model from the predictor variables, but is also unable to take advantage of partial pooling during

model fitting.

Here, we develop a Bayesian model that uses the truncated Pareto distribution to estimate λ

in response to both fixed and random predictor variables. The model extends the maximum

likelihood approach developed by Edwards et al. (2020) and allows for a flexible hierarchical

structure, including partial pooling, within the modeling language Stan (Stan Development Team

2022).

Methods

<u>Translating to Stan</u>

105   We first translated the probability density function described by Edwards et al. (2020) into

106   Stan by converting it to the log probability density function (lpdf). Stan is a probabilistic

107   modeling language that is capable of fitting complex models, including those with custom lpdf's.

108   The resulting lpdf is given as

$$lpdf = \begin{cases} \log\dfrac{\lambda+1}{x_{max}^{\lambda+1} - x_{min}^{\lambda+1}} + \lambda\log x, \lambda \neq -1 \\ \text{-}\log(\log x_{max} - \log x_{min}) - \log x, \lambda = -1 \end{cases}$$

109   with all variables as described above. We call this the $paretocustom$ distribution, which we can

110   now use to estimate λ of a given data set. For example, an intercept-only model would look like

111   this:

$$x_i \sim paretocustom(\lambda, x_{min}, x_{max})$$

$$\lambda = \alpha$$

$$\alpha \sim Normal(\mu, \sigma)$$

where x_i is the ith individual body size, λ is the size spectrum parameter (also referred to as the

exponent), x_{min} and x_{max} are as defined above, and α is the intercept with a prior probability

distribution. In this case, we specified a Normal prior since λ is continuous and can be positive or

negative, but this can be changed as needed.

The simple model above can be expanded to a generalized linear mixed model by including

fixed predictors ($\boldsymbol{\beta}X$) and/or varying intercepts ($\alpha_{[x]}$):

$$x_{ij} \sim paretocustom\left(\lambda_j, x_{min,j}, x_{max,j}\right)$$

$$\lambda = \alpha + \boldsymbol{\beta}X + \alpha_{[j]} + \alpha_{[x]}$$

$$\alpha \sim Normal(\mu_\alpha, \sigma_\alpha)$$

$$\beta \sim Normal\left(\mu_\beta, \sigma_\beta\right)$$

$$\alpha_{[j]} \sim Normal\left(0, \sigma_{[j]}\right)$$

$$\sigma_{[j]} \sim Exponential(\phi)$$

$$\alpha_{[x]} \sim Normal\left(0, \sigma_{[x]}\right)$$

$$\sigma_{[x]} \sim Exponential(\phi)$$

with one or more β regression parameters, represented by the vector $\boldsymbol{\beta}$, for one or more fixed

predictors X, and one or more varying intercepts α_x. We specify α_j separately because it is

needed to account for the non-independence of body sizes. In other words, each body size x_i is

clustered within each site and so they are not independent and identically distributed. The

addition of a varying intercept for each sample accounts for this non-independence. Prior

distributions are given as $Normal$ for the parameters and varying intercept and $Exponential$ for

$\sigma[x]$, but these can also be changed as needed.

The model above assumes that each body size x represents a single individual such that the

data set might have many repeats for individuals of the same size (e.g., x = {0.2, 0.2, 0.2, 0.4,

0.4, 0.5, 9.8}). However, when individual body sizes are repeated in a data set, they are often

accompanied by a count or density, such that the data set above might instead consist of two

columns with x = {0.2, 0.4, 0.5, 9.8} and $counts$ = {3, 2, 1, 1}. To analyze this more compact

data set, Edwards et al. (2020) developed a modification of the log probability density function

to include $counts$:

$$lpdf = \begin{cases} \text{counts}\left(\log\dfrac{\lambda+1}{x_{max}^{\lambda+1} - x_{min}^{\lambda+1}} + \lambda\log x\right), \lambda \neq -1 \\ \text{counts}(\text{-}\log(\log x_{max} - \log x_{min}) - \log x, \lambda = -1 \end{cases}.$$

132   We refer to this as $paretocounts$, such that the model can be fit by using

$$x_i \sim paretocounts(\lambda, x_{min}, x_{max}, counts)$$

$$\lambda = [\text{linear or non-linear model}] \text{ and}$$

$$[\text{priors}].$$

Aside from adding $counts$, the model is the same as presented beforeS. These models

($paretocustom$ and $paretocounts$) allow us to test how the size distribution parameter, λ,

varies in response to continuous or categorical predictors and to include hierarchical structure as

needed.

<u>Testing the models</u>

The $paretocustom$ and $paretocounts$ lpdfs give the same results, differing only in how the

data are aggregated. For simplicity, we demonstrate model performance here for the

$paretocounts$ distribution, since the empirical data we used (see *Case Study* below) contains

counts of individual body sizes. First, we tested for parameter recovery using data simulated

from a bounded power law with known values of λ. Second, we fit the model to fisheries trawl

data presented in Edwards et al. (2020) to estimate the hypothesis that λ declines over time.

<u>Parameter recovery from simulated data</u>

To ensure that the models could recover known parameter values, we simulated ten data sets

from a bounded power law using the inverse cumulative density function:

$$x_i = \left(u_i x_{max}^{(\lambda+1)} + (1 - u_i) x_{min}^{(\lambda+1)} \right)^{\frac{1}{(\lambda+1)}}$$

where x_i is the individual body size from the ith simulation, u_i is a unique draw from a

$Uniform(0,1)$ distribution, and all other variables are the same as defined above. We set $x_{min} =$

1, $x_{max} = 1000$, and simulated $i = 1000$ values from each of 10 λ's ranging from -2.2 to -1.2. To

generate $counts$, we rounded each simulated value to the nearest 0.001 and then tallied them.

We estimated the ten λ values in two ways. First, we fit a separate intercept-only model to

each of the ten data sets. Second, we fit a varying intercept model (Gelman et al. 2014). The

structure of this model is $\lambda = \alpha + \alpha_{[group]}$ where each group represents an offset from the mean

value of lambda.

Finally, we simulated data for a regression model with a single continuous predictor and a

varying intercept: $\lambda = \alpha + \beta x + \alpha_{[group]}$, where α = -1.5, β = -0.1, and σ_{group} = 0.3. The

predictor variable x was a continuous predictor. Using these parameters, we simulated 18 λ's,

with each λ coming from one of three x-values (-2, 0, 2), nested within 3 groups with each

replicated twice. From each λ, we simulated 1000 individuals using the procedure above, with

$x_{min} = 1$ and $x_{max} = 1000$. Using those 18,000 simulated body sizes (1000 sizes simulated from

18 λ's), we fit a $paretocounts$ regression model 40 times to measure variation in parameter

recovery among model runs.

<u>Sample Size</u>

We examined sensitivity to sample size (number of individual body sizes) across three λ

values (-2, -1.6, -1.2). For each λ, we varied the number of simulated individuals from 2 to 2048,

representing a 2^n sequence with n ranging from 1 to 11. Each of the 11 densities was replicated

10 times resulting in 110 datasets of individual body sizes. We fit each data set using separate

intercept-only $paretocounts$ models and then plotted the resulting λ values as a function of

sample size.

<u>Case Studies</u>

To examine model performance on empirical data, we re-ran a previously published analysis

from Edwards et al. (2020). In Edwards' study, size spectra parameters were first estimated

separately for each sample using maximum likelihood. Then the modeled parameters were used

as response variables in linear regression models. The goal was to test for linear changes in size

spectra over three decades using bi-yearly size data of marine fishes collected from the

International Benthic Trawl Survey (IBTS). The data set and original model results are available

in the sizeSpectra package (Edwards et al. 2017). We tested the same hypothesis as Edwards et

al. (2020), but instead of using a two-step process we fit a single model using the $paretocounts$

lpdf.

<u>Model Fitting</u>

We fit each of the above models in rstan (Stan Development Team 2022) using 2 chains each

with 1000 iterations. All models converged with R_{hat}'s <1.01. If a known parameter value fell

inside the 95% Credible Intervals, we considered parameter recovery successful. For the

replicated regression model, we also tallied the number of times that the known value fell outside

of the 95% CrI.

<u>Data Availability Statement</u>

All data, R code, and Stan code are available at <u>https://github.com/jswesner/stan_isd</u> (to be

permanently archived on acceptance).

Results

<u>Parameter Recovery</u>

For models fit to simulated individual data sets, all 95% credible intervals included the true

value of λ and posterior medians were no more than 0.05 units away from the true value (Table

1). Similarly, when the same data set was fit using a varying intercepts model, the posterior

median intercept α and group standard deviation σ_{group} were nearly identical to the true values

(Table 1). Using the varying intercept model to estimate group specific means yielded similar

results as using separate models per group (Figure 1), demonstrating that a single model can be

used to estimate multiple size spectra.

We also recovered regression parameters (α, β) along with the group-level standard

deviation (σ_{group} (Figure 2). Thirty-seven of the 40 models converged. Of those 37 models the

true value fell outside of the 95% CrI once for α and σ_{group} and three times for β (Figure 2).

Averaging the deviations (posterior median minus the true value) among the replicates indicated

no bias in the modeled estimates (mean bias $\pm$ sd: α = -0.01 $\pm$ 0.05, β = 0001 $\pm$ 0.004, σ_{group} =

0.02 $\pm$ 0.05).

<u>Sample Size</u>

Variation in modeled estimates was high for samples containing less than 100 individual

(Figure 3). For example, when the true λ value was -2, samples with just 8 individuals yielded

estimates ranging from -2.7 to -1.7. By contrast, all samples with more than 300 individuals

captured the true λ with less than 0.1 unit of error (Figure 3).

<u>Case Study</u>

Using IBTS data (Edwards et al. 2017) with a Bayesian hierarchical regression, we found a

negative trend over time. The ISD parameter of IBTS trawl data declined by ~0.001 units per

year, but with a 95% CrI ranging from -0.005 to 0.002. These values were nearly identical to

those reported by Edwards et al. (2020) using a two-step approach (Table 2). An advantage of

fitting the model in a single Bayesian hierarchical framework is that estimates for individual

groups are pulled toward the mean via partial pooling. This is apparent in comparing the

unpooled MLE estimates (Figure 4a) to the partially pooled Bayesian estimates in each year

(Figure 4b).

<u>Discussion</u>

The most important result of this work is the ability to analyze ISD parameters using fixed

and random predictors in a hierarchical model. Our approach allows ecologists to test hypotheses

about size spectra while avoiding the pitfalls of binning, which loses information and can lead to

biased estimates of λ (White et al. 2008). Maximum likelihood solves this problem by directly

estimating the ISD, but testing hypotheses with maximum likelihood is often done with a two-

step process in which λ is estimated individually for each sample and the results are then used as

response variables in linear or non-linear models (Edwards et al. 2020). Our approach merges

these steps, allowing for the incorporation of prior probabilities and hierarchical structure.

The ability to incorporate prior information using Bayesian updating has two practical

advantages over the two-step process described above. First, adding informative prior

distributions can improve model fit by limiting the MCMC sampler to reasonable sampling

space. In other words it would not be sensible to estimate the probability that λ is -1,234 or -9.

Without informative priors, those values (and more extreme values) are considered equally likely

and hence waste much of the algorithm's sampling effort on unlikely values (e.g., Wesner and

Pomeranz (2021)).

Second, and most importantly, ecologists have much prior information on the values that λ

can take. For example, global analysis of phytoplankton reveals values of -1.75, consistent with

prediction based on sub-linear scaling of metabolic rate with mass of -3/4 (Perkins et al. 2019).

Alternatively, Sheldon's conjecture suggests that λ is -2.05 (Andersen et al. 2006), a value

reflecting isometric scaling of metabolic rate and mass, with support in pelagic marine food webs

(Andersen and Beyer 2006). However, benthic marine systems typically have shallower

exponents (e.g., $\sim$ -1.4; Blanchard et al. (2009)), similar to those in some freshwater stream

ecosystems (-(Pomeranz et al. 2022). While the causes of these deviations from theoretical

predictions are debated, it is clear that values of λ are restricted to a relatively narrow range

between about -2.05 and -1.2. But this restriction is not known to the truncated Pareto, which has

no natural lower or upper bounds on λ (White et al. 2008). As a result, a prior that places most of

its probability mass on these values (e.g., $Normal(-1.75, 0.2)$ seems appropriate. Such a

continuous prior does not prevent findings of larger or smaller λ, but instead places properly

weighted skepticism on such values.

Similar to priors, partial pooling from varying intercepts provides additional benefits,

allowing for the incorporation of hierarchical structure and pulling λ estimates towards the global

mean (Gelman 2005, Qian et al. 2010). In the examples shown here, the amount of pooling is

relatively small because the sample sizes are large (>1000 individuals). However, the primary

benefit of pooling (both from varying effects and skeptical priors) is in prediction (Gelman 2005,

Hobbs and Hooten 2015). This becomes especially important when models are used to forecast

future ecosystem conditions. Forecasts are becoming more common in ecology (Dietze et al.

2018) and are likely to be easier to test with modern long-term data sets like NEON (National

Ecological Observatory Network) in which body size samples will be collected at the continental

scale over at least the next 20 years (Kuhlman et al. 2016). In addition, because the effects of

priors and pooling increase with smaller samples sizes, varying intercepts are likely to be

particularly helpful for small samples. In other words, priors and partial pooling contain built-in

skepticism of extreme values, ensuring the maxim that "extraordinary claims require

extraordinary evidence".

One major drawback to the Bayesian modeling framework here is time. Bayesian models of

even minimal complexity must be estimated with Markov Chain Monte Carlo techniques. In this

study, we used the No U-Turn sampling (NUTS) algorithm via rstan (Stan Development Team

2022). Stan can be substantially faster than other commonly used programs such as JAGS and

WinBUGS, which rely on Gibbs sampling. For example, Stan is 10 to 1000 times more efficient

than JAGS or WinBUGS, with the differences becoming greater as model complexity increases

(Monnahan et al. 2017). In the current study, intercept-only models for individual samples with

$\sim$ 300 to 1500 individuals could be fit quickly ($<$2 seconds total run time (warm-up + sampling

on a Lenovo T490 with 16GB RAM)) with as little as 1000 iterations and two chains. However,

the IBTS regression models took $>$2 hours to run with the same iterations and chains. These

times include the fact that our models used several optimization techniques, such as informative

priors, standardized predictors, and non-centered parameterization, each of which are known to

improve convergence and reduce sampling time (McElreath 2016). But if Bayesian inference is

desired, these run-times may be worth the wait. In addition, they are certain to become faster

with the refinement of existing algorithms and the introduction of newer ones like

Microcanonical HMC (Robnik et al. 2022).

Body size distributions in ecosystems have been studied for decades, yet comprehensive

analytical approaches to testing these hypotheses are lacking. We present a single analytical

approach that takes advantage of the underlying data structures of individual body sizes (Pareto

distributions) while placing them in a generalized (Non)-linear hierarchical modeling framework.

We hope that ecologists will adopt and improve on the models here to critically examine

hypotheses of size spectra or other power-law distributed data.

Acknowledgements

285   This material is based upon work supported by the National Science Foundation under Grant

286   Nos. 2106067 to JSW and 2106068 to JRJ. We especially thank Edwards et al. (2017) and

287   (2020) for placing their code and data in easily accessible repositories.

References

Andersen, K. H., and J. E. Beyer. 2006. Asymptotic Size Determines Species Abundance in the
Marine Size Spectrum. The American Naturalist 168:54–61.

Blanchard, J. L., S. Jennings, R. Law, M. D. Castle, P. McCloghrie, M.-J. Rochet, and E. Benoît.
2009. How does abundance scale with body size in coupled size-structured food webs? Journal
of Animal Ecology 78:270–280.

Brown, J. H., J. F. Gillooly, A. P. Allen, V. M. Savage, and G. B. West. 2004. Toward a
metabolic theory of ecology. Ecology 85:17711789.

Damuth, J. 1981. Population density and body size in mammals. Nature 290:699–700.

Dietze, M.C., A. Fox, L. M. Beck-Johnson, J. L. Betancourt, M. B. Hooten, C. S. Jarnevich, T.
H. Keitt, M. A. Kenney, C. M. Laney, L. G. Larsen, and H. W. Loescher. 2018. Iterative near-
term ecological forecasting: Needs, opportunities, and challenges. Proceedings of the National
Academy of Sciences, 115:1424-1432.

Edwards, A. M., J. P. W. Robinson, J. L. Blanchard, J. K. Baum, and M. J. Plank. 2020.
Accounting for the bin structure of data removes bias when fitting size spectra. Marine Ecology
Progress Series 636:19–33.

Edwards, A. M., J. P. W. Robinson, M. J. Plank, J. K. Baum, and J. L. Blanchard. 2017. Testing
and recommending methods for fitting size spectra to data. Methods in Ecology and Evolution
8:57–67.

Gelman, A. 2005. Analysis of variancewhy it is more important than ever. The Annals of
Statistics 33:1–53.

Hobbs, N. T., and M. B. Hooten. 2015. Bayesian models: A statistical primer for ecologists.
Princeton University Press.

Jennings, S., and J. L. Blanchard. 2004. Fish abundance with no fishing: Predictions based on
macroecological theory. Journal of Animal Ecology:632–642.

Kerr, S. R., and L. M. Dickie. 2001. The biomass spectrum: A predator-prey theory of aquatic
production. Columbia University Press.

Kuhlman, M. R., H. W. Loescher, R. Leonard, R. Farnsworth, T. E. Dawson, and E. F. Kelly.
2016. A New Engagement Model to Complete and Operate the National Ecological Observatory
Network. The Bulletin of the Ecological Society of America 97:283–287.

Martínez, A., A. Larrañaga, A. Miguélez, G. Yvon-Durocher, and J. Pozo. 2016. Land use
change affects macroinvertebrate community size spectrum in streams: the case of *Pinus radiata*
plantations. Freshwater Biology 61:69–79.

McElreath, R. 2016. Statistical rethinking: A bayesian course with examples in R and stan.
Second. CRC Press.

Monnahan, C. C., J. T. Thorson, and T. A. Branch. 2017. Faster estimation of Bayesian models
in ecology using Hamiltonian Monte Carlo. Methods in Ecology and Evolution 8:339–348.

Muller-Landau, H. C., R. S. Condit, K. E. Harms, C. O. Marks, S. C. Thomas, S.
Bunyavejchewin, G. Chuyong, L. Co, S. Davies, R. Foster, and others. 2006. Comparing tropical
forest tree size distributions with the predictions of metabolic ecology and equilibrium models.
Ecology letters 9:589–602.

O'Gorman, E. J., L. Zhao, D. E. Pichler, G. Adams, N. Friberg, B. C. Rall, A. Seeney, H. Zhang,
D. C. Reuman, and G. Woodward. 2017. Unexpected changes in community size structure in a
natural warming experiment. Nature Climate Change 7:659–663.

Perkins, D. M., I. Durance, F. K. Edwards, J. Grey, A. G. Hildrew, M. Jackson, J. I. Jones, R. B.
Lauridsen, K. Layer-Dobra, M. S. A. Thompson, and G. Woodward. 2018. Bending the rules:
Exploitation of allochthonous resources by a top-predator modifies size-abundance scaling in
stream food webs. Ecology Letters 21:1771–1780.

Perkins, D. M., A. Perna, R. Adrian, P. Cermeño, U. Gaedke, M. Huete-Ortega, E. P. White, and
G. Yvon-Durocher. 2019. Energetic equivalence underpins the size structure of tree and
phytoplankton communities. Nature Communications 10:255.

Peters, R. H., and K. Wassenberg. 1983. The effect of body size on animal abundance. Oecologia
60:89–96.

Pomeranz, J. P. F., J. R. Junker, and J. S. Wesner. 2022. Individual size distributions across
North American streams vary with local temperature. Global Change Biology 28:848–858.

Pomeranz, J. P. F., H. J. Warburton, and J. S. Harding. 2019. Anthropogenic mining alters
macroinvertebrate size spectra in streams. Freshwater Biology 64:81–92.

Pomeranz, J. P. F., J. S. Wesner, and J. S. Harding. 2020a. Changes in stream food-web structure
across a gradient of acid mine drainage increase local community stability. Ecology 101:e03102.

Pomeranz, J., J. S. Wesner, and J. S. Harding. 2020b. Changes in stream food-web structure
across a gradient of acid mine drainage increases local community stability. Ecology.

Qian, S. S., T. F. Cuffney, I. Alameddine, G. McMahon, and K. H. Reckhow. 2010. On the
application of multilevel modeling in environmental and ecological studies. Ecology 91:355–
361.

Reuman, D. C., C. Mulder, D. Raffaelli, and J. E. Cohen. 2008. Three allometric relations of
population density to body mass: Theoretical integration and empirical tests in 149 food webs.
Ecology Letters 11:1216–1228.

Robnik, J., G. B. De Luca, E. Silverstein, and U. Seljak. 2022, December. Microcanonical
Hamiltonian Monte Carlo. arXiv.

Sprules, W. G., J. M. Casselman, and B. J. Shuter. 1983. Size Distribution of Pelagic Particles in
Lakes. Canadian Journal of Fisheries and Aquatic Sciences 40:1761–1769.

Stan Development Team. 2022. RStan: The R interface to Stan.

Wesner, J. S., and J. P. F. Pomeranz. 2021. Choosing priors in Bayesian ecological models by
simulating from the prior predictive distribution. Ecosphere 12.

White, E. P., B. J. Enquist, and J. L. Green. 2008. On Estimating the Exponent of Power-Law
Frequency Distributions. Ecology 89:905–912.

White, E. P., S. K. M. Ernest, A. J. Kerkhoff, and B. J. Enquist. 2007. Relationships between
body size and abundance in ecology. Trends in Ecology & Evolution 22:323–330.

Tables

Table 1. Parameter recovery of the same data using two approaches. First,
ten separate models individually recapture known lambda values. Second,
the same ten data sets are estimated in a single hierarchical model. The
true values are compared to the posterior median and 95% Credible
Intervals.

Model	Parameter	True Value	q2.5	q50	q97.5
Separate Models	λ	-2.2	-2.22	-2.15	-2.09
Separate Models	λ	-2.09	-2.13	-2.06	-2
Separate Models	λ	-1.98	-2.05	-1.99	-1.92
Separate Models	λ	-1.87	-1.91	-1.85	-1.8
Separate Models	λ	-1.76	-1.83	-1.78	-1.72
Separate Models	λ	-1.64	-1.7	-1.66	-1.61
Separate Models	λ	-1.53	-1.59	-1.54	-1.5
Separate Models	λ	-1.42	-1.48	-1.44	-1.4
Separate Models	λ	-1.31	-1.31	-1.28	-1.24
Separate Models	λ	-1.2	-1.26	-1.23	-1.2
Single Model	α	-1.7	-1.96	-1.71	-1.5
Single Model	$\sigma_{[group]}$	0.34	0.23	0.36	0.58

Table 2. Slope values from a regression testing the relationship between the ISD exponent and year for IBTS trawl data (Edwards et al. 2020). The values are derived using the Bayesian hierarchical model presented here or from the maximum likelihood approach described in Edwards et al. (2020).

Model	Mean	q2.5	q97.5
Bayesian - one step	-0.001	-0.005	0.002
MLE - two steps	-0.001	-0.005	0.003

Figures

Figure 1. Modeled estimates (median +/- 95% Credible Intervals) of λ using either 10 separate
models or a single model with ten varying intercepts.

Figure 2. Posterior distributions of n = 40 modeled estimates of alpha, beta, and sigma_group for
a linear regression estimating the size spectrum exponent as a function of a continuous predictor.
All data were simulated. Gray densities indicate that the 95% CrI contains the true value, while
black densities indicate the true values fall outside of the CrI. The vertical lines indicate true
values.

Figure 3. Estimates of λ across 11 different sample sizes (ranging from 2 to 2048 individuals)
and three different true λ's (-2, -1.6, -1.2). Ten separate models were fit for each of the 11 sample
sizes. The horizontal lines show the true value of λ.

Figure 4. Regression results from a) Edwards et al. (2020) using maximum likelihood and linear
regression (two steps) and b) the Bayesian model with varying intercepts. In a) the points
represent maximum likelihood estimates calculated separately for each year. In b) they represent
hierarchical varying intercepts calculated from the model.